



Regime-dependent expression of regional aerosol signals in urban PM₁₀ and PM_{2.5} across arid northern China

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Abstract. Air pollution in arid regions emerges from the interaction between regional aerosol transport and local atmospheric processes, yet it remains unclear whether inter-city synchronization reflects direct transport or shared meteorological forcing. This ambiguity limits the interpretation of urban air quality and the design of effective control strategies in dust-influenced environments. Here, we analyze PM₁₀ and PM_{2.5} observations across 47 cities in northern China during 2015–2025. We quantify regional coupling using explained variance (R^2) and introduce a differential regional control metric (ΔR^2) to resolve the relative sensitivity of coarse and fine particles to regional forcing. We find that strong inter-city synchronization coexists with substantial heterogeneity in underlying processes. Although the regional signal explains comparable variability in PM₁₀ and PM_{2.5}, ΔR^2 varies widely across cities (−0.57 to +0.34), revealing distinct pollution regimes. These regimes are systematically linked to aerosol composition and proximity to desert sources, indicating that similar levels of synchronization can arise from fundamentally different mechanisms. Event-scale analysis further shows a nonlinear dependence of aerosol concentrations on wind speed, with extreme PM₁₀ occurring under both weak and strong winds. This reflects two contrasting regimes: accumulation under limited ventilation and transport-driven dust events. Together, these results demonstrate that regional synchronization is not a unique indicator of transport but an emergent property of multiple interacting processes. The proposed framework provides a process-oriented basis for distinguishing these regimes and improving the interpretation and management of air quality in arid regions.

Keywords: Aerosol pollution; Regional synchronization; Differential regional control; Arid regions; Dust transport.



1. Introduction

Air pollution in arid and semi-arid regions represents a complex system in which regional aerosol transport processes interact with local emission sources (Monks et al., 2009; Hassan et al., 2022; Liu et al., 2024). As demonstrated in studies of oasis cities along the southern margin of the Taklamakan Desert, up to 70% of PM₁₀ during dust storm periods originates from natural mineral dust, whereas during non-dust conditions anthropogenic sources such as industrial emissions and coal combustion become dominant (Yu et al., 2024). In contrast to humid and industrialized regions, where anthropogenic factors prevail, mineral dust plays a major role in arid environments, being generated in extensive desert systems and transported over long distances by atmospheric circulation (Shao et al., 2011; Ginoux et al., 2012). As noted by Querol et al. (2025) in a recent review, in receptor regions desert dust is typically mixed with locally and regionally emitted particulate matter, which complicates the interpretation of observations. This mixing process has also been confirmed for East Asia, where, according to the synthesis by Chen et al. (2017a), aerosols originating from the Taklamakan and Gobi deserts represent a dominant regional forcing. As a result, particulate matter concentrations in urban areas reflect not only local emissions but also the influence of large-scale processes, making both the interpretation of observations and the development of effective air quality management strategies more challenging.

However, the presence of a regional aerosol background does not necessarily imply that urban concentration variability is directly controlled by particle transport. In arid regions, atmospheric circulation may play a dual role, facilitating aerosol transport while also controlling ventilation and accumulation conditions (Miao et al., 2018; Huang et al., 2018; Liao et al., 2024). In particular, periods of weak wind and stable atmospheric stratification may promote the accumulation of both dust and anthropogenic components, whereas stronger winds can lead to dilution and dispersion. In this context, it remains an open question whether the observed regional coherence of concentrations is driven by direct aerosol transport or by the synchronous influence of meteorological conditions.

Northern China represents a unique natural and climatic system, encompassing one of the largest arid belts in the world, including the Taklamakan, Gobi, Badain Jaran, and Tengger deserts, as well as the Qaidam Basin (Chen et al., 2017a; Zhang et al., 2021; Mao et al., 2025). These regions are major sources of mineral dust, influencing air quality at both local and interregional scales (Uno et al., 2009; Zhang et al., 2018;



67 [Tian et al., 2021](#)). At the same time, urbanized areas located within or at the margins of
68 these deserts are affected by both regional transport and local factors, including
69 secondary aerosol formation, topographic effects, and urban emissions ([Zhang et al.,](#)
70 [2003](#); [Li et al., 2010](#); [Meng et al., 2019](#)).

71 Despite the substantial body of research on dust aerosols and air quality, most studies
72 have focused either on dust source characterization or on the temporal variability of
73 concentrations ([Li et al., 2025](#); [Zhang et al., 2025](#)). The spatial coherence of pollution
74 across cities and the mechanisms controlling differences in their degree of coupling to
75 the regional signal remain insufficiently explored ([Lin et al., 2022](#); [Zhang et al., 2022](#)).
76 In particular, it is still unclear to what extent the variability of PM_{10} and $PM_{2.5}$
77 concentrations is governed by regional processes and how this role varies depending on
78 geographic location and aerosol composition.

79 A particularly important aspect is the comparison between coarse (PM_{10}) and fine
80 ($PM_{2.5}$) aerosol fractions. In arid regions, PM_{10} is primarily associated with mineral dust
81 and regional sources, whereas $PM_{2.5}$ contains a substantial fraction of secondary and
82 anthropogenic components formed in the atmosphere and more sensitive to local
83 conditions ([Huang et al., 2014a](#); [Zhang and Cao, 2015](#); [Fomba et al., 2024](#); [Zhou et al.,](#)
84 [2026](#)). Field measurements in northwestern China have shown that the mixing of dust
85 and anthropogenic pollutants near desert sources leads to significant changes in aerosol
86 optical and microphysical properties ([Li et al., 2010](#)). This distinction provides a basis
87 for using the comparison of these fractions as a tool to disentangle regional and local
88 controls.

89 In this study, we propose a quantitative framework for analyzing regional
90 synchronization of air pollution and inter-city heterogeneity based on the
91 decomposition of concentration variability into regional and local components. The key
92 idea is to use explained variance (R^2) as a measure of the contribution of the regional
93 signal and to introduce a differential regional control metric (ΔR^2), reflecting the
94 relative dominance of processes affecting PM_{10} and $PM_{2.5}$. This approach enables a
95 transition from a descriptive characterization of pollution to the identification of
96 underlying mechanisms.

