



1 **Kinetic Partitioning with Effective Mass Accommodation for Secondary Organic Aerosol**
2 **Simulations in Chemical Transport Modeling**

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10 **Abstract**

11 Secondary organic aerosols (SOA) can adopt liquid, amorphous semi-solid, or glassy solid states
12 in the atmosphere. Previous model simulations have revealed that SOA adopt semisolid or glassy
13 solid phase states over dry regions and in the upper troposphere. A highly viscous phase state in
14 SOA particles can cause kinetic limitations in SOA partitioning, challenging traditional SOA
15 treatments with equilibrium partitioning in chemical transport modeling. In this study, we develop
16 a method to simulate kinetic partitioning of SOA accounting for the impacts of SOA phase state
17 by utilizing effective mass accommodation coefficient (α_{eff}). We initially test this method in box
18 modeling to calculate equilibration timescale of SOA partitioning for a wide parameter space for
19 bulk diffusivity, volatility, and total aerosol mass loading. Comparison with the chemical timestep
20 of 20 min in GEOS-Chem reveals that equilibrium partitioning is still applicable for particles with
21 bulk diffusivity $> 10^{-15} \text{ cm}^2 \text{ s}^{-1}$, whereas kinetic partitioning is required for partitioning of semi-
22 and low-volatile compounds into highly viscous particles. To reduce computational burden, we
23 developed analytical equations with gas-particle mass transfer coefficient by effectively
24 accounting for gas and bulk diffusion limitations. The application of the new α_{eff} kinetic
25 partitioning scheme into GEOS-Chem predicts lowers SOA mass over drylands (e.g., Western US)
26 in the boundary layer due to kinetically limited growth of viscous SOA particles, whereas SOA
27 mass can be described well with equilibrium partitioning in most other areas at the surface and at
28 850 hPa. In the upper troposphere at 500 hPa, the α_{eff} -kinetic scheme predicts up to 50% higher



29 SOA mass as evaporation is suppressed for glassy particles. The developed method is easily
30 applicable, improving description and treatments of SOA in chemical transport models.

31

32 **Short summary:** The occurrence of semisolid and glassy phase states in SOA challenges
33 traditional model treatments of equilibrium partitioning of SOA. We developed an effective
34 method to treat kinetic partitioning of SOA by resolving the effects of particle phase state in
35 chemical transport modeling. The application in GEOS-Chem reveals that SOA budget is impacted
36 significantly especially over drylands and in the upper troposphere.

37

38 1 Introduction

39 Secondary organic aerosols (SOA) are ubiquitous and represent a major fraction of the submicron
40 particulate matter in the troposphere, playing a central role in aerosol effects on climate, air quality,
41 and public health (Jimenez et al., 2009; Pöschl & Shiraiwa, 2015; Shrivastava et al., 2017). Unlike
42 primary organic aerosols which are directly emitted into the atmosphere, SOA are formed in the
43 atmosphere from oxidation of biogenic or anthropogenic volatile organic compounds (VOCs) into
44 lower-volatility products that may nucleate or condense onto pre-existing particles (Hallquist et
45 al., 2009; Kirkby et al., 2023; Kroll & Seinfeld, 2008). SOA formation has been extensively studied
46 in laboratory experiments, but there is still significant gap between ambient measurements and
47 model simulations of SOA concentrations and properties by chemical transport models (CTMs)
48 (Hallquist et al., 2009; Hodzic et al., 2016; Tsigaridis et al., 2014).

49 An important physical property of SOA particles is particle phase state which can vary
50 from liquid (characterized by dynamic viscosity $\eta < 10^2$ Pa s), semisolid ($10^2 < \eta \leq 10^{12}$ Pa s), to
51 glassy solid states ($\eta > 10^{12}$ Pa s) depending on the chemical composition, temperature, and relative
52 humidity (Koop et al., 2011; Reid et al., 2018; Virtanen et al., 2010). Traditionally, SOA particles
53 are assumed implicitly to be liquid-like in CTMs (Chung & Seinfeld, 2002; Pye & Seinfeld, 2010).
54 The assumption holds that the gas-particle equilibration timescale is rapid in comparison to other
55 atmospheric processes implying that SOA particles achieve instantaneous equilibrium partitioning
56 between the gas and particle phases (Pankow, 2003). However, there has been mounting evidence
57 of laboratory measurements and ambient observations showing that SOA can exist in glassy solid



58 or amorphous semisolid states as influenced by chemical composition and ambient conditions
59 (Jensen et al., 2025; O'Brien et al., 2014; Saukko et al., 2012; Vaden et al., 2011; Virtanen et al.,
60 2010).. A highly viscous phase state can lead to prolonged equilibration timescales of SOA
61 partitioning, depending on volatility of species and particle mass loadings (Li & Shiraiwa, 2019;
62 Schervish et al., 2024). We have developed methods to predict SOA phase state and viscosity
63 (DeRieux et al., 2018; Li et al., 2020; Shiraiwa et al., 2017); their applications in CTMs suggest
64 that SOA particles should adopt a viscous phase state over drylands and in the free troposphere
65 (Luu et al., 2025; Schmedding et al., 2020; Shiraiwa et al., 2017; Zhang et al., 2024). The major
66 limitation with most SOA phase state modeling studies is that the phase state prediction was
67 applied as offline and post-processing calculations, while there are several studies explicitly
68 resolving kinetic SOA partitioning (He et al., 2025; Shrivastava et al., 2024).

69 Implementing kinetic partitioning with explicit consideration of the effect of SOA phase
70 state is essential in CTMs to improve SOA budgets and thereby aerosol impacts on air quality,
71 climate, and public health. Process-level microphysics models such as the kinetic multi-layer
72 model of gas-particle interactions in aerosols and clouds (KM-GAP) (Shiraiwa et al., 2012) can
73 resolve kinetic limitations imparted by semi-solid or glassy solid phase states, but applying the
74 same methodology to a CTM would drastically increase computational costs. Therefore, in this
75 work, we develop a model scheme for the partitioning of organic compounds that is modulated by
76 online calculations of SOA phase state by utilizing an effective mass accommodation coefficient
77 (Shiraiwa & Pöschl, 2021). We investigate the influence of particle phase state, volatility, and total
78 organic aerosol concentration on gas-particle partitioning of organic compounds in SOA. Then,
79 we implement our partitioning scheme into the global CTM, GEOS-Chem, to evaluate the impacts
80 of SOA phase state on SOA budgets.

