



1 UAV-based method for measuring CO₂
2 emissions in forest ecosystems

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28 Abstract

29 This study addresses the difficulty of accurately quantifying CO₂ emissions in
30 forest ecosystems due to spatial heterogeneity, complex terrain, and the combined
31 effects of horizontal and vertical transport. A UAV-based platform, capable of three-
32 dimensional CO₂ emission observation, was developed, evaluated and applied to a
33 forest ecosystem. The system integrates a high-precision closed-path CO₂ analyzer with
34 a calibrated ultrasonic anemometer, and employs complementary box-pattern and
35 profile-pattern flight strategies within a mass-balance framework to simultaneously
36 resolve horizontal transport and vertical exchange. Validation against a long-term eddy-
37 covariance (EC) flux tower shows that vertical CO₂ fluxes derived using the gradient
38 method agree well with EC observations across seasons ($R^2 \approx 0.76\text{--}0.77$). Box-pattern
39 flights further reveal pronounced diurnal variation: lateral advection dominates in the
40 early morning, whereas vertical uptake prevails under well-mixed midday conditions.
41 Sensitivity analyses across scales of 50 m, 100 m, 150 m indicate that CO₂ emission
42 intensity is sensitive to control-volume dimensions and shows spatial heterogeneity
43 above the forest. Uncertainty assessment suggests that the total relative uncertainty of
44 UAV-derived emissions is typically ~15%, with wind-field calibration as the main error
45 source. Overall, this UAV approach provides a flexible and reliable complement for
46 analyzing near-surface three-dimensional CO₂ transport over complex and
47 heterogeneous forests, helping to overcome the limited spatial coverage of traditional
48 flux towers and remote sensing.

49 **Keywords:** UAV-based observations; CO₂ emissions; Mass balance; Three-
50 dimensional flux measurement; Spatial heterogeneity



51 **1 Introduction**

52 Carbon dioxide (CO₂) is the dominant anthropogenic greenhouse gas in the
53 atmosphere, contributing approximately 50–60% to the enhanced greenhouse effect
54 excluding water vapor (WMO, 2023). Atmospheric CO₂ budget, however, is strongly
55 impacted by biospheric carbon cycles. Forest ecosystems, in particular, play a critical
56 role in the terrestrial carbon cycle and are widely recognized as major carbon sinks,
57 covering about 40% of the global land surface and representing some of the most
58 extensive and structurally complex ecosystems on Earth (Bonan, 2008; IPCC, 2021;
59 Pan et al., 2011). Consequently, an accurate understanding of carbon exchange
60 processes between forests and the atmosphere is essential for assessing regional and
61 global carbon budgets and their feedbacks on climate change. However, despite decades
62 of carbon-cycle research to estimate CO₂ exchange across spatial scales, including eddy
63 covariance (EC) towers, satellite-based remote sensing, and manned aircraft
64 measurements (Baldocchi, 2003; Gioli et al., 2004; Jung et al., 2020), most existing
65 methods are inherently confined to either vertical or horizontal dimensions due to
66 theoretical or practical limitations, which restrict their ability to represent the full three-
67 dimensional structure of CO₂ transport in the lower atmosphere (Leuning et al., 2008).
68 In forest ecosystems, strong spatial heterogeneity, complex canopy–atmosphere
69 interactions, and non-negligible horizontal transport processes can introduce substantial
70 uncertainties into quantitative estimates of ecosystem–atmosphere CO₂ exchange
71 (Baldocchi, 2008; Schimel et al., 2015). As a result, approaches capable of resolving
72 both horizontal and vertical CO₂ transport are required.

73 Eddy covariance (EC) flux measurement technique is currently the most widely
74 used method for quantifying ecosystem-scale vertical CO₂ fluxes and has been
75 implemented globally through networks such as the FLUXNET (Aubinet et al., 2012;
76 Baldocchi et al., 2001; Baldocchi, 2003; Wilson et al., 2002). By design, EC provides
77 direct measurements of time-averaged vertical turbulent flux of CO₂ between the
78 surface and the atmosphere. However, the method relies on the validity of the Monin–
79 Obukhov Similarity Theory (MOST), which assumes a horizontally homogeneous, flat



80 terrain and stationary turbulence (Foken, 2008; Kaimal and Finnigan, 1994; Monin and
81 Obukhov, 1954; Stull, 1988). Over forested or topographically complex terrains, these
82 assumptions are often violated, calling into question the representativeness of EC flux
83 measurement results over such ecosystems. EC flux measurements do not explicitly
84 resolve horizontal CO₂ transport associated with advection, canopy heterogeneity,
85 terrain-induced flows, or mesoscale circulations (Aubinet et al., 2002; Feigenwinter et
86 al., 2004; Finnigan et al., 2003; Leuning et al., 2005). Horizontal fluxes are therefore
87 commonly neglected or assumed to be negligible in EC flux measurements, an
88 assumption that may not hold under weak or non-stationary turbulence conditions
89 (Mahrt, 1998, 2010; Vickers and Mahrt, 1997). In addition, EC towers provide flux
90 information at a fixed height, limiting their ability to characterize CO₂ transport at
91 different levels within the atmospheric boundary layer (Schmid, 2002), hence unable to
92 determine flux divergence. For large scale atmosphere-ecosystem CO₂ exchanges,
93 satellite remote sensing offers valuable information on horizontal distributions of
94 atmospheric pollutants (Crisp et al., 2017; Jacob et al., 2016), but it has limited
95 sensitivity to vertical gradients near the surface and cannot resolve canopy-scale
96 exchange processes with the accuracy of in situ measurements (Chevallier et al., 2014).
97 Likewise, manned aircraft campaigns provide regional-scale observations of pollutants
98 distributions and enable mass-balance approaches to emission estimation (Cambaliza
99 et al., 2014), but vertical fluxes are difficult to determine (Gordon et al., 2015).
100 Collectively, these limitations highlight a persistent observational gap in the
101 characterization of CO₂ three-dimensional transport in forest ecosystems.

102 Recent advances in rotating wing unmanned aerial vehicle (UAV) platforms and
103 lightweight sensor technologies have created new opportunities for atmospheric
104 measurements over complex environments. UAVs have emerged as practical tools for
105 complementing existing carbon observation networks by enabling flexible, high-
106 resolution measurements of greenhouse gases in both horizontal and vertical directions
107 (Bolek et al., 2024; Khan et al., 2012; Villa et al., 2016). Compared with fixed EC
108 towers, UAVs can directly measure concentration spatial variability over short time and



109 spatial scales without relying on assumptions of horizontal homogeneity (Witte et al.,
110 2017), enabling identification of emission or uptake hotspots that may be missed by
111 ground-based observations (Khan et al., 2012). Nonetheless, these UAV applications
112 have not demonstrated the feasibility of CO₂ flux measurements owing to their inability
113 to conduct accurate wind speed and greenhouse gas concentration measurements,
114 partially due to platform-induced airflow disturbances and rapidly changing flight
115 attitude (Soltaninezhad et al., 2025; Thielicke et al., 2021). To address these challenges,
116 various sensors, including multi-hole probes, pitot tubes, and sonic anemometers, have
117 been deployed on UAVs for wind measurements and correction algorithms have been
118 developed for disturbances (Langelaan et al., 2011; Niedzielski et al., 2017; Nolan et
119 al., 2018; Rogers and Finn, 2013; Soddell et al., 2004; Spiess et al., 2007). Among these,
120 sonic anemometers are considered particularly promising because of their high
121 accuracy and fast response (Thielicke et al., 2021). Nevertheless, the influence of rotor-
122 induced flow distortion, UAV attitude changes, and local turbulence on wind
123 measurements remains non-negligible and requires dedicated calibration strategies
124 (Soltaninezhad et al., 2025).

125 Building on our previous work on the development of an UAV platform for
126 integrated high sensitivity, high precision, fast response measurements of GHG
127 concentrations and 3-D wind speed, combined with a comprehensive wind calibration
128 algorithm to correct for UAV motion, attitude, and rotor effects (Yang et al., 2025), we
129 aim to improve quantification of atmospheric CO₂ transport in both horizontal and
130 vertical directions over forest ecosystems. In the present study, we evaluate the UAV
131 platform and the wind correction algorithm to assess its ability to determine three-
132 dimensional CO₂ transport including canopy-scale vertical CO₂ fluxes, and to evaluate
133 the robustness and uncertainty of the UAV-derived emission estimates. The evaluation
134 is based on specifically designed flight strategies with complementary box and profile
135 flight patterns and is conducted in field experiments at a subtropical forest site. Such
136 flight strategies are implemented within a mass-balance framework to resolve three-
137 dimensional CO₂ exchange in the lower atmosphere. While horizontal mass flux



138 measurements using this platform have been previously evaluated and validated in field
139 measurements (Han et al., 2024) vertical fluxes derived from such a platform have not
140 been evaluated until now, and a critical part of the present study is to validate the UAV
141 vertical flux measurement methodology against the collocated tower-based EC flux
142 measurements. As presented further below, the results demonstrate a highly versatile
143 methodology for determining net carbon fluxes out of complex ecosystem terrains that
144 not only complements existing carbon flux tower measurements but also resolves
145 underlying issues related to horizontal carbon transport not addressed in carbon cycle
146 research previously.

147 **2 Field site and platform description**

148 **2.1 Platform description**

149 **2.1.1 UAV platform**

150 The UAV platform used in this study is a battery-powered hexacopter (KWT-X6L-
151 15, ALLTECH, China), which provides stable flight performance and high
152 maneuverability, including vertical take-off and landing, hovering, and controlled
153 horizontal flight (Fig. 2b). This airframe was selected as the base platform for CO₂
154 measurements because of its ability to operate safely above forest canopies, its high
155 payload capacity, and its adaptability to multiple flight strategies. The UAV has a
156 maximum span of arm plus rotor blade of 2570 mm, and the height from the body stands
157 to the body top is 565 mm. Its maximum horizontal flight speed is 18 m s⁻¹, the
158 maximum ascending rate is 4 m s⁻¹, and the maximum wind resistance is 14.4 m s⁻¹, all
159 of which satisfy the requirements for atmospheric wind and CO₂ measurements in the
160 lower boundary layer. A detailed summary of the UAV specifications is provided in
161 Table 1.

