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23 additional mangrove sites.

24 **Keywords:** Mangrove forest monitoring; Eddy covariance; Apparent FAPAR; Tidal
25 inundation; Temporal upscaling

26 **1. Introduction**

27 Mangrove forests combine high carbon-storage capacity with direct exposure to
28 tidal forcing (Rani et al., 2023). Reliable monitoring of their canopy function is
29 consequently important for forest conservation, restoration assessment, and blue-
30 carbon accounting (Friess et al., 2019). Satellite gross primary productivity (GPP)
31 products provide regional coverage, but their accuracy depends on the fraction of
32 photosynthetically active radiation absorbed by vegetation (FAPAR). A recent global
33 synthesis for 1996–2020 identified FAPAR as a major predictor of mangrove GPP (Sun
34 et al., 2024), while multi-year eddy-covariance observations show that meteorological
35 and tidal drivers jointly regulate mangrove carbon exchange (Gou et al., 2023). An
36 unresolved monitoring problem is therefore whether an observation acquired at a fixed
37 satellite overpass time represents canopy function consistently when tidal phase
38 changes independently of the solar cycle.

39 Light-use-efficiency (LUE) models express canopy GPP as the product of FAPAR
40 and an efficiency scalar and are widely used to translate satellite reflectance into
41 productivity estimates (Pei et al., 2022; Zheng & Takeuchi, 2022). Their temporal
42 scaling commonly relies on solar geometry, although a single overpass may not
43 represent the daytime mean. Diffuse radiation penetrates canopy gaps more evenly than
44 direct radiation and can increase whole-canopy quantum yield by reducing local light



45 saturation ([Chakraborty et al., 2022](#)). Clearness-index models therefore separate direct
46 and diffuse radiation when modelling canopy photosynthesis ([Spitters et al., 1986](#)). In
47 monsoon mangroves, this radiation effect varies alongside vapour pressure deficit; net
48 ecosystem production at three subtropical sites declined above 2.50–2.95 kPa ([Gou et](#)
49 [al., 2024](#)). Solar geometry, radiation quality, and atmospheric demand can thus generate
50 within-day variation before tidal effects are considered.

51 Satellite overpass times are fixed by orbital mechanics, whereas tidal phase
52 follows a separate lunar cycle. The same mangrove stand may therefore be observed
53 with its aerial-root and lower-stem layer exposed or inundated, changing both the
54 background presented to an optical sensor and the environmental conditions under
55 which carbon exchange occurs. Eddy-covariance studies show that tidal and sea-breeze
56 interactions affect net ecosystem exchange ([Zhu et al., 2021a](#)), tidal energy advection
57 alters surface-energy partitioning ([Yong et al., 2024](#)), salinity modifies the coupling
58 between light absorption and carbon assimilation ([Zhu et al., 2021b](#)), and tidal pumping
59 contributes to sub-daily carbon-flux variability ([Zhu et al., 2024](#)). These findings
60 establish a process basis for tidal effects, but they do not show whether tidal phase
61 introduces a systematic error when an instantaneous canopy-function proxy is scaled to
62 a daily value. Resolving that distinction is important for forest monitoring: a repeatable
63 sampling bias should be corrected or flagged, whereas a site-specific ecological
64 response should not be generalized as a universal satellite algorithm.

65 UAV-LiDAR mapping has revealed that the relationship between forest structure
66 and tidal inundation is spatially heterogeneous at the sub-canopy scale ([Shih et al.,](#)



67 [2023](#)), and sun-induced chlorophyll fluorescence shows early promise as a remote
68 indicator of canopy photosynthetic capacity sensitive to salinity and tidal stress ([Zhu et](#)
69 [al., 2021c](#)). Extending tower-derived relationships to satellite pixels requires
70 confronting two sources of spatial uncertainty. Pixel-level observations in raster
71 products are not statistically independent: significance tests applied to full-resolution
72 imagery inflate the apparent number of degrees of freedom unless effective sample size
73 is corrected for spatial autocorrelation ([Meyer & Pebesma, 2021](#)). The SRTM digital
74 elevation model, widely used to delineate tidal inundation zones, carries a nominal
75 vertical error of ± 5 m ([Rodríguez et al., 2006](#)), potentially exceeding the width of
76 intertidal sensitivity zones by nearly an order of magnitude; DEM-based spatial
77 classifications therefore require formal Monte Carlo error propagation before area
78 estimates can be reported with confidence.

79 We used three safeguards to keep the tidal analysis separate from construction of
80 the response. Tidal height was reconstructed independently of the flux-derived proxy.
81 Maximum light-use efficiency was first calibrated without tidal classes, and class-
82 specific estimates were introduced only as a post-changepoint sensitivity analysis.
83 Tower time series and satellite imagery were also treated as different lines of evidence
84 because MODIS composites and three Sentinel-2 scenes could not validate a sub-daily
85 process independently. Changepoint analysis located an empirical transition without
86 fixing its elevation in advance ([Killick et al., 2012](#)), while spatial-autocorrelation
87 diagnostics and DEM error propagation limited the weight assigned to nominal pixel
88 counts ([Meyer & Pebesma, 2021](#)).



89 Here, we combine half-hourly eddy-covariance flux records from the Qinglangang
90 Mangrove Reserve with harmonically reconstructed tidal stages, MODIS FAPAR,
91 Sentinel-2 imagery, and Monte Carlo error propagation to (i) test whether tidal state
92 covaries with the tower-derived apparent FAPAR proxy after keeping the primary
93 efficiency calibration independent of tidal class, (ii) quantify the error from treating the
94 nominal satellite-overpass value as the daytime mean and determine whether the tidal
95 relationship follows a continuous gradient or a site-specific threshold, and (iii) assess
96 whether reserve-wide satellite patterns are directionally compatible with the tower
97 result after accounting for spatial autocorrelation and DEM uncertainty. This design
98 evaluates tidal sampling risk and a site-calibrated scaling ratio; it does not constitute an
99 independently validated tide-conditioned correction.

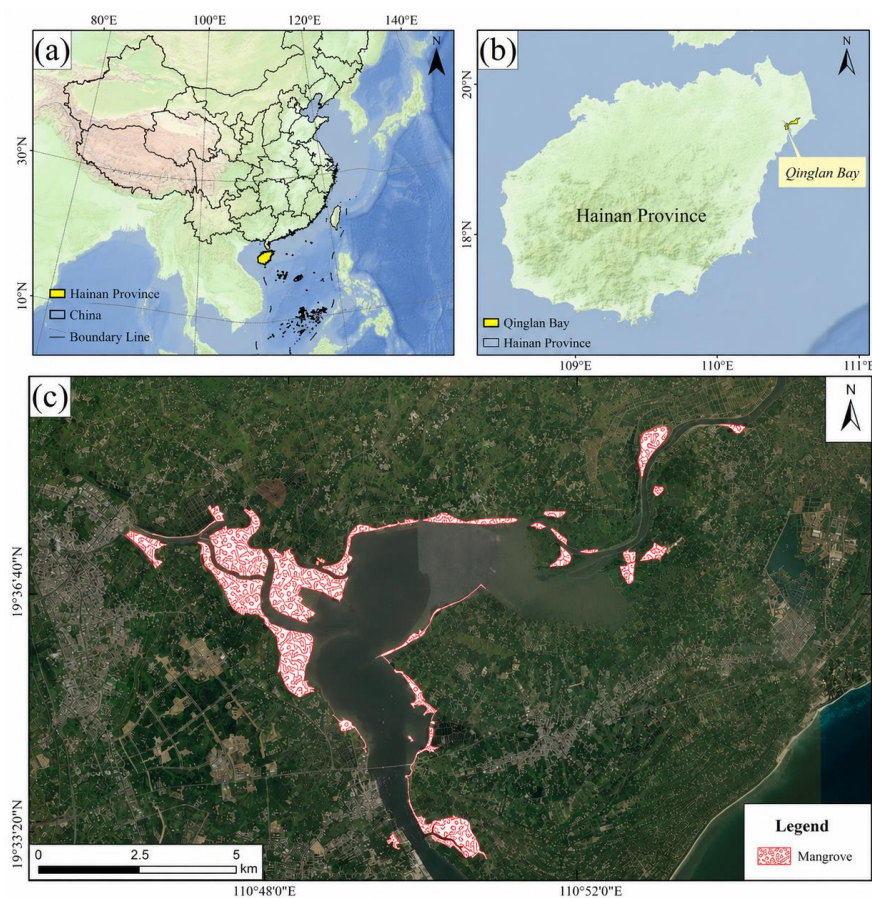
100 **2. Materials and Methods**

101 We used a four-stage workflow. Quality-controlled eddy-covariance GPP and
102 radiation observations first yielded a half-hourly apparent FAPAR proxy. Tidal
103 elevation was reconstructed independently, and changepoint analysis was used to locate
104 a shift in that proxy. The nominal overpass value was then compared with the daytime
105 mean to quantify temporal-scaling error. MODIS and Sentinel-2 products provided
106 benchmarks for seasonal magnitude and spatial pattern, respectively. Tidal
107 classification diagnosed a potential source of sampling error but did not enter the
108 constant median scaling ratio directly. Because the datasets differed in temporal support
109 and footprint, agreement among them was interpreted as corroboration rather than
110 independent validation.



