



1 **Soil contamination and soil-mediated human health risks associated with**
2 **household coal combustion in residential areas of Zavkhan Province,**
3 **Mongolia**

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5 Enkhchimeg Battsengel¹, Davaadorj Davaasuren², Baasansuren Gankhurel³,
6 Keisuke Fukushi³, Altankhuyag Davaajav⁴, Munkhzul Badamsuren²,
7 Sonomdagva Chonokhuu¹

8
9 ¹Department of Environment and Forest Engineering, School of Engineering
10 and Technology, National University of Mongolia, Ulaanbaatar 14201,
11 Mongolia

12 ²Department of Geography, School of Arts and Sciences, National University
13 of Mongolia, Ulaanbaatar 14201, Mongolia

14 ³Institute of Nature and Environmental Technology, Kanazawa University,
15 Kanazawa, Ishikawa 920-1192, Japan

16 ⁴School of Business and Information Technology, Zavkhan Branch, National
17 University of Mongolia, Uliastai 21000, Mongolia

18 Corresponding author: Sonomdagva Chonokhuu; ch_sonomdagva@num.edu.mn

19 **Abstract**

20 Soil contamination by heavy metals represents a growing environmental and public
21 health concern in cold–dry rural settlements where coal-based household heating
22 remains dominant. This study investigates how coal combustion alters soil element
23 dynamics and associated human health risks by applying a process-oriented,
24 integrated soil system assessment in a residential area of Uliastai city, western
25 Mongolia.

26 Surface soils (0–10 cm) from 38 sites were analyzed using ICP-OES and ICP-MS to
27 determine major and trace element concentrations. Multivariate statistical analysis
28 (principal component analysis, PCA) was combined with contamination indices
29 (enrichment factor and geo-accumulation index) and human health risk assessment to
30 explicitly link contamination sources, transport pathways, soil retention processes,
31 and potential human exposure.



32 Results reveal a clear separation between anthropogenically influenced metals (As,
33 Pb, Cd, Zn, and Cu) and elements predominantly controlled by geogenic background
34 conditions (Cr, Co, and Ni). Very high to extreme enrichment and geo-accumulation
35 levels for As, Pb, Zn, Cd, and Cu indicate substantial anthropogenic alteration of
36 surface soil metal pools. Comparison of soil, coal, and ash compositions identifies
37 coal combustion ash as the primary source of metal enrichment, acting as a
38 concentrated reservoir that is redistributed to soils via atmospheric deposition and
39 surface processes. Human health risk assessment shows that the most enriched metals,
40 particularly As and Pb, dominate both non-carcinogenic and carcinogenic risks, with
41 inhalation and ingestion pathways contributing most strongly to potential exposure.

42 The findings demonstrate that soil contamination in Uliastai reflects systemic changes
43 in soil functioning driven by household energy practices rather than isolated
44 concentration exceedances. By integrating source identification, contamination
45 intensity, and health risk within a unified soil system framework, this study provides
46 mechanistic insight into soil–human interactions and offers a transferable approach
47 for assessing soil impacts in coal-dependent rural environments.

48 Keywords: soil system impact; heavy metals; coal combustion; enrichment factor;
49 geo-accumulation index; human health risk; Mongolia

50 **1. Introduction**

51 Soil contamination has become an increasingly significant environmental and public
52 health concern in rural regions of Mongolia. In many countryside settlements, the
53 combined impacts of coal-based household heating, insufficient waste management
54 infrastructure, and the expansion of cultivated land have contributed to the
55 accumulation of pollutants in soils (Lkhagvasuren et al., 2019; Batjargal and Kim,
56 2020). Such contamination not only degrades soil quality but also poses potential
57 risks to human health, highlighting the need for integrated assessments that link
58 environmental contamination with human exposure pathways (Li et al., 2014).

59 Uliastai city, one of the oldest permanent settlements in western Mongolia with a
60 history spanning more than 300 years, provides a unique setting for evaluating the
61 long-term influence of anthropogenic activities and natural processes on soil quality.
62 The basin-like topography of the area promotes the accumulation of geochemical



63 materials transported by the Bogd and Chigestei rivers originating from the
64 Otgontenger mountain range. These geomorphological and hydrological conditions
65 favor the deposition and redistribution of sediments and associated trace elements
66 within the urban environment (Dorjgotov, 1976).

67 In addition to natural factors, land-use practices have played an important role in
68 shaping soil contamination patterns in Uliastai. Riverbank agricultural lands
69 surrounding the city have experienced decades of irrigation development and periodic
70 application of agrochemicals, which may contribute to the gradual buildup of
71 contaminants in surface soils (Purevsuren et al., 2017). The interaction between
72 historical land use, river dynamics, and urban expansion makes Uliastai an important
73 case study for understanding soil contamination processes in rural Mongolian
74 settlements.

75 Uliastai was first established in 1733 as a military administrative center during the
76 Manchu Qing Dynasty and subsequently evolved into a commercial and
77 administrative hub. These stages of urban development have influenced the sources
78 and spatial distribution of soil contaminants, including heavy metals such as Pb, Cu,
79 and Zn, as well as dust and household-derived pollutants within residential areas
80 (Dorjgotov, 1976).

