



Temperature and light regulated patterns of physiology, morphology and elemental stoichiometry in geographically distinct isolates of a cosmopolitan diatom

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10 **Abstract.** Anthropogenic influence on climate change has profound and diverse consequences on marine ecosystems. At the
base of the food web, phytoplankton, are experiencing altered temperature regimes. In south-east Australia, the southward
extension of warm waters, driven in part by the East Australian Current (EAC), is rapidly warming regional ocean
temperatures, leading to the intensification of marine heat waves (MHWs). In this study, we investigated thermally adapted
Leptocylindrus danicus strains isolated from four distinct latitudes to determine how silica production rates vary with
15 temperature and irradiance. We also explored how the intra-specific phenotypic variability affects physiology and silica
production. We found strong latitudinal effects on strain-specific cell volume (ranging from $313 \pm 22 \mu\text{m}^2$ to $2070 \pm 105 \mu\text{m}^3$)
and pigment quotas (chl *a* 1.04 ± 0.21 to $3.70 \pm 1.17 \text{ pg cell}^{-1}$; chl *c* 0.26 ± 0.07 to $2.09 \pm 1.50 \text{ pg cell}^{-1}$), both increasing with
increasing temperature. There was also a significant effect of temperature on silicification rates, which varied depending on
growth irradiance and cell normalisation. By identifying temperature and light regulated shifts in growth, morphology and
20 silicification in a cosmopolitan diatom, we can gain an improved understanding of the range in intraspecific phenotypic
variability of this key phytoplankton group. This study provides an assessment on how key diatom traits vary along a latitudinal
gradient, providing unique insight into how ocean warming may influence resilience and adaptation potential of *L. danicus*,
and how shift in physiology may impact diatom-regulated carbon and silicon cycling.



1. Introduction

Ocean warming is recognised as one of the major anthropogenic environmental stressors threatening marine ecosystem structure and functioning (Bindoff et al., 2013; Doney et al., 2012). It is well understood that temperature controls phytoplankton physiology and metabolic processes, and therefore defines the biogeographical range and distribution of a species (Boscolo-Galazzo et al., 2018; Finkel et al., 2010; Thomas et al., 2012). The Eastern Australian Current (EAC) is a dominant oceanographic feature of the east coast of Australia, and is identified as a future ‘hotspot’ for ocean warming, with models predicting up to an 8°C warming trend for the 2001–2050 time-period examined (Hobday & Pecl, 2014; Popova et al., 2016). As the EAC continues to strengthen with climate change due to a ‘spin-up’ of the South Pacific gyre, warm, nutrient-poor water is being driven southward along the east coast and is combined with ocean temperatures already increasing at a faster rate than projected, and as such, these trends are likely to have consequences for ecological processes and communities (Cai et al., 2005; Johnson et al., 2011; Phillips et al., 2020). Furthermore, warmer surface waters can cause a shallowing of the surface mixed layer, forming a barrier to vertical exchange of nutrients from depth and constraining phytoplankton in the upper surface waters, exposing them to high irradiances and reduced nutrient availability (Carmeño et al., 2008; Li et al., 2020). Therefore, understanding species-specific responses to temperature is crucial for predicting changes in phytoplankton abundance and community composition and how this will affect primary productivity, food web interactions, biogeochemical cycles and the biological pump.

Diatoms (Bacillariophyta) are siliceous organisms, responsible for an estimated 20% of global primary productivity (Field et al., 1998; Nelson et al., 1995). They are of significant ecological importance and recognised as one of the most diverse phytoplankton groups in the marine environment (Malviya et al., 2016). Their unique silicified cell walls (frustules) differentiates them from other phytoplankton and is the morphological feature that confers their central importance to global biogeochemical cycling of carbon and silicon in the oceans (Field et al., 1998; Martin-Jézéquel et al., 2000; Nelson et al., 1995). Diatoms are ubiquitously distributed, but tend to dominate dynamic coastal upwelling regions of the ocean where favourable conditions permit for the generation of blooms (Benoiston et al., 2017; Malviya et al., 2016; Smetacek, 1999). During these blooms, diatoms are major drivers of biological carbon sequestration in the oceans, however pronounced differences in growth, size, silicification and grazing pressure means that species abundance and composition is a determining factor of silica and carbon export in any given location at any point in time (Baines et al., 2010; Krause et al., 2010; Tréguer et al., 2018).

Leptocylindrus danicus is a cylindrical, chain-forming diatom with cosmopolitan distribution, and a major component of phytoplankton communities in coastal upwelling regions (Nanjappa et al., 2013). Along the south-east coast of Australia, *L. danicus* often dominates the coastal communities and is seen in greatest abundance during the spring (Ajani, et al., 2016). The continued success of *L. danicus* in these regions may be attributed to a combination of their widespread distribution (suggesting a broad thermal tolerance), rapid growth rates, large cell size (reduced pressure and predation by grazers) and life cycle strategy (reproduce both asexually and sexually via the formation of auxospores and resting spores), giving them a competitive



advantage to survive and adapt with each spring bloom due to their diverse phenotypic and genetic variability (Ajani et al., 2021; Ajani et al., 2016; Nanjappa et al., 2013).

60 Diatoms incorporate soluble silica from seawater, in the form of silicic acid, and transport it across their cell membrane into intracellular pools via silica deposition vesicles (SDVs) for frustule (cell wall) biosynthesis (Thamatrakoln & Hildebrand, 2008). This process, however, has been shown to vary with silicic acid concentrations in the environment, which coincide with changes in temperature and seasons, in turn influencing diatom growth because of its direct influence on their cell cycle (Hildebrand & Lerch, 2015; Shrestha et al., 2012; Shrestha & Hildebrand, 2015). Temperature-driven responses in
 65 physiological, morphological and biochemical traits were revealed using thermal performance curves (TPCs) on the diatom *Thalassiosira pseudonana*, and most notable of these trait changes was a reduction in silicification at higher temperatures (Baker et al., 2016; Sheehan et al., 2020). Previous work on *L. danicus* strains grown at a fixed temperature (18°C), on the other hand, demonstrated a latitudinal effect on silicification, whereby higher silicification rates were observed for strains isolated from lower latitudes compared to those from higher latitudes, likely due to strain-specific differences in thermal
 70 performance (Ajani et al., 2021). As such, while it is well known that temperature affects phytoplankton physiology and metabolic processes, there is a need to understand strain variability and the potential role of geographical adaptation in the phenotypic plasticity of individuals within a population in the context of ocean warming (Bennett et al., 2019). Here, we aim to identify temperature regulated patterns in silicification in multiple strains of *L. danicus*, and look at the intraspecific variability and thermal plasticity as a means of forecasting species-specific responses to warming. By looking at *L. danicus*
 75 strains with locally adapted thermal niches, we aim to provide new insights into how ocean warming may influence a species' growth and equally how diatom phenotypic plasticity provides resilience to rapid environmental change.

2. Materials and Methods

2.1 *Leptocylindrus danicus* isolates

80 Clonal culture isolates of the cosmopolitan diatom *Leptocylindrus danicus* used in this study were originally collected and established via single-cell isolation by Dr. P. Ajani between December 2015 to February 2016 (Table 1). For the present laboratory latitudinal study, strains were selected from four thermally distinct latitudes along the south-eastern Australian coastline, spanning from the north coast to south coast of New South Wales: Coffs Harbour (30.2986° S, 153.1094° E), Forster (32.1796° S, 152.5118° E), Maroubra (33.9495° S, 151.2437° E) and Twofold Bay (37.0844° S, 149.9277° E) (Table 1; Fig.
 85 1). Prior to acclimation for experiments (described below), the established *L. danicus* strains from Coffs Harbour, Maroubra and Twofold Bay were maintained in 33 ppt salinity fresh seawater (FSW, 0.2 µm filtered) enriched and autoclaved with f/2 medium (Guillard & Ryther, 1962). Batch cultures were maintained in 50 mL culture flasks at 18°C and illuminated at ~35 µmol photons m⁻² s⁻¹ on a 12 : 12 h light:dark cycle under white light. At the time of this study, the *L. danicus* Forster strain was discontinued from Dr. P. Ajani's culture collection, and instead retrieved from the Australian National Algae Culture
 90 Collection (ANACC), Hobart, Tasmania; CSIRO strain CS-1164 (Table 1). This strain was maintained in f/2 medium at 20°C,



with $\sim 60 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ illumination on a 12:12 h light:dark cycle using LED light panels. All cultures were transferred (1:100) into fresh medium at least every three weeks to maintain actively growing cells.

Table 1: List of *Leptocylindrus danicus* strains used in this study: originating *L. danicus* sampling locations along the coastline of New South Wales, strain designation, collection dates, mean annual and spring sea surface temperature (SST) ($^{\circ}\text{C}$) for the period 2015-19 at each location.

Location of isolation	Latitude, Longitude	Strain designation	Collection date	Mean annual SST	Mean spring SST
Coffs Harbour	30.2986° S, 153.1094° E	CH230116-5; synonym CS-1161	23 Jan 2016	22.58	21.30
Forster	32.1796° S, 152.5118° E	CS-1164; synonym FOS180216-7	18 Feb 2016	21.67	20.40
Maroubra	33.9495° S, 151.2437° E	MAR091215-6	09 Dec 2015	20.85	19.51
Twofold Bay	37.0844° S, 149.9277° E	TF250116-5	25 Jan 2016	18.02	16.33

Annual and spring mean SST calculated from monthly data at $1/12^{\circ}$ spatial resolution from 2015-19 (Jean-Michelet et al., 2021).

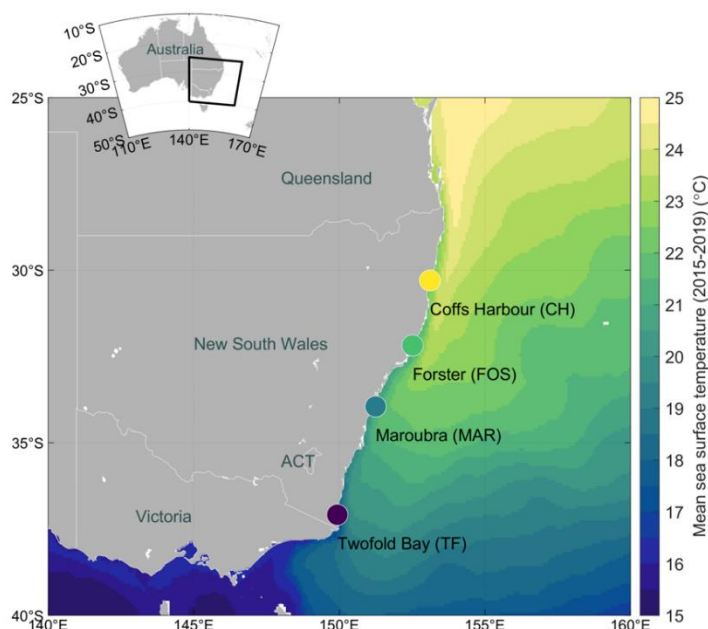


Figure 1: Mean annual sea surface temperature (SST) map with sampling locations. Locations sampled on the coastline of south-eastern Australia for *Leptocylindrus danicus* strain isolates used in the current latitudinal study, from the north to south coast of New South Wales: Coffs Harbour (CH), Forster (FOS), Maroubra (MAR) and Twofold Bay (TF) are shown. Sampling locations are overlayed on an annual mean SST map calculated from monthly data at 1/12° spatial resolution from 2015-19 (Jean-Michel et al., 2021). Colour bar shows temperature (°C).