111 2.1 Study design and site context

112 The Qinglangang Mangrove Reserve (19.615° N, 110.798° E) occupies the
113 northeast coast of Hainan Island, China (Fig. 1). Its tropical-monsoon climate and
114 irregular diurnal tide provide a natural test bed in which solar and tidal cycles vary
115 independently. The wet season (May–October) delivers approximately 80% of annual
116 precipitation and has a mean air temperature of 28.2°C; the dry season (November–
117 April) has a mean of 22.5°C. The reserve comprises mature mangrove stands
118 intersected by tidal creeks and channels. Dominant taxa within the tower fetch are
119 *Rhizophora stylosa*, *Bruguiera gymnorhiza*, *Kandelia obovata*, and *Avicennia marina*.
120 Mean canopy height is 6.0 m; displacement height (4.02 m) and roughness length (0.90
121 m) were estimated from the eddy-covariance wind profile. Mean spring and neap tidal
122 ranges are approximately 1.5 and 0.5 m, respectively. The tower location was selected
123 so that the 80% cumulative fetch (approximately 300 m) remained within mangrove
124 canopy for the dominant southeasterly summer and northeasterly winter wind sectors.
125 This configuration reduced land-cover mixing in the tower signal while retaining the
126 full local inundation cycle.



127

128 **Fig. 1. Geographic setting and mangrove distribution of the study area in Qinglan**

129 **Bay, Hainan Province, China. (a) Geographic location of Hainan Province in China.**

130 **(b) Location of Qinglan Bay within Hainan Province. (c) Spatial distribution of**

131 **mangrove forests in Qinglan Bay derived from imagery. Basemap data in panels (a) and**

132 **(b) © Resource and Environment Science and Data Center, Chinese Academy of**

133 **Sciences (RESDC). Satellite imagery in panel (c) © 2026 Google Earth.**

134 **2.2 Eddy covariance system and flux data processing**

135 An open-path eddy-covariance (EC) system operated 18.5 m above the sediment

136 surface from 8 March to 9 October 2023, yielding 10,298 half-hourly intervals over 215



137 days. Three-dimensional wind and sonic temperature were measured at 10 Hz with a
 138 Gill WindMaster Pro, and CO₂/H₂O densities were measured at 10 Hz with a LI-COR
 139 LI-7500DS. Photosynthetic photon flux density (PPFD; $\mu\text{mol m}^{-2} \text{s}^{-1}$) was recorded
 140 with a co-located LI-COR LI-190R quantum sensor. Raw data were processed in
 141 EddyPro v7.0 using despiking, double rotation, block averaging, time-lag
 142 compensation, Webb–Pearman–Leuning density correction, and high-frequency
 143 spectral correction (Webb et al., 1980; Wilczak et al., 2001; Moncrieff et al., 1997, 2004;
 144 Mauder et al., 2013; Fratini & Mauder, 2014). Net ecosystem exchange (NEE) was
 145 calculated as turbulent CO₂ flux plus storage flux. Intervals with SSITC quality flags
 146 ≥ 2 or more than 100 CO₂ spikes were removed before partitioning. Reporting the
 147 acquisition period, instrument heights, frequencies, corrections, and exclusion criteria
 148 defines the measurement domain to which subsequent FAPAR inference applies.

149 NEE was partitioned into gross primary production (GPP) and ecosystem
 150 respiration (Reco) in R v4.3 using REddyProc (Wutzler et al., 2018). Atmospheric
 151 turbulence sufficiency was assessed by the friction-velocity (u^*) threshold method of
 152 Papale et al. (2006); records below the site-specific threshold were excluded. Gaps were
 153 filled by the marginal distribution sampling algorithm (Reichstein et al., 2005).
 154 Nighttime Reco was parameterized using the Lloyd–Taylor (1994) equation fitted in
 155 15-day moving windows to quality-controlled nocturnal NEE (PPFD < 20 $\mu\text{mol m}^{-2}$
 156 s^{-1}):

$$157 \quad R_{eco} = R_{(ref)} \exp \left(E_0 \left(\frac{1}{T_{ref} - T_0} - \frac{1}{T_a - T_0} \right) \right) \quad (1)$$

158 where R_{ref} ($\mu\text{mol CO}_2 \text{ m}^{-2} \text{s}^{-1}$) and E_0 (K) were estimated by non-linear least squares



159 within each window. Daytime R_{eco} was extrapolated from (1), and GPP was obtained
 160 as:

$$161 \quad GPP = R_{eco} - NEE \quad (2)$$

162 GPP was set to zero at night. The attrition sequence was retained explicitly: of 10,298
 163 initial half-hourly intervals, 3,847 daytime records met PPFD $>20 \mu\text{mol m}^{-2} \text{s}^{-1}$ after
 164 flux processing; excluding $K_t < 0.05$ and physically inadmissible apparent-FAPAR
 165 estimates left 3,137 intervals for analysis. This transparent flow separates losses caused
 166 by flux quality control, the daytime definition, radiation screening, and inversion
 167 bounds. Daily summaries were calculated only for days with at least 20 valid daytime
 168 intervals, reducing sensitivity to uneven within-day coverage.

169 **2.3 Environmental, radiation, and tidal data**

170 Half-hourly meteorological variables recorded at the tower included air
 171 temperature (T_a ; °C) and relative humidity (RH; %) from a Vaisala HMP155 probe at
 172 18.5 m; precipitation from a TE525MM tipping-bucket gauge; and soil temperature and
 173 volumetric water content from CS655 sensors (Campbell Scientific) at 5 cm depth.
 174 Vapour pressure deficit (VPD; hPa) was derived from T_a and RH. Downwelling global
 175 solar radiation (R_g ; W m^{-2}) was measured by a Kipp & Zonen CMP6 pyranometer co-
 176 located with the tower. The clearness index was computed as:

$$177 \quad K_t = \frac{R_g}{R_0} \quad (3)$$

178 where R_0 is extraterrestrial irradiance on a horizontal surface, computed from site
 179 coordinates, day-of-year, and local time using standard solar geometry (Gueymard,
 180 2008). The diffuse fraction of incoming radiation was estimated from K_t using the



181 piecewise empirical model of [Spitters et al. \(1986\)](#). Records with $K_t < 0.05$ were
 182 excluded from all FAPAR and DF analyses to avoid physically unreliable diffuse-
 183 fraction estimates under near-opaque cloud.

184 Half-hourly tidal water levels at the tower coordinates were reconstructed from the
 185 FES2014 global ocean tide atlas ([Lyard et al., 2021](#)) by summing contributions of 11
 186 dominant constituents for the South China Sea region using the UTide package. Hourly
 187 predictions were linearly interpolated to 30-min resolution to produce a continuous tidal
 188 height record $h(t)$.

189 **2.4 Tide-robust retrieval of an apparent FAPAR proxy**

190 At each valid half-hour, an apparent canopy FAPAR proxy was derived by
 191 rearranging the Monteith light-use-efficiency framework ([Monteith, 1972, 1977](#)). The
 192 proxy linked observed GPP to incident PPFD through a season-specific maximum-
 193 efficiency term and was used to test temporal covariation with tidal state. We refer to it
 194 as ‘apparent’ FAPAR because EC observations constrain the product of absorbed
 195 radiation and realized efficiency rather than measuring optical absorption independently.

$$196 \quad FAPAR_{inst}(t) = \frac{GPP(t)}{\epsilon_{max}PPFD(t)} \quad (4)$$

197 GPP and PPFD are expressed in $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ and $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$,
 198 respectively, so the derived proxy is dimensionless. To prevent tidal classification from
 199 determining the quantity later tested against tide, the primary ϵ_{max} calibration pooled
 200 tidal states within each season and used only high-irradiance, clear-sky intervals
 201 ($PPFD > 800 \mu\text{mol m}^{-2} \text{ s}^{-1}$; $K_t > 0.6$). After the changepoint had been estimated, ϵ_{max}
 202 was recalibrated separately by tidal class as a sensitivity analysis; its coefficient of



variation quantified how strongly the inferred tidal contrast depended on the efficiency calibration. This sequence was intended to reduce circularity between tidal grouping and FAPAR inversion. It did not make the proxy equivalent to an independent optical FAPAR measurement, so satellite FPAR was used only as an external magnitude benchmark.