81 As of 2024, Zavkhan Province has a population of approximately 71 800, of which
82 about 22.5 % reside in Uliastai city (Zavkhan Province Statistical Yearbook, 2022).
83 The city comprises 4 871 households, and recent statistics indicate that approximately
84 83.5 % of households rely on coal for domestic heating. This heavy dependence on
85 coal represents a major source of both atmospheric and soil contamination.

86 Coal is extensively used in household stoves and centralized heating boilers in
87 Uliastai. Limited heating network coverage, high connection costs, substantial heat
88 loss, and inadequate thermal insulation of buildings all contribute to the continued
89 reliance on solid fuel combustion (Meteorology and Environmental Monitoring
90 Center, 2022). Coal combustion has been shown to enrich heavy metal concentrations
91 several-fold relative to natural geological materials, facilitating the accumulation of
92 these elements in air, soil, and water and increasing potential risks to human health
93 (Batjargal and Kim, 2020; Frankel et al., 2025).



94 The Uliastai Meteorology and Environmental Monitoring Center conducts biennial
95 monitoring of selected heavy metals (Cd, Pb, Zn, Co, and Sr) at 16 soil sampling
96 locations within the city. Monitoring results from 2011 to 2022 indicate elevated
97 concentrations of strontium and zinc in residential areas compared to background
98 sites, suggesting persistent environmental contamination.

99 Despite the long history of settlement, complex land-use patterns, and relatively high
100 population density, comprehensive assessments of soil contamination and associated
101 human health and ecological risks in Uliastai remain limited.

102 Therefore, the objectives of this study are to (i) quantify the concentrations of heavy
103 metals in surface soils from residential areas of Uliastai city and compare them with
104 regional background values and national soil quality standards, (ii) characterize the
105 contamination intensity and enrichment levels of individual metals using enrichment
106 factor and geo-accumulation index, (iii) identify potential anthropogenic and geogenic
107 sources of soil contamination through principal component analysis, (iv) evaluate the
108 contribution of coal ash and coal combustion-related activities to soil heavy metal
109 accumulation, and (v) assess associated non-carcinogenic and carcinogenic human
110 health risks within an integrated multivariate and multi-criteria assessment
111 framework.

112 **2. Methodology**

113 This study assessed the level of heavy metal contamination in the urban soils of
114 Uliastai, Zavkhan Province, by collecting and analyzing surface soil samples from a
115 total of 38 sites. Sampling locations were selected based on population density, land-
116 use intensity, and settlement structure, and were arranged in a grid pattern with an
117 interval of 500 meters between points. To enhance the reliability and comparability of
118 the results, 16 existing monitoring points from the Zavkhan Meteorology and
119 Environmental Monitoring Center (MEMC) were also included. Figure 1 shows the
120 study area and the distribution of sampling points.

121 **2.1 Field Survey and Soil Sampling**

122 The field survey was conducted between 20 and 27 March 2022 across the residential
123 zones of six khorooos (subdistricts) of Uliastai. Using the 16 regular monitoring



124 stations operated biennially by MEMC, a total of 38 soil samples were collected
125 (Figure 1).

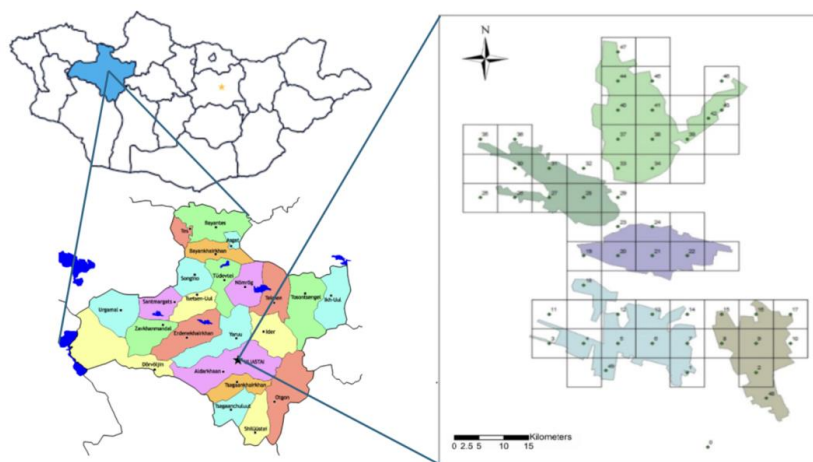


Fig1. Location of sampling points in Uliastai, Zavkhan Province

126

127 Soil sampling procedures followed the ISO 18400-104:2018 standard, using an
128 envelope sampling method to collect samples from a depth of 0–10 cm.

129 Approximately 300 g of soil was collected from each sampling point, placed in
130 polyethylene bags, securely sealed, and georeferenced using a Garmin GPS device.

131 Prior to laboratory transport, soil samples were pre-processed at the laboratory of the
132 National University of Mongolia (NUM). Samples were cleaned by removing stones,
133 roots, and coarse debris, then air-dried at room temperature (20°C) for three days. The
134 dried samples were sieved through a 2 mm mesh to achieve homogeneity and
135 subsequently stored in airtight polyethylene bags for one week to stabilize moisture
136 content. After pre-treatment, samples were shipped to Kanazawa University in Japan
137 for further laboratory analysis.