97 The analysis is based on ground-based monitoring data from 47 cities across arid
98 and semi-arid regions of northern China for the period 2015 to 2025. The main
99 objectives of this study are: (1) to assess the degree of regional synchronization of PM_{10}
100 and $PM_{2.5}$ concentrations, (2) to quantitatively separate the contributions of regional



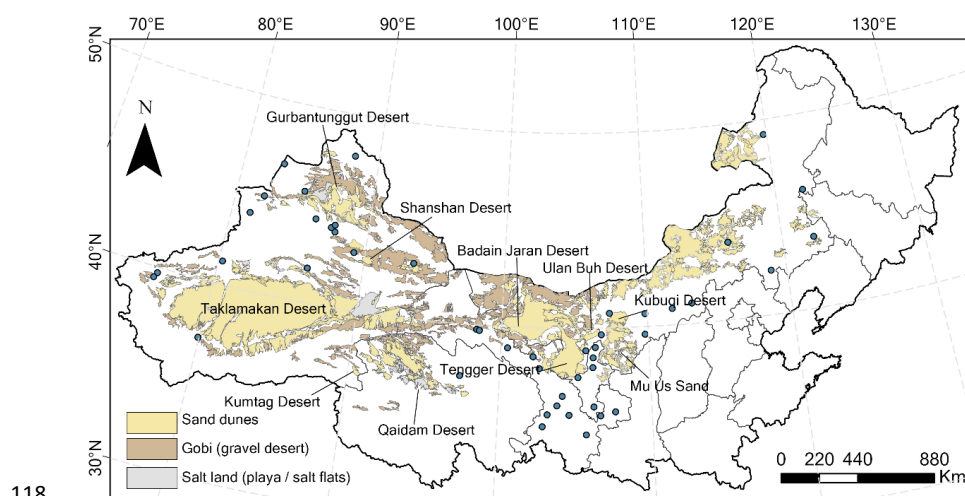
101 and local processes, (3) to classify cities according to types of regional control, and (4)
102 to identify the factors driving inter-city heterogeneity.

103 2. Methods

104 This study is based on a conceptual framework that separates observed variability in
105 urban particulate matter into regional and local components using a data-driven
106 decomposition approach. The framework combines leave-one-out estimation of a
107 regional signal with variance-based metrics to quantify the relative contribution of
108 large-scale processes and city-specific factors. Differences between PM_{10} and $PM_{2.5}$
109 responses are further used to diagnose regime-dependent mechanisms of aerosol
110 formation and amplification.

111 2.1. Study area and data sources

112 The study covers 47 cities located in arid and semi-arid regions of northwestern
113 China, including Xinjiang, Gansu, Ningxia, and Inner Mongolia, as well as inland
114 basins such as the Qaidam Basin. These regions are characterized by frequent dust
115 activity, strong climatic gradients, and diverse geomorphological conditions. The
116 spatial configuration of the study area, including the distribution of cities and major
117 desert systems, is shown in Figure 1.



118
119 **Figure 1.** Study area showing the spatial distribution of analyzed cities and major
120 desert regions across arid and semi-arid regions of northern China.

121 Ground-based air quality monitoring data were obtained from the national network
122 operated by the China National Environmental Monitoring Centre, including hourly



PM₁₀ and PM_{2.5} concentrations for the period from 1 January 2015 to 31 December 2025. Measurements are conducted in accordance with national standards and subject to centralized quality control procedures. Meteorological data were derived from the ERA5 reanalysis (European Centre for Medium-Range Weather Forecasts; [Hersbach et al., 2020](#)), using monthly fields at 0.25°×0.25° resolution, including 10 m wind speed. For each city, wind speed was extracted at the corresponding grid point and used to examine its role in aerosol variability and extreme pollution conditions.

Within this study, hourly data were aggregated to daily values and subsequently to monthly means. Geographic coordinates of the cities, including latitude and longitude, were used to analyze spatial gradients.

2.2. Data preprocessing and quality control

Initial data processing included format harmonization, removal of invalid values, and verification of time series continuity, with negative concentration values treated as measurement errors and excluded. Data aggregation was conducted in two stages, with hourly observations averaged to daily values and subsequently to monthly means; a monthly value was considered valid if at least 75% of daily observations were available. To ensure robustness, only cities with more than 85% valid monthly values over the entire study period for both PM₁₀ and PM_{2.5} were retained. After applying these criteria and matching with geographic coordinates, a total of 47 cities with time series of up to 132 months were included in the analysis.

2.3. Construction of regional signal

To estimate the regional signal, a leave-one-out approach was applied. For each city i , the regional signal was defined as the average concentration across all other cities:

$$R_i(t) = \frac{1}{N-1} \sum_{j \neq i} C_j(t)$$

where $C_j(t)$ - s the concentration in city j , and N - is the total number of cities.

This approach ensures that the regional signal is independent of the local dynamics of the target city and prevents artificial inflation of correlation estimates.

2.4. Standardization and anomaly analysis

To analyze synchronization of time series, standardized anomalies were used. For each city, the time series was normalized as follows:

$$Z_i(t) = \frac{C_i(t) - \mu_i}{\sigma_i}$$



154 where μ_i and σ_i — are the mean and standard deviation of concentrations for city i .

155 Standardization removes the influence of differences in absolute concentration
156 levels and allows the analysis to focus on temporal coherence. Spatial heterogeneity
157 was assessed based on the distribution of $Z_i(t)$ across all cities at each time step using
158 the interquartile range (IQR).

159 It should be noted that seasonal components were removed in this study, whereas
160 explicit detrending of long-term trends was not performed. As a result, part of the
161 detected synchronization may reflect not only short-term regional processes but also
162 synchronous changes associated with long-term emission trends, environmental
163 policies, and external factors such as the COVID-19 period.

164 **2.5. Correlation analysis and regional synchronization**

165 The relationship between concentration time series and the regional signal was
166 evaluated using the Pearson correlation coefficient. For each city, the correlation
167 between its time series and the corresponding leave-one-out regional signal was
168 calculated as:

$$169 \quad r_i = \text{corr}(C_i(t), R_i(t))$$

170 The inter-city correlation matrix was constructed using monthly data and Pearson
171 correlation. Calculations were performed pairwise using all available observations
172 (pairwise complete), allowing for missing data.