2.5 Data-driven threshold detection and tidal classification

To distinguish a threshold-like response from a continuous gradient without pre-specifying its elevation, the apparent FAPAR observations were ordered by independently reconstructed tidal height and analyzed with the PELT changepoint algorithm using a radial-basis-function cost. The tidal series was generated without using GPP, PPF, or FAPAR, preserving separation between the candidate driver and the response. Because adjacent half-hours and neighbouring tidal elevations are not independent replicates, the changepoint was interpreted as a site-level structural estimate rather than a universal population threshold.

$$\sum C(y_{\tau_{i-1}+1:\tau_i}) + \beta m \quad (5)$$

Here, m is the number of changepoints, τ_i their ordered locations, $C(\cdot)$ is the segment cost, and β is the modified-BIC penalty (Zhang & Siegmund, 2007). The detected breakpoint was examined using piecewise linear fits above and below the candidate elevation. Robustness was evaluated by stepping candidate thresholds from -0.5 to $+0.5$ m MSL in 0.1 -m increments and by checking whether the direction of the exposed–flooded contrast was retained across seasons. These analyses were used to assess threshold stability; they did not substitute for independent replication at another site or



225 year.

226 Only after threshold estimation were intervals classified as exposed [$h(t) < h_c$] or
 227 flooded [$h(t) \geq h_c$]. The ± 0.3 m band around h_c was defined as a spatial sensitivity zone
 228 to accommodate unresolved microtopographic variability; it was not treated as a
 229 precisely mapped ecological boundary. Exposed–flooded comparisons used Kruskal–
 230 Wallis and Bonferroni-adjusted Mann–Whitney tests, with effect magnitude and
 231 seasonal consistency considered alongside p values. Given serial dependence among
 232 half-hourly intervals, these tests characterized within-record separation and were not
 233 interpreted as replication across independent ecosystems.

234 **2.6 Instantaneous-to-daily scaling and tidal diagnostics**

235 The instantaneous-to-daily FAPAR upscaling ratio was defined as:

$$236 \quad R_{FAPAR} = \frac{FAPAR_{(inst)(t_{overpass})}}{(FAPAR)_{daily}} \quad (6)$$

237 where $FAPAR_{inst(t_{overpass})}$ is the tower-derived FAPAR at the Sentinel-2 overpass time
 238 (10:30 LST) and $FAPAR_{daily}$ is the daytime-mean FAPAR for that calendar day; R_{FAPAR}
 239 is dimensionless. The analogous GPP upscaling ratio is:

$$240 \quad R_{GPP} = \frac{GPP(t_{overpass})}{(GPP)_{daily}} \quad (7)$$

241 Two operational strategies were compared. The unity-ratio baseline assumed that
 242 FAPAR at 10:30 LST equalled the daytime mean. The site-calibrated strategy used the
 243 median ratio across valid days, with the daily mean recovered by applying the inverse
 244 of the defined instantaneous-to-daily ratio. The median was chosen before model
 245 comparison because it is less sensitive than the mean to days with short-term cloud or
 246 flux anomalies. Tidal class was analysed as a potential source of variation around this



247 scaling relationship but was not included as a predictor in the constant-ratio calibration.
 248 The analogous GPP ratio was retained as a process comparison rather than used to
 249 redefine FAPAR.

$$250 \quad \overline{FAPAR}_{daily}^{cal} = \frac{FAPAR_{(inst)(t_{overpass})}}{\bar{R}_{FAPAR}} \quad (8)$$

251 Performance was summarized by RMSE against observed daytime means for 163
 252 valid days and compared with the unity-ratio baseline. The same single-site record
 253 supplied both the median ratio and the evaluation days, so the reduction in error
 254 measured internal calibration performance rather than out-of-sample prediction. A ridge
 255 model (Hoerl & Kennard, 1970) was fitted as an exploratory test of whether the
 256 available covariates added information beyond the constant median. Neither the
 257 constant ratio nor the ridge model was evaluated with temporally blocked or external
 258 data; transferability was therefore not assessed.

259 **2.7 Cross-platform benchmarking and scale-aware spatial analysis**

260 MODIS MOD15A2H v061 FPAR/LAI (500 m, 8-day composites; Knyazikhin et
 261 al., 1998; Myneni et al., 2021) was obtained for March–November 2023, reprojected to
 262 WGS84, and clipped to the reserve. Twenty-seven tower–MODIS windows were
 263 matched at the tower footprint to test seasonal magnitude agreement between
 264 methodologically distinct retrievals. Tidal state at the nominal Terra overpass time was
 265 assigned from the reconstructed water level, but the 8-day composite was not treated as
 266 an instantaneous observation. Consequently, the tidal-class comparison was explicitly
 267 exploratory: a null contrast may reflect compositing, footprint mixing, retrieval
 268 uncertainty, or limited class counts.



269 Three cloud-limited Sentinel-2 Level-2A scenes (March, July, and October 2023;
 270 10 m; reported cloud cover <10%) were processed with the SNAP biophysical
 271 processor, which retrieves FAPAR through a PROSAIL-trained neural network
 272 (Jacquemoud et al., 2009). Vegetated, cloud-free pixels with NDVI >0.3 were retained.
 273 Sentinel-2 provided spatial pattern information at higher resolution, but only three
 274 seasonal snapshots; it was therefore used to examine directional consistency with the
 275 tower threshold rather than to estimate a temporally general tidal effect.

276 SRTM 30-m DEM tiles were bilinearly resampled to the 10-m Sentinel-2 grid.
 277 Using the changepoint elevation h_c identified in §2.5, pixels were assigned to three
 278 zones: elevated ($\text{DEM} > h_c + 0.3$ m), tidal-FAPAR sensitive ($h_c - 0.3$ m to $h_c + 0.3$ m),
 279 and low-lying ($\text{DEM} < h_c - 0.3$ m). Spatial autocorrelation of Sentinel-2 FAPAR was
 280 quantified by fitting an exponential semivariogram model:

$$281 \quad \gamma(d) = c_0 + c_1 \left(1 - \exp\left(-\frac{d}{a}\right)\right) \quad (9)$$

282 using the 'gstat' R package (Pebesma, 2004). The decorrelation distance d_{corr} equals the
 283 range parameter a . The effective number of spatially independent samples within the
 284 reserve was estimated as:

$$285 \quad N_{\text{eff}} \approx \frac{A}{\pi d_{\text{corr}}^2} \quad (10)$$

286 When N_{eff} was substantially smaller than the pixel count, formal pixel-level
 287 significance testing was not performed; the Sentinel-2 spatial patterns were instead
 288 reported as qualitative consistency checks.

289 SRTM's nominal ± 5 m vertical uncertainty exceeds the 0.6 m sensitivity-zone
 290 width. A Monte Carlo propagation was therefore used to report a distribution and 95%



291 interval for the detectable zone fraction rather than an unqualified area estimate. This
292 analysis quantified classification sensitivity to DEM error; it could not recover sub-
293 metre elevation information absents from SRTM.

294 **2.8 Exploratory spatial associations and statistical safeguards**

295 Random forests were used only to describe multivariate associations with scene-
296 specific Sentinel-2 FAPAR. Six candidate predictors were examined: atmospheric CO₂,
297 solar radiation, precipitation, land-surface temperature, PAR, and SRTM elevation.
298 Models used 500 trees implemented with ranger ([Wright & Ziegler, 2017](#)) and were
299 summarized by out-of-bag R² and SHAP values ([Lundberg & Lee, 2017](#)). NASA
300 POWER meteorological variables have 0.5° resolution and therefore provide broad
301 scene-level context rather than independent 10-m variation within the reserve. SHAP
302 rankings were interpreted as properties of the fitted FAPAR models under correlated
303 and scale-mismatched predictors, not as causal partitions, spatially independent
304 validation, or pixel-scale environmental effects.

305 Statistical interpretation followed the sampling support of each dataset. Non-
306 parametric group comparisons were used for skewed tower-derived quantities, with
307 exact sample sizes and effect magnitudes reported alongside p values. Calendar days,
308 rather than half-hours, were the support for upscaling performance. For Sentinel-2, an
309 exponential semivariogram was used to estimate decorrelation distance and effective
310 sample size; nominal pixel counts were not treated as independent replication. Cross-
311 sensor comparisons used matched temporal windows, while differences in footprint and
312 compositing were carried into interpretation. This hierarchy prevents large half-hourly



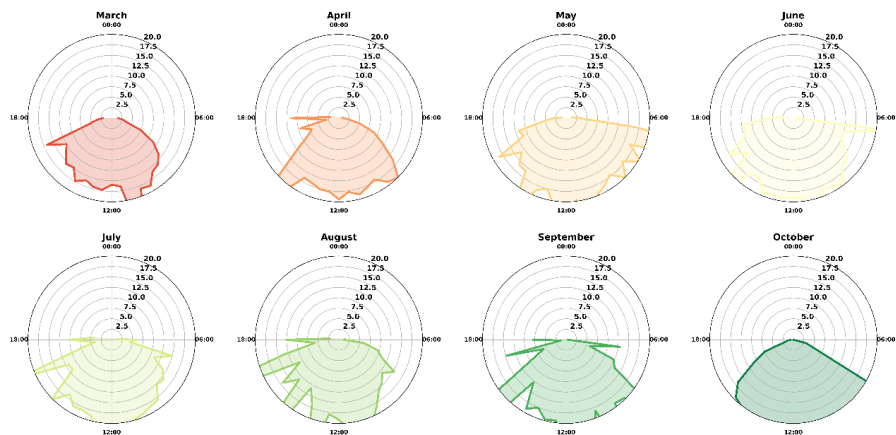
313 or pixel counts from generating spurious precision and keeps confirmatory, sensitivity,
314 and exploratory analyses distinct.