162

163 **Table 1.** Specifications and capabilities of the UAV.

hover endurance	≥30 min (for 15 kg payload)
Peak flight velocity	18 m s ⁻¹
Peak climb velocity	4 m s ⁻¹



Control radio range	≥ 7 km
Operating ambient temperature	-20 °C~ 55 °C
Maximum operating altitude	≥ 5000 m
Wind resistance capability	14.4 m s ⁻¹
Payload capacity	15 kg (without battery)
Rotor blade length	765 ± 5 mm
Symmetric motor center-to-center length	1635 ± 10 mm
Body stands to the body top length	565 ± 5 mm
Wing span	2570 ± 10 mm
Battery capacity/unit	30000 mAh (2 units)
Battery weight/unit	6 kg \pm 0.1 kg
Radio modem	902 MHz ~928 MHz
GPS accuracy	Horizontal direction: $\pm (8+10^{-6}D)$ mm; Vertical direction: $\pm (15+10^{-6}D)$ mm
Navigation	GPS (RTK system)

164

165 The airframe is constructed from carbon-fiber composite materials and houses an
 166 integrated flight control system, a power supply module, and a communication unit.
 167 The flight control system includes an inertial measurement unit (IMU), accelerometers,
 168 and a global positioning system (GPS) capable of real-time kinematics (RTK), and
 169 records UAV attitude and motion parameters (roll, pitch, yaw, and three-axis
 170 accelerations) which is critical for subsequent wind-speed correction. To improve
 171 positioning accuracy, the onboard RTK GPS receiver takes in real-time signals from a
 172 concurrently ground-deployed kinematic (RTK) GPS system (iRTK10, Hi-Target,
 173 China) for real time computation for centimetre- to millimetre-level accuracies in both
 174 the horizontal and vertical directions, which are particularly advantageous over a
 175 complex terrain. The power module consists of two lithium–polymer batteries
 176 connected in parallel, each with a capacity of 30,000 mAh, enabling a maximum
 177 payload of 15 kg and a typical flight endurance of approximately 30 min. The
 178 communication module operates in the 890-960 MHz frequency band, allowing reliable
 179 telemetry and control within a range of up to 10 km.

180 2.1.2 Scientific payloads

181 The integrated scientific payloads, with a total weight of 6.35 kg including the
 182 mounting chassis, were powered directly by the main UAV batteries. The payloads
 183 include a closed-path CO₂ analyzer and a 3-d ultrasonic anemometer, with their



184 specifications summarized in Table 2. The 3-d ultrasonic anemometer (Trisonica Mini,
185 LI-COR Environment, USA) is a compact sensor, with a measurement path of 35 mm
186 and a mass < 50 g, that measures three-dimensional wind speeds under its own
187 coordinate system, air temperature, and pressure. The anemometer was mounted 70 cm
188 above the rotor plane to minimize flow distortion caused by propeller blade downwash
189 (Fig. 1b) (Yang et al., 2025). This configuration has been shown to achieve acceptable
190 accuracy (± 2 %) in previous validation studies using reference-grade meteorological
191 instruments (Palomaki et al., 2017; Shimura et al., 2018; Thielicke et al., 2021). The
192 CO₂ analyzer (MIRA Ultra N₂O/ CO₂, Aeris Technologies, USA) was mounted beneath
193 the UAV body to measure CO₂ volume mixing ratios in air, with its air inlet set next to
194 the anemometer sonic sensors to ensure spatially consistent wind and CO₂
195 concentration measurements. A fixed time lag of 5 s, associated with air residence time
196 in the sampling line, was applied to synchronize the CO₂ measurements with the wind
197 data.

198 Data from all scientific payloads, together with UAV attitude and flight parameters
199 (roll, pitch, yaw, three-axis accelerations, and GPS position), were recorded by the
200 onboard main controller via UART communication and stored on an SD card (Fig. 1).
201 In addition, real-time telemetry, including sensor status, battery voltage, and flight
202 parameters, was transmitted to a ground station, allowing continuous monitoring during
203 flight operations. The CO₂ measurements were recorded at 1 Hz, while the anemometer
204 data were recorded at 5 Hz.

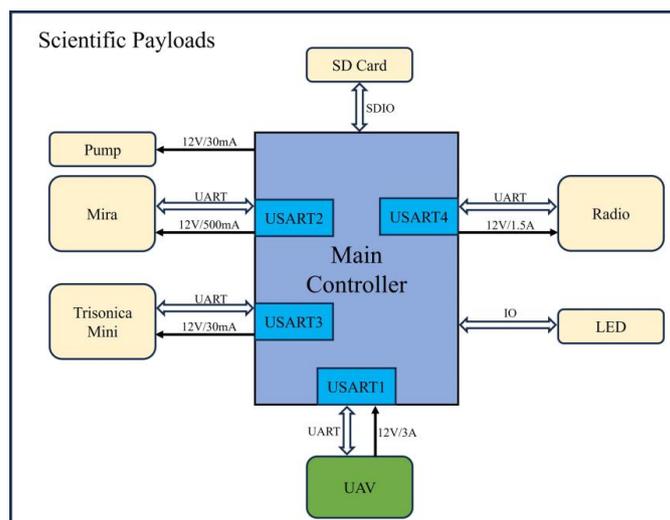
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Table 2. Detailed information of scientific payloads.

Instrument	Sample Frequency (Hz)	Weight (kg)	Measurement	Accuracy
MIRA Ultra	1	6	CO ₂	< 200 ppb/s
Trisonica Mini	5	0.05	Wind speed	(0 - 10 m/s): ± 0.2 m/s (11 - 30 m/s): $\pm 2\%$
			Wind direction	$\pm 1.0^\circ$
			T	± 2.0 °C
			P	± 1.0 kPa



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Figure 1. Diagram of the scientific payloads on the UAV-based flux measurement system.

209 2.2 Field site description

210 The UAV-based field measurements were conducted at the Qianyanzhou
211 Experimental Station (26°44'29" N, 115°03'29" E) located in Jiangxi Province,
212 southern China. This station is one of the core flux observation sites of the China Flux
213 Observation Research Network (ChinaFLUX) and is equipped with a comprehensive
214 eddy covariance (EC) measurement system (Yu et al., 2006). The EC tower has a height
215 of approximately 43 m above ground, with the EC system installed at 31 m above
216 ground level, while the mean forest canopy height is about 18 m.

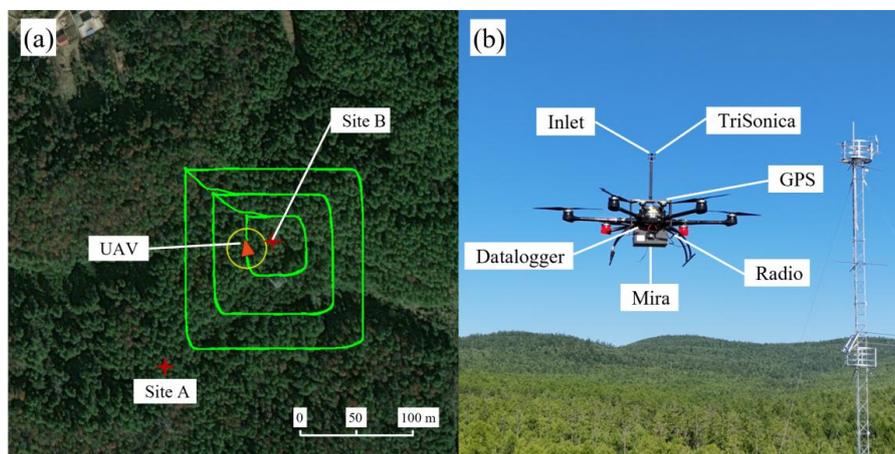
217 The long-term observation site, 1 ha in size and centred on the EC tower, is
218 dominated by plantation forests, primarily consisting of *Pinus massoniana*,
219 *Cunninghamia lanceolata*, and *Pinus elliottii* arranged in a strip-mixed pattern with an
220 approximate ratio of 49:48:3, respectively (Huang et al., 2007). The region experiences
221 a typical subtropical monsoon climate, characterized by prevailing southerly winds in
222 summer and northerly winds in winter. The mean annual air temperature is 17.9 °C, and
223 the mean annual precipitation is approximately 1490 mm, with most rainfall occurring
224 between March and June.
225



226 To evaluate the performance of the UAV platform for CO₂ vertical flux
227 measurements, measurement flights were conducted at Site B, located approximately
228 160 m southeast of the EC tower (Site A), and direct comparisons with tower-based
229 measurements were performed (Fig. 2a). Based on a comprehensive site survey, Site B
230 was selected as the UAV take-off and landing location because of its relatively open
231 forest structure, complex surrounding vegetation, and a lower elevation (~15 m lower)
232 than Site A, which facilitates safe UAV operations. Within a radius of approximately
233 350 m northwest of Site B, only a limited number of low-rise buildings below the forest
234 canopy used exclusively for scientific research are present, resulting in minimal
235 anthropogenic influence on the measured CO₂ fluxes. Such an experimental condition
236 provides a representative forest ecosystem setting in which UAV-derived fluxes can be
237 evaluated against EC tower observations.

238 From May 2023 to December 2024, a total of 209 UAV flights were conducted at
239 the site, corresponding to approximately 100 h of cumulative flight time. After data
240 quality controls, including the removal of records affected by rainfall and other
241 conditions that could compromise instrument performance, results from 168 flights
242 were retained for analysis, including 40 flights in spring, 55 in summer, 53 in autumn,
243 and 20 in winter.

