



1 **Unappreciated role of sulfate radicals in the aqueous aging**
2 **process of methoxyphenols derived from biomass burning**

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15 **Abstract:** Methoxyphenols are widely acknowledged tracers of biomass burning
16 emissions, undergo complex chemical transformations in atmospheric aqueous
17 environments that significantly modulate aerosol properties. The role of sulfate radicals
18 ($\text{SO}_4^{\cdot-}$), the highly potent electrophiles prevalent in cloud and fog waters, remains
19 poorly constrained. Herein, we investigated the aqueous-phase kinetics and aqueous
20 secondary organic aerosol (aqSOA) formation of three representative methoxyphenols
21 containing different substituents, namely vanillic acid (VAA), vanillin (VAL), and



22 coniferyl aldehyde (CFA), upon oxidation by $\text{SO}_4^{\bullet-}$. The determined second-order rate
23 constants for VAA, VAL, and CFA with $\text{SO}_4^{\bullet-}$ were in the range of $(2.40\text{-}3.37) \times 10^9 \text{ M}^{-1}$
24 s^{-1} , revealing that side-chain substituents critically govern reactivity through π -
25 electron density modulation. These reactions efficiently generated aqSOA with
26 maximum mass yields of 63.12%-70.49%, characterized by high oxidation degree
27 comparable to that of atmospheric low-volatility oxygenated organic aerosols (LV-
28 OOA). Notably, the oxidation process driven the generation of humic-like substances
29 (HULIS) with pronounced light-absorbing capabilities in the near-ultraviolet and
30 visible regions, thereby contributing to atmospheric brown carbon. Furthermore, the
31 oxidation products exhibited significantly elevated oxidative potentials compared to
32 their precursors, posing enhanced health risk. Our findings identify $\text{SO}_4^{\bullet-}$ -initiated
33 aqueous chemistry as a critical yet previously overlooked pathway that transforms
34 biomass burning emissions into toxic and light-absorbing secondary aerosols.
35 Therefore, it is necessary to incorporate $\text{SO}_4^{\bullet-}$ -related processes into atmospheric
36 models to accurately predict air quality and climate forcing.

37 **Keywords:** Substituted methoxyphenols, Sulfate radicals, Aqueous-phase reaction,
38 Rate constant, Secondary organic aerosol, Optical property, Oxidation potential



1 Introduction

Sulfate radicals ($\text{SO}_4^{\bullet-}$) are important atmospheric oxidants that exclusively exist in atmospheric aqueous media (Herrmann et al., 2010, 2015). Its estimated average concentration within cloud droplets and deliquescent particles spans 3.6×10^{-13} to $2.3 \times 10^{-15} \text{ mol L}^{-1}$, which is comparable to those of hydroxyl radicals ($^{\bullet}\text{OH}$) and nitrate radicals ($\text{NO}_3^{\bullet}$) (Herrmann et al., 2010). Given its high oxidation potential (2.5-3.1 V) and persistent reactivity over a broad pH range (2-9) (Liu et al., 2018). $\text{SO}_4^{\bullet-}$ can efficiently degrade a wide variety of soluble organic pollutants in atmospheric aqueous environments, including aromatics, alcohols, carboxylic acids, ethers, etc. (Herrmann et al., 2010, 2015; Rudziński and Szmigielski, 2019; Wojnárovits and Takács, 2019), and their corresponding second-order rate constants fall within the range of 10^7 to $10^9 \text{ M}^{-1} \text{ s}^{-1}$. In the past, it is well established that $\text{SO}_4^{\bullet-}$ is primarily generated via the autoxidation chain reaction of sulfur dioxide, alongside minor formation pathways involving the activation of sulfate or bisulfate ions by $^{\bullet}\text{OH}$, $\text{NO}_3^{\bullet}$, or iron species (Herrmann et al., 2010). In recent years, additional formation pathways of $\text{SO}_4^{\bullet-}$ in atmospheric aqueous media have been uncovered, chiefly including irradiation of sulfate-containing organic aerosols, $^{\bullet}\text{OH}$ -initiated oxidation of organosulfates (OSs), and autocatalytic reactions of sulfate anion (SO_4^{2-}) (Cope et al., 2022; Kwong et al., 2018; Pan et al., 2019). To date, the potentially critical role of $\text{SO}_4^{\bullet-}$ remains poorly elucidated. Therefore, more systematic investigations into $\text{SO}_4^{\bullet-}$ -initiated reactions of representative organic pollutants are urgently required, particularly to fill the gaps



60 regarding the corresponding kinetics and secondary organic aerosol (SOA) formation.

61 Methoxyphenols are broadly acknowledged as reliable tracers of biomass burning

62 emissions (Nolte et al., 2001; Liu et al., 2022). Formed during the pyrolysis process of

63 lignin, such species have been extensively identified within atmospheric gas, particulate,

64 and aqueous phases (Liu et al., 2022). 2-Methoxyphenol and 2,6-dimethoxyphenol

65 represent the basic structures of methoxyphenols, with their Henry's law constants

66 ranging from 10^3 to 10^4 M atm⁻¹ (Arciva et al., 2022). In contrast, substituted

67 methoxyphenols exhibit far larger Henry's law constants (10^6 - 10^9 M atm⁻¹) and

68 partition substantially into atmospheric aqueous phase, even under aerosol liquid water

69 conditions (McFall et al., 2020). In addition, the measured concentration of

70 methoxyphenols in fog water is 3-4 times higher than the values predicted by Henry's

71 law (Kitanovski et al., 2014). Therefore, aqueous-phase reactions of these compounds

72 constitute a critical pathway for atmospheric chemical transformation, especially in

73 regions affected by biomass burning. In recent years, aqueous-phase reactions of

74 methoxyphenols with [•]OH and triplet excited states (³C*) have garnered growing

75 attention within the research community. The existing relevant work has primarily

76 reported reaction rate constants, aqueous-phase SOA (aqSOA) formation, oxidation

77 products, and the underlying mechanisms (An et al., 2020; Arciva et al., 2022; Chen et

78 al., 2020; Huang et al., 2018; Lai et al., 2025; Li et al., 2021; Loisel et al., 2021; Ma et

79 al., 2021; Pang et al., 2019; Smith et al., 2014, 2015; Tang et al., 2020; Yang et al., 2023,

80 2024; Ye et al., 2020; Yu et al., 2016). However, aqueous-phase reactions of substituted



81 methoxyphenols with $\text{SO}_4^{\bullet-}$ have been scarcely reported even if they have potential
82 coexistence aqueous environment in the atmosphere (Herrmann et al., 2010; Kitanovski
83 et al., 2014). In particular, the corresponding kinetic data and possible aqSOA formation
84 are not well characterized.