173 **2.6. Variance decomposition and differential regional control**

174 To quantify the contribution of the regional signal, explained variance was used:

$$175 \quad R_i^2 = r_i^2$$

176 Values of R^2 were calculated separately for PM_{10} and $PM_{2.5}$. To assess differences
177 between them, a differential regional control metric was introduced:

$$178 \quad \Delta R_i^2 = R_{PM_{10},i}^2 - R_{PM_{2.5},i}^2$$

179 This metric provides a quantitative measure of the relative contribution of the
180 regional signal to variability in different aerosol fractions and allows identification of
181 inter-city variability in pollution formation mechanisms.

182 **2.7. City classification (typology of synchronization regimes)**

183 City classification was based on R^2 and ΔR^2 , taking into account both the absolute
184 strength of regional coupling and the differences between PM_{10} and $PM_{2.5}$:

- 185 • Strong regional control: $R_{PM_{10}}^2 \geq 0.64$ и $R_{PM_{2.5}}^2 \geq 0.64$;



- 186 • PM₁₀-dominated regional control: $\Delta R^2 \geq 0.10$;
- 187 • PM_{2.5}-dominated regional control: $\Delta R^2 \leq -0.10$;
- 188 • Moderate mixed control: all other cases.

189 This typology allows identification of groups of cities characterized by different
190 pollution formation mechanisms.

191 The threshold values used for classification (e.g., $R^2 \geq 0.64$) were selected empirically
192 based on the distribution of values in the dataset and represent the upper range of
193 regional synchronization. These thresholds are intended for illustrative separation of
194 cities into types and do not imply strict physical boundaries between regimes. Therefore,
195 the proposed typology is descriptive in nature and aims to reveal general patterns rather
196 than enforce rigid classification.

197 To assess the robustness of the typology, a sensitivity analysis was performed by
198 varying the thresholds of R^2 and ΔR^2 within ranges close to the baseline criteria. For
199 each alternative set of thresholds, the classification was repeated and its consistency
200 with the baseline scheme ($R^2=0.64$, $\Delta R^2=0.10$) was evaluated.

201 **2.8. Long-term mean concentrations and aerosol composition**

202 To analyze inter-city differences, mean concentrations over the entire observation
203 period were calculated as:

$$\begin{aligned} \overline{PM10}_i &= \frac{1}{T} \sum_{t=1}^T PM10_i(t) \\ \overline{PM2.5}_i &= \frac{1}{T} \sum_{t=1}^T PM2.5_i(t) \end{aligned}$$

206 Based on these values, the ratio was computed as:

$$Ratio_i = \frac{\overline{PM2.5}_i}{\overline{PM10}_i}$$

208 which was used as an integrated indicator of aerosol composition.

209 **2.9. Drivers of inter-city heterogeneity**

210 To analyze the drivers of inter-city heterogeneity, the ΔR^2 metric was compared with:

- 211 • the PM_{2.5}/PM₁₀ ratio
- 212 • the mean PM₁₀ concentration
- 213 • geographic coordinates (latitude and longitude)

214 Statistical relationships were assessed using the Spearman rank correlation
215 coefficient:



216
$$p = \text{Spearman}(X, \Delta R^2)$$

217 This approach is robust to outliers and does not require the assumption of linear
218 relationships. Statistical significance was evaluated at the level of $p < 0.05$.

219 **2.10. Wind regime analysis and extreme events**

220 To investigate the role of atmospheric circulation in shaping regional
221 synchronization, a regime-based analysis combining extreme event identification and
222 wind-dependent statistics was performed. Regional mean values of PM_{10} , $\text{PM}_{2.5}$, and
223 10 m wind speed were calculated by averaging across all cities, and extreme pollution
224 events were defined as months with PM_{10} exceeding the 90th percentile.

225 The dependence of extreme event probability on wind conditions was assessed by
226 grouping wind speed values into intervals and calculating, for each interval, the fraction
227 of months classified as extreme. In parallel, PM_{10} distributions were compared between
228 contrasting wind regimes, defined using wind speed quantiles (below the 25th
229 percentile for low-wind conditions and above the 75th percentile for high-wind
230 conditions), with differences evaluated using the Mann–Whitney U test.

231 Finally, the relationship between wind speed and pollution intensity during extreme
232 events was quantified using the Spearman rank correlation coefficient, allowing
233 identification of robust monotonic dependencies without assuming linearity.

234 **3. Results**

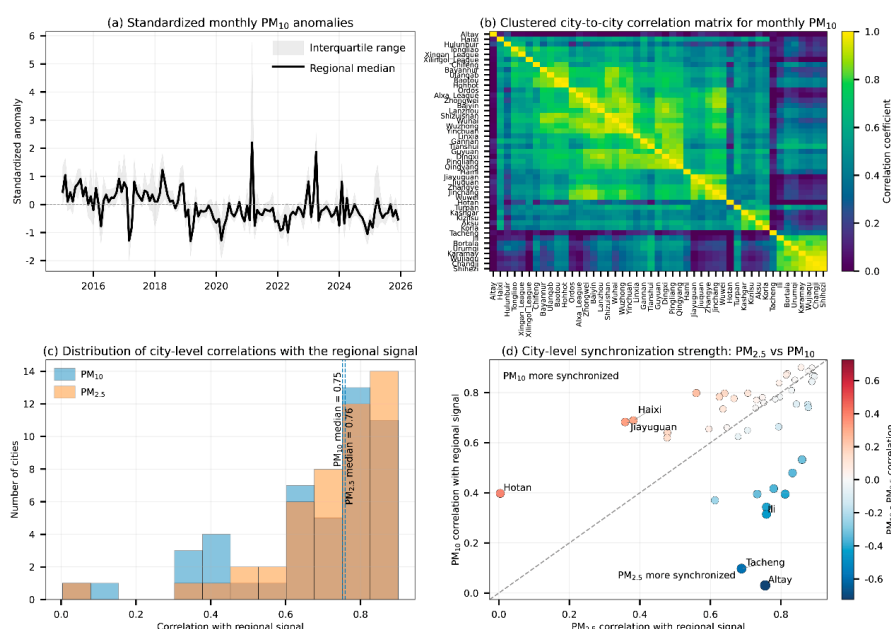
235 **3.1. Regional synchronization of PM_{10} and $\text{PM}_{2.5}$ concentrations**