244



245
246 **Figure 2.** (a) A top-down view of the UAV's flight tracks. The green flight tracks represent the box



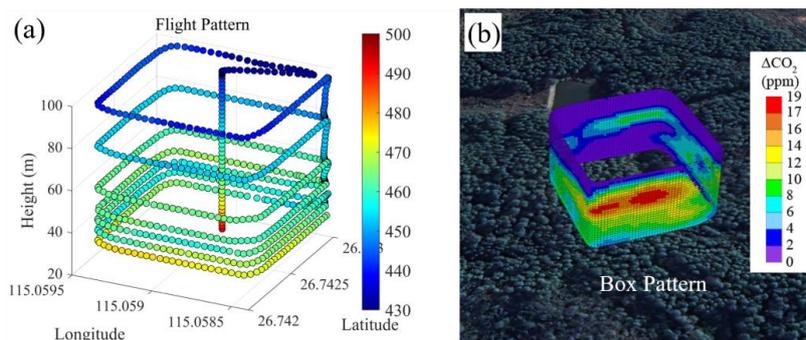
247 pattern, categorized into three box sizes: 50×50 m, 100×100 m, and 150×150 m. Site A marks the
248 location of the flux tower where EC measurements of vertical CO₂ fluxes are continuously carried
249 out, while Site B denotes the take-off point for the profile pattern. The straight-line distance
250 between the two sites is 150 m. (b) The instrumented UAV platform; the flux tower can be seen in
251 the background.

252 **2.3 Flight patterns**

253 Two complementary flight strategies were executed in this study: a box pattern
254 and a profile pattern (Fig. 3a). The box pattern consists of a closed rectangular flight
255 path flown above the forest canopy at multiple altitudes, enabling sampling of
256 horizontal CO₂ transport across a defined control volume. Box flights were conducted
257 at heights of 35, 40, 45, 50, 60, 80, and 100 m above ground level. Using flight-path
258 fitting and screen interpolation (Han et al., 2024), the three-dimensional spatial
259 distributions of CO₂ volume mixing ratio, wind vectors, and air density were
260 reconstructed within the box domain (Fig. 3b). Flight-path fitting refers to the
261 construction of a smooth, closed loop constructed from the discrete UAV trajectories at
262 all altitudes, thereby defining the four virtual vertical surfaces of the box. Screen
263 interpolation refers to the mapping of CO₂ mixing ratios and corrected wind speed
264 measurements observed along the flight path onto a regular two-dimensional grid on
265 the vertical surfaces, which enables numerical integration of fluxes across each box
266 wall.

267 The profile pattern was designed to capture vertical gradients of CO₂ mixing ratios,
268 wind speed, and air density from near the surface to the top measurement height. Each
269 box-pattern flight was paired with a vertical profile, and all measurements were
270 completed within approximately 20 min to ensure consistency of the sampled air mass
271 in the controlled volume. Vertical CO₂ fluxes were computed from the profile data using
272 the gradient method (Sect. 3.2.2), with the analysis height range defined by the lowest
273 and highest box levels (35–100 m). Because the EC system is installed at 31 m, UAV-
274 derived CO₂ vertical fluxes used data from 28 and 33 m to enable direct comparison
275 with tower-based EC measurements.

276



277

278 **Figure 3.** Overview of flight patterns. (a) Two distinct flight strategies: box pattern and the profile
279 pattern. (b) The spatial distribution of CO₂ mixing ratios obtained through flight-path fitting and
280 screen interpolation for flights during 2024-08-08 7:00 – 7:30, morning.

281 3 Methodologies

282 3.1 Wind speed calibration

283 To obtain accurate and reliable wind speed measurements, the raw data acquired
284 by the ultrasonic anemometer were processed using a dedicated wind speed calibration
285 algorithm. This algorithm, developed by [Yang et al. \(2025\)](#), employs computational
286 fluid dynamics (CFD) simulations to characterize the flow-field distortion by the UAV
287 rotor-induced airflows. Hundreds of simulation scenarios were conducted,
288 encompassing a wide range of flight altitudes, wind directions, and ambient wind
289 speeds, in order to quantify the influence of rotor-induced airflow disturbances on wind
290 measurements. The simulated disturbance fields were subsequently combined with
291 UAV motion and attitude compensation schemes, yielding a comprehensive correction
292 algorithm that simultaneously accounts for UAV translation, rotation, and rotor effects.
293 The algorithm was validated with 3-D wind measurements at multiple levels on an 80-
294 meter met tower [Yang et al. \(2025\)](#). All wind measurements used in this study were
295 corrected using this algorithm, ensuring consistency and comparability with reference-
296 grade wind observations.

297 3.2 CO₂ emission estimation based on UAV measurements

298 CO₂ emissions from within the control volume defined by the virtual box were



299 quantified using a mass-balance framework that integrates measurements obtained
300 from box-pattern and profile-pattern UAV flights. This approach assumes that CO₂
301 emitted within the control volume is transported by the mean wind field and becomes
302 sufficiently mixed, such that the net emission rate can be inferred from the fluxes
303 crossing the boundaries of the volume. Under steady or quasi-steady wind conditions,
304 the mass flux of a chemical compound through a plane perpendicular to the mean wind
305 direction can be considered approximately constant.

306 According to the divergence theorem, the rate of change of mass of a stable
307 compound within a control volume equals the surface integral of the mass flux through
308 its enclosing surfaces. Consequently, the total CO₂ emission rate within the volume
309 equals the net outward flux integrated across all box wall surfaces. Based on the integral
310 form of the continuity equation, the emission rate within the control volume enclosed
311 by the box pattern can be expressed as

$$312 \quad E_C = E_{C,H} + E_{C,HT} + E_{C,V} + E_{C,VT} - E_{C,M} \quad (3.1)$$

313 where E_C denotes the total CO₂ emission rate within the control volume, $E_{C,H}$ and
314 $E_{C,V}$ denote the integrals of horizontal and vertical advective fluxes through the box
315 walls, respectively; $E_{C,HT}$ and $E_{C,VT}$ represent the integrated turbulent fluxes in the
316 horizontal and vertical directions, and $E_{C,M}$ accounts for changes in CO₂ mass
317 associated with air-density variations. Eq. (3.1) forms the basis of the TERRA emission
318 inversion model (Gordon et al., 2015).

319 To investigate the contributions of each component to the emissions, the TERRA
320 model was updated here, in particular the treatment of the vertical flux algorithm, to
321 directly compute vertical turbulent fluxes based on the UAV measurements instead of
322 relying on parametrization assuming entrainment. Here, the mean wind vectors (\bar{u} , \bar{v} ,
323 \bar{w}) measured with the UAV sonic anemometer over a 30-minute period and the
324 corresponding wind fluctuations (u' , v' , w') are used to determine fluxes in the three
325 directions. For the horizontal terms, the mean horizontal wind speed \bar{u} can be regarded
326 as the main contributor to the $E_{C,H}$ term, typically 2–3 m/s. The horizontal turbulent
327 velocity (u') is the main contributor to $E_{C,HT}$. Using the 95th percentile of u' , the
328 turbulent velocity is less than 0.02 m/s. The horizontal advection term ($E_{C,H}$) is larger



329 than the turbulence term ($E_{C,HT}$) by more than 100 times, so that the $E_{C,HT}$ term can
330 be neglected from Eq. (3.1). For the vertical flux terms, the reverse is true; the mean
331 vertical wind speed \bar{w} , at typically <0.02 m/s, makes $E_{C,V}$ a minute contribution
332 compared to the turbulent flow contribution $E_{C,VT}$, which depends on the vertical
333 turbulent velocity (w'). Using the 95th percentile of w' , the turbulent velocity can
334 reach 0.3 m/s. As such, vertical advection term can also be neglected from Eq. (3.1).

335 The density-related term can be expressed as

$$336 \quad E_{C,M} = M_R \iiint \chi_C \frac{d\rho_{air}}{dt} dx dy dz \quad (3.2)$$

337 Because the maximum vertical (≈ 65 m), horizontal (≈ 150 m) extents, and duration
338 of the UAV flights are limited, air-density variation over time $\frac{d\rho_{air}}{dt}$ within the control
339 volume can be assumed to be negligible, and $E_{C,M}$ was therefore neglected. Accurate
340 estimation of the remaining flux terms ($E_{C,H}$ and $E_{C,VT}$) enables computation of the
341 net CO₂ emission rate. The reliability of the box-pattern approach for resolving CO₂
342 emission is further supported by [Han et al. \(2024\)](#), who conducted a series of controlled
343 UAV-based CO₂ validation experiments over an industrial coking facility.

344

345 3.2.1 Computation of CO₂ horizontal emission component

346 The horizontal emission component was derived from box-pattern measurements
347 by integrating the advective CO₂ flux across the four vertical walls of the virtual box.

348 The horizontal CO₂ emission rate can be written as

$$349 \quad E_{C,H} = M_R \oint \chi_C(s, z) \rho_{air}(s, z) U_{\perp}(s, z) ds dz \quad (3.3)$$

350 For each altitude level, the line flux $F(z)$ was first calculated from the box-
351 pattern data as

$$352 \quad F(z) = M_R \sum_0^L (\chi_C(s, z) \cdot \rho_{air}(s, z) \cdot u_{\perp}(s, z)) \cdot \Delta s \quad (3.4)$$

353 where Δs denotes the distance travelled during each 1 s sampling interval, and L
354 represents the total length of the closed flight path at that altitude.

355 To suppress small-scale variability, a moving average was applied to consecutive
356 line-flux values, yielding layer-averaged fluxes for discrete altitude intervals. These



357 layer-integrated values were then vertically integrated to obtain the total horizontal CO₂
358 emission rate

$$359 \quad E_{C,H} = \sum_0^N F(z) \cdot \Delta z = M_R \sum_0^N (\sum_0^L (\chi_C(s,z) \cdot \rho_{air}(s,z) \cdot u_n(s,z)) \cdot \Delta s) \cdot \Delta z \quad (3.5)$$

360 where Δz denotes altitude intervals and N denotes N sets of line flux values.

361

362 3.2.2 Computation of vertical emission component

363 The vertical CO₂ emission component was estimated from profile measurements
364 using the gradient method, which relates turbulent scalar fluxes to vertical gradients of
365 the scalar concentration. In practice, vertical profiling over forested terrain is
366 operationally challenging, and UAV measurements were conducted at a single
367 representative location. The resulting vertical fluxes were assumed to be representative
368 of the broader domain, an assumption evaluated through direct comparison with eddy
369 covariance (EC) tower measurements (see Section 4.2.1).

