



1 **Linking the environmental factors and heterotrophic bacteria to the variability of carbon isotope**
2 **($\delta^{13}\text{C}$) across crustacean zooplankton in shallow freshwater ecosystems**

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15



16 **Abstract**

17 Planktonic crustaceans are important part of carbon cycling in in shallow, temperate freshwater
18 habitats. The objective of the study was to examine the link between selected environmental variables,
19 densities of various groups of bacteria and $\delta^{13}\text{C}$ of calanoid copepods, cyclopoid copepods and *Daphnia*
20 sp. Generalized Linear Models (GLMs) with stepwise-selected environmental predictors showed that
21 $\delta^{13}\text{C}$ of calanoid copepods was positively correlated with the Day of the year and Other bacteria density
22 and negatively correlated with NH_4^+ and NO_3^- concentrations. Cyclopoid copepods' $\delta^{13}\text{C}$ was correlated
23 exclusively with water temperature, which showed a positive effect on the stable carbon isotopes ratio.
24 GLM for *Daphnia* sp. was the most complex and retained the Day of the year, Other bacteria density
25 and PO_4^{3-} as positively correlated predictors, while β -proteobacteria NH_4^+ and NO_3^- were included in
26 the model as negatively correlated predictors for $\delta^{13}\text{C}$. Our study highlights the complex impact of
27 environmental conditions, microbial processes and functional trait-driven responses on $\delta^{13}\text{C}$ values in
28 planktonic crustaceans. Thanks to integration of plankton ecology, microbiology and the analysis of
29 stable isotopes the study improves the understanding of carbon flow in shallow freshwater ecosystems

30 **1. Introduction**

31 Functioning of freshwater ecosystems is strongly reliant on pelagic food webs with planktonic
32 crustaceans (copepods and cladocerans; members of zooplankton community) interconnecting
33 primary producers (bacteria, algae) and higher trophic levels (invertebrates, fish). Filter-feeding
34 cladocerans of genus *Daphnia* are widely accepted keystone taxon, constituting phytoplankton-
35 herbivorous zooplankton link of energy transfer (Ger et al., 2016; Sommer et al., 1986; Straile, 2015).
36 *Daphnia* is also recognized as an efficient bacteria grazer which obviously is an advantage, under
37 shortage of algal food (Martin-Creuzburg et al., 2011). However, currently observed global climatic
38 changes, forcing alterations in freshwater food webs, may lead to loss of significance of cladocerans of
39 genus *Daphnia*, as conduits of energy flow and to destabilization of freshwater food webs (revised by
40 Wojtal-Frankiewicz, 2012). Findings of de Senerpont Domis et al. (2007) suggest that far-reaching
41 climate warming (+6°C) may trigger phenological mismatches between *Daphnia* and its prey. Thus, the
42 'center of gravity' in functioning of freshwater food webs may be switched to copepods which
43 frequently are an abundant component of zooplankton community and beside ability of suppressing
44 biomass of phytoplankton, extensively forage on protozoans (De Meester et al., 2023; Krztoń et al.,
45 2022; Sommer et al., 2003).

46 Planktonic crustaceans are important part of "fast carbon cycle", which refers to carbon cycling in
47 biosphere. These animals rely on three distinct carbon sources: carbon dioxide (CO_2), dissolved organic
48 carbon (DOC), and methane (CH_4), mobilized by phytoplankton, heterotrophic bacteria and



49 methanotrophic bacteria (methane-oxidizing bacteria; MOB), respectively (Jones and Grey, 2011).
50 Substantial part of the carbon flowing to crustaceans may originate from bacteria (Wylie and Currie,
51 1991), which strongly contribute to biogeochemical cycle of carbon (Baltar et al., 2015). This is a reason
52 why bacteria are recognized as an important part of in the energy flow through the food web. Some
53 authors showed that the bacteria's significance vary depending on temporary environmental
54 conditions (Hanashiro et al., 2023; Louati et al., 2023). Study of Hall and Cotner (2007) demonstrates
55 strong dependence of bacteria-mediated carbon cycling on temperature and nutrients availability,
56 which is a significant observation given the increase in temperature caused by climate change.

57 The detailed assessment of carbon sources in the diet of planktonic crustaceans and studies over the
58 pathways of carbon flow in pelagic freshwater food webs remain important topics in the field
59 hydrobiological researches. Therefore, we measured stable carbon isotope ratio ($\delta^{13}\text{C}$) which is a
60 powerful indicator used for tracing carbon flow, and its values vary in tissues of diverse organisms
61 depending on the pathway that carbon is incorporated into the food web (Hobson, 2023). One of the
62 pathways can be constituted by bacteria (microbial loop) that might provide nutritional benefits for
63 crustacean grazers, feeding both directly (cladocerans) and indirectly via transitional consumers
64 (copepods and cladocerans) (Wenzel et al., 2021).

65 Previous studies showed that isotopic niches of planktonic crustaceans increase in area and overlap
66 during cyanobacterial bloom compared to pre-bloom conditions (Krztoń et al., 2025). Here, we follow
67 this background and aim to explore variability of $\delta^{13}\text{C}$ in planktonic crustacean communities.

68 The objective of the study was to examine the link between selected environmental variables, densities
69 of various groups of bacteria and $\delta^{13}\text{C}$ of planktonic crustaceans (calanoid copepods, cyclopoid
70 copepods and *Daphnia* sp.) in shallow, temperate freshwater habitats. Based on the fact that copepods
71 are not likely to feed directly on bacteria, we hypothesized that among studied crustaceans, only
72 *Daphnia's* $\delta^{13}\text{C}$ is driven by densities of various bacteria groups. In turn, we expected $\delta^{13}\text{C}$ of copepod
73 to be dependent on physical and/or chemical parameters of water. Addressing these research
74 questions we aimed to explore the role of planktonic crustaceans in biogeochemical cycles in shallow
75 freshwater systems.