2.2 Experimental design and culture maintenance

For the experiment, *Leptocylindrus danicus* strains from each of the four locations were acclimated and grown in locally sourced offshore sea water (0.2 µm filtered and autoclaved) enriched with f/10 medium (Guillard & Ryther, 1962) and modified with the addition of silicate (acidified with 0.5 M HCl in a 1.5:1 ratio). Cultures were maintained under controlled light and temperature conditions in Plant Growth Cabinets (CLIMATRON-1100-SL-H, Thermoline Scientific) fitted overhead with warm white, fluorescent lights. Light was supplied on a 12:12 h light:dark cycle, with cultures grown under two light intensities provided at approximately 100 and 55 µmol photons m⁻² s⁻¹ (photosynthetically active radiation; PAR, using a LI-COR Light Meter, LI-250A, equipped with a 2Pi Quantum Sensor). To achieve the lower light intensity, growth cabinets were fitted with neutral density filters (90% shading; LEE Filters). Cultures were grown at four different temperatures: 16, 18, 20, or 22°C (Table 2), selected based on the natural average spring sea surface temperature (SST) for the period 2015-19 at the approximate location of which the strains were isolated (see Table 1). Temperature was kept constant and monitored every 15 min throughout the acclimation and experimental period using iButton® temperature loggers (Thermochron) (Table 2).



At each light intensity, quadruplicate batch cultures (200 mL) of each strain were acclimated and grown for more than two months prior to measurement. Cultures were inoculated at ~ 5000 cells mL^{-1} into sterile f/10 medium and grown in 250 mL Erlenmeyer flasks at their respective temperatures and irradiances. Throughout the acclimation and experimental periods, all cultures were transferred into fresh media every 10-14 days under sterile (biosafety cabinet) conditions to maintain healthy growing cells. The Twofold Bay *L. danicus* cultures were unable to establish at $100 \mu\text{mol photons m}^{-2} \text{s}^{-1}$, and therefore, only low light ($55 \mu\text{mol photons m}^{-2} \text{s}^{-1}$) incubations were completed for this strain. All measurements were performed on cultures during their exponential growth.

Table 2: Study temperatures. Assigned target experimental temperature and mean measured temperature ($^{\circ}\text{C}$) \pm standard deviation (SD). Data were collected every 15 min throughout acclimation period and experiment using temperature loggers.

Location of isolation	Target temperature	Mean temperature	SD
Coffs Harbour	22	22.6	0.51
Forster	20	20.39	0.56
Maroubra	18	18.14	0.38
Twofold Bay	16	15.72	0.38

2.3 Cell enumeration, growth rates and morphology

To establish growth rates, aliquots of 1 mL were sub-sampled daily from each flask and fixed with glutaraldehyde (2% v/v final concentration). Cell counts were performed manually using a Sedgewick Rafter counting chamber by counting to a minimum of 400 cells as per Ajani *et al.* (2016), using an upright Nikon Eclipse Ci fluorescence microscope ($\leq 200\times$ magnification). Growth rates were monitored via light microscopy until the stationary phase was reached. Maximum average growth rates (μ_{Max} in divisions per day) for each strain were calculated during the exponential growth phase according to Guillard (1973).

Cell counts and cell volume measurements for data normalisation were taken on the day of experimental sampling for each strain when cultures were in exponential growth using a fixed 2 mL sub-sample (2% glutaraldehyde). Cells were imaged under a microscope (Nikon Eclipse Ci fluorescence microscope) at $200\times$ magnification using the Infinity Analyze 7 software, (version 7.1.0.1215; Teledyne Lumenera). To estimate cell sizes for each strain and replicate, the first 50 cells encountered were measured using the approximate cell diameters (width) and length, in the image processing package Fiji (Fiji is Just ImageJ) (Schindelin *et al.*, 2012). Cell volume was then calculated assuming cylindrical cells, as per Hillebrand *et al.* (1999).

2.4 PDMPO labelling protocol and sampling



To assess biogenic silica (bSi) content and silica incorporation rates, exponentially growing cultures were harvested and inoculated in quadruplicate at $10,000 \text{ cells mL}^{-1}$ into new 250 mL Falcon Tissue Culture Flasks in 155 mL (final volume) of filtered seawater amended with f/10 media. Cultures were inoculated with the fluorescent stain PDMPO (LysoSensor Yellow/Blue DND-160; ThermoFisher Scientific, Australia) to a final concentration of $0.125 \mu\text{M}$ (Leblanc & Hutchins, 2005) and incubated under experimental conditions for 24 h. Following 24 h incubation, labelled cultures were sampled for cell density, bSi content, PDMPO incorporation, and measurement of the photosynthetic health of cells (F_v/F_m).

To determine bSi content and silica incorporation rates from PDMPO incubations, 100 mL aliquots from each incubation flask were filtered under low vacuum onto a 47 mm, $0.4 \mu\text{m}$ Nucleopore Polycarbonate Track-Etch (PCTE) Membrane filter (Cytiva, Whatman). Filters were rinsed with $0.2 \mu\text{m}$ FSW (1 mL, three times) to remove any residual and unbound PDMPO. Using sterile tweezers, filters were folded into quarters and placed toward the bottom of a 1.8 mL cryovial. Immediately after sampling, samples were flash frozen in liquid N_2 and stored at -80°C until further processing. Samples were prepared for processing by first thawing and removing filters from cryovials, unfolded onto a filter tower, and covered in 2 mL of 10% HCl for 2 min to disrupt cell membranes. To further osmotically lyse cells and remove any unbound PDMPO and cellular debris, filters were subsequently rinsed with milli-Q water (1 mL, three times). HCl washed filters were then transferred to 15 mL Falcon tubes to digest frustules in a NaOH-HCl matrix for analyses using established methods (Leblanc & Hutchins, 2005). Briefly, 8 mL of 0.2 M NaOH was added to sample tubes before digestion in a hot water bath ($\sim 95^\circ\text{C}$) for 3 h. Samples were cooled in an ice bath and neutralised by adding 2 mL of 0.8 M HCl and mixed well. Filters were removed from sample tubes before centrifugation at $4000 \times g$ for 5 min to pellet any remaining debris. Samples were then split for processing of bSi and PDMPO analyses.

Bsi production in samples were determined by withdrawing 2.5 mL of supernatant for analysis using the colorimetric analysis of reactive silicate, as in Strickland & Parsons (1972). To ensure samples fell within the range of the standard curve, samples (supernatant) were diluted with milli-Q water (2.5 mL final volume) prior to the addition of reagents for bSi content assays on both high light (1:10 dilution) and low light (1:2 dilution) samples. Absorbance was measured on a UV-VIS Spectrophotometer (Shimadzu UV-1280) at 810 nm and compared against a standard curve ($R^2 \geq 0.997$) prepared with a sodium metasilicate stock (10 mM) to a maximum concentration of $200 \mu\text{mol L}^{-1}$. Another 2.5 mL supernatant was withdrawn for measurement of PDMPO in the solution using a spectrofluorometer (Shimadzu RF-6000), equipped with a xenon lamp. Excitation was set at 380 nm (5 nm bandwidth) and emission peaks measured at 542 nm (10 nm bandwidth) with a 200 nm/min scanning speed. A standard curve ($R^2 \geq 0.998$) was prepared using a 1/1000 dilution of the original PDMPO stock culture (1 mM) for calibration standards (i.e. $1 \mu\text{M}$ stock concentration) in the NaOH-HCl digestion matrix to a maximum concentration of $0.1 \mu\text{mol L}^{-1}$. All PDMPO data were converted to bSi production using the recommended ratio of 2916 mol bSi per mol PDMPO (McNair et al., 2015) to enable quantitative comparisons. All measurements were made using a quartz cuvette (10 mm light path), results corrected against blanks and normalised to cell density and surface area.

2.5 Macronutrient determination



To estimate initial concentrations of cellular uptake and drawdown of phosphate and nitrate/nitrite (NO_x) following 24 h PDMPO incubations, filtrate from each replicate was collected from experimental flasks (30 mL) and PDMPO incubation flasks (100 mL) using 0.2 µm PCTE filters. Macronutrients were analysed using the methods of Ringuelet et al. (2011) for phosphate assays, and NO_x assays based on Schnetger & Lehnert (2014). Briefly, phosphate assays were prepared by adding 200 µL of sample to a 96 Well Tissue Culture Plate with 50 µL reducing agent and gently vortexed on a microplate shaker for 30 min. Absorbance was measured at 880 nm and results compared to a standard curve ($R^2 \geq 0.99$) prepared with a potassium phosphate monobasic (KH₂PO₄) stock solution. NO_x (NO₂⁻ + NO₃⁻) assays were prepared by adding 180 µL of samples to a 96 Well Tissue Culture Plate with 120 µL reducing agent and incubated at 45°C for 60 min. Absorbance was measured at 540 nm and results compared to a standard curve ($R^2 \geq 0.99$) prepared with a potassium nitrate (KNO₃) stock solution. All samples were conducted using triplicate technical samples and absorbance measurements made on a TECAN Spark® Multimode Microplate Reader (Spark Control version 3.1).

2.6 Photophysiology

The photo-physiological condition of exponentially growing *L. danicus* cultures were analysed via a multi-step (eight light levels) steady state light curve, using a pulse amplitude modulated (PAM) fluorometer (Water PAM, Walz GmbH, Effeltrich, Germany) following the protocol outlined in Petrou et al. (2011). Prior to analyses, live subsamples (2 mL) were dark-adapted under subdued light for 15 min. Following dark acclimation, a saturating pulse of light for 0.8 s with an intensity of 8 was applied for determination of minimum and maximum fluorescence. To estimate the efficiency of photosystem II (PSII), a series of saturating pulses (30 sec apart) of light under actinic illumination (4 min) at eight intensities for both high light (72, 107, 163, 240, 360, 540, 764, 1050 µmol photons m⁻² s⁻¹) and low light adapted cultures (48, 72, 107, 163, 240, 360, 540, 764 µmol photons m⁻² s⁻¹) were conducted, and the average of the last three measurements at each light step recorded. Here, the minimum (F_o') and maximum fluorescence (F_m') in the light-adapted state were obtained and the maximum quantum efficiency of PSII (F_v/F_m) was calculated, where: $F_v/F_m = (F_m - F_o)/F_m$. The resulting fluorescence parameters: effective quantum yield of PSII [$Y(II)$] and electron transport rate (ETR) were calculated automatically from the light response curves by the Water PAM (Walz) in the WinControl-3 software. Using a custom-made macro in SigmaPlot version 14.5 (Systat Software, Inc.), a curve fitting model was applied to rETR as a function of irradiance (PAR; µmol photons m⁻² s⁻¹), and the maximum relative electron transport rate (rETR_{max}), light utilisation efficiency (alpha; α) and minimum saturating irradiance (I_k) calculated as in Ralph & Gademann (2005). To determine whether PDMPO negatively affected cells, the maximum quantum efficiency of PSII (F_v/F_m) was determined before and following 24 h incubations.

