



1 **Mineralization Mechanism and Application Potential of Vaterite**
2 **Induced by Two Bacterial Strains**

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15



16 **Abstract**

17 Clarifying the bacterial mechanisms underlying vaterite biosynthesis is important for
18 advancing biomineralization theory and expanding its applications in material science,
19 environmental protection, and biomedicine. In this study, *Bacillus velezensis* LB002
20 and *Pseudomonas putida* KT2440 were used as representative bacterial strains to
21 compare strain-dependent vaterite formation and environmental remediation
22 performance. The biomineralization processes in the liquid medium were dynamically
23 monitored, and the induced biominerals were characterized, with particular attention
24 to their organic components. The differences in vaterite characteristics between both
25 bacterial systems were evaluated, and their remediation potential was assessed using
26 Cu^{2+} and tetracycline removal as model applications. Both strains induced vaterite
27 formation under identical culture conditions, while *B. velezensis* exhibited a stronger
28 biomineralization capacity than *P. putida*. In a 100 mL culture system, the Ca^{2+}
29 mineralization percentage of *B. velezensis* reached 35.06%, and the dry weights of
30 biomineral aggregates produced by *B. velezensis* and *P. putida* were 0.36 and 0.31 g,
31 respectively. SEM observations revealed that the vaterite induced by *B. velezensis* had
32 a loose and porous surface, whereas that induced by *P. putida* had a relatively smooth
33 surface. For the high concentration of Cu^{2+} at 50 mg/L, the removal rate reached
34 84.93% for *B. velezensis*-induced vaterite, compared with 71.86% for *P.*
35 *putida*-induced vaterite under the same conditions. Both biominerals demonstrated
36 similar tetracycline removal performance, with removal rates of approximately 60%.
37 These findings provide comparative evidence for strain-dependent vaterite formation
38 and support the potential application of biogenic vaterite in environmental
39 remediation.

40 **Keywords:** Bacteria; Biomineralization; Vaterite; Mineralization Mechanism;
41 Cu^{2+} -TC Combined Pollution

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45 **1. Introduction**

46 Microbially induced calcium carbonate precipitation (MICP) is a widespread
47 biosedimentary process in both terrestrial and marine environments, where it
48 contributes substantially to elemental biogeochemical cycling and the formation of
49 marine sedimentary rocks (Dupraz et al., 2009; Diaz, et al., 2019; Wang et al., 2023).
50 Under suitable environmental conditions, most naturally occurring microorganisms
51 can mediate calcium carbonate synthesis (Hou et al., 2011; Li et al., 2023a; Xiao, et
52 al., 2016; Whitaker et al., 2020). Among bacterially induced calcium carbonate
53 biominerals, calcite, aragonite–calcite assemblages, and vaterite–calcite assemblages
54 are the most frequently reported forms (Lyu et al., 2021; Dhami et al., 2013). The
55 diversity of crystal polymorphs and morphologies produced during bacterial
56 mineralization is controlled by several biological and geochemical processes.
57 Bacterial cell walls may provide mineralization templates that promote calcium
58 carbonate nucleation and crystal growth (Castanier et al., 2000; Li et al., 2023a).
59 Carbonic anhydrase (CA) secreted by bacteria accelerates the hydration of CO₂ to
60 HCO₃⁻, which can subsequently be converted into CO₃²⁻ under weakly alkaline
61 conditions and incorporated into biomineralization (Han et al., 2013; Zhang et al.,
62 2023; Lyu et al., 2021). Extracellular polymeric substances (EPS) can regulate
63 mineral formation because their negatively charged functional groups, including
64 carboxyl, hydroxyl, and phosphate groups, can chelate Ca²⁺ and provide nucleation
65 sites for biogenic calcium carbonate formation (Zhuang et al., 2018; Zhang et al.,
66 2023). The type and abundance of organic matter generated by different bacteria
67 during cultivation influence both the polymorph composition and structural stability
68 of the resulting calcium carbonate biominerals (Zhu et al., 2021; Li et al., 2023a). For
69 instance, calcite and aragonite have relatively stable crystal structures and can be
70 formed in the presence of comparatively simple organic components (Yu et al., 2018).
71 In contrast, vaterite is thermodynamically less stable and more difficult to induce
72 under conventional conditions. Its initial nucleation commonly depends on amino
73 acids or small-molecule proteins (Zhang et al., 2020; Gao et al., 2023). In most
74 calcium carbonate mineralization systems, vaterite behaves as a metastable phase and



75 readily transforms into calcite or aragonite (Zhu et al., 2021; Li et al., 2019). However,
76 bacterially induced vaterite has been shown to possess comparatively high structural
77 stability. Under specific high-temperature and high-pressure conditions, it can retain
78 its original morphology and crystal structure for extended periods without
79 transforming into calcite or aragonite (Liu et al., 2021; Liu et al., 2023). In addition to
80 its structural stability, *Bacillus velezensis*-induced vaterite has demonstrated strong
81 adsorption capacities for high concentrations of heavy metals, including Cd²⁺, Ni²⁺,
82 and Cu²⁺, with maximum adsorption capacities (Q_{max}) of 172.41, 270.27, and 178.57
83 mg/g, respectively. These capacities are higher than those reported for conventional
84 adsorbents such as biochar and graphene oxide (Park et al., 2016; Tan et al., 2015; Liu
85 et al., 2019a, 2019b). Biogenic vaterite (BV) has been reported to adsorb antibiotic
86 contaminants such as doxorubicin hydrochloride, with a Q_{max} of 447.58 mg/g (Zhang
87 et al., 2022), suggesting that it may be applicable for the removal of typical organic
88 pollutants. Its large specific surface area, loose porous surface structure, and favorable
89 structural stability further support its potential use in drug loading (Li et al., 2023b)
90 and environmental remediation (Liu et al., 2019a, 2019b; Jin et al., 2024). However,
91 further development of BV for environmental engineering requires clarification of
92 several issues related to biomineralization synthesis, controlling factors, and practical
93 applicability. In particular, the key factors governing bacteria-induced BV formation,
94 mechanisms regulating BV morphology and particle size, and feasibility of BV
95 production by different bacterial sources as environmental remediation agents remain
96 insufficiently defined.