76 **2. Materials and methods**

77 *2.1 Study area*

78 The study was conducted in five shallow, eutrophic water bodies: three oxbow lakes of the Vistula River:
79 Tyniec 1 (50° 01'47" N, 19° 49' 39.8" E), Tyniec 2 (50°01'28.1" N, 19°48'47.7" E), Wołowice 1
80 (49°59'18.0"N 19°45'01.3"E) and two artificial ponds: Podkamycze 1 (50° 05'11" N, 19° 50' 01.6" E) and



81 Podkamycze 2 (50° 04' 59.6" N, 19° 50'05.4" E) located in the area of Cracow and nearby (Southern
82 Poland).

83 *2.2 Sampling and environmental factors analyses*

84 Samplings were carried out bi-weekly in the vegetation season (April - October) in 2019. Samplings
85 were performed at the central point of each waterbody with the use of a 5 L sampler. Integrated water
86 samples covered the whole column of the water (sampling interval: each 0.5 m starting from surface
87 to the bottom). Each sample included: physico-chemical parameters of water (temperature, pH,
88 conductivity, PO_4^{3-} , NH_4^+ , NO_3^- concentration), bacterioplankton composition and density and samples
89 for analyses of stable isotopes of carbon ($\delta^{13}\text{C}$) in tissues of planktonic crustaceans (see below). In
90 addition to collecting samples, maximum depths and transparency were measured in situ, each time, in
91 each reservoir. In total 75 environmental samples were collected.

92 Water transparency was measured using Secchi disc. Water temperature, pH and conductivity were
93 measured using YSI ProDSS sonde. Samples for ions concentrations (PO_4^{3-} , NH_4^+ , NO_3^-) were cold-stored
94 and immediately transported to laboratory of the Institute of Nature Conservation, Polish Academy of
95 Sciences. The samples were determined in the laboratory using Dionex ion chromatograph (DIONEX,
96 IC25 Ion Chromatograph; ICS-1000, Sunnyvale, CA, USA).

97 *2.3 Bacterial communities analysis*

98 The water samples (vol. 10mL) for bacteria community analysis were fixed on filters (0.2 μm pore size;
99 type GTTP; Millipore). Bacterial community composition was determined using fluorescent in situ
100 hybridization (FISH) with Cy3-labelled oligonucleotide probes, following the protocols described by
101 Glöckner et al. (1996) and Pernthaler et al. (2001). The probe set targeted the major phylogenetic
102 groups of Proteobacteria, including Alpha- (α -), Beta- (β -), Gamma- (γ -), and Delta- (δ -) subclasses.
103 Probe target sites and hybridization conditions followed the recommendations of Amann et al. (1990),
104 Daims et al. (1999), Alm et al. (1996), Manz et al. (1992) and (Lücker et al., 2007). To assess
105 autofluorescence and non-specific probe binding, all samples were simultaneously hybridized with the
106 NON338 probe, complementary to EUB338. The proportion of non-specifically stained cells did not
107 exceed 5% in any sample. Filter sections were examined using an epifluorescence microscope (Olympus
108 BX61) equipped with a $\times 100$ oil-immersion objective, UV illumination, and a CCD camera with DAPI and
109 Cy3 filter sets (Olympus, Poland). Images were processed using Cell F software (Olympus, Poland). The
110 abundance of FISH-positive cells was quantified in triplicate by counting at least 1,000 DAPI-stained
111 cells per sample.

112 *2.4 Analysis of stable isotopes*



113 The samples for analysis of stable isotopes of carbon ($\delta^{13}\text{C}$) in tissues of planktonic crustaceans were
114 collected using vertical hauls performed with plankton net of mesh size 50 μm . In the laboratory, using
115 a light microscope and micromanipulator, samples were segregated to the following fractions: 1)
116 Calanoid copepods (N = 36), 2) cyclopoid copepods (N = 29), 3) and *Daphnia* sp. (N = 12). Single samples
117 consisted of 30 - 100 adult specimens. In total, $\delta^{13}\text{C}$ were determined in 77 samples of planktonic
118 crustaceans.

119 The stable carbon isotope analyses were conducted at the Stable Isotope Laboratory of the Institute of
120 Geological Sciences, Polish Academy of Sciences (Warsaw, Poland), using a Flash 1112 HT elemental
121 analyzer coupled to a Delta V Advantage isotope ratio mass spectrometer (Thermo Scientific, Waltham,
122 MA, USA) operating in continuous-flow mode with helium as the carrier gas. Carbon isotope ratios are
123 expressed in δ notation as per mil (‰) relative to the Vienna Pee Dee Belemnite (VPDB) standard.
124 Duplicate measurements were calibrated against three international reference materials: USGS 41a (L-
125 glutamic acid, $\delta^{13}\text{C} = +36.55 \pm 0.08$ ‰), USGS 40 (L-glutamic acid, $\delta^{13}\text{C} = -26.389 \pm 0.042$ ‰) and IAEA
126 600 (caffeine, $\delta^{13}\text{C} = -27.771 \pm 0.043$ ‰). Instrument performance and drift were monitored using USGS
127 65 (glycine, $\delta^{13}\text{C} = -20.29 \pm 0.04$ ‰) as an internal laboratory standard. Analytical precision and
128 reproducibility, based on repeated measurements of the laboratory standard during each analytical
129 run, were better than ± 0.1 ‰ (1σ) for $\delta^{13}\text{C}$. Carbon-to-nitrogen (C:N) ratios were determined from the
130 same analyses with an accuracy of ± 0.1 . Because lipids are depleted in ^{13}C and may bias $\delta^{13}\text{C}$
131 measurements in planktonic crustaceans, carbon isotope values were corrected for lipid content using
132 the elemental C:N ratio following the equation proposed by (Syväranta and Rautio, 2010).