81 **2 Methods**

82 **2.1 Validation of partitioning method with a box model**

83 Based on absorptive partitioning theory (Pankow, 1994), the mass transfer rates between gas and
84 particle phases can be expressed for gas- (C^g) and particle-phase (C^p) mass concentrations (in μg
85 m^{-3}) of semi-volatile compounds with the following ordinary differential equations (Shiraiwa &
86 Pöschl, 2021; Trump et al., 2014):



$$\frac{dC^g}{dt} = -k_{gp} \left(C^g - \frac{C^p}{C_{OA}} C^0 \right) \quad (1)$$

$$\frac{dC^p}{dt} = k_{gp} \left(C^g - \frac{C^p}{C_{OA}} C^0 \right) \quad (2)$$

Partitioning is also dependent on the ratio of particle-phase concentration, C^p , and C_{OA} ($\mu\text{g m}^{-3}$), the total organic aerosol concentration. C^0 ($\mu\text{g m}^{-3}$) is the temperature-dependent equilibrium vapor concentration and is described via the Clausius-Clapeyron relation:

$$C^0(T) = C^0(298 \text{ K}) \exp \left(\frac{\Delta H_{\text{vap}}}{R} \left(\frac{1}{298 \text{ K}} - \frac{1}{T} \right) \right) \quad (3)$$

where R is the molar gas constant in $\text{J mol}^{-1} \text{K}^{-1}$ and T refers to the ambient temperature in K. Species with a low volatility tend to have a higher enthalpy of vaporization, ΔH_{vap} (kJ mol^{-1}), as described by the semi-empirical parameterization (Stolzenburg et al., 2018; Stolzenburg et al., 2022):

$$\Delta H_{\text{vap}} = -5.7 \times \log_{10} C^0(298 \text{ K}) + 129 \quad (4)$$

k_{gp} (s^{-1}) is the first-order gas-particle mass transfer coefficient or otherwise termed as a condensation sink, which we approximate by assuming a monodisperse aerosol population:

$$k_{gp}(\beta) = 2\pi D_g D_p N_p \beta \quad (5)$$

$$\beta(Kn, \alpha_{\text{eff}}) = \frac{0.75\alpha_{\text{eff}}(1 + Kn)}{Kn^2 + Kn + 0.283Kn\alpha_{\text{eff}} + 0.75\alpha_{\text{eff}}} \quad (6)$$

Where D_g ($\text{cm}^2 \text{s}^{-1}$) is the gas-phase diffusivity and D_p (cm) is the particle diameter. The term β is the Fuchs-Sutugin correction factor for gas diffusion in the transition regime, which is a function of the Knudsen number (Kn) and the effective mass accommodation coefficient (α_{eff}). Kn is the ratio of the mean free path in the gas phase (λ) as calculated using the mean thermal velocity (ω), and D_p : $Kn = 2\lambda / D_p = 6 D_g / (\omega D_p)$ (Pöschl et al., 2007). We assume a monodisperse particle population initialized at $D_p = 100 \text{ nm}$ with particle growth/loss considered only through condensation/evaporation as the number concentration is held constant while C_{OA} may vary.



109 α_{eff} describes the probability of a gas molecule colliding with a particle surface to enter the
110 condensed phase, considering potential kinetic limitation of bulk diffusion in the particle phase as
111 described as follows (Shiraiwa & Pöschl, 2021):

$$112 \quad \alpha_{\text{eff}} = \alpha_s \frac{1}{1 + \frac{\alpha_s \omega C^0 D_p}{4 D_b \rho_p} \frac{1}{10} \times 10^{-12}} \quad (7)$$

113 where α_s is the surface accommodation coefficient, which is assumed to be 1 (Julin et al., 2014),
114 and ρ_p (g cm^{-3}) is the particle density. Note that most studies assume certain fixed values for mass
115 accommodation coefficient (often 1, 0.5, or 0.1) without considering the effect of particle phase
116 state (Zaveri et al., 2020). α_{eff} provides an efficient way of accounting for the influence of species
117 volatility and bulk diffusivity, D_b ($\text{cm}^2 \text{s}^{-1}$) on SOA partitioning (Shiraiwa & Pöschl, 2021). The
118 conversion factor 10^{-12} allows volatility to be inserted in the commonly used $\mu\text{g m}^{-3}$ units. The
119 application of α_{eff} in traditional mass transfer equations yielded consistent simulation results with
120 a kinetic multilayer model (KM-GAP) and two-film model (MOSAIC) (Zaveri et al., 2014) and
121 α_{eff} successfully explained experimentally measured accommodation coefficient in SOA
122 partitioning (Liu et al., 2019; Shiraiwa & Pöschl, 2021).

123 To solve the stiff ODE system of Eqs. (1) and (2), we use the Livermore Solver for Ordinary
124 Differential Equations (LSODE) in double precision (DLSODE) (Hindmarsh, 1983). The use of
125 an ODE solver can be computationally expensive especially when implemented in a CTM;
126 therefore, we also treat Eq. (1) with an exponential approximation accompanied with the
127 assumption that the equilibrium concentration is known *a priori*. The derivation of this
128 approximation is as follows: at equilibrium, the net mass transfer rate between the gas and particle
129 phase is zero.

$$130 \quad \frac{dC^g}{dt} = 0 \quad (9)$$

131 Substituting Eq. (1) into Eq. (9) gives the following equation knowing that the condensation sink
132 cannot be zero.

$$133 \quad \frac{C^p}{C_{\text{OA}}} C^0 = C_{\text{eq}}^g \quad (10)$$



At equilibrium, the gas-phase concentration would be the product of the particle-phase mass fraction of a species and its volatility. Substitution of Eq. (10) into Eq. (1) gives:

$$\frac{dC^g}{dt} = -k_{gp} (C^g - C_{eq}^g) \quad (11)$$

Eq. (11) can be solved analytically with integration by separation of variables to obtain the following analytical solution of the gas-phase concentration at time, t , when initial and equilibrium concentrations are known as well as the condensation sink of that species.

$$C^g(t) = C_0^g - (C_0^g - C_{eq}^g)(1 - e^{-k_{gp}t}) \quad (12)$$

We conducted model simulations of SOA partitioning using a standalone zero-dimensional box model to calculate and compare two timescales of SOA partitioning, equilibration and mixing timescales, under various environmental conditions and mass loadings. The e -folding timescale to achieve gas-particle equilibrium or otherwise known as the equilibration timescale, τ_{eq} is calculated for the first instance (t) when the following inequality is satisfied (Li & Shiraiwa, 2019):

$$\frac{|C_p(t) - C_{p,eq}|}{|C_{p,0} - C_{p,eq}|} < \frac{1}{e} \quad (13)$$

$C_{p,0}$ and $C_{p,eq}$ refer to the initial and equilibrium mass concentration of a diffusing compound in the particle phase, respectively. τ_{eq} of a partitioning species also corresponds to approximately the inverse of its k_{gp} , given that k_{gp} is held constant. We simulated partitioning under a closed system in which condensation of a species depletes its gas-phase mass concentration (C_g) while the particle-phase mass concentration increases (C_p) whereas the opposite occurs under evaporation conditions. A closed system configuration is best suited for integration into the GEOS-Chem partitioning module as partitioning occurs within a chemical time step where the gas phase is not replenished.

