



1 **High rates of N and C mineralization in upper permafrost of a**
2 **thawing palsamire and their implications for GHG dynamics**

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20



21 **Abstract**

22 Rapid warming accelerates permafrost thaw in Arctic peatlands, exposing their large stocks of soil
23 organic matter (SOM) to microbial decomposition. Despite the importance of permafrost peatlands for
24 ecosystem-climate feedback, their coupled C and N mineralization potential and associated greenhouse
25 gas (GHG) dynamics still remain poorly understood. Particularly interesting in this context is the upper
26 permafrost, which is expected to thaw in near future and has been linked to substantial CO₂ and N₂O
27 production in previous studies.

28 Here, we investigated depth patterns of SOM mineralization and GHG production in a subarctic pals
29 mire in a factorial incubation experiment, where peat from active layer and upper permafrost was
30 exposed to three temperatures (4, 10, and 16 °C) and two oxygen conditions (aerobic, anaerobic).

31 The highest rate of CO₂ production with the lowest temperature sensitivity were found in the surface
32 peat and upper permafrost. Surface peat showed CH₄ consumption under aerobic condition, while CH₄
33 production was negligible across all peat depths and oxygen conditions. The highest rate and lowest
34 temperature sensitivity of net N mineralization were found in upper permafrost and active layer-
35 permafrost interface, but this did not translate into high N₂O production; instead, the maximum N₂O
36 production occurred under anaerobic condition in the lower active layer.

37 Overall, our study revealed distinct depth patterns of C and N dynamics in pals profiles, with high
38 mineralization potential in the upper permafrost, supported by labile SOM. Relatively high
39 mineralization rates at low temperatures resulted in low temperature sensitivity in the upper permafrost,
40 and indicate high potential for C and N losses under realistic temperature range expected with gradual
41 active layer deepening in the deep peat layers.

42



43 **1 Introduction**

44 Permafrost peatlands contain one of the largest terrestrial reservoirs for soil organic carbon (415 ± 147
45 Pg C) and nitrogen (10 ± 7 Pg N) and represent therefore a key component of the global permafrost-
46 climate feedback (Hugelius et al., 2020; Schuur et al., 2015). With an area of $1.7 \pm 0.5 \times 10^6$ km², they
47 comprise about half of the total northern peatland area (Hugelius et al., 2020). The large C and N stocks
48 in permafrost peatlands have accumulated under cold and wet conditions, where decomposition is slow
49 (Davidson and Janssens, 2006). With ongoing warming and permafrost thaw (Beilman and Robinson,
50 2003; Borge et al., 2017; Sannel et al., 2016; Swindles et al., 2015), decomposition processes in
51 permafrost peatlands become less temperature limited, exposing their C and N stocks to microbial
52 activities. While almost half of the permafrost peatland C stocks and most of the N stocks are currently
53 in a frozen state within permafrost (Hugelius et al., 2020), gradual active layer deepening is widely
54 documented (Streletskiy et al., 2026). Furthermore, the high ground ice content (Karjalainen et al.,
55 2022) makes permafrost peatlands vulnerable for abrupt thaw, which causes rapid mobilization of C
56 and N. Depending on thaw trajectory and environmental conditions prevailing after thaw, permafrost
57 degradation may result in increased emissions of the greenhouse gases (GHGs) carbon dioxide (CO₂),
58 methane (CH₄), and/or nitrous oxide (N₂O), thereby potentially creating a positive feedback to climate
59 warming (Helbig et al., 2017; Schuur et al., 2008; Voigt et al., 2017a, b). However, predicting the
60 magnitude and even direction of post-thaw GHG fluxes remains challenging, because warming and
61 permafrost thaw alter various environmental factors simultaneously.

62 Temperature and moisture are primary controls of microbial decomposition and other microbial
63 activities, together regulating GHG production and consumption rates in soil. While the response of C
64 mineralization in permafrost peatlands to temperature and moisture has been extensively studied by
65 incubation experiments (Baysinger et al., 2025; Diáková et al., 2016; Kjær et al., 2024; Schädel et al.,
66 2016 and references therein), fewer studies have investigated N mineralization (Diáková et al., 2016;
67 Keuper et al., 2012; Song et al., 2018) and N₂O production (Song et al., 2022; Spiller et al., 2025).
68 Incubation studies combining both C and N mineralization together with production of all three GHGs



69 are remarkably scarce (Song et al., 2022). Within the temperature range naturally occurring in northern
70 soils, temperature generally enhances microbial activity including C and N mineralization (Liu et al.,
71 2017; Schädel et al., 2014). However, temperature does not necessarily increase production of N₂O or
72 CH₄, since these gases represent the net balance of production and consumption processes, each of
73 which is differentially regulated by temperature (Dunfield et al., 1993; Holtan-Hartwig et al., 2002).
74 Furthermore, positive temperature effect is strongly modified by oxygen availability and the quantity
75 and quality of C and N substrates.

76 Moisture content largely determines oxygen diffusion to the peat profile, thereby affecting the rate of
77 C mineralization and the fraction of CO₂ and CH₄ in total C gas release Click or tap here to enter text..
78 Oxygen availability is also a key factor regulating mineral N transformation pathways which contribute
79 to N₂O production and consumption Click or tap here to enter text.. Well-drained conditions typically
80 prevail in palsas and peat plateaus—ombrotrophic peat mounds in permafrost peatlands, uplifted above
81 the surrounding mire surface by a permafrost core Click or tap here to enter text.. High oxygen
82 availability in palsa peat enhances CO₂ production and N mineralization compared to permafrost-free
83 wetlands with water-logged conditions Click or tap here to enter text.. Drier, oxygen-rich conditions
84 also support other aerobic pathways such as CH₄ oxidation Click or tap here to enter text. and
85 nitrification, as long as N substrate availability is not limiting Click or tap here to enter text.. Because
86 of prevailing aerobic conditions, palsa surfaces typically do not show substantial CH₄ emissions *in-situ*
87 but can act as sinks of atmospheric CH₄ Click or tap here to enter text.. A typical redox gradient from
88 comparatively dry surface peat to water-saturated peat deeper in the profile supports N₂O production
89 via coupled nitrification-denitrification Click or tap here to enter text.. While gradual deepening of the
90 active layer (i.e., seasonally thawing layer above permafrost) can initially promote aerobic pathways
91 and overall decomposition rates by increasing the oxic soil volume, continued permafrost degradation
92 or abrupt thaw leads to palsa collapse and initiation of water-logged conditions. Under wet anoxic
93 conditions, decomposition is slower, and anaerobic microbial pathways, such as methanogenesis Click
94 or tap here to enter text. and N₂O reduction via denitrification can occur Click or tap here to enter text..



95 The availability of labile C and N compounds is another control on GHG production and consumption
 96 in permafrost peatlands (Pengerud et al., 2013). Susceptibility of SOM for mineralization, commonly
 97 referred to as SOM quality, differs greatly between and within peatlands but also according to the depth
 98 along the peat profile. Earlier studies have reported high biodegradability and content of dissolved
 99 organic carbon (DOC) (Fouché et al., 2020) and high mineral N content (Keuper et al., 2012; Wickland
 100 et al., 2018) in permafrost peat compared to the active layer. High contents of labile C and N substrates
 101 in permafrost layers are relevant for the permafrost-climate feedback because rapid C and N
 102 mineralization after permafrost thaw can promote emissions of CO₂, CH₄ and N₂O (Knoblauch et al.,
 103 2018; Voigt et al., 2017a, 2019). Variability in SOM quality results mainly from two factors: I) the
 104 initial peat characteristics, determined by nutrient status and vegetation type during peat accumulation,
 105 II) the decomposition history, largely affected by permafrost aggradation-degradation history and the
 106 conditions before and after the frozen state (Jones et al., 2017; Treat et al., 2016). Further, downward
 107 leaching of water-soluble C and N compounds during the thaw season can contribute to the enrichment
 108 of labile C and N in the upper permafrost (Hansen and Elberling, 2023; Voigt et al., 2017a). High SOM
 109 quality affects not only the rate, but also the temperature sensitivity of mineralization: recalcitrant
 110 compounds are considered more sensitive to temperature increase than labile compounds (Conant et al.,
 111 2011; Fierer et al., 2005).

112 In this study, our aim was to assess the vulnerability of a permafrost peatland for C and N mineralization
 113 and associated GHG production, and investigate the dependence of these processes on temperature,
 114 oxygen availability and peat depth. For this purpose, we used sample material from an earlier permafrost
 115 thawing experiment, where we exposed intact peat mesocosms from a subarctic palsamire to gradual
 116 thaw. When thawing reached upper permafrost in this previous mesocosm experiment, we observed
 117 enhanced N₂O fluxes (Voigt et al., 2017a) and a larger contribution from deeper, old peat layers to
 118 surface CO₂ fluxes (Voigt et al., 2019), suggesting high lability of organic matter in this layer. We
 119 hypothesized that: I) C and N mineralization rates are higher in permafrost than in the active layer, but
 120 less sensitive to temperature increase, owing to the high content of labile organic matter in permafrost;



121 II) high mineralization rates in permafrost support high production of all three GHGs, CO₂, CH₄ and
 122 N₂O; III) C and N mineralization rates are significantly higher under aerobic than anaerobic conditions.

123 2 Materials and methods

124 2.1 Study site and soil sampling

125 In October 2012 peat cores were sampled at Peera Palsa Mire located near Kilpisjärvi, Finland (68° 89'
 126 N, 21° 05' E) (Fig. 1a). The mean annual temperature in the region is −1.9 °C and the mean annual
 127 precipitation is 487 mm (1981-2010) (Pirinen et al., 2012). The sampling site is situated in the subarctic
 128 sporadic permafrost zone, also described as the palsa zone (Seppälä, 1988, 2006). The Peera Palsa is
 129 raised by frost heave up to ~3 m above the surrounding mire area. At the time of sampling, it showed
 130 signs of an early collapsing state (Kohout et al., 2014). Palsas in this area mainly developed between
 131 3000 BP and 1000 BP (Oksanen, 2006; Seppälä, 1988, 2005). The palsa surface on Peera has typical
 132 ombrotrophic peatland vegetation including dwarf shrubs and herbaceous plants (e.g., *Betula nana* L.,
 133 *Rubus chamaemorus*, *Empetrum nigrum*) and lichens (e.g., *Cladoniaceae*). Due to frost heave and wind
 134 abrasion, parts of the palsa surface are bare of vegetation (Seppälä, 2003).

135 The peat cores for this study were sampled from naturally bare peat surfaces on the top of the palsa
 136 (Fig. 1c, d), because high N₂O emissions from such surfaces are well-documented by previous studies
 137 (Marushchak et al., 2011; Repo et al., 2009; Voigt et al., 2017b). A detailed description of the sampling
 138 site and procedure is given in the supplementary materials and methods section and in our previous
 139 publication by Voigt et al. (2017a). In brief: Four intact peat cores including the whole active layer plus
 140 15 cm of the upper permafrost (total length of 71–92 cm) and a diameter of 10 cm were sampled using
 141 a custom-made coring device consisting of a stainless-steel corer with a replaceable polypropylene
 142 lining (Fig. S1a-c). Immediately after sampling, the peat cores were transferred within the plastic lining
 143 to a mild freezing temperature of −5 °C (±1 °C; SD) and stored for about five months before the



144 mesocosm thawing experiment. During the mesocosm experiment of 33-weeks, the peat cores were
145 sequentially thawed from top to bottom, the native moisture content was contained by regular water
146 additions, and fluxes of CO₂, CH₄, and N₂O were continuously monitored (Voigt et al., 2017a, 2019).
147 After the mesocosm experiment, the cores were refrozen to mild freezing temperatures and stored for
148 six months before the start of the follow-up incubation experiment presented here.

149 **2.2 Layer separation and preparation of peat for incubation experiment**

150 Each intact peat core was divided into five sections. From the 15 cm of permafrost at the bottom of each
151 core, we separated the lower 10 cm to the *permafrost table section (PFT)* and upper 5 cm with the
152 bottom 5 cm of the active layer to the *permafrost–active layer interface section (PFI)* (Fig. 1b). The
153 remaining active layer was divided into three equally high layers labelled from surface down as *active*
154 *layer sections 1, 2 and 3 (AL1, AL2, and AL3)* (Fig. 1b). The active layer sections had different lengths
155 (17–24 cm), depending on the total peat core length (Table S1). After cutting the cores sections in a
156 frozen state (see supplementary information) (Fig. S2), they were thawed to +4°C (±2 °C) and
157 homogenized by hand-mixing. Living roots, stones and big pieces of undecomposed wood were
158 removed, if present.

159 **2.3 Set-up of incubation experiment**

160 The incubation experiment was carried out for 161 days (23 weeks) under aerobic and anaerobic
161 conditions at three temperatures (4 °C, 10 °C, and 16 °C) (Fig. S3, S4; Table S3), following the
162 incubation procedure described by Diáková et al. (2016). Prior to the start of the incubation experiment,
163 we adjusted the water content of peat to 60% water holding capacity (WHC) for aerobic incubations
164 and to 100% WHC for anaerobic incubations. After WHC adjustments, 30 g fresh weight (FW)
165 subsamples of peat were weighed into air-tight 550 mL incubation flasks (Jouco Oy, Finland) which
166 were closed with a rubber septum (Dätwyler, Switzerland) and tightened with aluminium crimp seals.



167 Aerobic incubation flasks contained ambient air as headspace gas and anaerobic flasks contained an
 168 oxygen-free, slightly over-pressured atmosphere consisting of 99 % dinitrogen (N_2) gas with 1 %
 169 carbon dioxide (CO_2). For each temperature and oxygen treatment set, we included one blank control
 170 (incubation flask without peat). More details are given in the supplementary materials and methods.

171 2.4 Gas sampling and analyses

172 The incubation flasks were kept closed during the incubation experiment for both aerobic and anaerobic
 173 treatments. The accumulation of CO_2 , CH_4 , N_2O was measured weekly for the first five weeks and bi-
 174 weekly for the residual 18 weeks. The gas concentrations were measured at the start and end of each
 175 sampling interval (in total 13 intervals) (Fig. S4), and the difference between these concentrations was
 176 used to calculate the rate of gas production or consumption. In the start of each sampling interval, the
 177 incubation flasks were flushed with the headspace gas. The start gas samples (30 ml) were taken within
 178 15 min from sealing the incubation flasks, and similar sampling took place in the end of each interval.
 179 25 ml of the gas samples were transferred into 12 ml gas exetainers (Labco Exetainer®, UK) for storage
 180 until concentration analysis. The sampled gas volume was compensated with air for aerobic flasks and
 181 with N_2 gas (5.0, AGA Industrial Gases, Finland) for anaerobic flasks. To avoid introducing O_2 into
 182 anaerobic flask, additional 30 ml of N_2 gas (5.0) was added to the headspace to establish a slight
 183 overpressure. Concentrations of CO_2 , CH_4 , N_2O in the gas samples were analysed with a gas
 184 chromatograph (7890B GC System, Agilent Technologies, USA) coupled with an auto sampler (GX-
 185 271 Liquid Handler, Gilson, USA) as described in (Voigt et al., 2017a, 2019). Gas production rates
 186 were calculated per kg dry weight (DW) based on the ideal gas law and the concentration change
 187 between the start and end of each sampling interval. Cumulative GHG fluxes were calculated by
 188 summing up the fluxes for the whole incubation period of 161 days. Minor negative CO_2 and CH_4 fluxes
 189 under anaerobic conditions, which were occasionally observed during the incubation period, were
 190 considered negligible and set to zero for the calculation of the cumulative gas production. Although



191 possibility of chemoautotrophic CO₂ consumption or anaerobic CH₄ oxidation cannot be ruled out
 192 (Sabrekov et al., 2024), our incubation system was not designed for studying these processes.