370 The vertical CO₂ emission component was obtained from integrating the vertical
371 flux over the horizontal area enclosed by the box pattern and can be approximated by
372 the product of vertical flux and the horizontal area, expressed as

$$373 \quad E_{C,VT} = \iint F_v dx dy = A \cdot F_v \quad (3.6)$$

374 where F_v denotes the vertical flux and A is the horizontal area of the control volume.

375 According to the flux–gradient relationship, turbulent scalar fluxes are
376 proportional to the vertical gradient of the scalar quantity:

$$377 \quad F_v = -\rho_{air} K_C \frac{\partial C}{\partial z} \quad (3.7)$$

378 where K_C denotes turbulent exchange coefficient, \bar{C} represents average CO₂
379 concentration, converted from measured mixing ratio at the measured T and P.

380 Following the turbulent mixing-length theory and the friction velocity definition

$$381 \quad u_* = (-\overline{u'w'})^{1/2},$$

$$382 \quad K = l u_* \quad (3.8)$$

$$383 \quad u_* = l \frac{\partial U}{\partial z} = \kappa z \frac{\partial U}{\partial z} \quad (3.9)$$

384 where l is the mixing length and κ is the von Karman constant ($\kappa=0.4$).



385 Stratification effects were incorporated using a stability correction, given
386 by

$$387 \quad l = mkz \quad (3.10)$$

388 where m is the stratification correction parameter, determined as a function of the
389 gradient Richardson number (R_i)

$$390 \quad m = (1 - R_i)^{1/2} \quad (3.11)$$

391 Under neutral stratification conditions ($R_i = 0$), $m = 1$. For computational convenience,
392 the gradient Richardson number R_i is approximated by the bulk Richardson number in
393 finite-difference form

$$394 \quad R_i \approx R_B = \frac{g\Delta\bar{\theta}}{\bar{\theta}\Delta z} \frac{\Delta\bar{U}}{(\Delta U)^2} = \frac{g}{\bar{\theta}} \frac{\Delta\bar{\theta}\Delta z}{(\Delta U)^2} \quad (3.12)$$

395 where g is the gravitational acceleration constant; $\Delta\bar{\theta}$ and $\Delta\bar{U}$ represent the mean
396 temperature change and wind speed change between the layers (Δz); $\bar{\theta}$ is the mean
397 temperature between the layers.

398 Because forest canopies consist of complex underlying surfaces, wind profiles
399 may deviate from the ideal logarithmic form. The surface-layer wind profile was
400 therefore expressed as

$$401 \quad \bar{U} = \frac{u_*}{\kappa} \ln \frac{z-d}{z_0} \quad (3.13)$$

402 where d is the zero-plane displacement height, typically equal to 70%-80% of the
403 vegetation height, and z_0 is the surface roughness length.

404 Assuming neutral atmospheric stratification and treating u_* , d , and z_0 as
405 constants over the same time period, u_* can be derived from wind measurements at
406 two heights thereby eliminating z_0 from Eq. (3.13) and rearranging the equation so
407 that

$$408 \quad u_* = \frac{\kappa(\bar{U}_2 - \bar{U}_1)}{\ln [(z_2 - d)/(z_1 - d)]} \quad (3.14)$$

409 where \bar{U}_1 and \bar{U}_2 are the mean horizontal wind speeds at heights z_1 and z_2 ,
410 respectively. In the present study, the two mean horizontal wind speeds were taken from
411 the horizontal wind measurements at any two heights as the UAV ascended during the
412 vertical profile measurements. Substituting Eq. (3.10), Eq. (3.11) and Eq. (3.14) into



413 Eq. (3.8) yields

$$414 \quad K = \kappa^2(1 - R_i)^{1/2} \frac{(\overline{U_2} - \overline{U_1})}{\ln [(z_2 - d)/(z_1 - d)]} (z - d) \quad (3.15)$$

415 Substitute Eq. (3.15) into Eq. (3.7) and integrate to obtain the vertical flux

$$416 \quad F_v = \rho_{air} \kappa^2 (1 - R_i)^{1/2} \frac{(\overline{U_2} - \overline{U_1})(\overline{C_1} - \overline{C_2})}{\{\ln [(z_2 - d)/(z_1 - d)]\}^2} \quad (3.16)$$

417 where $\overline{C_1}$ and $\overline{C_2}$ represent the mean concentrations of CO₂ at heights z_1 and z_2 ,
418 respectively, converted from measured CO₂ mixing ratios at the measured T and P at
419 both heights during the UAV ascent.

420 The vertical CO₂ emission ($E_{C,VT}$) is finally expressed as

$$421 \quad E_{C,VT} = A \rho_{air} \kappa^2 (1 - R_i)^{1/2} \frac{(\overline{U_2} - \overline{U_1})(\overline{C_1} - \overline{C_2})}{\{\ln [(z_2 - d)/(z_1 - d)]\}^2} \quad (3.17)$$

422 In summary, horizontal CO₂ emission components were derived from box-pattern
423 measurements using the divergence theorem, whereas vertical emission components
424 were obtained from profile-pattern measurements using the gradient method. The
425 reliability of the vertical flux estimates, computed using Eq (3.18), was evaluated
426 through direct comparisons with results from EC tower observations (see Section 4.2.1).

427 **3.3 EC system data processing**

428 The EC system consisted of a three-dimensional ultrasonic anemometer (CSAT3,
429 Campbell Scientific, USA), an open-path CO₂/H₂O analyzer (LI-7500, LI-COR
430 Biosciences, USA), and a datalogger (CR3000, Campbell Scientific, USA). Raw data
431 were acquired at a sampling frequency of 10 Hz. The time lag between wind and gas
432 measurements was corrected following the method of [Sahlée et al. \(2007\)](#). Standard
433 quality-control procedures were applied to the turbulence data (*SI Appendix, S1*).

434 Fluxes were computed over 30 min averaging intervals, and coordinate rotation,
435 virtual temperature correction, and frequency-response correction were applied to
436 ensure high-quality vertical flux estimates. The friction velocity (u_*) and scalar fluxes
437 (F_v) were computed as

$$438 \quad u_* = (-\overline{u'w'})^{1/2} \quad (3.18)$$

$$439 \quad F_v = \rho_{air} \overline{w'c'} \quad (3.19)$$

440 where the primes denote turbulent fluctuations of the corresponding variables.



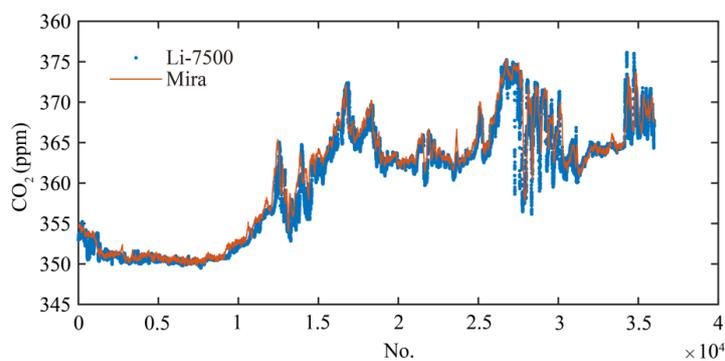
441

442 **4 Results and discussion**

443 **4.1 Gas analyzer comparison test**

444 To evaluate the consistency between different CO₂ analyzers, ground-based
445 intercomparison experiments were conducted using an open-path eddy covariance
446 analyzer (LI-7500) and the closed-path analyzer (MIRA) used on the UAV. The
447 experiment was performed from 21:00 to 22:00 local time on 12 August 2024 (Fig. 4).
448 The air inlets of the two analyzers were positioned at the same height to ensure sampling
449 of the same air mass, with the LI-7500 operating at 10 Hz and the MIRA analyzer at 1
450 Hz. The MIRA analyzer operates in the mid-infrared (MIR) spectral range, where CO₂
451 absorption is substantially stronger than in the near-infrared, enabling a high sensitivity
452 within a compact optical cell. A fixed time lag of 5 s, associated with air flow through
453 the sampling line, was applied to synchronize the MIRA measurements with the LI-
454 7500 data.

455



456

457 **Figure 4.** The comparative chart of CO₂ mixing ratio measured by the LI-7500 and Mira.

458

459 As shown in Fig. 4, the CO₂ mixing ratios measured by the two instruments exhibit
460 highly consistent temporal variabilities, although small differences in absolute values
461 and short-term fluctuations are present. Both datasets show a pronounced increase in



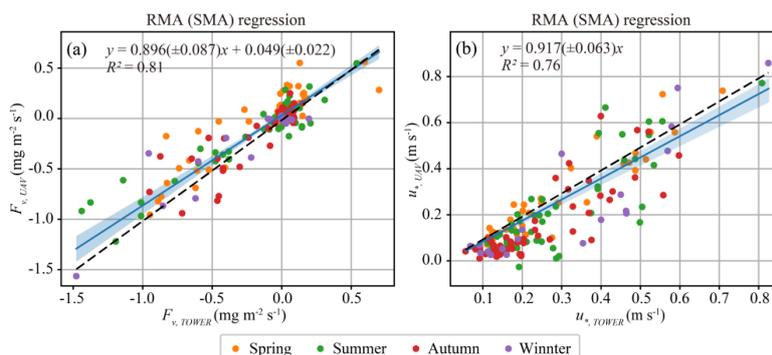
462 CO₂ mixing ratio during the initial phase of the experiment, followed by enhanced
463 variability with multiple peaks, and sustained elevated concentrations during the later
464 period. The overall temporal coherence between the two datasets is strong, indicating
465 that both instruments capture the same underlying CO₂ fluctuation dynamics. The
466 MIRA measurements tend to report slightly lower CO₂ mixing ratios than the LI-7500.
467 This attenuation is primarily attributed to high-frequency losses associated with the
468 closed-path sampling system, including viscous damping within the tubing, interactions
469 with the sampling line material, and signal smoothing induced by the sampling line.
470 Such attenuation is a well-documented characteristic of closed-path systems and does
471 not compromise the ability of the analyzer to resolve mean concentration levels and
472 mesoscale variability. In summary, despite minor attenuations at high frequencies, the
473 closed-path MIRA analyzer demonstrates a strong agreement with the EC-based LI-
474 7500 measurements, supporting its suitability for CO₂ measurements from the UAV-
475 based platform.