133 2.5 Statistical analyses

134 All statistical analyses and data visualizations were performed with R V. 4.2.2 and RStudio statistical
135 software (R Core Team, 2022). Values of $\delta^{13}\text{C}$ were tested for differences among studied groups of
136 planktonic crustaceans using ANOVA ('stats' package; R Core Team, 2022). Tukey's Honest Significant
137 Difference test (HSD; 'stats' package; R Core Team, 2022) was applied in order to calculate confidence
138 intervals between means of $\delta^{13}\text{C}$ among particular crustacean groups.

139 To identify environmental and microbial drivers of carbon isotopic variation ($\delta^{13}\text{C}$) in zooplankton,
140 generalized linear models (GLMs) with Gaussian error distribution and identity link function were fitted
141 separately for Calanoid copepods, Cyclopoid copepods, and *Daphnia* sp. For each group, the initial (full)
142 model included sampling day, bacterial abundance variables (total Proteobacteria, α -proteobacteria,
143 β -proteobacteria, γ -proteobacteria, δ -proteobacteria, and the group Other bacteria), water
144 temperature, and nutrient concentrations (NH_4^+ , NO_3^- , PO_4^{3-}) as explanatory variables. Prior to model
145 selection, multicollinearity among predictors was assessed using variance inflation factors (VIFs; 'car'



146 package, Fox et al., 2019), which remained below commonly accepted thresholds. Model selection was
 147 performed using stepwise Akaike Information Criterion (AIC) optimization with both forward and
 148 backward selection (*stepAIC*, Venables and Ripley, 2013) to identify the most parsimonious models
 149 explaining variation in $\delta^{13}\text{C}$. The results of stepwise regression were visualized with use of ‘visreg’
 150 package (Breheny and Burchett, 2017).

151 **3. Results**

152 *3.1 Environmental variables*

153 Water temperature in studied lakes ranged from 10.8 to 25.6 °C (mean: 18.2°C), reaching highest values
 154 in July and August. Concentrations of NO_3^- ranged between <0.01 and 18.8 mg/L (mean: 4.6 mg/L).
 155 Highest mean concentrations of NO_3^- were detected in the period of beginning of April to end of May,
 156 while the lowest concentrations were noted between beginning of June to mid-August. PO_4^{3-}
 157 concentrations measured in collected samples ranged between <0.01 and 0.44 mg/L (mean: 0.42
 158 mg/L). The highest mean concentrations of PO_4^{3-} were detected in the period of mid-May to end of
 159 the August. Concentrations of NH_4^+ ranged from <0.01 to 0.99 mg/L (mean: 0.18 mg/L). Highest mean
 160 concentrations of NH_4^+ were detected from beginning of May to end of June, and then from beginning
 161 of September to end of October. Lake specific ranges, mean values, standard deviations and coefficients
 162 of variance of studied parameters are given in Table 1.

163

164 Table 1. Lake specific ranges, mean values, standard deviations (SD) and coefficients of variance (CV) of
 165 water temperature and concentrations of NO_3^- , PO_4^{3-} and NH_4^+ .

		Waterbody				
		Tyniec 1	Tyniec 2	Wołowice 1	Podkamycze 1	Podkamycze 2
Temperature [°C]	Mean	18.89	19.73	14.85	17.27	19.86
	Range (min- max)	11.8 – 25.5	11.2 – 25.5	10.8 – 18.3	10.9 – 23.7	12.1 – 25.6
	SD	4.62	5.03	2.16	4.01	5.19
	CV	24.43	25.5	14.56	23.24	26.13
	Mean	1.44	1.39	8.99	11.43	1.99



NO ₃ ⁻ [mg/L]	Range (min- max)	0.003 – 4.91	0.02 – 6.17	0.05 – 18.83	9.43 – 15.03	0.19 – 6.13
	SD	1.5	1.59	6.44	1.47	2.21
	CV	103.95	114.59	71.76	12.88	110.31
PO ₄ ³⁻ [mg/L]	Mean	0.02	0.05	0.077	0.08	0.003
	Range (min- max)	0 – 0.1	0 – 0.11	0 – 0.44	0 – 0.18	0 – 0.02
	SD	0.033	0.04	0.12	0.07	0.003
	CV	205.31	80.72	154.89	83.52	126.87
NH ₄ ⁺ [mg/L]	Mean	0.1	0.2	0.45	0.03	0.02
	Range (min- max)	0.01 – 0.2	0.001 – 0.99	0.06 – 0.97	0.003 – 0.09	0 – 0.09
	SD	0.06	0.29	0.32	0.03	0.02
	CV	59.96	145.29	71.24	104.62	98.81

166

167 *3.2 Bacterial communities*

168 Analyses of bacteria revealed presence of following groups of bacteria: Proteobacteria, α-
 169 proteobacteria, β-proteobacteria, γ-proteobacteria, δ-proteobacteria and Other bacteria. Total
 170 number of planktonic bacteria cells ranged from 5.55×10^6 to 8.58×10^7 cells per milliliter. The largest
 171 mean densities were observed for Proteobacteria and Other bacteria, while the lowest for α-
 172 proteobacteria group. The largest variability of densities (based on CV) was found for the group of
 173 Other bacteria while the lowest for β-proteobacteria (Tab. 2).

