



1 **Riverine Organic Matter Dynamics in the Headwaters of the China's Yellow River**

2 **under Climate Change**

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13 **Abstract:** Alpine headwater streams play a crucial role in the global carbon cycle and are
14 particularly sensitive to climate change. Riverine organic matter (ROM) mediates the transport
15 and transformation of terrestrial carbon across aquatic systems. However, the response of ROM in
16 headwater streams on the Qinghai–Tibetan Plateau (QTP) to climate change remains poorly
17 understood due to scarce in situ measurements. In this study, we used machine learning models
18 combining satellite data and geographical variables to reconstruct historical variations in Chemical
19 Oxygen Demand (COD, a proxy for ROM) along the Yellow River’s Headstream (YRHS), and to
20 predict future changes under typical Shared Socioeconomic Pathways. The results indicate that
21 COD levels in the midstream region of the YRHS, characterized by greater precipitation, higher
22 soil organic matter, and denser vegetation, were relatively high ($2.73 \pm 1.63 \text{ mg L}^{-1}$) and exhibited
23 an increasing trend ($+0.44 \text{ mg L}^{-1}$) over the past decade. Driven primarily by increasing
24 precipitation and temperature, COD levels are projected to rise in upstream and downstream areas
25 but decline at midstream sites under SSP126, SSP245, and SSP585 by the 2100s. The annual
26 export of COD from the midstream of the YRHS is expected to increase from 62.5 Gg to 81.6 Gg
27 by 2100s due to projected increase in COD concentrations and discharge. Our findings identify the
28 midstream region of the YRHS as a critical and climate-sensitive region for organic matter
29 dynamics. Nevertheless, substantial uncertainties remain in the future ROM changes owing to the
30 complex interactions among precipitation, warming, and their combined effects on carbon cycles
31 in alpine catchments. Therefore, further research is required to improve our understanding of
32 catchment-scale carbon dynamics on the QTP in the context of climate change.

33 **Key words:** Alpine River; Carbon cycle; Climate change; Remote sensing; Machine learning



34 **1 Introduction**

35 Alpine regions represent some of the most vulnerable ecosystems to climate change, experiencing
36 temperature increases nearly twice the global average (Aryal and Pokhrel, 2025; Kotlarski et al.,
37 2023). Although these high-altitude environments cover only a small fraction of the Earth's
38 surface, they provide essential ecosystem services, including freshwater supply, carbon storage,
39 and biodiversity support (Aryal and Pokhrel, 2025; Chen et al., 2022; Hotaling et al., 2017).
40 Among them, the Qinghai–Tibetan Plateau (QTP), often referred to as the “Third Pole”, is the
41 highest and one of the most climate-sensitive regions on Earth (Chen et al., 2022; Wang et al.,
42 2023). In recent decades, the accelerated warming has triggered widespread environmental
43 changes on the QTP, which are expected to significantly alter the sources, forms, and fluxes of
44 carbon exports associated with permafrost degradation (Chen et al., 2022; Hong et al., 2025; Xu et
45 al., 2024). Studies have shown that parts of the QTP have already shifted from carbon sinks to
46 carbon sources and lateral transport of carbon has increased the riverine carbon fluxes on the QTP
47 (L. Li et al., 2025). Riverine organic matter (ROM) plays a pivotal role in the biogeochemical
48 cycles of carbon (Beusen et al., 2022; Giri, 2021; Hu et al., 2020; Regnier et al., 2022) and
49 increase in ROM will pose threats to the water quality of rivers on the QTP. Despite growing
50 concern regarding the ROM and its ecological consequences, our understanding of ROM
51 dynamics across the QTP remains limited. Therefore, there is an urgent need for high spatial
52 resolution and long-term observations of ROM dynamics and their exports on the QTP.

53 Headwater streams account for approximately 61% of total riverine ROM efflux despite
54 representing only 34–38% of total stream surface area (Ran et al., 2021). Over recent decades,
55 climate change has substantially influenced the ROM dynamics of headstreams on the QTP (Yao



56 et al., 2022). Long-term records show that the annual streamflow of headstreams in this region has
57 generally increased over the past six decades (1962–2019) (Z. Zhang et al., 2024). Warming and
58 wetting trends have contributed to higher ROM in the headwaters (Xu et al., 2024). Furthermore,
59 climate-induced changes have intensified sediment erosion and enhanced associated ROM
60 transport processes across the QTP (J. Li et al., 2023; Li et al., 2024; J. Li et al., 2025; Zhao et al.,
61 2023). Nevertheless, fewer than 30% of headstreams across the QTP are consistently monitored
62 (Li et al., 2024), leaving the responses of ROM dynamics in headwaters to climate change poorly
63 constrained.

64 The limited understanding of ROM dynamics in alpine headwaters is largely due to the
65 scarcity of long-term, continuous in-situ measurements, where data collection remains logically
66 challenging. In recent years, advances in remote sensing and machine learning have provided
67 powerful alternatives for addressing these limitations, offering new opportunities to predict water
68 quality and infer nutrient-related parameters across extensive spatial and temporal scales (Adegun
69 et al., 2023; Zeng et al., 2023; Zhi et al., 2024). Satellite-derived indices such as reflectance ratios,
70 the Normalized Difference Chlorophyll Index (NDCI), the Suspended Sediment Index (SSI), etc.,
71 have been increasingly employed to assess riverine organic matter, particularly when integrated
72 with machine learning algorithms (Deng et al., 2024; Liu et al., 2021; Yan et al., 2025a).
73 Consequently, integrating remote sensing data with in-situ measurements and machine learning
74 holds significant potential for improving our understanding of ROM in headwaters and their
75 responses to ongoing climate changing.

76 The headstreams of the Yellow River, located on the eastern QTP, constitute a crucial
77 component of QTP's water resources and carbon cycling systems and have undergone substantial



78 environmental changes in recent decades (Wen et al., 2024). Current study indicated that
79 approximately 35% of the Yellow River's source region serves as a net carbon source, and the
80 lateral transport of carbon accounts for 31% of net ecosystem production in the Yellow River's
81 source region (L. Li et al., 2025). Although numerous studies have been carried out across the
82 broader Yellow River Basin, they have not focused on the historical and future dynamics of ROM
83 in the source region of the Yellow River (Deng et al., 2025; Lan et al., 2025), which is largely
84 attributable to limited monitoring network coverage in this region. Therefore, given the critical
85 role of the Yellow River headwaters in carbon cycling and downstream water quality and its
86 sensitivity to climate change, there is an urgent need to investigate the dynamics of riverine
87 organic matter in the headwaters of the Yellow River under climate change.