138 2.2 Chemical Analysis of Soil

139 Chemical analysis of the soil samples was conducted at the Geochemistry Laboratory,
140 Institute of Environmental Technology, Kanazawa University, Japan. Heavy metal
141 concentrations were determined using inductively coupled plasma mass spectrometry
142 (ICP-MS; iCAP RQ, Thermo Fisher Scientific, Waltham, MA, USA) and inductively



143 coupled plasma optical emission spectrometry (ICP-OES; ES-710, Varian Inc., Palo
144 Alto, CA, USA).

145 Sample preparation for instrumental analysis followed a standardized analytical
146 protocol consisting of five sequential steps. The procedure was adapted from a
147 previously published method (Battsengel Enkhchimeg, 2020), with minor
148 modifications to accommodate the characteristics of the studied soils.

149 Soil samples (0.05 g) were subjected to sequential acid digestion using 3 mL of 60%
150 HNO₃ and 3 mL of 48% HF at 120 °C until dryness, followed by a second digestion
151 with 3 mL of 30% HCl. The digested residue was extracted with 10 mL of 0.6 vol%
152 HNO₃ and shaken for 24 h. The extracts were then filtered through a 0.20 µm
153 membrane filter and diluted 50-fold with HNO₃ prior to elemental analysis by ICP-
154 MS and ICP-OES.

155 **2.3 Contamination Index Assessment**

156 To assess the level of heavy metal contamination in soils, widely used international
157 pollution indices such as the Enrichment Factor (EF) and the Geo-accumulation Index
158 (I_{geo}) were applied. These indices are effective tools for distinguishing
159 anthropogenic contributions from natural background concentrations and for
160 evaluating the extent to which metal levels deviate from their geochemical baseline
161 values (Müller, 1969; Abdullah, M.I.C., 2000).

162 The I_{geo} is calculated using the following equation (Equation 1).

$$163 \quad \text{Log}_2 = \left(\frac{C_n}{1.5 \times B_n} \right)$$

164 The Enrichment Factor (EF) is used to assess the degree of contamination and identify
165 the influence of anthropogenic sources on the concentration of specific elements in
166 soil. In contrast, the Geo-accumulation Index (I_{geo}) allows for a more detailed
167 evaluation of anthropogenic impacts by comparing current concentrations with natural
168 background levels.

169 In EF calculations, elements such as aluminum (Al) or iron (Fe), which are stable and
170 strongly associated with soil mineral composition, are typically used as normalizing
171 elements. In this study, iron (Fe) was selected as the reference element for the



172 following reasons: it is widely distributed in nature, occurs consistently in soil mineral
 173 structures, it is sensitive to anthropogenic disturbances, it is strongly associated with
 174 clay minerals, making Fe a widely recommended normalizer in international studies
 175 (Barbieri, M. et al., 2016; Taylor, S.R et al., 1964).

176 The Enrichment Factor is calculated using the following equation (Equation 2).

177
$$EF = \frac{\left(\frac{M}{Fe}\right)_{sample}}{\left(\frac{M}{Fe}\right)_{background}}$$

178 For the EF computation, the Fe background concentration was derived from the
 179 control sampling site within the study area, and a value of 1125 mg/kg was used. The
 180 categories applied for evaluating contamination severity based on enrichment factor
 181 (EF) and geo-accumulation index (I_{geo}) are presented in **Table 1**.

182 Table 1. Category of EF (enrichment factor) and I_{geo} (geo-accumulation).

EF		I_{geo}	
EF < 2	no enrichment	$0 < I_{geo}$	uncontaminated
EF = 3–5	moderate enrichment	$0 < I_{geo} < 1$	uncontaminated to moderately uncontaminated
EF = 5–10	moderately severe enrichment	$1 < I_{geo} < 2$	moderately contaminated
EF = 10–25	severe enrichment	$2 < I_{geo} < 3$	moderately to heavily contaminated
EF = 25–50	very severe enrichment	$3 < I_{geo} < 4$	heavily contaminated
EF > 50	extremely severe enrichment	$4 < I_{geo} < 5$	heavily to extremely contaminated
		$I_{geo} > 5$	extremely contaminated

183

184 **2.4 Soil–Coal–Ash Comparative Study Using PCA**

185 Principal Component Analysis (PCA) is a multivariate statistical technique widely
 186 used to reduce the dimensionality of complex datasets while preserving the maximum
 187 amount of variance. In environmental studies, PCA is particularly effective for
 188 identifying relationships among heavy metals (e.g., Cr, Cu, Pb, Zn, Cd), determining
 189 contamination patterns, and distinguishing between natural (geogenic) and
 190 anthropogenic sources.



191 In this study, PCA was employed to compare the heavy metal concentrations
192 measured in: Surface soils from residential areas of Uliastai, Coal samples from the
193 “Mogoin Gol” coal deposit, commonly used by residents, and Coal ash produced
194 during combustion.

195 The objective was to detect common patterns, correlations, and possible source
196 linkages among soil, coal, and ash samples. PCA was used to: Identify whether heavy
197 metals found in soil are associated with coal or coal ash; Determine clustering
198 patterns across soil, coal, and ash samples; Assess the degree of metal enrichment due
199 to coal combustion; Differentiate geogenic and anthropogenic contributions; Reveal
200 dominant sources of contamination using statistical evidence, etc.