476 **4.2 In-flight comparison test**

477 4.2.1 Seasonal variability of vertical fluxes

478 A key part of the present study is to determine reliable vertical flux measurements
479 using the UAV platform over a forest ecosystem on complex terrains. To quantitatively
480 evaluate the agreement between UAV-based profile retrievals and EC tower
481 measurements, the UAV-derived vertical CO₂ flux (F_v) and friction velocity (u_*) against
482 their EC counterparts (Fig. 5). Because both datasets contain measurement uncertainties,
483 the Reduced Major Axis (RMA) regression (Smith, 2009; Warton et al., 2006), which
484 is more appropriate than ordinary least squares when neither datasets can be treated as
485 error-free. The season-specific RMA slope, intercept, and coefficient of determination
486 (R^2) are summarized in Table 3.

487



488
 489 **Figure 5.** Reduced major axis (RMA/SMA) regressions comparing UAV-based measurements
 490 with EC tower observations across all seasons. Panel (a) shows CO₂ flux (F_c) and panel (b) shows
 491 friction velocity (u_*). Points are colored by season (Spring, Summer, Autumn, Winter). The solid
 492 blue line denotes the RMA fit, the shaded band indicates the 95% confidence interval (CI) of the
 493 regression line, and the dashed black line represents the 1:1 reference. Regression equations (slope
 494 \pm 95% CI and intercept \pm 95% CI) and R^2 are reported in each panel.

495
 496 **Table 3.** Seasonal reduced major axis (RMA/SMA) regression results comparing UAV-derived
 497 measurements with EC tower observations for CO₂ flux (F_v) and friction velocity (u_*). For each
 498 season (spring, summer, autumn, winter), the table reports the sample size (n), coefficient of
 499 determination (r^2), RMA slope and its 95% confidence interval (CI), and the intercept of the
 500 relationship $y_{UAV} = \text{slope } x_{TOWER} + \text{intercept}$. The u_* regressions are constrained to pass through
 501 the origin (intercept fixed at 0).

Variables	Season	n	r^2	Slope	95% CI of Slope	Intercept
F_v	Spring	40	0.792	0.928	[0.777, 1.074]	0.127
	Summer	52	0.880	0.801	[0.694, 0.929]	0.034
	Autumn	55	0.744	0.997	[0.771, 1.281]	0.022
	Winter	20	0.840	0.940	[0.551, 1.079]	0.013
u_*	Spring	40	0.888	1.012	[0.926, 1.093]	0
	Summer	52	0.714	0.915	[0.790, 1.021]	0
	Autumn	55	0.710	0.819	[0.679, 0.951]	0
	Winter	20	0.782	0.914	[0.677, 1.059]	0

502
 503 As shown in Fig. 5a, UAV-derived $F_{v,UAV}$ exhibits a clear linear relationship with
 504 EC $F_{v,TOWER}$, with points generally distributed around the 1:1 line, indicating that the
 505 UAV approach captures the temporal variability of canopy-scale vertical CO₂ exchange.



506 Table 3 further demonstrates consistently strong coefficients of determination across
507 seasons ($R^2=0.744-0.880$). The slope, slightly lower than unity, suggests a modest
508 “range compression” in the UAV-derived fluxes, meaning that the magnitudes tend to
509 be slightly attenuated under strong emission or strong uptake conditions (large
510 positive/negative fluxes). In spring, the slope is 0.928 (95% CI: [0.777, 1.074],
511 $R^2=0.792$), and in autumn the slope is 0.997 (95% CI: [0.771, 1.281], $R^2=0.744$); both
512 are close to unity and their confidence intervals encompass 1, suggesting good
513 consistency in flux magnitude. In summer, the slope decreases to 0.801 (95% CI: [0.694,
514 0.929], $R^2=0.880$), and the confidence interval does not include 1, indicating a smaller
515 flux magnitude during periods of stronger exchange. Hence the slightly lower $F_{v,UAV}$
516 than $F_{v,TOWER}$ likely reflects enhanced nonstationarity and rapid scalar fluctuations
517 during peak growing season, which increase sensitivity to sampling height from ground
518 and platform-related perturbations. In winter, the slope is 0.940 (95% CI: [0.551, 1.079],
519 $R^2=0.840$), showing strong correlation but larger uncertainty due to the smaller sample
520 size ($n=20$).

521 Figure 5b shows a clear linear relationship between friction velocity $u_{*,UAV}$ and
522 $u_{*,TOWER}$. Because u_* should physically approach zero under very weak turbulence,
523 an origin-constrained regression (intercept fixed at 0) was used to avoid non-physical
524 extrapolation from a free intercept. The full-sample RMA regression is:

$$525 \quad u_{*,UAV}=0.917(\pm 0.063)u_{*,TOWER}, \quad R^2=0.76 \quad (4.1)$$

526 This result indicates that the UAV platform captures the variations in turbulent
527 intensity well ($R^2 \approx 0.76$), while the slope below unity suggests a slight multiplicative
528 underestimation of u_* relative to tower observations. Regressions for different seasons
529 (Table 3) further indicate that the best agreement occurs in spring, with a slope of 1.012
530 (95% CI: [0.926, 1.093], $R^2 = 0.888$). In contrast, $u_{*,UAV}$ shows a slight
531 underestimation relative to $u_{*,TOWER}$ in summer, autumn, and winter, respectively,
532 with a slope of 0.915, 0.819, 0.914.

533 Taken together, Fig. 5 and Table 3 demonstrate that the UAV platform reproduces
534 the tower-observed variability of both canopy-scale CO_2 exchange and key turbulence

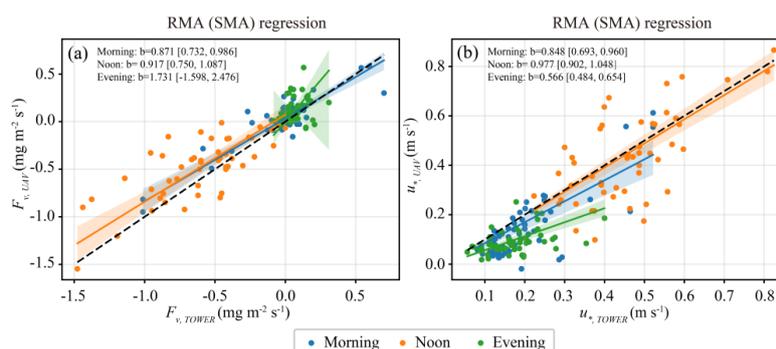


535 metrics across all seasons ($R^2=0.81$ for F_v and $R^2=0.76$ for u_*), supporting the
536 feasibility and robustness of such UAV profile measurements for monitoring above-
537 canopy exchange. Nevertheless, two systematic features warrant attention: (i) F_v
538 magnitudes tend to be slightly underestimated in summer (slope significantly < 1), and
539 (ii) u_* shows a slight overall underestimation under the origin-constrained regression.
540

541 4.2.2 Diurnal variability of vertical fluxes

542 To further evaluate the applicability and consistency of the UAV-based profiling
543 method under different diurnal conditions, the UAV-derived F_v estimates were
544 grouped and compared with results from EC observations in three time periods: early
545 morning (06:00–08:00 LST), noon (13:00–15:00 LST), and evening (19:00–21:00
546 LST). Reduced Major Axis (RMA/SMA) regression was applied to both the vertical
547 CO₂ flux (F_v) and the friction velocity (u_*) for each time period (Fig. 6), with the
548 corresponding statistical results summarized in Table 4.

549



550

551 **Figure 6.** Reduced major axis (RMA/SMA) regressions comparing UAV-based measurements
552 with EC tower observations for (a) CO₂ flux (F_v) and (b) friction velocity (u_*) grouped by time of
553 day. Data are separated into morning (06:00–08:00 LST), noon (13:00–15:00 LST), and evening
554 (19:00–21:00 LST) and shown with different colors. Solid colored lines represent the RMA fits for
555 each time period, and the shaded bands indicate the 95% confidence intervals. The dashed black
556 line denotes the 1:1 reference. Reported in each panel are the RMA slopes and their 95%
557 confidence intervals; the u_* regressions in panel (b) are constrained to pass through the origin
558 (intercept fixed at 0).
559

560 As shown in Table 4, the RMA regressions for F_v exhibit relatively strong



561 correlations during the early morning and midday periods ($R^2=0.806$, $n=63$ and
 562 $R^2=0.752$, $n=550$, respectively). The corresponding slopes are 0.871 (95% CI: [0.732,
 563 0.986]) for early morning and 0.916 (95% CI: [0.750, 1.087]) for midday, both close to
 564 unity but slightly lower than 1, indicating a modest amplitude compression. This
 565 suggests that the UAV-based method is generally capable of reproducing the variability
 566 of CO_2 fluxes observed by the EC system, while tending to underestimate flux
 567 magnitudes under stronger exchange conditions. This behavior is consistent with the
 568 seasonal-scale results presented in Section 4.2.1 and likely reflects the increased
 569 sensitivity of the gradient method to strong flux events under limited vertical sampling
 570 resolution and short-term nonstationary conditions.

571

572 **Table 4.** Reduced major axis (RMA/SMA) regression statistics for comparisons between UAV-
 573 based measurements and EC tower observations, grouped by time of day (morning, noon, and
 574 evening). Results are reported for CO_2 flux (F_v) and friction velocity (u_*), including sample size
 575 (n), coefficient of determination (r^2), RMA slope, the 95% confidence interval (CI) of the slope,
 576 and the intercept for the relationship $y_{\text{UAV}} = \text{slope } x_{\text{TOWER}} + \text{intercept}$. For u_* , the regression is
 577 constrained to pass through the origin (intercept fixed at 0).