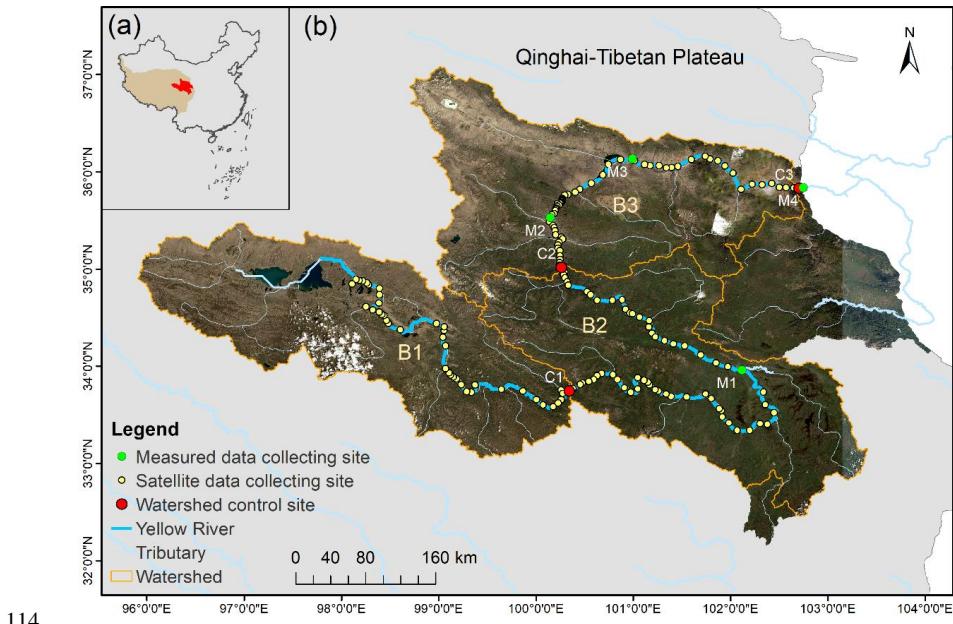
88 According to previous studies, ROM is commonly estimated using Chemical Oxygen
89 Demand (COD) as a proxy, which reflects the amount of oxygen required to oxidize organic
90 carbon in river water (Liu et al., 2023; Wang et al., 2026). Accordingly, we focus on the source
91 region of the Yellow River to develop COD predictive models by combining in situ water quality
92 observations with satellite reflectance data and to assess future changes under the SSP126,
93 SSP245, and SSP585 scenarios. The objectives are to: (1) quantify the decadal variability of COD
94 in the headstreams of the Yellow River and identify its driving factors; (2) predict future changes
95 in COD under projected climate change scenarios; and (3) estimate COD exports and assess their
96 potential impacts on downstream ecosystems. This study aims to provide a comprehensive and
97 scalable understanding of how climate change is reshaping riverine organic matter dynamics
98 within one of the world's most critical headwater regions.

99 **2 Methods and materials**



100 **2.1 Study area**

101 The source region of the Yellow River is situated in the eastern QTP (Fig. 1a), covering
102 approximately 195,000 km², with elevations ranging from 2650 m to 6250 m. The mean annual
103 temperature in this area ranges between -3.5 and 7.5 °C. The annual average precipitation in this
104 area ranges from 420 to 705 mm, with the highest values occurring in the midstream region (Yang
105 et al., 2023). The dominant land cover consists of alpine meadows and alpine steppes (Fig. S1).
106 The area covers extensive permafrost and seasonally frozen ground, together covering more than
107 85% of the region (Z. Li et al., 2025; M. Yang et al., 2025). Under the combined influence of
108 climate warming and anthropogenic disturbances, permafrost in the upper reaches of the Yellow
109 River source region has shown a significant trend of degradation in recent decades (Yang et al.,
110 2023). The headwater region of the Yellow River contains a dense network of tributaries and
111 contributes nearly 49% of the Yellow River's total discharge, making it one of China's most
112 important water sources. However, within the national surface water quality monitoring network,
113 only four sites (M1–M4) are situated at the headwaters of the Yellow River (Fig. 1b).



114

115 Fig. 1 Location of the source region of the Yellow River (a), river network within the study area

116 and data collection sites along the mainstream of the Yellow River (b). B1–B3 represent the up-,
117 mid-, and downstream region of the Yellow River's headstream. A total of 160 monitoring sites
118 span from downstream to upstream, with Site 1 located at the basin outlet.

119

120 **2.2 Data collection**

121 **2.2.1 In-situ Water Quality Data**

122 Riverine COD data for model training were obtained from 107 monitoring sites distributed across
123 the QTP and surrounding regions (Fig. S2), including four sites (M1–M4) positioned along the
124 mainstream of the Yellow River (Fig. 1b). COD data for each site was obtained from the China
125 National Environmental Monitoring Centre (CNEMC; <https://www.cnemc.cn/>) for the period
126 2021–2023. Measurements were performed every four hours (00:00, 04:00, 08:00, 12:00, 16:00,
127 and 20:00). Following the *Environmental Quality Standards for Surface Water* (GB 3838–2002),



128 COD was determined using the acidic potassium permanganate method, in which 50 ml water
129 sample was oxidized by standardized KMnO₄ solution under controlled heating conditions. After
130 the reaction, the residual permanganate was titrated, and COD_{mn} was calculated from the amount
131 of KMnO₄ consumed.

132 **2.2.2 Satellite Reflectance Data**

133 Surface reflectance data for model training was acquired for the 107 monitoring sites across the
134 QTP and its surrounding areas from 2021 to 2023 using the Google Earth Engine (GEE) platform,
135 based on Landsat-8 OLI/TIRS and Landsat-9 OLI-2/TIRS-2 imagery. For historical COD
136 estimation in the source region of the Yellow River, surface reflectance data from Landsat-8
137 OLI/TIRS was also collected for 2014–2024 at 160 sites located at the headwater of the Yellow
138 River (Fig. 1b). The Landsat-8/9 OLI datasets include five visible and near-infrared (VNIR) bands
139 and two shortwave infrared (SWIR) bands (J. Li et al., 2025). Atmospheric correction for
140 Landsat-8 and Landsat-9 imagery was performed using the Land Surface Reflectance Code
141 (LaSRC, version 1.5.0). Pixels affected by clouds, cloud shadows, snow, ice, or non-water
142 surfaces were removed based on the Landsat reflectance quality assessment bands.

143 **2.2.3 Geographical Feature Data**

144 Monthly precipitation and temperature data were obtained from the 1-km *Monthly Precipitation*
145 *Dataset for China* (1901–2024) (Peng, 2025a, 2025b). Surface soil properties, including soil
146 texture, pH, soil organic carbon, and bulk density, were extracted from the *Harmonized World Soil*
147 *Database* (FAO, 2021). Vegetation coverage data were sourced from the *China Regional 250-m*
148 *Fractional Vegetation Cover Dataset* (2000–2024) (Gao et al., 2025). Land use data were derived
149 from the *Land Use Dataset of China* (1980–2015) provided by the Resource and Environmental



150 Science Data Center, Chinese Academy of Sciences (2020). Human Activity Intensity Index (HAI)
151 were obtained from the *Human Activity Intensity Dataset of the Qinghai–Tibet Plateau* (2000–
152 2020) (Liu, 2023). The Human Activity Intensity Index was calculated based on key human
153 activity data, including agricultural and animal husbandry practices, industrial and mining
154 development, urbanization, tourism, major ecological engineering projects, and pollutant
155 discharge. Digital Elevation Model (DEM) data were retrieved from
156 <https://viewfinderpanoramas.org>. Historical and projected Yellow River discharge data were
157 compiled from previously published studies (Long et al., 2024; L. Liu et al., 2025; M. Ma et al.,
158 2023; Wang, 2024).