236 Analysis of ground-based monitoring data from 47 cities in arid regions of
237 northwestern China over the period 2015–2025 (132 months after quality control)
238 reveals that variability in aerosol concentrations is characterized by pronounced
239 regional coherence. After removal of the seasonal component and normalization of time
240 series (Figure 2a), the interquartile range (IQR), representing spatial variability among
241 cities, remains relatively narrow throughout most of the period (median IQR = 0.75).
242 This indicates that fluctuations in PM_{10} concentrations occur synchronously across
243 most cities and therefore cannot be explained solely by local sources. The observed
244 coherence suggests a dominant influence of large-scale factors; however, it does not
245 uniquely determine the underlying mechanism of their nature. Regional
246 synchronization may arise either from the coherent influence of atmospheric circulation,
247 including aerosol transport (Shao et al., 2011; Zhao et al., 2015), or from simultaneous
248 changes in ventilation and accumulation conditions across cities. Thus, the detected



249 synchronicity reflects the presence of a common regional regime but does not directly
250 define its underlying mechanism.

251 At the same time, temporal dynamics show a clear relationship between the intensity
252 of the regional signal and spatial heterogeneity. In particular, in March 2021, a
253 simultaneous increase in both the amplitude of the regional anomaly (up to 2.20) and
254 spatial variability (IQR = 6.01) is observed (Figure 2a), reflecting the impact of an
255 intense large-scale dust event (Liang et al., 2022; Yu et al., 2023). Such a synchronous
256 response across a large number of cities indicates that extreme concentration levels are
257 primarily driven by regional-scale processes (Wang et al., 2021), while local conditions
258 modulate the magnitude of the response in individual cities.



259

260 **Figure 2.** Regional synchronization of particulate pollution across arid cities of China.

261 The spatial structure of inter-city relationships supports this interpretation. The PM₁₀
262 correlation matrix (Figure 2b) shows a consistently positive relationship among most
263 cities (mean = 0.48, median = 0.50), indicating the presence of a common regional
264 regime. However, the wide range of values (from -0.35 to 0.98) reflects substantial
265 spatial heterogeneity. These differences are likely associated with a combination of
266 factors, including geographic location, orographic barriers, and the degree of
267 involvement of cities in major dust transport pathways (Ge et al., 2014; Kang et al.,



268 2021). Thus, the regional signal is not uniform but modulated, with local conditions
269 determining the degree of deviation of individual cities from the common regime.

270 Quantitative assessment of the relationship between each city and the regional signal
271 (Figure 2c) shows that for PM₁₀ the median correlation reaches 0.75 (mean = 0.67),
272 indicating a high degree of synchronization. For PM_{2.5}, similar values (median = 0.76,
273 mean = 0.72) suggest comparable coherence; however, lower variability ($\sigma = 0.17$
274 compared to 0.21 for PM₁₀) indicates a more homogeneous spatial distribution of fine
275 particles. This difference reflects the physical nature of the particles, where PM₁₀ in
276 arid regions is predominantly associated with mineral dust and regional sources, while
277 PM_{2.5} includes a substantial fraction of secondary and anthropogenic components
278 formed locally and less sensitive to extreme dust events (Zhang and Cao, 2015).

279 Comparison of synchronization strength between PM₁₀ and PM_{2.5} at the city level
280 (Figure 2d) shows no dominant aerosol type. In 24 out of 47 cities, PM₁₀ exhibits a
281 stronger relationship with the regional signal, whereas in 23 cities the opposite pattern
282 is observed. The mean difference remains close to zero (−0.05), indicating a balance
283 between dust-related and secondary processes in shaping regional synchronization. At
284 the same time, the wide range of differences (from −0.72 to +0.39) reflects substantial
285 inter-city variability in pollution formation mechanisms.

286 The inclusion of Haixi Prefecture (Qaidam Basin) reveals an important transitional
287 regime within the regional synchronization system. Despite its location in an arid
288 environment and proximity to potential dust sources, Haixi exhibits relatively low
289 correlation with the regional signal for PM_{2.5} (0.382) and is among the cities with the
290 strongest dominance of PM₁₀ (difference between PM₁₀ and PM_{2.5} correlations ≈ 0.31).
291 This indicates the dominance of local dust processes and limited coupling with the
292 broader regional dynamics, reflecting the characteristics of inland basins with sparse
293 urbanization (Zhu et al., 2026). In contrast to more integrated urban systems, such as
294 cities in Gansu and Ningxia, Haixi demonstrates reduced participation in the common
295 regional regime, representing a transitional type between locally controlled and
296 regionally synchronized systems.

297 To further characterize spatial heterogeneity, cities were classified according to their
298 synchronization regimes with the regional signal (Table 1). The classification accounts
299 for both the absolute level of correlation and differences between PM₁₀ and PM_{2.5},
300 allowing identification of groups of cities with distinct dominant pollution formation
301 mechanisms. In total, 11 cities exhibit strong regional synchronization, 9 are PM₁₀-



dominated, 13 are PM_{2.5}-dominated, and 14 show moderate mixed control. The full list of cities and corresponding correlation values is provided in Table S1.

Table 1. Typology of cities based on synchronization with the regional particulate pollution signal, defined using R^2 and ΔR^2 thresholds.