174 Table 2. Range (Min-Max), Mean, standard deviation (SD) and coefficient of variation (CV) of densities
 175 of studied bacteria groups in samples collected through the sampling season

	Range (Min-Max) [Cells/mL]	Mean [Cells/mL]	SD [Cells/mL]	CV
Proteobacteria	2.94×10^6 - 5.98×10^7	1.49×10^7	9.65×10^6	0.65



α -proteobacteria	$1 \times 10^5 - 3.01 \times 10^6$	1.03×10^6	7.43×10^5	0.72
β -proteobacteria	$4.34 \times 10^5 - 5.45 \times 10^6$	2.15×10^6	1.02×10^6	0.47
γ -proteobacteria	$1.34 \times 10^5 - 5.77 \times 10^6$	1.51×10^6	1.18×10^6	0.78
δ -proteobacteria	$2.11 \times 10^5 - 5.70 \times 10^6$	2.45×10^6	1.49×10^6	0.61
Other bacteria	$2.37 \times 10^5 - 4.80 \times 10^7$	9.57×10^6	1.01×10^7	1.06
Total bacteria	$5.55 \times 10^6 - 8.58 \times 10^7$	3.16×10^7	1.68×10^7	0.53

176

177

178 3.3 Crustacean zooplankton

179 Composition and distribution of studied planktonic crustaceans varied throughout the sampling period
 180 (Figure 1). Calanoid copepods were found in all of the samples collected throughout the year (beginning
 181 of April - end of October). Cyclopoid copepods and *Daphnia* sp. showed similar frequency, with
 182 exception for the beginning of April (when *Daphnia* sp. was absent) and mid-July (when cyclopoid
 183 copepods were absent).

184 3.4 Stable isotopes

185 Mean values of $\delta^{13}\text{C}$ were found in calanoid copepods accounted for -31.8‰ with range between -
 186 41.5‰ to -23.7. Mean $\delta^{13}\text{C}$ values of cyclopoid copepods was -30.6‰ and range: -40‰ to -22.2‰).
 187 $\delta^{13}\text{C}$ measurements of *Daphnia* sp. showed mean value of -31.2‰ and range between -42.5‰ and -
 188 24.4‰)

189 No statistically significant differences among crustacean groups in mean $\delta^{13}\text{C}$ values were found among
 190 Calanoid copepods, Cyclopoid copepods ($p < 0.01$) and *Daphnia* sp. (Supplementary material, Tables S1
 191 and S2).

192 For calanoid copepods, the full GLM model explained a substantial proportion of deviance relative to
 193 the null model. Stepwise AIC selection retained a reduced model including day of the year, β -
 194 proteobacteria, Other bacteria, and concentrations of NH_4^+ , NO_3^- , PO_4^{3-} . $\delta^{13}\text{C}$ values in calanoid
 195 copepods increased significantly over time, as indicated by a positive effect of day of the year. NH_4^+ and
 196 NO_3^- concentrations were strong predictors, showing significant negative relationships with $\delta^{13}\text{C}$,



197 whereas PO_4^{3-} exhibited a weak positive effect. Among bacterial predictors, β -proteobacteria showed
 198 a marginal negative association with $\delta^{13}\text{C}$, while Other bacteria were positively associated. Remaining
 199 tested bacterial groups as well as water temperature did not significantly improve model fit and were
 200 excluded during model selection (Tab. 3.; Fig. 2.).

201 Tab. 3. Results of Generalized Linear Model with stepwise-selected environmental predictors for
 202 Calanoid copepods, Cyclopoid copepods and *Daphnia* sp. $\delta^{13}\text{C}$ values (dependent variable)

		Estimate	Std. Error	T value	P value
Calanoid copepods	Intercept	-34.60	2.332	-14.84	<0.001
	Day of the year	2.97×10^{-2}	9.45×10^{-3}	3.14	<0.01
	β -proteobacteria density	-1.05×10^{-6}	5.83×10^{-7}	-1.81	0.079
	Other bacteria density	1.29×10^{-7}	6.34×10^{-8}	2.03	<0.05
	NH_4^+	-7.86	2.54	-3.10	<0.01
	NO_3^-	-3.17×10^{-1}	1.18×10^{-1}	-2.70	<0.05
	PO_4^{3-}	1.58×10^1	8.51	1.86	0.071
Cyclopoid copepods	Intercept	-38.83	2.40	-16.14	<0.001
	Day of the year	0.015	0.008	1.90	0.064
	Water temperature	0.30	0.115	2.61	<0.05
<i>Daphnia</i> sp.	Intercept	-41.36	6.49	-6.37	<0.001
	Day of the year	4.59×10^{-2}	1.80×10^{-2}	2.55	<0.05
	Proteobacteria	7.22×10^{-8}	6.71×10^{-8}	1.08	0.300
	α -proteobacteria	2.31×10^{-6}	1.17×10^{-6}	1.97	0.069
	β -proteobacteria	-3.21×10^{-6}	1.23×10^{-6}	-2.61	<0.05
	γ -proteobacteria	-1.68×10^{-6}	9.15×10^{-7}	-1.83	0.088
	δ -proteobacteria	1.17×10^{-6}	6.56×10^{-7}	1.79	0.095
	Other bacteria density	2.90×10^{-7}	8.15×10^{-8}	3.56	<0.01
	Water temperature	2.41×10^{-1}	2.03×10^{-1}	1.19	0.253
	NH_4^+	-1.32×10^1	4.33	-3.06	<0.01
	NO_3^-	-4.66×10^{-1}	2.05×10^{-1}	-2.28	<0.05
	PO_4^{3-}	4.68×10^1	1.74×10^1	2.69	<0.05

203

204 In contrast, $\delta^{13}\text{C}$ variation in cyclopoid copepods was weakly explained by the full model, and stepwise
 205 AIC selection resulted in a highly simplified final model including only day of the year and water
 206 temperature. Cyclopoid $\delta^{13}\text{C}$ showed a significant positive relationship with water temperature,



207 whereas the effect of day of the year was marginal. Neither bacterial groups abundances nor nutrient
208 concentrations contributed significantly to explaining variation in cyclopoid $\delta^{13}\text{C}$, and all were removed
209 during model selection (Tab. 3.; Fig. 3.).