159 Future monthly temperature and precipitation data with spatial resolution of 1 km in the
160 Yellow River source region were obtained from climate projections generated by the Beijing
161 Climate Center Climate System Model, version 2 (BCC-CSM2-MR) (Hu et al., 2025). This data
162 was generated under three Shared Socioeconomic Pathways (SSPs) aligned with climate scenarios:
163 SSP126 (sustainability-focused, low radiative forcing), SSP245 (intermediate, moderate
164 stabilization), and SSP585 (fossil-fueled development, high radiative forcing).

165 **2.3 Model Development for Historical COD Reconstruction**

166 **2.3.1 Matching Measured COD with Satellite Reflectance**

167 COD data and satellite reflectance data (Landsat-8 OLI/TIRS and Landsat-9 OLI-2/TIRS-2)
168 acquired from 107 sites across the eastern QTP and adjacent regions during 2021–2023 were used
169 for model development (Fig. S2). Since Landsat-8/9 overpasses the study region at approximately
170 10:30 local time, COD measurements obtained at 08:00 or 12:00 on the same day were paired
171 with corresponding satellite scenes. When multiple valid matches were available, only the 12:00



172 COD measurement was retained to ensure temporal consistency. A total of 3048 valid COD–
173 reflectance matchups were obtained for the 107 sites during 2021–2023. These matched datasets
174 covered a broad COD concentration range (0.25–23.2 mg L⁻¹) and were used to develop models
175 for COD prediction.

176 **2.3.2 Input and Prediction Variables of Models**

177 The 3048 valid COD–reflectance matchups were further processed to generate input variables.
178 Input features included raw spectral bands and derived indices, with COD serving as the
179 prediction variable. The raw spectral bands comprised Landsat-8/9 reflectance from Bands 1–7
180 and selected band ratios (SR_{nir}/SR_{red}, SR_{red}/SR_{green} and SR_{red}/SR_{blue}). Additionally, several spectral
181 indices known to be sensitive to organic matter were calculated (Table S1), namely the
182 Normalized Difference Chlorophyll Index (NDCI), Organic Carbon Index (OC_Index),
183 Normalized Suspended Material Index (NSMI), and Hue Angle (Yan et al., 2025b). Longitude and
184 latitude were also incorporated as input variables.

185 **2.3.3 Model Training**

186 The AutoGluon-Tabular algorithm, implemented in the Anaconda3 environment, was employed to
187 train models for COD prediction with Landsat-derived spectral features mentioned above.
188 AutoGluon-Tabular is an open-source automated machine learning (AutoML) framework that
189 builds high-accuracy predictive models for tabular data through multi-layer model ensembling and
190 stacking. AutoGluon’s *TabularPredictor* was configured in regression mode and trained using
191 multiple base learners (e.g., Gradient Boosting Machines, Random Forests, Neural Networks).
192 The framework automatically optimized model selection, hyperparameters, and ensemble weights.
193 The matched dataset (3048 pairs) described in Section 2.3.2 was randomly divided into training



194 (80%) and testing (20%) subsets. A total training time limit of 600s was assigned to ensure
195 sufficient exploration of model configurations.

196 **2.3.4 Model Evaluation**

197 Model performance was evaluated using the negative Root Mean Squared Error (RMSE) for both
198 the validation and test datasets (Eqs. 1–2). Lower (i.e., closer-to-zero) test and validation scores
199 indicated superior model performance. The performance metrics for all models are summarized in
200 Table S2.

201
$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (y_i - \hat{y}_i)^2} \quad (1)$$

202
$$\text{Test /Validation score} = -RMSE \quad (2)$$

203 where, N denotes the size of the observations and y_i and \hat{y}_i denote the in-situ values and
204 satellite-retrieved values, respectively.

205 The importance of different input variables affecting the prediction of COD was also
206 assessed based on the importance scores (Eq.3).

207
$$Importance(f_i) = M(D) - M(D_{shuffled(f_i)}) \quad (3)$$

208 where, $M(\cdot)$, model performance metric (e.g., accuracy, log-loss, RMSE); D , original evaluation
209 dataset; $D_{shuffled(f_i)}$, same dataset, but with feature f_i randomly permuted; $Importance(\hat{f})$, Drop in
210 performance when f_i is destroyed.

211 **2.3.5 Reconstruct of Historical COD and Calculation of COD Changes**

212 The model exhibiting the highest test and validation scores was applied to reconstruct COD
213 concentrations at 160 sites located in the headwater region of the Yellow River from 2014 to 2024
214 using Landsat-8 surface reflectance data. To evaluate model applicability and reliability, the
215 reconstructed COD concentrations were compared with in situ measurements at two representative



216 headwater sites. Model accuracy was assessed for the two sites using the Mean Absolute
217 Percentage Difference (MAPD), RMSE, and Bias, following the approach of Yan et al. (2025).
218 Historical COD changes were assessed by comparing averages from 2014–2015 and 2023–2024.

219 **2.4 Prediction of Future COD and Its Changes in the Headstream of the Yellow River**

220 To predict future COD levels at 160 sites in the headwaters of the Yellow River, the machine
221 learning approach was also employed to model the relationships between geographical variables
222 and the satellite-based estimates of COD levels. Using the Anaconda3 platform, soil texture (sand,
223 silt, clay), bulk density, soil organic carbon, soil pH, altitude, human activity intensity, vegetation
224 cover, land use, and ten-year averages of annual mean precipitation and temperature for each site
225 during 2014-2024 were used as input features, while the ten-year average COD concentration for
226 each site during 2014-2024 served as the target variable. The best-performing model, selected
227 based on test and validation scores, was then used to predict future COD concentrations,
228 incorporating both static variables (e.g., soil properties, vegetation cover, land use) and dynamic
229 variables (future annual mean precipitation and temperature), depending on data availability.

230 In order to predict the future COD using the trained model, future monthly temperature and
231 precipitation data was collected from Hu et al., (2025). Future annual mean precipitation and
232 temperature were then calculated for 2025-2040, 2041-2070, and 2071-2100, representing near-,
233 mid-, and long-term horizons. Future COD changes in the Yellow River source region for
234 2025-2040, 2041-2070, and 2071-2100 under SSP126, SSP245, and SSP585 scenarios were
235 calculated relative to the mean values from 2014–2024.

236 **2.5 Calculation of COD flux**

237 The COD flux in the source region of the Yellow River was calculated using Eq. 4:



238
$$F = C_i \times Q_i / 10 \quad (4)$$

239 where, F (Gg), nutrient export; C_i (mg/L), COD concentration; Q_i (10^8 m 3), river discharge.