315 All analyses were organized around three levels of evidence. Half-hourly tower
316 observations supported process-level and changepoint analyses; calendar days were the
317 units for instantaneous-to-daily scaling; and effective, rather than nominal, pixel counts
318 governed interpretation of spatial imagery. Results derived from the same record used
319 for calibration were labelled internal or exploratory. Parameters, thresholds, exclusion
320 rules, software versions, and sample attrition were reported to facilitate reproduction.
321 This organization separated estimation, sensitivity analysis, and external benchmarking
322 and avoided treating agreement among non-independent analyses as independent
323 validation.

324 **3. Results**

325 **3.1 Seasonal diurnal GPP and environmental context**

326 Half-hourly GPP varied seasonally in both magnitude and diurnal shape (Fig. 2).
327 The midday peak increased from approximately $12 \mu\text{mol m}^{-2} \text{s}^{-1}$ in March to $18 \mu\text{mol}$
328 $\text{m}^{-2} \text{s}^{-1}$ in July and then declined through October. The diurnal envelope was narrowest
329 in March and broadest in July–August. Monthly GPP amplitude covaried with day
330 length ($r = 0.89$), consistent with a strong seasonal influence of solar geometry. Tidal
331 differences in the apparent FAPAR proxy were examined against this seasonal
332 background.



333
334 **Fig. 2. Monthly diurnal GPP patterns.** Polar coordinates: angle represents time of
335 day, radius represents mean half-hourly GPP ($\mu\text{mol m}^{-2} \text{s}^{-1}$). The midday GPP peak
336 strengthens from March to July and contracts by October, reflecting seasonal variation
337 in photoperiod and photosynthetic capacity.

338 Several environmental variables differed between tidal states (Fig. 3). Exposed
339 periods had higher water-use efficiency than flooded periods (Kruskal–Wallis $p < 0.05$;
340 Fig. 3a). The H/LE ratio was also higher during flooding (Fig. 3b), while friction
341 velocity declined weakly with tidal height (Fig. 3c). Near-surface CO_2 concentration
342 and the Bowen ratio were higher during flooded intervals (Fig. 3d,e). Diffuse radiation
343 followed similar diurnal courses in the two tidal classes (Fig. 3f). These comparisons
344 describe covariation within the tower record; they do not by themselves identify the
345 processes responsible for the differences.

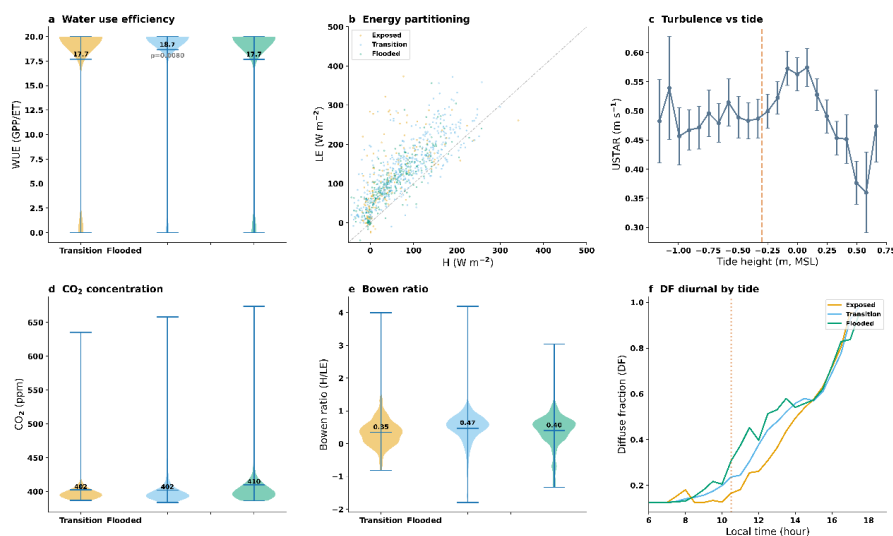


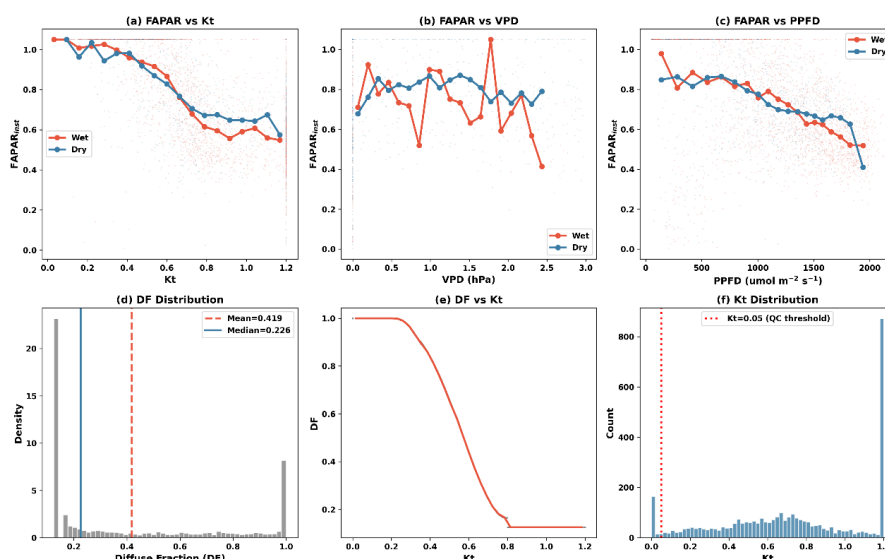
Fig. 3. Environmental responses to tidal state. (a) Water use efficiency ($\text{WUE} = \text{GPP}/\text{ET}$). (b) Energy partitioning (H vs LE) colored by tidal state. (c) Friction velocity (U_{STAR}) as a function of continuous tide height; dashed line marks the -0.3 m changepoint. (d) Near-surface CO_2 concentration by tidal state. (e) Bowen ratio (H/LE) by tidal state. (f) Diurnal course of the diffuse fraction (DF) stratified by tidal state.

3.2 Tidal covariation in apparent FAPAR and temporal-scaling error

FAPAR_{inst} declined with increasing clearness index (K_t) in all three seasons (Fig. 4a); binned means fell from approximately 0.85 at $K_t < 0.3$ to approximately 0.65 at $K_t > 0.8$. FAPAR_{inst} also decreased across the observed VPD range (Fig. 4b) but showed little additional response to PPFD above approximately $500 \mu\text{mol m}^{-2} \text{s}^{-1}$ (Fig. 4c). The diffuse-fraction distribution was right-skewed (mean = 0.46, median = 0.27; Fig. 4d), and its relationship with K_t followed the Spitters curve (Fig. 4e). Records with $K_t < 0.05$, representing 6.9% of daytime observations, were excluded (Fig. 4f). Because the proxy was derived from GPP, PPFD, and calibrated efficiency, its associations with



361 Kt and VPD could reflect changes in realized light-use efficiency as well as optical
 362 absorption.



363

364 **Fig. 4. Environmental associations and radiation diagnostics for FAPAR_{inst}.** (a–c)

365 FAPAR_{inst} in relation to Kt, VPD, and PPFD by season (binned means $\pm$ SE). (d)
 366 Distribution of diffuse fraction (DF). (e) DF in relation to Kt, with binned mean shown
 367 by the solid line. (f) Distribution of Kt; the dashed line indicates the exclusion threshold
 368 (Kt < 0.05).

369 Instantaneous FAPAR at the nominal overpass time was positively associated with
 370 the daytime mean ($r = 0.76$; Fig. 5a). Across 163 valid days, the instantaneous-to-daily
 371 ratio had a median of 0.83 (Fig. 5b), so the 10:30 value averaged 83% of the daytime
 372 mean. Applying this ratio reduced RMSE by 28% relative to the unity-ratio assumption
 373 (Fig. 5c). This was an in-sample improvement because the ratio and its performance
 374 were estimated from the same record.

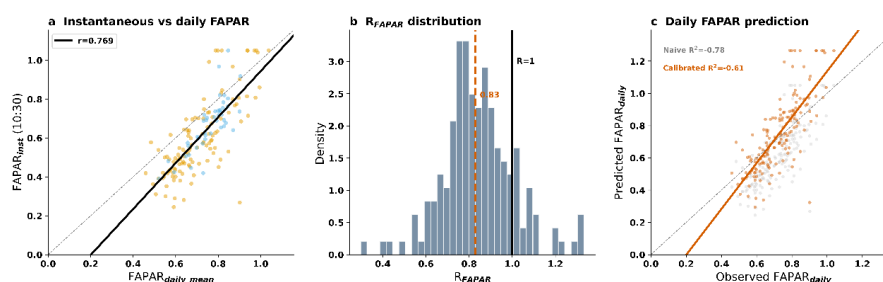


Fig. 5. Instantaneous-to-daily FAPAR upscaling. (a) FAPAR at Sentinel-2 overpass (10:30) versus daytime-mean FAPAR. (b) Distribution of the upscaling ratio $R_{FAPAR} = FAPAR_{inst} / FAPAR_{daily\ mean}$ (median = 0.83). (c) Daily FAPAR predicted by the naive assumption ($R = 1$, grey) and the calibrated ratio ($R = 0.83$, orange).