97 Copper pollution in agricultural production systems can be derived from multiple
98 sources. Chemical fertilizers and pesticides can introduce copper into soils, whereas
99 certain feed additives contribute to copper accumulation in livestock breeding farms.
100 Copper-containing compound fertilizers and organic fertilizers have particularly
101 pronounced effects on soil copper content (Xiao et al., 2022). Long-term and
102 excessive application of organic fertilizers, such as livestock manure, not only
103 intensifies copper accumulation in soils but also increases antibiotic concentrations,
104 thereby threatening soil ecological quality and human health (Wang et al., 2013; Xu,



105 et al., 2023 Wang et al., 2021). To evaluate the potential application of BV in
106 agricultural production, *B. velezensis* LB002 (Gram-positive, G⁺) and *P. putida*
107 KT2440 (Gram-negative, G⁻) were selected as experimental strains. Both strains are
108 typical environmental functional bacteria widely distributed in natural soils, with
109 similar ecological niches and strong environmental adaptability. Meanwhile, they
110 exhibit distinct inherent differences in cell wall structure, as well as the composition
111 and secretion characteristics of extracellular polymeric substances (EPS). This
112 controllable comparative system enables an in-depth investigation into the effects of
113 bacterial Gram characteristics and extracellular properties on the biosynthesis,
114 micromorphology and mineralization efficiency of vaterite. Their capacity to induce
115 BV synthesis and the temporal characteristics of mineralization were examined under
116 identical culture conditions. The dry weight, micromorphology, and organic
117 composition of BV produced by the two bacterial strains were compared. The roles of
118 fermentation broth composition, physicochemical properties, and CA activity in
119 bacterial biomineralization were analyzed. Finally, the removal efficiencies of BV
120 induced by the two strains for organic and inorganic pollutants were assessed. The
121 resulting evidence provides a basis for understanding strain-dependent BV formation
122 and evaluating the environmental application potential of biogenic vaterite.

123 **2. Materials and Methods**

124 **2.1 Experimental Strains and Culture**

125 *B. velezensis* LB002 (NCBI accession no. CP037417.1), isolated from soil, and *P.*
126 *putida* KT2440 (ATCC 33015), obtained as a model strain, were used in this study.
127 Both strains were preserved at the Key Laboratory of Microbiology and Functional
128 Genes, Nanjing Normal University. Luria–Bertani (LB) medium was used for
129 bacterial cultivation and mineralization induction. The medium was prepared with 1%
130 (m/v) tryptone, 0.5% (m/v) yeast extract, and 1% (m/v) NaCl, and the initial pH was
131 adjusted to 7.0–7.3. Bacterial cultures were incubated at 30°C and 180 rpm in a
132 shaking incubator. Each treatment was performed with three biological replicates.

133 **2.2 Analysis of Carbonate Mineralization Characteristics Induced by Two** 134 **Bacterial Strains**



135 Before mineralization experiments, both bacterial strains were activated in LB
136 medium until the OD₆₀₀ of the seed cultures reached approximately 1.2. The viable
137 cell concentrations determined by plate counting were approximately 5×10^8
138 CFU/mL for *B. velezensis* and 4×10^8 CFU/mL for *P. putida*. For each mineralization
139 treatment, 2 mL of the activated bacterial suspension was inoculated into 100 mL of
140 LB medium supplemented with 0.8 g CaCl₂. This calcium concentration was selected
141 based on the previous vaterite induction results obtained by our research group (Liu et
142 al., 2021). An abiotic control was established using LB medium containing 0.8 g
143 CaCl₂ without any bacterial inoculation. All cultures and controls were incubated at
144 30°C and 180 rpm for 14 d. Samples were collected after 1, 3, 5, 7, 10, and 14 d of
145 incubation. The pH of each fermentation broth was measured using a pH meter
146 (PHSJ-4F, Leici, China). The collected samples were centrifuged at 8000 rpm, and the
147 supernatants were retained for Ca²⁺ analysis. The Ca²⁺ concentrations in the
148 experimental and control groups were determined using atomic absorption
149 spectroscopy (AA-6300C, Shimadzu, Japan). The Ca²⁺ mineralization percentage was
150 calculated using the following equation (Li et al., 2023a).

151 Percent of Ca²⁺ mineralization (%) =
$$\frac{C_{Ca^{2+} \text{ in control group}} - C_{Ca^{2+} \text{ in experimental group}}}{C_{Ca^{2+} \text{ in control group}}} \times 100$$

152

153 The mixture was centrifuged at 8000 rpm for 5 min to separate the supernatant
154 and precipitate. The precipitate was washed three times with double-distilled water,
155 then dried in an oven at 55° C for 12 h until constant weight was achieved. The dry
156 mass of each precipitate was measured using an analytical balance. The mineral
157 phases of the dried products were identified using X-ray diffraction (XRD; BTX-526,
158 Olympus, USA) with a Co target over a scanning range of 5–55°.

158

159 **2.3 Comparative analysis of carbonate mineralization ability and pollutant** 160 **immobilization performance induced by two bacterial strains**

160

2.3.1 Comparison of organic composition and content of two biogenic vaterite



161

162 Preliminary tests using samples collected at different incubation stages confirmed that
163 the calcium carbonate phase stabilized on day 7. At this stage, only slight changes in
164 dry weight were observed, and the Ca²⁺ mineralization percentage reached an
165 approximately steady state. Accordingly, the day-7 mineral products were selected for
166 subsequent comparative analysis. After 7 d of incubation, the precipitates were
167 collected, washed, and dried to a constant weight. The organic elemental
168 compositions of the two biomineral aggregates were determined using an elemental
169 analyzer (vario EL cube, Elementar, Germany). Fourier transform infrared
170 spectroscopy (FTIR; Nicolet 8700, Thermo Nicolet, USA) and
171 thermogravimetry–derivative thermogravimetry (TG-DTG; STA7300, Perkin-Elmer,
172 USA) were used to identify the organic matter composition and quantify its content in
173 the biominerals (Liu et al., 2019; Li et al., 2023a).

173

174 **2.3.2 Comparison of Morphology of Two Biogenic Vaterite Samples**

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175 Following centrifugation and lyophilization, the biomineral powders were
176 characterized using scanning electron microscopy coupled with energy-dispersive
177 spectroscopy (SEM–EDS; Supra 55, Zeiss, Germany). Transmission electron
178 microscopy (TEM; HT7700, Hitachi, Japan) was used to examine the microstructure
179 and crystal morphology of the induced biominerals. For TEM preparation, 1.5 mL of
180 bacterial fermentation broth was centrifuged, and the supernatant was discarded. The
181 collected bacterial cells were fixed in 2.5% glutaraldehyde fixative for electron
182 microscopy, embedded in epoxy resin, and sectioned into ultrathin slices using a
microtome (EMUC7, Leica, Germany). The prepared sections were subsequently used



183

for TEM observations and analyses.

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2.4 Analysis of Biochemical Characteristics of Bacterial Culture Solution during

185

Bacteria-Induced Mineralization

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After 7 d of cultivation, the fermentation broths from the two bacterial strains were
187 centrifuged, and the collected supernatants were subjected to biochemical analysis.

188

Polysaccharide content was quantified using the phenol–sulfuric acid method, and
189 protein content was measured using a Bradford Protein Assay Kit (Shanghai Sangon

190

Biotech, China) (Li et al., 2023a). The ability of the fermentation supernatants to
191 induce calcium carbonate formation was examined using a biomimetic mineralization

192

system established according to previous methods (Lian et al., 2006; Liu et al., 2021).