2.7 Chlorophyll *a* and *c* determination

To determine chlorophyll *a* and *c* ($c_1 + c_2$) content, 50 mL aliquots from each replicate of exponentially growing culture were filtered onto 25 mm Whatman GF/F filters (0.7 µm nominal pore size) under low vacuum (≤ 15 mm Hg). Filters were placed at the bottom of 15 mL Falcon tubes and immediately stored at -30°C until further processing. Exposure of chlorophyll extracts



to light was avoided to evade degradation of products. Pigments were extracted from cells by adding 3 mL of 90% acetone (v/v) to thawed samples, vortexed, filters pressed to the bottom of the tube, then allowed to stand in a refrigerator in the dark for at least 15 h at 4°C. Following overnight extraction, samples were vortexed again, filters removed, then centrifuged at 4000 × g for 5 min at 4°C to remove any remaining cellular debris. Chlorophyll content from each sample was determined by loading 2 mL of supernatant into a quartz cuvette (10 mm light path) and absorbance measured at 630 nm and 664 nm using a UV-VIS spectrophotometer (Shimadzu, UV-1280). Prior to sample measurements, a 90% acetone blank was run to correct for background fluorescence. Chlorophyll *a* and *c* concentrations were calculated ($\mu\text{g mL}^{-1}$) using the equations of Jeffrey & Humphrey (1975), later refined by Ritchie (2006), then normalised to cell density and cell volume.

2.8 Particulate organic carbon (POC) and nitrogen (PON)

For POC/PON determination, samples (100 mL) of exponentially growing cultures (one per biological replicate) were collected by vacuum filtration (<20 mm Hg) onto pre-combusted (450°C for 5h) 47 mm Whatman GF/F filters (0.7 μm nominal pore size). Individual filters were then wrapped in pre-combusted aluminium foil and immediately stored at -30°C until further processing. Filters were prepared for analysis by first thawing, then acidifying with 1M HCl dropwise until the filter was saturated. Acidified filters were then left in a drying oven at 60°C for at least 48 h before being cut into quarters, wrapped and pelletised in 9 mm tin capsules (Elemental Microanalysis, UK). To quantify and correct for background concentrations, four ‘blanks’ (filters with 100 mL of f/10 medium only) were also collected onto pre-combusted GF/F filters and processed, as above. Carbon and nitrogen analyses were conducted by the Biogeochemical Stable Isotope Facility at the University of Hawai‘i at Mānoa using a Costech Elemental Combustion System with a Zero Blank Autosampler (ECS 4010) coupled with a Thermo Scientific Delta V Advantage, connected through a Thermo Scientific Conflo IV isotope ratio mass spectrometer.

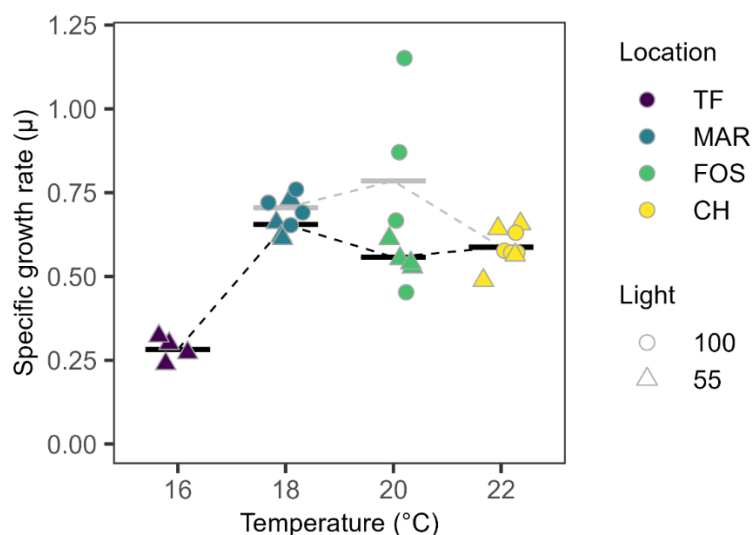
2.9 Data analyses

All figures and data analyses were performed in RStudio version 2024.12.0+467 (R Core Team, 2024). Morphological and physiological trait data were tested for temperature-driven responses using linear regression. When no significant linear relationship was found, a one-way analysis of variance (ANOVA) was used to test for significant differences in phenotypic traits between strains at each light intensity. To ensure assumptions of normality and equal variance were satisfied, a Levene’s test for homogeneity of variance was applied to all data, *a priori*. When treatment groups varied significantly, a Tukey’s honestly significant difference (HSD) post hoc was used to isolate which groups differed significantly from each other. In cases where the assumption of equal variances was violated, a Bartlett’s test was performed, followed by a Welch’s ANOVA. If the *F* ratio was found to be significant, a Games-Howell post hoc test was performed to determine which treatment groups differed. Statistical significance was determined at $\alpha = 0.05$ for all samples. All figures were generated using add-on package ggplot2 v. 3.5.1 (Wickham, 2016).



3. Results

265 Growth rates did not differ with growth irradiance for *L. danicus* strains acclimated to 18-22°C (Fig. 2). Under high light conditions, we saw a significant effect of temperature on mean specific growth rate (μ) between *L. danicus* strains ($F_{2, 5.023} = 8.688, P = 0.023$) where the Maroubra strains grew significantly faster ($P = 0.018$) than the Coffs Harbour strains; $0.71 \pm 0.05 \text{ d}^{-1}$ and $0.59 \pm 0.03 \text{ d}^{-1}$, respectively (Fig. 2). Under low light conditions, we saw a significant temperature-dependent response ($F_{3, 12} = 37.13, P < 0.001$) with significantly lower growth rates recorded for Twofold Bay strains (16°C ; $0.28 \pm 0.03 \text{ d}^{-1}$) compared with the other strains ($P < 0.001$; Fig. 2).



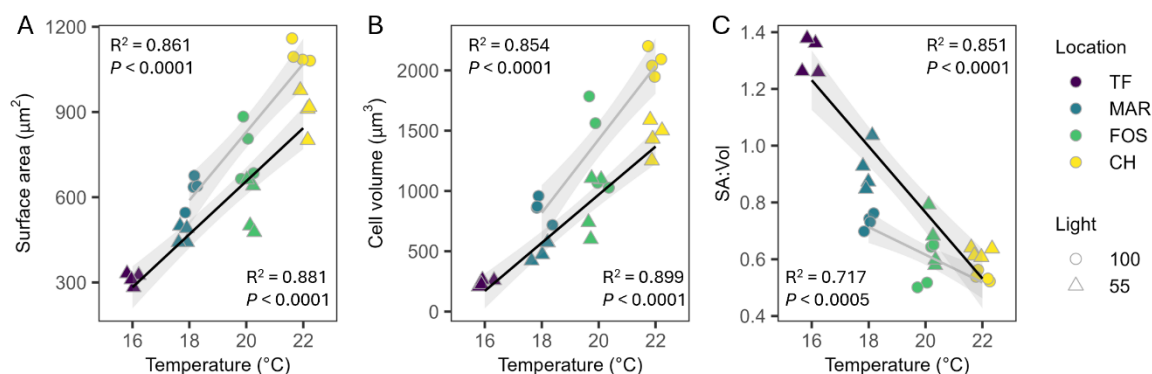
275 **Figure 2: Temperature and light dependent growth rates.** Maximum specific growth rates (μ_{Max} in divisions per day) of *Leptocylindrus danicus* strains in response to temperature under $100 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (circles) and $55 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (triangles). Data are individual replicates ($n = 4$) and bars represent their mean value, stippled lines between means are there as a visual guide.

We saw a clear latitudinal pattern with cell size measurements during exponential growth in both light treatments (Fig. 3A, B). Mean cell surface area under high light, ranged from $624 \pm 55 \mu\text{m}^2$ for Maroubra strains to $1104 \pm 37 \mu\text{m}^2$ for Coffs Harbour strains ($R^2 = 0.861, P < 0.0001$; Fig. 3A), while low light cell volume ranged from $313 \pm 22 \mu\text{m}^3$ for Twofold Bay strains to $902 \pm 73 \mu\text{m}^3$ for Coffs Harbour strains ($R^2 = 0.881, P < 0.0001$; Fig. 3A). Mean cell volume more than doubled from 18 to 22°C under high light, from $853 \pm 99 \mu\text{m}^3$ for Maroubra strains to $2070 \pm 105 \mu\text{m}^3$ for Coffs Harbour strains ($R^2 = 0.854, P < 0.0001$; Fig. 3B). For low light strains, there was a six-fold increase in mean cell volume across the temperature range



from $239 \pm 27 \mu\text{m}^3$ for Twofold Bay strains to $1444 \pm 143 \mu\text{m}^3$ for Coffs Harbour strains ($R^2 = 0.899$, $P < 0.0001$; Fig. 3B). An inverse response was observed in the surface area to volume ratio, which was negatively correlated with increasing temperature for both high ($R^2 = 0.717$, $P = 0.0005$; Fig. 3C) and low light treatments ($R^2 = 0.851$, $P < 0.0001$; Fig. 3C). We also detected a significant interaction ($P = 0.0025$) with growth irradiance and strain (temperature), where the difference in SA:V diminishes at higher light (Fig. 3C).

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Figure 3: Cell size and physiological trait variability. **A** Surface area (μm^2), **B** cell volume (μm^3), **C** surface area to volume ratio (SA: Vol) of *Leptocylindrus danicus* strains in their respective temperatures, under high (circles) and low (triangles) light. Data are independent replicates ($n = 4$). The lines show the linear regression and the grey shaded area represent 95% confidence intervals.

Steady-state light curves revealed temperature differentiation in effective quantum yield ($Y(II)$) and relative electron transport rates ($rETR$), which were amplified at low irradiance (Fig. 4A-D). Under high growth irradiance, maximum relative electron transport rates ($rETR_{\text{max}}$) declined with increasing temperature ($F_{2,9} = 8.312$, $P = 0.009$; Fig. 4E), being significantly higher in Maroubra strains compared to Coffs Harbour strains ($P = 0.007$; Fig. 4E). Under low light conditions, both Coffs Harbour and Forster strains demonstrated a significantly higher $rETR_{\text{max}}$ compared to Maroubra and Twofold Bay strains ($F_{3,5.4} = 122.09$, $P < 0.001$; Fig. 4F). Light utilisation efficiency (α) differed with temperature under high ($F_{2,9} = 12.88$, $P = 0.002$; Fig. 4G) and low light ($F_{3,5.87} = 84.376$, $P < 0.001$; Fig. 4H) conditions, but without a clear temperature trend. Minimum saturating irradiances (I_k) mirrored $rETR_{\text{max}}$, varying significantly between strains under high ($F_{2,9} = 4.781$, $P = 0.039$; Fig. 4I) and low light ($F_{3,5.37} = 56.145$, $P < 0.001$; Fig. 4J).