Category	Criteria	Number of cities	Interpretation	Example cities
Strong regional synchronization	$R^2 (PM_{10}) \geq 0.64$ and $R^2 (PM_{2.5}) \geq 0.64$	11	Dominant regional control for both fractions	Wuzhong, Guyuan
PM ₁₀ -dominated synchronization	$\Delta R^2 \geq 0.10$	9	Stronger regional coupling for coarse particles (dust influence)	Jinchang, Alxa_ League
PM _{2.5} -dominated synchronization	$\Delta R^2 \leq -0.10$	13	Stronger regional coupling for fine particles (secondary/anthropogenic influence)	Tianshui, Turpan
Moderate synchronization	Otherwise	14	Mixed and intermediate control regimes	Zhongwei, Ulanqab

Cities with strong regional synchronization (e.g., Wuzhong and Guyuan) are characterized by high correlations for both PM₁₀ and PM_{2.5}, indicating strong coherence with the regional signal. PM₁₀-dominated cities (e.g., Jinchang, Alxa League, and Haixi) exhibit substantially stronger coupling of PM₁₀ with the regional signal compared to PM_{2.5}, consistent with the enhanced role of coarse particles associated with mineral dust. In contrast, PM_{2.5}-dominated cities (e.g., Tianshui and Turpan) show stronger synchronization in fine particles, reflecting a greater influence of secondary and anthropogenic processes. Moderately synchronized cities (e.g., Zhongwei and Ulanqab) display intermediate correlations, indicating mixed control where regional and local factors interact.

Overall, the results indicate that air pollution in arid urban systems is characterized by a pronounced hierarchical structure of synchronization, in which the regional signal defines the overall temporal regime, while local conditions modulate its manifestation at the city level. The inclusion of Haixi further demonstrates the existence of transitional regimes where coupling to the regional signal weakens and the role of local dust sources increases. This highlights the necessity of distinguishing between regional and local mechanisms of aerosol concentration formation, which is addressed in the subsequent analysis.

It is important to note that the leave-one-out regional signal represents a statistical construct reflecting temporal coherence among cities rather than a direct measure of



326 physical aerosol transport processes. The observed coherence may arise either from a
327 common regional forcing, such as dust storms and atmospheric circulation, or from the
328 superposition of multiple partially independent modes of variability.

329 In this context, the regional signal should be interpreted as an integrated indicator of
330 large-scale forcing rather than a direct proxy for transport alone. Distinguishing
331 between transport-driven and regime-driven synchronization therefore requires
332 additional process-based analysis, which is addressed in the following sections.

333 **3.2. Differential regional control of PM₁₀ and PM_{2.5}**

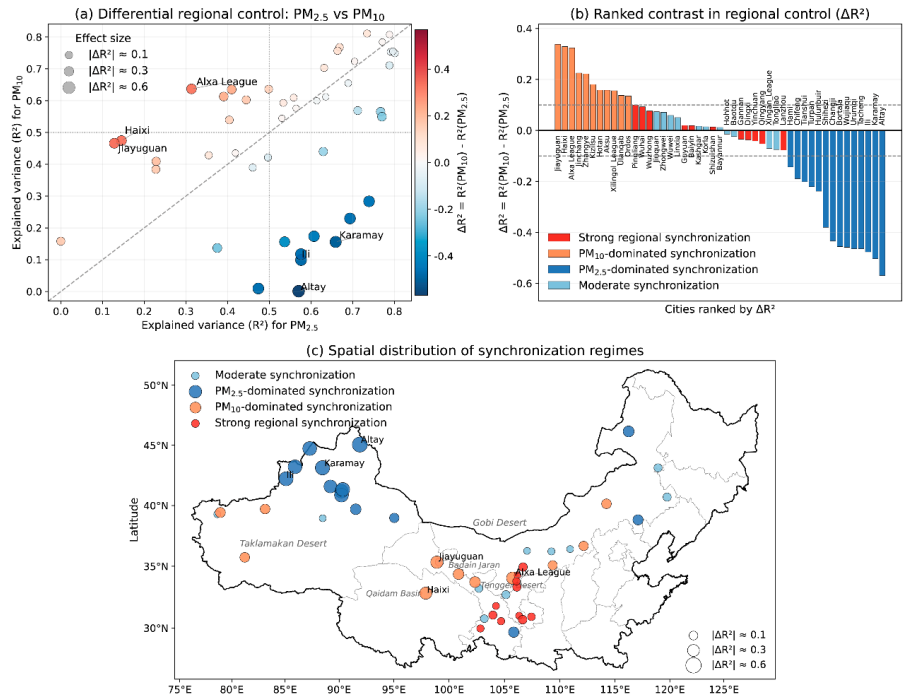
334 The transition from correlation analysis to explained variance decomposition (R^2)
335 enables a quantitative assessment of the contribution of the regional signal to aerosol
336 concentrations and reveals differences in the control mechanisms of PM₁₀ and PM_{2.5}.
337 On average, the regional signal explains a comparable fraction of variability for both
338 particle types, but with fundamentally different spatial structures. The mean R^2 value
339 for PM₁₀ is 0.49 (median 0.57), whereas for PM_{2.5} it is 0.55 (median 0.58). Despite
340 similar aggregate values, the distribution of the difference (ΔR^2) exhibits pronounced
341 asymmetry and substantial inter-city variability (from -0.57 to $+0.34$), indicating the
342 existence of distinct pollution control regimes.

343 At the level of individual cities, differences between PM₁₀ and PM_{2.5} are systematic
344 rather than random (Figure 3a). A substantial fraction of cities is distributed along the
345 line of equality, corresponding to a balanced influence of regional processes on both
346 aerosol types. However, deviations from this line form two contrasting regimes, namely
347 PM₁₀-dominated and PM_{2.5}-dominated control. In the former case ($\Delta R^2 > 0$), the
348 contribution of the regional signal to PM₁₀ variability substantially exceeds that for
349 PM_{2.5}, reflecting the dominance of mineral dust and direct dust emissions. In the latter
350 case ($\Delta R^2 < 0$), the opposite pattern is observed, indicating a stronger role of secondary
351 and anthropogenic processes in shaping PM_{2.5} variability.

352 Ranking of cities by ΔR^2 (Figure 3b) confirms the absence of a single dominant
353 mechanism. The mean ΔR^2 is slightly negative (-0.06), while the median remains close
354 to zero (0.01), indicating an overall balance between the two regimes in the presence of
355 pronounced extremes. In particular, cities such as Jiayuguan ($\Delta R^2 = 0.34$), Alxa League
356 (0.32), and Haixi (0.33) exhibit the strongest PM₁₀ dominance, whereas Altay (-0.57),
357 Karamay (-0.50), and Ili (-0.48) are characterized by strong PM_{2.5} dominance. This



358 structured distribution indicates that inter-city differences are driven by persistent
359 physical processes rather than random fluctuations.