Variables	Period	n	r^2	Slope	Slope of 95% CI	Intercept
F_v	Moring	63	0.806	0.871	[0.732, 0.986]	0.034
	Noon	50	0.752	0.916	[0.750, 1.087]	0.075
	Evening	54	0.055	1.730	[-1.598, 2.476]	0.002
u_*	Moring	63	0.786	0.848	[0.926, 1.093]	0
	Noon	50	0.748	0.977	[0.790, 1.021]	0
	Evening	54	0.180	0.566	[0.679, 0.951]	0

578

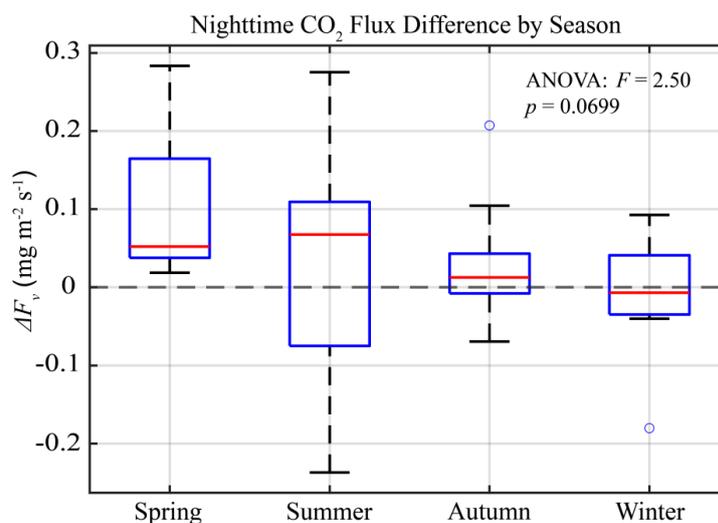
579 In contrast, the agreement for F_v deteriorates substantially during the evening
 580 period, with a low coefficient of determination ($R^2=0.055$, $n=54$) and a highly
 581 uncertain slope (1.730, with a 95% confidence interval spanning both positive and
 582 negative values). The primary reason for such a comparison result is that low nighttime
 583 turbulence leads to vertical fluxes, for which the UAV methodology produces large
 584 uncertainties in the derived vertical fluxes with an apparent low bias in the estimates



585 compared to EC measurements. It should be noted, however, that EC vertical flux
586 measurements under such low turbulence conditions also bear large uncertainties, and
587 hence this comparison only implies a low bias in the UAV methodology relative to the
588 EC methodology, and does not indicate the absolute errors in the UAV-determined F_v .
589 In the late afternoon to evening, as radiative forcing weakens and a stable stratification
590 develops, turbulence above the canopy diminishes, leading to much reduced vertical
591 fluxes. Consequently, the relative contributions of flux source–sink terms, storage, and
592 horizontal transport—particularly advection and non-stationary transport under weak-
593 wind conditions—increase, rendering the stationarity and horizontal homogeneity
594 assumptions of the eddy covariance (EC) method more susceptible to violation. [Aubinet
595 et al. \(2002\)](#) and [Gu et al. \(2005\)](#) also reported that under low-turbulence nighttime
596 conditions, the EC method substantially underestimates ecosystem respiration fluxes.

597 Figure 7 presents the seasonal distributions of nighttime CO₂ flux differences,
598 defined as $\Delta F_v = F_{v,UAV} - F_{v,Tower}$, for spring, summer, autumn, and winter. Overall,
599 the distributions of nighttime ΔF_v show similar central tendencies across the four
600 seasons. The mean values are small, with absolute values generally smaller than 0.1,
601 indicating that the nighttime CO₂ flux differences between UAV-based estimates and
602 EC tower measurements are small in magnitude for all seasons. The one-way analysis
603 of variance (ANOVA) further confirms that the seasonal differences in nighttime ΔF_v
604 are not statistically significant. In the analysis, $F=2.50$, which represents the variability
605 among seasons, is approximately 2.50 times larger than the random variability within
606 seasons; $p=0.0699 > 0.05$, representing the differences among seasons, is not
607 statistically significant. These ANOVA results are consistent with the distribution
608 patterns illustrated in the boxplots, which show no evident systematic seasonal shifts in
609 nighttime flux differences.

610



611
612 **Figure 7.** Boxplots of nighttime differences in CO₂ fluxes (ΔF_v) between UAV and EC tower
613 observations across four seasons. The boxes indicate the interquartile range with median values
614 shown as horizontal lines, small circles represent outliers and whiskers represent 1.5 times the
615 interquartile range.

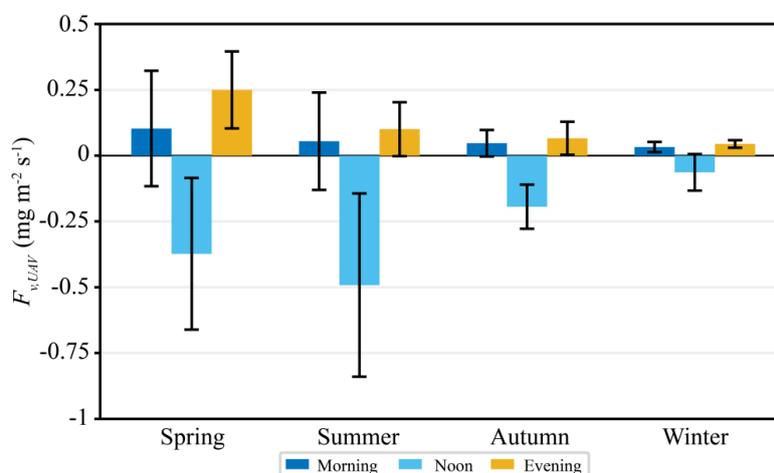
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617 Overall, the nighttime ANOVA analysis further confirms, from a statistical
618 perspective, the consistency between UAV-based and EC tower vertical flux
619 observations at the seasonal scale. When combined with the results obtained from the
620 RMA regression analysis presented above, it can be seen that although flux
621 measurements under nighttime conditions are associated with relatively higher
622 uncertainties, the overall differences between UAV-derived and tower-based
623 observations remain small and do not exhibit pronounced seasonal shifts. This indicates
624 that, despite the apparent low bias using the RMA regression analysis under weak
625 nighttime turbulence conditions, CO₂ fluxes derived from the UAV platform are not
626 statistically different from those derived from tower-based measurements.

627 Figure 8 illustrates the seasonal and diurnal variability of the UAV-derived vertical
628 CO₂ fluxes across morning, noon, and evening periods for all four seasons. Distinct
629 diurnal patterns are evident in every season. Midday periods consistently exhibit the
630 strongest downward CO₂ fluxes, corresponding to maximum ecosystem uptake,
631 particularly in spring, summer, and autumn. In contrast, morning and evening fluxes



632 remain close to neutral or weakly positive, reflecting reduced turbulent mixing and
633 lower photosynthetic activity. This diurnal structure aligns with the canonical coupling
634 between solar radiation, canopy photosynthesis, and boundary-layer turbulence.
635



636
637 **Figure 8.** Seasonal variations of UAV-derived vertical CO₂ fluxes during morning, noon, and
638 evening across spring, summer, autumn, and winter. Bars represent the mean flux for each period,
639 and error bars indicate one standard deviation. Positive values denote upward (emission) fluxes,
640 while negative values represent downward (uptake) fluxes.

641
642 Seasonal contrasts are equally pronounced. Summer and spring display the
643 strongest net CO₂ uptake, consistent with high leaf area index, favorable moisture
644 conditions, and strong radiative forcing in this subtropical forest. Autumn fluxes are
645 weaker, reflecting declining physiological activity, while winter fluxes cluster near zero,
646 with negligible uptake or slight emissions across all periods. The ordering of seasonal
647 uptake strength - summer > spring > autumn >> winter - is consistent with long-term
648 EC observations at this site (Yu et al., 2008).

649 The observed diurnal and seasonal structures in the UAV-based vertical fluxes
650 closely match those reported in previous tower-based studies at the Qianyanzhou station
651 (Yu et al., 2006; Zhang et al., 2014) and in other temperate and subtropical forest
652 ecosystems (Baldocchi, 2008; Goulden et al., 1996). These results therefore not only
653 summarize the observed variability but also provide an ecologically meaningful



654 consistency check for the UAV-based vertical flux approach.

655 **4.3 Spatial heterogeneity of net CO₂ emission**

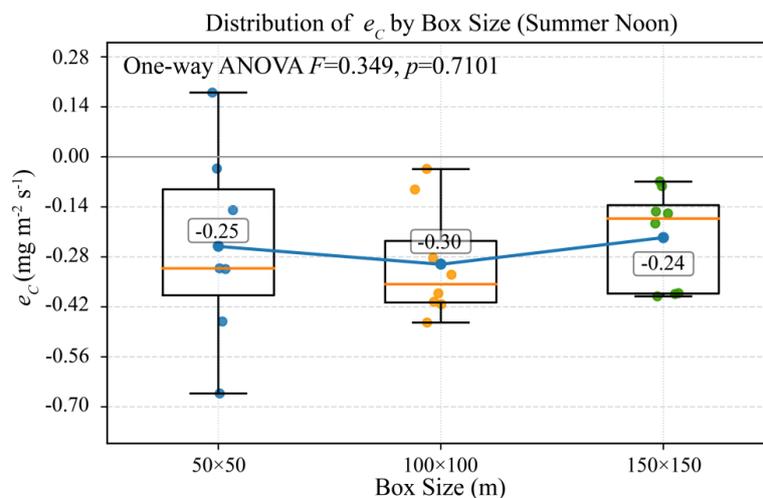
656 To verify whether net emissions (E_C) calculated using the mass-balance approach
657 differ across spatial scales, emission computations were made for three box-shaped
658 control volumes constructed from the UAV flight tracks with different horizontal scales
659 (50×50, 100×100, and 150×150 m; height=50 m). The vertical contribution term $E_{C,VT}$
660 and the horizontal transport contribution term $E_{C,H}$ within each control volume were
661 separately computed as per the detailed outlines above. Net emission is defined as $E_C =$
662 $E_{C,VT} + E_{C,H}$, with uncertainties estimated at 15% (see Section 4.4 for details). To
663 remove effects arising from differences in areas covered by the boxes, the net emissions
664 were normalized by the areas of the boxes to give area-normalized emission intensity
665 $e_C = E_C/A$, where A denotes the bottom area of the box. To minimize the influence of
666 different time periods and seasons on the statistics, and according to the analysis in
667 Section 4.2, the UAV measurements collected in summer morning and noon, which
668 show better agreement with the EC tower comparisons, were selected for the spatial
669 heterogeneity analysis.