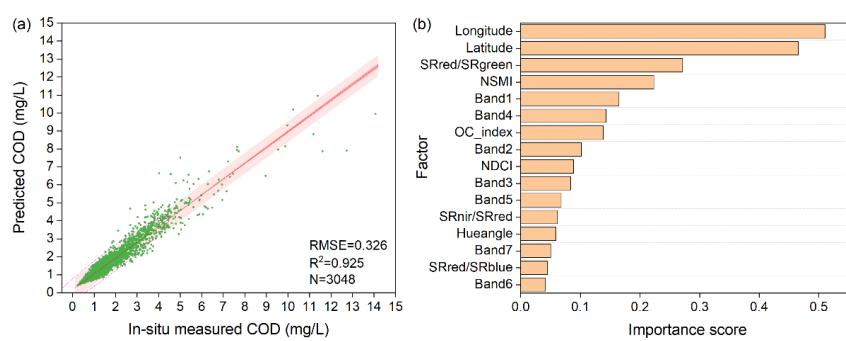
240 **3 Results**

241 **3.1 Performance of developed models**

242 **3.1.1 Predicting Historical COD Using Landsat Reflectance Data**

243 A total of 24 models were developed and evaluated on the Anaconda3 platform to predict COD
244 levels from Landsat reflectance data, with test and validation scores summarized in Table S2. The
245 LightGBMLarge model demonstrated the highest predictive accuracy for COD across the QTP
246 and surrounding regions, achieving test and validation scores of -0.22325 and -0.66913 ,
247 respectively. Predicted versus in-situ measured COD levels for 2021–2023 based on training
248 dataset are shown in Fig. 2a, with RMSE and R^2 of 0.326 and 0.925, respectively, indicating
249 strong predictive performance.

250 The relative importance of input variables in the LightGBMLarge model was quantified
251 using importance scores (Fig. 2b). The five most influential predictors were longitude, latitude,
252 SR_{Red}/SR_{Green}, NSMI, and Band 1. Based on its performance, the LightGBMLarge model was
253 selected to estimate historical COD levels in the source region of the Yellow River.



254

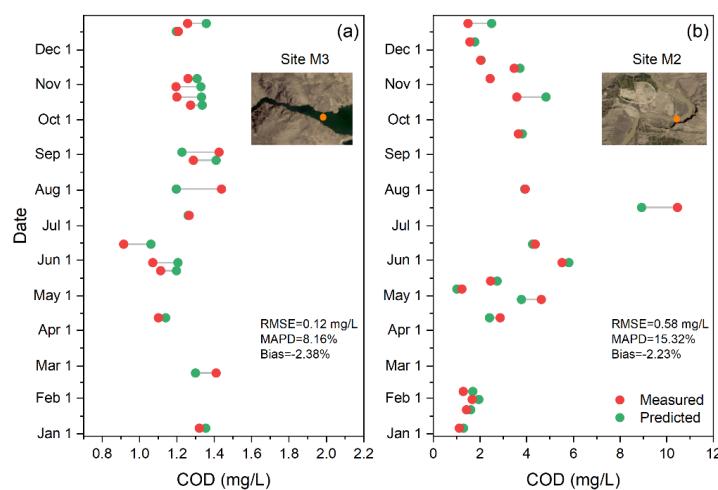
255 Fig.2 Model performance for COD prediction based on training set using Landsat 8/9 reflectance



256 data (a), and importance of input variables in COD prediction (b). Bands 1-7 are surface
257 reflectance data produced by the Landsat 8 OLI/TIRS sensors; SR_{nir}/SR_{red}, SR_{red}/SR_{green} and
258 SR_{red}/SR_{blue} are band ratios; NSMI is Normalized Suspended Material Index; OC_Index is
259 Organic Carbon Index; NDCI is Normalized Difference Chlorophyll Index; Hue Angle is used to
260 reflect watercolor and water quality.

261

262 To assess model reliability in the study area, two representative sites (M2 and M3) in the
263 Yellow River headwaters were used to compare modeled and in-situ COD time series (Fig. 3). The
264 results indicate that the model can accurately capture both spatial and temporal variability of
265 riverine COD in the headstreams of the Yellow River. At site M3, RMSE, MAPD, and Bias were
266 0.12 mg L⁻¹, 8.16%, and -2.38%, respectively, and RMSE, MAPD, and Bias at site M2 were 0.58
267 mg L⁻¹, 15.32%, and -2.23%, respectively.



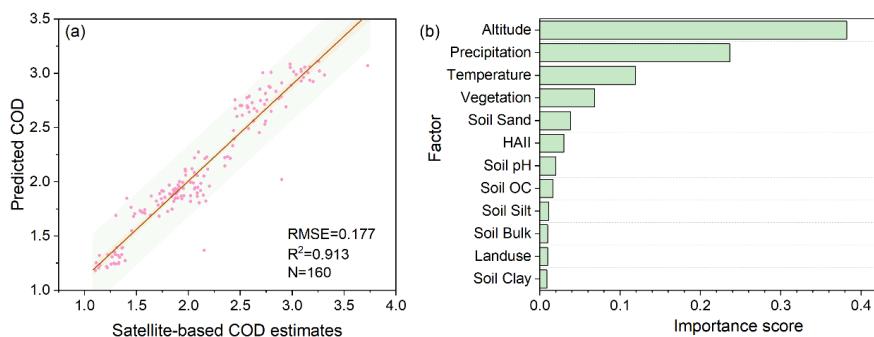
268
269 Fig.3 Comparison between predicted and measured COD values in 2023 at sites M3 (a) and M2 (b)
270 within the Yellow River source region. For sites M2 and M3, 24 and 20 matched data points were
271 collected, respectively.



272

273 **3.1.2 Prediction of Future COD Using Geographical Features**

274 Models for predicting future COD based on geographical features were also developed using the
275 Anaconda3 platform. A total of 36 models were evaluated (Table S3), and the LightGBMLarge
276 model exhibited the highest performance, with test and validation scores of -0.081 and -0.264 ,
277 respectively. Predicted versus satellite-based estimates of average COD levels based on the
278 training dataset in the Yellow River source region are presented in Fig. 4a, achieving an R^2 of
279 0.913 and an RMSE of 0.177, indicating strong predictive capability of this model. Analysis of
280 feature importance revealed that altitude, precipitation, temperature, and vegetation contributed
281 most significantly to COD variability in the study area (Fig. 4b).



282

283 Fig.4 Performance of the developed model for COD prediction based on the training set using
284 geographical feature data (a), and importance of variables in the predictive model (b). HAI is
285 Human Activity Intensity Index; Soil OC represents soil organic carbon; Soil Bulk is soil bulk
286 density; Soil Sand, Clay and Silt are soil texture.