360
361 **Figure 3.** Differential regional control of PM₁₀ and PM_{2.5} across arid cities.
362

363 Haixi Prefecture, located within the Qaidam Basin, represents a particularly
364 illustrative case. Despite arid conditions and proximity to potential dust sources, Haixi
365 exhibits a moderate contribution of the regional signal to PM₁₀ ($R^2 = 0.47$) and an
366 extremely low contribution to PM_{2.5} ($R^2 = 0.15$), resulting in one of the highest ΔR^2
367 values (0.33) among all cities. This pattern indicates the dominance of local dust
368 processes and weak integration into the regional regime, in contrast to more urbanized
369 areas in Gansu and Ningxia. Haixi thus represents a transitional system in which the
370 regional signal weakens and local sources become dominant.

371 The spatial distribution of control regimes (Figure 3c) reveals a clear geographic
372 organization. Cities with strong regional control and balanced contributions of PM₁₀
373 and PM_{2.5} are predominantly located along the Gansu–Ningxia corridor, where
374 conditions favor efficient aerosol transport. PM₁₀-dominated cities are concentrated
375 near desert and semi-desert regions (e.g., Alxa and western areas), reflecting the
376 influence of active dust sources. In contrast, PM_{2.5}-dominated cities are more frequently



found in the northern and eastern parts of the study region, where the role of secondary aerosols and anthropogenic emissions increases.

Overall, the variance decomposition demonstrates that regional synchronization does not imply uniform underlying processes. Instead, it reflects a superposition of distinct, regime-dependent mechanisms governing aerosol variability. The ΔR^2 framework therefore provides a quantitative basis for disentangling these mechanisms and linking observed synchronization to physically interpretable controls, including dust transport and local aerosol formation processes.

3.3. Observed drivers of inter-city heterogeneity in differential regional control

To identify factors associated with inter-city heterogeneity in differential regional control, the ΔR^2 metric was compared with indicators reflecting aerosol composition and geographic positioning relative to major dust source regions (Figure 4).

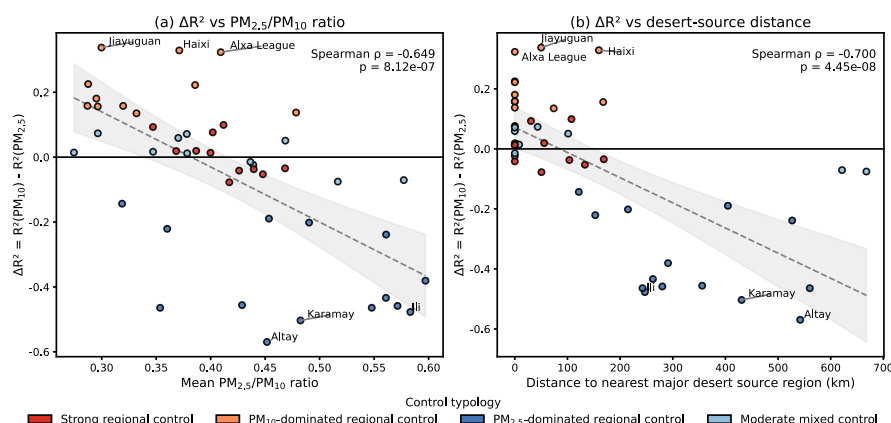


Figure 4. Observed correlates of inter-city heterogeneity in differential regional control.

The strongest statistical relationship is observed with the mean $PM_{2.5}/PM_{10}$ ratio ($\rho = -0.649$, $p < 0.001$; Figure 4a). As the contribution of the fine aerosol fraction increases, ΔR^2 systematically shifts toward negative values, indicating a relatively stronger role of the regional signal in $PM_{2.5}$ compared to PM_{10} . Conversely, lower $PM_{2.5}/PM_{10}$ ratios are associated with positive ΔR^2 values, reflecting enhanced regional control of coarse particles. This relationship highlights aerosol composition as an integrated indicator of the dominant processes shaping particulate variability across cities.

Importantly, a similarly strong relationship is identified with the distance to the nearest major desert source region ($\rho = -0.700$, $p < 0.001$; Figure 4b). Cities located



402 closer to major dust sources tend to exhibit higher ΔR^2 values, consistent with stronger
403 regional control of PM_{10} driven by mineral dust emissions and transport. In contrast,
404 more distant cities show increasingly negative ΔR^2 values, indicating a relative shift
405 toward stronger regional expression in $PM_{2.5}$. Compared to latitude-based
406 representations, this distance-based metric provides a more direct geographic proxy of
407 dust-source exposure and transport influence.

408 Together, these results demonstrate that inter-city heterogeneity in differential
409 regional control is not random but systematically associated with both aerosol
410 composition and spatial positioning relative to major dust-emission regions. The
411 combination of these factors defines a continuum of regimes ranging from dust-
412 dominated systems with strong regional control of PM_{10} to environments where fine
413 particles exhibit stronger regional coherence.

414 It should be noted that the relationship between ΔR^2 and the $PM_{2.5}/PM_{10}$ ratio may
415 include a mathematical component, as both metrics involve characteristics of $PM_{2.5}$
416 variability. However, the consistency of this relationship with an independent
417 geographic gradient, represented by distance to desert sources, supports a physically
418 meaningful interpretation and argues against a purely statistical artifact.

419 Crucially, ΔR^2 should not be interpreted as a direct measure of aerosol transport, but
420 rather as a response metric that quantifies how urban particulate matter variability
421 reflects regional forcing. In this framework, high ΔR^2 values indicate systems in which
422 coarse particles (PM_{10}) respond more strongly to regional-scale variability, consistent
423 with transport-dominated dust regimes, whereas negative ΔR^2 values indicate a relative
424 enhancement of the regional signal in fine particles ($PM_{2.5}$), reflecting accumulation
425 and secondary processes. Thus, ΔR^2 captures the regime-dependent realization of a
426 common regional signal within different urban environments, rather than the transport
427 process itself.