670 If the CO₂ source-sink distribution in this region were spatially uniform, then by
671 mass conservation the area-normalized emission intensity would be equal, i.e., $e_C(150)$
672 $\approx e_C(100) \approx e_C(50)$ would be true. Based on the mean annotations of three box scales
673 in Fig. 9 are -0.25, -0.30, and -0.24 (units: mg m⁻² s⁻¹). This indicates a pronounced net
674 CO₂ uptake (i.e., negative emissions) of the system as a whole at summer noon, while
675 the differences among spatial scales are clearly smaller than the 15% systematic
676 uncertainty. This result may suggest relatively weak spatial heterogeneity during the
677 summer-noon period. Furthermore, one-way ANOVA shows the mean differences in
678 e_C among the three scales were not significant ($F=0.349$, $p=0.710$). Overall, under
679 summer-noon conditions, the area-normalized net exchange intensity is insensitive to
680 the horizontal scales examined, and within-group variability dominates (Fig. 9).

681 In addition, the median values of $E_{C,H}/E_{C,VT}$ in Table 5 are very small ($\approx 0.5\%$ –
682 2.4%), indicating that the system at summer noon is primarily vertical-exchange



683 dominated with horizontal transport being nearly negligible. This pattern is readily
 684 explained: under summer-noon conditions, when turbulent mixing is stronger and
 685 horizontal concentration gradients are comparatively weak, the net budget (E_C) of the
 686 control volume is governed mainly by vertical turbulent exchange ($E_{C,VT}$), whereas
 687 horizontal transport ($E_{C,H}$) contributes only a minor fraction. The negative sign of the
 688 ratio suggests that the horizontal and vertical terms often have opposite signs (e.g., the
 689 vertical term indicates uptake and is negative, while the horizontal term may represent
 690 a weak compensating inflow/outflow, yielding a small positive or negative
 691 contribution), but its magnitude remains far smaller than that of the vertical term.



692
 693 **Figure 9.** Distribution of e_C across box sizes (50×50 m, 100×100 m, 150×150 m) of
 694 summer noon. Boxes represent the interquartile range (IQR), the center line is the median, and
 695 whiskers extend to $1.5 \times IQR$. Colored points are individual samples. The blue solid polyline
 696 connects the mean e_C of each group, with boxed numerical annotations.

697
 698 **Table 5.** Summary statistics of net CO₂ emission estimates (E_C) derived from three box sizes
 699 during summer noon and autumn morning. For each box size, n denotes the number of valid
 700 samples. Reported metrics include the mean and median of E_C (kg h^{-1}), the interquartile range
 701 (IQR), and the median ratio $E_{C,H}/E_{C,V}$, where $E_{C,H}$ and $E_{C,V}$ represent the horizontal-transport
 702 and vertical-exchange contributions to the net emission, respectively.

	Box Size (m)	n	E_C Mean (kg h^{-1})	E_C Median (kg h^{-1})	IQR (kg h^{-1})	Median $E_{C,H}/E_{C,V}$
Summer Noon	50×50	7	-2.235	-2.787	2.648	-0.024
	100×100	8	-10.732	-12.722	6.149	-0.005



	150×150	8	-18.148	-13.889	19.859	-0.019
	50×50	7	0.166	0.138	1.849	3.214
Autumn	100×100	8	7.720	7.292	2.875	0.847
Morning	150×150	8	9.278	11.950	14.92	1.213

703

704 Another case study of spatial heterogeneity is based on the autumn-morning
 705 dataset (Fig.10). The mean values in Fig.10 indicate that the emission intensities (e_C)
 706 derived for different box sizes (0.02, 0.21, and 0.11, respectively) differ to some extent.
 707 The differences among box sizes are substantially larger than the uncertainty of the
 708 method used in this study (15%), representing e_C differ of various sizes. One-way
 709 ANOVA analysis also shows the differences in e_C among the three scales were
 710 significant ($F=4.482$, $p=0.025$). This providing direct evidence that the vegetation–
 711 atmosphere CO₂ source-sink distribution in the upper layer is spatially heterogeneous.

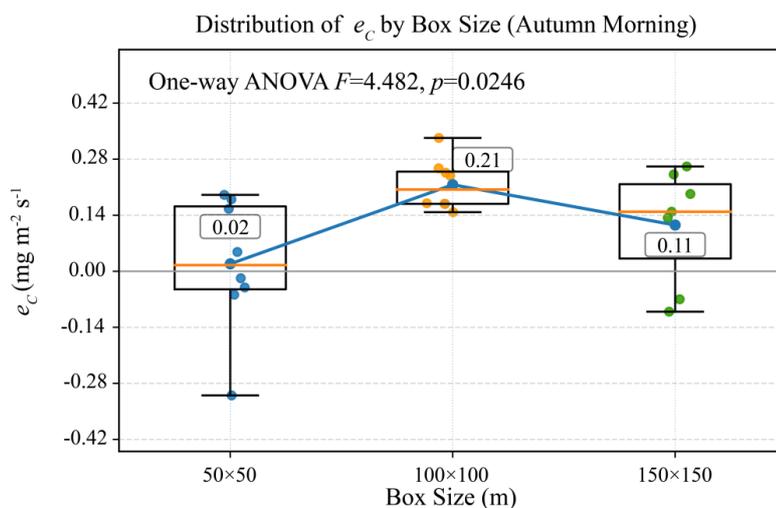
712 In addition, the component-wise contribution analysis (Table 5) indicates that the
 713 relative contribution of the horizontal transport term is markedly enhanced. The median
 714 metric, median $E_{C,H}/E_{C,VT}$, is 3.214 for 50×50, 0.847 for 100×100, and 1.213 for
 715 150×150. These results suggest that there is a significant spatial heterogeneity in the
 716 CO₂ emissions/uptake in the range covered by the different boxes leading to horizontal
 717 CO₂ concentration gradients, and as a result near-surface horizontal CO₂ gradients
 718 under the measured horizontal winds are non-negligible within small control volumes.
 719 As the control-volume scale increases, local “hotspots/coldspots” and horizontal
 720 gradient structures are spatially averaged, weakening the relative influence of $E_{C,H}$
 721 and thereby causing E_C to exhibit scale sensitivity.

722 This scale sensitivity further highlights the true advantage of UAV observations
 723 for characterizing spatial heterogeneity. UAV measurements can provide multipoint,
 724 three-dimensional CO₂ distributions and associated wind speed gradient information
 725 within a short time window, enabling the simultaneous estimation of $E_{C,H}$ (horizontal
 726 transport) and $E_{C,V}$ (vertical transport) within a consistent data framework. Moreover,
 727 by varying the box sizes, the same dataset can be re-aggregated into different spatial
 728 scales to facilitate a multi-scale comparison. In contrast, flux towers (eddy covariance)



729 essentially measure point-based vertical turbulent fluxes, whose representativeness is
 730 determined by a footprint that varies with wind direction and atmospheric stability. As
 731 a result, EC tower observations yield a footprint-weighted effective flux, making it
 732 difficult to construct a spatial control volume and impossible to explicitly resolve the
 733 lateral transport term $E_{C,H}$. Therefore, box-size dependence driven by spatial
 734 heterogeneity is more readily captured by UAV observations, whereas EC towers are
 735 limited in both observation geometry and the types of physical quantities they can
 736 diagnose.

737



738

739

740 **Figure 10.** Distribution of e_c across box sizes (50×50 m, 100×100 m, 150×150 m) of
 741 autumn morning. Boxes represent the interquartile range (IQR), the center line is the median, and
 742 whiskers extend to $1.5 \times IQR$. Colored points are individual samples. The blue solid polyline
 743 connects the mean e_c of each group, with boxed numerical annotations.

743

744 4.4 Computation of uncertainties

745 The uncertainty of the final CO₂ emission estimates was quantified by considering
 746 three independent sources: instrumental errors (δ_M), wind-calibration uncertainty (δ_V),
 747 and thermodynamic variability associated with temperature and pressure (δ_{dens}). These
 748 contributions were combined in quadrature:

$$749 \quad \delta^2 = \delta_M^2 + \delta_V^2 + \delta_{dens}^2 \quad (4.7)$$



750 Table 6 presents all uncertainties of the CO₂ emissions calculated for the morning,
751 noon, and evening periods on selected days in four seasons.

752

753 **Table 6.** Relative uncertainties (%) of CO₂ emissions derived from the UAV for four
754 representative days in different seasons (Spring: 12 May 2024, Summer: 20 August 2023,
755 Autumn: 3 December 2024, and Winter: 26 December 2024). Columns labeled M, N, and E
756 correspond to morning, noon, and evening periods, respectively. The total relative uncertainty δ
757 is obtained by combining the individual contributions in quadrature for each period.

	12 May 2024			20 August 2023			3 December 2024			26 December 2024		
	M	N	E	M	N	E	M	N	E	M	N	E
δ_M	1	2	1	1	4	2	1	2	4	2	5	1
δ_V	6	21	13	8	11	3	5	13	8	9	2	4
δ_{dens}	0	1	0	2	1	0	0	2	0	0	1	0
δ	6	21	13	8	12	4	5	13	9	9	5	4

758

759 4.4.1 Instrumental errors

760 Instrumental errors are comparatively small. The ultrasonic anemometer used on
761 the UAV has a wind speed accuracy of ± 0.2 m s⁻¹ and a wind direction accuracy of $\pm 1^\circ$.
762 For typical flight-level wind speeds of 4-6 m s⁻¹, this corresponds to a relative error of
763 about 5% in the horizontal wind components. The closed-path CO₂ analyzer has a
764 precision of better than 0.2 ppm, whereas the plume enhancements resolved in the box
765 walls frequently exceed tens of ppm (e.g. up to ~ 40 ppm in the morning box on 8 August;
766 Fig. 11), implying a relative CO₂ uncertainty of approximately 0.5%. Propagating these
767 errors through the flux equations leads to an instrumental contribution to the emission
768 rate uncertainty on the order of $\delta_M \approx 5\%$, which may be regarded as a minor term
769 compared to sampling and methodological effects.