193

Excess NH_4HCO_3 was placed at the bottom of the mineralization device, and 20 mL
194 of a mixed solution containing 0.2 mol/L CaCl_2 and NaHCO_3 was introduced into the
195 upper layer of each group. Parallel mineralization systems were supplemented with 5

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mL of fermentation supernatant from each bacterial strain or 5 mL of double-distilled
197 water. The sealed devices were maintained at room temperature for 14 d. The minerals

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formed during incubation were collected and analyzed using XRD for crystal phase
199 identification (Lian et al., 2006; Liu et al., 2021). The carbonic anhydrase (CA)

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activity in the bacterial supernatants was also determined to further compare the
201 mineralization-inducing capacity of the two strains (Li et al., 2013). After

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centrifugation, the biomineral precipitates were dried to a constant weight, and their
203 carbonate content was determined by acid–base titration (Li et al., 2023a).

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2.5 Comparison on Removal Efficiency of Simulated Single and Combined



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Pollutants by Biogenic Vaterite

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207 Biominerals harvested after 7 d of incubation were used to evaluate the removal

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209 performance of biogenic vaterite for simulated inorganic and organic pollutants,

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211 represented by Cu^{2+} and tetracycline (TC), respectively. Single-pollutant solutions

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213 were prepared at Cu^{2+} concentrations of 10, 20, 30, 40, and 50 mg/L and TC

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215 concentrations of 5, 10, 15, 20, and 25 mg/L. Combined pollutant solutions containing

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217 both Cu^{2+} and TC were prepared using the same concentration gradients. For each

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219 treatment, 20 mL of the pollutant solution was placed in a 50 mL centrifuge tube, and

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221 0.05 g of biogenic vaterite induced by either bacterial strain was added. The mixtures

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223 were shaken at 25°C and 120 rpm for 12 h, followed by centrifugation at 8000 rpm to

224

225 separate the supernatant and precipitate. The residual Cu^{2+} concentration in the

226

supernatant was measured using atomic absorption spectroscopy (AAS), and the

residual TC content was calculated from the absorbance at 415 nm using a

spectrophotometer (Rathod, et al., 2015). The adsorption-related textural properties of

vaterite induced by the two bacteria were preliminarily examined through N_2

adsorption–desorption isotherm analysis using an automatic gas adsorption analyzer

after degassing at 80°C for 12 h. The specific surface area was calculated using the

BET model, the core theoretical model for specific surface area determination,

whereas the pore size distribution was obtained using the BJH model, a classical

model for 2–50 nm mesoporous pore size analysis (Wang et al., 2014; Sheng et al.,

2009).

3. Results



227

3.1 Synthesis Process and Capacity Analysis of Vaterite Induced by Two Bacterial Strains

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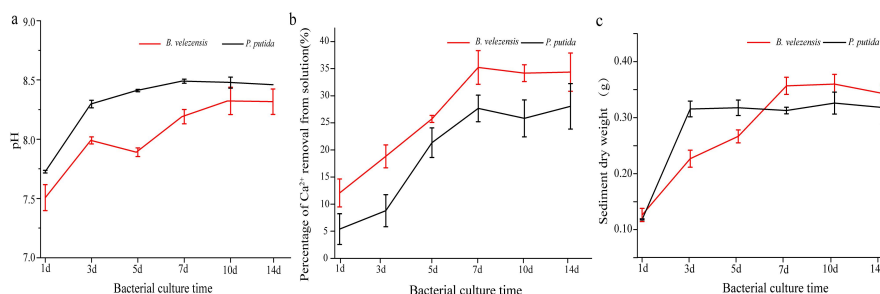
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The fermentation media of the two bacterial strains were dynamically monitored to characterize vaterite formation. The pH of both fermentation broths increased gradually from day 1 to day 7 and subsequently remained at a weakly alkaline level. The pH reached approximately 8.10 for *B. velezensis* and approximately 8.30 for *P. putida* (Fig. 1a). Consistent with this pH evolution, the Ca²⁺ mineralization percentages in both bacterial systems approached equilibrium by day 7 (Fig. 1b). The measured Ca²⁺ balance indicated that only a small fraction of the consumed Ca²⁺ was directly assimilated by bacterial cells, whereas most of the consumed Ca²⁺ was immobilized in newly formed biominerals. The dry weight of the precipitates increased continuously during the first 7 d and then remained nearly constant (Fig. 1c). Based on the results shown in Fig. 1b, the Ca²⁺ mineralization percentage reached approximately 35.06% for *B. velezensis* on day 7, compared with 27.50% for *P. putida*. The corresponding dry weights of the biomineral aggregates were 0.36 and 0.31 g, respectively (Fig. 1c). Therefore, the temporal variation in precipitate dry weight was consistent with the change in the Ca²⁺ mineralization percentage (Fig. 1b and c).

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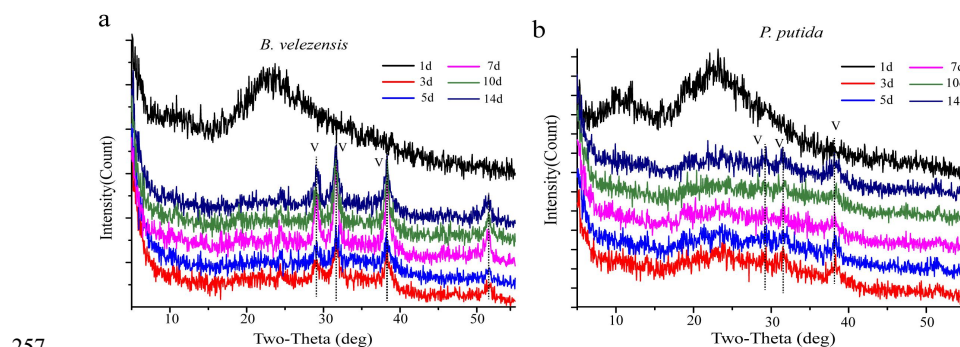


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246 Fig. 1 Dynamic changes in key mineralization parameters induced by the two
247 bacterial strains. (a) Variation in pH during the bacterial-induced mineralization
248 process; (b) comparison of Ca^{2+} mineralization rates induced by the two bacterial
249 strains; (c) comparison of the dry weights of biominerals formed by the two bacterial
250 strains.

250

251 XRD analysis was conducted to identify the mineral phases of the biominerals
252 induced by the two bacterial strains. No biomineral products were detected in either
253 bacterial system on day 1. From day 3 onward, both *B. velezensis* and *P. putida*
254 induced vaterite formation, although the vaterite diffraction peaks in the *P. putida*
255 were relatively weak. With increasing incubation time, the crystal phase of the
256 induced carbonate minerals remained unchanged, and vaterite was maintained as the
dominant phase throughout the cultivation period (Fig. 2a and b).



257

258 Fig. 2 XRD patterns of mineral products induced by different bacterial strains during
259 1–14 d of incubation.

260 (a) Identification of mineral products induced by *B. velezensis*; (b) identification of
261 mineral products induced by *P. putida*.