201 **2.5 Human health risk assessment**

202 Residents living in the urban area of Uliastai, Zavkhan Province, were evaluated for
203 potential human health risks using the United States Environmental Protection
204 Agency (US EPA) risk assessment framework. For each heavy metal detected in soil,
205 exposure pathways were quantified based on the Average Daily Intake (ADI),
206 calculated using Equations 3-5.

$$ADI_{Ingestion} = \frac{C_x \text{ IngR} \times EF \times ED \times CF}{BW \times AT} \quad (3)$$

$$ADI_{dermalcontact} = \frac{C_x SA \times FE \times ABS \times EF \times ED \times CF}{BW \times AT} \quad (4)$$

$$ADI_{inhalation} = \frac{C_x \text{ InhR} \times EF \times ED \times PEF}{BW \times AT} \quad (5)$$

207 Here, ADI represents the average daily intake of a heavy metal from soil (mg/kg·day),
208 and C denotes the concentration of the metal in soil (mg/kg). The assessment
209 considered exposure through three primary pathways: (i) ingestion of soil particles,
210 (ii) inhalation of resuspended soil dust, and (iii) dermal absorption through skin
211 contact. Table 2 presents the exposure parameters used in the human health risk
212 assessment, differentiated between children and adults.



213 **Table 2** presents the exposure parameters used for the human health risk assessment,
 214 categorized for both children and adults.

215 Table 2. Parameters of risk assessment.in soil

Parameters	Adult	Children	Unit
ADI, average daily intake	-	-	[mg/kg day]
IngR, soil ingestion rate	100	200	[mg/day]
EF, exposure frequency	350	350	[day/year]
ED, exposure duration	30	6	[year]
BW, body weight	70	15	[kg]
SF, skin area exposed to soil contact	5700	2800	[cm ²]
AF, soil to skin adherence factor	0.07	0.2	[kg/cm day]
ABS, contact factor	0.1	0.1	none
InhR, inhalation rate	15	10	[m ³ /day]
PEF, particle emission factor	1.36 × 10 ⁹	1.36 × 10 ⁹	[m ³ /kg]
AT, average time non-carcinogenic	10,950	2190	[days]
AT, average time carcinogenic	25,550	25,550	[days]
CR, Conversion factor	1 × 10 ⁻⁶	1 × 10 ⁻⁶	[mg/kg]
FE, Dermal exposure ratio	0.61	0.61	-

216
 217 Non-carcinogenic risk was assessed using the Hazard Index (HI), which is
 218 calculated based on the Average Daily Intake (ADI) of each element and its
 219 corresponding Reference Dose (RfD). In contrast, carcinogenic risk (CR) was
 220 evaluated using the ADI together with the Slope Factor (SF), provided by
 221 USEPA (1997).

222 The equations used to calculate non-carcinogenic and carcinogenic risks are presented
 223 in Equations (6)–(8). The non-carcinogenic risk of each metal was expressed as the
 224 Hazard Quotient (HQ). To obtain the total carcinogenic risk, the risks associated with
 225 each heavy metal through the three exposure pathways were calculated according to
 226 Equation (8).

227 For risk assessment, the acceptable threshold for non-carcinogenic effects (HI) was
 228 set at 1, whereas the acceptable limit for carcinogenic risk was set at 10⁻⁴ (Ma et al.,
 229 2018).

230
$$HQ = \frac{ADI}{RfD} \quad (6)$$

231
$$HI = \sum HQ = \sum \frac{ADI_i}{RfDi} \quad (7)$$



232 $Risk (total) = \sum Risk (inh) + Risk (ing) + Risk (dermal)$ (8)

233 Among the 26 elements analyzed in this study, only eight had available RfD values;
 234 therefore, these elements were included in the non-carcinogenic risk assessment (HI).
 235 For carcinogenic risk assessment, SF values were available for only four elements,
 236 and thus carcinogenic risk (CR) was calculated for these four elements.

237 The RfD and SF values used in this study are presented in Table 3.

238 Table 3. RfD (reference dose) and SF (slope factor) values

Elements		RfD [mg/kg day]		
		Ingestion	Dermal	Inhalation
1	Cr	$3 \times 10^{-3(a)}$	$3 \times 10^{-3(a)}$	$2.86 \times 10^{-5(a)}$
2	Pb	$1.4 \times 10^{-3(a)}$	$5.25 \times 10^{-4(a)}$	$3.52 \times 10^{-3(a)}$
3	Zn	$3 \times 10^{-1(c)}$	$6 \times 10^{-2(c)}$	$3 \times 10^{-1(a)}$
4	Cd	$3 \times 10^{-1(a)}$	$2.3 \times 10^{-5(a)}$	$1 \times 10^{-5(a)}$
5	As	$1 \times 10^{-4(a)}$	$1.23 \times 10^{-4(a)}$	$1.23 \times 10^{-4(a)}$
6	Co	$2 \times 10^{-2(a)}$	NA	NA
7	Cu	$4 \times 10^{-2(a)}$	$1.2 \times 10^{-2(a)}$	$4 \times 10^{-2(a)}$
8	Mo	$5 \times 10^{-3(a)}$	NA	NA