For the evaluation of kinetic limitations of bulk diffusion in a single particle, the characteristic mixing timescale, τ_{mix} , can be calculated as (Seinfeld & Pandis, 2016; Shiraiwa et al., 2011):

$$\tau_{mix} = \frac{D_p^2}{4\pi^2 D_b} \quad (14)$$



159 where D_p is particle diameter which is assumed to be 100 nm and τ_{mix} is the timescale for a diffusing
160 molecule to diffuse and establish a uniform concentration within the particle (Li & Shiraiwa, 2019).
161 Note that τ_{eq} is fundamentally different from τ_{mix} , as a homogeneously mixed bulk phase is not a
162 necessary requirement to achieve gas-particle partitioning equilibrium. Additionally, while τ_{mix} is
163 solely determined by D_b and particle diameter, τ_{eq} is also affected by gas diffusion, volatility and
164 absorptive mass concentration (Li & Shiraiwa, 2019; Maclean et al., 2017; Schervish & Shiraiwa,
165 2023; Shiraiwa et al., 2012). Comparison of the two timescales with regards to modeling
166 partitioning is explored in Sect. 3.1.

167 The box model was designed to emulate all the relevant processes used in the standard
168 GEOS-Chem complex scheme partitioning. Temperature, $C^0(298\text{K})$, particle D_b , and initial
169 concentrations of the gas- and particle-phase species are all prescribed for the box model
170 simulations. A non-volatile particle-phase species is used to establish initial background aerosol
171 C_{OA} . The volatilities of all species are temperature-corrected using Eq. (3) and (4).

172 2.2 GEOS-Chem Global Model Set up

173 We used the global chemical transport model, GEOS-Chem (ver. 14.1.1) to simulate
174 organic aerosols in the troposphere (Community, 2024b; Lin et al., 2024). The model system was
175 used to simulate the atmospheric composition in the year 2020 with a horizontal grid resolution of
176 $4^\circ \times 5^\circ$ and 47 vertical layers extending to about 80 km altitude (0.1 hPa) using terrain-following
177 (sigma) coordinates near the surface. The model spin-up was six months, after which the following
178 12-month outputs were recorded and analyzed. The recommended timesteps for this model
179 resolution are 10 min for transport and 20 min for chemistry. Details on the meteorology and
180 emission inventories used are the same as in our previous study, Luu et al., 2025.

181 The full chemistry mechanism with the complex SOA scheme including semi-volatile POA
182 was applied. In this scheme, all SOA components are represented by a volatility-basis set (VBS)
183 based on Pye et al. (2010)(Pye et al., 2010) apart from isoprene SOA (ISOA) and organo-nitrate
184 SOA, which are modeled through a reactive uptake scheme (Fisher et al., 2016; Marais et al., 2016).
185 Thus, ISOA, excluded from the VBS bins, are also excluded from contributing to the absorbing
186 SOA mass for partitioning purposes, while we do include ISOA for particle phase state calculations.
187 D_b is calculated based on assigned glass transition temperature (T_g) values, temperature, relative
188 humidity, and particle mass concentrations. Assigned T_g values were based on our previous study



(Luu et al., 2025). Briefly, T_g of SOA component in each volatility bin ($T_{g,i}$) is calculated based on each component's volatility and oxygen-to-carbon (O:C) ratio (Table S1) (Li et al., 2020):

$$T_{g,i} = 289.10 - 16.50 \times \log_{10}(C^0) - 0.29 \times [\log_{10}(C^0)]^2 + 3.23 \times \log_{10}(C^0) (O:C) \quad (8)$$

T_g of dry SOA is calculated considering mass fraction of each component and T_g of SOA at ambient RH is calculated considering water uptake using the Gordon-Taylor equation. Using the modified Vogel-Fulcher-Tammann (VFT) equation, T_g is converted to particle viscosity, which is then converted to D_b with the fractional Stokes-Einstein equation. The detailed steps including key equations and parameters can be found in our previous studies (Li et al., 2020; Luu et al., 2025; Shiraiwa et al., 2017).

For the current work, only reversibly partitioning species are modified which includes the oxidation products of monoterpenes and sesquiterpenes (TSOA) and oxidation products of light aromatics (ASOA) and intermediate volatility organic compounds (IVOCs). The VBS for TSOA covers C^0 of 0.1, 1, 10, and 100 $\mu\text{g m}^{-3}$, while ASOA uses C^0 of 10^{-10} , 1, 10, and 100 $\mu\text{g m}^{-3}$ (Community, 2024a). In addition, we updated the heat of vaporization, previously assumed as 42 kJ mol^{-1} for all organic species, to be temperature and volatility dependent using Eq. (4) (Pye et al., 2010; Pye & Seinfeld, 2010). To maintain consistency with our previous methodology (Luu et al., 2025), we adjusted the molecular weights of all SOA components based on molecular corridors (Shiraiwa et al., 2014), as shown in Table S1.

While the aforementioned kinetic theory behind gas-particle partitioning has been applied in process-level modeling, large-scale transport models such as GEOS-Chem primarily rely on a thermodynamic partitioning method known as absorptive partitioning theory (Chung & Seinfeld, 2002). An alternative configuration for GEOS-Chem coupled with an online aerosol microphysics module (TOMAS) does not utilize thermodynamic partitioning as SOA is treated as non-volatile. GEOS-Chem-TOMAS, however, portions SOA based on effective surface area with the $\alpha = 1$ assumption (D'Andrea et al., 2013; Pierce & Adams, 2009). The details of absorptive partitioning method solving for the equilibrium gas and particle concentration values via a root finding algorithm is further explained in Chung & Seinfeld, (2002). This necessitates that chemical equilibrium is reached within the model's chemical time step (Kanakidou et al., 2005). This assumption is, however, not always true, especially when the particle phase state is highly viscous.



218 Thus, we use α_{eff} to account for these kinetic limitations to SOA partitioning with an assumption
219 of particle diameter of 100 nm.

220 Composition-dependent D_b is used to determine which algorithm the gas-particle
221 partitioning proceeds through: (1) default GEOS-Chem instantaneous equilibrium or the (2)
222 analytical solution as elaborated on in Sect. 3.1. If the analytical solution is selected, the
223 condensation sink is calculated as shown in Eq. (5, 6) using α_{eff} and Eq. (12) is used to determine
224 final gas and particle concentrations at the end of the chemical time step.

225 3 Results and Discussion

226 We first show the box model results of six representative cases to illustrate how different
227 partitioning treatments of the condensing species affect the time evolution of its gas-phase and
228 particle-phase concentrations. Then we show GEOS-Chem results on the impact of the updated
229 partitioning scheme on simulated global SOA budgets.