Half-hourly $FAPAR_{inst}$ differed between tidal states, with exposed-period values 9.8% higher than flooded-period values (0.780 ± 0.239 vs. 0.711 ± 0.271 ; Kruskal–Wallis $p = 0.0075$; Bonferroni-adjusted Mann–Whitney U $p = 0.003$; Fig. 6a). These p values quantify separation within the serially dependent tower record and do not represent replication across independent ecosystems. The direction of the contrast was consistent across spring, summer, and autumn. State-specific LUE calibration under $PPFD > 800 \mu mol m^{-2} s^{-1}$ and $K_t > 0.6$ yielded a coefficient of variation of 3.7% (Fig. 6b), indicating limited sensitivity to the selected calibration while not independently excluding inversion bias. The exposed–flooded separation persisted across daylight hours and was greatest from mid-morning to early afternoon (Fig. 6c). Changepoint detection identified one break at -0.3 m MSL (Fig. 6d). Piecewise fits showed little additional trend on either side, consistent with a threshold-like pattern within this dataset. The interval from -0.6 to 0 m MSL was retained as the site-specific tidal-FAPAR sensitivity zone.

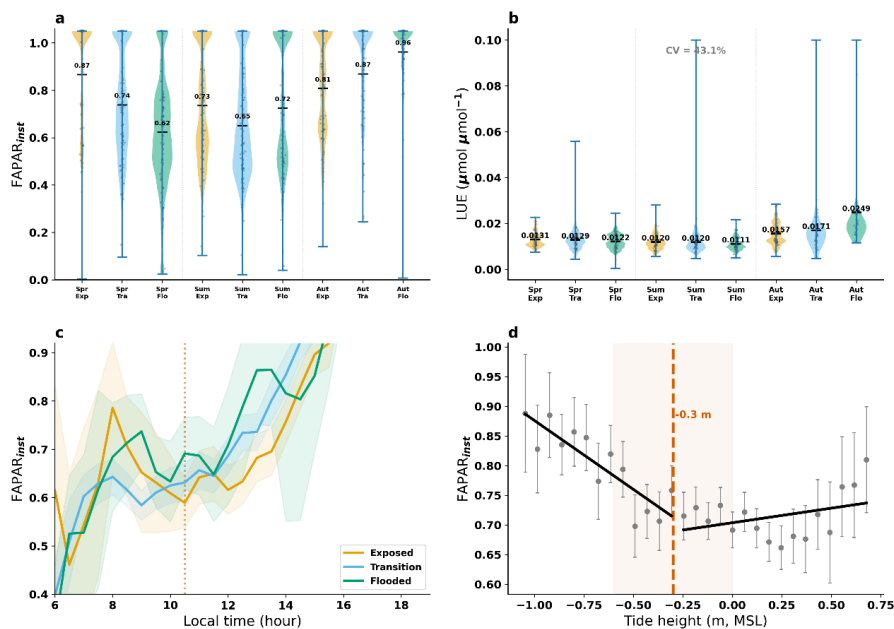


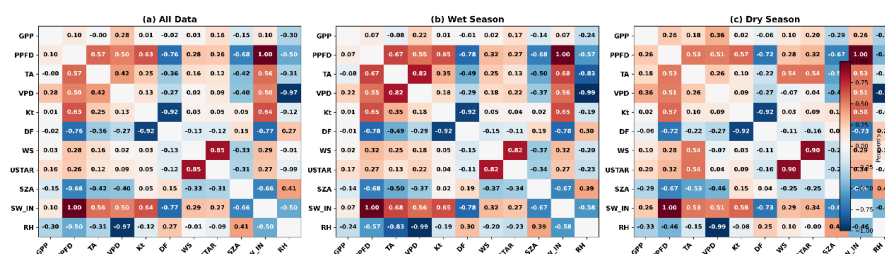
Fig. 6. Tidal modulation of instantaneous FAPAR. (a) FAPAR_{inst} by season and tidal state. (b) LUE under optimal conditions, stratified by season and tidal state (CV = 3.7%). (c) Diurnal FAPAR_{inst} by tidal state (shading: 95% CI). (d) Change-point detection at -0.3 m MSL; red band: sensitivity zone.

3.3 Cross-platform benchmarking and seasonal stability

Correlation matrices were used to describe the seasonal stability of environmental associations (Fig. 7). GPP correlated most strongly with PPFD ($r = 0.65\text{--}0.72$), followed by Kt ($r = 0.45\text{--}0.55$). The GPP–DF correlation was negative in both seasons and stronger during the wet season ($r = -0.38$ versus -0.21). Correlations with VPD ($r = -0.15$ to -0.22) and air temperature ($r = 0.18\text{--}0.25$) were weaker. Kt and DF were strongly negatively correlated ($r = -0.72$ to -0.78). The rank order of these associations was similar in wet and dry subsets, although their magnitudes differed. These descriptive results supported use of the pooled daytime record but did not establish



causal environmental controls.



409

410 **Fig. 7. Variable correlation matrices by season.** Pearson correlation coefficients
 411 among eleven half-hourly variables for (a) all data, (b) wet season, and (c) dry season.
 412 GPP is most strongly correlated with PPFD ($r = 0.65$ – 0.72) and Kt ($r = 0.45$ – 0.55). The
 413 GPP–DF correlation is negative and stronger in the wet season ($r = -0.38$) than the dry
 414 season ($r = -0.21$), consistent with greater cloud–radiation variability during the
 415 monsoon

416 Tower-derived FAPAR and MODIS MOD15A2H FPAR were rank-correlated
 417 across 27 matched 8-day windows at the tower pixel (Spearman $\rho = 0.47$, $p = 0.013$;
 418 RMSE = 0.095; bias = -0.02 ; Fig. 8a). Both series declined from March to October
 419 (Fig. 8b). The half-hourly tower record retained the 9.8% exposed–flooded contrast
 420 reported in Section 3.2, and a representative 3-day interval ranged from 0.63 to 0.98
 421 (Fig. 8c). MODIS composites did not differ detectably among tidal states (Kruskal–
 422 Wallis $p = 0.36$); exposed and flooded means were 0.71 ± 0.07 ($n = 7$) and 0.76 ± 0.04
 423 ($n = 5$), respectively (Fig. 8f). The MODIS value for 9 March (0.76) was close to the
 424 corresponding 3-day tower mean (0.77), but the comparison cannot resolve sub-daily
 425 tidal variation.

426 Stratified by DEM-derived tidal sensitivity zone, wet-season FAPAR was $0.60 \pm$



0.29 ($n = 18$) for elevated pixels (> 0 m), 0.71 ± 0.19 ($n = 3$) for sensitive pixels (-0.6
 to 0 m), and 0.59 ± 0.09 ($n = 8$) for low-lying pixels (< -0.6 m); dry-season values were
 0.57 ± 0.28 ($n = 21$), 0.61 ± 0.18 ($n = 3$), and 0.54 ± 0.18 ($n = 9$), respectively (Fig. 8d).
 Differences among zones were not statistically significant in either season (wet:
 Kruskal–Wallis $p = 0.81$; dry: $p = 0.98$). The tower-pixel MODIS FAPAR fell within
 the central mass of the reserve-wide distribution in both seasons (Fig. 8e).

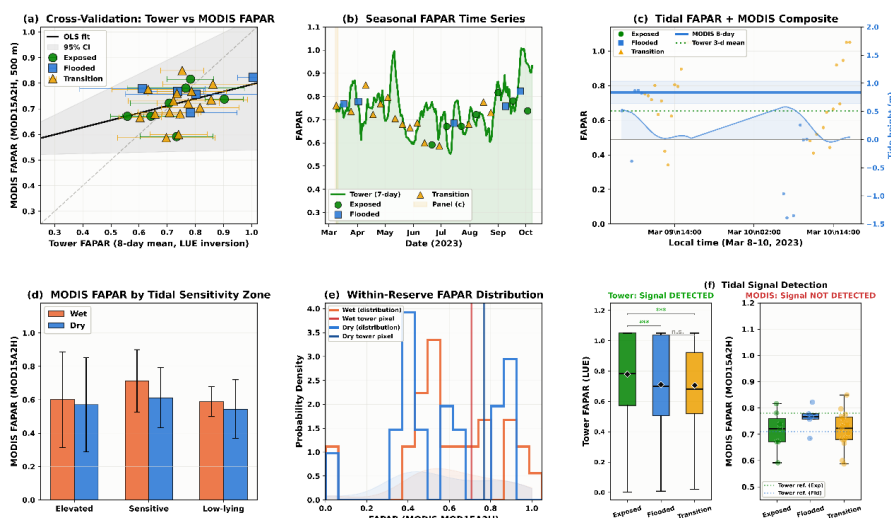


Fig. 8. Comparison of tower-derived and MODIS FAPAR. (a) Tower-derived FAPAR versus MODIS MOD15A2H at the tower pixel, classified by tidal state at nominal overpass. (b) Seasonal series of tower and 8-day MODIS values. (c) Three-day interval (8–10 March) with the coincident MODIS composite. (d) MODIS FAPAR by DEM-derived tidal-sensitivity zone. (e) Reserve-wide distributions with the tower pixel indicated. (f) Tidal-class comparisons for the tower record and MODIS composites.