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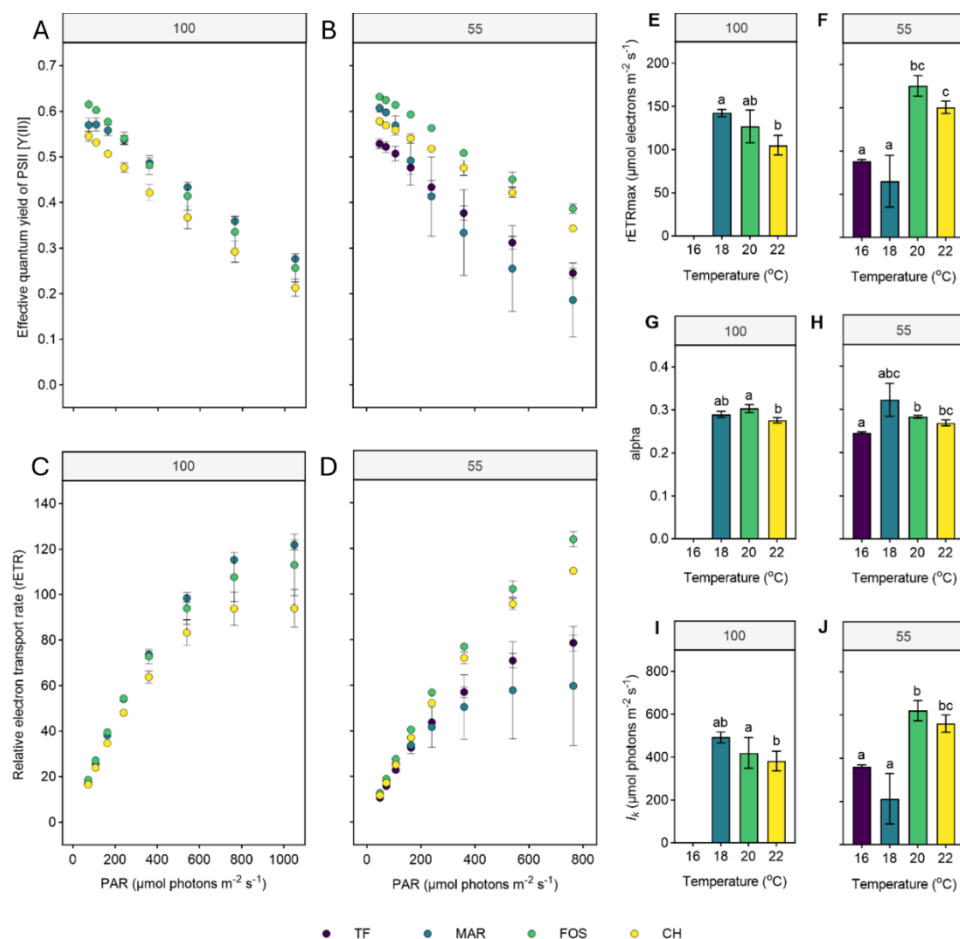
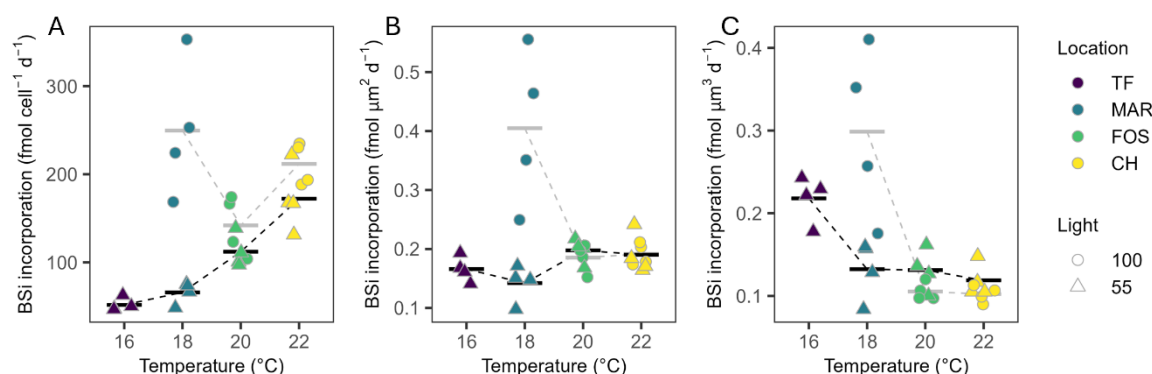


Figure 4: Chlorophyll fluorescence parameters. A, B Effective quantum yield of PSII [Y(II)], C, D relative electron transport rates (rETR) as a function of irradiance (PAR), E, F relative maximum electron transport rate (rETR_{max}), G, H light utilisation efficiency (alpha; α) and I, J minimum saturating irradiance (I_k) determined for *Leptocylindrus danicus* strains acclimated to temperature conditions under each light treatment. Data represent the mean of four biological replicates \pm SD. Different letters indicate significant differences between strains ($\alpha = 0.05$).

Silicification rates (bSi incorporation rate measured from PDMPO incorporation) per cell differed significantly across strains under high light ($F_{2,9} = 4.651$, $P = 0.041$), where at 18°C, Maroubra cells silicified more ($249.69 \pm 77.33 \text{ fmol cell}^{-1} \text{ d}^{-1}$) than Forster strains at 20°C ($141.98 \pm 33.77 \text{ fmol cell}^{-1} \text{ d}^{-1}$; Fig. 5A). Conversely, there was a significant temperature effect on silicification rates under low light, with cells silicifying more with increased temperature ($R^2 = 0.804$, $F_{1,14} = 57.27$, $P < 0.0001$; Fig. 5A). Normalised to surface area, however, we saw a significant temperature-dependent response on silicification in high light exposed strains ($F_{2,9} = 10.05$, $P = 0.005$), with the highest silicification seen at the lowest temperature in Maroubra strains ($0.4 \pm 0.13 \text{ fmol } \mu\text{m}^2 \text{ d}^{-1}$), while both Forster and Coffs Harbour strains silicified less than half this amount at $0.19 \pm 0.02 \text{ fmol}$



$\mu\text{m}^2 \text{d}^{-1}$ (Fig. 5B). Under low irradiance, the temperature-dependent relationship seen in the per cell data was less pronounced, with rates ranging from $0.14 \pm 0.03 \text{ fmol } \mu\text{m}^2 \text{d}^{-1}$ to $0.20 \pm 0.13 \text{ fmol } \mu\text{m}^2 \text{d}^{-1}$ (Fig. 5B). When normalised to cell volume, bSi showed a similar pattern for both high and low light cultures (Fig. 5C), with a significantly higher incorporation rate in the Maroubra strain grown at high irradiance ($F_{2,9} = 13.90$, $P = 0.0018$) compared with the others, and a higher incorporation rate for the Twofold Bay strain ($F_{3,12} = 10.80$, $P = 0.001$), when compared with the other strains grown at low light (Fig. 5C). We found no negative effect of PDMPO addition on the F_V/F_M , confirming no potential toxicity of the stain at the concentration used (Table 3).



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Figure 5: Silicification rates. Biogenic silicate (bSi) incorporation from 24 h incubations, normalised to **A** cell density ($\text{fmol cell}^{-1} \text{d}^{-1}$), **B** surface area ($\text{fmol } \mu\text{m}^2 \text{d}^{-1}$), and **C** cell volume (μm^3) of *Leptocylindrus danicus* strains for high (circles) and low (triangles) light treatment. Data are individual measurements ($n = 4$), and horizontal bars represent their mean value. Stippled lines between means are provided as a visual guide only.

335 Total carbon per cell was differed across strains grown under high light, showing an increase with increasing temperature ($F_{2,9} = 14.78$, $P = 0.001$), where both Forster and Coffs Harbour strains had significantly higher carbon content ($61.60 \pm 11.11 \text{ pg cell}^{-1}$ and $72.58 \pm 8.24 \text{ pg cell}^{-1}$, respectively) than Maroubra strains ($31.31 \pm 13.41 \text{ pg cell}^{-1}$; Fig. 6A). Under low light, mean carbon content varied significantly across strains ($F_{3,12} = 16.06$, $P < 0.001$), with higher total carbon seen in Forster strains ($78.81 \pm 20.69 \text{ pg cell}^{-1}$) compared to the strains from Twofold Bay ($17.40 \pm 1.72 \text{ pg cell}^{-1}$), Maroubra ($32.18 \pm 4.48 \text{ pg cell}^{-1}$), and Coffs Harbour ($31.23 \pm 16.25 \text{ pg cell}^{-1}$; Fig. 6A). Cell specific nitrogen content showed a similar pattern to carbon, with higher nitrogen content in both Forster ($14.62 \pm 0.57 \text{ pg cell}^{-1}$) and Coffs Harbour ($14.05 \pm 0.90 \text{ pg cell}^{-1}$) strains under high light ($F_{2,9} = 104.1$, $P < 0.001$; Fig. 6B). Likewise, mean nitrogen content was greatest under warmer conditions under low light ($F_{3,12} = 73.73$, $P < 0.001$; Fig. 6B), with higher values for Forster ($19.46 \pm 2.38 \text{ pg cell}^{-1}$) and Coffs Harbour strains ($14.92 \pm 2.22 \text{ pg cell}^{-1}$), compared with strains from Maroubra ($6.90 \pm 0.74 \text{ pg cell}^{-1}$) and Twofold Bay ($3.85 \pm 0.26 \text{ pg cell}^{-1}$). Mean carbon to nitrogen ratios (C:N) showed no significant difference or latitudinal trend between strains in the high light

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treatment (Fig. 6C), whereas low light C:N differed across strains ($F_{3,12} = 19.14$, $P < 0.001$), with significantly lower C:N for the Coffs Harbour strain (Fig. 6B).

Table 3: Maximum quantum yield of photosystem II (F_V/F_M). Comparison of initial (T0) and final (T24; following 24 h incubations) F_V/F_M of *Leptocylindrus danicus* strains for all temperature and light treatments. Data represent the mean of four biological replicates \pm SD. Bold values denote statistical significance of F_V/F_M following 24 h incubations within strains ($\alpha = 0.05$).