210 For *Daphnia* sp., the full model was retained as the best-supported model according to AIC, indicating
211 strong explanatory power without further simplification. *Daphnia* sp. $\delta^{13}\text{C}$ increased significantly with
212 the day of the year. β -proteobacteria abundance was negatively associated with $\delta^{13}\text{C}$, whereas the
213 “Other bacteria” showed a strong positive effect. Positive α -, and δ -proteobacteria effects and negative
214 γ -proteobacteria effect remained marginal in final model.

215 Nutrient concentrations were also significant predictors, with negative effects of NH_4^+ and NO_3^-
216 concentrations and a strong positive effect of PO_4^{3-} concentration. Proteobacteria abundance and
217 water temperature did not significantly explain additional variation in *Daphnia* sp. $\delta^{13}\text{C}$ (Tab. 3.; Fig. 4.).

218 4. Discussion

219 During the last few decades great progress has been made in understanding carbon cycling in
220 freshwaters (Tranvik et al., 2018), but it is still not entirely understood. Our study is another step
221 forward, as we provide an evidence for environmental factors – bacteria – zooplankton link, which
222 significantly contributes to the understanding of carbon flux within planktonic networks. Values of $\delta^{13}\text{C}$
223 in grazers community (zooplankton) is strongly influenced by $\delta^{13}\text{C}$ values of their food sources (Tiselius
224 and Fransson, 2016). However, conducted analyses showed that the variability of $\delta^{13}\text{C}$ values in
225 planktonic crustaceans was driven by distinct environmental and microbial factors and the set of the
226 factors was ultimately linked to specific crustacean group. Planktonic crustaceans bear diverse
227 functional traits (e.g. the method of food acquisition) and therefore their interaction with the
228 environment and response for its’ changes is different (De Meester et al., 2023).

229 Day of the year was a significant predictor, showing positive effects on $\delta^{13}\text{C}$ values of calanoid copepods
230 and *Daphnia* sp. In case of $\delta^{13}\text{C}$ values of cyclopoid copepods, day of the year was retained as a
231 predictor (with positive effect) within the final model, however its’ effect remained marginal. This
232 finding corroborates with similar studies addressing the seasonality of $\delta^{13}\text{C}$ in freshwater
233 phytoplankton (Lammers et al., 2017), herbivorous zooplankter *Daphnia middendorffiana* (Gu et al.,
234 1999) and particulate organic matter (POM; Moschen et al., 2009). To the certain extent this can be an
235 effect of water temperature but also the seasonal succession in source of food, e.g. phytoplankton
236 community since algal $\delta^{13}\text{C}$ values greatly influence grazers’ isotope signal.

237 However, the only factor that positively affected $\delta^{13}\text{C}$ values of cyclopoid copepods, in the present
238 studies, was water temperature. Copepods tend to decrease their body size with increasing



239 temperatures and response of specific taxa to this effect is dependent on food uptake method (Horne
240 et al., 2016). The response of cyclopoids is weaker compared to calanoids (Horne et al., 2016), however
241 any decrease in body size also decreases the optimal food particle size (Hansen et al., 1994) increasing
242 the share of algae in the diet of cyclopoid copepods (Adrian and Frost, 1993), which explains positive
243 effects of temperature on $\delta^{13}\text{C}$ values of the animals. In present study, water temperature was in fact
244 retained as the predictor in the final model for $\delta^{13}\text{C}$ in *Daphnia* sp. however not as a significant one.
245 This corroborates with the findings of (Schilder et al., 2015) who concluded that temperature is
246 negligible predictor of *Daphnia* $\delta^{13}\text{C}$ values.

247 Availability of nutrients (i.e. NH_4^+ , NO_3^- and PO_4^{3-} concentrations) in water also showed significant
248 effects on crustacean $\delta^{13}\text{C}$ values. In final models NH_4^+ and NO_3^- concentrations were step-wise
249 selected as significant predictors with negative effect on $\delta^{13}\text{C}$ values of *Daphnia* and calanoid copepods.
250 NH_4^+ and NO_3^- are crucial for phytoplankton development and in shallow lakes often become a limiting
251 factor, especially under availability of phosphorous (as in studied lakes; Søndergaard et al., 2017). Some
252 algal groups, such as green algae, cyanobacteria and flagellates show preferences toward NH_4^+ while
253 diatoms are more likely to utilize NO_3^- (Domingues et al., 2011; Glibert et al., 2016). Furthermore,
254 $\text{NO}_3^-:\text{NH}_4^+$ ratio is recognized to impact community composition by influencing growth rates of certain
255 taxa (Trommer et al., 2020). Increased growth rates and factors controlling them (such as nutrient
256 availability) are demonstrated to negatively affect fractionation of carbon stable isotopes, and thus
257 leading to more positive $\delta^{13}\text{C}$ values in phytoplankton (Burkhardt et al., 1999; Magozzi et al., 2017).
258 Assuming this mechanism to be in force in waterbodies considered in present study, observed negative
259 effect of NH_4^+ and NO_3^- on $\delta^{13}\text{C}$ suggest higher concentrations of these nutrients may be associated
260 with slower phytoplankton growth rates, resulting in lower $\delta^{13}\text{C}$ values in crustacean consumers. This
261 pattern is opposite to what would be expected if nutrient availability directly stimulated phytoplankton
262 growth, but seasonal dynamics and declining water temperatures effects might be involved. In shallow
263 lakes, increased nitrogen availability is typically observed in early spring and late autumn, when low
264 temperatures constrain primary production despite nutrient abundance (Søndergaard et al., 2017).
265 Furthermore, this result might also involve nutrient recycling by crustacean grazers, which depending
266 on N and P requirements by certain taxa, can differently affect entire food web (Branco et al., 2018).
267 Cladocerans of genus *Daphnia* are demonstrated to excrete substantial amounts of NH_4^+ while grazing,
268 increasing availability of NH_4^+ for phytoplankton (Karjalainen et al., 1997; Wojtal-Frankiewicz and
269 Frankiewicz, 2011). This phenomenon is an additional explanation for effects of NH_4^+ on crustacean
270 $\delta^{13}\text{C}$ values throughout the season.