428 Additional relationships with latitude and mean PM_{10} concentration are presented in
429 the Supplement (Figure S3). While both variables show statistically significant
430 associations with ΔR^2 , they are weaker than those observed for aerosol composition
431 and desert-source distance, and likely reflect indirect or proxy effects related to
432 geographic positioning and aerosol characteristics rather than primary controls.

433 3.4. Ventilation versus accumulation regimes in regional particulate variability



To refine the physical interpretation of the regional signal, an additional analysis of the relationship between aerosol concentrations and wind characteristics was performed. Analysis of mean relationships between regional PM values and wind speed (Figure S1) shows weak and ambiguous associations for both PM₁₀ and PM_{2.5}. The observed correlations do not exhibit a consistent increase in concentrations with increasing wind speed, and the use of absolute anomalies also does not reveal a clear positive relationship. This indicates that mean relationships have limited explanatory power for interpreting pollution formation mechanisms.

To address this limitation, an event-based analysis focusing on extreme PM₁₀ concentrations (upper 10% of the distribution) was conducted, allowing identification of regimes associated with the highest pollution levels (Figure 5). The results show that extreme concentrations occur across a wide range of wind speeds and are not confined to low-wind conditions (Figure 5a). Moreover, the probability of extreme events does not exhibit a monotonic dependence on wind speed (Figure 5b), ruling out interpretation in terms of a single dominant mechanism.

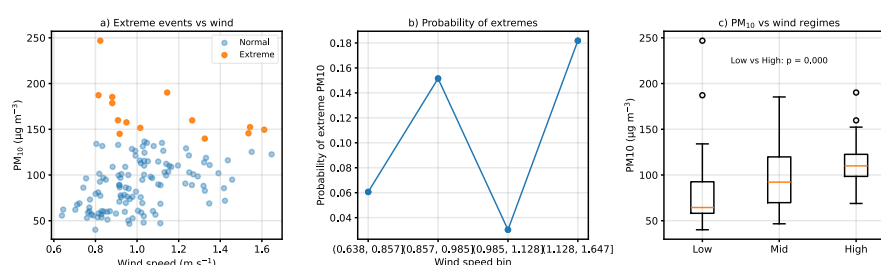


Figure 5. Wind-regime dependence of extreme particulate pollution in arid regions.

Comparison of concentration distributions under different wind regimes (Figure 5c) reveals statistically significant differences between low- and high-wind conditions (Mann–Whitney U test, $p < 0.05$). Under low-wind conditions, higher concentrations are observed on average, indicating accumulation of aerosols due to limited ventilation. At the same time, extreme values are also recorded at higher wind speeds, which are likely associated with episodes of intensified mineral dust transport.

Overall, these results demonstrate the existence of two physically distinct pollution formation regimes. The first regime is associated with accumulation under low-wind conditions and reduced ventilation, while the second regime is driven by enhanced transport of dust aerosols under higher wind speeds. In this framework, atmospheric circulation acts as a dual-role driver, simultaneously governing both accumulation and



463 transport processes. This regime-dependent behavior provides a mechanistic
464 explanation for the coexistence of strong regional synchronization and substantial inter-
465 city heterogeneity, linking statistical patterns directly to underlying physical processes.

466 **4. Discussion**

467 The results demonstrate that regional synchronization of particulate matter
468 concentrations in arid urban systems cannot be interpreted solely as a direct
469 consequence of aerosol transport. The absence of a monotonic relationship between
470 wind speed and pollution levels, together with the occurrence of extreme PM₁₀
471 concentrations across a wide range of wind conditions, indicates that variability is
472 governed by a regime-dependent structure rather than a single dominant mechanism.
473 This interpretation is consistent with previous studies showing that atmospheric
474 circulation can simultaneously regulate pollutant transport and boundary-layer
475 ventilation, producing nonlinear responses under contrasting synoptic conditions (Miao
476 et al., 2018; Huang et al., 2018; Liao et al., 2024).

477 Two physically distinct regimes emerge from the analysis. The first is associated
478 with accumulation under low-wind conditions and reduced ventilation, leading to
479 elevated concentrations of both coarse and fine particles. The second corresponds to
480 transport-driven conditions, in which stronger winds enhance the emission and long-
481 range transport of mineral dust, producing short-term peaks in PM₁₀. Such a dual role
482 of atmospheric circulation is consistent with the established understanding that wind
483 can either suppress pollution through dispersion or intensify particulate loading through
484 dust mobilization, particularly in arid environments (Knippertz and Todd, 2012; Shao
485 et al., 2011; Wang et al., 2025). This duality implies that similar levels of regional
486 synchronization may arise from fundamentally different physical processes.

487 The comparison between PM₁₀ and PM_{2.5} further reveals that the regional signal is
488 not uniformly expressed across aerosol fractions but undergoes a regime-dependent
489 transformation. Coarse particles are primarily associated with mineral dust emission
490 and transport (Knippertz and Todd, 2012; Kok et al., 2012; Wang et al., 2025), whereas
491 fine particles more strongly reflect secondary formation and anthropogenic
492 contributions (Huang et al., 2014b; Zhang and Cao, 2015; Zhou et al., 2019).
493 Observations from northwestern China indicate that dust and anthropogenic aerosols
494 are frequently mixed near source regions, producing fraction-dependent responses in
495 urban environments (Li et al., 2010; Fomba et al., 2024). As a result, the same regional



496 forcing can manifest differently depending on aerosol composition, generating
497 substantial inter-city variability despite overall synchronization.

498 In this context, the ΔR^2 metric provides a quantitative representation of how urban
499 systems respond to regional forcing. Importantly, ΔR^2 should not be interpreted as a
500 direct proxy for aerosol transport, but rather as a response metric describing the relative
501 sensitivity of different particle fractions to regional variability. Positive ΔR^2 values
502 indicate systems in which coarse particles respond more strongly to regional forcing,
503 consistent with dust-dominated regimes, whereas negative values reflect a relatively
504 stronger regional expression in fine particles, associated with accumulation and
505 secondary processes. This interpretation shifts the focus from transport itself to the
506 realization of regional forcing within urban environments.