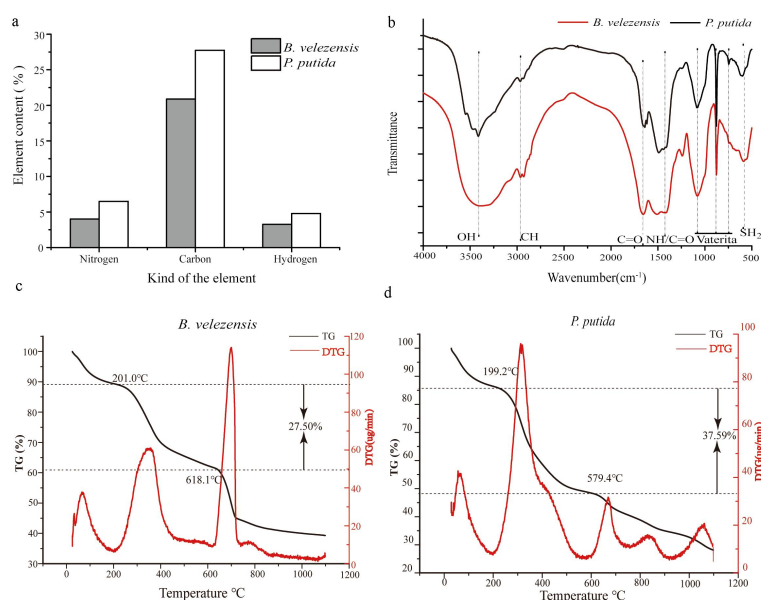
262 3.2 Organic Matter Content and Mineral Characterization of Vaterite Induced 263 by Two Bacterial Strains



264 Biologically induced vaterite usually contains variable amounts of organic matter (Liu
265 et al., 2023; Li et al., 2023a; Lyu et al., 2017). In this study, the N, C, and H contents
266 of the biomineral samples were used to evaluate the organic fraction within the
267 mineral aggregates (Fig. 3a). The higher N and C contents suggest a higher proportion
268 of unmineralized bacterial cells or organic constituents retained in the
269 organic–inorganic mineral composites. This composition was accompanied by a lower
270 relative calcium carbonate fraction and reflected weaker Ca^{2+} mineralization capacity.
271 Elemental analysis of the day-7 biomineral aggregates indicated that the *B.*
272 *velezensis*-induced vaterite contained 4.02% N, 20.89% C, and 3.27% H, whereas the
273 *P. putida*-induced vaterite contained 6.49% N, 27.74% C, and 4.78% H. The higher
274 elemental content measured in the *P. putida*-derived product indicated greater organic
275 matter retention in this biomineral (Fig. 3a). In addition, FTIR characterization
276 supported the incorporation of organic components into both biogenic vaterites.
277 Characteristic bands assigned to -OH, -CH, and -SH were observed at 3412, 2968,
278 and 579 cm^{-1} , respectively (Fig. 3b), confirming the presence of abundant
279 polysaccharides and proteins. Meanwhile, typical vaterite bands were detected at 1083,
280 876, and 747 cm^{-1} (Liu et al., 2019). The organic matter content of the biomineral
281 aggregates collected after 7 d of incubation was quantified using TG-DTG analysis.
282 Because large numbers of bacterial cells were present in the precipitates, the detected
283 organic fraction included both organic substances incorporated into the mineral matrix
284 and residual free bacterial cells that had not been fully mineralized. The
285 corresponding results are shown in Fig. 3c and d. The thermal mass-loss profiles of
286 the two biominerals were divided into three stages. The first stage, occurring near
287 200°C, corresponded to the evaporation of bound water. The second stage, extending
288 from 201 to 618°C, was mainly associated with the combustion of organic matter. The
289 third stage appeared at approximately 600°C and represented biomineral
290 decomposition, during which calcium carbonate began to decompose into CaO and
291 CO_2 . According to the mass losses recorded across these stages, the organic matter
292 contents of the biominerals induced by *B. velezensis* and *P. putida* were calculated as
293 27.50% and 37.59%, respectively. The higher organic matter content in the *P.*



294 *putida*-induced biomineral aggregates (Fig. 3a) suggests a weaker mineralization
 295 capacity and greater retention of residual organic components, including bacterial
 296 cells, in the precipitates. In contrast, *B. velezensis* exhibited a stronger mineralization
 297 ability and produced precipitates with a higher calcium carbonate proportion. This
 298 pattern was consistent with the elemental analysis results (Fig. 3a).



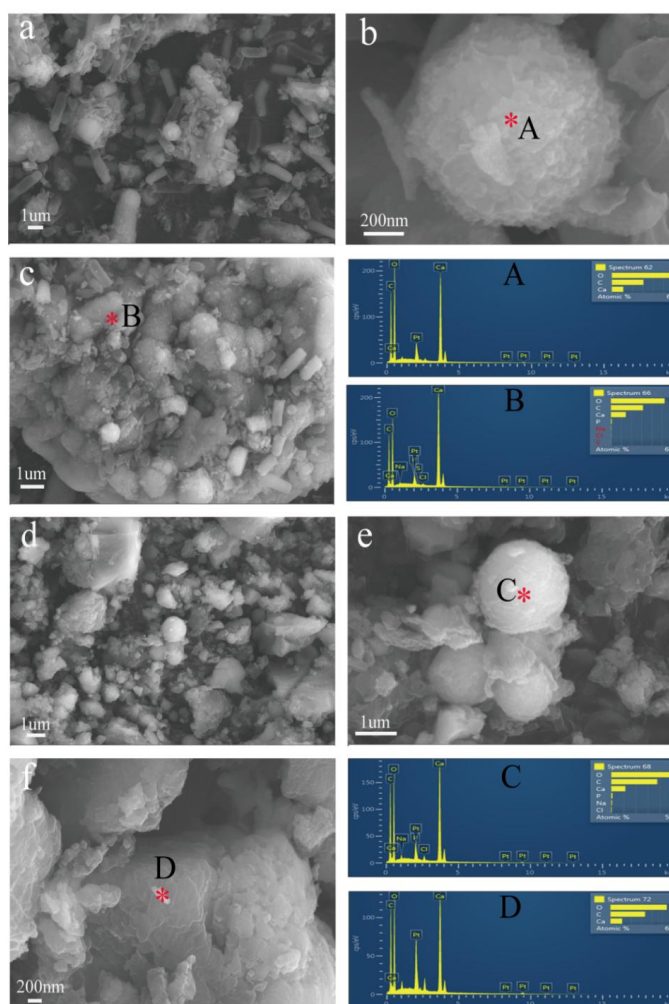
299
 300 Fig. 3 Characterization of the types and contents of organic matter in biogenic vaterite.
 301 (a) Major elemental compositions of the two biomineral samples; (b) FTIR spectra of
 302 the two biomineral samples; (c) thermogravimetric analysis of biominerals induced by
 303 *B. velezensis*; (d) thermogravimetric analysis of biominerals induced by *P. putida*.