Light treatment ($\mu\text{mol photons m}^{-2} \text{ s}^{-1}$)	Temperature ($^{\circ}\text{C}$)	Strain isolation location	T0	T24
100	22	Coffs Harbour	0.651 ± 0.031	0.670 ± 0.014
	20	Forster	0.688 ± 0.011	0.718 ± 0.014
	18	Maroubra	0.644 ± 0.006	0.666 ± 0.006
	16	Twofold Bay	n.d.	n.d.
55	22	Coffs Harbour	0.663 ± 0.011	0.688 ± 0.020
	20	Forster	0.716 ± 0.019	0.693 ± 0.059
	18	Maroubra	0.686 ± 0.009	0.694 ± 0.007
	16	Twofold Bay	0.583 ± 0.012	0.627 ± 0.009

Chlorophyll content followed a latitudinal trend at which chlorophyll *a* and *c* pigments increased with increasing temperature under both light treatments (Table 4). Under high light, we saw significantly higher concentrations of chlorophyll *a* per cell at Coffs Harbour than the other two sites ($F_{2,5.211} = 9.604$, $P = 0.018$). Significantly higher chlorophyll *a* content per cell was also seen for lower latitude strains exposed to low light ($F_{3,12} = 18.41$, $P < 0.001$). Chlorophyll *c* per cell followed the same latitudinal trend with higher mean chlorophyll *c* concentrations in Coffs Harbour strains ($2.09 \pm 1.50 \text{ pg cell}^{-1}$) than Maroubra strains ($0.26 \pm 0.07 \text{ pg cell}^{-1}$; $P = 0.037$) (Table 4), and higher chlorophyll *c* per cell in the Coffs Harbour strains ($0.98 \pm 0.17 \text{ pg cell}^{-1}$) compared to all other latitudes grown under low light ($F_{3,12} = 23.24$, $P < 0.001$; Table 4). When normalised to cell volume, mean chlorophyll *a* concentration differed across strains ($F_{2,9} = 4.775$, $P = 0.039$), with significantly more chlorophyll *a* per unit volume in Coffs Harbour strains ($1.78 \pm 0.53 \text{ fg } \mu\text{m}^3$) than Forster strains ($0.97 \pm 0.18 \text{ fg } \mu\text{m}^3$; $P = 0.034$) under high light (Fig. 6D). Low light cultures showed a latitudinal trend ($F_{3,12} = 20.69$, $P < 0.001$), with higher chlorophyll *a* per volume in Twofold Bay strains ($4.62 \pm 0.83 \text{ fg } \mu\text{m}^3$) compared to all other strains, ranging between 1.91 to $2.69 \text{ fg } \mu\text{m}^3$ (Fig. 3D). Per volume mean chlorophyll *c* content did not differ across strains under high light (Table 4). However, in low light conditions, chlorophyll *c* per cell volume decreased with warmer temperatures ($F_{3,12} = 20.69$, $P < 0.001$; Table 4).



Table 4: Chlorophyll pigments. Chlorophyll *a* and *c* content normalised per cell (pg cell^{-1}) and per volume ($\text{fg } \mu\text{m}^3$) for *Leptocylindrus danicus* strains in response to temperature under each light treatment. Data represent the mean of four biological replicates \pm SD. Different letters indicate significant differences between strains within light condition ($\alpha = 0.05$).

Light treatment ($\mu\text{mol photons m}^{-2}$ s^{-1})	Temperature ($^{\circ}\text{C}$)	Strain isolation location	Chlorophyll <i>a</i> content		Chlorophyll <i>c</i> content	
			pg cell^{-1}	$\text{fg } \mu\text{m}^3$	pg cell^{-1}	$\text{fg } \mu\text{m}^3$
100	22	Coffs Harbour	3.70 ± 1.17^a	1.78 ± 0.53^a	2.09 ± 1.50^a	0.99 ± 0.66^a
	20	Forster	1.28 ± 0.15^b	0.97 ± 0.18^b	0.47 ± 0.07^{ab}	0.35 ± 0.04^a
	18	Maroubra	1.04 ± 0.21^b	1.24 ± 0.33^a	0.26 ± 0.07^b	0.30 ± 0.06^a
	16	Twofold Bay	n.d.	n.d.	n.d.	n.d.
55	22	Coffs Harbour	2.73 ± 0.35^a	1.91 ± 0.30^a	0.98 ± 0.17^a	0.69 ± 0.16^a
	20	Forster	2.33 ± 0.47^a	2.69 ± 0.33^a	0.62 ± 0.13^b	0.72 ± 0.10^a
	18	Maroubra	1.38 ± 0.39^b	2.67 ± 0.40^a	0.42 ± 0.11^{bc}	0.83 ± 0.24^{ab}
	16	Twofold Bay	1.10 ± 0.15^{bc}	4.62 ± 0.83^b	0.29 ± 0.07^{cd}	1.23 ± 0.28^b

Biogenic silica to carbon ratios (bSi:C) differed across strains ($F_{2,9} = 19.94$, $P < 0.001$), with the highest ratio for Maroubra (1.13 \pm 0.38) followed by the Coffs Harbour (0.35 \pm 0.04), and Forster (0.24 \pm 0.06) strains (Fig. 6E). In the low light acclimated strains, bSi: C was highest in Coffs Harbour strain (0.32 \pm 0.12) compared to the Twofold Bay (0.24 \pm 0.06) and Maroubra (0.15 \pm 0.04) strains, and lowest in the Forster (0.09 \pm 0.02) strain ($F_{3,12} = 8.63$, $P = 0.003$; Fig. 6E).

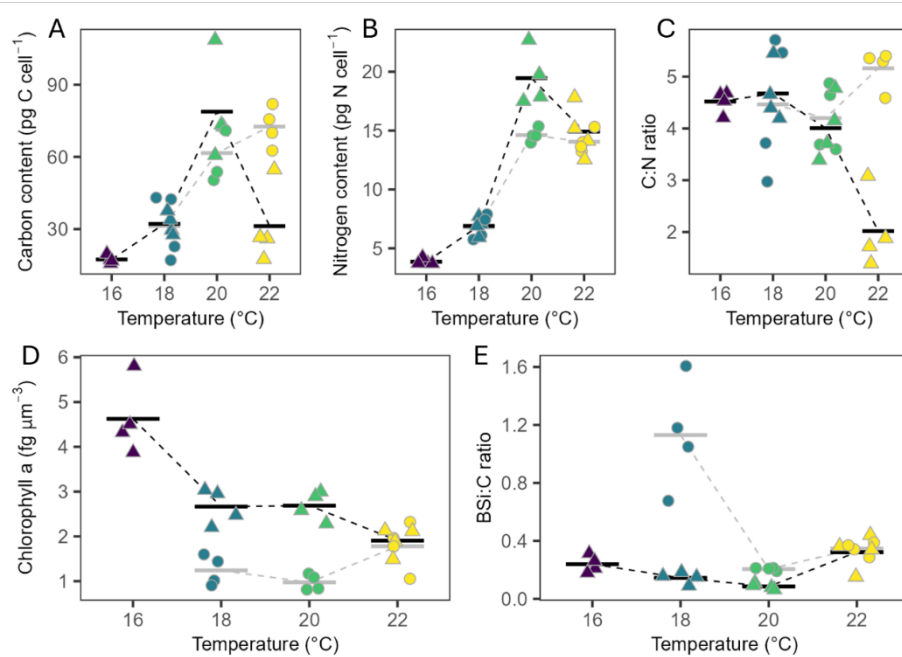


Figure 6: Elemental composition. **A** total carbon per cell (pg), **B** total nitrogen per cell (pg), **C** carbon to nitrogen ratio (C:N), **D** chlorophyll *a* per cell volume (fg μm⁻³), **E** carbon to biogenic silicate (bSi) ratio of *Leptocylindrus danicus* strains from high (circles) and low (triangles) light. Data are individual replicates (n = 4) and horizontal bars represent their mean value. Stippled lines between means are provided as a visual guide.

Principal component analysis (PCA) of physiological and morphological traits revealed strain separation associated with temperature and light, with PC1 explaining 37.6% of the variation, while PC2 explained 22.8% of variation (Fig. 7A). Cell size (SA:V) was anticorrelated with growth rate, having the greatest influence along PC1 (Fig. 7B). A key trait driving separation along PC2 was bSi:C, closely aligning with silicification rate, both of which were anti-correlated with photosynthetic efficiency (F_v/F_M).

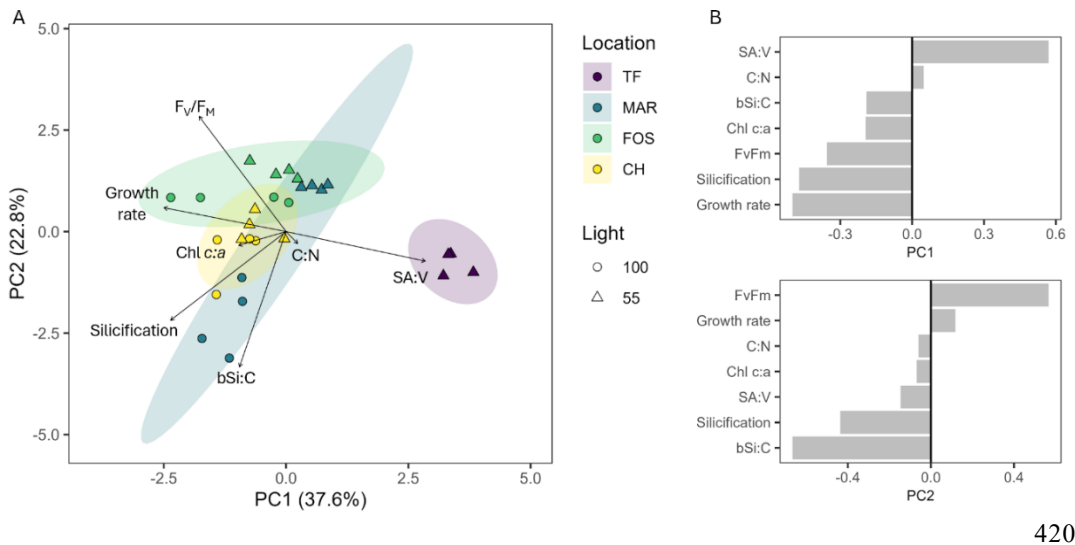


Figure 7: Principal component analysis and loadings. **A** PCA with vectors of parameters contributing to differences in *Leptocylindrus danicus* strains grown at 16, 18, 20 and 22°C under 100 and 55 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$; morphological (growth rate, cell surface area to volume ratio), and physiological traits (chlorophyll *a*, F_v/F_m , silicification rate, carbon to nitrogen ratio (C: N), carbon to biogenic silica ratio (bSi:C)). **B** PCA loadings showing the contributions of each trait to PC1 and PC2 variation.

4. Discussion

Ocean warming is re-shaping phytoplankton communities, differentially impacting species' productivity and physiology (Behrenfeld et al., 2016; Venegas et al., 2023). South-east Australia, in particular, has experienced accelerated warming of near-surface waters in the EAC over the past few decades (Cai et al., 2005; Hobday & Pecl, 2014; Oliver & Holbrook, 2014; Ridgway, 2007), driving more frequent and extreme marine heatwaves (MHW), such as that experienced in the Tasman Sea in 2015/16 (Frölicher et al., 2018; Hobday & Pecl, 2014; Venegas et al., 2023). With increased frequency, duration and intensity of MHWs in this region expected to continue with the strengthening of the EAC (Cai et al., 2005; Ridgway, 2007), investigation of functional traits is therefore relevant for understanding plasticity, physiological trade-offs and the adaptation potential of individual taxa within a community already undergoing change. Here we investigated thermally adapted *L. danicus* strains isolated from four distinct latitudes to determine how strain-specific physiology, morphology and silica production rates differ with temperature and irradiance.