193 **2.5 Physical and chemical peat characteristics**

194 Dry weight (DW), fresh bulk density (BD), water holding capacity (WHC), pH, and extractable mineral
 195 N content (ammonium (NH₄⁺) and nitrate (NO₃⁻)) were determined on fresh peat before the incubation
 196 experiment. Further physical and chemical analyses, including soil organic matter (SOM) content, total
 197 organic carbon (C) and nitrogen (N) content, and water-filled pore space (WFPS) were carried out on
 198 oven-dried peat (+60 °C, 48 h) using standard methods. More details of the peat analyses can be found
 199 in the supplementary materials and methods.

200 **2.6 DOC, DON, microbial biomass carbon and aromaticity of DOC**

201 Before the incubation, we determined the dissolved organic carbon (DOC) and total dissolved nitrogen
 202 (TDN) contents from 0.5 M K₂SO₄ extracts (1:4 w/v) with a TOC/TN analyser (LiquicTOC II,
 203 Elementar, Germany). Dissolved organic nitrogen (DON) was calculated by subtracting inorganic
 204 nitrogen as ammonium (NH₄⁺) and nitrate (NO₃⁻) from TDN.

205 Microbial biomass carbon (C_{mic}) was determined using the chloroform fumigation extraction method
 206 (Vance et al., 1987). The content of C_{mic} was calculated as the difference in extractable C (0.5 M K₂SO₄,
 207 1:4 w/v) between the fumigated and non-fumigated sub-samples. The yield of the fumigation extraction
 208 method was corrected for incomplete extraction of microbial carbon using factor K_{EC} = 0.38 (Vance et
 209 al., 1987).

210 The aromaticity of DOC as a proxy for DOC recalcitrance was determined from filtrated (0.45 µm) peat
 211 water extracts (ratio 1:2 v/v) based on the specific ultraviolet absorbance at a wavelength 254 nm
 212 (SUVA₂₅₄), which was normalized with the DOC concentration determined with a TOC/TN analyser



(Multi N/C 2100S, Analytic Jena AG, Germany). The aromaticity as % from DOC was calculated following (Weishaar et al., 2003).

2.7 Net nitrogen mineralization and nitrification

Net nitrogen mineralization ($N_{\min}$) during aerobic and anaerobic incubation was calculated over the whole incubation period (161 days) as the difference of the initial and end content of extractable inorganic nitrogen ($NH_4^+ + NO_3^-$), divided by the time. For aerobic incubations, we further calculated the net nitrification (N_{nit}).

2.8 Temperature sensitivity of CO₂ production and net nitrogen mineralization

To assess the temperature sensitivity of aerobic and anaerobic respiration and net nitrogen mineralization, we calculated the temperature dependency coefficient Q_{10} for each process and peat layer. The Q_{10} is a commonly used index, representing the change in the rate ratio of a biological reaction with a temperature increase of 10 °C (Hamdi et al., 2013; Kirschbaum, 1995; Liu et al., 2017; Moni et al., 2015; Mundim et al., 2020). We did not calculate the Q_{10} for CH₄ production due to mostly negligible production rates. The temperature sensitivity of CO₂ production (Q_{10CO_2}) was calculated by fitting to an exponential function to the cumulative CO₂ production rates over the full incubation period as a function of temperature using Equation 1, where T is the incubation temperature and R and α are fitted parameters (Fierer et al., 2005; Hamdi et al., 2013). The Q_{10CO_2} was then calculated using the parameter α in equation 2.

231

$$CO_2 \text{ production} = R e^{\alpha T} \quad (1)$$

$$Q_{10CO_2} = e^{10\alpha} \quad (2)$$

234



235 The temperature sensitivity of net nitrogen mineralization ($Q_{10}N_{\min}$) was derived from the net nitrogen
 236 mineralization rates according to equation 3 (Liu et al., 2017). In the equation, N_1 and N_2 represent net
 237 nitrogen mineralization rates at given incubation temperatures, here 10 °C and 16 °C.

238

$$239 \quad Q_{10}N_{\min} = (N_2/N_1) e^{[10/(T_2-T_1)]} \quad (3)$$

240 **2.9 Data processing and statistical analyses**

241 Data processing, statistical analyses and graphical representation were performed using the R software
 242 (R version 4.5.1) (R Core Team, 2025). All graphics were prepared using *ggplot2* (Wickham, 2016).
 243 Data normality was visually inspected by histograms, density and Q-Q plots and tested using Shapiro-
 244 Wilk test. For non-normally distributed data, we performed $\log_{10}$ transformation, and used non-
 245 parametric tests if normality was not achieved.

246 Pairwise differences between peat layers were tested using one-way ANOVA with Tukey's HSD post
 247 hoc test (parametric test) or Kruskal-Wallis test with Dunn's test (non-parametric test) (*FSA*; (Ogle et
 248 al., 2025). This included peat characteristics, GHG production, net N mineralization rates within each
 249 temperature, as well as $Q_{10}CO_2$ and $Q_{10}N_{\min}$. The peat layers were grouped based on significant
 250 differences ($P \leq 0.05$) using *agricolae* (de Mendiburu, 2020) or *rcompanion* (Mangiafico, 2020).
 251 Similar testing was used for differences between incubation temperatures within each peat layer.

252 Further, we tested differences between peat layers and incubation temperatures in cumulative GHG
 253 production, net N mineralization ($N_{\min}$), net nitrification (N_{nitr}) the $Q_{10}CO_2$ and $Q_{10}N_{\min}$ with linear
 254 mixed effect models using *lme4* (Bates et al., 2015). The modelling was done 1) for all data combined
 255 including both aerobic and anaerobic conditions (temperature, peat depth and oxygen condition as fixed
 256 factors and the core number as random factor; Table 2), and 2) separately for each oxygen conditions
 257 (temperature and peat depth as fixed factors and the core number as random factor; Table 3, 4). The *P*-



values for model coefficients were obtained via Kenward–Roger approximations for the degrees of freedom with R package lmerTest (Kuznetsova et al., 2017).

The statistical significance of changes of DOC, DN, DON, NH_4^+ and NO_3^- content during the incubation and differences between peat layers were evaluated separately for aerobic and anaerobic incubation conditions with linear mixed-effects models as described above (Tables S11–13).

The depth-specific differences in GHG production, N_{min} , Q_{10} and peat characteristics and associations of these variables with each other were analyzed with principal component analyses (PCA) and Pearson’s correlation analyses. The principal component analyses (PCA) were visualized using *factoextra* (Kassambara and Mundt, 2020) and the correlations using *ggcorrplot* (Kassambara, 2023).

3 Results

3.1 Physical and chemical peat characteristics

Many of the peat characteristics, determined before the incubation experiment, varied distinctly across the peat profile (ANOVA; Table 1). Active layer peat was significantly drier ($F(4, 15) = 8.30$, $P < 0.001$) compared to permafrost peat, while the permafrost interface had an intermediate moisture content. A similar increasing depth pattern was seen in DOC concentrations with significant differences between layers ($F(4, 15) = 3.76$, $P = 0.026$), and in DN and DON, although without significant differences. Aromaticity of DOC showed an opposite depth pattern, with the highest aromaticity in surface peat and the lowest in permafrost ($F(4, 15) = 7.69$, $P = 0.001$). The total C and N contents and C:N ratio did not show significant depth-dependent differences; instead, the variability of C:N ratio among replicate cores was high. While NH_4^+ content tended to increase with depth, and elevated NO_3^- content was found in the top peat and permafrost, these depth patterns were not statistically significant due to high variability among replicates. Overall, NH_4^+ clearly dominated over NO_3^- across the peat profile, with a 1–2 orders of magnitude higher content.



281 Additionally, we tested the changes of DOC, C_{mic} , DN, DON, NH_4^+ , NO_3^- and pH between the start and
 282 end of the incubation with linear mixed effects models (Fig. S7–S10; Tables S11–13). Under aerobic
 283 conditions, we found significant net accumulation of NH_4^+ , especially in the permafrost layer (Fig. S9a).
 284 In the same treatment, NO_3^- accumulated throughout the peat profile at 10 °C but only in the active
 285 layer at 16 °C, while it was completely consumed at 4 °C at all depths (Fig. S9b). DOC or DN did not
 286 change during aerobic conditions while C_{mic} , DON and pH significantly decreased (Fig. S7, 8 and 10).
 287 Under anaerobic conditions, we found significant increases during the incubation in DOC and NH_4^+
 288 contents while DON content decreased (Fig. S7, 8, 9; Table S12). No significant changes were found
 289 in C_{mic} , DN, NO_3^- and pH during the anaerobic incubation.

290 3.2 CO₂ production

291 Carbon dioxide production showed a positive temperature dependency under both oxygen treatments
 292 (Fig. 2; linear mixed effect model, $P < 0.001$; Table 3). We found higher rates under aerobic than
 293 anaerobic conditions (linear mixed effect model; $P < 0.001$; Fig. 2, Table 3), and this difference between
 294 oxygen treatments increased with temperature from 2.5-fold at 4 °C to 6-fold at 16 °C (Fig. 2, Table
 295 S9). Accordingly, the temperature sensitivity of CO₂ production ($Q_{10}CO_2$) was significantly higher
 296 under aerobic than anaerobic conditions (4.0 ± 1.0 vs. 1.9 ± 0.8 , respectively; linear mixed effect model,
 297 $P < 0.001$). The depth pattern of CO₂ production was roughly similar under both oxygen treatments,
 298 with the highest rates in the top active layer and permafrost, and the lowest rates at variable depths in
 299 the active layer and permafrost interface (linear mixed effect model: aerobic: $P < 0.001$; anaerobic: $P <$
 300 0.001 – 0.003 ; Table 3). The depth pattern of $Q_{10}CO_2$ was opposite to the depth pattern of CO₂ production
 301 rate under both oxygen treatments, meaning that high CO₂ production rates were generally associated
 302 with low temperature sensitivities. The $Q_{10}CO_2$ under aerobic conditions decreased almost linearly with
 303 depth from 5.0 ± 0.9 in the top active layer to 2.9 ± 0.9 in permafrost. Under anaerobic conditions, the
 304 lowest $Q_{10}CO_2$ was similarly found in permafrost (1.2 ± 0.5), but the highest in the mid profile (2.7 ± 1.0)
 305 (Fig. 4).



306 3.3 CH₄ flux

307 The CH₄ flux during the incubation varied between significant net consumption to small net production
 308 (Fig. 2c, d; Table S4). Under aerobic conditions, we found net CH₄ consumption in the top active layer
 309 at all three incubation temperatures, resulting in a considerable, although statistically only marginally
 310 significant difference between oxygen treatments (linear mixed effect model, $P = 0.053$; Table 2). We
 311 did not observe any positive temperature dependency of CH₄ consumption (Table 3). Surprisingly, we
 312 observed small but persistent net production of CH₄ in permafrost interface and permafrost in aerobic
 313 conditions, but not in anaerobic conditions, where CH₄ production was negligible to marginal
 314 throughout the 23-week-long incubation throughout the peat profile (Fig. 2c, d; Table 3, Table S5).

315 3.4 Net N mineralization

316 We found 2 to 4-fold higher net N mineralization ($N_{\min}$) under aerobic than anaerobic conditions (linear
 317 mixed effect model, $P < 0.001$; Table 2) (Fig. 3c, d, Table S7). The depth pattern of $N_{\min}$ was similar
 318 under both oxygen treatments with uniform rates across the active layer, followed by an increase
 319 towards the permafrost layer (linear mixed effect model; $P < 0.001$ for both treatments; Table 4). A
 320 significant positive temperature dependency of $N_{\min}$ was found in both oxygen treatments (linear mixed
 321 effect model; aerobic, 16 °C vs. 4 °C: $P = 0.002$; anaerobic, both 10 and 16 °C vs. 4 °C: $P < 0.001$;
 322 Table 4). The temperature sensitivity of $N_{\min}$ ($Q_{10}N_{\min}$) was similar in both oxygen treatments with
 323 overall means of 2.1 ± 1.3 under aerobic and 2.0 ± 1.4 under anaerobic conditions (linear mixed effect
 324 model, $P = 0.682$; Table 2) (Fig. 4, Table S9). The $Q_{10}N_{\min}$ value decreased with peat depth under both
 325 oxygen conditions (linear mixed effect model; aerobic: $P < 0.001$ – 0.004 ; anaerobic: $P < 0.001$ – 0.002 ;
 326 Table 4).

327 Net nitrification rates (N_{nit}) in aerobic incubations were highly variable among replicates, resulting in
 328 ranges overlapping with zero and lack of significant differences between layers in most of the cases



(Fig. S6). A linear mixed effect model revealed a significantly higher N_{nit} at 10 °C than at 4 °C ($P = 0.001$; Table 4).

331

332 **3.5 N₂O production**

During the 161-day incubation, we found higher N₂O production under anaerobic than aerobic conditions (linear mixed effect model, $P = 0.017$; Table 2) (Fig. 3a, b; Table S4, S6). Aerobic N₂O production was negligible throughout the peat profile, while anaerobic N₂O production peaked in the mid (AL2) and deep active layer (AL3), although the differences between layers were not statistically significant due to high variability among replicates (Fig. 3b, Table 4). Under anaerobic conditions, we observed a clear, negative temperature dependency of N₂O production with the highest production rates at 4 °C, intermediate at 10 °C and lowest at 16 °C (Fig. 3b; Table S6; linear mixed effect model, $P = 0.011$; Table 4). A clear temporal pattern was observed in anaerobic N₂O production, which peaked in the beginning of the incubation at all temperatures, followed by net N₂O consumption as the incubation progressed, especially at 16 °C (data not shown). The maximum rates during the early incubation were substantially (1-2 orders of magnitude) higher than the mean rates over the whole duration (Fig. S5).

344 **3.6 Abiotic controls of C and N mineralization and GHG production**

A principal component analysis (PCA) revealed distinct differences in peat characteristics, C and N mineralization rates, and greenhouse gas production between peat layers. More specifically, we found ordination of peat layers according to depth along the first principal component (PC1), with active layer and permafrost peat at the extreme ends of this axis (Fig. 5). Overall, the PCA results were largely similar in all three incubation temperatures, and the ordination was driven by the same variables, with only slight differences in variable contributions. The PCA of 4 °C data showed the clearest depth



351 separation, suggesting that the differences in the process rates were most pronounced at low
 352 temperatures (Fig. 5, Table S10).