85 To address the aforementioned knowledge gaps, here we investigated aqueous-
86 phase reactions of $\text{SO}_4^{\bullet-}$ with three representative methoxyphenols containing different
87 substituents on the 2-methoxyphenol backbone, namely vanillic acid (VAA), vanillin
88 (VAL), and coniferyl aldehyde (CFA), focusing on kinetics, aqSOA mass yields, optical
89 properties, and oxidation potentials. The findings of this work provide new insights into
90 the aqueous-phase chemical transformation of methoxyphenols in the atmosphere and
91 furnish updated kinetic and aqSOA mass yield datasets for application in atmospheric
92 modeling.

93 **2 Experimental section**

94 **2.1 Aqueous-phase reactions**

95 Fresh solution of 0.3 mM methoxyphenol sample was prepared using deionized
96 water. 300 mL of sample solution was introduced into a 500 mL cylindrical glass reactor
97 (Fig. S1), and its initial pH was adjusted to 2 by diluted sulfuric acid and sodium
98 hydroxide. Subsequently, 50 mM peroxydisulfate (PDS, $\text{Na}_2\text{S}_2\text{O}_8$) was added and
99 applied to produce $\text{SO}_4^{\bullet-}$ in this system via gently thermal decomposition at 50 °C.
100 Meanwhile, 50 mM *tert*-butyl alcohol (TBA) was added in order to prevent the possible
101 interference of $^{\bullet}\text{OH}$. Lastly, the aqueous-phase reactions of methoxyphenols with $\text{SO}_4^{\bullet-}$



102 were initiated. The solution was first stirred at 450 rpm for 2 min, after which stirring
103 was maintained at 50 rpm for the entire duration of the experiment. At predetermined
104 time intervals, solution samples were collected for the following analysis. In addition,
105 the possible impacts of temperature and ion strength on aqueous-phase reactions were
106 not taken into account in this work and discussed in detail in the Supplement.

107 **2.2 Relative rate method**

108 The relative rate method was used to determine the second-order rate constants of
109 VAA, VAL, and CFA in the aqueous-phase reactions with $\text{SO}_4^{\bullet-}$. The corresponding
110 equation is listed as follows: $\ln\left(\frac{C_{M0}}{C_{Mt}}\right) = \frac{k_{2,M}}{k_{2,R}} \ln\left(\frac{C_{R0}}{C_{Rt}}\right)$, where C_{M0} and C_{R0} are the initial
111 concentrations of methoxyphenol sample and reference compound, respectively; C_{Mt}
112 and C_{Rt} are their concentrations at reaction time t . $k_{2,M}$ and $k_{2,R}$ are the second-order rate
113 constants of methoxyphenol sample and reference compound. In this work, ferulic acid
114 was adopted as the reference compound, which is a typical kind of methoxyphenols
115 with a second-order rate constant of $3.2 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$ toward $\text{SO}_4^{\bullet-}$ (Herrmann et al.,
116 2010). In the experiments, methoxyphenols and ferulic acid were both used at an
117 identical initial concentration of 0.3 mM, with all other experimental conditions
118 consistent with those described in the preceding section.

119 **2.3 Analytical methods**

120 A double-beam UV-vis spectrophotometer (M9, Shanghai Mapada, China) was
121 employed to record the UV-vis absorption spectra of solution samples within the
122 wavelength range of 200-800 nm at predetermined intervals. Deionized water was used



123 as reference solution. During data processing, the signals of absorption spectra were
124 corrected to eliminate the possible background interferences caused by TBA and
125 inorganics in the solution. The fluorescence spectrophotometer (F-7000, Hitachi, Japan)
126 was adopted to conduct three-dimensional excitation-emission matrix (EEM)
127 fluorescence spectral scanning of solution samples. Its emission wavelength scanning
128 range was 220-700 nm with a scanning step size of 5 nm, and corresponding excitation
129 wavelength scanning range was 200-600 nm with a scanning step size of 10 nm. The
130 scanning speed was 2400 nm min⁻¹. The concentrations of methoxyphenol sample and
131 ferulic acid were measured by gas chromatography-mass spectrometer (GC-MS)
132 analysis. Before analysis, solution samples were extracted using CH₂Cl₂ and then
133 derivatized by *N,O*-bis(trimethylsilyl)trifluoroacetamide. Humic-like substances
134 (HULIS) were isolated from reaction products using solid-phase extraction cartridges
135 (CNW Poly-Sery HLB) and quantified via high performance liquid chromatography
136 (HPLC) analysis. In addition, an improved dithiothreitol (DTT) assay method was
137 adopted to analyze the oxidation potential of reaction products (Li et al., 2022). 2-Nitro-
138 5-thiobenzoic acid (TNB) produced from the reaction between DTT and 5,5'-dithiobis
139 (2-nitrobenzoic acid) (DNTB) has a characteristic absorption at 412 nm. Thus, the
140 consumption rate of DTT can be calculated by its corresponding changes in absorbance.
141 An Aerodyne high-resolution time-of-flight aerosol mass spectrometer (HR-ToF-AMS)
142 was applied to characterize the mass concentration and chemical composition of aqSOA
143 (Decarlo et al., 2006). The mass data acquired from the HR-ToF-AMS were analyzed



144 using the toolkit SQUIRREL 1.57I/PIKA 1.61I version in Igor Pro version 6.37.
145 Detailed descriptions of the UV-vis, EEM, GC-MS, DTT, and HULIS measurements,
146 HR-ToF-AMS analysis, and the calculation of aqSOA mass yields are provided in the
147 Supplement.