4.1 Temperature and light have a positive effect on growth and cell size

Cell size is considered a key phenotypic trait influencing the extent at which phytoplankton respond to changes in the environment (Litchman et al., 2010). This study revealed significant intraspecific morphological variability between *L. danicus*



isolates in response to acclimation to their locally adapted thermal niche (as experienced during spring). We confirmed a broad thermal niche in *L. danicus*, as in Ajani et al. (2021), and uncovered a strong temperature-dependent and latitudinal effect on growth rate and cell size (surface area and volume) for each strain. Under low light, low temperature conditions (Twofold Bay), we saw a significant reduction in mean growth rate which was anticorrelated with higher cell surface area to volume ratio. This higher surface area to volume ratio, typical of smaller cells, provides a competitive advantage over their larger counterparts, such as improving nutrient uptake through the reduced diffusive boundary layer thickness, and lower nutrient requirements to achieve maximum growth (Aksnes & Egge, 1991). Generally, phytoplankton are expected to decrease in size with increased surface warming, due to resource constraints and increased reproductive rates (Aksnes & Egge, 1991; Hattich et al., 2024; Sommer et al., 2017). The strength of these shifts toward smaller species are likely to intensify with longer and more intense MHWs, and reduced nutrient supply from weakened coastal upwelling (Zhan et al., 2024). However, the response of diatoms to increasing temperatures is conflicting, with some species increasing in cell size, likely due to evolving different optimal growth temperatures (Liang et al., 2019). The increase in cell size of *L. danicus* at higher temperatures may be a direct physiological response to warming due to differences in phenotypic thermal performance (Baker et al., 2016; Schulte et al., 2011; Sheehan et al., 2020), or driven by an increased need for resources (Hattich et al., 2024).

4.2 Light is a stronger driver than temperature of carbon allocation and elemental stoichiometry

Climate-induced changes in phytoplankton physiology and community structure are likely to alter the efficiency of the biological carbon pump (Basu & Mackey, 2018). This is because the nutritional quality (elemental and biochemical composition) and quantity of organic matter that ultimately influences the transfer of energy to higher trophic levels and carbon export to depth is largely dependent on species composition, their abundance, and their individual elemental stoichiometry (Duncan et al., 2022; Finkel et al., 2010; Li et al., 2012). We found that under low growth irradiance, Coffs Harbour strains (22°C) reduced their chlorophyll *a* content per unit area, compared to the other strains that displayed increased intracellular chlorophyll *a* with decreasing temperature. We also saw that low light acclimated Coffs Harbour strains had enhanced photosynthetic capacity through increased $rETR_{max}$ and I_k . The relatively lower intracellular chlorophyll *a* content of the larger, low light grown Coffs Harbour cells can be explained by the limitation of self-shading of pigments within the cell due to packaging effects with cell size, therefore providing absorptive efficiency of photosynthetic pigments (Finkel et al., 2004). Under the same conditions, the C:N ratios in Coffs Harbour strains were lower than that of other *L. danicus* strains, and when grown under higher irradiance, a finding consistent with Saito & Tsuda (2003). Nitrogen starvation is known to limit phytoplankton growth, due to decreased chlorophyll and protein synthesis, resulting in the accumulation of carbon-rich carbohydrates and therefore an increased carbon to nitrogen ratio (Hipkin et al., 1983; Liefer et al., 2019), but the strains used in this study were grown under nutrient replete conditions, precluding nitrogen limitation as driving this response (Supplementary Table S1). Since these differences in elemental composition did not seem to be linked to temperature and growth rates, but instead shifted in response to growth irradiance in Coffs Harbour strains, this suggests that the stoichiometric plasticity of these strains is in response to low light acclimation (Sauterey & Ward, 2022; Yang et al., 2020). As such, under



anticipated conditions caused by climate change induced stratification, our results indicate increased C:N with higher light availability could lead to decreased food quality (lower nitrogen supply) for grazing zooplankton (Finkel et al., 2010; Li et al., 2012; Marinov et al., 2010; Mitra & Flynn, 2005).

4.3 Temperature effects on silica production

480 Diatom size and silicification are both important phenotypic traits driving species performance, in turn influencing the structure and functioning of diatom assemblages and food web dynamics (Finkel et al., 2010; Sommer et al., 2017). Similarly, silica content, which varies within and among a diatom species (Martin-Jézéquel et al., 2000), can influence susceptibility to grazing pressure and sinking capacity (Smetacek, 1999). Our study revealed a clear latitudinal trend in silicification with temperature, and revealed strong size-dependent relationships. We observed a reduction in silica deposition per cell with lower temperature
 485 (under low growth irradiance), however when differences in cell surface area were accounted for, there was little difference in silicification rates across strains. When normalised to cell volume, strong trends appeared with lower temperature strains showing higher incorporation rates. Together, these results infer that under low light conditions, *L. danicus* from Twofold Bay (16°C) have thicker frustules, which may in part be due to the lower growth rates of this strain, and therefore the increased length of time cultures underwent the cell wall synthesis phase, i.e. allowed for maximum silica incorporation into the frustule
 490 (Martin-Jézéquel et al., 2000; Su et al., 2018). The higher silica content in these cells would result in higher density and therefore more efficient sinking, as well as increased mechanical protection (in response to copepod cues), which are key ecological strategies for the smaller, slower growing diatoms that may otherwise be easily grazed in surface waters (Grønning & Kjørboe, 2020; Hamm et al., 2003; Liu et al., 2016).

On the other hand, under high light conditions, *L. danicus* showed the highest silica incorporation per cell, per surface area
 495 and per volume under the lowest temperature (18°C, Maroubra strain), confirmed by the highest bSi:C ratio, contrasting with the typical inverse correlation of silicification to growth rate (Flynn & Martin-Jézéquel, 2000). This strong increase in silicification of strains grown at lower temperature is consistent with previous studies looking at temperature-regulated responses on silicification using the diatom *T. pseudonana* (Baker et al., 2016; Sheehan et al., 2020). These data suggest that while the mechanistic uptake of silica, polymerisation and deposition of frustules may be thermally constrained, and
 500 independent of cell division (Sheehan et al., 2020), a combination of both increased temperature and light may elicit a similar diminished rate of silica deposition as seen in other stressors, such as seawater acidification, as seen in some species of an Antarctic diatom community (Petrou et al., 2019), high irradiance and pCO₂ in *T. weissflogii* (Fitzgerald-Lowry et al., 2025), or osmotic stress at higher salinities (Vrieling et al., 1999). Considering the phenotypic differences observed here, in the context of future oceans with combined warmer temperature and increased pH where diatoms are larger with reduced silica content,
 505 these data would suggest that diatoms in coastal upwelling systems may be subjected to increased grazing pressure, making them less efficient vectors for silica export (via sinking), and influencing trophic energy transfer to higher levels (Smetacek, 1999; Tréguer et al., 2018; Zhang et al., 2017).

4.4 Ecological implications of a warmer ocean



510 As temperatures in the EAC continue to warm, these data suggest that we may see changes to primary productivity and a
reduction in silica production rates in *L. danicus* in this region. Such responses could have important implications for local
silicon cycling and carbon sequestration given the dominant role of *L. danicus* in coastal communities and spring blooms on
the Australian coast (Ajani et al., 2016). The large strain variability observed for *L. danicus*, a non-isolated cosmopolitan
species, indicate that strain-specific responses to temperature are unlikely to be uniform (Ajani et al., 2021; Ajani et al., 2016;
515 Koester et al., 2013). Therefore, further studies are necessary to isolate whether such responses are solely due to strain
variability or whether the responses we are seeing are in fact temperature-driven (owing to their phenotypic plasticity). While
L. danicus is known to encompass a number of functional traits that allow them to rapidly diversify and adapt with fluctuating
conditions in upwelling systems (Ajani et al., 2021; Ajani et al., 2016), the potential reductions in cell density under warmer
conditions may counteract the success of this species in this region. As such, further studies into species-specific responses
520 with temperature are required to provide a better understanding of the major taxonomic groups that are likely to be affected
under future oceanic conditions. This will enhance our understanding of how warmer oceans will alter species composition,
abundance, distribution and food web interactions, which are essential for shaping marine nutrient cycling and carbon export.

Data Availability

525 All data is available as a supplementary file

Team list

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Author contributions

AMT: methodology, investigation, data curation, data visualisation, formal analysis, writing original draft; DAN:
530 methodology, data visualisation; PA: conceptualisation; KP: conceptualisation, methodology, investigation, data visualisation,
formal analysis, funding acquisition, supervision, writing – review and editing.

Competing interests

The authors declare that there are no competing interests.



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References

- 545 Ajani, P. A., Armbrrecht, L. H., Kersten, O., Kohli, G. S., & Murray, S. A. (2016). Diversity, temporal distribution and physiology of the centric diatom *Leptocylindrus* Cleve (Bacillariophyta) from a southern hemisphere upwelling system. *Diatom Research*, 31(4), 351–365. <https://doi.org/10.1080/0269249X.2016.1260058>
- Ajani, P. A., Petrou, K., Larsson, M. E., Nielsen, D. A., Burke, J., & Murray, S. A. (2021). Phenotypic trait variability as an indication of adaptive capacity in a cosmopolitan marine diatom. *Environmental Microbiology*, 23(1), 207–223. <https://doi.org/10.1111/1462-2920.15294>
- 550 Aksnes, D. L., & Egge, J. K. (1991). A theoretical model for nutrient uptake in phytoplankton. *Marine Ecology Progress Series*, 70(1), 65–72. <https://doi.org/10.3354/meps070065>
- Baines, S. B., Twining, B. S., Brzezinski, M. A., Nelson, D. M., & Fisher, N. S. (2010). Causes and biogeochemical implications of regional differences in silicification of marine diatoms. *Global Biogeochemical Cycles*, 24(4), 1–15. <https://doi.org/10.1029/2010GB003856>
- 555 Baker, K. G., Robinson, C. M., Radford, D. T., McInnes, A. S., Evenhuis, C., & Doblin, M. A. (2016). Thermal performance curves of functional traits aid understanding of thermally induced changes in diatom-mediated biogeochemical fluxes. *Frontiers in Marine Science*, 3(APR), 1–14. <https://doi.org/10.3389/fmars.2016.00044>
- Basu, S., & Mackey, K. R. M. (2018). Phytoplankton as key mediators of the biological carbon pump: Their responses to a changing climate. *Sustainability (Switzerland)*, 10(3). <https://doi.org/10.3390/su10030869>
- 560 Behrenfeld, M. J., O'Malley, R. T., Boss, E. S., Westberry, T. K., Graff, J. R., Halsey, K. H., Milligan, A. J., Siegel, D. A., & Brown, M. B. (2016). Revaluating ocean warming impacts on global phytoplankton. *Nature Climate Change*, 6(3), 323–330. <https://doi.org/10.1038/nclimate2838>
- Bennett, S., Duarte, C. M., Marbà, N., & Wernberg, T. (2019). Integrating within-species variation in thermal physiology into