230 3.1 Zero-dimensional Box Model Simulations

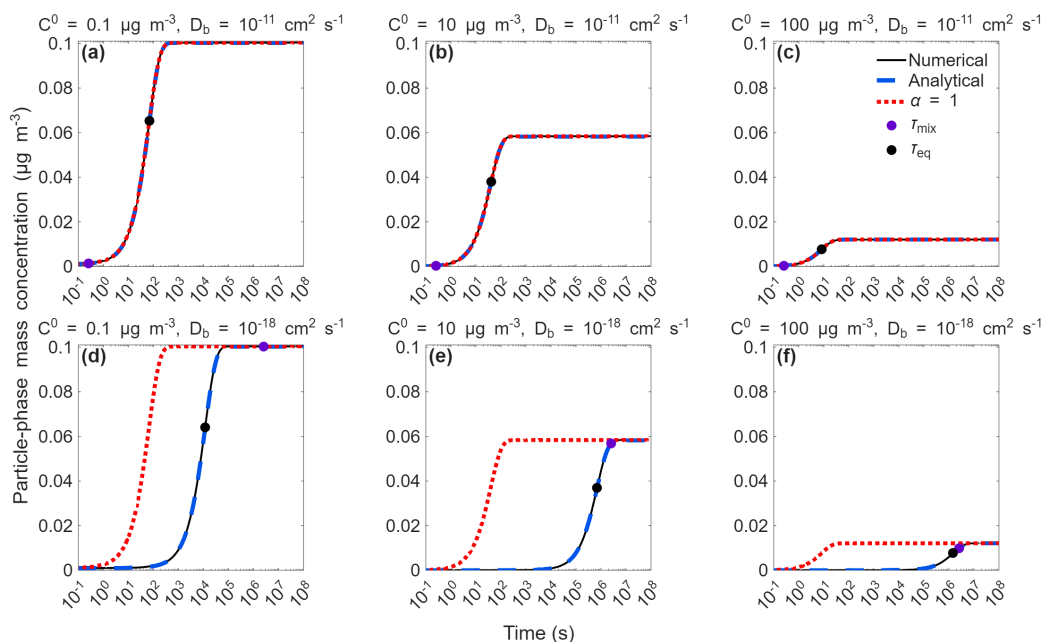
231 Figure 1 shows the exemplary simulation results for the partitioning of a semi-volatile
232 species of either a $C^0 = 0.1 \mu\text{g m}^{-3}$ (a, d), $10 \mu\text{g m}^{-3}$ (b, e), or $100 \mu\text{g m}^{-3}$ (c, f). The initial seed
233 concentration is assumed to be $10 \mu\text{g m}^{-3}$ with a number concentration of $1 \times 10^4 \text{ cm}^{-3}$,
234 representative of polluted conditions, and the initial particle diameter is 100 nm. Initial mass
235 concentrations of semi-volatile species in the gas and particle phases are set to 0.1 and $0 \mu\text{g m}^{-3}$,
236 respectively. Two D_b cases are tested for the particle phase: $D_b = 10^{-11} \text{ cm}^2 \text{ s}^{-1}$ (a, b, c) or 10^{-18} cm^2
237 s^{-1} in (d, e, f). Per case, three simulations are shown where the treatment of partitioning differs: (1)
238 by solving the system of ODEs numerically, Eq. (1 and 2) with α_{eff} , (2) by solving the analytical
239 solution, Eq. (12), or (3) same as method (1) except with the assumption of perfect mass
240 accommodation ($\alpha = 1$). An internal time stepping loop is used to iteratively recalculate the
241 condensation sink for both the numerical and analytical solutions.

242 For partitioning into liquid particles (Fig. 1a-c), particle-phase concentrations reach
243 equilibrium concentrations rapidly with τ_{eq} of $\sim 10^2 \text{ s}$, equivalent to the condensation sink timescale
244 set by gas-phase diffusion, indicating little kinetic limitations of bulk diffusion and a well-mixed
245 particle bulk with τ_{mix} of $\sim 0.03 \text{ s}$. Note that τ_{eq} is longer than τ_{mix} as the partitioning is governed by
246 gas-phase diffusion and the condensation sink rather than D_b . The numerical solutions whether



247 using α_{eff} (black lines) or $\alpha = 1$ (red dotted lines) give identical results as α_{eff} is almost 1 with high
 248 D_b . The analytical solution (blue dashed lines) also yields practically identical results. Figure 1 d-
 249 f present simulations for highly viscous particles ($D_b = 10^{-18} \text{ cm}^2 \text{ s}^{-1}$) with τ_{mix} of $\sim 2 \times 10^6 \text{ s}$, showing
 250 that partitioning proceeds with much longer equilibration timescales ($\tau_{\text{eq}} > 10^4 \text{ s}$) due to strong
 251 kinetic limitations of bulk diffusion. The simulations with $\alpha = 1$ (red dotted lines) overestimates
 252 partitioning rates substantially. For the partitioning of low volatility species with $C^0 = 0.1 \mu\text{g m}^{-3}$
 253 (Fig. 1d), τ_{eq} is shorter at $\sim 10^4 \text{ s}$ as partitioning occurs relatively quicker due to low volatility nature
 254 of the condensing species rapidly establishing a local equilibrium between the gas phase and the
 255 near-surface bulk (Li & Shiraiwa, 2019). With higher volatility ($C^0 = 10$ and $100 \mu\text{g m}^{-3}$ in Fig. 1e
 256 and f, respectively), τ_{eq} becomes longer as re-evaporation becomes relevant due to their semi-
 257 volatile nature. However, in this case τ_{mix} overestimates τ_{eq} as partitioning is finished before a well-
 258 mixed bulk is established; the use of τ_{mix} to represent partitioning timescale would lead to
 259 significant underestimation of SOA mass. This further exemplifies how the two timescales cannot
 260 be interchangeable as they are physically distinct. For all cases represented in Fig. 1, both the
 261 numerical and analytical solutions are identical in simulating particle growth. Overall, τ_{eq} can be
 262 longer than the CTM chemical timestep of 20 min under a multitude of ambient conditions,
 263 indicating that proper consideration of kinetic limitations of bulk diffusion is required.

264



265

266 **Figure 1.** Temporal evolution of partitioning of semi-volatile species with $C^0 = 0.1 \mu\text{g m}^{-3}$ (a, d),
 267 $C^0 = 10 \mu\text{g m}^{-3}$ (b, e), or $C^0 = 100 \mu\text{g m}^{-3}$ (c, f) into seed particles with initial concentrations of 10
 268 $\mu\text{g m}^{-3}$ with $D_b = 10^{-11} \text{ cm}^2 \text{ s}^{-1}$ in (a-c) or $D_b = 10^{-18} \text{ cm}^2 \text{ s}^{-1}$ in (d-f). The temperature was set to
 269 298 K, and the initial particle diameter was 100 nm. τ_{mix} and τ_{eq} are marked by purple and green
 270 dots, respectively.