353 The PC1 explained 36.8% of the overall variability in the data (Fig. 5). The main contributors to PC1
 354 were $N_{\min}$, NH_4^+ , DN, DON, DOC, aromaticity of DOC, CO_2 , $Q_{10}CO_2$, and WFPS (Fig. S11a). While
 355 the negative loadings of NH_4^+ , $N_{\min}$, WFPS, DOC, DN and DON indicate that these were higher in the
 356 permafrost than in the active layer, the opposite was true for $Q_{10}CO_2$, $Q_{10}N_{\min}$, and aromaticity of DOC
 357 (Fig. 4; Table 1, S9). The second dimension (PC2) was mainly related to spatial differences between
 358 biological replicates. PC2 explained an additional 15.8% of variability (Fig. 5). The variables with
 359 strongest positive contributions to PC2 were C:N, NO_3^- , aerobic $Q_{10}CO_2$ and anaerobic $Q_{10}N_{\min}$, while
 360 the total N content, C_{mic} and aerobic N_{nit} had negative contributions (Fig. S11b).

361 The results of the Pearson's correlation analysis were in a good agreement with the PCA findings (Fig.
 362 6). We found strong positive correlations between CO_2 production and DOC content ($r = 0.6, p < 0.05$),
 363 Fig. 6), and between net N mineralization ($N_{\min}$) and DOC, DN, DON, and NH_4^+ ($r = 0.5-0.9, p < 0.05$;
 364 Fig. 6). Statistically significant negative correlations were found between aerobic $Q_{10}CO_2$ and $N_{\min}$ ($r =$
 365 -0.5 to $-0.6, p < 0.05$; Fig. 6), as well as between aromaticity of DOC and $N_{\min}$ ($r = -0.7, p < 0.05$; Fig.
 366 6).

367 4 Discussion

368 4.1 Labile organic matter in permafrost supports high C and N mineralization rates

369 Our incubation study on peat cores from a subarctic palsamire revealed distinct depth-related variability
 370 in C and N mineralization and GHG production. Our first hypothesis about high bioavailability of C
 371 and N in the permafrost layer, but low temperature sensitivity, was confirmed: We found high
 372 mineralization rates in the upper permafrost, similar (in the case of C) or higher (in the case of N) than
 373 the rates observed in the surface peat (Fig. 2, 3). The soil chemical analysis confirmed the highly



374 available C and N pool in the upper permafrost. Contents of extractable C (DOC) and N (DN, DON,
 375 NH_4^+) species were elevated in the deeper part of the profile and drove the separation between the peat
 376 layers in the PCA. As a further confirmation, the extractable C and N pools correlated positively with
 377 both CO_2 production and net N mineralization (N_{min}) rates irrespective of the oxygen treatment (Fig. 6).

378 Our findings of high N mineralization rates and elevated mineral N content in permafrost as compared
 379 to the active layer corroborate previous findings from various permafrost sediments (Hansen and
 380 Elberling, 2023; Strauss et al., 2022 and references therein). The accumulation of reactive N may result
 381 from downward leaching of soluble N to the permafrost table (Voigt et al., 2017b) and/or low C:N ratios
 382 at depth (Palmtag et al., 2022), supporting high N mineralization rates (Booth et al., 2005; Klemetsson
 383 et al., 2005). In our case, the C:N ratio or the total N content did not show clear depth patterns, but the
 384 mean DON content in the beginning of the incubation was almost twice as high in the upper permafrost
 385 than other layers, suggesting a higher fraction of easily accessible organic N at depth (Table 1). This
 386 DON was supporting high N mineralization during the incubation, which resulted in a reduction of
 387 DON coupled with an increase of NH_4^+ . Importantly, the high bioavailability of N in the upper
 388 permafrost layer points towards a fertilizer effect on plant productivity with permafrost thaw, potentially
 389 promoting C sequestration in a thawing palsamire (Keuper et al., 2012). On the other hand, high supply
 390 of N from the upper permafrost will also support microbial decomposition, and thereby CO_2 emissions
 391 from deeper peat layers (Song et al., 2022; Wild et al., 2014). Predicting the net effect of these
 392 mechanisms, associated with opposite C fluxes, is a complex modelling task and beyond the scope of
 393 this study. However, our data can be used in such work for model validation purposes.

394 Due to negligible to marginal CH_4 production under both oxygen conditions in our study, CO_2 was
 395 almost the sole product of C mineralization. We observed the highest CO_2 production rates under both
 396 aerobic and anaerobic conditions in the surface peat but also at depth, in the active layer-permafrost
 397 interface and the upper permafrost (Fig. 2a, b; Tables S4, S5). A similar bimodal depth pattern of CO_2
 398 production was previously observed in Norwegian palsamires by Kjær et al. (2024), while other
 399 permafrost peatland studies have reported a decrease from the surface towards permafrost (Song et al.,



2014, 2026; Wang et al., 2013; Zuo et al., 2025), or no differences between active layer and permafrost (Treat et al., 2014). These differences in depth patterns between studies likely result from differences among peat types and decomposition stages. In our study, the depth pattern of the CO₂ production resembled the depth pattern of the DOC content (Table 1, Fig. 2a, b). The permafrost layer contained almost twice as much DOC than the active layer, and the permafrost DOC was more labile (as indicated by a lower SUVA₂₅₄ value). Accumulation of labile DOC at the permafrost table has been previously observed (Fouché et al., 2020) and, as discussed above for DON, can result from vertical leaching (Voigt et al., 2017b). The significant positive correlation of both aerobic and anaerobic CO₂ production with DOC confirms the importance of the water-soluble C pool for CO₂ production (Fig. 6).

The temperature sensitivities of C and N mineralization showed similar depth patterns, opposite to those of the process rates. For both processes, the Q₁₀ values declined progressively with depth, reaching the lowest values in the upper permafrost (Fig. 4). The negative correlations of Q₁₀ values with DOC content and positive correlations with DOC aromaticity suggest that the depth trends can be explained by differences in substrate quality (Fig. 6). Our findings align with the common theory of lower temperature sensitivity of C decomposition for labile than recalcitrant SOM (Conant et al., 2011; Fierer et al., 2005). According to a global meta-analysis, this theory applies also to N mineralization (Liu et al., 2017). The mean temperature sensitivities of ~4 and ~2 in our study for aerobic and anaerobic CO₂ production, respectively, compare well with the ranges reported in previous subarctic and boreal peatland studies (Diáková et al., 2016; Hamdi et al., 2013; Liu et al., 2024; Tarkhov et al., 2019). However, two synthesis studies have reported lower aerobic Q₁₀ values for peatlands (Schädel et al., 2016; Wang et al., 2019) which suggests that our peat was relatively recalcitrant. Similarly high temperature sensitivity of aerobic CO₂ production was previously observed on bare permafrost peatland surfaces (Diáková et al., 2016). The lower Q₁₀ values of both C and N mineralization in the upper permafrost than surface peat resulted from comparatively high rates already at low temperatures and a proportionally lower increase with warming. This has important implications for post-thaw mobilization of C and N. It suggests that microbes in permafrost peat are cold-adapted and, thus, active under the



426 low temperature range expected in the deep active layer after gradual thaw, enhancing the possibility
 427 for CO₂ losses and mineral N turnover.

428 **4.2 Low N₂O and CH₄ production in permafrost despite high mineralization rates**

429 Contrary to our second hypothesis, high C and N mineralization rates in the upper permafrost did not
 430 translate into high rates of CH₄ or N₂O production under either of the oxygen treatments. Despite the
 431 high mineral N content and net N mineralization rate, we did not find any N₂O production in the upper
 432 permafrost or the permafrost interface; instead, the highest N₂O production was observed in the middle
 433 to lower part of the active layer (Fig. 3b). This can be explained by initial differences in moisture content
 434 and their effects on the microbial assembly mediating N turnover. Before normalising the moisture for
 435 the incubations, the water-filled pore space (WFPS) in the permafrost layer was ~80% while it was only
 436 ~40–50% in the active layer (Table 1). The lower moisture content in the active layer favoured nitrifier
 437 activity, as indicated by our net nitrification data which peaked in the active layer, not permafrost (Fig.
 438 S6). Based on substantial N₂O production only under anaerobic condition, denitrification was the
 439 dominant N₂O production pathway in our study, similar to previous permafrost peatland studies (Gil et
 440 al., 2022; Palmer and Horn, 2012). However, the denitrifier N₂O production is dependent on NO₃⁻ and
 441 NO₂⁻ supply from nitrification (Lamprecht et al., 2026). Other studies have shown that N₂O production
 442 in permafrost can be limited by a low abundance of ammonia-oxidizers, responsible for the first step of
 443 nitrification (Marushchak et al., 2021; Monteux et al., 2020). It is possible that such microbial limitation
 444 was taking place also in our deeper peat layers; however, we are lacking direct microbial data to confirm
 445 this.

446 The high importance of coupled nitrification-denitrification process for N₂O production in unfertilized
 447 soils poses a challenge for conventional soil incubation experiments, like ours, that target fully aerobic
 448 or anaerobic conditions. In our anaerobic incubation, we observed a temporary peak of N₂O production,
 449 followed by net N₂O consumption after the initial NO₃⁻ was depleted and denitrifiers switched to using
 450 N₂O as their electron acceptor. This may partially explain the lack of N₂O production in the permafrost



451 layer: under the reductive conditions, the initial NO_3^- was so quickly consumed that we did not capture
 452 any N_2O with our incubation set-up with weekly to biweekly sampling intervals. Similarly, the negative
 453 temperature dependency of N_2O production in our experiment can be explained by quicker consumption
 454 of the initial NO_3^- at higher temperatures than 4 °C.

455 An important question remains: what was causing the post-thaw increase in N_2O emissions in the
 456 preceding mesocosm experiment (Voigt et al., 2017a), if the upper permafrost did not produce any N_2O ?
 457 Increased N_2O emissions following permafrost thaw may have resulted from a larger peat volume with
 458 a favourable intermediate moisture content for coupled nitrification-denitrification, or from vertical
 459 leaching of NO_3^- produced in the upper, oxic peat layers to the deeper, anoxic layer, where it was
 460 denitrified. These mechanisms could not be captured in our incubation experiment, where homogenized
 461 peat from individual layers was incubated separately under either fully aerobic or fully anaerobic
 462 conditions. Also, without microbial data on functional gene abundance and activity, the processes
 463 underlying the increased N_2O emissions in the mesocosms experiment remain partly inconclusive.
 464 Despite this, our results demonstrate the added value of conducting parallel investigations using
 465 mesocosm approaches and conventional soil incubations with discrete samples but also highlight the
 466 importance of considering interactions between soil layers and gas transport processes.

467 In the oxic incubation, we found CH_4 consumption in the surface peat (Fig. 2c, Table 3, S4). This is
 468 consistent with the net uptake of CH_4 we found in the preceding mesocosm experiment in the dry
 469 scenario, imitating gradual active layer deepening (in contrast to a wet scenario, imitating palsa
 470 collapse) (Voigt et al., 2019), as well as field measurements in palsa mires (Voigt et al., 2023).
 471 Similarly, the lack of CH_4 production in our anaerobic treatment (Fig. 2d) is in accordance with the
 472 negligible CH_4 emissions under the wet scenario of the mesocosm experiment (Voigt et al., 2019). The
 473 lack of CH_4 production under the 5-month anaerobic incubation is likely a consequence of a very low
 474 abundance or complete absence of active methanogens. In the uplifted palsa surfaces, the active layer
 475 is well-aerated and rich with oxygenated compounds, such as NO_3^- (Table 1), SO_4^{2-} , Fe^{3+} , and
 476 oxygenated organic compounds, which can act as alternative electron acceptors and need to be reduced



477 before methanogenesis becomes favourable (Acht nich et al., 1995). This, together with the low initial
 478 abundance of methanogens, may prevent CH₄ production even under extended anoxia. In the upper
 479 permafrost, on the other hand, the long-term frozen conditions can cause limited microbial functionality
 480 due to lack of certain microbial groups or plant derived organics. For this reason, long lag-phases or
 481 complete absence of CH₄ production are commonly reported for oxic soil layers and permafrost soils
 482 (Knoblauch et al., 2018; Treat et al., 2016). Under anaerobic conditions, aerobic methanotrophs could
 483 also use produced N₂O as source of oxygen as shown earlier (Awala et al., 2024; Hao et al., 2022), this
 484 could be a potential reason for limited N₂O and CH₄ production under anoxic conditions. However, we
 485 do not have microbial data to support this theory. Interestingly, we found marginal CH₄ production rates
 486 in aerobic conditions in the deeper peat layers (Fig. 2c), similar to another subarctic peatland study
 487 (Diáková et al., 2016). This can be explained by CH₄ production in anaerobic microsites in trace
 488 amounts which could not be detected in the anaerobic treatment with CH₄-free headspace but were
 489 detectable in the aerobic treatment with ambient CH₄ concentration. Also, we cannot completely rule
 490 out the possibility for oxic CH₄ production (Knief, 2019).

491 **4.3 Evidence for stronger warming impact under gradual than abrupt permafrost thaw**

492 In confirmation of our third hypothesis, our results suggest different responses of C and N
 493 mineralization and GHG release from palsas mires to initial active layer deepening and continued or
 494 abrupt permafrost thaw. While both permafrost thaw trajectories will unlock the highly bioavailable C
 495 and N substrates in the upper permafrost, they will be differently affected by the contrasting
 496 hydrological and redox conditions. Active layer deepening will initially increase the oxic peat volume,
 497 whereas continued or abrupt permafrost thaw will result in palasa collapse and water saturated
 498 conditions. The higher overall GHG production in our study under aerobic than anaerobic conditions,
 499 strongly dominated by CO₂ production, suggests that gradual active layer deepening will result in higher
 500 GHG release. We found substantially higher CO₂ production under aerobic than anaerobic conditions,
 501 in accordance with the previous pan-arctic incubation synthesis (Schädel et al., 2016). Furthermore, the



502 response of CO₂ production to a temperature increase was stronger under aerobic than anaerobic
 503 conditions (Fig. 4, Table 2), a finding which has also been reported from other peatlands (Liu et al.,
 504 2024; Szafranek-Nakoneczna and Stępniewska, 2014). The higher temperature sensitivity suggests
 505 stronger diurnal and seasonal dynamics under dry oxic conditions. Also, N mineralization rates was
 506 remarkably higher under aerobic than anaerobic conditions, suggesting higher reactive N supply with
 507 gradual active layer deepening compared to abrupt thaw (Fig. 3c, d). In this context it must be noted,
 508 however, that while gradual active layer deepening will expose the upper permafrost with high content
 509 of labile C for microbial decomposition, the peat will likely remain wet and poorly oxygenated in
 510 comparison to the surface peat, restricting the rate of decomposition compared to fully aerobic
 511 conditions in our aerobic treatment.

512 Under the different scenario that the palsa collapses, the decomposition of peat and associated CO₂
 513 production will slow down. On the other hand, highly productive vegetation usually establishes in
 514 thermokarst wetlands and produces labile carbon, which could cause increased CO₂ and CH₄ production
 515 (Heffernan et al., 2022; Hodgkins et al., 2014). It is important to note that our anaerobic incubation
 516 likely underestimated the CH₄ production due to the microbial limitation of methanogenesis, which will
 517 be alleviated quickly in the real environment after the palsa collapses and gets exposed to highly reduced
 518 conditions and microbial invasion from the surrounding wetland. The same is true for any soil
 519 incubations with discrete samples, limiting their value for predicting ecosystem-climate feedback
 520 (Knoblauch et al., 2018). At the same time, classical soil incubations produce depth-resolved data and
 521 allow soil characteristics and GHG production to be linked by soil layer. As such, they serve as a
 522 valuable complement for field observations and mesocosm studies.