- 565 climate change ecology. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 374(1778).
<https://doi.org/10.1098/rstb.2018.0550>
- Benoiston, A. S., Ibarbalz, F. M., Bittner, L., Guidi, L., Jahn, O., Dutkiewicz, S., & Bowler, C. (2017). The evolution of diatoms and their biogeochemical functions. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 372(1728). <https://doi.org/10.1098/rstb.2016.0397>
- 570 Bindoff, N. L., Stott, P. A. A., AchutaRao, K. M., Allen, M. R. R., Gillett, N., Gutzler, D., Hansingo, K., Hegerl, G., Hu, Y., Jain, S., Sebbari, R., Zhang, X., Aldrin, M., Sarojini, B. B., Beer, J., Boucher, O., Braconnot, P., Browne, O., Chang, P., ... Zhang, R. (2013). Detection and Attribution of Climate Change: from Global to Regional. In P. Stocker, T. F. Qin, D. Plattner, G. K. Tignor, M. M. B. Allen, S. K. Boschung, J. Nauels, A. Xia, Y. Bex, V. Midgley (Ed.), *Climate change 2013: The physical science basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate*
 575 *Change* (pp. 867–952). Cambridge University Press. <https://doi.org/10.1038/ncomms13676>
- Boscolo-Galazzo, F., Crichton, K. A., Barker, S., & Pearson, P. N. (2018). Temperature dependency of metabolic rates in the upper ocean: A positive feedback to global climate change? *Global and Planetary Change*, 170(March), 201–212. <https://doi.org/10.1016/j.gloplacha.2018.08.017>
- Cai, W., Shi, G., Cowan, T., Bi, D., & Ribbe, J. (2005). The response of the Southern Annular Mode, the East Australian
 580 Current, and the southern mid-latitude ocean circulation to global warming. *Geophysical Research Letters*, 32(23), 1–4. <https://doi.org/10.1029/2005GL024701>
- Cermeño, P., Dutkiewicz, S., Harris, R. P., Follows, M., Schofield, O., & Falkowski, P. G. (2008). The role of nutricline depth in regulating the ocean carbon cycle. *Proceedings of the National Academy of Sciences*, 105(51), 20344–20349.
- Doney, S. C., Ruckelshaus, M., Emmett Duffy, J., Barry, J. P., Chan, F., English, C. A., Galindo, H. M., Grebmeier, J. M.,
 585 Hollowed, A. B., Knowlton, N., Polovina, J., Rabalais, N. N., Sydeman, W. J., & Talley, L. D. (2012). Climate change impacts on marine ecosystems. *Annual Review of Marine Science*, 4, 11–37. <https://doi.org/10.1146/annurev-marine-041911-111611>
- Duncan, R. J., Nielsen, D. A., Sheehan, C. E., Deppeler, S., Hancock, A. M., Schulz, K. G., Davidson, A. T., & Petrou, K. (2022). Ocean acidification alters the nutritional value of Antarctic diatoms. *New Phytologist*, 233(4), 1813–1827. <https://doi.org/10.1111/nph.17868>
- 590 Field, C. B., Behrenfeld, M. J., Randerson, J. T., & Falkowski, P. (1998). Primary production of the biosphere: Integrating terrestrial and oceanic components. *Science*, 281(5374), 237–240. <https://doi.org/10.1126/science.281.5374.237>
- Finkel, Z. V., Beardall, J., Flynn, K. J., Quigg, A., Rees, T. A. V., & Raven, J. A. (2010). Phytoplankton in a changing world: Cell size and elemental stoichiometry. *Journal of Plankton Research*, 32(1), 119–137. <https://doi.org/10.1093/plankt/fbp098>
- Finkel, Z. V., Irwin, A. J., & Schofield, O. (2004). Resource limitation alters the 3/4 size scaling of metabolic rates in
 595 phytoplankton. *Marine Ecology Progress Series*, 273(June 2014), 269–279. <https://doi.org/10.3354/meps273269>
- FitzGerald-Lowry B, Nielsen DA, Duncan RJ, Theseira AM, Thompson G, Petrou K (2025) Multi-trait responses in two marine diatoms to pH and irradiance reveal interactive effect of light and acidification, mediated by silicification. *Limnology and Oceanography*. In Press. doi.org/10.1002/lno.70014.



- Flynn, K. J., & Martin-Jézéquel, V. (2000). Modelling Si–N-limited growth of diatoms. *Journal of Plankton Research*, 22(3), 447–472. <https://doi.org/10.1093/plankt/22.3.447>
- Frölicher, T. L., Fischer, E. M., & Gruber, N. (2018). Marine heatwaves under global warming. *Nature*, 560(7718), 360–364. <https://doi.org/10.1038/s41586-018-0383-9>
- Grønning, J., & Kiørboe, T. (2020). Diatom defence: Grazer induction and cost of shell-thickening. *Functional Ecology*, 34(9), 1790–1801. <https://doi.org/10.1111/1365-2435.13635>
- Guillard, R. R. L. (1973). Division rates. *Handbook of Phycological Methods: Culture Methods and Growth Measurements*.
 Guillard, R. R. L., & Ryther, J. H. (1962). Studies of marine planktonic diatoms: I. *Cyclotella nana* Hustedt, and *Detonula confervacea* (Cleve) Gran. *Canadian Journal of Microbiology*, 8(2), 229–239. <https://doi.org/10.1139/m62-029>
- Hamm, C. E., Merkel, R., Springer, O., Jurkojc, P., Maiert, C., Prechtelt, K., & Smetacek, V. (2003). Architecture and material properties of diatom shells provide effective mechanical protection. *Nature*, 421(6925), 841–843. <https://doi.org/10.1038/nature01416>
- Hattich, G. S. I., Jokinen, S., Sildever, S., Gareis, M., Heikkinen, J., Junghardt, N., Segovia, M., Machado, M., & Sjöqvist, C. (2024). Temperature optima of a natural diatom population increases as global warming proceeds. *Nature Climate Change*, 14(5), 518–525. <https://doi.org/10.1038/s41558-024-01981-9>
- Hildebrand, M., & Lerch, S. J. L. (2015). Diatom silica biomineralization: Parallel development of approaches and understanding. *Seminars in Cell and Developmental Biology*, 46, 27–35. <https://doi.org/10.1016/j.semcdb.2015.06.007>
- Hillebrand, H., Dürselen, C. D., Kirschtel, D., Pollinger, U., & Zohary, T. (1999). Biovolume calculation for pelagic and benthic microalgae. *Journal of Phycology*, 35(2), 403–424. <https://doi.org/10.1046/j.1529-8817.1999.3520403.x>
- Hipkin, C. R., Thomas, R. J., & Syrett, P. J. (1983). Effects of nitrogen deficiency on nitrate reductase, nitrate assimilation and photosynthesis in unicellular marine algae. *Marine Biology*, 77(2), 101–105. <https://doi.org/10.1007/BF00396306>
- Hobday, A. J., & Pecl, G. T. (2014). Identification of global marine hotspots: Sentinels for change and vanguards for adaptation action. *Reviews in Fish Biology and Fisheries*, 24(2), 415–425. <https://doi.org/10.1007/s11160-013-9326-6>
- Jean-Michel, L., Eric, G., Romain, B.-B., Gilles, G., Angélique, M., Marie, D., Clément, B., Mathieu, H., Olivier, L. G., Charly, R., Tony, C., Charles-Emmanuel, T., Florent, G., Giovanni, R., Mounir, B., Yann, D., & Pierre-Yves, L. T. (2021). The Copernicus Global 1/12° Oceanic and Sea Ice GLORYS12 Reanalysis. *Frontiers in Earth Science*, 9. <https://doi.org/10.3389/feart.2021.698876>
- Jeffrey, S. W., & Humphrey, G. F. (1975). New spectrophotometric equations for determining chlorophylls a, b, c1 and c2 in higher plants, algae and natural phytoplankton. *Biochimie Und Physiologie Der Pflanzen*, 167(2), 191–194. [https://doi.org/10.1016/S0015-3796\(17\)30778-3](https://doi.org/10.1016/S0015-3796(17)30778-3)
- Johnson, C. R., Banks, S. C., Barrett, N. S., Cazassus, F., Dunstan, P. K., Edgar, G. J., Frusher, S. D., Gardner, C., Haddon, M., Helidoniotis, F., Hill, K. L., Holbrook, N. J., Hosie, G. W., Last, P. R., Ling, S. D., Melbourne-Thomas, J., Miller, K., Pecl, G. T., Richardson, A. J., ... Taw, N. (2011). Climate change cascades: Shifts in oceanography, species' ranges and subtidal marine community dynamics in eastern Tasmania. *Journal of Experimental Marine Biology and Ecology*, 400(1–2),



- 17–32. <https://doi.org/10.1016/j.jembe.2011.02.032>
- Koester, J. A., Swanson, W. J., & Armbrust, E. V. (2013). Positive selection within a diatom species acts on putative protein
 635 interactions and transcriptional regulation. *Molecular Biology and Evolution*, 30(2), 422–434.
<https://doi.org/10.1093/molbev/mss242>
- Krause, J. W., Brzezinski, M. A., Landry, M. R., Baines, S. B., Nelson, D. M., Selph, K. E., Taylor, A. G., & Twining, B. S.
 (2010). The effects of biogenic silica detritus, zooplankton grazing, and diatom size structure on silicon cycling in the euphotic
 zone of the eastern equatorial Pacific. *Limnology and Oceanography*, 55(6), 2608–2622.
 640 <https://doi.org/10.4319/lo.2010.55.6.2608>
- Leblanc, K., & Hutchins, D. A. (2005). New applications of a biogenic silica deposition fluorophore in the study of oceanic
 diatoms. *Limnology and Oceanography: Methods*, 3(10), 462–476. <https://doi.org/10.4319/lom.2005.3.462>
- Li, G., Cheng, L., Zhu, J., Trenberth, K. E., Mann, M. E., & Abraham, J. P. (2020). Increasing ocean stratification over the
 past half-century. *Nature Climate Change*, 10(12), 12
- 645 Li, W., Gao, K., & Beardall, J. (2012). Interactive Effects of Ocean Acidification and Nitrogen-Limitation on the Diatom
Phaeodactylum tricornutum. *PLoS ONE*, 7(12), 1–8. <https://doi.org/10.1371/journal.pone.0051590>
- Liang, Y., Koester, J. A., Liefer, J. D., Irwin, A. J., & Finkel, Z. V. (2019). Molecular mechanisms of temperature acclimation
 and adaptation in marine diatoms. *ISME Journal*, 13(10), 2415–2425. <https://doi.org/10.1038/s41396-019-0441-9>
- Liefer, J. D., Garg, A., Fyfe, M. H., Irwin, A. J., Benner, I., Brown, C. M., Follows, M. J., Omta, A. W., & Finkel, Z. V. (2019).
 650 The macromolecular basis of phytoplankton C:N:P under nitrogen starvation. *Frontiers in Microbiology*, 10(MAR), 1–16.
<https://doi.org/10.3389/fmicb.2019.00763>
- Litchman, E., de Tezanos Pinto, P., Klausmeier, C. A., Thomas, M. K., & Yoshiyama, K. (2010). Linking traits to species
 diversity and community structure in phytoplankton. *Hydrobiologia*, 653(1), 15–28. <https://doi.org/10.1007/s10750-010-0341-5>
- 655 Liu, H., Chen, M., Zhu, F., & Harrison, P. J. (2016). Effect of diatom silica content on copepod grazing, growth and
 reproduction. *Frontiers in Marine Science*, 3(JUN), 1–8. <https://doi.org/10.3389/fmars.2016.00089>
- Malviya, S., Scalco, E., Audic, S., Vincent, F., Veluchamy, A., Poulain, J., Wincker, P., Iudicone, D., De Vargas, C., Bittner,
 L., Zingone, A., & Bowler, C. (2016). Insights into global diatom distribution and diversity in the world's ocean. *Proceedings*
of the National Academy of Sciences of the United States of America, 113(11), E1516–E1525.
 660 <https://doi.org/10.1073/pnas.1509523113>
- Marinov, I., Doney, S. C., & Lima, I. D. (2010). Response of ocean phytoplankton community structure to climate change
 over the 21st century: Partitioning the effects of nutrients, temperature and light. *Biogeosciences*, 7(12), 3941–3959.
<https://doi.org/10.5194/bg-7-3941-2010>
- Martin-Jézéquel, V., Hildebrand, M., & Brzezinski, M. A. (2000). Silicon metabolism in diatoms: Implications for growth.
 665 *Journal of Phycology*, 36(5), 821–840. <https://doi.org/10.1046/j.1529-8817.2000.00019.x>
- McNair, H. M., Brzezinski, M. A., & Krause, J. W. (2015). Quantifying diatom silicification with the fluorescent dye, PDMPO.