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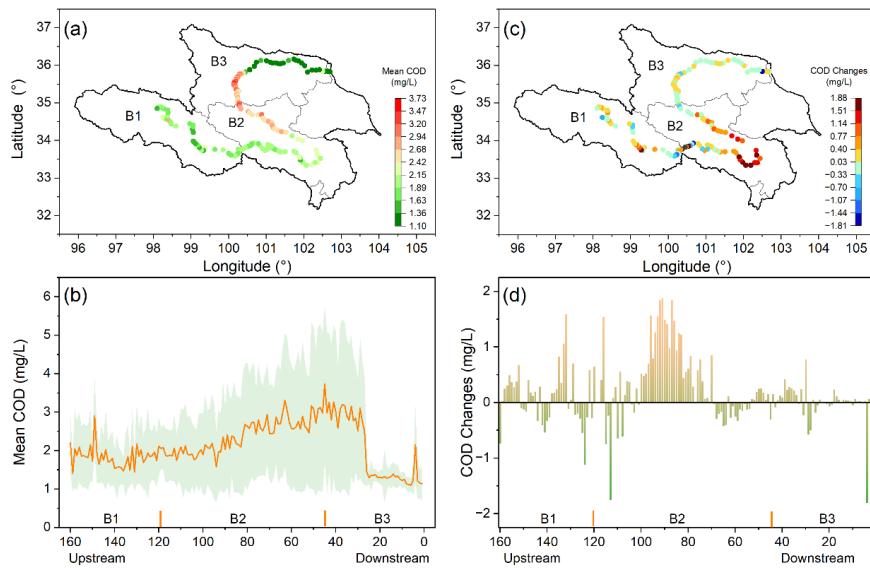
288 **3.2 Spatiotemporal Changes of COD in the Headstream of the Yellow River**

289 The developed model estimated an annual mean COD concentration of $2.14 \pm 1.06 \text{ mg L}^{-1}$ in the
15



290 Yellow River source region from 2014 to 2024, with pronounced spatial variability. In the
291 upstream area, the mean COD was $1.89 \pm 0.81 \text{ mg L}^{-1}$, while higher concentrations were observed
292 in the mid-downstream region ($2.73 \pm 1.63 \text{ mg L}^{-1}$). The downstream area, which includes large
293 reservoirs such as Longyangxia and Lijiaxia, exhibited significantly lower COD levels ($1.54 \pm$
294 0.50 mg L^{-1}) (Fig. 5a,b).

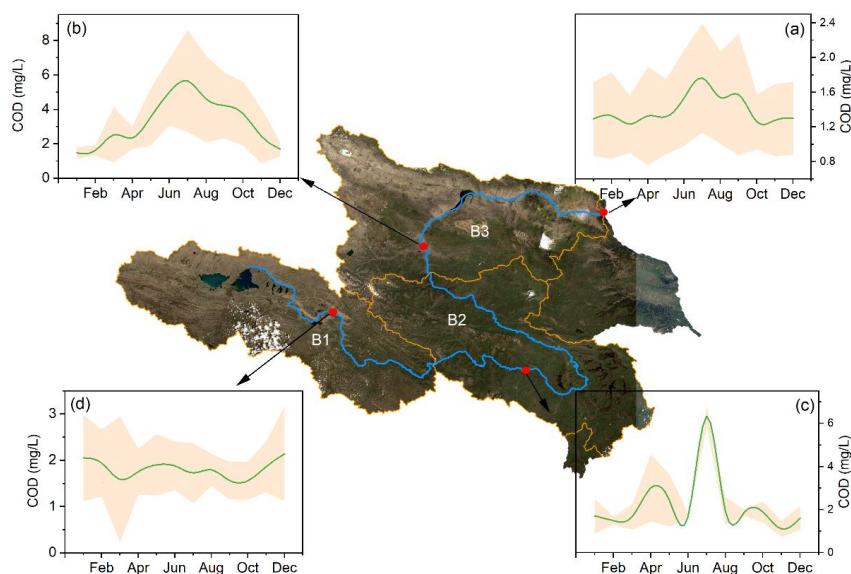
295 Over the past decade, the midstream region showed a clear increasing trend in COD, with an
296 average rise of 0.441 mg L^{-1} (Fig. 5c,d). In contrast, the upstream area exhibited both increasing
297 and decreasing trends, with a mean change of 0.099 mg L^{-1} , whereas COD concentrations in the
298 downstream region remained relatively stable (mean change: -0.002 mg L^{-1}).



299
300 Fig. 5 Spatial variation of mean COD and its change between 2014–2024 in the Yellow River
301 source region. Changes were assessed by comparing averages from 2014–2015 and 2023–2024.
302
303 Seasonal variations of COD were also evident during our study. Average monthly COD



304 concentrations ranged from 0.8 to 8 mg L⁻¹ (Fig. 6). The downstream site maintained low (0.8-2.4
305 mg L⁻¹) and showed higher COD levels in summer (Fig. 6a), while the midstream site exhibited a
306 pronounced increase from March, peaking in July (~8 mg L⁻¹), followed by a gradual decline (Fig.
307 6b). The upper–midstream site displayed two distinct peaks in late spring and summer, suggesting
308 episodic organic inputs associated with ice melt and rainfall (Fig. 6c). Upstream sites remained
309 low to moderate (1-3 mg L⁻¹) with minimal seasonal variation (Fig. 6d).



310
311 Fig.6 Averaged monthly variations of COD during 2014–2024 within the Yellow River source
312 region.

313

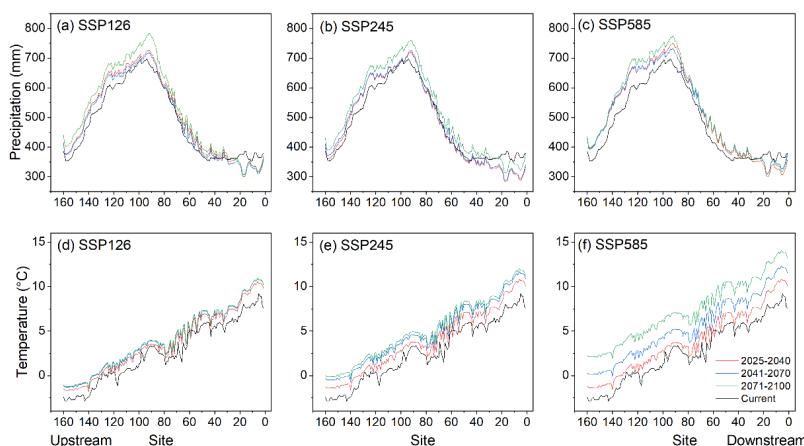
314 3.3 Future Climate and COD Changes in the Source Region of the Yellow River

315 Across all scenarios and periods, precipitation exhibits a consistent spatial pattern, increasing from
316 upstream reaches to a midstream maximum and subsequently decreasing toward downstream



317 sections (Fig. 7a-c). This pattern is preserved under all SSPs, although the magnitude of
318 precipitation varied among scenarios and time slices. Relative to current conditions, precipitation
319 increases slightly in the midstream region under SSP126 and SSP245, particularly during the mid-
320 and late-century periods. Under SSP585, precipitation displays enhanced increase, with higher
321 midstream values during 2041–2070 and 2071–2100 and lower precipitation toward downstream
322 reaches.