523 **5 Conclusion**

524 Our study demonstrates that the upper permafrost of palsa peatlands contains highly bioavailable C and
 525 N, making it particularly vulnerable to microbial mineralization and greenhouse gas production



526 following permafrost thaw. In addition to elevated concentrations of labile dissolved carbon and
527 nitrogen, the significantly lower aromaticity of dissolved organic matter provides independent evidence
528 that the upper permafrost contains a less recalcitrant, more bioavailable carbon and nitrogen pools than
529 the overlying active layer. Despite the relatively low temperature sensitivity, the high substrate
530 availability in the upper permafrost resulted in the highest CO₂ production rates. This suggests that
531 substrate availability rather than temperature sensitivity will dominate the initial CO₂ response to
532 permafrost thaw, and that comparatively high CO₂ production can be expected in the lower part of peat
533 profile despite cold conditions at depth. In contrast, N₂O production was primarily controlled by oxygen
534 availability and coupled oxic-anoxic transformations rather than substrate availability, indicating that
535 future N₂O emissions cannot be predicted from peat characteristics alone. These findings highlight, that
536 reliable predictions of the permafrost–climate feedback require consideration of both, substrate quality
537 and the spatial organization of redox conditions. It is important to consider the possibility of functional
538 limitations due to absence of certain microbial groups in permafrost when designing experiments and
539 interpreting data; however, in the real environment, such limitations are likely quickly removed.

540



541 **6 Code availability**

542 The R code for the data analyses is available upon request.

543

544 **7 Data availability**

545 The data of this study, analyzed and generated, is provided in full within the article and its
546 supplementary information.

547

548 **8 Supplementary information**

549 The supplementary information text (Supplementary_Information.pdf) comprises of more detailed
550 descriptions of the peat core sampling, the incubation experiment procedure and analyses performed.
551 Furthermore, the supplementary information includes additional figures and tables supporting the text.

552 **9 Author contribution**

553 C.B., H.S., P.J.P., K.D. and R.E.L planned and designed the experiment. R.E.L., C.V. and M.E.M.
554 sampled the peat cores in the field. R.E.L and K.D. performed the incubation experiment and relating
555 chemical analyses, gas analyses and data calculations. R.E.L., with the support and supervision of C.V.,
556 H.M.P.S. and M.E.M., further processed data, performed statistical analyses and prepared the graphs.
557 R.E.L., with supervision of H.M.P.S., C.B. and M.E.M. wrote the first draft of the manuscript. All
558 authors have discussed and revised the manuscript.



559 **10 Competing interests**

560 The authors declare that they have no conflict of interest.

561

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577 **12 References**

- 578 Achtnich, C., Bak, F., and Conrad, R.: Competition for electron donors among nitrate reducers, ferric
 579 iron reducers, sulfate reducers, and methanogens in anoxic paddy soil, *Biol. Fertil. Soils*, 19, 65–72,
 580 <https://doi.org/10.1007/BF00336349/METRICS>, 1995.
- 581 Awala, S. I., Gwak, J. H., Kim, Y., Jung, M. Y., Dunfield, P. F., Wagner, M., and Rhee, S. K.: Nitrous
 582 oxide respiration in acidophilic methanotrophs, *Nat. Commun.*, 15, 4226–,
 583 <https://doi.org/10.1038/s41467-024-48161-z>, 2024.
- 584 Bates, D., Mächler, M., Bolker, B., and Walker, S.: Fitting Linear Mixed-Effects Models Using lme4,
 585 *J. Stat. Softw.*, 67, 1–48, <https://doi.org/10.18637/jss.v067.i01>, 2015.
- 586 Baysinger, M. R., Laurent, M., Verdonen, M., Reif, J., Kumpula, T., Liebner, S., and Treat, C. C.:
 587 Abrupt Thaw in a Finnish Palsa: Potential CH₄ Production Driven by Vegetation Adaptation in the
 588 Transition From Permafrost to Post-Thaw Soils, *J. Geophys. Res. Biogeosci.*, 130, e2025JG008847,
 589 <https://doi.org/10.1029/2025JG008847>, 2025.
- 590 Beilman, D. W. and Robinson, S. D.: Peatland permafrost thaw and landform type along a climatic
 591 gradient, in: *Proceedings of the 8th International Conference on Permafrost*, 61–65, 2003.
- 592 Booth, M. S., Stark, J. M., and Rastetter, E.: CONTROLS ON NITROGEN CYCLING IN
 593 TERRESTRIAL ECOSYSTEMS: A SYNTHETIC ANALYSIS OF LITERATURE DATA, *Ecol.*
 594 *Monogr.*, 75, 139–157, <https://doi.org/10.1890/04-0988>, 2005.
- 595 Borge, A. F., Westermann, S., Solheim, I., and Etzelmüller, B.: Strong degradation of palsas and peat
 596 plateaus in northern Norway during the last 60 years, *Cryosphere*, 11, 1–16, [https://doi.org/10.5194/tc-](https://doi.org/10.5194/tc-11-1-2017)
 597 11-1-2017, 2017.



- 598 Butterbach-Bahl, K., Philippot, L., Serra, J., Silver, W. L., Ogle, S., and Abalos, D.: Three Decades of
 599 Soil N₂O Research—Insights and Gaps, *Glob. Chang. Biol.*, 32, e71038,
 600 <https://doi.org/10.1111/GCB.71038>, 2026.
- 601 Conant, R. T., Ryan, M. G., Ågren, G. I., Birge, H. E., Davidson, E. A., Eliasson, P. E., Evans, S. E.,
 602 Frey, S. D., Giardina, C. P., Hopkins, F. M., Hyvönen, R., Kirschbaum, M. U. F., Lavelle, J. M.,
 603 Leifeld, J., Parton, W. J., Megan Steinweg, J., Wallenstein, M. D., Martin Wetterstedt, J. Å., and
 604 Bradford, M. A.: Temperature and soil organic matter decomposition rates - synthesis of current
 605 knowledge and a way forward, <https://doi.org/10.1111/j.1365-2486.2011.02496.x>, 1 November 2011.,
 606 2011.
- 607 Davidson, E. A. and Janssens, I. A.: Temperature sensitivity of soil carbon decomposition and
 608 feedbacks to climate change, *Nature*, 440, 165–173, <https://doi.org/10.1038/nature04514>, 2006.
- 609 Diáková, K., Čapek, P., Kohoutová, I., Mpamah, P. A., Bárta, J., Biasi, C., Martikainen, P. J., and
 610 Šantrůčková, H.: Heterogeneity of carbon loss and its temperature sensitivity in East-European
 611 subarctic tundra soils, *FEMS Microbiol. Ecol.*, 92, fiw140, <https://doi.org/10.1093/femsec/fiw140>,
 612 2016.
- 613 Dunfield, P., knowles, R., Dumont, R., and Moore, T. R.: Methane production and consumption in
 614 temperate and subarctic peat soils: Response to temperature and pH, *Soil Biol. Biochem.*, 25, 321–326,
 615 [https://doi.org/10.1016/0038-0717\(93\)90130-4](https://doi.org/10.1016/0038-0717(93)90130-4), 1993.
- 616 Fan, Z., Neff, J. C., Waldrop, M. P., Ballantyne, A. P., and Turetsky, M. R.: Transport of oxygen in soil
 617 pore-water systems: implications for modeling emissions of carbon dioxide and methane from
 618 peatlands, *Biogeochemistry*, 121, 455–470, <https://doi.org/10.1007/S10533-014-0012-0/FIGURES/5>,
 619 2014.
- 620 Fierer, N., Craine, J. M., Mclauchlan, K., and Schimel, J. P.: Litter quality and the temperature
 621 sensitivity of decomposition, *Ecology*, 86, 320–326, <https://doi.org/10.1890/04-1254>, 2005.



- 622 Fouché, J., Christiansen, C. T., Lafrenière, M. J., Grogan, P., and Lamoureux, S. F.: Canadian
 623 permafrost stores large pools of ammonium and optically distinct dissolved organic matter, *Nat.*
 624 *Commun.*, 11, 1–11, <https://doi.org/10.1038/s41467-020-18331-w>, 2020.
- 625 Gil, J., Marushchak, M. E., Rütting, T., Baggs, E. M., Pérez, T., Novakovskiy, A., Trubnikova, T.,
 626 Kaverin, D., Martikainen, P. J., and Biasi, C.: Sources of nitrous oxide and the fate of mineral nitrogen
 627 in subarctic permafrost peat soils, *Biogeosciences*, 19, 2683–2698, [https://doi.org/10.5194/bg-19-2683-](https://doi.org/10.5194/bg-19-2683-2022)
 628 2022, 2022.
- 629 Hamdi, S., Moyano, F., Sall, S., Bernoux, M., and Chevallier, T.: Synthesis analysis of the temperature
 630 sensitivity of soil respiration from laboratory studies in relation to incubation methods and soil
 631 conditions, *Soil Biol. Biochem.*, 58, 115–126, <https://doi.org/10.1016/j.soilbio.2012.11.012>, 2013.
- 632 Hansen, H. F. E. and Elberling, B.: Spatial Distribution of Bioavailable Inorganic Nitrogen From
 633 Thawing Permafrost, *Global Biogeochem. Cycles*, 37, e2022GB007589,
 634 <https://doi.org/10.1029/2022GB007589>, 2023.
- 635 Hao, Q., Wang, O., Jiao, J. Y., Xiao, L., Zhang, Y., Li, W. J., and Liu, F.: *Methylobacter* couples
 636 methane oxidation and N₂O production in hypoxic wetland soil, *Soil Biol. Biochem.*, 175, 108863,
 637 <https://doi.org/10.1016/J.SOILBIO.2022.108863>, 2022.
- 638 Heffernan, L., Cavaco, M. A., Bhatia, M. P., Estop-Aragonés, C., Knorr, K. H., and Olefeldt, D.: High
 639 peatland methane emissions following permafrost thaw: enhanced acetoclastic methanogenesis during
 640 early successional stages, *Biogeosciences*, 19, 3051–3071, <https://doi.org/10.5194/BG-19-3051-2022>,
 641 2022.
- 642 Helbig, M., Chasmer, L. E., Desai, A. R., Kljun, N., Quinton, W. L., and Sonnentag, O.: Direct and
 643 indirect climate change effects on carbon dioxide fluxes in a thawing boreal forest–wetland landscape,
 644 *Glob. Chang. Biol.*, 23, 3231–3248, <https://doi.org/10.1111/gcb.13638>, 2017.
- 645 Hodgkins, S. B., Tfaily, M. M., McCalley, C. K., Logan, T. A., Crill, P. M., Saleska, S. R., Rich, V. I.,
 646 and Chanton, J. P.: Changes in peat chemistry associated with permafrost thaw increase greenhouse gas



647 production, Proceedings of the National Academy of Sciences, 111, 5819–5824,
 648 <https://doi.org/10.1073/pnas.1314641111>, 2014.

649 Holtan-Hartwig, L., Dörsch, P., and Bakken, L. R.: Low temperature control of soil denitrifying
 650 communities: kinetics of N₂O production and reduction, Soil Biol. Biochem., 34, 1797–1806,
 651 [https://doi.org/10.1016/S0038-0717\(02\)00169-4](https://doi.org/10.1016/S0038-0717(02)00169-4), 2002.

652 Hugelius, G., Loisel, J., Chadburn, S., Jackson, R. B., Jones, M., MacDonald, G., Marushchak, M.,
 653 Olefeldt, D., Packalen, M., Siewert, M. B., Treat, C., Turetsky, M., Voigt, C., and Yu, Z.: Large stocks
 654 of peatland carbon and nitrogen are vulnerable to permafrost thaw, Proc. Natl. Acad. Sci. U. S. A., 117,
 655 20438–20446, <https://doi.org/10.1073/pnas.1916387117>, 2020.

656 Jones, M. C., Harden, J., O'Donnell, J., Manies, K., Jorgenson, T., Treat, C., and Ewing, S.: Rapid
 657 carbon loss and slow recovery following permafrost thaw in boreal peatlands, Glob. Chang. Biol., 23,
 658 1109–1127, <https://doi.org/10.1111/gcb.13403>, 2017.

659 Karjalainen, O., Aalto, J., Kanevskiy, M. Z., Luoto, M., and Hjort, J.: High-resolution predictions of
 660 ground ice content for the Northern Hemisphere permafrost region, Earth System Science Data
 661 Discussion [preprint], 1–40 pp., <https://doi.org/10.5194/essd-2022-144>, 2022.

662 Kassambara, A.: Visualization of a Correlation Matrix using “ggplot2” [R package ggcorrplot version
 663 0.1.4.1], CRAN: Contributed Packages, <https://doi.org/10.32614/CRAN.PACKAGE.GGCORRLOT>,
 664 2023.

665 Kassambara, A. and Mundt, F.: factoextra: Extract and Visualize the Results of Multivariate Data
 666 Analyses, <https://cran.r-project.org/package=factoextra>, 2020., 2020.

667 Keuper, F., van Bodegom, P. M., Dorrepaal, E., Weedon, J. T., van Hal, J., van Logtestijn, R. S. P., and
 668 Aerts, R.: A frozen feast: Thawing permafrost increases plant-available nitrogen in subarctic peatlands,
 669 Glob. Chang. Biol., 18, 1998–2007, <https://doi.org/10.1111/j.1365-2486.2012.02663.x>, 2012.



- 670 Kirschbaum, M. U. F.: The temperature dependence of soil organic matter decomposition, and the effect
 671 of global warming on soil organic C storage, *Soil Biol. Biochem.*, 27, 753–760,
 672 [https://doi.org/10.1016/0038-0717\(94\)00242-S](https://doi.org/10.1016/0038-0717(94)00242-S), 1995.
- 673 Kjær, S. T., Westermann, S., Nedkvitne, N., and Dörsch, P.: Carbon degradation and mobilisation
 674 potentials of thawing permafrost peatlands in northern Norway inferred from laboratory incubations,
 675 *Biogeosciences*, 21, 4723–4737, <https://doi.org/10.5194/bg-21-4723-2024>, 2024.
- 676 Klemetsson, L., Von Arnold, K., Weslien, P., and Gundersen, P.: Soil CN ratio as a scalar parameter
 677 to predict nitrous oxide emissions, *Glob. Chang. Biol.*, 11, 1142–1147, [https://doi.org/10.1111/J.1365-](https://doi.org/10.1111/J.1365-2486.2005.00973.X)
 678 [2486.2005.00973.X](https://doi.org/10.1111/J.1365-2486.2005.00973.X), 2005.
- 679 Knief, C.: Diversity of methane-cycling microorganisms in soils and their relation to oxygen, *Curr.*
 680 *Issues Mol. Biol.*, 33, 23–56, <https://doi.org/10.21775/CIMB.033.023>, 2019.
- 681 Knoblauch, C., Beer, C., Liebner, S., Grigoriev, M. N., and Pfeiffer, E. M.: Methane production as key
 682 to the greenhouse gas budget of thawing permafrost, *Nat. Clim. Chang.*, 1–4,
 683 <https://doi.org/10.1038/s41558-018-0095-z>, 2018.
- 684 Kohout, T., Bučko, M. S., Rasmus, K., Leppäranta, M., and Matero, I.: Non-Invasive Geophysical
 685 Investigation and Thermodynamic Analysis of a Palsa in Lapland, Northwest Finland, *Permafr.*
 686 *Periglac. Process.*, 25, 45–52, <https://doi.org/10.1002/ppp.1798>, 2014.
- 687 Kuznetsova, A., Brockhoff, P. B., and Christensen, R. H. B.: lmerTest Package: Tests in Linear Mixed
 688 Effects Models, *J. Stat. Softw.*, 82, 1–26, <https://doi.org/10.18637/jss.v082.i13>, 2017.
- 689 Lamprecht, R. E., Marushchak, M. E., Voigt, C., Bagnoud, A., Alves, R. J. E., Schleper, C.,
 690 Martikainen, P., Biasi, C., and Siljanen, H. M. P.: Response of ammonia-oxidizing archaea to
 691 experimental warming in high-N₂O-emitting permafrost peatland habitats, *Arct. Sci.*,
 692 <https://doi.org/10.1139/as-2025-0075>, 2026.