271 In addition to negative effects of NH_4^+ and NO_3^- we found positive significant effects of PO_4^{3-} on
272 *Daphnia's* $\delta^{13}\text{C}$ values. For optimal performance, *Daphnia* requires high amounts of phosphorus (P) in



273 relation to nitrogen (N) and carbon (C; Hessen, 1992; Hessen et al., 2013), which is a reason why in
274 some systems *Daphnia* can act as a phosphorus trap (Branco et al., 2018; Karjalainen et al., 1997). The
275 reason can be the fact highlighted by Vrede et al. (1999), who showed that approximately 14% of P in
276 *Daphnia*'s body is allocated within the carapace, that is lost during the molting (which can happen every
277 2 – 3 days at 20°C; Lynch et al., 1986). This mechanism strongly drains *Daphnia* individuals from P
278 content (Vrede et al., 1999). However, thanks to the ability of P uptake from dissolved phase (Boersma,
279 2000; Urabe et al., 1997), *Daphnia* is not only reliant on algal P but also on the P compounds dissolved
280 in the water. Phosphorus is known to have regulatory effects on energy transfer within food webs
281 (Hessen, 2008; Welts et al., 2017), however in our study PO_4^{3-} concentration was significantly correlated
282 only with $\delta^{13}\text{C}$ in *Daphnia*. PO_4^{3-} concentration was also retained in the final model for calanoid
283 copepods, however its effect was marginal. Copepods are recognized to store lower amounts of
284 phosphorus in dry weight (P:DW) than *Daphnia* (Vrede et al., 1999) and nitrogen : phosphorus ratio
285 (N:P) is higher in copepods than in *Daphnia* (Sommer and Sommer, 2006). Growth rates of copepods
286 are slower than *Daphnia*'s therefore their P requirement is lower (Lorenz et al., 2019), which explains
287 no effects of PO_4^{3-} concentration on $\delta^{13}\text{C}$ values of both copepod groups.

288 Beyond phytoplankton, which is substantial part of planktonic crustaceans' diet, also bacterial
289 communities play an important role in cycle of carbon in shallow freshwater ecosystems. Bacterial
290 communities are recognized to acquire majority of carbon from phytoplankton (Lammers et al., 2017),
291 and can serve as a food source for crustacean zooplankton, directly (filter-feeders) or through trophic
292 cascades (via protozoans; predatory animals) (Kalinowska et al., 2015; Vrede and Vrede, 2005; Work
293 and Havens, 2003). In our studies, we found presence of several groups of bacteria: Proteobacteria, α -
294 proteobacteria, β -proteobacteria, γ -proteobacteria, δ -proteobacteria and "Other bacteria". The
295 models established for calanoid copepods and *Daphnia* indicated positive effects of "Other bacteria"
296 density on $\delta^{13}\text{C}$ values. This indicates that "Other bacteria" constitute a pathway (direct or indirect) of
297 incorporation of photosynthetically fixed carbon to diet of these animals, since photosynthesis tends
298 to drive $\delta^{13}\text{C}$ to less negative values in autotrophs and therefore in all consumers relying on them (Finlay
299 and Kendall, 2007). The final model showed that $\delta^{13}\text{C}$ values in *Daphnia* underwent significant negative
300 effect of β -proteobacteria density. β -proteobacteria are able to quickly utilize organic matter and thus
301 significantly contribute element cycling in the freshwater ecosystems and are often related to nitrogen
302 cycle (Bouvier and Del Giorgio, 2002; Hayden and Beman, 2014; Schulhof et al., 2020) which explains
303 similar negative effects of NO_3^- , NH_4^+ and β -proteobacteria on crustaceans' $\delta^{13}\text{C}$. However, β -
304 proteobacteria are also recognized as methanotrophs (Amin et al., 2017; van Grinsven et al., 2020;
305 Kuloyo et al., 2020; Redmond et al., 2010; Somers et al., 2020) and multiple β -proteobacteria members
306 have the genomic capacity to the biological cleavage of methylphosphonate, which simultaneously



307 releases phosphate and methane (Peoples et al., 2023). Methane-oxidizing bacteria can be grazed by
308 zooplankton and therefore constitute a conduit for methane incorporation into the food web (Agasild
309 et al., 2014; Jones and Grey, 2011). Sanseverino et al. (2012) demonstrated that $\delta^{13}\text{C}$ of grazers foraging
310 on methane oxidizing bacteria can be driven to low values (-42.0‰), which is comparable to the lowest
311 $\delta^{13}\text{C}$ values found in *Daphnia* spp. (-42.5‰) in our study. Methane oxidation is also identified in α - and
312 γ -proteobacteria (Borrel et al., 2011; Op den Camp et al., 2009), but our analyses showed no
313 relationship between densities of these groups and $\delta^{13}\text{C}$ in planktonic crustaceans considered in
314 present study.

315 Overall, our study highlights the complex interplay of environmental conditions, microbial processes
316 and functional trait-driven response of $\delta^{13}\text{C}$ values in planktonic crustaceans. A clear limitation of the
317 study is the lack of $\delta^{13}\text{C}$ measurements in crustacean dietary sources (e.g. algae, bacteria), which
318 appear as a future research direction. Insight into the diets of planktonic crustaceans and linking $\delta^{13}\text{C}$
319 values found in our samples to specific food sources would require subsequent measurements of $\delta^{13}\text{C}$
320 in lower trophic levels and application stable isotope mixing models. This approach was not considered
321 here and is a clear limitation of this study but will be examined in the future research. Nevertheless,
322 this study involves integrative perspective improving the understanding of carbon flow in shallow
323 freshwater ecosystems.