148 **3 Results and discussion**

149 **3.1 Rate constants**

150 The second-order rate constants for VAA, VAL, and CFA in the aqueous-phase
151 reactions with $\text{SO}_4^{\bullet-}$ were determined by the relative rate method in this work. As shown
152 in Fig. 1, data points were well fitted using a linear regression ($R^2 > 0.97$). Table 1 lists
153 the slopes, rate constants, and corresponding atmospheric lifetimes (τ). The
154 uncertainties of the slopes (i.e., $k_{2,M}/k_{2,R}$) correspond to the standard deviations obtained
155 from linear regression analysis. As shown in Table 1, the second-order rate constants
156 for VAA, VAL, and CFA were determined to be $(2.40 \pm 0.19) \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$, (2.56 ± 0.10)
157 $\times 10^9 \text{ M}^{-1} \text{ s}^{-1}$, and $(3.37 \pm 0.22) \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$, respectively. This observation indicates
158 that the rate constants are strongly affected by variations in the side-chain substituents
159 of 2-methoxyphenol backbone, as detailed below. Table 2 summarizes $k_{2,M}$ values for
160 aqueous-phase reactions of methoxyphenols with $\text{SO}_4^{\bullet-}$, $\cdot\text{OH}$, and $^3\text{C}^*$, compiled from
161 published literatures. The $k_{2,M}$ values of methoxyphenols reacting with $\text{SO}_4^{\bullet-}$ and $^3\text{C}^*$
162 are comparable and on the order of $10^9 \text{ M}^{-1} \text{ s}^{-1}$, approximately 2-5 times lower than
163 those for reactions with $\cdot\text{OH}$. In addition, the $k_{2,M}$ values obtained in this work are
164 comparable to those reported for other aromatic compounds (e.g., nitrophenol and



165 benzaldehyde) reacting with $\text{SO}_4^{\cdot-}$ and $\text{NO}_3^{\cdot}$ (Herrmann et al., 2010). Therefore, $\text{SO}_4^{\cdot-}$
166 -initiated degradation reasonably serves as an important transformation pathway of
167 methoxyphenols in atmospheric aqueous environments.

168 $\text{SO}_4^{\cdot-}$ is an effective electrophilic oxidant and capable of undergoing single electron
169 transfer (SET), addition to a double bond, and hydrogen abstraction from aliphatic parts
170 of organic molecules (Luo et al., 2017; Neta et al., 1988; Tong et al., 2020; Wojnárovits
171 and Takács, 2019). During reactions with organic molecules containing C=C bond,
172 radical adducts may potentially form, but no such adducts have been detected to date
173 (Paul et al., 2014; Wojnárovits and Takács, 2019). However, SET is overwhelmingly
174 dominant mechanism for its reactions with phenolic aromatics, simultaneously
175 accompanied by the formation of SO_4^{2-} and radical cations of aromatic molecules (Luo
176 et al., 2017; Tong et al., 2020; Wojnárovits and Takács, 2019). Therefore, the reaction
177 rate constant strongly depends on the π -electron density and the extent of π -conjugation
178 of organic molecules. In this work, VAA, VAL, and CFA share an identical 2-
179 methoxyphenol backbone, and their structural differences arise from distinct side-chain
180 substituents attached to the same C position on the benzene ring. CFA bears a –
181 $\text{CH}=\text{CH}-\text{CHO}$ side chain with a Hammett substituent constant (σ_p) of approximately
182 +0.31, which forms an extended conjugated system spanning the aromatic ring, C=C
183 bond, and terminal $-\text{CHO}$ group (Gould, 1959). The expanded delocalization of π
184 electrons, together with the reactive C=C bond serving as an additional electrophilic
185 attack site, leads CFA to become most reactive toward $\text{SO}_4^{\cdot-}$. By contrast, VAA and



186 VAL possess -COOH and -CHO side chains, respectively. Both substituents act as
187 strong electron-withdrawing groups, with comparable σ_p values (+0.45 for -COOH and
188 +0.42 for -CHO) (Jaffé, 2002), and markedly reduce the π -electron density of the
189 aromatic ring (Chataigner et al., 2007; Pang et al., 2019). Consequently, the SET
190 pathway initiated by $\text{SO}_4^{\bullet-}$ is suppressed, which results in low and similar rate constants
191 for VAA and VAL. Consistent with these observations, Li et al. (2021) and He et al.
192 (2019) have recently reported that electron-withdrawing substituents (e.g., -COOH) on
193 methoxyphenols markedly reduce their reactivity toward $\text{OH}^\bullet$ under aqueous-phase
194 conditions, thereby decreasing the corresponding rate constants. On the contrary,
195 electron-donating substituents (e.g., -OCH_3 and -OH) attached to methoxyphenols
196 elevate the π -electron density of the aromatic ring, facilitating their aqueous-phase
197 reactions (He et al., 2019; Li et al., 2021; Jain et al., 2026).

198 **3.2 Aqueous-phase SOA formation**

199 Earlier studies have reported that aqueous-phase reactions of representative
200 methoxyphenols with $\text{OH}^\bullet$ and $^3\text{C}^*$ produce substantial low-volatility products and
201 possess high aqSOA mass yields (Huang et al., 2018; Ma et al., 2021; Smith et al., 2014,
202 2015; Ye et al., 2020; Yu et al., 2016). To clarify whether $\text{SO}_4^{\bullet-}$ -initiated reactions of
203 VAA, VAL, and CFA also produce considerable amounts of aqSOA, time-dependent
204 variations in aqSOA mass yields were investigated in this work and shown in Fig. 2a.
205 It is noticeable that VAA, VAL, and CFA had significant aqSOA formation potentials,
206 and their corresponding aqSOA mass yields at 30 min were 66.41%, 70.49%, and



207 63.12%, respectively. As illustrated in Fig. S2, aqSOA mass yields for five
208 methoxyphenols (i.e., vanillyl alcohol, guaiacyl acetone, ferulic acid, syringic acid, and
209 syringyl acetone) upon reactions with $\cdot\text{OH}$ range from approximately 67% to 90% (Ma
210 et al., 2021). By contrast, their corresponding yields from reactions with $^3\text{C}^*$ vary
211 between 59% and 99% (Arciva et al., 2022). In addition, the maximum aqSOA mass
212 yields for 4-methylsyringol and eugenol in reactions with $\cdot\text{OH}$ can exceed 150%,
213 followed by significant declines as the reactions proceed (Ye et al., 2020). Overall,
214 aqSOA mass yields obtained in this work are comparable to those for other
215 methoxyphenols in aqueous-phase reactions with $\cdot\text{OH}$ and $^3\text{C}^*$. Thus, aqueous-phase
216 reactions between methoxyphenols and $\text{SO}_4^{\cdot-}$ constitute a non-negligible pathway for
217 aqSOA formation and should deserve greater attention in future studies.

218 As shown in Fig. 2a, the rapid aqSOA formation observed in the aqueous-phase
219 reactions of the three methoxyphenols can be reasonably attributed to the efficient
220 generation of substantial low-volatility products, including functionalized derivatives
221 and oligomers (Arciva et al., 2022; Huang et al., 2018; Smith et al., 2014; Yu et al.,
222 2016). In addition, a pronounced decrease in aqSOA mass yields was observed as the
223 reaction time extends to 180 min, with corresponding yields of 24.20% for VAA, 23.59%
224 for VAL, and 19.11% for CFA. This can be explained by the fact that fragmentation
225 processes dominated the reactions during this period, leading to the substantial
226 formation of highly volatile fragmented products (Chen et al., 2020; Huang et al., 2018;
227 Ye et al., 2020). Such species are difficult to retain in the particulate phase, making



228 them unfavorable for aqSOA formation and unable to be efficiently detected by the HR-
229 ToF-AMS (Decarlo et al., 2006). Consistent trends have also been reported for aqueous-
230 phase reactions of 4-methylsyringol and eugenol with $\cdot\text{OH}$ and $^3\text{C}^*$ (Chen et al., 2020;
231 Ye et al., 2020). In our recent study, mild declining trends in aqSOA mass yields were
232 also observed for guaiacol and syringol upon reactions with $\text{SO}_4^{\cdot-}$ (Chen et al., 2026).