- 693 Liu, H., Rezanezhad, F., Zhao, Y., He, H., Van Cappellen, P., and Lennartz, B.: The apparent
 694 temperature sensitivity (Q10) of peat soil respiration: A synthesis study, *Geoderma*, 443, 116844,
 695 <https://doi.org/10.1016/j.geoderma.2024.116844>, 2024.
- 696 Liu, Y., Wang, C., He, N., Wen, X., Gao, Y., Li, S., Niu, S., Butterbach-Bahl, K., Luo, Y., and Yu, G.:
 697 A global synthesis of the rate and temperature sensitivity of soil nitrogen mineralization: latitudinal
 698 patterns and mechanisms, *Glob. Chang. Biol.*, 23, 455–464, <https://doi.org/10.1111/gcb.13372>, 2017.
- 699 Mangiafico, S.: rcompanion: Functions to Support Extension Education Program Evaluation,
 700 <https://cran.r-project.org/package=rcompanion>, 2020., 2020.
- 701 Marushchak, M. E., Pitkämäki, A., Koponen, H., Biasi, C., Seppälä, M., and Martikainen, P. J.: Hot
 702 spots for nitrous oxide emissions found in different types of permafrost peatlands, *Glob. Chang. Biol.*,
 703 17, 2601–2614, <https://doi.org/10.1111/j.1365-2486.2011.02442.x>, 2011.
- 704 Marushchak, M. E., Kerttula, J., Diáková, K., Faguet, A., Gil, J., Grosse, G., Knoblauch, C.,
 705 Lashchinskiy, N., Martikainen, P. J., Morgenstern, A., Nykamb, M., Ronkainen, J. G., Siljanen, H. M.
 706 P., van Delden, L., Voigt, C., Zimov, N., Zimov, S., and Biasi, C.: Thawing Yedoma permafrost is a
 707 neglected nitrous oxide source, *Nature Communications* 2021 12:1, 12, 1–10,
 708 <https://doi.org/10.1038/s41467-021-27386-2>, 2021.
- 709 de Mendiburu, F.: agricolae: Statistical Procedures for Agricultural Research, [https://cran.r-](https://cran.r-project.org/package=agricolae)
 710 [project.org/package=agricolae](https://cran.r-project.org/package=agricolae), 2020., 2020.
- 711 Moni, C., Lerch, T. Z., Knoth de Zarruk, K., Strand, L. T., Forte, C., Certini, G., and Rasse, D. P.:
 712 Temperature response of soil organic matter mineralisation in arctic soil profiles, *Soil Biol. Biochem.*,
 713 88, 236–246, <https://doi.org/10.1016/j.soilbio.2015.05.024>, 2015.
- 714 Monteux, S., Keuper, F., Fontaine, S., Gavazov, K., Hallin, S., Juhanson, J., Krab, E. J., Revaillet, S.,
 715 Verbruggen, E., Walz, J., Weedon, J. T., and Dorrepaal, E.: Carbon and nitrogen cycling in Yedoma
 716 permafrost controlled by microbial functional limitations, *Nat. Geosci.*, 13, 794–798,
 717 <https://doi.org/10.1038/s41561-020-00662-4>, 2020.



- 718 Mundim, K. C., Baraldi, S., Machado, H. G., and Vieira, F. M. C.: Temperature coefficient (Q10) and
 719 its applications in biological systems: Beyond the Arrhenius theory, *Ecol. Modell.*, 431, 109127,
 720 <https://doi.org/10.1016/j.ecolmodel.2020.109127>, 2020.
- 721 Nykänen, H., Heikkinen, J. E. P., Pirinen, L., Tiilikainen, K., and Martikainen, P.: Annual CO₂
 722 exchange and CH₄ fluxes on a subarctic palsa mire during climatically different years, *Global*
 723 *Biogeochem. Cycles*, 17, 1018, <https://doi.org/10.1029/2002GB001861>, 2003.
- 724 Ogle, D. H., Doll, J. C., Wheeler, A. P., and Dinno, A.: FSA: Simple Fisheries Stock Assessment
 725 Methods, <https://fishr-core-team.github.io/FSA/>, 2025., 2025.
- 726 Oksanen, P. O.: Holocene development of the Vaisjeäggi palsa mire, Finnish Lapland, *Boreas*, 35, 81–
 727 95, <https://doi.org/10.1111/j.1502-3885.2006.tb01114.x>, 2006.
- 728 Palmer, K. and Horn, M. A.: Actinobacterial nitrate reducers and proteobacterial denitrifiers are
 729 abundant in N₂O-metabolizing palsa peat, *Appl. Environ. Microbiol.*, 78, 5584–5596,
 730 <https://doi.org/10.1128/AEM.00810-12>, 2012.
- 731 Palmtag, J., Obu, J., Kuhry, P., Richter, A., Siewert, M. B., Weiss, N., Westermann, S., and Hugelius,
 732 G.: A high spatial resolution soil carbon and nitrogen dataset for the northern permafrost region based
 733 on circumpolar land cover upscaling, *Earth Syst. Sci. Data*, 14, 4095–4110,
 734 <https://doi.org/10.5194/essd-14-4095-2022>, 2022.
- 735 Pengerud, A., Cécillon, L., Johnsen, L. K., Rasse, D. P., and Strand, L. T.: Permafrost Distribution
 736 Drives Soil Organic Matter Stability in a Subarctic Palsa Peatland, *Ecosystems*, 16, 934–947,
 737 <https://doi.org/10.1007/s10021-013-9652-5>, 2013.
- 738 Pirinen, P., Simola, H., Aalto, J., Kaukoranta, J.-P., Karlsson, P., and Ruuhela, R.: Tilastoja suomen
 739 ilmastosta 1981-2010, Raportteja, 96 pp., <http://hdl.handle.net/10138/35880>, 2012.
- 740 R Core Team: R: A Language and Environment for Statistical Computing, <https://www.r-project.org/>,
 741 2025., 2025.



- 742 Repo, M. E., Susiluoto, S., Lind, S. E., Jokinen, S., Elsakov, V., Biasi, C., Virtanen, T., and Martikainen,
 743 P. J.: Large N₂O emissions from cryoturbated peat soil in tundra, *Nat. Geosci.*, 2, 189–192,
 744 <https://doi.org/10.1038/ngeo434>, 2009.
- 745 Sabrekov, A. F., Semenov, M. V., Terentieva, I. E., Krasnov, G. S., Kharitonov, S. L., Glagolev, M. V.,
 746 and Litt, Y. V.: Anaerobic methane oxidation is quantitatively important in deeper peat layers of boreal
 747 peatlands: Evidence from anaerobic incubations, in situ stable isotopes depth profiles, and microbial
 748 communities, *Science of the Total Environment*, 916, 170213,
 749 <https://doi.org/10.1016/j.scitotenv.2024.170213>, 2024.
- 750 Sannel, A. B. K., Hugelius, G., Jansson, P., and Kuhry, P.: Permafrost Warming in a Subarctic Peatland
 751 – Which Meteorological Controls are Most Important?, *Permafr. Periglac. Process.*, 27, 177–188,
 752 <https://doi.org/10.1002/ppp.1862>, 2016.
- 753 Schädel, C., Schuur, E. A. G., Bracho, R., Elberling, B., Knoblauch, C., Lee, H., Luo, Y., Shaver, G.
 754 R., and Turetsky, M. R.: Circumpolar assessment of permafrost C quality and its vulnerability over time
 755 using long-term incubation data, *Glob. Chang. Biol.*, 20, 641–652, <https://doi.org/10.1111/gcb.12417>,
 756 2014.
- 757 Schädel, C., Bader, M. K. F., Schuur, E. A. G., Biasi, C., Bracho, R., Čapek, P., De Baets, S., Diáková,
 758 K., Ernakovich, J., Estop-Aragones, C., Graham, D. E., Hartley, I. P., Iversen, C. M., Kane, E.,
 759 Knoblauch, C., Lupascu, M., Martikainen, P. J., Natali, S. M., Norby, R. J., O'Donnell, J. A.,
 760 Chowdhury, T. R., Šantrůčková, H., Shaver, G., Sloan, V. L., Treat, C. C., Turetsky, M. R., Waldrop,
 761 M. P., and Wickland, K. P.: Potential carbon emissions dominated by carbon dioxide from thawed
 762 permafrost soils, *Nat. Clim. Chang.*, 6, 950–953, <https://doi.org/10.1038/nclimate3054>, 2016.
- 763 Schuur, E. A. G., Bockheim, J., Canadell, J. G., Euskirchen, E., Field, C. B., Goryachkin, S. V.,
 764 Hagemann, S., Kuhry, P., Lafleur, P. M., Lee, H., Mazhitova, G., Nelson, F. E., Rinke, A., Romanovsky,
 765 V. E., Shiklomanov, N., Tarnocai, C., Venevsky, S., Vogel, J. G., and Zimov, S. A.: Vulnerability of



- 766 permafrost carbon to climate change: Implications for the global carbon cycle, *Bioscience*, 58, 701–
 767 714, <https://doi.org/10.1641/B580807>, 2008.
- 768 Schuur, E. A. G., McGuire, A. D., Schädel, C., Grosse, G., Harden, J. W., Hayes, D. J., Hugelius, G.,
 769 Koven, C. D., Kuhry, P., Lawrence, D. M., Natali, S. M., Olefeldt, D., Romanovsky, V. E., Schaefer,
 770 K., Turetsky, M. R., Treat, C. C., and Vonk, J. E.: Climate change and the permafrost carbon feedback,
 771 *Nature*, 520, 171–179, <https://doi.org/10.1038/nature14338>, 2015.
- 772 Seppälä, M.: Palsas and related forms, in: *Advances in Periglacial Geomorphology*, edited by: Clark,
 773 M. J., John Wiley & Sons, Ltd, 247–278, 1988.
- 774 Seppälä, M.: Surface abrasion of palsas by wind action in Finnish Lapland, *Geomorphology*, 52, 141–
 775 148, [https://doi.org/10.1016/S0169-555X\(02\)00254-4](https://doi.org/10.1016/S0169-555X(02)00254-4), 2003.
- 776 Seppälä, M.: Dating of palsas, *Special Paper of the Geological Survey of Finland*, 79–84, 2005.
- 777 Seppälä, M.: Palsa mires in Finland, *The Finnish Environment*, 23, 155–162, 2006.
- 778 Seppälä, M.: Synthesis of studies of palsa formation underlining the importance of local environmental
 779 and physical characteristics, *Quat. Res.*, 75, 366–370, <https://doi.org/10.1016/j.yqres.2010.09.007>,
 780 2011.
- 781 Siljanen, H. M. P., Alves, R. J. E., Ronkainen Jussi, G., Lamprecht, R. E., Bhattarai, H. R., Bagnoud,
 782 A., Marushchak, M. E., Martikainen, P. J., Schleper, C., and Biasi, C.: Archaeal nitrification is a key
 783 driver of high nitrous oxide emissions from arctic peatlands, *Soil Biol. Biochem.*, 107539,
 784 <https://doi.org/10.1016/j.soilbio.2019.107539>, 2019.
- 785 Song, C., Wang, X., Miao, Y., Wang, J., Mao, R., and Song, Y.: Effects of permafrost thaw on carbon
 786 emissions under aerobic and anaerobic environments in the Great Hing'an Mountains, China, *Science*
 787 *of the Total Environment*, 487, 604–610, <https://doi.org/10.1016/j.scitotenv.2013.09.083>, 2014.



- 788 Song, Y., Song, C., Hou, A., Ren, J., Wang, X., Cui, Q., and Wang, M.: Effects of temperature and root
 789 additions on soil carbon and nitrogen mineralization in a predominantly permafrost peatland, *Catena*
 790 (Amst)., 165, 381–389, <https://doi.org/10.1016/J.CATENA.2018.02.026>, 2018.
- 791 Song, Y., Cheng, X., Song, C., Li, M., Gao, S., Liu, Z., Gao, J., and Wang, X.: Soil CO₂ and N₂O
 792 emissions and microbial abundances altered by temperature rise and nitrogen addition in active-layer
 793 soils of permafrost peatland, *Front. Microbiol.*, 13, 1093487,
 794 <https://doi.org/10.3389/fmicb.2022.1093487>, 2022.
- 795 Song, Y., Feng, H., Luo, S., Qi, J., Wang, X., Mei, W., Chen, N., Feng, Y., and Song, C.: Vertical
 796 profiles of soil CO₂ and CH₄ emissions and their driving mechanisms in different types of permafrost
 797 peatlands, *J. Hydrol. (Amst).*, 665, 134754, <https://doi.org/10.1016/j.jhydrol.2025.134754>, 2026.
- 798 Spiller, A., Kallenbach, C. M., Burnett, M. S., Olefeldt, D., Schulze, C., Maranger, R., and Douglas, P.
 799 M. J.: Gradual drying of permafrost peat decreases carbon dioxide production in drier peat plateaus but
 800 not in wetter fens and bogs, *SOIL*, 11, 371–379, <https://doi.org/10.5194/soil-11-371-2025>, 2025.
- 801 Strauss, J., Biasi, C., Sanders, T., Abbott, B. W., von Deimling, T. S., Voigt, C., Winkel, M.,
 802 Marushchak, M. E., Kou, D., Fuchs, M., Horn, M. A., Jongejans, L. L., Liebner, S., Nitzbon, J.,
 803 Schirmer, L., Walter Anthony, K., Yang, Y., Zubrzycki, S., Laboor, S., Treat, C., and Grosse, G.:
 804 A globally relevant stock of soil nitrogen in the Yedoma permafrost domain, *Nat. Commun.*, 13, 1–9,
 805 <https://doi.org/10.1038/s41467-022-33794-9>, 2022.
- 806 Streletskiy, D. A., Nyland, K. E., Shiklomanov, N. I., Haggerty, E., Mann, M. L., Tolmanov, V., Nelson,
 807 F. E., Klene, A. E., Pretis, F., Chen, L., Christiansen, H., de Pablo, M. A., Fyodorov-Davydov, D.,
 808 Guglielmin, M., Hrbáček, F., Isaksen, K., Johansson, M., Leibman, M., Trombotto Liaudat, D.,
 809 Maslakov, A., Mastepanov, M., Noetzli, J., Romanovsky, V. E., Sharkhuu, A., Smith, S. L., Abramov,
 810 A., Babkina, E., Boike, J., Burn, C. R., Cannone, N., Cao, W., Davydov, S., Davydova, A., Delaloe, R.,
 811 Douglas, T., Duchesne, C., Grebenets, V., Hauck, C., Hilbich, C., Hölzle, M., Kaufmann, R. K.,
 812 Kholodov, A. L., Khomutov, A. V., Kňázková, M., Konstantinov, P., Kuziakin, L., Lambiel, C., Li, G.,