507 The observed relationships indicate that this response is systematically associated
508 with both aerosol composition and geographic context. The $PM_{2.5}/PM_{10}$ ratio serves as
509 an integrated indicator of dominant physicochemical processes, while proximity to
510 major desert source regions constrains exposure to dust emissions and transport
511 pathways. This interpretation is consistent with previous studies demonstrating that
512 dust contributions to urban aerosol loading depend strongly on distance from source
513 regions and transport pathways linking the Taklamakan and Gobi deserts to receptor
514 areas across northern China (Chen et al., 2017a; Yu et al., 2019; Han et al., 2025). Their
515 combined influence defines a continuum of regimes ranging from dust-dominated
516 systems near source regions to environments in which fine particles exhibit stronger
517 regional coherence.

518 Although the relationship between ΔR^2 and aerosol composition may partly reflect
519 mathematical coupling, its consistency with an independent geographic gradient
520 supports a physically meaningful interpretation. Additional variables such as latitude
521 and mean PM_{10} concentration, which show weaker associations, are therefore better
522 interpreted as proxy indicators rather than primary controls. This distinction highlights
523 the importance of separating physically meaningful drivers from secondary correlations
524 when analyzing large-scale observational datasets.

525 The existence of transitional regimes further indicates that urban systems form a
526 continuous spectrum rather than discrete categories. Such heterogeneity is consistent
527 with previous work demonstrating strong spatial variability in aerosol sources and
528 mixing processes across arid regions (Chen et al., 2017a; Kang et al., 2017; Li et al.,
529 2025; Zhang et al., 2025). In this framework, transitional systems illustrate how



530 regional forcing can be expressed predominantly through one aerosol fraction while
531 remaining weakly coupled in another.

532 It is important to emphasize that the regional signal used in this study is a statistical
533 construct and does not represent a direct measurement of physical transport processes.
534 The observed synchronization therefore reflects coordinated variability arising from
535 multiple factors, including atmospheric circulation, dust emissions, and large-scale
536 changes in anthropogenic activity. In addition, the absence of explicit detrending
537 implies that part of the detected coherence may be associated with long-term emission
538 reductions and policy-driven improvements in air quality across China (Cheng et al.,
539 2019; Li et al., 2019; Zhang et al., 2019).

540 From a practical perspective, the results indicate that air-quality management in arid
541 regions must explicitly account for regime-dependent pollution formation. Under
542 accumulation-dominated conditions, mitigation should focus on local emissions and
543 processes controlling urban ventilation. Under transport-dominated conditions, greater
544 emphasis should be placed on dust source regions and land–atmosphere interactions
545 governing dust mobilization (Shao et al., 2011; Tian et al., 2021). The coexistence of
546 these regimes implies that uniform control strategies are unlikely to be effective in dust-
547 influenced environments.

548 5. Conclusions

549 This study provides a quantitative analysis of inter-city synchronization of
550 particulate matter across 47 cities in arid and semi-arid regions of northern China for
551 the period 2015–2025. The proposed framework, based on explained variance and the
552 differential regional control metric (ΔR^2), enables the separation of regional and local
553 contributions to PM_{10} and $PM_{2.5}$ variability.

554 The results show that air pollution in arid urban systems is characterized by strong
555 regional synchronization driven by large-scale forcing, while simultaneously exhibiting
556 pronounced spatial heterogeneity in underlying mechanisms. This indicates that
557 observed coherence cannot be attributed to a single process but instead reflects the
558 coexistence of multiple pollution formation regimes.

559 A key finding is that the relationship between aerosol concentrations and wind speed
560 is nonlinear and regime-dependent. Extreme PM_{10} concentrations occur under both low-
561 and high-wind conditions, reflecting two physically distinct regimes: accumulation
562 under weak ventilation and transport-driven dust events. This demonstrates that



563 atmospheric circulation acts as a dual-role driver, simultaneously regulating both
564 accumulation and transport processes.

565 The analysis further shows that inter-city differences in regional control are
566 systematically associated with aerosol composition and geographic positioning relative
567 to major dust sources. The $PM_{2.5}/PM_{10}$ ratio reflects the dominant physicochemical
568 processes, while proximity to desert source regions constrains exposure to dust
569 emissions and transport pathways. Together, these factors define a continuum of urban
570 response regimes across arid environments.

571 Importantly, the ΔR^2 metric should be interpreted as a response metric rather than a
572 direct proxy for aerosol transport. It quantifies how urban particulate matter variability
573 reflects regional forcing under different environmental conditions, thereby linking
574 statistical patterns to physically interpretable regimes.

575 Methodologically, the results demonstrate that combining variance decomposition
576 with regime-based analysis provides a more robust framework for interpreting regional
577 air pollution dynamics than correlation-based approaches alone.

578 From a practical perspective, the findings highlight the need for regime-specific air
579 quality management strategies in arid regions. Measures targeting dust emission and
580 transport should be complemented by controls on pollution accumulation under weak
581 ventilation conditions. Future work should focus on integrating independent transport
582 diagnostics and further constraining the physical processes underlying the identified
583 regimes.

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587 **Conflict of Interest**

588 The authors declare no conflicts of interest relevant to this study.

589 **Data Availability Statement**

590 Air quality data used in this study are publicly available from the China National
591 Environmental Monitoring Center (CNEMC, <https://www.cnemc.cn/>), including daily
592 concentrations of $PM_{2.5}$, PM_{10} for 47 cities in northern China during 2015–2025. Wind
593 speed data at 10 m height were obtained from the European Centre for Medium-Range



594 Weather Forecasts ERA5 reanalysis dataset, accessed via the Copernicus Climate Data
595 Store (<https://cds.climate.copernicus.eu/>). City location data were compiled from
596 publicly available geographic datasets. The processed datasets generated in this study,
597 including monthly aggregated time series, regional signal estimates, and ΔR^2 metrics,
598 are available from the corresponding author upon reasonable request. All code used for
599 data processing, statistical analysis, and figure generation is available from the
600 corresponding author upon request.

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