323 Compared with current conditions, future temperature rises across the entire study area (Fig.
324 7d-f), with the magnitude of warming increasing over time and from SSP126 to SSP585. Under
325 SSP126, temperature increases are relatively moderate, whereas SSP245 shows stronger warming,
326 especially during 2071–2100. The largest temperature increases occurred under SSP585, with
327 pronounced downstream variability during the late-century period, resulting in an amplified
328 longitudinal temperature gradient.

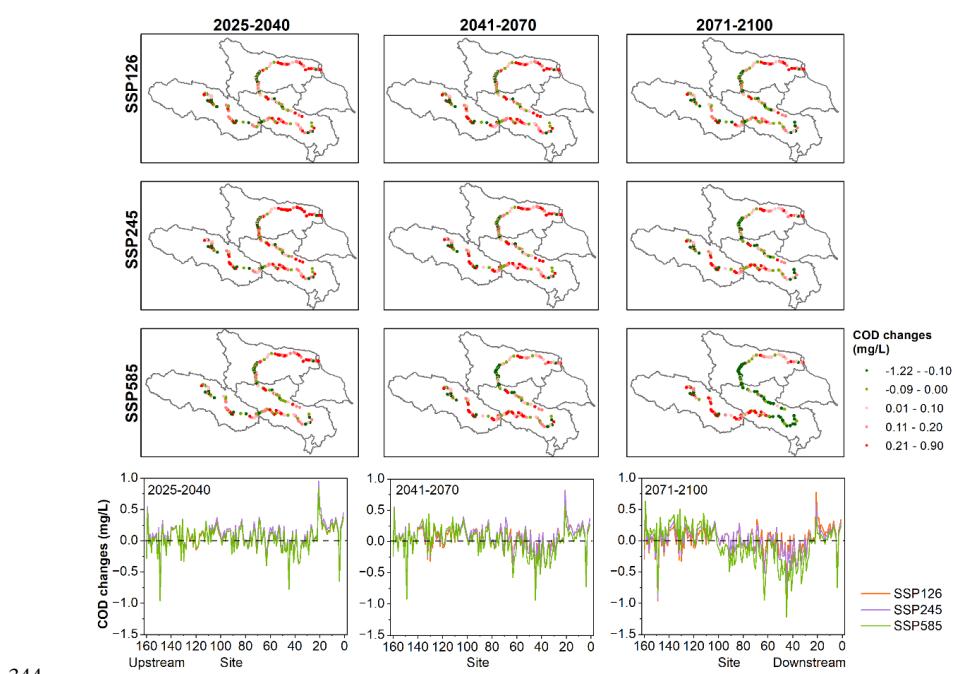


329
330 Fig.7 Projected precipitation and temperature variations for 2025-2040, 2041-2070, and
331 2071-2100 under SSP126, SSP245, and SSP585 scenarios based on CMIP6 data.

332



333 Projected changes in COD across the study area show pronounced spatial and temporal
334 variability under all three SSP scenarios (Fig. 8). During the near-term period (2025–2040), COD
335 changes are generally modest, with most monitoring sites exhibiting values close to zero. In the
336 mid-term period (2041–2070), the COD changes show an overall increase, with mean value of
337 0.05 mg L⁻¹. Specifically, midstream sites more frequently experience decreases in COD, whereas
338 upstream and downstream regions display a mixture of small increases and decreases. By the
339 late-century period (2071–2100), the average change in COD is -0.01 mg L⁻¹. COD shows a
340 pronounced decrease trend in the midstream area (-0.12 mg L⁻¹) under SSP585. Upstream and
341 downstream areas show a higher occurrence of positive COD changes. Overall, COD shows a
342 slight increasing trend across the three scenarios over different periods, with 877 positive changes
343 (61%) and 563 negative changes (39%).



345 Fig.8 Projected spatial variation of COD changes in the Yellow River source region for 2025-2040,
19



346 2041–2070, and 2071–2100 under SSP126, SSP245, and SSP585 scenarios. COD changes were

347 calculated relative to the mean values from 2014–2024.

348

349 **3.4 Historical and Future COD Exports in the Source Region of the Yellow River**

350 Based on the annual average discharge at control site C3 (1987–2021), approximately 28.1 Gg of

351 COD was exported from the entire study area. In contrast, the annual mean COD export at site M2

352 was 62.5 Gg, indicating that approximately 34 Gg of COD was either deposited or decomposed

353 between C3 and M2 along the downstream of the Yellow River’s source region. The export of

354 COD primarily occurred during the wet season (June–October) (Fig. 9a), which contributes

355 substantially to the annual COD export from the source region (Fig. 9b). Additionally, the

356 observed increase in COD export between sites M1 and M2 highlights the midstream region as a

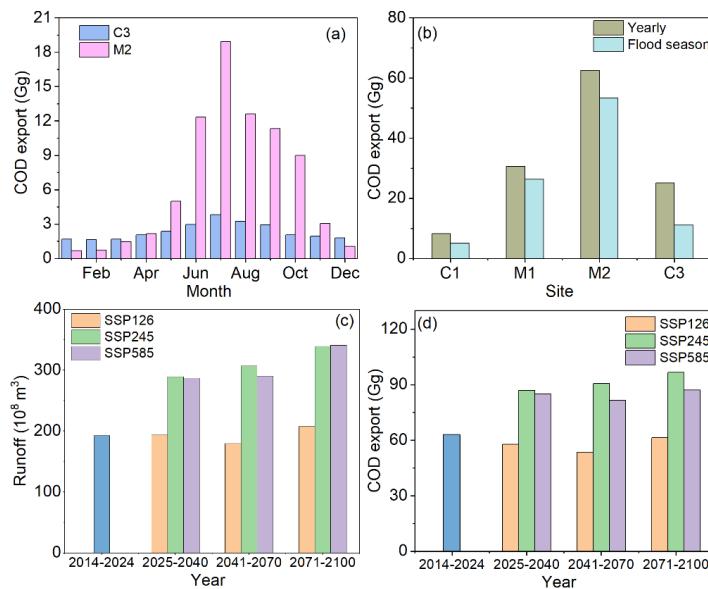
357 major source of COD (Fig. 9b).

358 Affected by the increase in the projected runoff (Fig. 9c), future projections at site M2

359 indicate a slight declining trend in COD export from 2025 to 2100 under SSP126 scenarios, and

360 an increase trend under SSP245 and SSP585 scenarios (Fig. 9d). By 2100, the average COD

361 export at M2 under different scenarios is expected to increase from 62.5 to 81.6 Gg.