Elements		SF [mg/kg day]		
		Ingestion	Dermal	Inhalation
1	Cr	$5 \times 10^{-1(a)}$	20 ^(a)	42 ^(a)
2	Pb	$8.5 \times 10^{-3(a)}$	NA	$4.2 \times 10^{-2(c)}$
3	Cd	NA	NA	6.3 ^(a)
4	As	1.5 ^(a)	3.66 ^(a)	$4.3 \times 10^{-3(a)}$

239 NA represents data not available. ^a USEPA, 2007; ^b Huang, 2017; and ^c Kamunda,
 240 2016 .

241

242

243 **3. Results**

244 **3.1. Concentrations of heavy metals in surface soils**

245 Concentrations of 11 heavy metals (Cr, Co, Cu, Zn, As, Se, Cd, Pb, V, Sr, and Ni)
 246 measured in 38 surface soil samples (0–10 cm) exhibited substantial spatial variability
 247 across the residential areas of Uliastai. Mean concentrations of all analyzed elements
 248 were below the national soil quality limits (MNS 5850:2019); however, maximum
 249 values of Zn, Cu, Pb, and Co showed pronounced local enrichments, indicating the
 250 presence of contamination hotspots rather than uniform area-wide pollution. This
 251 contrast between mean and extreme values highlights the importance of spatially



252 explicit assessment for identifying site-specific soil degradation that may be masked
 253 in aggregated statistics.

254 Descriptive statistics, including minimum, maximum, mean, median, standard
 255 deviation, and range values, are summarized in Table 4, together with the
 256 corresponding threshold values defined in the Mongolian national soil quality
 257 standard (MNS 5850:2019).

258 Table 4. Concentrations of heavy metals in surface soils

№	Elements	Min	Max	Mean	Median	Std.	Range	Mongolian Standard (MNS) (5850:2019)		
		(mg/ kg)	(mg/ kg)		Std. Dev.	Dev.		Acceptable limit	Hazardous	Dangerous
1	Cr	0.8	13.4	2.8	1.8	2.7	12.6	150	400	1500
2	Co	0.7	5.9	1.9	1.7	0.9	5.1	50	500	1000
3	Cu	0.2	155	11	4.3	26	155	100	500	1000
4	Zn	8.7	2757	147	38	456	2749	300	600	1000
5	As	0.2	29	2.8	0.7	5.4	29	20	50	100
6	Se	0.1	3.5	0.5	0.4	0.6	3.4	10	50	100
7	Cd	0.04	1.6	0.2	0.2	0.3	1.5	3	10	20
8	Pb	0.1	90	12.0	6.7	17.3	90.4	100	500	1200
9	V	2.1	89	9.9	5.6	14.7	86.8	150	600	1000
10	Sr	0.2	4.2	1.0	0.7	0.9	4.0	800	3000	6000
11	Ni	1.60	14.20	1.9	1.7	0.9	5.1	150	1000	1800

259 Elevated concentrations of Zn, Cu, Pb, and As in soils collected near ger districts,
 260 central heating facilities, and densely populated residential zones. These areas are
 261 characterized by intensive household heating, frequent ash handling, and limited
 262 surface sealing, all of which facilitate direct deposition and accumulation of
 263 combustion-derived particulates in surface soils. In contrast, samples from peripheral
 264 and less densely inhabited areas showed lower and more homogeneous
 265 concentrations, reflecting weaker anthropogenic pressure and stronger influence of
 266 natural background conditions.

267 Elements such as Cr, Co, and Ni displayed relatively narrow concentration ranges and
 268 weak spatial gradients. Their distributions showed limited association with settlement
 269 structure or heating facilities, suggesting dominant control by parent material and
 270 regional lithology. This spatial divergence between anthropogenically enriched metals



271 and geogenically controlled elements provides initial evidence for multiple
272 controlling processes within the soil system.

273 3.2. Multivariate structure of soil contamination

274 Principal Component Analysis (PCA) extracted two principal components explaining
275 the majority of variance in the dataset, indicating that a limited number of dominant
276 processes govern soil metal variability. The first component (PC1) was characterized
277 by strong positive loadings of As, Pb, Cd, Zn, and Cu, indicating a coherent
278 assemblage of elements with similar spatial behavior and shared controlling factors.
279 These elements are commonly associated with coal combustion residues and
280 settlement-related activities, suggesting a dominant anthropogenic signal.

281 The second component (PC2) was dominated by Cr, Co, and Ni, reflecting a
282 contrasting geochemical association. These elements are typically linked to soil parent
283 material and mineralogical composition, and their grouping along PC2 indicates a
284 predominantly lithogenic origin. The clear separation between PC1 and PC2
285 demonstrates that anthropogenic and geogenic controls on soil chemistry can be
286 statistically distinguished despite spatial overlap within the settlement.

287 Figure 2 shows Principal Component Analysis (PCA) biplot of soil heavy metals. PC1
288 is dominated by As, Pb, Cd, Zn, and Cu, representing an anthropogenic component
289 associated with coal combustion and settlement activities. PC2 is dominated by Cr,
290 Co, and Ni, reflecting geogenic control by natural background conditions.