- 813 Lupachev, A., Magnin, F., Sharkhuu, N., Orekhov, P., Pellet, C., Phillips, M., Pogliotti, P., Ponti, S.,
 814 Rudd, D. A., Tregubov, O., Urban, F., Vieira, G., Vieli, A., Yu, W., Zamolodchikov, D., and Irrgang,
 815 A.: Long-term monitoring of active layer thickness confirms global permafrost degradation,
 816 Communications Earth & Environment 2026 7:1, 7, 671-, [https://doi.org/10.1038/s43247-026-03824-](https://doi.org/10.1038/s43247-026-03824-1)
 817 1, 2026.
- 818 Swindles, G. T., Morris, P. J., Mullan, D., Watson, E. J., Turner, T. E., Roland, T. P., Amesbury, M. J.,
 819 Kokfelt, U., Schoning, K., Pratte, S., Gallego-Sala, A., Charman, D. J., Sanderson, N., Garneau, M.,
 820 Carrivick, J. L., Woulds, C., Holden, J., Parry, L., and Galloway, J. M.: The long-term fate of permafrost
 821 peatlands under rapid climate warming, Sci. Rep., 5, 17951, <https://doi.org/10.1038/srep17951>, 2015.
- 822 Szafranek-Nakoneczna, A. and Stępniewska, Z.: Aerobic and anaerobic respiration in profiles of
 823 Polesie Lubelskie peatlands, Int. Agrophys., 28, 219–229, <https://doi.org/10.2478/intag-2014-0011>,
 824 2014.
- 825 Tarkhov, M. O., Matyshak, G. V., Ryzhova, I. M., Goncharova, O. Y., Bobrik, A. A., Petrov, D. G.,
 826 and Petrzhik, N. M.: Temperature Sensitivity of Soil Respiration in Palsa Peatlands of the North of
 827 Western Siberia, Eurasian Soil Science, 52, 945–953, <https://doi.org/10.1134/S1064229319080155>,
 828 2019.
- 829 Treat, C. C., Wollheim, W. M., Varner, R. K., Grandy, A. S., Talbot, J., and Frohling, S.: Temperature
 830 and peat type control CO₂ and CH₄ production in Alaskan permafrost peats, Glob. Chang. Biol., 20,
 831 2674–2686, <https://doi.org/10.1111/gcb.12572>, 2014.
- 832 Treat, C. C., Jones, M. C., Camill, P., Gallego-Sala, A., Garneau, M., Harden, J. W., Hugelius, G.,
 833 Klein, E. S., Kokfelt, U., Kuhry, P., Loisel, J., Mathijssen, P. J. H., O'Donnell, J. A., Oksanen, P. O.,
 834 Ronkainen, T. M., Sannel, A. B. K., Talbot, J., Tarnocai, C., and Väiranta, M.: Effects of permafrost
 835 aggradation on peat properties as determined from a pan-Arctic synthesis of plant macrofossils, J.
 836 Geophys. Res. Biogeosci., 121, 78–94, <https://doi.org/10.1002/2015JG003061>, 2016.



- 837 Vance, E. D., Brookes, P. C., and Jenkinson, D. S.: An extraction method for measuring soil microbial
 838 biomass C, *Soil Biol. Biochem.*, 19, 703–707, [https://doi.org/10.1016/0038-0717\(87\)90052-6](https://doi.org/10.1016/0038-0717(87)90052-6), 1987.
- 839 Voigt, C., Marushchak, M. E., Lamprecht, R. E., Jackowicz-Korczyński, M., Lindgren, A., Mastepanov,
 840 M., Granlund, L., Christensen, T. R., Tahvanainen, T., Martikainen, P. J., and Biasi, C.: Increased
 841 nitrous oxide emissions from Arctic peatlands after permafrost thaw, *Proceedings of the National*
 842 *Academy of Sciences*, 114, 6238–6243, <https://doi.org/10.1073/pnas.1702902114>, 2017a.
- 843 Voigt, C., Lamprecht, R. E., Marushchak, M. E., Lind, S. E., Novakovskiy, A., Aurela, M., Martikainen,
 844 P. J., and Biasi, C.: Warming of subarctic tundra increases emissions of all three important greenhouse
 845 gases – carbon dioxide, methane, and nitrous oxide, *Glob. Chang. Biol.*, 23, 3121–3138,
 846 <https://doi.org/10.1111/gcb.13563>, 2017b.
- 847 Voigt, C., Marushchak, M. E., Mastepanov, M., Lamprecht, R. E., Christensen, T. R., Dorodnikov, M.,
 848 Jackowicz-Korczyński, M., Lindgren, A., Lohila, A., Nykänen, H., Oinonen, M., Oksanen, T., Palonen,
 849 V., Treat, C. C., Martikainen, P. J., and Biasi, C.: Ecosystem carbon response of an Arctic peatland to
 850 simulated permafrost thaw, *Glob. Chang. Biol.*, 25, 1746–1764, <https://doi.org/10.1111/gcb.14574>,
 851 2019.
- 852 Voigt, C., Virkkala, A. M., Hould Gosselin, G., Bennett, K. A., Black, T. A., Detto, M., Chevrier-Dion,
 853 C., Guggenberger, G., Hashmi, W., Kohl, L., Kou, D., Marquis, C., Marsh, P., Marushchak, M. E.,
 854 Nesic, Z., Nykänen, H., Saarela, T., Sauheitl, L., Walker, B., Weiss, N., Wilcox, E. J., and Sonnentag,
 855 O.: Arctic soil methane sink increases with drier conditions and higher ecosystem respiration, *Nat. Clim.*
 856 *Chang.*, 13, 1095–1104, <https://doi.org/10.1038/s41558-023-01785-3>, 2023.
- 857 Wang, Q., Zhao, X., Chen, L., Yang, Q., Chen, S., and Zhang, W.: Global synthesis of temperature
 858 sensitivity of soil organic carbon decomposition: Latitudinal patterns and mechanisms, *Funct. Ecol.*,
 859 33, 514–523, <https://doi.org/10.1111/1365-2435.13256>, 2019.



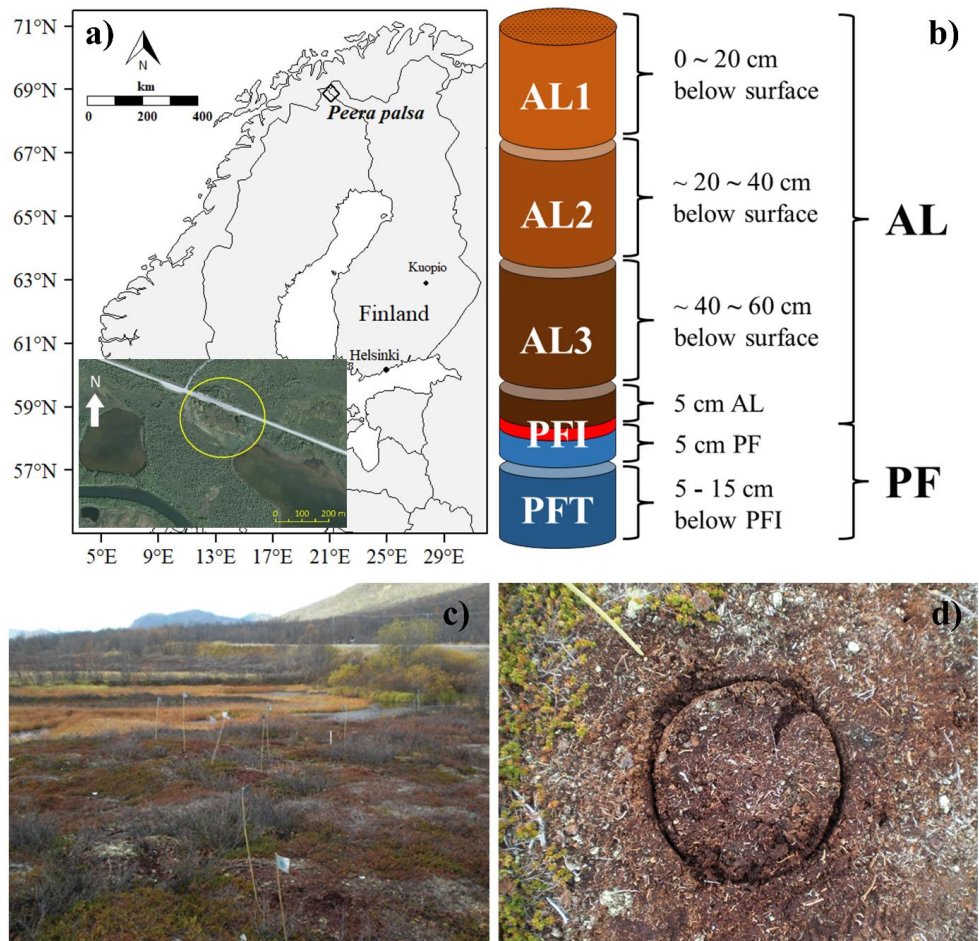
- 860 Wang, X., Song, C., Wang, J., Miao, Y., Mao, R., and Song, Y.: Carbon release from Sphagnum peat
 861 during thawing in a montane area in China, *Atmos. Environ.*, 75, 77–82,
 862 <https://doi.org/10.1016/j.atmosenv.2013.04.056>, 2013.
- 863 Weishaar, J. L., Aiken, G. R., Bergamaschi, B. A., Fram, M. S., Fujii, R., and Mopper, K.: Evaluation
 864 of specific ultraviolet absorbance as an indicator of the chemical composition and reactivity of dissolved
 865 organic carbon, *Environ. Sci. Technol.*, 37, 4702–4708, <https://doi.org/10.1021/es030360x>, 2003.
- 866 Wickham, H.: *ggplot2: Elegant Graphics for Data Analysis*, 2016., 2016.
- 867 Wickland, K. P., Waldrop, M. P., Aiken, G. R., Koch, J. C., Jorgenson, M. T., and Striegl, R. G.:
 868 Dissolved organic carbon and nitrogen release from boreal Holocene permafrost and seasonally frozen
 869 soils of Alaska, *Environmental Research Letters*, 13, 065011, [https://doi.org/10.1088/1748-](https://doi.org/10.1088/1748-9326/aac4ad)
 870 [9326/aac4ad](https://doi.org/10.1088/1748-9326/aac4ad), 2018.
- 871 Wild, B., Schneckner, J., Alves, R. J. E., Barsukov, P., Bárta, J., Čapek, P., Gentsch, N., Gittel, A.,
 872 Guggenberger, G., Lashchinskiy, N., Mikutta, R., Rusalimova, O., Šantrůčková, H., Shibistova, O.,
 873 Urich, T., Watzka, M., Zrazhevskaya, G., and Richter, A.: Input of easily available organic C and N
 874 stimulates microbial decomposition of soil organic matter in arctic permafrost soil, *Soil Biol. Biochem.*,
 875 75, 143–151, <https://doi.org/10.1016/J.SOILBIO.2014.04.014>, 2014.
- 876 Zhu, X., Burger, M., Doane, T. A., and Horwath, W. R.: Ammonia oxidation pathways and nitrifier
 877 denitrification are significant sources of N₂O and NO under low oxygen availability, *Proc. Natl. Acad.*
 878 *Sci. U. S. A.*, 110, 6328–6333, <https://doi.org/10.1073/pnas.1219993110>, 2013.
- 879 Zuo, Y., Song, Y., Jiang, L., Li, Y., Wang, Y., Chen, N., Jiang, P., Zheng, S., Song, C., Xu, X., Yuan,
 880 F., and Sun, L.: Long-term carbon release of peatland soil after permafrost thaw in northmost China,
 881 *Catena (Amst.)*, 261, 109551, <https://doi.org/10.1016/j.catena.2025.109551>, 2025.

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884 13 Figures

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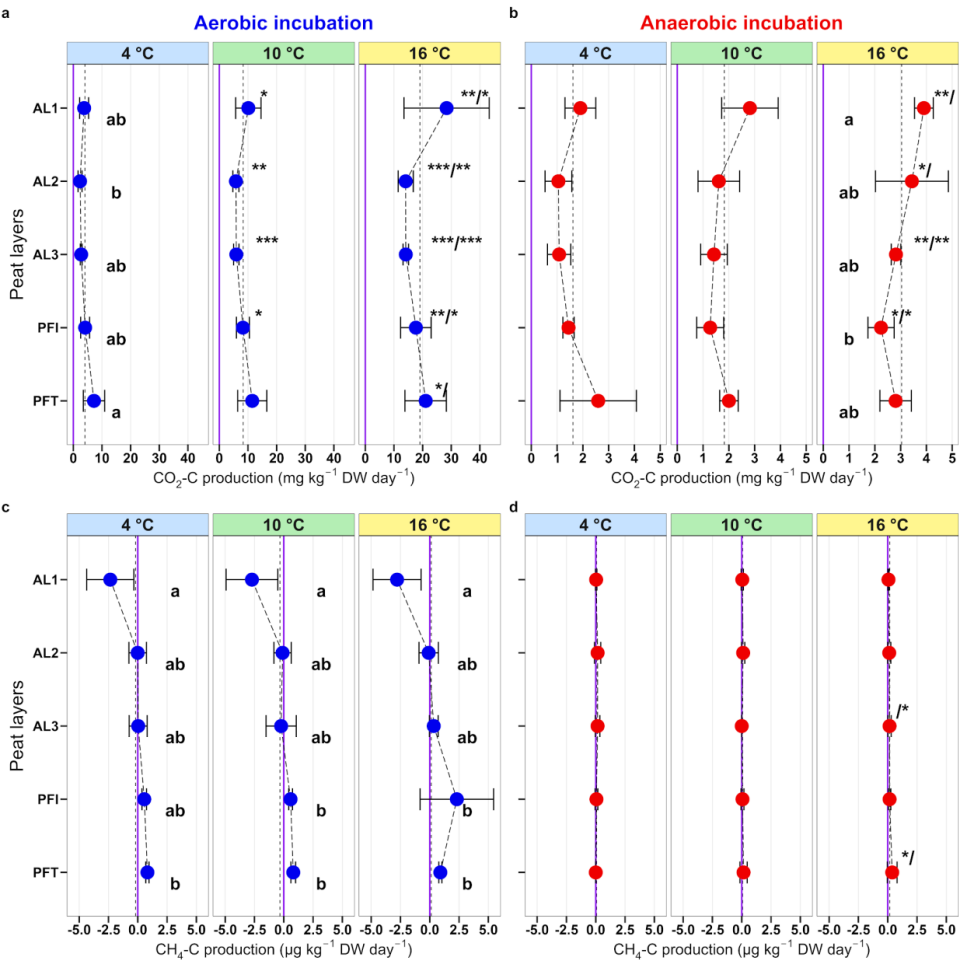
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887 **Fig. 1: Intact peat soil core sampling site “Peera palsa”.** In a) the sampling site location in northern Finland and a
888 satellite image of the palsa complex (insert image source: © 2019 Google Maps, Airbus, CNES / Airbus, Lantmäteriet
889 / Metria, Maxar Technologies). b) Scheme of the peat layer separation of the intact peat cores. The peat layer height of
890 the active layer fragments AL1, AL2 and AL3 varied by a few cm depending on the total depth of the active layer (see
891 supplementary information table S1). The permafrost interface layer (PFI) and the permafrost layer (PFT) had the
892 same height of 10 cm. The peat core diameter was 10 cm. c) Photograph of the palsa complex taken during the peat
893 core sampling. d) Photograph of a bare peat surface where the peat cores were sampled. Pictures were taken by C.
894 Voigt (2012).