362

363 Fig.9 Monthly average COD export during 2014-2024 at sites M2 and C3 (a); comparison of
364 annual average COD export and flood-season COD export during 2014-2024 among sub-basins
365 (b); projected annual runoff (c) and COD export at M2 for 2025-2040, 2041-2070, and 2071-2100
366 under SSP126, SSP245, and SSP585 scenarios (d). C1-C3 are the control sites for sub-basins B1-
367 B3, respectively. Historical annual and seasonal runoff was obtained from L. Liu et al. (2025) and
368 Wang (2024). The projected runoff under SSP126 was derived from M. Ma et al. (2023), while the
369 runoff under SSP245 and SSP585 was calculated using data from Long et al. (2024). Different
370 sources of runoff data may result in discrepancies in the predicted COD exports across scenarios.

371

372 **4 Discussion**

373 **4.1 Factors Driving COD Dynamics in the Headstream of the Yellow River**

374 **4.1.1 Spatial Variations**

375 Primary sources of riverine organic carbon include losses from soils, plant litter, root exudates in



376 grasslands, and permafrost thaw via surface runoff (Yan et al., 2023). In this study, precipitation,
377 temperature, vegetation, and soil texture were identified as key drivers of COD variation (Fig. 4b).
378 Notably, the overall soil organic carbon, vegetation cover, fine particle content, and precipitation
379 are higher in the midstream region of the headwater (Fig. S3), providing abundant carbon sources
380 and facilitating transport to nearby rivers, resulting in elevated COD levels and fluxes in the
381 mid-downstream areas (Figs. 5a, 9a,b). Additionally, bare land in the mid-downstream areas of the
382 headstream may also contribute to the soil erosion and the associated organic carbon losses (Fig.
383 S1). L. Li (2025) identified the higher lateral fluxes of river organic matter in the midstream area
384 of the Yellow River's source region. Consequently, the midstream area represents a critical source
385 of organic carbon for the Yellow River headwaters.

386 In contrast, COD levels decreased markedly in the downstream region (Figs. 5a,b). This
387 decline is likely driven by damming, as three large reservoirs, Longyangxia, Laxiwa, and Lijiaxia
388 have been constructed in the downstream headwater region since the 1970s. Reservoirs increase
389 water residence time, promoting deposition and decomposition of organic carbon, greenhouse gas
390 emissions, and uptake by aquatic plants (Maavara et al., 2017, 2020). Therefore, the deposition or
391 decomposition of organic matter along the downstream areas may contribute to the decline of
392 COD levels. The deposition of riverine organic matter can be inferred from the decrease in COD
393 fluxes at the outlet (C3) of the study area relative to Site M2 (Figs. 9a, b). L. Li et al. (2025) also
394 reported negative dissolved organic matter fluxes in the main channels of the Yellow River's
395 headstreams. Additionally, the decomposition of the deposited organic matter is evidenced by CO₂
396 outgassing from the river water. Previous study has found that CO₂ outgassing from the cascade
397 reservoirs of Yellow River's source region was estimated at $131.02 \pm 156.77 \text{ mmol m}^{-2} \text{ d}^{-1}$ in dry



398 seasons and 466.10 ± 366.67 mmol m⁻² d⁻¹ in flood seasons (Wang et al., 2022). Therefore, these
399 observations suggest that the sedimentation and degradation of organic matter contributes to lower
400 COD concentrations in the downstream area.

401 The COD concentration remained relatively stable along the upper to middle reaches of the
402 Yellow River's headwaters. This region is characterized by permafrost and seasonally frozen
403 ground, with low precipitation and limited runoff (Yang et al., 2023). Surface runoff in this area
404 primarily originates from precipitation, snowmelt, and glacier melt (G. Li et al., 2025).
405 Consequently, due to the relatively uniform and stable runoff and land cover, COD levels
406 exhibited a homogeneous spatial distribution along the upper to middle reaches of the Yellow
407 River's headwaters.

408 **4.1.2 Temporal Variations**

409 Climate-driven changes in runoff and temperature strongly influence catchment-scale carbon
410 dynamics, altering riverine organic matter fluxes globally (Costa et al., 2023). Over the past
411 decade, air temperature in the source region of the Yellow River has shown an increasing trend,
412 with a basin-averaged increase of 0.67 °C, and the precipitation also presented an over increase
413 trend though decrease slightly after 2019 (Fig. S4). The study area is particularly sensitive to
414 warming-induced permafrost thaw, glacier melting, soil erosion, and intensified human activities
415 over recent decades (Chen et al., 2022; Deng et al., 2025; Lan et al., 2025), which modify ROM
416 sources and lateral transport between terrestrial and aquatic systems, affecting ROM levels and
417 fluxes in rivers (Chen et al., 2022; Wang et al., 2020, 2025). Previous studies highlighted that
418 permafrost thaw was projected to release 129.39 ± 21.02 Tg soil organic carbon annually (Lan et
419 al., 2025). Glacier melting contributes an estimated 0.19 Tg C yr⁻¹ as dissolved organic carbon



420 (Chen et al., 2022). Intensified rainfall has increased soil erosion by 42% over the past four
421 decades, resulting in an average soil carbon loss of $4 \pm 1.2 \text{ Tg C yr}^{-1}$ (Deng et al., 2025).

422 Therefore, affected by the increased temperature and precipitation, COD concentrations in
423 the source region of the Yellow River exhibited an overall increasing trend over the past decade,
424 with an average rise of 0.186 mg L^{-1} (Figs. 5c,d). Especially the midstream region experienced the
425 most pronounced increase (0.441 mg L^{-1}). The higher precipitation combined with increased
426 runoff likely enhance organic matter leaching from grasslands, contributing to the rise of riverine
427 COD in the midstream area (Chen et al., 2025; Yang et al., 2023).

428 In addition to climate effects, growing human activities such as land use changes, grazing,
429 and construction on the QTP also influence carbon mobilization (Chen et al., 2022).
430 Human-induced land-use changes, grassland degradation, construction activities, and overgrazing
431 can lead to organic matter losses in the study area. In addition, the construction and urbanization
432 of natural and agricultural ecosystems may further accelerate organic matter losses within the
433 watershed. All these human activities may contribute to the rising of COD in the headstream of the
434 Yellow River.

435 **4.2 Future Changes in COD under Climate Change**

436 Rapid climate change in alpine regions is altering carbon cycling in both terrestrial and aquatic
437 ecosystems (Chen et al., 2022). As carbon dynamics are highly sensitive to climatic variations,
438 warming and wetting trends are expected to influence the stability of riverine organic matter in
439 alpine catchments. Numerous studies indicate that precipitation and temperature on the QTP are
440 projected to increase throughout this century, particularly in high-elevation areas (Karim et al.,
441 2025; Meng et al., 2023). Our projections also suggest an overall increase in precipitation,



442 especially in the up-midstream area of the Yellow River source region (Fig. 7), and significant
443 warming is projected in the whole study area by 2100, especially under SSP 245 and SSP585.