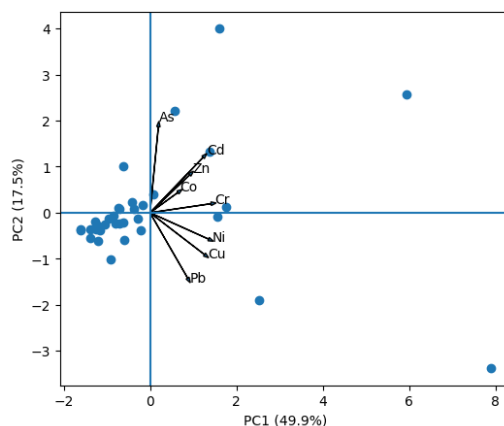




Figure 2. Multivariate structure of soil contamination

291 Sample score distributions further showed that sites with high PC1 scores correspond
292 to locations with elevated concentrations of anthropogenically enriched metals,
293 particularly near heating facilities and densely populated ger areas. Samples clustered
294 along PC2 represent areas where soil chemistry remains largely governed by natural
295 background signatures. This multivariate structure confirms that soil contamination in
296 the study area results from the superposition of anthropogenic inputs on an underlying
297 geogenic baseline.

298 3.3. Contamination intensity based on pollution indices

299 3.3.1. Enrichment factor (EF)

300

301 Enrichment factor analysis revealed very high to extreme EF values for As, Pb, Zn,
302 Cd, and Cu at the majority of sampling sites, indicating substantial enrichment
303 relative to background conditions (Fig. 3). In contrast, Cr, Co, and Ni exhibited low to
304 moderate EF values, suggesting minimal anthropogenic influence. The spatial
305 distribution of EF values highlights distinct enrichment hotspots coinciding with
306 residential heating zones and areas influenced by coal ash deposition.

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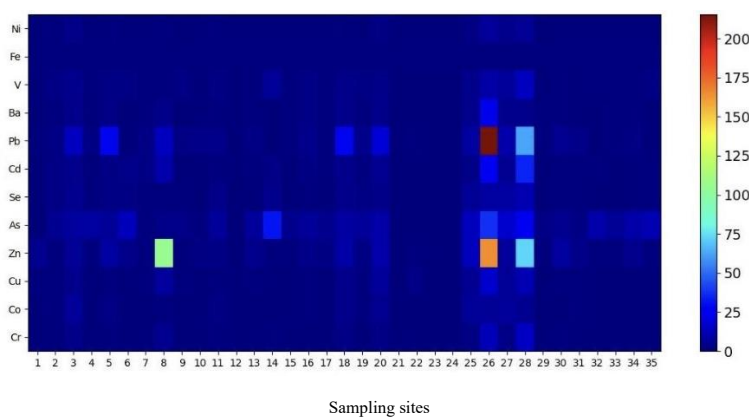


Figure 3. Spatial distribution of enrichment factor (EF) values for selected heavy metals

308



309 The contamination patterns identified using EF are further examined using the geo-
310 accumulation index (Igeo) to evaluate contamination intensity and to cross-validate
311 the observed anthropogenic enrichment.

312

313 3.3.2. Geo-accumulation index (Igeo)

314

315 The Igeo values classified As, Pb, Zn, Cd, and Cu predominantly within the heavily to
316 extremely contaminated categories, whereas Cr, Co, and Ni were generally classified
317 as uncontaminated to moderately contaminated (Fig. 4). Elevated Igeo values for As,
318 Pb, Zn, Cd, and Cu were observed at multiple sampling sites, indicating pronounced
319 spatial variability and the presence of localized contamination hotspots. The clear
320 contrast between anthropogenically influenced and geogenically controlled elements,
321 together with the strong agreement between EF and Igeo results, confirms the
322 robustness of the contamination intensity assessment and reinforces the differentiation
323 between anthropogenic and geogenic element groups.

324

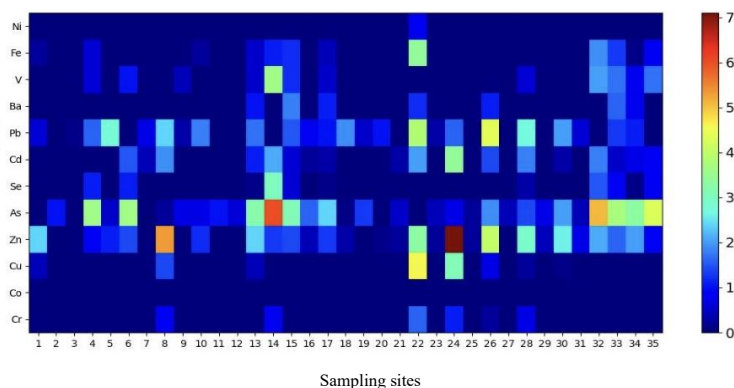


Figure 4. Classification of soil contamination intensity based on the geo-accumulation index (Igeo) for selected heavy metals.

325 3.4 Human Health Risk Assessment Results

326 3.4.1 Non-carcinogenic Risk Assessment

327 The results of the non-carcinogenic risk assessment for Cr, Co, Cu, Zn, As, Cd, and
328 Pb are summarized in Table 5. Hazard index (HI) values were calculated for children
329 and adults through oral ingestion, dermal contact, and inhalation exposure pathways.