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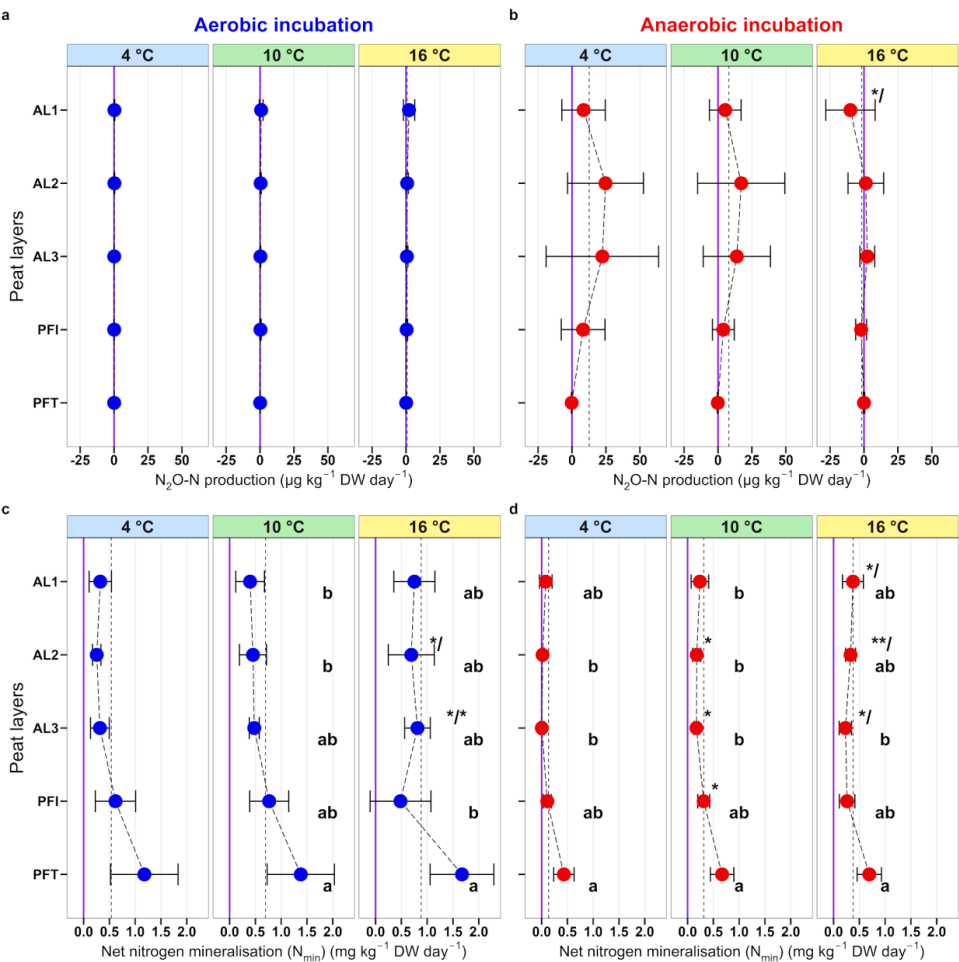
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898 Fig. 2: Daily aerobic (a) and anaerobic (b) CO₂-C production rate (mg kg⁻¹ DW (dry weight) day⁻¹) and aerobic (c)
899 and anaerobic (d) CH₄-C production rate (μg kg⁻¹ DW (dry weight) day⁻¹) at 4, 10 and 16 °C during the incubation
900 period of 161 days. Data points represent mean values; error bars show standard deviation (n = 4). Letters indicate
901 statistical differences between peat layers within the given temperature (One-way ANOVA and Tukey's HSD, P ≤ 0.05).
902 Asterisks indicate significant differences between incubation temperatures: for 10 °C, significant differences to 4 °C;
903 for 16 °C significant difference to 4 °C/10 °C (One-way ANOVA). Dashed grey line in the background represents the
904 mean value across all five peat layers at the given temperature. Level of significance: ***significant at P ≤ 0.001, **
905 significant at P ≤ 0.01, *significant at P ≤ 0.05. Note the different scales on X-axis for aerobic and anaerobic CO₂-C
906 production rate.

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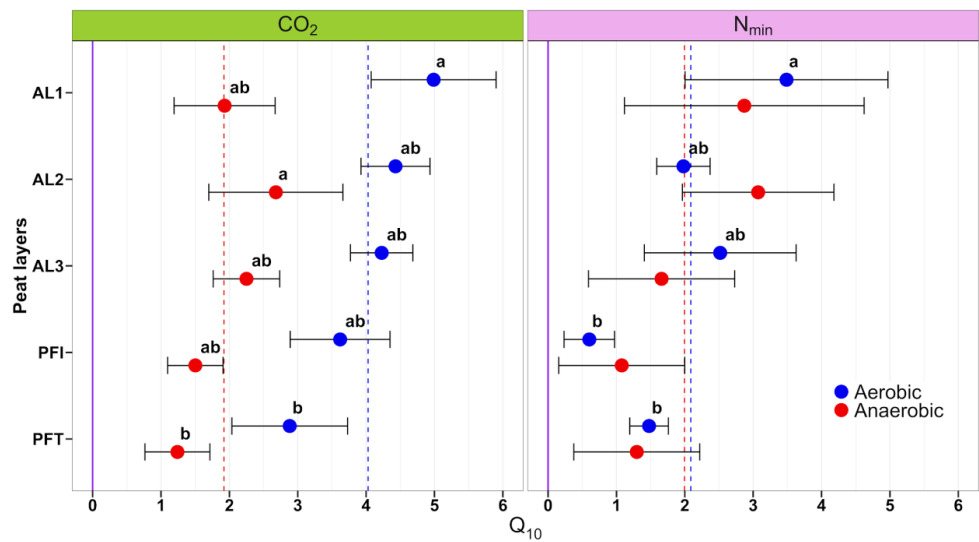
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Fig. 3: Daily aerobic (a) and anaerobic (b) N_2O-N production rate ($\mu g\ kg^{-1}\ DW$ (dry weight) day $^{-1}$) and aerobic (c) and anaerobic (d) net nitrogen mineralization rate ($mg\ kg^{-1}\ DW$ (dry weight) day $^{-1}$) at 4, 10 and 16 °C during the incubation period of 161 days. Data points represent mean values; error bars show standard deviation (n = 4). Letters indicate statistical differences between peat layers within the given temperature (One-way ANOVA and Tukey's HSD, $P \leq 0.05$). Asterisks indicate significant differences between incubation temperatures: for 10 °C, significant differences to 4 °C; for 16 °C significant difference to 4 °C/10 °C (One-way ANOVA). Dashed grey line in the background represents the mean value across all five peat layers at the given temperature. Level of significance: ***significant at $P \leq 0.001$, ** significant at $P \leq 0.01$, *significant at $P \leq 0.05$.

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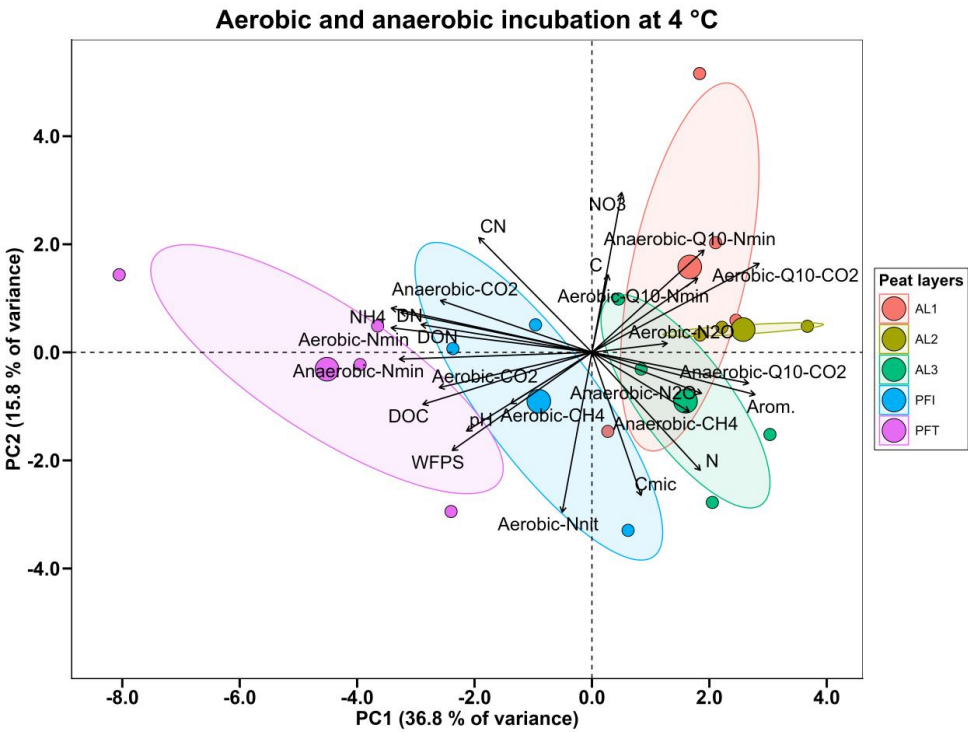
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921 Fig. 4: Temperature dependence of the five peat layers shown as Q10 values calculated for CO₂-C production (CO₂ -
922 left panel) and net nitrogen mineralization (N_{min} - right panel) in aerobic (blue) and anaerobic (red) incubation
923 conditions. Data points represent mean values; error bars show standard deviation (n = 4). Letters indicate statistical
924 differences (One-way ANOVA and Tukey's HSD, P ≤ 0.05) between the peat layers. Dashed coloured lines represent
925 the mean Q10 values across all peat layers at aerobic and anaerobic incubation conditions.

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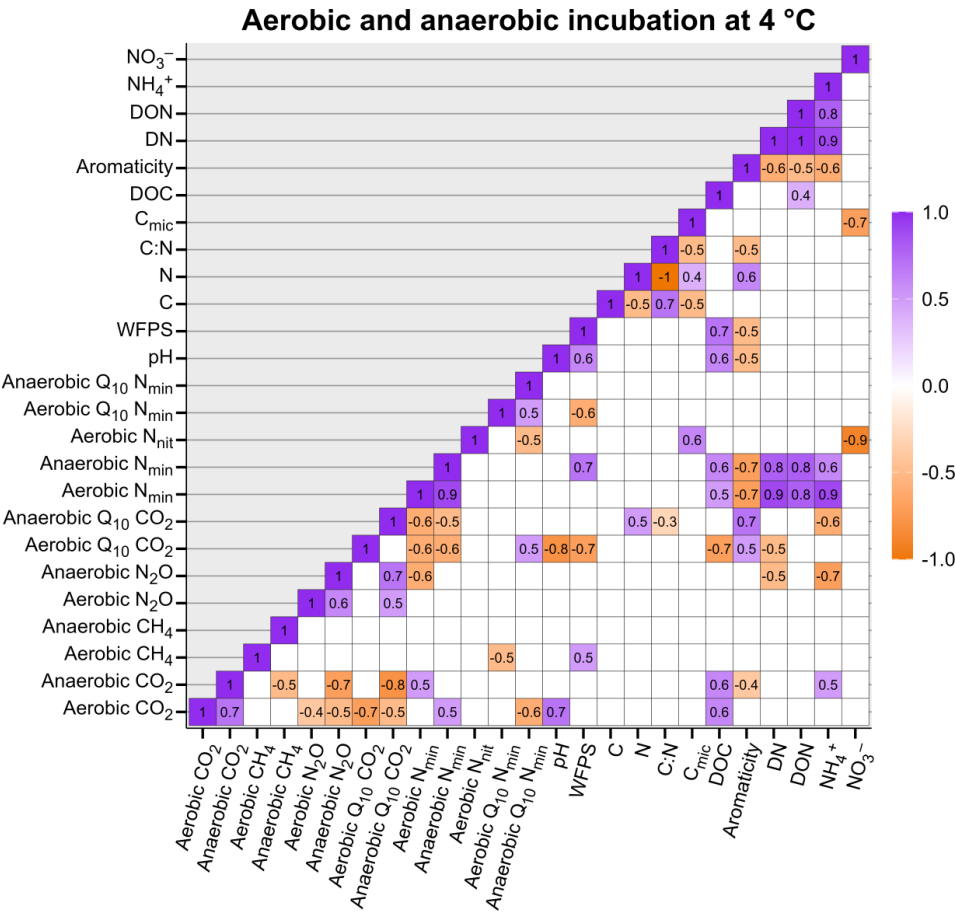
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929 **Fig. 5: Principal component analysis (PCA) bi-plot of aerobic and anaerobic incubations at 4 °C. The analysed data**
930 **includes the cumulative GHG gas production over 161 incubation days; the temperature sensitivity of CO₂ production,**
931 **Q₁₀CO₂, and nitrogen mineralization, Q₁₀N_{min}; the net nitrogen mineralization rate (N_{min}) and net nitrification rate**
932 **(N_{nitr}) (aerobic conditions only); in combination with soil chemical and physical parameters. Largest variability were**
933 **shown in the incubation temperature of 4 °C. Individual contributions of different variables to PC1 and PC2 are shown**
934 **in Figure S11.**

935



936

937 **Fig. 6: Pearson's correlation matrix of aerobic and anaerobic incubations at 4 °C. The analysed data includes the**
938 **cumulative GHG gas production over 161 incubation days; the temperature sensitivity of CO₂ production, Q₁₀CO₂ and**
939 **nitrogen mineralization, Q₁₀N_{min}; the net nitrogen mineralization rate (N_{min}) and net nitrification rate (N_{nit}) (aerobic**
940 **conditions only); in combination with soil chemical and physical parameters.**



14 Tables

Table 1: Peat characteristics of the peat layers analysed before the incubation experiment. Dry bulk density (BD), soil organic matter content (SOM - loss of ignition), water content (H₂O) water saturation from water holding capacity (WHC), water filled pore space (WFPS), total carbon (C) and total nitrogen (N) content, C:N ratio, pH, K₂SO₄ extractable dissolved organic carbon (DOC), total dissolved nitrogen (DON), aromaticity of dissolved organic carbon, microbial biomass carbon (C_{mic}), K₂SO₄ extractable ammonium (NH₄⁺) and H₂O extractable nitrate (NO₃⁻). Values are mean ± standard deviation (n = 4). Different letters indicate statistical differences between the peat layers (One-way ANOVA and Tukey's HSD, P ≤ 0.05).