271 To further probe the parameter space for kinetic partitioning, we computed contour plots
 272 of τ_{eq} as a function of D_b , C^0 , and C_{OA} of $0.1 \mu\text{g m}^{-3}$, $1 \mu\text{g m}^{-3}$ and $10 \mu\text{g m}^{-3}$. Figure 2 shows the
 273 contour plots of τ_{eq} as calculated using the numerical solution. Note that the simulations were also
 274 conducted for evaporation, yielding very similar timescales and the results of Fig. 2 are directly
 275 applicable to evaporation (Shiraiwa & Seinfeld, 2012; Shiraiwa et al., 2013). In the low bulk
 276 diffusivity ($D_b < 10^{-15} \text{ cm}^2 \text{ s}^{-1}$) regime, τ_{eq} becomes prolonged with lower D_b due to kinetic
 277 limitations of bulk diffusion, while τ_{eq} shortens for lower C^0 as partitioning becomes largely
 278 determined by the condensation sink as there is efficient condensation with very slow re-
 279 evaporation of lower volatility compounds (Li & Shiraiwa, 2019). τ_{eq} is shorter for higher C_{OA} due
 280 to higher condensation sink, while this dependence is weaker compared to D_b and C^0 . With an
 281 increased mass loading, the effect of bulk diffusivity on τ_{eq} is slightly lessened by the larger
 282 condensation sink. Figure S1 shows the contour plots of τ_{eq} calculated using the analytical solution



283 to partitioning and reproduces very similar predicted τ_{eq} as the numerical solution in Fig. 2. Using
 284 the average case of $C_{OA} \sim 1 \mu\text{g m}^{-3}$, τ_{eq} is shorter than or equal to 20 min for all C^0 when $D_b \geq 10^{-15}$
 285 $\text{cm}^2 \text{s}^{-1}$; therefore, we can apply the instantaneous equilibrium solution for particles with little
 286 bulk diffusion limitations (demarked in green) (Fig. 3b). In this equilibrium regime, τ_{eq} exhibits
 287 sensitivity to C^0 where lower C^0 corresponds to a prolonged τ_{eq} as more mass transfer must occur
 288 to reach equilibrium conditions. In more pristine conditions where the background aerosol
 289 concentrations are lower, *i.e.* $C_{OA} < 1 \mu\text{g m}^{-3}$, and τ_{eq} is greater than 20 minutes when $D_b > 10^{-15}$
 290 $\text{cm}^2 \text{s}^{-1}$ then there may be slight overestimation in the partitioned mass however this is not expected
 291 to be large as the aerosol burden in these areas would be minimal.

292 When particles are highly viscous with $D_b < 10^{-15} \text{cm}^2 \text{s}^{-1}$, kinetic partitioning needs to be
 293 applied. The analytical solution (Fig. S1) is an adequate representation of partitioning behavior as
 294 benchmarked by the numerical solution (Fig. 2b); thus, partitioning can be approximated by Eq.
 295 (12) where the instantaneous analytical solution can be applied at $D_b < 10^{-15} \text{cm}^2 \text{s}^{-1}$. For integration
 296 into GEOS-Chem, additional internal time stepping for partitioning may be undesirable as only
 297 the partitioned mass is needed at the end of the chemical time step. Figure S2 shows a contour plot
 298 of the final ($t = 20$ min) particle-phase concentration which is calculated through two approaches:
 299 (1) the numerical and analytical solutions with internal time stepping and (2) the instantaneous
 300 analytical solution without time stepping and without iteratively recalculating C_{OA} and k_{gp} . For the
 301 three tested mass loadings, when $D_b < 10^{-15} \text{cm}^2 \text{s}^{-1}$ the instantaneous analytical solution exhibits a
 302 linear trend that coincides with the linear portion of the numerical (and analytical) solution contour.
 303 Thus, when $D_b < 10^{-15} \text{cm}^2 \text{s}^{-1}$, the instantaneous analytical solution can be used to accurately
 304 simulate the final particle-phase (and gas-phase) concentration at 20 minutes.

305 With the assumption that the average background aerosol concentration is within the
 306 magnitude of $C_{OA} = 1 \mu\text{g m}^{-3}$, we have implemented the algorithm as summarized: (1) if $D_b \geq 10^{-15}$
 307 $\text{cm}^2 \text{s}^{-1}$, equilibrium partitioning is applied or (2) if $D_b < 10^{-15} \text{cm}^2 \text{s}^{-1}$, then the instantaneous
 308 analytical solution is applied. We term this kinetic partitioning implementation as the “ α_{eff} scheme”
 309 to replace the instantaneous equilibrium gas-particle partitioning scheme (“base” case) in GEOS-
 310 Chem.

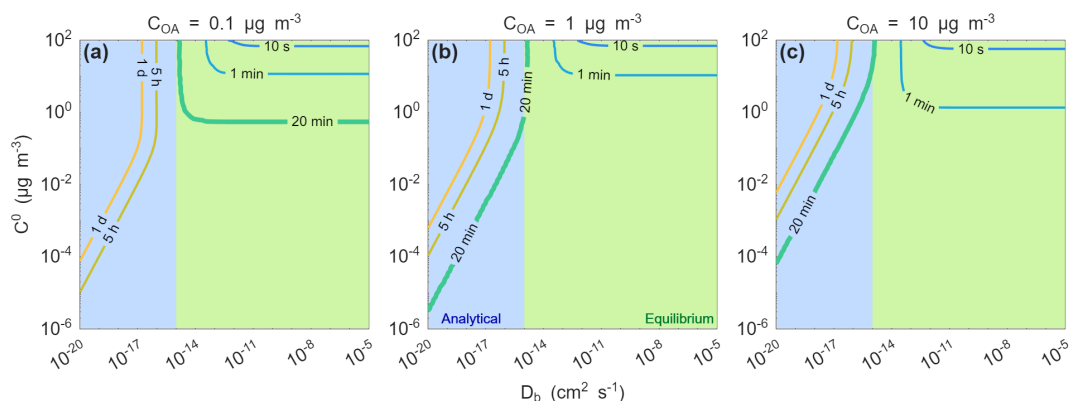


Figure 2. Contour plot of equilibration timescale (τ_{eq}) calculated using the numerical solution as a function of bulk diffusivity (D_b), saturation mass concentration (C^0), and initial pre-existing particle concentrations (C_{OA}) of (a) $0.1 \mu\text{g m}^{-3}$, (b) $1 \mu\text{g m}^{-3}$, and (c) $10 \mu\text{g m}^{-3}$. The τ_{eq} is calculated using the numerical solution. Two colored regimes are depicted: (1) when $D_b \geq 10^{-15} \text{ cm}^2 \text{ s}^{-1}$, equilibrium partitioning is applied (in green) and (2) if $D_b < 10^{-15} \text{ cm}^2 \text{ s}^{-1}$, then the instantaneous analytical solution is applicable (in blue). The initial gas-phase concentration was $0.1 \mu\text{g m}^{-3}$ (to simulate condensation), or the initial particle-phase concentration was $0.1 \mu\text{g m}^{-3}$ (to simulate evaporation), and the temperature was set to 298K.