233 To further unravel variation trend of aqSOA composition along with aging process,
234 time-dependent changes of element ratios (H/C and O/C) and fractions of organic ion
235 groups including CH, CHO, and CHOgt1 (containing more than one oxygen atom) were
236 investigated. Meanwhile, the average carbon oxidation state ($\text{OS}_\text{C} = 2\text{O/C} - \text{H/C}$) was
237 calculated based on the element ratios of aqSOA determined by the HR-ToF-AMS.
238 Compared with the single O/C ratio, OS_C is widely regarded as a more accurate
239 indicator of the oxidation degree of atmospheric organic matters, because it accounts
240 for the saturation state of carbon atoms in SOA constituents (Kroll et al., 2011). As
241 shown in Figs. 2b and S3, the O/C ratios and OS_C values of aqSOA formed from the
242 aqueous-phase reactions of the three methoxyphenols with $\text{SO}_4^{\cdot-}$ exhibited a gradual
243 increasing trend, whereas the H/C ratios displayed an opposite trend as the reactions
244 proceeded. Similar observations have been reported in earlier studies about aqueous-
245 phase reactions of other five methoxyphenols (i.e., eugenol, 4-ethylguaiacol, 4-
246 methylsyringol, syringaldehyde, and acetosyringone) with $\cdot\text{OH}$ and $^3\text{C}^*$ (Chen et al.,
247 2020; Huang et al., 2018; Ye et al., 2020). The maximum O/C ratios and OS_C values
248 were within the ranges of 0.86-1.20 and 0.53-1.43, respectively, which fall within the



249 corresponding observed ranges for ambient low-volatility oxygenated organic aerosols
250 (LV-OOA) (Ye et al., 2020; Kroll et al., 2011; Yu et al., 2016). The pronounced increases
251 in O/C ratios and OS_C values indicate continuous oxidation of aqSOA constituents with
252 the prolongation of reactions, along with the formation of more oxygen-containing
253 functional groups in the products (Chen et al., 2020). The oxidation degree of aqSOA
254 compositions were also well supported by the variations in mass spectra of aqSOA at
255 different reaction time obtained by the HR-ToF-AMS. As shown in Fig. S4, the
256 fractions of CH groups in aqSOA constituents decreased substantially for VAA, VAL,
257 and CFA, falling from 35.2%, 63.0%, and 41.1% at 30 min to 15.8%, 36.0%, and 30.2%
258 at 360 min, respectively. In contrast, significant increases were observed in the total
259 fractions of CHO and CHOgt1 groups, rising from 64.8%, 37.0%, and 58.9% to 84.2%,
260 64.0%, and 69.8% for VAA, VAL, and CFA, respectively. In addition, we found that
261 aqSOA mass yields declined with increasing OS_C values. This phenomenon is
262 consistent with previous reports, which demonstrate that higher OS_C values correspond
263 to more extensive aging of SOA (Ortega et al., 2016). Such aging process typically
264 generates abundant highly volatile fragmented products via the C–C bond scission of
265 low-generation products and thereby causes an overall reduction in SOA mass (Ortega
266 et al., 2016; Ye et al., 2020; Huang et al., 2018).

267 The Van Krevelen (VK) triangle diagram of aqSOA is depicted in Fig. 2c, revealing
268 that the H/C and O/C ratios of aqSOA constituents clustered within a narrow slope
269 range from -1 to -0.5 . This trend is consistent with previous observations for



270 atmospheric OA and aqueous-phase reactions of methoxyphenols with $\cdot\text{OH}$ (Heald et
271 al., 2010; Huang et al., 2018; Chhabra et al., 2011). These results suggest carboxylic
272 acid or equal amounts of hydroxy and carbonyl functional groups are generally added
273 during the reaction process (Heald et al., 2010). Meanwhile, the data acquired after 120
274 min fall within the characteristic region of LV-OOA in VK-triangle diagram (Ng et al.,
275 2011). Given that VAA, VAL, and CFA contain unsaturated carbon chains, the
276 substantial loss of hydrogen suggests their extensive oxidation by $\text{SO}_4^{\cdot-}$. In general,
277 aging trajectories shown in VK-triangle diagram are dependent on precursor species
278 (e.g., VAL and isoprene) and its classes (e.g., biogenic and anthropogenic volatile
279 organic compounds) (Chhabra et al., 2011).

280 **3.3 Light-absorbing property**

281 Biomass burning emissions are widely recognized as an important source of
282 atmospheric brown carbon (BrC) (Li et al., 2023), which absorbs incoming solar
283 shortwave radiation predominantly in the near-ultraviolet and short visible wavelengths,
284 thereby perturbing Earth's radiative budget (Laskin et al., 2025). In this work,
285 pronounced color changes were observed for the solution samples, all of which turned
286 from transparent to brownish-yellow following the reactions, shown in Fig. S5.
287 Therefore, time-series variations in UV-vis spectra of solution samples were analyzed.
288 As shown in Fig. 3, CFA displayed a characteristic absorption peak at 340 nm. In
289 contrast, VAA and VAL exhibited their characteristic peaks at 260 nm and 292 nm, and
290 280 nm and 308 nm, respectively. These absorptions are resulted from $\pi\text{-}\pi^*$ ($> 180 \text{ nm}$)



291 and $n-\pi^*$ (270-350 nm) electronic transitions of benzene ring and its substituted groups
292 (Li et al., 2021; Tang et al., 2020; Santos et al., 2016). The substitutes are either
293 electron-withdrawing groups (i.e., C=C, -CHO, and -COOH) or electron-donating
294 groups (i.e., -OH and -OCH₃) (Santos et al., 2016). The absorbance of the characteristic
295 peaks for VAA and VAL first increased and then declined to nearly zero as the reaction
296 proceeded up to 360 min. Similar trends have been documented in previous
297 investigations regarding $\cdot\text{OH}$ -initiated aqueous oxidation of 4-methylsyringol and
298 2,4,6-trimethylphenol (Ye et al., 2020). This phenomenon can be interpreted as a
299 combined effect between the degradation of parent compounds and the generation of
300 complex products (e.g., oligomerized and functionalized species) (Ye et al., 2020). In
301 contrast, the absorbance of CFA at 340 nm decreased steadily over the entire reaction
302 period, with an obvious drop observed after 60 min. This behavior is likely attributable
303 to the degradation of CFA by $\text{SO}_4^{\cdot-}$ (Ye et al., 2020).