444 Climate-driven changes are expected to induce the largest variability in COD concentrations
445 in the midstream region (Fig. 8). COD levels are projected to increase in most parts of the source
446 region of the Yellow River, and annual average COD exports will increase from 62.5 Gg to 81.6
447 Gg (Fig. 9d). ROM dynamics are jointly regulated by organic matter supply and hydrological
448 processes within catchments (Li et al., 2022; Wu et al., 2024). Increased precipitation can enhance
449 surface runoff, leading to losses of organic matter from the catchment, even though COD
450 concentrations may be diluted as discharge continues to increase (Li et al., 2022; Wu et al., 2024).

451 The projected warming in the headwater area of the Yellow River will lead to the continuing
452 permafrost degradation, promoting the loss of organic matter within the catchment and leading to
453 the increase of the COD level in the headstreams (Z. Li et al., 2025). In addition, damming
454 prolongs water residence time, promoting organic matter deposition in cascade reservoirs in the
455 downstream area of the Yellow River's source region. Under warm conditions, the degradation of
456 deposited organic matter in the bottom of the reservoir may be enhanced, releasing additional
457 organic carbon from reservoir sediments and contributing to the rise of COD levels in the
458 downstream area (Battin et al., 2023; Vachon et al., 2021).

459 However, the extent to which organic matter reaches rivers depends on hydrological
460 processes and connectivity between the source region and river channels. The characteristics of
461 rainfall events may shift the ROM response to storms from transport-limited to supply-limited,
462 thereby reducing ROM levels (Li et al., 2022). Consequently, the projected increase in
463 precipitation by the 2100s is expected to lead to a decrease in COD levels within the midstream



464 areas (Fig. 8). In addition, permafrost thaw induced by the warming can also alter runoff pathways,
465 promoting subsurface flow and potentially reducing surface discharge to rivers (Yang et al., 2023).
466 Increased vegetation cover can also reduce transport of organic matter through reducing soil
467 erosion and surface runoff (Liu et al., 2024; W. Ma et al., 2023). Therefore, these factors may
468 jointly contribute to the decrease of riverine organic matter at some specific sites on the
469 headstream of the Yellow River.

470 **4.3 Uncertainty in Riverine Organic Matter Dynamics on the QTP**

471 Understanding riverine organic matter dynamics is essential for assessing aquatic ecosystem
472 feedback to climate change. Ongoing warming and wetting on the QTP are expected to alter
473 carbon stability, increasing uncertainty in watershed organic matter losses (Bai et al., 2025). For
474 instance, increased precipitation may enhance surface runoff (Li et al., 2022), however, rising
475 temperatures can also decrease catchment runoff by enhancing evaporation. The influence of
476 hydrological processes on ROM dynamics depends on the characteristics of the catchment (Li et
477 al., 2022). Previous studies have predicted both increasing and decreasing streamflow trends on
478 the QTP under future climate scenarios (G. Li et al., 2025; San et al., 2025; Yang et al., 2023).
479 Increased streamflow may elevate COD in rivers, or dilute COD due to higher streamflow.
480 Conversely, reduced streamflow may decrease organic matter transport to rivers or concentrate
481 COD in river water.

482 Additionally, warming-induced permafrost thaw may accelerate topsoil organic matter
483 mobilization, stimulate microbial activity, and promote soil organic matter decomposition,
484 releasing more dissolved organic matter (Z. Li et al., 2025). However, the frozen ground
485 degradation can also result in the decreased surface runoff by increasing the subsurface runoff



486 (Yang et al., 2023). Vegetation greening on the QTP may both enhance soil carbon storage and
487 reduce organic matter losses to river networks through mitigating soil erosion (Deng et al., 2025).

488 Therefore, climate change can further differentially affect terrestrial carbon stocks and carbon
489 losses, introducing additional complexity and uncertainty in catchment-scale organic matter
490 sources and transport (Chen et al., 2022; Liu et al., 2024; W. Ma et al., 2023).

491 In-stream carbon biogeochemical processes are also highly complex under changing climatic
492 conditions. Organic matter decomposition depends on temperature, redox conditions, hydrology,
493 microbial activity, and organic matter characteristics (Battin et al., 2023). The climate change is
494 projected to change the hydrological processes and the associated biogeochemical processes,
495 further altering the carbon cycling in the alpine rivers. However, interactions among these factors
496 are not fully understood in alpine rivers, making predictions of alpine river metabolism and carbon
497 budget under climate change uncertain (L. Li et al., 2025).

498 Anthropogenic activities on the QTP further contribute to uncertainty. Warming and wetting
499 may expand farming, increasing soil erosion and carbon losses (Chen et al., 2022). Future dam
500 construction alters organic matter transport by promoting deposition, influencing greenhouse gas
501 emissions, and facilitating organic matter degradation (Battin et al., 2023). Conversely, ecological
502 restoration, such as afforestation, can increase soil organic matter accumulation and litterfall
503 (Chen et al., 2022; Yan et al., 2023), leading to increase in organic matter and the COD levels in
504 the nearby headwaters of the Yellow River.

505 In this study, we project a general increase in levels and fluxes of COD based on modeled
506 temperature, precipitation, and discharge. However, future climatic trends may deviate from
507 current projections, and the predicted temperature and precipitation may not represent the actual



508 future patterns. Furthermore, our developed model does not incorporate changes in land covers,
509 soil carbon stocks, hydrological processes, or in-stream carbon biogeochemistry, all of which may
510 significantly influence riverine carbon cycling. Therefore, future research should integrate these
511 factors to develop more precise, process-based models and improve understanding of carbon
512 dynamics in alpine river catchments.

513 **5 Conclusions**

514 In this study, LightGBM-Large models were developed to reconstruct the historical variations of
515 COD in the source region of the Yellow River and to predict its future changes. Our findings
516 identify the midstream region of the Yellow River's headstream, with higher COD levels ($2.73 \pm$
517 1.63 mg L^{-1}), is sensitive to climate change and exhibited an increasing trend of COD ($+0.44 \text{ mg$
518 L^{-1}) in the past decade. Precipitation and temperature were identified as the key factors
519 influencing COD dynamics in the headwaters of the Yellow River. Future projections suggest that
520 increased precipitation and temperature may lead to an overall rise of COD levels, with decrease
521 in some specific sites. The export of organic matter from the headwaters of the Yellow River is
522 also projected to increase based on the anticipated increase in COD levels and discharge. However,
523 the limited spatial and temporal resolution of Landsat data and lack of geophysical data, still poses
524 challenges for accurately capturing rapid variations in riverine COD in the source region of the
525 Yellow River. Therefore, future research should focus on developing more robust models by
526 integrating higher-quality satellite observations and incorporating a broader range of
527 environmental variables.

528 **Data availability**

529 The data will be available upon request.



530 **Author contributions**

531 XW: Conceptualization, Methodology, Writing—original draft, Funding acquisition; JW:
532 Methodology, Data curation, Review and editing; TR: Investigation, Formal analysis, Validation;
533 JZ: Validation, Conceptualization, Review and editing.

534 **Competing interests**

535 The authors declare that they have no conflict of interest.

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