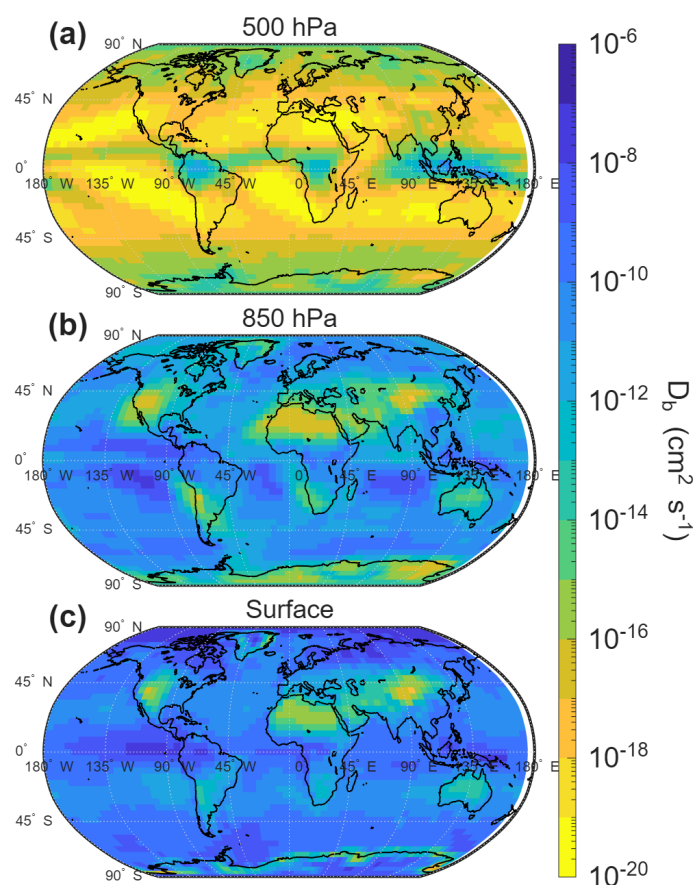
3.2 Global SOA Budget Simulations

In this section, we explore the effects of kinetic partitioning on simulated global SOA budgets. We implemented the α_{eff} kinetic partitioning scheme to simulate the two sets of TSOA and ASOA species. As we have incorporated the volatility-dependent ΔH_{vap} parameterization into our methods, to accurately visualize the change due to purely partitioning scheme differences, the “base” simulation includes effects due to the change in ΔH_{vap} calculation. This is because the assumption of ΔH_{vap} in combination with the initial gas-phase concentration modulates the potential aerosol formation (or evaporation) upon cooling (heating) (Fig. S3) (Cappa & Jimenez, 2010). Lower volatility species, *e.g.* $C^0 = 0.1 \mu\text{g m}^{-3}$, will have a higher heat of vaporization corresponding to a stronger temperature dependence therefore cooling enhances their preference for the particle phase compared to higher volatility compounds. Overall, the update to volatility-



332 dependent ΔH_{vap} resulted in global SOA burden increase driven by increased SOA concentrations
333 especially in lower temperatures (Fig. S3).

334 Figure 3 presents the annually averaged D_b in SOA particles (using the α_{eff} scheme in
335 GEOS-Chem) from the surface (vertical level 1), and two elevated hybrid-coordinate model
336 altitudes (vertical levels 11 and 23). GEOS-Chem's hybrid sigma-pressure grid is terrain-following
337 rather than being constant-pressure surfaces or fixed altitudes. Vertical levels 11 and 23 were
338 selected as to representative tropospheric model levels which correspond to 850 hPa (above the
339 boundary layer) and 500 hPa (in the free troposphere) for a column at atmospheric sea level
340 (GEOS-Chem, 2026). Primarily at the surface level, SOA particles have typically $D_b \geq 10^{-10} \text{ cm}^2$
341 s^{-1} , indicating liquid-like particles with little kinetic limitations to partitioning. Western US,
342 northern Africa, and the Tibetan Plateau are predicted to have slightly lower D_b representative of
343 semisolid particles. Above the standard 850 hPa layer in these regions, lower D_b becomes more
344 predominant at higher altitudes. At 500 hPa, D_b is expected to be mostly smaller than $\sim 10^{-15} \text{ cm}^2$
345 s^{-1} in highly viscous particles apart from equatorial continents such as northern South America,
346 sub-Saharan Africa, and Southeast Asia where SOA particles are less viscous with higher D_b . The
347 spatial distribution of D_b agrees well with previous phase state predictions for large-scale models
348 (Luu et al., 2025; Shiraiwa et al., 2017).



349

350 **Figure 3.** Simulated annually averaged bulk diffusivities (D_b) in SOA particles at selected
 351 representative tropospheric model levels which correspond to (a) 500 hPa and (b) 850 hPa for a
 352 column at atmospheric sea level. While (c) at the surface, corresponds to the true fixed terrain-
 353 following pressure surface.

354

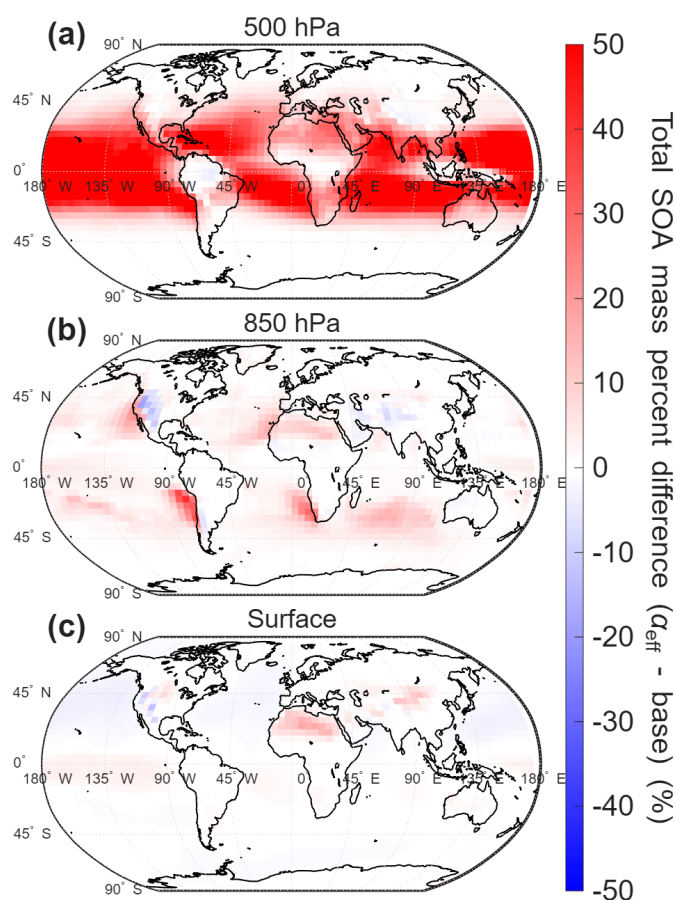
355 Figure 4 shows the percent difference between annually predicted SOA budget with the α_{eff}
 356 kinetic partitioning scheme (see actual SOA mass loadings in Fig. S4) compared to the base case
 357 (Fig. S5). At the surface level, both schemes predict very similar concentrations, showing little-to-
 358 no changes in most regions. The α_{eff} scheme predicts up to 10% less mass over the western and
 359 central US due to kinetically limited condensational growth on viscous particles (Fig. 3c). Northern
 360 Africa and Tibetan Plateau are predicted to have roughly 20% higher SOA mass with the α_{eff}