304 3.3 Morphological Comparison of Vaterite Induced by Two Bacterial Strains

305 After vaterite formation was confirmed by XRD, SEM analysis was conducted to
 306 compare the morphology of the mineral products induced by the two bacterial strains.
 307 The *B. velezensis*-induced products were composed of micro- to nanoscale vaterite
 308 particles with smaller particle sizes and a loose surface texture (Fig. 4a and b). The
 309 participation of extracellular organic matter in the mineralization process was
 310 reflected by the incorporation of bacterial cells and spores into the mineral phase,
 311 resulting in compact bacteria–mineral complexes (Fig. 4a and c). In contrast, the
 312 products induced by *P. putida* were dominated by relatively larger particles at the



313 micron scale, and the corresponding biomineral complexes contained a higher
314 proportion of organic matter (Fig. 4d and e). EDS analysis indicated that the *P.*
315 *putida*-induced vaterite was characterized by higher carbon content and lower calcium
316 content than the *B. velezensis*-induced vaterite (Fig. 4f). This revealed relatively lower
317 degree of carbonate mineralization and confirmed the stronger
318 mineralization-inducing capacity of *B. velezensis*.



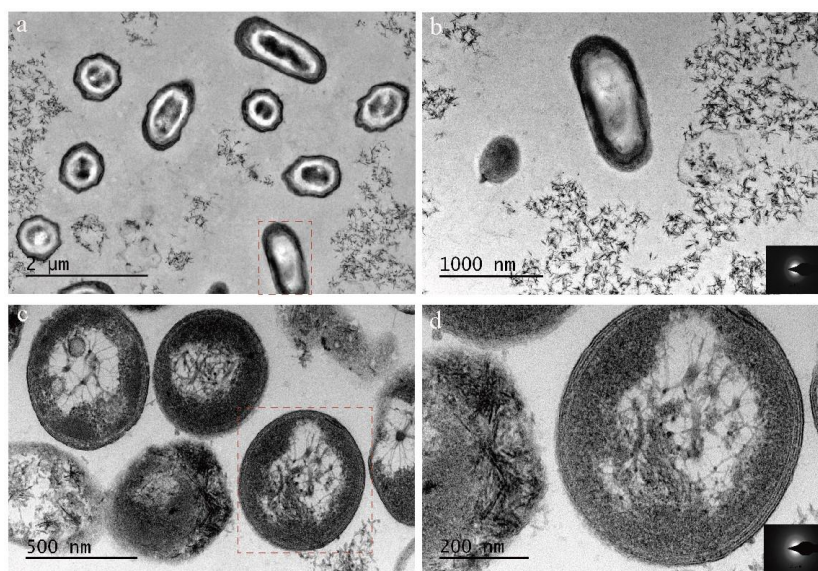
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320 Fig. 4 SEM-EDS characterization of biogenic vaterite induced by the two bacterial
321 strains. (a–c) Morphological characterization of biominerals induced by *B. velezensis*;
322 (A–B) corresponding EDS spectra of the selected regions shown in panels (a–c); (d–f)



323 morphological characterization of biominerals induced by *P. putida*; (C–D)
324 corresponding EDS spectra of the selected regions shown in panels (d–f).

325 Surface-associated mineralization of the two bacterial strains was examined
326 using TEM. *B. velezensis* retained a rod-shaped morphology, and numerous acicular
327 mineral particles were distributed around the bacterial cells (Fig. 5a and b). The *P.*
328 *putida* cells appeared spherical, with biomineral material attached to their surface (Fig.
329 5c and d). However, the selected area electron diffraction (SAED) analysis of the
330 mineralized surfaces of both bacterial strains did not reveal characteristic diffraction
331 patterns corresponding to vaterite.



332
333 Fig. 5 TEM and SAED analyses of bacterial surface mineralization. (a) TEM
334 observation of surface mineralization in *Bacillus velezensis*; (b) cellular morphology
335 of *Bacillus velezensis*; (c) cellular morphology of *Pseudomonas putida*; (d) selected
336 area electron diffraction (SAED) analysis of bacterial surface mineralization.

337 3.4 Analysis of Biochemical Characteristics of Culture Solution Related to 338 Bacterial Mineralization

339 The carbonic anhydrase (CA) activity differed markedly between the two bacterial
340 fermentation broths. The CA activity measured in the *B. velezensis* broth was
341 approximately five times higher than that in the *P. putida* broth (Fig. S1a). Consistent



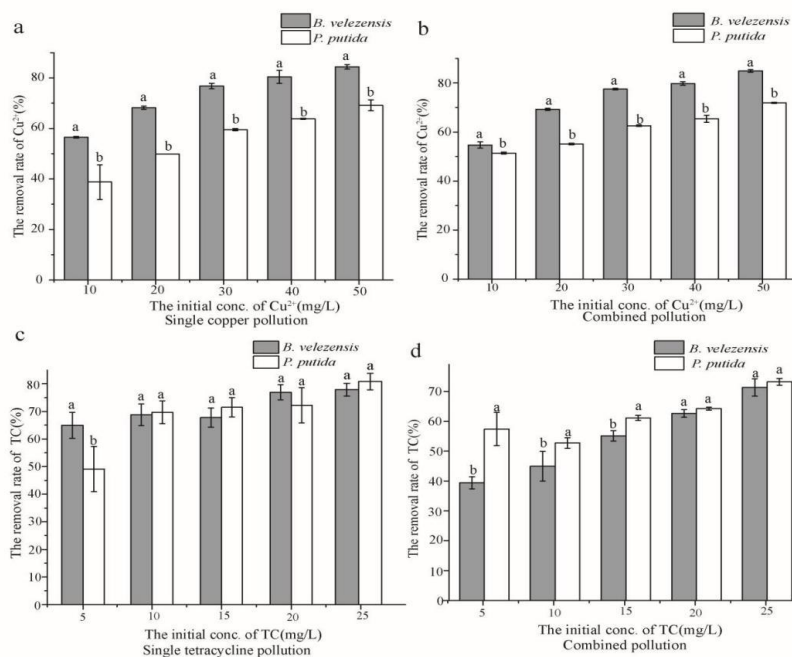
342 with this difference, titration analysis demonstrated that the carbonate proportions in
343 the biominerals induced by *B. velezensis* and *P. putida* were 38.93% and 24.26%,
344 respectively (Fig. S1b). The biochemical compositions of the 7-day fermentation
345 broths were analyzed. *B. velezensis* secreted a significantly higher polysaccharide
346 content than *P. putida*, with values of 0.53 and 0.33 g/L, respectively (Fig. S1c).
347 Conversely, the protein content in *P. putida* reached 0.024 g/L, which was higher than
348 that of 0.014 g/L in *B. velezensis* (Fig. S1d). Subsequently, biomimetic mineralization
349 experiments were performed using fermentation supernatants from both bacterial
350 strains, along with dextran and bovine serum albumin. The fermentation broths of
351 both *B. velezensis* and *P. putida* partially converted calcite to vaterite, whereas
352 ultrapure water did not induce this phase transformation (Fig. S2). These results
353 suggest that CA promotes the generation of carbonate ions, thereby facilitating their
354 subsequent binding with calcium ions to form calcium carbonate. Protein components
355 may contribute to vaterite biosynthesis and stabilization during mineralization.