304 As illustrated in Fig. 3, the overall spectral profiles of VAA, VAL, and CFA
305 changed markedly over the corresponding reactions. Notably, absorbance rose within
306 the near-ultraviolet and visible regions (320-420 nm), where the three methoxyphenols
307 exhibited no light absorption. The results confirm the generation of new light-absorbing
308 products. Similar redshifts in absorption peaks of VAA, VAL, and other
309 methoxyphenols (e.g., syringic acid and guaiacol) have also been observed during their
310 aqueous-phase reactions with $\cdot\text{OH}$ (Li et al., 2021; Tang et al., 2020; Yang et al., 2023).
311 Furthermore, the spectral profiles obtained in this work are consistent with those of



312 atmospheric HULIS (Baduel et al., 2010; Fan et al., 2016). Therefore, aqueous-phase
313 reactions of methoxyphenols with $\text{SO}_4^{\cdot-}$ serve as a potential source of HULIS in
314 atmospheric aqueous environments. The strong absorbance of HULIS at longer
315 wavelengths mainly originates from highly condensed, high-molecular-weight
316 aromatic structures (Chen et al., 2002). In addition, all absorbance at longer
317 wavelengths (320 nm for VAA, 356 nm for VAL, and 410 nm for CFA) initially
318 increased and subsequently declined over the course of the reactions. Concurrently, the
319 color of the reaction solutions gradually lightened over time. The observations indicate
320 that novel chromophores were first formed and gradually degraded in the ensuing
321 reactions (Li et al., 2021; Saleh et al., 2013; Santos et al., 2016; Tang et al., 2020; Wong
322 et al., 2017; Zhao et al., 2015). Analogous trends have been reported for the aqueous-
323 phase reactions of VAA, VAL, and syringic acid with $\cdot\text{OH}$ (Li et al., 2021; Santos et al.,
324 2016; Tang et al., 2020), as well as guaiacol and syringol with $\text{SO}_4^{\cdot-}$ (Chen et al., 2026).
325 Additionally, this evolutionary trend resembles the photochemical aging process of BrC
326 emitted from biomass burning (Saleh et al., 2013; Wong et al., 2017; Zhao et al., 2015).

327 To further characterize the fluorescence property of solution samples, fluorescence
328 spectroscopy was employed herein, owing to its excellent performance in identifying
329 atmospheric HULIS (Fan et al., 2016). As shown in Fig. S6, the initial EEM spectra of
330 VAA, VAL, and CFA solutions exhibited characteristic fluorescence bands centered at
331 $\text{Ex/Em} = (200\text{--}300)/(300\text{--}400)$ nm, $\text{Ex/Em} = (200\text{--}350)/(320\text{--}480)$ nm, and $\text{Ex/Em} =$
332 $(250\text{--}300)/(350\text{--}450)$ nm, respectively. Such features align with the representative



333 fluorescence characteristics of HULIS in smoke particles (Fan et al., 2016). Following
 334 reactions, the fluorescence bands of the three methoxyphenols redshifted. New distinct
 335 fluorescence bands for VAA, VAL, and CFA emerged at $Ex/Em = (250–400)/(300–500)$
 336 nm, $(250–350)/(320–520)$ nm, and $(280–380)/(380–500)$ nm, respectively. Such
 337 redshifts in fluorescence bands are generally linked to the generation of more complex
 338 products featuring greater conjugation (Chang and Thompson, 2010). Compared with
 339 the fluorescence regions identified in biomass-burning smoke $PM_{2.5}$, the post-reaction
 340 fluorescence peaks measured in this work fall within the typical HULIS region (C peak)
 341 (Fan et al., 2016; Tang et al., 2020). Furthermore, Chen et al. (2016, 2019) have reported
 342 that atmospheric organic aerosol-associated HULIS exhibit characteristic fluorescence
 343 emission spanning from 400 nm to 500 nm.

344 Ng et al. (2010) have reported that the mass spectral signatures of atmospheric LV-
 345 OOA closely resemble those of HULIS, because highly oxidized LV-OOA comprises
 346 polycarboxylic acids or acid-derived structures analogous to HULIS. Given the above-
 347 mentioned oxidation degree of aqSOA as well as the VK-triangle diagram results,
 348 aqSOA formed from the three methoxyphenols in this work can be classified as LV-
 349 OOA. In AMS mass spectra, the dominant fragment ions for HULIS are $m/z = 44$,
 350 along with other minor fragments (e.g., $m/z = 29$ and 40) (Paglione et al., 2014;
 351 Hawkins et al., 2016). Accordingly, time series profiles of the organic mass fractions of
 352 $m/z = 44$ (f_{44}) and $m/z = 29$ (f_{29}) are presented in Fig. S7. The gradual increases in both
 353 f_{44} and f_{29} suggest that progressive accumulation of HULIS-like components within the



354 products as the reactions progressed. This observation is well supported by the time-
355 dependent trends in HULIS concentrations (Fig. S8). HULIS concentrations increased
356 gradually and reached 17.91 mg L⁻¹ for VAA, 18.67 mg L⁻¹ for VAL, and 25.78 mg L⁻¹
357 for CFA at 240 min, respectively, before remaining relatively stable for the subsequent
358 reaction period. In this work, the HULIS fraction isolated via polymer-based solid-
359 phase extraction is primarily composed of low-molecular-weight organic acids, fulvic
360 acids, and other humic substances (Li et al., 2022). According to the above observations
361 and discussion, the aqueous-phase reactions of the three methoxyphenols with SO₄^{•-}
362 are a potential source of HULIS in atmospheric aqueous environments.