361 scheme due to kinetic limitations to evaporation as low C_{OA} and high temperature conditions drive
362 net evaporation in these regions. At 850 hPa, western US is predicted to have up to 10% less SOA
363 mass predicted by the α_{eff} scheme, while offshore is predicted to have roughly 20% more SOA
364 mass. Other regions such as off the coasts of Northwestern Africa, South Africa, and Chile are also
365 predicted to have increased SOA mass up to 40%.

366 At 500 hPa, the effects of kinetic limitations on partitioning become significant, reaching
367 upwards of 50% greater predicted SOA mass along the equatorial and mid-latitudes as evaporation
368 is suppressed for glassy particles. Over the remote oceans (Indian and Pacific Oceans) and central
369 US, we predict up to an order of magnitude SOA budget increase in the upper troposphere (Fig.
370 S4a and S5a). This occurrence may be driven by the Hadley cell in the region that lifts SOA and
371 its precursors to the tropical tropopause, particularly in the tropical rainforest regions in the
372 Amazon, Congo, and Oceania. These continental detainment regions are more humid than
373 elsewhere in the tropical upper troposphere and hence have higher D_b values (Fig. 3a) allowing
374 SOA formation in cold temperatures. As the tropical tropopause air undergoes subsidence in the
375 Hadley cell, the temperature increases, which drives evaporation; however, the RH simultaneously
376 decreases, which lowers D_b more than the increasing temperature increases D_b (Fig. S6). Hence,
377 the descending regions of the Hadley cell in the subtropics have $D_b < 10^{-16} \text{ cm}^2 \text{ s}^{-1}$. While the
378 warming temperatures drive evaporation in the base simulations, the low D_b values greatly delay
379 evaporation with the α_{eff} scheme, leading to higher concentrations in this simulation. This is
380 expected as solid particles are more prevalent under free troposphere conditions (low average
381 temperatures and low C_{OA}) throughout the year (Fig. 3 and S4). During each season at 500 hPa,
382 the predicted SOA mass difference trend prevails because of kinetic limitations on evaporation
383 when $D_b < 10^{-16} \text{ cm}^2 \text{ s}^{-1}$ (Fig. S7). The suppression of evaporation is more prominent for higher
384 volatility compounds with $C^*(298 \text{ K}) = 10$ and $100 \mu\text{g m}^{-3}$ with up to 50% higher mass predicted
385 with the α_{eff} scheme for a majority of the mid- to equatorial latitudes (Fig. S8). The annual lower
386 troposphere averaging approximately 1% increase while the upper troposphere is about 15%
387 increase in the SOA budget (Table S2).

388



389

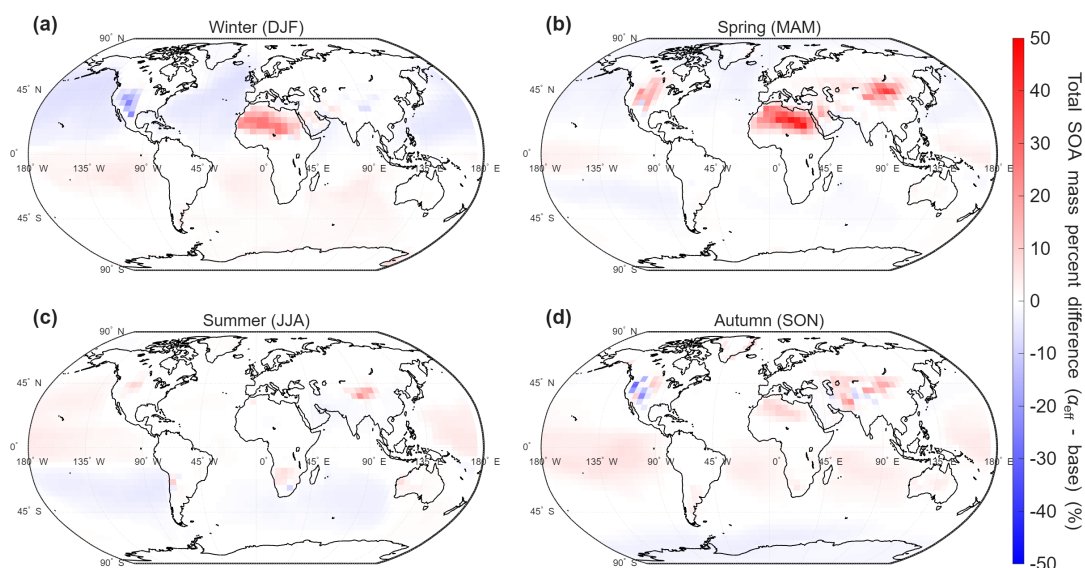
390 **Figure 4.** Predicted annual percent SOA mass loading differences **(a)** at the free troposphere, **(b)**
391 above the boundary layer, and **(c)** at the surface between the α_{eff} kinetic and base case partitioning
392 scheme simulations.

393

394 We show the seasonality of the SOA budget percent differences at the surface in Figure 5.
395 In Northern hemisphere winter, the α_{eff} kinetic partitioning scheme predicts roughly a 40%
396 decrease in SOA mass in the Western US but an approximately 30% increase in Saharan Africa
397 (Fig. 5a). As for spring, there is a mixture of an increase in predicted SOA mass in Northern Great
398 Plains of the US extending into Canada with a small decrease for Southwestern US (Fig. 5b). There
399 is also up to a 50% increase in SOA budget in Saharan Africa and as much as 40% in the Tibetan
400 Plateau. With increased average surface temperatures in the summer months, there are virtually no



401 differences in predicted SOA mass apart from the Tibetan Plateau as the equilibrium partitioning
402 assumption is valid when $D_b > 10^{-15} \text{ cm}^2 \text{ s}^{-1}$ (Fig. S9c). The following autumn months show a
403 mixture of enhancement and reduction in the SOA budget in western US and a slight increase
404 across Saharan Africa, Central Asia, and the Tibetan Plateau. From these plots, we observe that
405 each season demonstrates distinct spatial patterns of SOA budget discrepancies when kinetic
406 limitations are applied to the partitioning scheme. It is projected that on a finer temporal resolution
407 on the span of weeks or days, typically simulated for field campaign comparisons, weather
408 variability may make these predicted differences more significant, yet this remains to be further
409 investigated.