Peat layer *	pH	BD (g cm ⁻³)	H2O (%)	WHC (%)	WFPS (%)	SOM (%)	C (%)	N (%)	C:N
AL1	3.7 ± 0.2 b	0.16 ± 0.05 a	68.8 ± 3.2 c	46.6 ± 12.5 b	40.7 ± 15.2 b	99.94 ± 0.08	47.8 ± 5.0	1.3 ± 0.4	30.8 ± 8.3
AL2	3.6 ± 0.1 b	0.16 ± 0.03 ab	72.0 ± 1.9 bc	52.5 ± 6.6 b	45.4 ± 9.7 b	99.90 ± 0.10	47.2 ± 6.8	1.8 ± 0.3	27.5 ± 6.8
AL3	3.9 ± 0.3 ab	0.12 ± 0.03 ab	76.5 ± 3.2 b	49.2 ± 7.5 b	43.7 ± 8.3 b	99.91 ± 0.10	47.1 ± 6.6	1.7 ± 0.2	29.4 ± 7.6
PFI	4.2 ± 0.5 ab	0.10 ± 0.01 ab	83.9 ± 2.2 a	57.4 ± 8.9 ab	58.5 ± 9.9 ab	99.91 ± 0.07	47.5 ± 4.4	1.5 ± 0.4	33.3 ± 9.7
PFT	4.7 ± 0.7 a	0.09 ± 0.01 b	88.9 ± 1.7 a	73.3 ± 8.1 a	76.9 ± 6.9 a	99.92 ± 0.04	47.7 ± 1.7	1.4 ± 0.4	37.5 ± 11.9
Average	4.0 ± 0.6	0.13 ± 0.04	78.0 ± 8.0	55.8 ± 12.6	53.1 ± 16.5	99.92 ± 0.07	47.4 ± 4.7	1.6 ± 0.3	31.7 ± 8.8
Peat layer *	C _{mic} (mg C kg ⁻¹ DW)	DOC (mg C kg ⁻¹ DW)	Aromaticity (% DOC)	DN (mg N kg ⁻¹ DW)	DON (mg N kg ⁻¹ DW)	NH ₄ ⁺ (mg N kg ⁻¹ DW)	NO ₃ ⁻ (mg N kg ⁻¹ DW)		
AL1	710.8 ± 295.4	854.9 ± 186.7 b	28.9 ± 9.6 a	166.2 ± 94.5	131.0 ± 82.8	29.8 ± 16.9	5.4 ± 5.1		
AL2	847.6 ± 243.8	964.4 ± 322.7 ab	32.8 ± 2.7 a	117.9 ± 55.5	100.7 ± 57.5	14.2 ± 5.2	3.4 ± 2.7		
AL3	1337.4 ± 1022.9	838.3 ± 142.9 b	24.7 ± 4.4 ab	140.8 ± 77.3	111.2 ± 58.8	28.1 ± 19.2	1.5 ± 1.5		
PFI	1275.8 ± 707.6	915.3 ± 133.1 ab	20.8 ± 4.1 ab	177.7 ± 81.3	103.1 ± 31.0	73.6 ± 50.0	0.9 ± 0.8		
PFT	880.9 ± 305.2	1525.4 ± 481.9 a	12.9 ± 4.2 b	359.1 ± 199.3	233.1 ± 107.1	123.3 ± 95.7	2.6 ± 3.2		
Average	1010.5 ± 589.5	1019.7 ± 366.1	24.0 ± 8.6	192.3 ± 133.9	135.8 ± 82.0	53.8 ± 60.3	2.7 ± 3.1		

* Active layer 1, 2, 3 (AL1, 2, 3), permafrost interface (PFI), permafrost table (PFT) (see Fig. 1b)



948 **Table 2: Aerobic and anaerobic incubation differences: linear mixed effects model estimates of fixed effects, SE, t-**
 949 **value, lower and upper confidence intervals and P-values for cumulative production of CO₂, CH₄ and N₂O, net N**
 950 **mineralization (N_{min}) rates and the Q₁₀ of CO₂ production and net N mineralization.**

Fixed effects	Estimate	SE	t-value	2.5% CI	97.5% CI	P-value	Significance
CO ₂ production							
(Intercept)	3.099	0.054	57.552	2.992	3.206	<0.001	***
Oxic condition: Anaerobic	-0.614	0.052	-11.795	-0.717	-0.511	<0.001	***
CH ₄ production							
(Intercept)	-0.037	0.019	-1.914	-0.075	0.001	0.058	
Oxic condition: Anaerobic	0.044	0.023	1.959	-0.001	0.089	0.053	
N ₂ O production							
(Intercept)	0.076	0.358	0.213	-0.632	0.785	0.832	
Oxic condition: Anaerobic	0.948	0.389	2.433	0.176	1.719	0.017	*
Net N mineralization (N _{min})							
(Intercept)	0.213	0.022	9.559	0.169	0.257	<0.001	***
Oxic condition: Anaerobic	-0.113	0.018	-6.168	-0.149	-0.076	<0.001	***
Q ₁₀ CO ₂ production (Q ₁₀ CO ₂)							
(Intercept)	4.029	0.154	26.107	3.723	4.334	<0.001	***
Oxic condition: Anaerobic	-2.108	0.155	-13.599	-2.415	-1.801	<0.001	***
Q ₁₀ net N mineralization (Q ₁₀ N _{min})							
(Intercept)	2.091	0.213	9.836	1.670	2.513	<0.001	***
Oxic condition: Anaerobic	-0.096	0.233	-0.411	-0.557	0.366	0.682	

Level of significance: * p < 0.05, ** p < 0.01, *** p < 0.001

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Table 3: Linear mixed effects model estimates of fixed effects, SE, t-value, lower and upper confidence intervals and P-values of aerobic and anaerobic cumulative production of CO₂, CH₄ and the Q₁₀ of CO₂ production.

Fixed effects	Estimate	SE	t-value	2.5% CI	97.5% CI	P-value	Significance
Aerobic CO₂ production							
(Intercept)	2.835	0.071	39.853	2.692	2.978	<0.001	***
Incubation temperature: 10 °C	0.338	0.037	9.205	0.264	0.411	<0.001	***
Incubation temperature: 16 °C	0.701	0.037	19.103	0.627	0.775	<0.001	***
Peat layer: AL2	-0.223	0.047	-4.716	-0.319	-0.128	<0.001	***
Peat layer: AL3	-0.190	0.047	-4.001	-0.285	-0.094	<0.001	***
Peat layer: PFI	-0.067	0.047	-1.405	-0.162	0.029	0.166	
Peat layer: PFT	0.066	0.047	1.394	-0.029	0.161	0.169	
Anaerobic CO₂ production							
(Intercept)	2.503	0.061	40.998	2.381	2.626	<0.001	***
Incubation temperature: 10 °C	0.066	0.051	1.294	-0.036	0.168	0.202	
Incubation temperature: 16 °C	0.313	0.051	6.168	0.211	0.415	<0.001	***
Peat layer: AL2	-0.208	0.066	-3.171	-0.340	-0.076	0.003	**
Peat layer: AL3	-0.227	0.066	-3.468	-0.359	-0.096	0.001	**
Peat layer: PFI	-0.231	0.066	-3.518	-0.362	-0.099	<0.001	***
Peat layer: PFT	-0.055	0.066	-0.840	-0.187	0.077	0.405	
Aerobic CH₄ production							
(Intercept)	-0.292	0.047	-6.252	-0.386	-0.198	<0.001	***
Incubation temperature: 10 °C	-0.021	0.035	-0.600	-0.092	0.049	0.551	
Incubation temperature: 16 °C	0.009	0.035	0.267	-0.061	0.080	0.791	
Peat layer: AL2	0.288	0.045	6.360	0.197	0.379	<0.001	***
Peat layer: AL3	0.295	0.045	6.513	0.204	0.386	<0.001	***
Peat layer: PFI	0.361	0.045	7.961	0.270	0.452	<0.001	***
Peat layer: PFT	0.351	0.045	7.754	0.260	0.442	<0.001	***
Anaerobic CH₄ production							
(Intercept)	0.001	0.004	0.296	-0.007	0.009	0.768	
Incubation temperature: 10 °C	0.000	0.004	-0.082	-0.008	0.007	0.935	
Incubation temperature: 16 °C	0.006	0.004	1.705	-0.001	0.014	0.094	
Peat layer: AL2	0.006	0.005	1.211	-0.004	0.016	0.232	
Peat layer: AL3	0.004	0.005	0.782	-0.006	0.014	0.438	
Peat layer: PFI	0.003	0.005	0.634	-0.007	0.013	0.529	
Peat layer: PFT	0.009	0.005	1.751	-0.001	0.018	0.086	
Aerobic Q₁₀ CO₂ production							
(Intercept)	4.986	0.258	19.312	4.468	5.504	<0.001	***
Peat layer: AL2	-0.559	0.213	-2.627	-0.987	-0.132	0.011	*
Peat layer: AL3	-0.756	0.213	-3.551	-1.183	-0.329	<0.001	***
Peat layer: PFI	-1.367	0.213	-6.418	-1.794	-0.939	<0.001	***
Peat layer: PFT	-2.105	0.213	-9.886	-2.532	-1.678	<0.001	***
Anaerobic Q₁₀ CO₂ production							
(Intercept)	1.932	0.189	10.237	1.553	2.310	<0.001	***
Peat layer: AL2	0.750	0.232	3.236	0.285	1.214	0.002	**
Peat layer: AL3	0.319	0.232	1.375	-0.146	0.783	0.175	
Peat layer: PFI	-0.43	0.232	-1.856	-0.895	0.035	0.069	
Peat layer: PFT	-0.693	0.232	-2.991	-1.157	-0.228	0.004	**

Level of significance: * p < 0.05, ** p < 0.01, *** p < 0.001



Table 4: Linear mixed effects model estimates of fixed effects, SE, t-value, lower and upper confidence intervals and P-values of aerobic and anaerobic cumulative production of N₂O, the net N mineralization and net nitrification rates and the Q₁₀ of net N mineralization.

Fixed effects	Estimate	SE	t-value	2.5% CI	97.5% CI	P-value	Significance
Aerobic N₂O production							
(Intercept)	0.127	0.066	1.916	-0.006	0.259	0.061	
Incubation temperature: 10 °C	0.038	0.057	0.668	-0.076	0.153	0.507	
Incubation temperature: 16 °C	0.113	0.057	1.977	-0.002	0.227	0.053	
Peat layer: AL2	-0.095	0.074	-1.286	-0.242	0.053	0.204	
Peat layer: AL3	-0.120	0.074	-1.625	-0.267	0.028	0.110	
Peat layer: PFI	-0.128	0.074	-1.743	-0.276	0.019	0.087	
Peat layer: PFT	-0.161	0.074	-2.183	-0.308	-0.013	0.034	*
Anaerobic N₂O production							
(Intercept)	1.234	1.049	1.176	-0.872	3.339	0.245	
Incubation temperature: 10 °C	-0.754	0.871	-0.865	-2.503	0.995	0.391	
Incubation temperature: 16 °C	-2.313	0.871	-2.655	-4.062	-0.564	0.011	*
Peat layer: AL2	2.104	1.125	1.870	-0.154	4.361	0.067	
Peat layer: AL3	1.865	1.125	1.658	-0.393	4.123	0.103	
Peat layer: PFI	0.324	1.125	0.288	-1.934	2.582	0.774	
Peat layer: PFT	-0.232	1.125	-0.206	-2.490	2.026	0.838	
Aerobic net N mineralization (N_{min})							
(Intercept)	0.121	0.04	3.043	0.041	0.201	0.004	**
Incubation temperature: 10 °C	0.044	0.026	1.684	-0.008	0.096	0.098	
Incubation temperature: 16 °C	0.084	0.026	3.236	0.032	0.137	0.002	**
Peat layer: AL2	-0.006	0.034	-0.186	-0.074	0.061	0.853	
Peat layer: AL3	0.016	0.034	0.475	-0.052	0.084	0.637	
Peat layer: PFI	0.031	0.034	0.913	-0.037	0.098	0.366	
Peat layer: PFT	0.203	0.034	6.026	0.135	0.271	<0.001	***
Anaerobic net N mineralization (N_{min})							
(Intercept)	0.036	0.017	2.084	0.001	0.071	0.042	*
Incubation temperature: 10 °C	0.063	0.012	5.207	0.039	0.088	<0.001	***
Incubation temperature: 16 °C	0.084	0.012	6.949	0.06	0.109	<0.001	***
Peat layer: AL2	-0.015	0.016	-0.954	-0.047	0.017	0.345	
Peat layer: AL3	-0.033	0.015	-2.134	-0.064	-0.002	0.038	*
Peat layer: PFI	0.001	0.015	0.077	-0.03	0.032	0.939	
Peat layer: PFT	0.113	0.015	7.343	0.082	0.144	<0.001	***
Aerobic net nitrification (N_{nit})							
(Intercept)	-1.198	0.09	-13.254	-1.379	-1.017	<0.001	***
Incubation temperature: 10 °C	0.278	0.082	3.409	0.114	0.442	0.001	**
Incubation temperature: 16 °C	0.159	0.082	1.953	-0.004	0.323	0.056	
Peat layer: AL2	0.161	0.105	1.531	-0.05	0.373	0.132	
Peat layer: AL3	0.141	0.105	1.34	-0.07	0.353	0.186	
Peat layer: PFI	0.174	0.105	1.654	-0.037	0.386	0.104	
Peat layer: PFT	0.025	0.105	0.24	-0.186	0.237	0.812	

Level of significance: * p < 0.05, ** p < 0.01, *** p < 0.001

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959 **Table 4: continued**

Aerobic Q_{10} net N mineralization							
(Intercept)	3.486	0.240	14.538	3.005	3.968	<0.001	***
Peat layer: AL2	-1.506	0.325	-4.627	-2.160	-0.852	<0.001	***
Peat layer: AL3	-0.969	0.325	-2.978	-1.623	-0.315	0.004	**
Peat layer: PFI	-2.870	0.353	-8.128	-3.579	-2.161	<0.001	***
Peat layer: PFT	-2.010	0.325	-6.175	-2.663	-1.356	<0.001	***
Anaerobic Q_{10} net N mineralization							
(Intercept)	2.868	0.408	7.028	2.049	3.686	<0.001	***
Peat layer: AL2	0.204	0.377	0.540	-0.553	0.960	0.592	
Peat layer: AL3	-1.206	0.377	-3.196	-1.962	-0.449	0.002	**
Peat layer: PFI	-1.789	0.377	-4.743	-2.546	-1.033	<0.001	***
Peat layer: PFT	-1.570	0.377	-4.162	-2.327	-0.814	<0.001	***

Level of significance: * $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$

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