363 **3.4 Oxidation potential**

364 Oxidation potential describes the capacity of substances to generate reactive
365 oxygen species (ROS) or deplete antioxidants, and is widely adopted as a metric to
366 assess biological oxidative stress (Bates et al., 2019). In general, DTT assay is
367 commonly used to characterize the oxidation potential level of reaction products,
368 because its consumed amounts are directly proportional to the quantity of redox-active
369 substances (Jiang et al., 2019). As shown in Fig. 4a, the DTT consumed mass (M_{DTT})
370 of VAA, VAL, CFA, and blank samples exhibited linear correlations with incubation
371 time ($R^2 > 0.95$). The corresponding consumption rates (R_{DTT}) represent the oxidation
372 potential, which were 0.069 μM min⁻¹ for VAA, 0.073 μM min⁻¹ for VAL, 0.079 μM
373 min⁻¹ for CFA, and 0.041 μM min⁻¹ for blank sample, respectively. The observed M_{DTT}
374 value of blank sample was likely due to dissolved oxygen in deionized water (Li et al.,



2022). The differences in R_{DTT} values of the three methoxyphenols shown in Fig. 4a are attributed from the distinct substituents on the 2-methoxyphenol backbone. As described in Section 3.1, CFA is more susceptible to oxidation than VAA and VAL, which accounts for its higher R_{DTT} value. This trend is consistent with their second-order rate constants. As illustrated in Fig. 4b, the DTT activity of the three methoxyphenols followed a similar trend, which rose rapidly at the initial stage of the reactions and then gradually declined as the reactions proceeded. Overall, the DTT activity measured over the course of reactions is higher than that of parent compounds. The findings indicate that oxidation products generated from aqueous-phase reactions of methoxyphenols with $SO_4^{\cdot-}$ possessed higher oxidation potentials, which posed greater health risks than their precursors (Bates et al., 2019; Jiang et al., 2019). Comparable trends have been reported for eugenol undergoing aqueous-phase reactions with $\cdot OH$ and $^3C^*$ (Li et al., 2022). The enhanced oxidation potential of reaction products is a widely-documented phenomenon associated with atmospheric aging processes (Lei et al., 2023; Bates et al., 2019). As illustrated in Figs. 2b and 4b, higher O/C ratios and OS_C values did not correspond to higher R_{DTT} values, underscoring that precursor identity governs the oxidation potential of aqSOA (Tuet et al., 2017). This observation holds for both laboratory-generated SOA and ambient atmospheric OA (Tuet et al., 2017; Xu et al., 2015a; Xu et al., 2015b; Verma et al., 2015). In general, R_{DTT} reaches a peak when highly reactive oxidation products accumulate via functionalization during the initial aging stage, and subsequently declines owing to the



396 evaporation of smaller, higher-volatility redox-active molecules generated from
397 fragmentation (Wang et al., 2018).

398 **4 Conclusions and atmospheric implications**

399 This work provides updated kinetic data for aqueous-phase reactions of three
400 representative substituted methoxyphenols with $\text{SO}_4^{\bullet-}$, and further elucidates aqSOA
401 mass yields, optical properties, and oxidation potentials. The kinetic data have
402 important implications for the fate of methoxyphenols in atmospheric aqueous
403 environments. The atmospheric lifetimes of VAA, VAL, and CFA were in the ranges of
404 0.32-50.54 h, 0.30-47.22 h, and 0.23-35.94 h, respectively, calculated according to the
405 $k_{2,M}$ values obtained in this work and a typically average $\text{SO}_4^{\bullet-}$ concentration ranging
406 from 3.6×10^{-13} to 2.3×10^{-15} mol L⁻¹ in atmospheric cloud droplets and deliquescent
407 particles (Herrmann et al., 2010). Their lifetimes may be further shortened when other
408 aqueous-phase degradation pathways (e.g., reactions with $\cdot\text{OH}$ and $^3\text{C}^*$) are taken into
409 account. Given that the obtained second-order rate constants were comparable to those
410 reported for other methoxyphenols reacting with $\cdot\text{OH}$ and $^3\text{C}^*$, $\text{SO}_4^{\bullet-}$ -initiated
411 degradation reasonably represents a vital transformation pathway of substituted
412 methoxyphenols in atmospheric aqueous environments.

413 Significant aqSOA mass yields and its light-absorbing property observed in this
414 work suggest that $\text{SO}_4^{\bullet-}$ -initiated aqueous-phase reactions of substituted
415 methoxyphenols are a potential pathway for formation of aqSOA and HULIS in
416 atmospheric aqueous environments. Recent field observations have suggested that



417 aqSOA formed from biomass burning emissions contributes substantially to haze
418 pollution in China (Xiao et al., 2022; Xue et al., 2025). Notably, methoxyphenols
419 account for up to 80% of the total mass of aqSOA generated from phenolic precursors
420 (Xiao et al., 2022). Given that the concentration of phenols and substituted phenols can
421 exceed 100 μM in fog waters impacted by biomass burning (Yu et al., 2016), the mass
422 concentration of aqSOA produced via aqueous-phase reactions of phenolic species with
423 $\text{SO}_4^{\bullet-}$ can be crudely estimated to approximately 6 mg L^{-1} . This calculation adopts the
424 minimum aqSOA mass yield (63.12%) at 30 min determined in this work and the lowest
425 molecular weight among phenolic compounds ($\text{MW} = 98$). Therefore, more attention
426 should be paid to aqSOA formation in regions impacted by biomass burning and
427 wildfires. In addition, the oxidation potentials of oxidation products for all three
428 methoxyphenols were higher than those of parent compounds. This indicates that
429 products generated during $\text{SO}_4^{\bullet-}$ -initiated oxidation of methoxyphenols posed a
430 relatively high potential risk of oxidative stress to human health (Bates et al., 2019; Lei
431 et al., 2023). Considering that the DTT assay alone cannot fully evaluate the toxicity of
432 complex atmospheric mixtures, future studies should incorporate multiple biological
433 indicators to achieve a more comprehensive toxicological assessment. Overall,
434 aqueous-phase reactions of substituted methoxyphenols with $\text{SO}_4^{\bullet-}$ substantially
435 modulate the physicochemical properties of atmospheric particles, especially in regions
436 impacted by biomass burning. In addition, incorporating the kinetic parameters and
437 aqSOA mass yields into current atmospheric models would help improve the predictive



438 performance of model simulations.

439 **Data availability.** The experimental data are available upon request to the
440 corresponding authors.

441 **Supplement.** The supplement related to this article is available.

442 **Author contributions.** R.C., T.C., and C.L. conceived the study, R.C. and X.L.
443 performed the experiments and conducted the data analyses. R.C., X.L., X.H., and R.L.
444 interpreted and discussed the data results. R.C., T.C., and C.L. wrote and revised the
445 manuscript. All coauthors contributed to discussions and suggestions in finalizing the
446 manuscript.