410

411 **Figure 5.** Predicted seasonal variations in percent SOA budget differences at the surface, during
412 Northern hemisphere (a) winter (Dec-Jan-Feb), (b) spring (Mar-Apr-May), (c) summer (Jun-Jul-
413 Aug), and (d) autumn (Sep-Oct-Nov).

414

415 **4 Method Limitations and Future Needs**

416 We address limitations in our methods and future needs. In the complex scheme,
417 oxygenated primary organic aerosol (OPOA) is formed through OH oxidation of emitted primary
418 organic gases (EPOG), producing oxidized primary organic gases (OPOG) that reversibly partition



419 into OPOA (Pai et al., 2020). Therefore, previous studies aimed at quantifying simulated OA
420 burden have categorized OPOA differently either as POA (Pai et al., 2020; Pye & Seinfeld, 2010)
421 or SOA (Brewer et al., 2023; Murphy et al., 2014). In this study, we have chosen to omit OPOA
422 from our main SOA phase state calculations, as only two volatility bins each are used to represent
423 both EPOG and their oxidation products, OPOA, which may not be sufficient constraints
424 especially for phase state predictions which have been validated for SOA (DeRieux et al., 2018;
425 Shiraiwa et al., 2014). Both POA and OPOA are considered for total organic aerosol mass
426 concentration (C_{OA}) as the absorbing mass for partitioning as in the base scheme. Additionally,
427 potential phase state impact on isoprene-derived SOA (ISOA) is not resolved in the present work.
428 ISOA is not treated as a reversibly partitioning species but treated as non-reversible reactive uptake
429 in GEOS-Chem. Further development is needed to include kinetic limitations to the reactive uptake
430 of ISOA.

431 In the current GEOS-Chem configuration, SOA particles are implicitly assumed to be
432 externally mixed from inorganic compounds or phase-separated, if internally mixed, with the
433 predicted phase state representing only the organic phase. Phase separation could occur between
434 SOA and POA (Huang et al., 2021), but they are treated as ideally well-mixed as both SOA and
435 POA are included for the total absorbing mass (C_{OA}). We partially account for phase separation in
436 the present work, by determining using SOA phase state to determine absorptive partitioning
437 dynamics of all species. However, the effect of non-ideality determining phase separation or well-
438 mixed bulk phase on SOA partitioning should be implemented in future studies, as has been
439 explored in several studies (Schmedding et al., 2020; Serrano Damha et al., 2025). Additionally,
440 SOA evaporation and heterogeneous chemistry can lead to the formation of a surface crust that
441 may hinder further evaporation (Schervish et al., 2026; Zhou et al., 2019). Such complex effects
442 should be further explored for potential treatments in CTMs in future studies.

443 α_{eff} is particle size dependent as α_{eff} is inversely related to D_p , whereas we assumed $D_p =$
444 100 nm in this study. The lack of particle size resolution with the GEOS-Chem complex scheme
445 limits our results to one particle size, but GEOS-Chem-TOMAS' size tracking capabilities presents
446 a potential for further investigation on the size-dependent kinetic limitations effects on the SOA
447 budget. Note that recently a size-dependent α_{eff} has been implemented in GEOS-Chem-TOMAS
448 to portion the SOA mass accordingly across size distribution, without altering the SOA budget



449 (O'Donnell et al., 2025). This development, though not impacting the total partitioning mass,
450 should be included in further advancements in implementation of kinetic partitioning in GEOS-
451 Chem-TOMAS.

452

453 **5 Conclusions**

454 In summary, we developed a computationally efficient method to simulate SOA phase state
455 and its impacts on kinetic partitioning utilizing the effective mass accommodation coefficient in
456 GEOS-Chem. This method effectively treats the effect of gas diffusion (i.e. condensation sink) as
457 well as phase state and associated kinetic limitations on partitioning. We note that τ_{mix} represents
458 kinetic limitations of bulk diffusion within SOA particles; however, τ_{mix} is not an accurate
459 representation to describe SOA partitioning in aerosol modeling. A recent study by Li et al. (2026)
460 implemented a phase-state dependent partitioning scheme using τ_{mix} as partitioning timescale;
461 however, the use of τ_{mix} as an equilibration timescale of partitioning is fundamentally incorrect,
462 leading to inaccurate representation of phase state impacts on SOA, as it ignores the effects of gas
463 diffusion, volatility, and mass loadings on partitioning timescale, as discussed in detail in the box
464 modeling section.

465 With implementation of α_{eff} -kinetic partitioning scheme in GEOS-Chem, we predict that
466 SOA particles adopt viscous semisolid and glassy phase states in arid regions below the boundary
467 layer and is the dominant phase state in the upper troposphere. We simulate that the kinetic
468 limitations of SOA partitioning can lead to annual SOA mass deviation up to 50% for mid- to
469 equatorial latitudes compared to instantaneous equilibrium partitioning above 500 hPa. Overall,
470 we predict annually generally little change at the surface and at 850 hPa but significant perturbation
471 in predicted SOA mass in dry regions and in the free troposphere. We anticipate stronger deviations
472 for finer temporal resolution simulations as we expect a reduced dilution effect on the mass
473 difference between the α_{eff} and base case with shorter time averaging. The development of this α_{eff}
474 kinetic partitioning scheme enables comparison with aircraft observations to quantify impacts of
475 SOA phase state in future studies. Observations of SOA mass loading in the upper troposphere
476 would be especially compelling to investigate as predictions of high-altitude new particle
477 formation may be improved from implementing the α_{eff} scheme for SOA partitioning. Addressing



478 this well-known knowledge gap may help achieve closure between SOA model simulations and
479 field observations.

480

481 **Code and Data Availability.** The GEOS-Chem model is maintained by the Harvard Atmospheric
482 Chemistry Modelling Group and the specific version (ver. 14.1.1) used in this work is permanently
483 archived at <https://doi.org/10.5281/zenodo.7696632> (Yantosca et al., 2023). The modified
484 carbon_mod.f90 used in this work and the 0-D box model are archived at
485 <https://doi.org/10.5281/zenodo.22099483> (Luu et al., 2026). GEOS-Chem simulation outputs are
486 available by contacting the corresponding author.

487 **Author Contributions.** RL, MSc, JP and MSh designed the research. RL developed the box model,
488 GEOS-Chem implementation, and ran the simulations. RL and MSc conducted data analysis and
489 all authors were involved in discussions. RL drafted the manuscript, MSh edited the manuscript,
490 and all authors reviewed and commented on the paper.

491 **Supplement.** The supplement related to this article is available online.

492 **Competing interests.** We declare no competing interests.

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495

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