3.4 Spatial variation of FAPAR across seasons, elevations, and offshore distances

Sentinel-2 FAPAR varied spatially in all three scenes (Fig. 9). Mean values were



0.137 ± 0.111 in spring, 0.193 ± 0.156 in summer, and 0.187 ± 0.139 in autumn, based on 145,649, 144,980, and 138,775 valid pixels, respectively. Pixels above mean sea level had the highest mean FAPAR (0.154–0.218), whereas the DEM-defined sensitivity zone had lower values (0.081–0.116), a difference of 44–48% with Cohen's κ ranging from −0.65 to −0.67. Low-lying pixels were intermediate (0.097–0.156). Because pixel values were spatially autocorrelated and elevation was uncertain, these contrasts were treated as descriptive spatial patterns rather than independent tests of tidal effects.

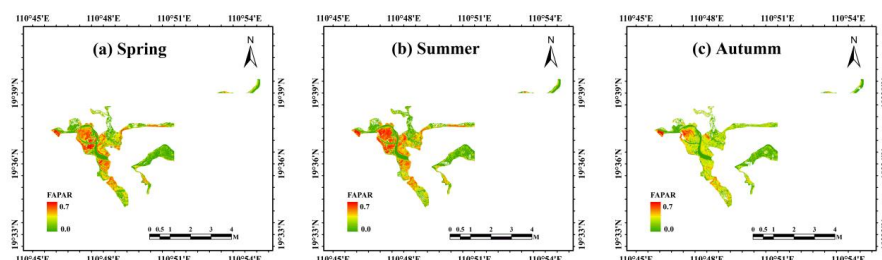


Fig. 9. Seasonal FAPAR spatial maps. Sentinel-2 FAPAR for (a) Spring, (b) Summer, and (c) Autumn.

Binned FAPAR increased with elevation in each Sentinel-2 scene (Fig. 10). Within −2 to +2 m MSL, fitted slopes ranged from 0.008 to 0.012 m^{−1} and descriptive correlations ranged from $r = 0.71$ to 0.79. Values approached 0.25–0.30 above +4 m. FAPAR also increased with distance from mapped tidal channels, from 0.10–0.12 within 20 m to 0.18–0.22 beyond 200 m, although this association was weaker ($r = 0.42$ –0.55). Distance to a mapped channel is an imperfect proxy because hydrological connections vary among channels. The elevation and distance patterns were therefore interpreted as spatial context for the tower result, not as confirmation of its mechanism.

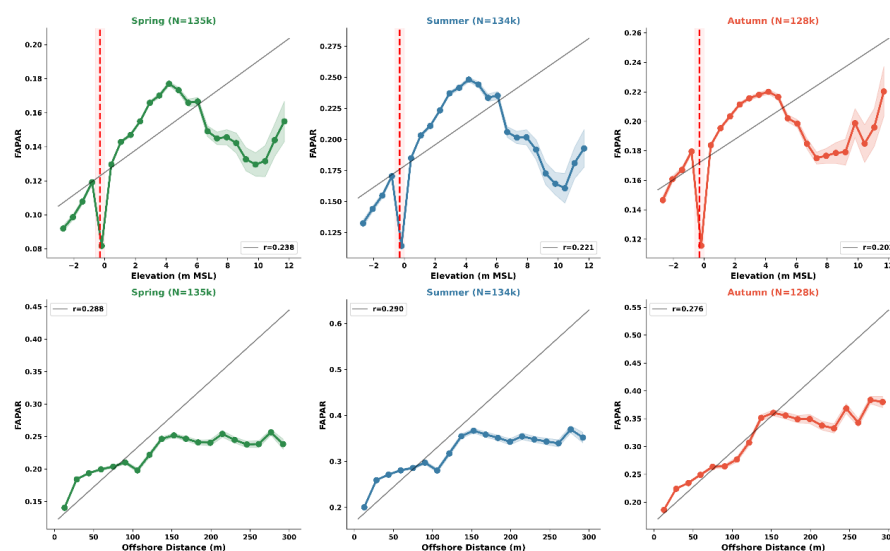


Fig. 10. FAPAR as a function of elevation and offshore distance. Binned means ($\pm$ SE) of Sentinel-2 FAPAR against (a–c) elevation for Spring, Summer, and Autumn, and (d–f) offshore distance for the same seasons. The red dashed line marks the -0.3 m change point; the red band spans the tidal-FAPAR sensitivity zone.

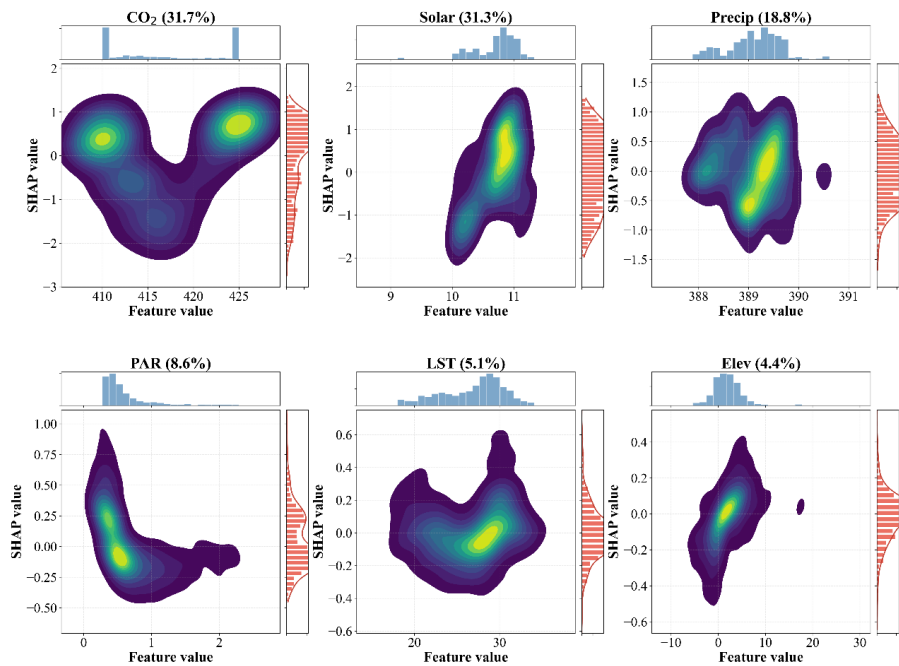
3.5 Exploratory environmental associations with mapped FAPAR

Random-forest models fitted to scene-specific Sentinel-2 FAPAR with six environmental predictors yielded out-of-bag R^2 values of 0.807 in spring, 0.692 in summer, and 0.740 in autumn (Figs. 11–13). These values describe model fit within the sampled imagery; because pixels were spatially autocorrelated and several meteorological predictors were available only at 0.5° resolution, they should not be interpreted as spatially independent predictive performance.

In spring, SHAP importance was 31.7% for CO_2 concentration, 31.3% for solar radiation, 18.8% for precipitation, 8.6% for PAR, 5.1% for land-surface temperature, and 4.4% for elevation (Fig. 11). The fitted FAPAR response was positive at low-to-



476 moderate CO₂ concentrations and flattened above approximately 410 ppm, while solar
477 radiation showed a positive monotonic association. These patterns describe the fitted
478 model under correlated and scale-mismatched predictors and should not be interpreted
479 as causal effects.