356 **3.5 Comparison of Removal Efficiency of Organic-Inorganic Pollutants by** 357 **Vaterite Induced by Two Bacterial Strains**

358 The environmental application potential of the two biogenic vaterites was assessed by
359 comparing their removal performance for inorganic and organic pollutants. In both
360 single- and combined-pollutant systems, *B. velezensis*-induced vaterite showed a
361 stronger capacity for Cu^{2+} removal than *P. putida*-induced vaterite. At a high Cu^{2+}
362 concentration of 50 mg/L, the removal efficiencies of the *B. velezensis*-induced
363 vaterite reached 84.37% and 84.93%, respectively (Fig. 6a and b), whereas those of
364 the *P. putida*-induced vaterite were 69.16% and 71.86% under the same conditions.
365 The limited difference between the single and combined systems indicates that the
366 introduction of TC exerted only a minor effect on Cu^{2+} adsorption by vaterite. At
367 lower Cu^{2+} concentrations (10–20 mg/L), the *B. velezensis*-induced vaterite
368 maintained higher removal efficiencies than the *P. putida*-induced vaterite. This
369 stronger heavy metal removal performance was associated with its larger specific
370 surface area and pore structure, with a specific surface area of 13.19 m^2/g and an
371 average pore size of 4.88 nm, compared with 2.56 m^2/g and 4.13 nm for the *P.*



372 *putida*-induced vaterite, respectively (Table S1). In the single TC system, the *P.*
 373 *putida*-induced vaterite presented a higher TC removal rate at the low TC
 374 concentration of 5 mg/L, reaching 64.94% compared with 49.08% for the *B.*
 375 *velezensis*-induced vaterite. As the TC concentration increased, the removal
 376 efficiencies of both biominerals increased and gradually approached approximately
 377 70% (Fig. 6c). In the combined-pollutant system, Cu²⁺ addition reduced TC removal
 378 by both types of biogenic vaterite (Fig. 6d). This reduction was mainly attributed to
 379 the preferential occupation of adsorption sites by Cu²⁺, whereas TC, with its relatively
 380 larger molecular size, exhibited weaker competitive access to these sites. Overall, *B.*
 381 *velezensis*-induced vaterite exhibited more favorable removal performance,
 382 particularly for Cu²⁺ immobilization, supporting its potential application in pollutant
 383 remediation.



384
 385 Fig. 6 Comparison of pollutant removal performance by biogenic vaterite induced by
 386 the two bacterial strains. (a) Removal efficiency of Cu²⁺ under single-pollutant
 387 conditions; (b) removal efficiency of tetracycline (TC) under single-pollutant
 388 conditions; (c) removal efficiency of Cu²⁺ under combined-pollution conditions; (d)



389 removal efficiency of TC under combined-pollution conditions.

390 **4. Discussion**

391 **4.1 Key Controlling Factors for Bacterially Induced Vaterite Synthesis**

392 *B. velezensis* is a Gram-positive (G^+) bacterium with surface flagella and the ability to
393 form endospores. It commonly colonizes soil and plant surfaces and is widely used as
394 a probiotic strain in agricultural production. *P. putida* is a Gram-negative (G^-), ovoid,
395 lophotrichous bacterium that is broadly distributed in natural environments. Although
396 both strains differ markedly in cell type and physiological characteristics, they are
397 capable of inducing biogenic vaterite formation in Ca^{2+} -supplemented culture systems.
398 During mineralization, the pH increased from day 1 to day 3 of incubation (Fig. 1a).
399 This early alkaline shift was accompanied by the CA-mediated conversion of CO_2
400 into HCO_3^- , which subsequently participated in Ca^{2+} binding and calcium carbonate
401 formation under alkaline conditions. The most rapid increase in precipitate dry weight
402 also occurred during this period (Fig. 1c), followed by stabilization on day 7 (Fig. 1).
403 However, the mineralization efficiency differed substantially between these two
404 strains. The Ca^{2+} mineralization ratio reached 35.06% in the *B. velezensis* system,
405 whereas only 27.50% was achieved in the *P. putida* (Fig. 1b). This difference
406 indicates that *B. velezensis* provides a more favorable biological environment for
407 biogenic vaterite formation. The stronger mineralization performance of *B. velezensis*
408 was closely associated with differences in bacterial cell wall structure and
409 extracellular polymeric substances (EPS). The cell wall properties and EPS
410 composition can affect both the crystal type and micromorphology of biogenic
411 calcium carbonate. Among EPS components, proteins are particularly important
412 because they provide nucleation sites for vaterite formation (Zhang et al., 2021; Gao
413 et al., 2023). Although both strains induced vaterite under identical culture conditions,
414 their mineral products exhibited evident morphological differences. *B.*
415 *velezensis*-induced vaterite exhibited a loose, porous, and layered surface structure.
416 The polysaccharides secreted by *B. velezensis* appeared to promote the binding of
417 vaterite particles to bacterial cells, resulting in compact bacteria–mineral complexes
418 (Fig. 4c). In contrast, the *P. putida*-induced vaterite presented a smoother surface and



419 lacked a layered microstructure. The lower polysaccharide content in the *P. putida*
420 culture system weakened cell–mineral adhesion and restricted the formation of large
421 bacteria–mineral aggregates (Fig. 4d). These observations suggest that the
422 morphology of biogenic vaterite is largely regulated by the composition, type, and
423 abundance of EPS in the fermentation broth. In addition to EPS composition, bacterial
424 culture pH and CA activity were involved in regulating calcium carbonate synthesis.
425 During vaterite formation, proteins act mainly as nucleation sites for crystal initiation,
426 whereas polysaccharides mediate adhesion between bacterial cells and mineral
427 particles, thereby promoting the development of bacteria–mineral aggregates.
428 Biomimetic mineralization experiments showed that protein molecules were the
429 dominant contributors to biogenic vaterite formation (Fig. S2), while polysaccharides
430 primarily promoted mineral-particle adhesion and aggregate enlargement (Fig. 4b).
431 This pattern is consistent with the findings of Zhang et al. (2020) and supports the
432 involvement of EPS in fermentation broth-mediated biomineralization. Given the high
433 compositional complexity of bacterial fermentation broth, components other than
434 polysaccharides and proteins, such as organic acids, may participate in regulating the
435 formation of different calcium carbonate polymorphs. Therefore, further studies are
436 needed to clarify how the content and composition of polysaccharides and proteins
437 modulate calcium carbonate morphology and bacteria–mineral complex formation.