324

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332

333 **Competing interests**

334 The authors declare that they have no conflict of interest.

335

336 **Author contributions**



337 WK and EWW conceived the ideas and designed the study; WK and EW collected the data; WK and IG
338 processed biological samples; WK analyzed the data; WK and EWW led the writing of the manuscript;
339 EW and IG reviewed and edited the manuscript. All authors contributed critically to the drafts and gave
340 final approval for publication.

341

342 **Data availability**

343 The dataset used within this study will be made available on a reasonable request

344

345

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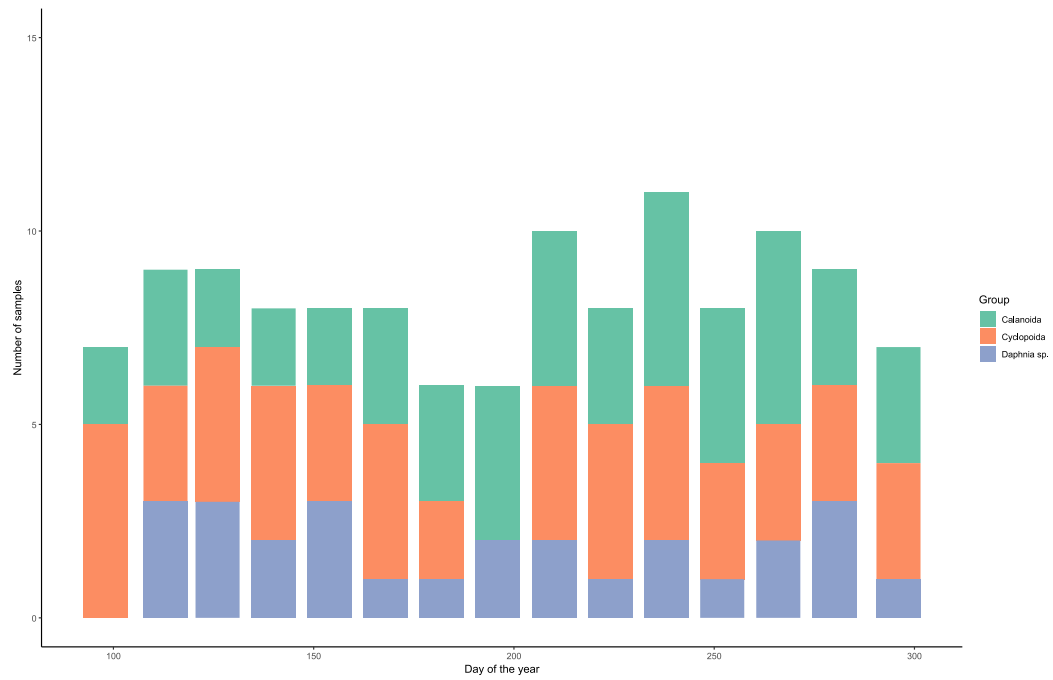
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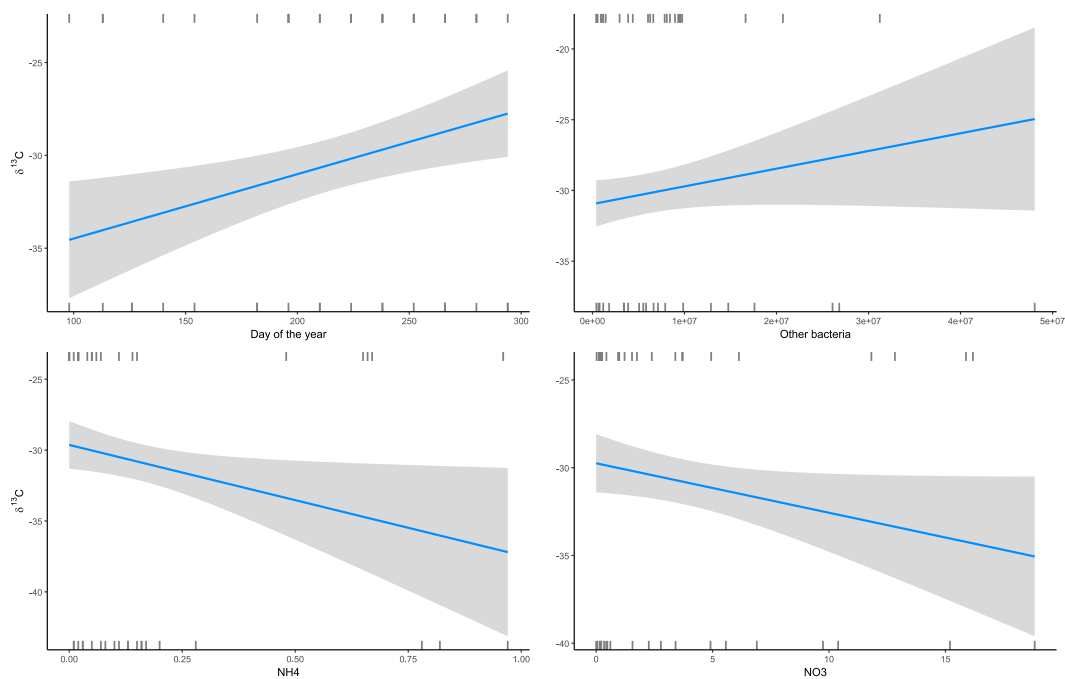
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561 Fig. 1. Distribution of studied planktonic crustacean groups throughout the sampling period