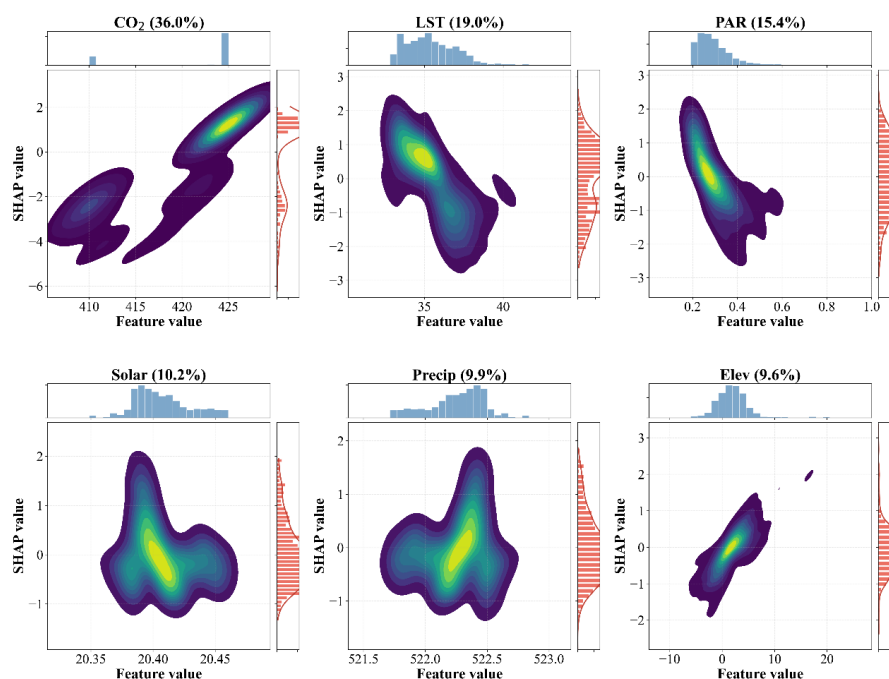
480
481 **Fig. 11. Exploratory environmental associations with mapped FAPAR in spring.**

482 SHAP KDE for six predictors (out-of-bag $R^2 = 0.807$). CO₂ (31.7%) and solar radiation
483 (31.3%) received the largest shares of model importance.

484 In summer, SHAP importance was 36.0% for CO₂, 19.0% for land-surface
485 temperature, 15.4% for PAR, 10.2% for solar radiation, 9.9% for precipitation, and 9.6%
486 for elevation (Fig. 12). The land-surface-temperature range was 24–38°C, and elevation
487 reached its highest relative contribution among the three scenes. The positive elevation
488 association was directionally consistent with the mapped inundation pattern, but the



489 model cannot separate elevation effects from correlated spatial structure.



490

491 **Fig. 12. Exploratory environmental associations with mapped FAPAR in summer.**

492 SHAP KDE for six predictors (out-of-bag $R^2 = 0.692$). CO₂ received the largest share
 493 of model importance (36.0%), followed by LST (19.0%).

494 In autumn, SHAP importance was 44.3% for solar radiation, 22.3% for
 495 precipitation, 15.6% for CO₂, 10.4% for land-surface temperature, and less than 5%
 496 each for PAR and elevation (Fig. 13). The rise in solar-radiation importance from 31.3%
 497 in spring to 44.3% in autumn coincided with the seasonal transition toward clearer
 498 conditions. Because the coarse meteorological predictors mainly varied between scenes
 499 rather than among 10-m pixels, these changes are reported as exploratory model
 500 behavior rather than mechanistic attribution.

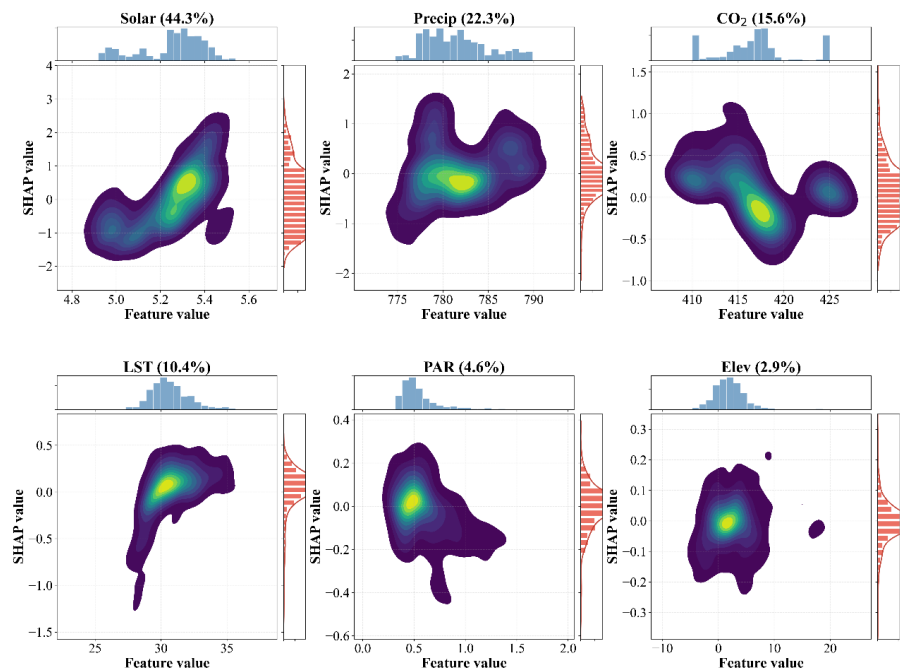


Fig. 13. Exploratory environmental associations with mapped FAPAR in autumn.

SHAP KDE for six predictors (out-of-bag $R^2 = 0.740$). Solar radiation received the largest share of model importance (44.3%), followed by precipitation (22.3%).

Across the three scenes, elevation contributed approximately 3–10% of SHAP importance. Elevation SHAP values were generally positive above approximately 0 m MSL and negative or near zero below this level, directionally matching the tower-derived pattern near -0.3 m MSL. Its contribution remained smaller than those assigned to CO_2 , solar radiation, and precipitation. This consistency is descriptive rather than an independent validation because elevation was included directly in the landscape model and both analyses represent the same site.

4. Discussion

4.1 Ecological interpretation of the -0.3 m tidal threshold



514 Piecewise analysis identified a site-specific break near -0.3 m MSL (Fig. 6d). The
 515 apparent FAPAR proxy was 9.8% higher during exposed than flooded intervals, and the
 516 direction of this contrast was retained across spring, summer, and autumn. The
 517 breakpoint therefore marked a repeatable transition within the tower record rather than
 518 a difference confined to one season. Its ecological meaning must nevertheless be kept
 519 local: the elevation is referenced to the site datum and may vary with stand structure,
 520 microtopography, species composition, and tidal range.

521 The estimated elevation overlaps the reported vertical range of aerial-root
 522 structures in some mangrove taxa (Nguyen et al., 2023), but root height was not
 523 measured at the tower footprint. Two non-exclusive processes could contribute. Water
 524 may obscure part of the aerial-root and understory stratum, thereby changing the
 525 fraction of plant area exposed to incident radiation, while root-zone inundation may
 526 alter gas exchange and realized light-use efficiency (Zhu et al., 2021b; Gou et al., 2023).
 527 Because the response variable was inferred from GPP and PPFD, the analysis could not
 528 separate an optical change from a physiological response. The threshold is consequently
 529 compatible with these processes but does not demonstrate either one.

530 Several diagnostics narrowed the range of alternative explanations. State-specific
 531 ϵ_{max} estimates varied by only 3.7% (Fig. 6b), indicating that the exposed–flooded
 532 contrast was not highly sensitive to the efficiency calibration used after changepoint
 533 detection. Diffuse fraction showed no clear tidal-state separation (Fig. 3f), whereas
 534 water-use efficiency, the H/LE ratio, friction velocity, near-surface CO_2 , and the Bowen
 535 ratio covaried with tidal state (Fig. 3). These concurrent changes suggest that inundation



536 was embedded in a broader micrometeorological transition. They also explain why the
537 observed contrast should be described as tidal covariation in apparent FAPAR, rather
538 than as a direct measurement of canopy absorption or a single physiological mechanism.

539 **4.2 Consequences for instantaneous-to-daily scaling**

540 The temporal-scaling analysis showed that the nominal 10:30 overpass did not
541 represent the daytime mean uniformly. Across 163 valid days, the median
542 instantaneous-to-daily FAPAR ratio was 0.83 (Fig. 5b), and applying this empirical
543 ratio reduced RMSE by 28% relative to the unity-ratio baseline (Fig. 5c). The size of
544 the improvement indicates that assuming equality between an instantaneous
545 observation and the daytime mean introduced appreciable error at this site. Because the
546 same record supplied both the ratio and its evaluation, however, the 28% reduction
547 measures internal calibration performance and should not be read as an expected out-
548 of-sample gain.

549 This reduction was larger than the 7–8% reported for temporal upscaling of
550 MODIS FAPAR across 78 FLUXNET forest sites (Zhang et al., 2023). Differences in
551 canopy structure, sampling time, radiation regime, tidal exposure, and validation design
552 could all contribute to the contrast. The present study cannot assign the difference
553 specifically to tide, but it identifies tidal state as an additional source of within-day
554 variation that is absent from conventional solar-geometry scaling. A fixed overpass may
555 therefore sample not only a particular point on the diurnal radiation curve but also a
556 different hydrological state from one date to another.

557 Radiation quality provides a second source of variation. Apparent FAPAR declined



558 with increasing clearness index and VPD (Fig. 4), and diffuse fraction was retained in
 559 the exploratory ridge model of the scaling ratio. These patterns are consistent with the
 560 known influence of diffuse radiation and atmospheric demand on canopy
 561 photosynthesis (Chakraborty et al., 2022; Gou et al., 2024), although they partly reflect
 562 variation in realized efficiency because the proxy was derived through an LUE
 563 formulation. For operational scaling, the practical implication is that tide and radiation
 564 should be evaluated jointly. Their statistical contributions cannot be treated as
 565 independent when cloud conditions, atmospheric mixing, and tidal phase covary within
 566 the available record.