438 Taken together, vaterite synthesis induced by the two bacterial strains was
439 governed by the combined effects of bacterial cell wall properties, extracellular
440 metabolites, and mineralization conditions (Fig. 7). Both strains were able to induce
441 biogenic vaterite under suitable culture conditions, and the mineralization process was
442 closely related to pH variation and CA activity. Nevertheless, *B. velezensis* presented
443 a stronger mineralization tendency and higher mineralization capacity. Extracellular
444 metabolites, particularly proteins and polysaccharides, are key regulators of
445 extracellular bacterial mineralization. Proteins promoted vaterite nucleation and
446 formation, whereas polysaccharides strengthened biomineral particle adhesion and
447 supported the development of larger bacteria–mineral complexes. Consequently,
448 although both strains produced biogenic vaterite, the resulting minerals differed



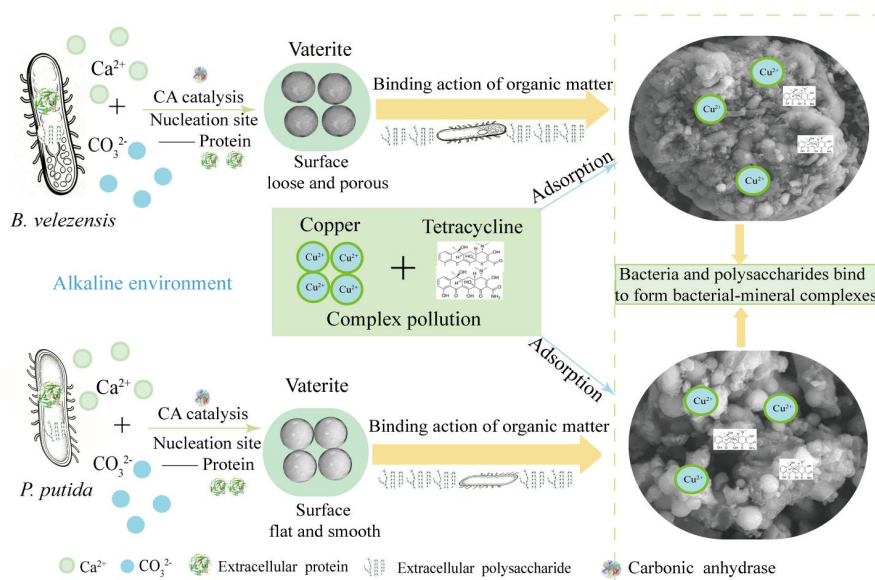
449 markedly in terms of surface microstructure and pollutant immobilization
450 performance. The higher Ca^{2+} mineralization rate and larger specific surface area of *B.*
451 *velezensis*-induced vaterite indicate that this biomineral has greater potential for
452 practical application.

453 **4.2 Synthesis Mode of Biogenic Vaterite**

454 Compared with chemically synthesized vaterite, bacterially induced biogenic
455 vaterite generally exhibits more complex morphology and higher structural stability,
456 making it less prone to transformation into aragonite or calcite (Liu et al., 2021). This
457 difference is mainly associated with the involvement of abundant organic matter,
458 especially protein components, which can function as nucleation sites during vaterite
459 formation. Both *B. velezensis* and *P. putida* secreted substantial EPS, including
460 polysaccharides and proteins. However, the higher CA activity of *B. velezensis*
461 promoted greater Ca^{2+} precipitation as calcium carbonate in the culture system. In
462 addition, the dense peptidoglycan layer in the *B. velezensis* cell wall could provide a
463 mineralization template, whereas the carboxyl and phosphate groups in teichoic acid
464 could bind environmental metal cations and immobilize more Ca^{2+} in the culture
465 system (Zhang et al., 2021; Li et al., 2023a). These findings suggest that
466 Gram-positive bacteria have a stronger cell-wall-mediated mineralization capacity and
467 are more likely to form bacteria–mineral complexes through the combined effects of
468 cell walls and extracellular polysaccharides (Fig. 7). The weaker mineralization
469 performance of *P. putida* was supported by organic matter and carbonate analyses.
470 TG-DTG results showed that *P. putida*-induced biogenic vaterite contained a higher
471 organic matter content, indicating that its bacteria–mineral aggregates retained more
472 unmineralized bacterial cells or organic components. This higher organic fraction
473 corresponded to a lower mineralization-inducing capacity in the *P. putida* system.
474 Quantitative carbonate titration showed that the carbonate content of *P.*
475 *putida*-induced vaterite was approximately 14.7% lower than that of *B.*
476 *velezensis*-induced vaterite, confirming that organic components accounted for a
477 larger proportion of the former mineral structure (Figs. 2a and S1b). Biogenic vaterite
478 induced by the two bacterial strains has been explored to show the potential for heavy



479 metal and organic pollutant treatment. For Cu^{2+} pollution, *B. velezensis*-induced
 480 biogenic vaterite had micro- to nanoscale particles and a large specific surface area,
 481 which contribute to its effective removal of heavy metal ions (Liu et al., 2019). In
 482 contrast, the *P. putida*-induced vaterite had a smaller specific surface area and weaker
 483 Cu^{2+} adsorption performance. For TC pollution, the removal efficiencies of biogenic
 484 vaterite induced by *B. velezensis* and *P. putida* differed only slightly, and both
 485 materials effectively reduced the TC concentration (Fig. 6a and b). The large surface
 486 pore size and specific surface area of bacterially induced biogenic vaterite support its
 487 potential application in pollutant remediation. However, the differences in bacterial
 488 sources produced significant variations in mineral morphology and surface structure,
 489 which affected pollutant adsorption performance. Therefore, *B. velezensis*-induced
 490 biogenic vaterite is more competitive for applications such as agricultural soil
 491 improvement. The proposed model by which the two bacteria induced calcium
 492 carbonate mineralization and retained combined pollutants is shown in Fig. 7.



494 Fig. 7 Proposed schematic model of calcium carbonate mineralization and combined
 495 pollutant immobilization induced by the two bacterial strains.

496 4.3 Environmental Application Potential of Biogenic Vaterite

497 Compared with chemically synthesized vaterite, which is prone to phase



498 transformation into aragonite and calcite, bacterially induced biogenic vaterite has a
499 more stable structure and can sustain long-term adsorption capacity. This stability is
500 mainly associated with the participation of organic matter, particularly proteins, in
501 nucleation (Liu et al., 2023; Sarkar et al., 2010; Wang et al., 2009). The specific
502 surface area of *B. velezensis*-induced vaterite reached 13.19 m²/g, which was
503 substantially higher than that of *P. putida*-derived vaterite (2.56 m²/g). The loose and
504 porous structure of *B. velezensis*-induced vaterite provided abundant adsorption sites
505 and enabled a removal rate of 84.93% for high concentrations of Cu²⁺ at 50 mg/L.
506 This removal performance was higher than that reported for several conventional
507 adsorbents, including biochar and graphene oxide (Park et al., 2016; Tan et al., 2015).