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Table 1. Second-order rate constants of VAA, VAL, and CFA in the aqueous-phase reactions with $\text{SO}_4^{\bullet-}$.

methoxyphenols	structures	$k_{2,M}/k_{2,R}$	$k_{2,M} (10^9 \text{ M}^{-1} \text{ s}^{-1})$	$\tau \text{ (h)}^a$
vanillic acid (VAA, $\text{C}_8\text{H}_8\text{O}_4$, MW 168)		0.75 ± 0.06	2.40 ± 0.19	0.32-50.54
vanillin (VAL, $\text{C}_8\text{H}_8\text{O}_3$, MW 152)		0.80 ± 0.03	2.56 ± 0.10	0.30-47.22
coniferyl aldehyde (CFA, $\text{C}_{10}\text{H}_{10}\text{O}_3$, MW 178)		1.06 ± 0.08	3.37 ± 0.22	0.23-35.94

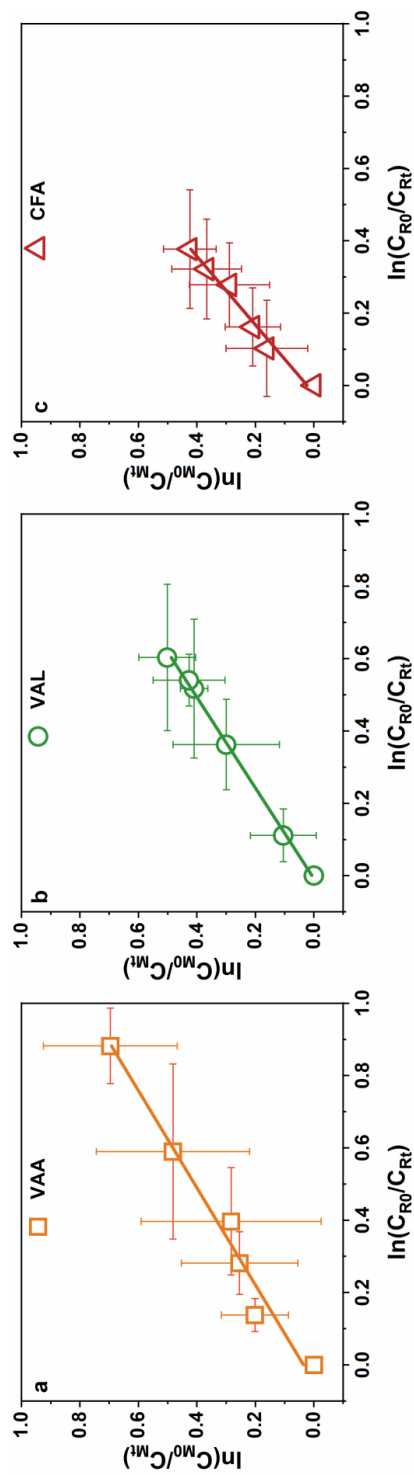
^a Atmospheric lifetime (τ) is calculated using the equation $\tau = 1/k_{2,M}[\text{SO}_4^{\bullet-}]$, where $[\text{SO}_4^{\bullet-}]$ is the average concentration of $\text{SO}_4^{\bullet-}$ in atmospheric aqueous environment, ranging from 3.6×10^{-13} to $2.3 \times 10^{-15} \text{ mol L}^{-1}$ (Herrmann et al., 2010).



Table 2. Summary of second-order rate constants of methoxyphenols in the aqueous-phase reactions with $\text{SO}_4^{\bullet-}$, $\cdot\text{OH}$, and $^3\text{C}^*$.

oxidants	methoxyphenols	$k_{2,M}$ ($\text{M}^{-1} \text{s}^{-1}$)	pH	τ (h) ^a	references
$\text{SO}_4^{\bullet-}$	VAA	$(2.40 \pm 0.19) \times 10^9$	2	4.82	this work
	VAL	$(2.56 \pm 0.10) \times 10^9$	2	4.52	this work
	CFA	$(3.37 \pm 0.22) \times 10^9$	2	3.43	this work
	ferulic acid	3.2×10^9	—	3.62	(Zhu et al., 2006)
	sinapic acid	3.2×10^9	—	3.62	(Zhu et al., 2006)
	syringic acid	$(1.7 \pm 0.1) \times 10^9$	3	6.81	(Tong et al., 2020)
$\cdot\text{OH}$	VAA	$(9.8 \pm 1.5) \times 10^9$	2	1.29	(Tang et al., 2020)
	guaiacol	$(6.8 \pm 3.1) \times 10^9$	2	1.86	(Smith et al., 2015)
	syringol	$(1.7 \pm 0.4) \times 10^{10}$	2	0.74	(Smith et al., 2015)
	syringaldehyde	3.70×10^{10}	2	0.34	(An et al., 2020)
	vanillyl alcohol	$(8.2 \pm 3.8) \times 10^9$	2	1.54	(Arciva et al., 2022)
	guaiacyl acetone	$(8.8 \pm 4.2) \times 10^9$	2	1.43	(Arciva et al., 2022)
	ferulic acid	$(13 \pm 6) \times 10^9$	2	0.97	(Arciva et al., 2022)
	syringic acid	$(9.4 \pm 4.4) \times 10^9$	2	1.34	(Arciva et al., 2022)
	syringyl acetone	$(14 \pm 7) \times 10^9$	2	0.90	(Arciva et al., 2022)
	5-methoxysalicylic acid	1.2×10^{10}	6-8	1.05	(O'Neill and Steenken, 1977)
	3-methoxyphenol	2.2×10^{10}	6-7	0.57	(O'Neill and Steenken, 1977)
	4-methoxyphenol	1.8×10^{10}	6-7	0.70	(O'Neill and Steenken, 1977)
	3,5-dimethoxyphenol	1.5×10^{10}	6-7	0.84	(O'Neill and Steenken, 1977)
	2,3-dimethoxyphenol	1.4×10^{10}	6-7	0.90	(O'Neill and Steenken, 1977)
	creosol	$(1.7 \pm 0.4) \times 10^{10}$	5.6-6.0	0.74	(He et al., 2019)
	3-methoxycatechol	$(1.9 \pm 0.4) \times 10^{10}$	5.6-6.0	0.66	(He et al., 2019)
$^3\text{C}^*$	guaiacol	$(3.2 \pm 0.7) \times 10^9$	2	1.74	(Smith et al., 2015)
	syringol	$(6.7 \pm 1.5) \times 10^9$	2	0.83	(Smith et al., 2015)
	vanillyl alcohol	$(3.0 \pm 0.75) \times 10^9$	2	1.85	(Ma et al., 2021)
	guaiacyl acetone	$(3.3 \pm 0.83) \times 10^9$	2	1.68	(Ma et al., 2021)
	ferulic acid	$(3.4 \pm 0.85) \times 10^9$	2	1.63	(Ma et al., 2021)
	syringic acid	$(3.2 \pm 0.80) \times 10^9$	2	1.74	(Ma et al., 2021)
	syringyl acetone	$(4.5 \pm 1.1) \times 10^9$	2	1.23	(Ma et al., 2021)

^a Atmospheric lifetime (τ) is calculated using average $\text{SO}_4^{\bullet-}$, $\cdot\text{OH}$, and $^3\text{C}^*$ concentrations in atmospheric aqueous environment of 2.4×10^{-14} M (Herrmann et al., 2010), 2.2×10^{-14} M (Herrmann et al., 2010), and 5.0×10^{-14} M (Kaur and Anastasio, 2018), respectively.