330 Table 5. Non-carcinogenic risk (hazard index, HI) for selected heavy metals

Risk	Pathways	Elements						
		Cr	Co	Cu	Zn	As	Cd	Pb
Noncarcinogenic	Child							
	Oral	1.2E-02	1.2E-03	3.2E-03	6.2E-03	3.5E-01	1.1E-05	3.7E-02
	Dermal	2.1E-03		1.8E-03	5.3E-03	4.2E-03	2.0E-03	3.2E-04
	Inhalation	2.6E-05		2.5E-02	1.8E-02	2.8E-04	1.3E-04	2.1E-03
	Adult							
	Oral	1.3E-03	1.3E-04	3.5E-04	6.7E-04	3.8E-02	1.1E-06	3.9E-03
	Dermal	3.2E-04		2.8E-04	8.1E-04	3.2E-03	1.5E-03	9.9E-04
	Inhalation	1.9E-06		7.6E-02	5.7E-02	8.8E-04	3.9E-04	6.6E-03

331 For both population groups, children exhibited higher non-carcinogenic risk values
 332 than adults across all exposure pathways. Among the analyzed elements, arsenic (As)
 333 and lead (Pb) were the dominant contributors to total non-carcinogenic risk,
 334 accounting for the majority of cumulative HI values. In contrast, Cr, Co, Cu, Zn, and
 335 Cd contributed relatively lower proportions to the overall risk.

336 Across exposure pathways, oral ingestion of contaminated soil and dust was the
 337 primary contributor to non-carcinogenic risk, followed by inhalation of resuspended
 338 particulates, while dermal contact contributed comparatively little to total HI values.
 339 Elevated HI values were spatially associated with sampling sites characterized by
 340 higher metal concentrations in surface soils, indicating a close linkage between
 341 localized soil contamination and potential human exposure.

342 3.4.2 Carcinogenic Risk Assessment

343 The results of the carcinogenic risk assessment for Cr, As, Cd, and Pb are presented in
 344 Table 6. For both children and adults, inhalation was the dominant exposure pathway,
 345 contributing substantially higher carcinogenic risk values than oral ingestion or
 346 dermal contact. Among the analyzed elements, Cr and As exhibited the highest
 347 carcinogenic risk values, particularly via inhalation, whereas Cd and Pb contributed
 348 comparatively lower risks.

349 Table 6. Carcinogenic risk (CR) for selected heavy metals

Risk	Pathways	Elements			
		Cr	As	Cd	Pb



Carcinogenic	Child				
	Oral	1.6E-06	4.5E-06		1.1E-07
	Dermal	1.1E-05	1.9E-06		
	Inhalation	1.3E+01	1.3E-02	1.7E+00	5.5E-01
	Adult				
	Oral	8.4E-07	2.4E-06		6.0E-08
	Dermal	8.1E-06	1.4E-06		
	Inhalation	7.0E+00	7.0E-03	9.2E-01	3.0E-01

350 Carcinogenic risk values estimated for children were consistently higher than those
 351 for adults across all pathways, reflecting differences in exposure parameters. Oral and
 352 dermal pathways generally resulted in carcinogenic risk values below the acceptable
 353 benchmark, while inhalation-related risks for Cr and As exceeded the threshold (10^{-4})
 354 at several sites, indicating potential long-term health concern associated with soil-
 355 derived airborne particulates.

356

357

358 4. Discussion

359 The integrated interpretation of multivariate analysis, contamination indices, and
 360 human health risk assessment provides a comprehensive understanding of soil-related
 361 environmental risks in Uliastai. PCA clearly distinguishes between anthropogenic and
 362 geogenic sources, with As, Pb, Cd, Zn, and Cu strongly associated with the dominant
 363 anthropogenic component. This pattern is consistent with EF and Igeo results, which
 364 demonstrate severe enrichment and pollution for the same elements.

365 Coal combustion residues, ash deposition, and settlement-related activities are
 366 identified as the primary drivers of soil contamination. These findings are consistent
 367 with studies from other cold-climate regions where solid fuel combustion is a
 368 dominant source of heavy metals in residential soils. In contrast, Cr, Co, and Ni are
 369 largely controlled by natural geological background conditions and contribute
 370 minimally to overall health risk, highlighting the importance of distinguishing
 371 between anthropogenic and geogenic contributions.



372 The source apportionment and contamination patterns identified in this study are
373 consistent with findings reported from other cold-climate and coal-dependent regions
374 worldwide. Previous studies have demonstrated that coal combustion and ash
375 deposition are dominant sources of arsenic, lead, cadmium, zinc, and copper in
376 residential soils, particularly in areas with intensive household heating and limited
377 pollution control (e.g., Liu et al., 2016; Chen et al., 2018). Similar enrichment patterns
378 and anthropogenic signatures have been observed in urban and peri-urban soils in
379 northern China and Central Asia, where solid fuel use remains prevalent.

380 The predominance of geogenic control for chromium, cobalt, and nickel observed in
381 this study is also in agreement with international soil contamination studies, which
382 report that these elements are often closely associated with parent material and
383 lithogenic background rather than direct anthropogenic inputs (Alloway, 2013;
384 Reimann and Garrett, 2005). The separation between anthropogenic and geogenic
385 components identified through PCA in the present study reflects a common
386 multivariate pattern reported in soil geochemistry literature.

387 In terms of health implications, the elevated non-carcinogenic and carcinogenic risks
388 associated with arsenic and lead are comparable to those reported in residential soil
389 studies from other developing and transition economies, where soil exposure
390 pathways contribute significantly to overall human health risk (Li et al., 2014; Hu et
391 al., 2017). These similarities highlight that soil contamination driven by household
392 energy practices represents a broader environmental and public health challenge
393 beyond the local context of Mongolia.