770

771 4.4.2 Wind-calibration uncertainties

772 The wind correction algorithm for platform-induced flow distortions and UAV
773 attitude changes introduces an additional source of uncertainty. Our wind-correction
774 algorithm, derived from extensive CFD simulations and in-flight validation (Yang et al.,
775 2025), corrects for rotor downwash, UAV attitude changes, and background wind over
776 a wide flight envelope conditions. Residual biases are most likely during periods of



777 strong turbulence or rapid maneuvers, but the close agreement between UAV-derived
778 friction velocity and EC-tower estimates (Fig. 5b) suggests that any remaining
779 systematic error in wind speed is modest. The uncertainty due to the wind correction
780 algorithm for the E_c term is estimated at 0.025 kg h^{-1} (2.1 % of E_c , minimal impact)
781 for the 12 May 2024 flight and 0.643 kg h^{-1} (13.8 %, maximum impact) for the 3
782 December 2024 flight. A conservative $\delta_V \approx 5\text{-}14\%$ is assigned to this uncertainty term.

783

784 4.4.3 Temperature and pressure impacts

785 The horizontal emission term in Eq. (3.4) depends explicitly on air density, which
786 depends on air temperature and pressure, and we use the time-averaged temperature
787 and pressure over the duration of a box–profile pair to derive a single representative air
788 density for the entire control volume. To assess the impact of temporal variability in
789 thermodynamic conditions, the emission calculations were repeated using the minimum
790 and maximum densities observed during each flight period as alternative scenarios.
791 Table 4 shows that the uncertainties arising from variations in air density and
792 temperature are very small ($\delta_{dens} \approx 1\text{-}2\%$) and can be considered negligible.

793

794 4.4.4 Total estimated uncertainties

795 By combining the uncertainties from the three independent components -
796 instrumental errors (δ_M), wind-calibration uncertainties (δ_V), and thermodynamic
797 (temperature–pressure) variability (δ_{dens}) - the total relative uncertainty of the UAV-
798 derived CO_2 emission rates can be obtained through quadrature addition. As shown in
799 Table 6, the total uncertainty typically falls within 5-15%, with the exact magnitude
800 depending on atmospheric stability, time of day, and season.

801

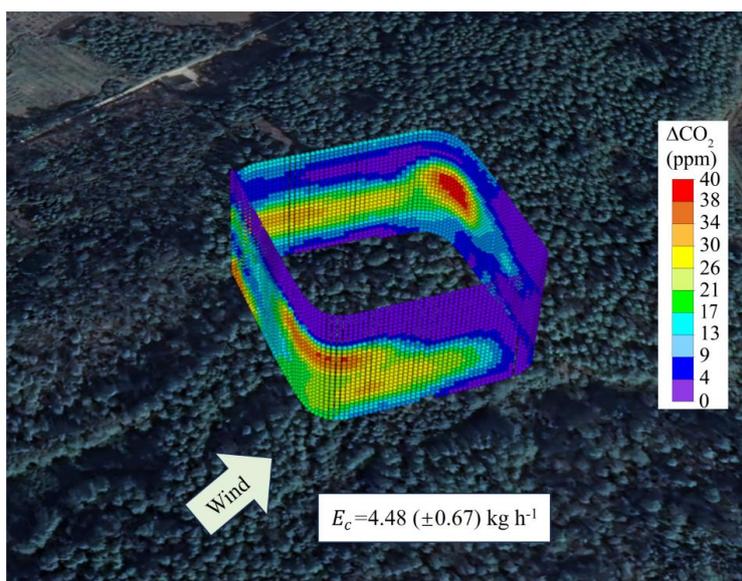
802 4.5 UAV measurement results

803 Figure 11 provides a three-dimensional visualization of the CO_2 enhancement field
804 (ΔCO_2) measured along the walls of the $100 \times 100 \text{ m}$ control volume during the morning
805 flight on 8 August 2024. The reconstructed ΔCO_2 distribution exhibits pronounced



806 spatial heterogeneity, with localized high-concentration patches embedded within
807 relatively low-background regions. In particular, a distinct hotspot reaching ~ 40 ppm
808 is observed near the upper-right sector of the box, while elevated ΔCO_2 values also
809 form elongated bands along portions of the lateral walls. This non-uniform pattern
810 indicates that the near-canopy CO_2 field was far from horizontally homogeneous during
811 the sampling window, consistent with the expectation of enhanced stratification and
812 limited mixing in early morning conditions.

813



814
815 **Figure 11.** Three-dimensional CO_2 concentration distribution measured by the UAV
816 (100×100 m) during the morning measurement on 8 August 2024.
817

818 The wind vector shown in Figure 11 indicates a prevailing southwesterly flow. The
819 spatial alignment of the ΔCO_2 enhancements with the box boundaries suggests that
820 lateral advection and flux divergence played a central role in shaping the observed
821 concentration gradients: CO_2 -enriched air masses are transported into the control
822 volume on the upwind side and redistributed along the walls, producing strong cross-
823 boundary gradients that dominate the mass balance. Consistent with this interpretation,
824 the box-mass balance calculation yields a positive net exchange rate of $E_c = 4.48 \pm 0.67$



825 kg h⁻¹ for the 100×100 m control volume. Breaking down the individual mass balance
826 terms indicates that the horizontal component ($E_{C,H} = 4.22 \pm 0.63$ kg h⁻¹) accounts for
827 nearly all of the net emission in the control volume, whereas the vertical component is
828 small ($E_{C,VT} = 0.26 \pm 0.04$ kg h⁻¹). The near-zero vertical term is also consistent with the
829 concurrent eddy-covariance observations similarly indicating a negligible turbulent
830 vertical CO₂ flux. Therefore, the positive E_c primarily reflects net CO₂ inflow driven
831 by horizontal transport (i.e., lateral convergence) rather than local vertical emission
832 from the canopy.

833 Overall, the three-dimensional Δ CO₂ structure shown in Figure 11 highlights that,
834 at this moment, CO₂ exchange above the forest canopy was controlled predominantly
835 by horizontal transport under heterogeneous concentration fields. This underscores the
836 importance of resolving spatial gradients and lateral boundary fluxes—features that can
837 be captured by UAV sampling but are not directly observable from a single-point flux
838 tower measurement.

839 **5 Conclusions**

840 This study developed and field-evaluated a UAV-based methodology to quantify
841 three-dimensional CO₂ transport over a forest ecosystem by explicitly resolving both
842 vertical exchange and horizontal advection within a three-dimensionally unified mass-
843 balance approach, making use of an integrated UAV measurement platform and a
844 computational algorithm that evaluates both horizontal and vertical fluxes based on a
845 specifically designed flight strategy. Extensive applications of this methodology at the
846 Qianyanzhou subtropical plantation forest site demonstrate that the methodology can
847 robustly determine canopy-scale CO₂ exchange. Vertical CO₂ fluxes derived from UAV
848 vertical profiles show excellent agreement with those determined from collocated eddy
849 covariance (EC) tower observations across seasons, indicating that the UAV-based
850 gradient method reproduces both the magnitude and variability of vertical turbulent
851 exchange under typical daytime conditions. The UAV-based vertical fluxes showed that
852 the diurnal and seasonal patterns over the forest ecosystem at Qianyanzhou - strongest



853 uptake around midday during the growing season and near-neutral exchange in winter
854 – are ecologically consistent with long-term site behavior. However, vertical fluxes of
855 UAV and EC diverge under weak turbulence during the evening stable period,
856 highlighting increased uncertainties for both profile-based retrievals and tower-based
857 estimates under low vertical flux conditions.

858 Box-pattern UAV measurements further reveal that horizontal transport can be a
859 dominant term under certain conditions. Enhanced CO₂ concentrations on the virtual
860 walls of the flight boxes and wind vectors show pronounced spatial heterogeneity and
861 lateral gradients, especially in the morning, indicating that advection-driven
862 convergence/divergence may substantially shape the local CO₂ budget, processes that
863 cannot be directly resolved from single-tower measurements. A multi-scale analysis
864 using different box sizes (50×50, 100×100, and 150×150 m) shows that net emission
865 estimates can exhibit scale dependence, reflecting the spatial heterogeneity of
866 sources/sinks and the changing relative contribution of horizontal transport with
867 control-volume size. These results highlight the value of UAV observations for
868 determining spatial heterogeneity in short term ecosystem carbon exchange with the
869 atmosphere, a task unfit for single-site measurements, for diagnosing when and where
870 the assumption of horizontal homogeneity becomes invalid, and for quantifying lateral
871 flux contributions explicitly.

872 Overall, the combined box–profile UAV framework provides a flexible and
873 spatially explicit complement to conventional EC monitoring, offering an effective
874 pathway to bridge the observational gap between point-scale tower fluxes and larger-
875 scale heterogeneities, particularly over heterogeneous canopies and complex terrain. In
876 conclusion, UAV-based platforms offer a transformative opportunity for observing
877 ecosystem-atmosphere CO₂ exchange in three dimensions. With continued
878 methodological refinement and broader application, such systems are well positioned
879 to become an essential component of next-generation carbon observation strategies in
880 forest ecosystems and beyond.

881



882 *Data availability.* Data are available upon request to the corresponding author upon
883 reasonable request.

884

885 *Author contributions.* HS built the unmanned aerial vehicle platform, developed the
886 corresponding emission inversion model, and carried out the data processing, and
887 drafted the initial version of the manuscript. SML supervised the development of the
888 inversion model and the experimental validation in the paper, and carried out a second
889 round of revisions to the manuscript. YW provided some key data used in the paper.
890 YY and TH provided the wind speed correction method used in the inversion model
891 and the validation results for horizontal emissions, respectively. KC, CL, and JZ
892 provided assistance with the field observation experiments and revisions to the
893 manuscript. ZM and FY provided all flux tower data from the Qianyanzhou station. JG
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895

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909

910



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