567 **4.3 Cross-platform agreement and scale mismatch**

568 The tower and MODIS series agreed moderately in seasonal rank across 27
 569 matched composite windows (Spearman $\rho = 0.47$, $p = 0.013$; Fig. 8a), and both declined
 570 from March to October (Fig. 8b). This correspondence provides a useful magnitude
 571 check for the EC-derived proxy, but it is not a validation of the inferred sub-daily
 572 process. MODIS did not resolve a tidal-state difference ($p = 0.36$; Fig. 8f), despite the
 573 9.8% contrast in the half-hourly tower record. Eight-day compositing can average
 574 observations made under different tidal states, while retrieval uncertainty and the
 575 mismatch between a 500-m pixel and the variable EC footprint further weaken direct
 576 comparison.

577 Sentinel-2 supplied finer spatial detail but did not remove the scale problem. Mean
 578 FAPAR was lowest in the DEM-defined -0.6 to 0 m sensitivity zone, and the difference
 579 from elevated pixels was 44–48% across the three scenes (Cohen's $d = -0.65$ to -0.67 ;



580 [Figs. 9, 10](#)). FAPAR also increased with elevation more strongly than with distance
 581 from mapped tidal channels ($r = 0.71\text{--}0.79$ versus $0.42\text{--}0.55$; [Fig. 10](#)). These patterns
 582 are consistent with an inundation gradient, but elevation and channel distance also
 583 covary with drainage, salinity, nutrient supply, vegetation structure, and mapping
 584 geometry ([Gou et al., 2023](#); [Zhu et al., 2021b](#)). The imagery therefore adds spatial
 585 context rather than a separate test of the tower-derived threshold.

586 Nominal pixel counts considerably overstated the information available for spatial
 587 inference. Decorrelation distances of approximately 550–800 m reduced the effective
 588 sample size to about 24–30 within the 31.9 km² reserve, so pixel-level significance was
 589 not interpreted as independent replication ([Dutilleul et al., 1993](#); [Meyer & Pebesma,](#)
 590 [2021](#)). Vertical uncertainty was even more restrictive: the nominal ± 5 m uncertainty of
 591 SRTM ([Rodríguez et al., 2006](#); [Farr et al., 2007](#)) was much larger than the 0.6 m
 592 sensitivity-zone width. The Monte Carlo reduction in detectable area from 19.1% to
 593 3.9% quantified this vulnerability but could not recover sub-metre topography.
 594 Agreement among the three platforms should thus be judged by compatible direction
 595 and scale, not by treating their observations as equivalent.

596 **4.4 Seasonal spatial associations and model interpretation**

597 The random-forest models described different seasonal allocations of predictor
 598 importance ([Figs. 11–13](#)). Out-of-bag R^2 declined from 0.807 in spring to 0.692 in
 599 summer and recovered to 0.740 in autumn. CO₂ and solar radiation had similar
 600 importance in spring, CO₂ ranked first in summer, and solar radiation ranked first in
 601 autumn. This reordering is plausible in a monsoon mangrove system where radiation,



602 atmospheric demand, precipitation, and tidal exchange vary seasonally. It should not,
603 however, be interpreted as a causal partition of controls because SHAP values allocate
604 importance within the fitted predictor set and correlated predictors can exchange
605 importance ([Aas et al., 2021](#); [Pei et al., 2022](#)).

606 The relatively high CO₂ importance illustrates this interpretive problem. Near-
607 surface CO₂ differed between tidal states in the tower record ([Fig. 3d](#)), and tidal cycles
608 can influence mangrove stem, soil, and ecosystem carbon exchange ([Yong et al., 2024](#);
609 [Zhu et al., 2024](#)). At the spatial resolution used here, CO₂ may also represent
610 atmospheric mixing, ventilation, seasonal meteorology, or covariance with radiation
611 rather than a direct control on canopy absorption. The fitted FAPAR response therefore
612 cannot distinguish a biological CO₂ association from tidal and atmospheric covariation.
613 Vertically resolved CO₂ profiles, turbulence observations, optical FAPAR
614 measurements, and in situ water levels would be required to evaluate these alternatives.

615 Elevation contributed only about 3–10% of SHAP importance, reaching 9.6% in
616 summer ([Fig. 12](#)). Its association with FAPAR was directionally compatible with the
617 tower changepoint, but this agreement was structurally non-independent because
618 elevation defined the spatial zones and entered the landscape model directly. In addition,
619 model performance within the sampled predictor domain does not establish map
620 accuracy in poorly represented environmental conditions ([Meyer & Pebesma, 2021](#);
621 [Wadoux et al., 2021](#)). The seasonal models are therefore most useful for identifying
622 hypotheses and measurement priorities. They indicate when radiation, CO₂ covariance,
623 and elevation deserve closer observation, rather than ranking universal controls on



624 mangrove FAPAR.

625 **4.5 Monitoring implications, limitations, and priorities for validation**

626 The immediate monitoring implication is procedural rather than numerical.
627 Optical acquisitions in intertidal forests should be paired with water level referenced to
628 a local vertical datum, and observations close to a locally estimated transition should
629 be flagged or assigned additional uncertainty. Such screening could reduce the chance
630 that a change in inundation state is mistaken for a change in canopy condition. The -0.3
631 m threshold and the 0.83 scaling ratio are not transferable constants; they are site
632 calibrations for a low-stature mangrove stand with exposed aerial roots. Other species
633 assemblages, geomorphic settings, and tidal ranges require independent estimation.

634 The principal limitation is that one site and one partial year cannot establish
635 generality. Tidal height was reconstructed from FES2014 rather than measured at the
636 tower, leaving unquantified classification error near the changepoint. The apparent
637 FAPAR proxy was constrained by EC-derived GPP and a calibrated efficiency term, not
638 by an independent optical instrument, and the same temporal record was used for
639 calibration and evaluation. In the landscape analysis, coarse meteorological fields,
640 uncertain elevation, predictor collinearity, and spatial dependence further limited
641 attribution. These restrictions do not negate the within-record patterns, but they narrow
642 the inference to tide-related sampling risk at Qinglangang.

643 A decisive validation would combine several measurements that were unavailable
644 here. In situ water levels and local vertical control would test the changepoint directly;
645 optical FAPAR or absorbed-radiation measurements would separate canopy absorption



646 from realized efficiency; and multi-year observations would permit temporally blocked
647 evaluation of the scaling ratio. Repeated Sentinel-2 acquisitions at contrasting tidal
648 stages, together with UAV, LiDAR, or waterline-constrained elevation models, would
649 improve spatial classification (Shih et al., 2023). Replication across mangrove types
650 would then show whether tidal information improves daily FAPAR and GPP estimates
651 beyond solar geometry and radiation quality alone. Until those tests are available, the
652 workflow is best used as a transparent quality-control framework rather than a universal
653 correction model.

654 **5. Conclusion**

655 Tidal state covaried with an LUE-derived apparent FAPAR proxy at Qinglangang:
656 exposed-period values were 9.8% higher than flooded-period values, and the estimated
657 changepoint occurred near -0.3 m MSL. Across 163 valid days, a site-calibrated
658 instantaneous-to-daily ratio of 0.83 reduced in-sample RMSE by 28% relative to the
659 unity-ratio baseline. This constant ratio did not encode tidal state, and its performance
660 was not evaluated outside the calibration record. MODIS captured part of the seasonal
661 variation but not the tidal contrast, while Sentinel-2 provided only directionally
662 compatible spatial context under substantial elevation and spatial-dependence
663 constraints. The defensible contribution is therefore a workflow for diagnosing tide-
664 related sampling uncertainty, not a transferable tidal correction. Independent optical
665 FAPAR, measured water levels, blocked temporal validation, and replication across
666 mangrove settings are required before the approach can support operational adjustment
667 of satellite observations.



668 **Code availability**

669 The code that support the findings of this study are available from the
670 corresponding author upon request.

671 **Data availability**

672 The data that support the findings of this study are available from the
673 corresponding author upon request

674 **Author Contributions**

675 Demei Zhao: Writing – original draft, Visualization, Software, Methodology,
676 Investigation, Formal analysis. Haolin Chen: Writing – review & editing, Supervision,
677 Methodology, Funding acquisition. Wenmeijun Wang: Writing – review & editing,
678 Methodology, Data curation. Wenhao Huang: Writing – review & editing, Supervision,
679 Methodology. Kangwen Zhu: Funding acquisition, Data curation.

680 **Competing interests**

681 The authors declare that they have no conflict of interest.

682 **Acknowledgments**

683 This work was supported by National Natural Science Foundation of China
684 (42401417), China Postdoctoral Science Foundation Project (2025MD774150),
685 Guangdong Major Project of Basic and Applied Basic Research (2023B0303000017),
686 Shenzhen Science and Technology Program (JCYJ20241202124510015,
687 JCYJ20250604182203005) and Guangdong Basic and Applied Basic Research
688 Foundation (2025A151551008).

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