394 The integration of contamination indices with human health risk assessment reveals a
395 direct linkage between pollution severity and potential adverse health outcomes.
396 Metals exhibiting the highest enrichment and pollution levels are also responsible for
397 elevated non-carcinogenic and carcinogenic risks, emphasizing the relevance of soil
398 contamination for public health in residential environments.

399 Spatial heterogeneity in contamination and risk patterns reflects uneven distribution of
400 anthropogenic emissions and site-specific accumulation processes. Localized
401 contamination hotspots underscore the need for targeted mitigation measures rather
402 than uniform soil management strategies. From a policy perspective, the findings



403 highlight the importance of promoting cleaner household energy alternatives,
404 improving coal ash management, and implementing focused soil remediation in high-
405 risk areas.

406 Overall, this study demonstrates that soil contamination in Uliastai is a multifaceted
407 environmental and public health challenge. The integrated multivariate and multi-
408 criteria framework applied here offers a robust and transferable approach for soil risk
409 assessment in rural settlements facing similar environmental pressures.

410 Despite the comprehensive nature of the integrated assessment, several limitations
411 should be acknowledged. First, the health risk assessment relied on USEPA default
412 exposure parameters, which may not fully represent local lifestyle characteristics, soil
413 ingestion rates, or behavioral patterns in rural Mongolian communities. Second, the
414 study focused on surface soils collected during a single sampling campaign, and
415 therefore temporal variability in contamination levels could not be assessed.

416 In addition, while PCA, EF, and Igeo provide robust indicators of contamination
417 sources and intensity, uncertainties remain regarding the relative contributions of
418 overlapping anthropogenic activities, such as traffic emissions and small-scale
419 industrial operations. Future studies incorporating seasonal sampling, site-specific
420 exposure parameters, and isotopic or source-tracing techniques would further refine
421 source attribution and risk estimation.

422 **Conclusion**

423 This study applied an integrated multivariate and multi-criteria environmental and
424 human health risk assessment framework to evaluate soil heavy metal contamination
425 in a rural settlement of Mongolia. By combining principal component analysis,
426 enrichment factor, geo-accumulation index, and human health risk assessment, the
427 study provides a coherent and systematic interpretation linking contamination sources,
428 pollution intensity, and potential human health impacts.

429 The results demonstrate that arsenic, lead, zinc, cadmium, and copper are the
430 dominant contaminants in the investigated soils. These metals exhibit strong
431 anthropogenic signatures, characterized by very high to extreme enrichment and
432 heavy to extreme pollution levels. The strong agreement between multivariate source



433 identification and contamination indices confirms that coal combustion residues, ash
434 deposition, and settlement-related human activities are the primary drivers of soil
435 contamination. In contrast, chromium, cobalt, and nickel are largely controlled by
436 geogenic background conditions and contribute comparatively little to overall human
437 health risk.

438 Human health risk assessment further reveals that the most enriched and polluted
439 metals are also responsible for elevated non-carcinogenic and carcinogenic risks.
440 High hazard index values for arsenic, lead, and cadmium, together with carcinogenic
441 risks exceeding acceptable thresholds for arsenic and lead, indicate that soil
442 contamination poses a tangible public health concern in rural communities. These
443 findings highlight a direct linkage between the severity of environmental
444 contamination and adverse human health outcomes.

445 Overall, the integrated assessment framework proved effective in synthesizing
446 complex datasets and providing a holistic understanding of soil-related environmental
447 and human health risks. The results underscore the need for targeted mitigation
448 measures, including the promotion of cleaner household energy practices, improved
449 coal ash management, and focused remediation of identified high-risk areas. The
450 methodological approach and findings of this study offer a valuable reference for
451 environmental risk assessment and evidence-based decision-making in rural regions
452 facing similar contamination challenges.

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525 **Code and data availability**

526 The data supporting the findings of this study are available from the corresponding
527 author upon reasonable request. Heavy metal concentration data for surface soils,
528 together with associated sampling coordinates, were generated during the field
529 campaign conducted in Uliastai, Zavkhan Province, Mongolia. All data processing,
530 statistical analyses, and figure preparation were carried out using standard scientific
531 software. No proprietary code was used in this study.

532 If applicable, derived datasets used for multivariate analysis, contamination indices,
533 and human health risk assessment can be provided to facilitate reproducibility of the
534 results.

535 **Supplement**

536 The supplement related to this article includes additional tables and figures supporting
537 the main text, including detailed statistical outputs and supplementary spatial
538 visualizations. The supplement is available online upon publication.

539 **Author contributions**

540 E.Btse. conceived and designed the study, conducted field sampling, performed
541 laboratory analysis, contributed to data analysis, statistical interpretation, and
542 visualization, and led data interpretation and manuscript preparation. C.S. contributed
543 to the conceptualization of the study. B.G and K.F were responsible for laboratory



544 analysis and instrumental measurements. D.D. and A.D. E.Btu assisted with
545 fieldwork, soil sample preparation for measurement, data validation, and mapping. All
546 authors reviewed and approved the final manuscript.

547 **Competing interests**

548 The authors declare that they have no competing interests.

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