746

747 **Figure 1.** Plots of $\ln(C_{M0}/C_{Mt})$ versus $\ln(C_{R0}/C_{Rt})$ for aqueous-phase reactions of $\text{SO}_4^{\cdot-}$ with VAA (a), VAL (b), CFA (c), and reference compound

748 (ferulic acid).

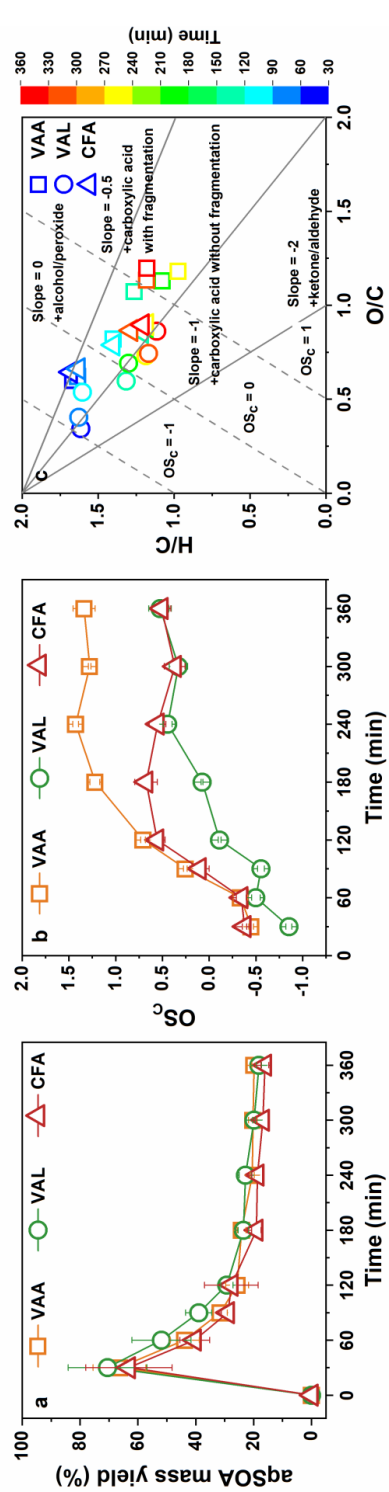


Figure 2. Time-series variations in aqSOA mass yields (a) and OS_C values (b) during aqueous-phase oxidation of VAA, VAL, and CFA by $\text{SO}_4^{\cdot-}$, and Van Krevelen diagram of the formed aqSOA (c).

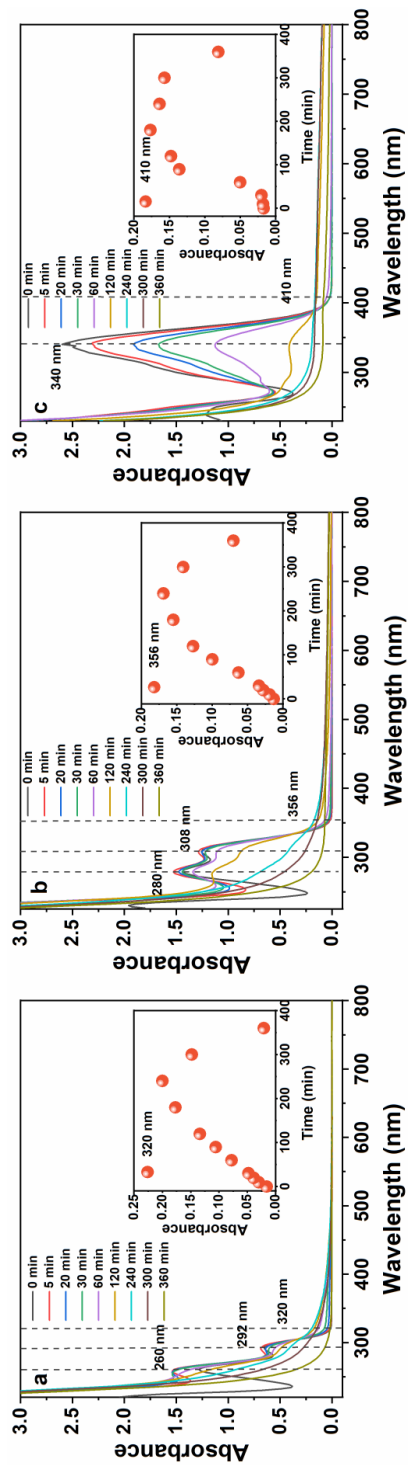
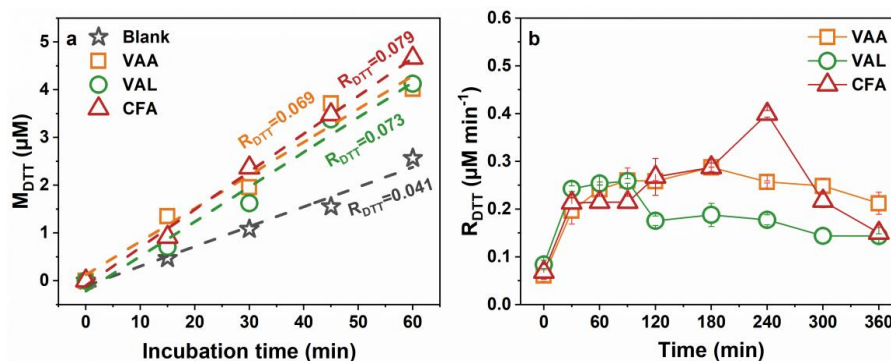


Figure 3. UV-vis absorption spectra of VAA (a), VAL (b), and CFA (c) in the aqueous-phase reactions with $\text{SO}_4^{\bullet-}$.



754

755 **Figure 4.** Correlation between DTT consumed mass (M_{DTT}) and incubation time (a),

756 and correlation between DTT consumption rate (R_{DTT}) and reaction time (b).