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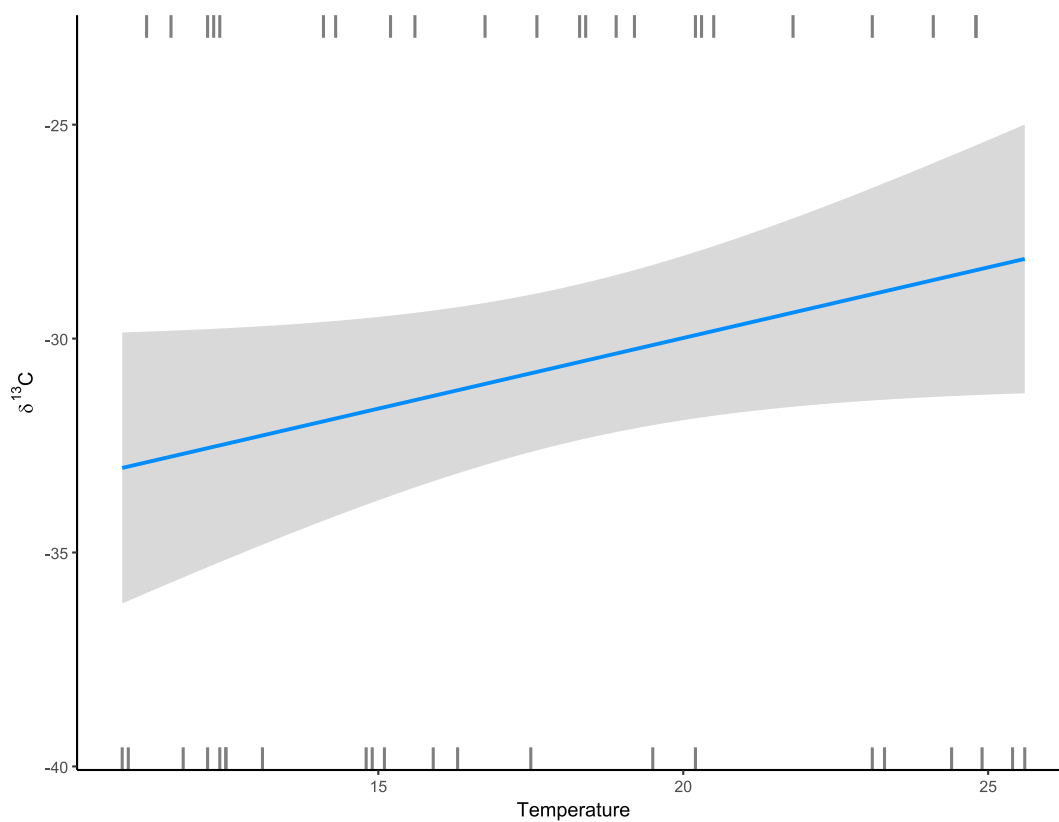
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565 Fig. 2. Visualization of stepwise-selected significant predictors of Calanoid copepods' $\delta^{13}\text{C}$

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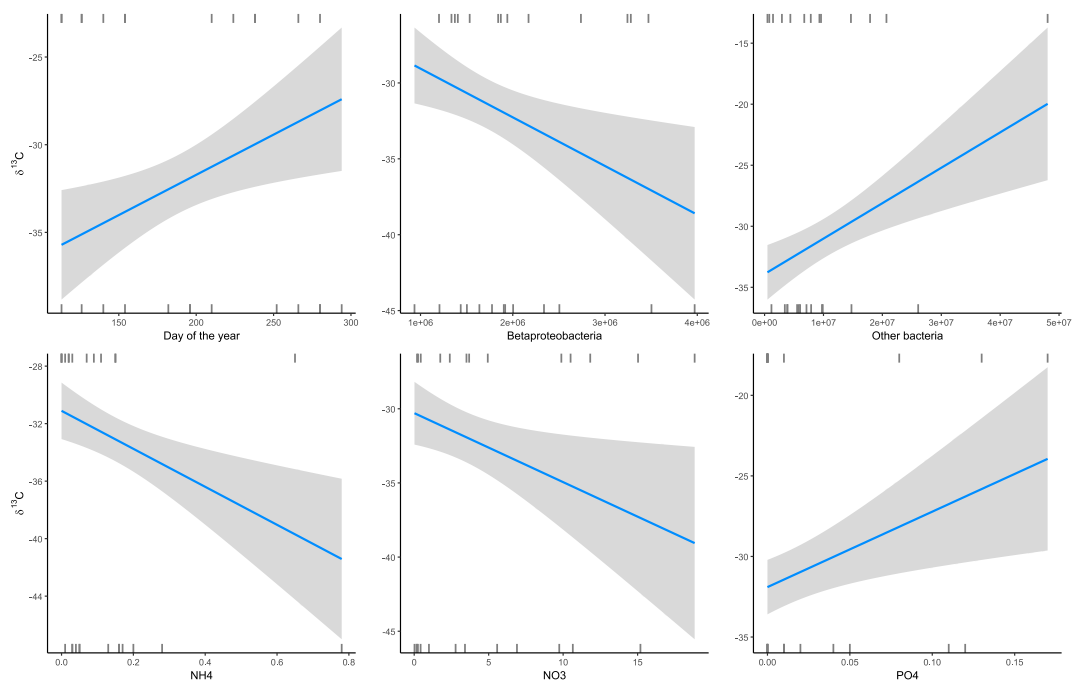
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569 Fig. 3. Visualization of stepwise-selected significant predictors of Cyclopoid copepods' $\delta^{13}\text{C}$

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573 Fig. 4. Visualization of stepwise-selected significant predictors of *Daphnia* $\delta^{13}\text{C}$

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