- Limnology and Oceanography: Methods*, 13(10), 587–599. <https://doi.org/10.1002/lom3.10049>
- Mitra, A., & Flynn, K. J. (2005). Predator-prey interactions: Is “ecological stoichiometry” sufficient when good food goes bad? *Journal of Plankton Research*, 27(5), 393–399. <https://doi.org/10.1093/plankt/fbi022>
- 670 Nanjappa, D., Kooistra, W. H. C. F., & Zingone, A. (2013). A reappraisal of the genus *Leptocylindrus* (Bacillariophyta), with the addition of three species and the erection of *Tenuicylindrus* gen. nov. *Journal of Phycology*, 49(5), 917–936. <https://doi.org/10.1111/jpy.12102>
- Nelson, D. M., Tréguer, P. J., Brzezinski, M. A., Leynaert, A., & Quéguiner, B. (1995). Production and dissolution of biogenic silica in the ocean: Revised global estimates, comparison with regional data and relationship to biogenic sedimentation. *Global*
- 675 *Biogeochemical Cycles*, 9(3), 359–372. <https://doi.org/10.1029/95GB01070>
- Oliver, E. C. J., & Holbrook, N. J. (2014). Journal of Geophysical Research : Oceans Modeling the East Australian Current in a future climate. *Journal of Geophysical Research: Oceans*, 119, 2788–2805. <https://doi.org/10.1002/2013JC009591>. Received
- Petrou, K., Baker, K. G., Nielsen, D. A., Hancock, A. M., Schulz, K. G., & Davidson, A. T. (2019). Acidification diminishes
- 680 diatom silica production in the Southern Ocean. *Nature Climate Change*, 9(10), 781–786. <https://doi.org/10.1038/s41558-019-0557-y>
- Petrou, K., Doblin, M. A., & Ralph, P. J. (2011). Heterogeneity in the photoprotective capacity of three Antarctic diatoms during short-term changes in salinity and temperature. *Marine Biology*, 158(5), 1029–1041. <https://doi.org/10.1007/s00227-011-1628-4>
- 685 Phillips, L. R., Carroll, G., Jonsen, I., Harcourt, R., & Roughan, M. (2020). A Water Mass Classification Approach to Tracking Variability in the East Australian Current. *Frontiers in Marine Science*, 7(June), 1–10. <https://doi.org/10.3389/fmars.2020.00365>
- Popova, E., Yool, A., Byfield, V., Cochrane, K., Coward, A. C., Salim, S. S., Gasalla, M. A., Henson, S. A., Hobday, A. J., Pecl, G. T., Sauer, W. H., & Roberts, M. J. (2016). From global to regional and back again: Common climate stressors of
- 690 marine ecosystems relevant for adaptation across five ocean warming hotspots. *Global Change Biology*, 22(6), 2038–2053. <https://doi.org/10.1111/gcb.13247>
- R Core Team. (2024). *R: A Language and Environment for Statistical Computing* (2024.12.0+467). R Foundation for Statistical Computing.
- Ralph, P. J., & Gademann, R. (2005). Rapid light curves: A powerful tool to assess photosynthetic activity. *Aquatic Botany*,
- 695 82(3), 222–237. <https://doi.org/10.1016/j.aquabot.2005.02.006>
- Ridgway, K. R. (2007). Long-term trend and decadal variability of the southward penetration of the East Australian Current. *Geophysical Research Letters*, 34(13), 1–5. <https://doi.org/10.1029/2007GL030393>
- Ringuet, S., Sassano, L., & Johnson, Z. I. (2011). A suite of microplate reader-based colorimetric methods to quantify ammonium, nitrate, orthophosphate and silicate concentrations for aquatic nutrient monitoring. *Journal of Environmental*
- 700 *Monitoring*, 13(2), 370–376.



- Ritchie, R. J. (2006). Consistent sets of spectrophotometric chlorophyll equations for acetone, methanol and ethanol solvents. *Photosynthesis Research*, 89(1), 27–41. <https://doi.org/10.1007/s11120-006-9065-9>
- Saito, H., & Tsuda, A. (2003). Influence of light intensity on diatom physiology and nutrient dynamics in the Oyashio region. *Progress in Oceanography*, 57(3–4), 251–263. [https://doi.org/10.1016/s0079-6611\(03\)00100-9](https://doi.org/10.1016/s0079-6611(03)00100-9)
- 705 Sauterey, B., & Ward, B. A. (2022). Environmental control of marine phytoplankton stoichiometry in the North Atlantic Ocean. *Proceedings of the National Academy of Sciences of the United States of America*, 119(1), 1–6. <https://doi.org/10.1073/pnas.2114602118>
- Schindelin, J., Arganda-Carreras, I., Frise, E., Kaynig, V., Longair, M., Pietzsch, T., Preibisch, S., Rueden, C., Saalfeld, S., Schmid, B., Tinevez, J.-Y., White, D. J., Hartenstein, V., Eliceiri, K., Tomancak, P., & Cardona, A. (2012). Fiji: an open-
710 source platform for biological-image analysis. *Nature Methods*, 9, 676. <https://doi.org/10.1038/nmeth.2019>
- Schnetger, B., & Lehnert, C. (2014). Determination of nitrate plus nitrite in small volume marine water samples using vanadium (III) chloride as a reduction agent. *Marine Chemistry*, 160, 91–98.
- Schulte, P. M., Healy, T. M., & Fangue, N. A. (2011). *Thermal Performance Curves, Phenotypic Plasticity, and the Time Scales of Temperature Exposure*. 51(5), 691–702. <https://doi.org/10.1093/icb/acr097>
- 715 Sheehan, C. E., Baker, K. G., Nielsen, D. A., & Petrou, K. (2020). Temperatures above thermal optimum reduce cell growth and silica production while increasing cell volume and protein content in the diatom *Thalassiosira pseudonana*. *Hydrobiologia*, 847(20), 4233–4248. <https://doi.org/10.1007/s10750-020-04408-6>
- Shrestha, R. P., & Hildebrand, M. (2015). Evidence for a regulatory role of diatom silicon transporters in cellular silicon responses. *Eukaryotic Cell*, 14(1), 29–40. <https://doi.org/10.1128/EC.00209-14>
- 720 Shrestha, R. P., Tesson, B., Norden-Krichmar, T., Federowicz, S., Hildebrand, M., & Allen, A. E. (2012). Whole transcriptome analysis of the silicon response of the diatom *Thalassiosira pseudonana*. *BMC Genomics*, 13(1). <https://doi.org/10.1186/1471-2164-13-499>
- Smetacek, V. (1999). Diatoms and the ocean carbon cycle. *Protist*, 150(1), 25–32. [https://doi.org/10.1016/s1434-4610\(99\)70006-4](https://doi.org/10.1016/s1434-4610(99)70006-4)
- 725 Smetacek, V. (2012). Making sense of ocean biota: How evolution and biodiversity of land organisms differ from that of the plankton. *Journal of Biosciences*, 37(4), 589–607. <https://doi.org/10.1007/s12038-012-9240-4>
- Sommer, U., Peter, K. H., Genitsaris, S., & Moustaka-Gouni, M. (2017). Do marine phytoplankton follow Bergmann’s rule sensu lato? *Biological Reviews*, 92(2), 1011–1026. <https://doi.org/10.1111/brv.12266>
- Strickland, J. D. H., & Parsons, T. R. (1972). A practical handbook of sea-water analysis, 2nd edn. 167, 311. In *Journal of Fisheries Research Board of Canada* (Vol. 167). Bulletin of Fisheries Research Board of Canada.
- 730 Su, Y., Lundholm, N., & Ellegaard, M. (2018). The effect of different light regimes on diatom frustule silicon concentration. *Algal Research*, 29(November 2017), 36–40. <https://doi.org/10.1016/j.algal.2017.11.014>
- Thamatrakoln, K., & Hildebrand, M. (2008). Silicon uptake in diatoms revisited: A model for saturable and nonsaturable uptake kinetics and the role of silicon transporters. *Plant Physiology*, 146(3), 1397–1407.



- 735 <https://doi.org/10.1104/pp.107.107094>
 Thomas, M. K., Kremer, C. T., Klausmeier, C. A., & Litchman, E. (2012). A global pattern of thermal adaptation in marine
 phytoplankton. *Science*, 338(6110), 1085–1088. <https://doi.org/10.1126/science.1224836>
 Tréguer, P. J., Bowler, C., Moriceau, B., Dutkiewicz, S., Gehlen, M., Aumont, O., Bittner, L., Dugdale, R., Finkel, Z., Iudicone,
 D., Jahn, O., Guidi, L., Lasbleiz, M., Leblanc, K., Levy, M., & Pondaven, P. (2018). Influence of diatom diversity on the ocean
 740 biological carbon pump. *Nature Geoscience*, 11(1), 27–37. <https://doi.org/10.1038/s41561-017-0028-x>
 Venegas, R. M., Acevedo, J., & Tremblé, E. A. (2023). Three decades of ocean warming impacts on marine ecosystems: A
 review and perspective. *Deep-Sea Research Part II: Topical Studies in Oceanography*, 212(August), 105318.
<https://doi.org/10.1016/j.dsr2.2023.105318>
 Vrieling, E., Poort, L., Beelen, T., & Gieskes, W. (1999). Growth and silica content of the diatoms *Thalassiosira weissflogii*
 745 and *Navicula salinarum* at different salinities and enrichments with aluminium. *European Journal of Phycology*, 34(3), 307–
 316. <https://doi.org/10.1080/09670269910001736362>
 Wickham, H. (2016). *ggplot2: Elegant Graphics for Data Analysis*. Springer-Verlag New York. <https://ggplot2.tidyverse.org>
 Yang, X., Liu, L., Yin, Z., Wang, X., Wang, S., & Ye, Z. (2020). Quantifying photosynthetic performance of phytoplankton
 based on photosynthesis–irradiance response models. *Environmental Sciences Europe*, 32(1). [https://doi.org/10.1186/s12302-](https://doi.org/10.1186/s12302-020-00306-9)
 750 [020-00306-9](https://doi.org/10.1186/s12302-020-00306-9)
 Zhan, W., Feng, M., Zhang, Y., Shen, X., Zhan, H., & He, Q. (2024). Reduced and smaller phytoplankton during marine
 heatwaves in eastern boundary upwelling systems. *Communications Earth and Environment*, 5(1), 1–11.
<https://doi.org/10.1038/s43247-024-01805-w>
 Zhang, S., Liu, H., Ke, Y., & Li, B. (2017). Effect of the silica content of diatoms on protozoan grazing. *Frontiers in Marine*
 755 *Science*, 4(JUN). <https://doi.org/10.3389/fmars.2017.00202>