508 Biogenic vaterite is derived from microbial metabolism and is expected to
509 exhibit favorable potential compatibility with environmental matrices, such as soil
510 and water. Unlike chemically synthesized materials, its production does not introduce
511 residual toxic additives from the synthesis process. Because it is composed mainly of
512 calcium carbonate and natural organic matter, biogenic vaterite has the potential to
513 undergo natural degradation or transformation in the environment with a low risk of
514 secondary pollution. In the Cu²⁺–TC combined pollution system, the *B.*
515 *velezensis*-induced vaterite maintained a Cu²⁺ removal rate above 84%, whereas TC
516 removal decreased slightly because of competitive adsorption. This dual-removal
517 behavior is important because conventional single-function adsorbents often have
518 limited ability to treat inorganic and organic contaminants simultaneously, thereby
519 expanding the potential application scenarios of biogenic vaterite (Fig. 6d). However,
520 its practical implementation is still restricted by high cultivation costs and low
521 production efficiency. The LB medium used in this study was costly, and the
522 medium-related raw material costs would represent a major proportion of the total
523 expenses during scale-up. In addition, *B. velezensis*-induced mineralization required 7
524 d to reach Ca²⁺ mineralization equilibrium, and only 0.36 g of biogenic vaterite was
525 obtained from a 100 mL culture system (Fig. 1c). This low yield limits its direct
526 utilization for large-scale pollution remediation. To reduce production costs, future
527 studies should explore the substitution of LB medium with agricultural waste, such as



528 straw hydrolysate and livestock manure fermentation broth (Li et al., 2023c). For
529 practical applications, biogenic vaterite may be prioritized for small-scale,
530 high-concentration, and high-risk pollution scenarios, including heavy
531 metal–antibiotic co-contaminated soils around livestock farms and industrial
532 wastewater pretreatment. Its combination with existing remediation technologies,
533 such as soil leaching and biological composting, may further improve overall
534 remediation efficiency through functional complementation.

535 **5. Conclusion**

536 This study compared the characteristics, formation mechanisms, and simulated
537 environmental applications of vaterite induced by *B. velezensis* and *P. putida* under
538 identical culture conditions. As a Gram-positive strain, *B. velezensis* showed a
539 stronger calcium carbonate mineralization capacity than the Gram-negative strain *P.*
540 *putida*, with carbonate mineralization percentages of 35.06% and 27.50%,
541 respectively. The mineralization capacity was mainly associated with EPS
542 composition and carbonic anhydrase activity, whereas the crystal type of bacterially
543 induced calcium carbonate was regulated by extracellular proteins. The *B.*
544 *velezensis*-induced vaterite developed a loose and porous surface structure with a
545 relatively large specific surface area. The removal rates for high-concentration Cu^{2+}
546 and TC reached 84.37% and 80.81%, respectively, indicating its potential
547 applicability in the remediation of heavy metal and organic pollutant contamination.

548 **Data availability.**

549 All datasets obtained for this study are included in the manuscript/Supplementary
550 Material.

551 **Author contributions.**

552 Xiaofang Li: Conceptualization, Methodology, Formal analysis, Writing original draft
553 – review & editing, Funding acquisition. Wenjun Nie: Conceptualization,
554 Methodology, Validation. Zhen Guo: Formal analysis, Investigation. Lin Tang:
555 Methodology, Formal analysis. Fuguo Liu: Formal analysis, Investigation. Bin Lian:
556 Methodology, Formal analysis, Writing – review & editing



557 **Competing interests.**

558 The contact author has declared that none of the authors has any competing interests.

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563 **References**

564 Castanier, S., Métayer-Levrel, G.L., and Perthuisot, J.P.: Bacterial roles in the
565 precipitation of carbonate minerals. Springer Berlin Heidelberg. 2000

566 Diaz, M.R., and Eberli, G. P.: Decoding the mechanism of formation in marine ooids:
567 A review, *Earth-Science Reviews.*, 190, 536-556,
568 <https://doi.org/10.1016/j.earscirev.2018.12.016>, 2019.

569 Dhami, N.K., Reddy, M.S., and Mukherjee, A.: Biomineralization of calcium
570 carbonates and their engineered applications: a review. *Front. Microbiol.*, 4.
571 <https://doi.org/10.3389/fmicb.2013.00314>, 2013.

572 Dupraz, C., Reid, R.P., Braissant, O., Decho, A.W., Norman, R.S., and Visscher, P.T.:
573 Processes of carbonate precipitation in modern microbial mats. *Earth-Sci. Rev.*,
574 96(3), 141-162. <https://doi.org/10.1016/j.earscirev.2008.10.005>, 2009.

575 Gao, X., Li, J., Hu, K., Zhao, Y., Han, Y., Liu, F., Tucker, M.E., and Han, Z.Z.:
576 Calcification of cell membranes: From ions to minerals. *Chem. Geol.*, 617,
577 121266, <https://doi.org/10.1016/j.chemgeo.2022.121266>, 2023.

578 Han, J., Lian, B., and Ling, H.: Induction of calcium carbonate by *Bacillus cereus*.
579 *Geomicrobiol. J.*, 30(8), 682-689, 10.1080/01490451.2012.758194, 2013.

580 Hou, W., Lian, B., and Zhang X.: CO₂ mineralization induced by fungal nitrate
581 assimilation. *Bioresource Technol.*, 102, 1562-1566,
582 [10.1016/j.biortech.2010.08.080](https://doi.org/10.1016/j.biortech.2010.08.080), 2011.

583 Jin, B., Wang, S., Lei, Y., Jia, H., Niu, Q., Dapaah, M., Gao, Y., and Cheng, L.: Green
584 and effective remediation of heavy metals contaminated water using CaCO₃
585 vaterite synthesized through biomineralization. *J. Environ. Manage.*, 353, 120136,



- 586 <https://doi.org/10.1016/j.jenvman.2024.120136>, 2024,
- 587 Lian, B., Hu, Q., Chen, J., Ji, J., and Teng, H.: Carbonate biomineralization induced by
588 soil bacterium *Bacillus megaterium*. *Geochim. Cosmochim. Ac.*, 70(22),
589 5522-5535. [10.1016/j.gca.2006.08.044](https://doi.org/10.1016/j.gca.2006.08.044), 2006.
- 590 Li, H., Yao, Q., Wang, F., Huang, Y., Fu, S., and Zhou, G.: Insights into the formation
591 mechanism of vaterite mediated by a deep-sea bacterium *Shewanella*
592 *piezotolerans* WP3. *Geochim. Cosmochim. Ac.*, 256, 35-48,
593 [10.1016/j.gca.2018.06.011](https://doi.org/10.1016/j.gca.2018.06.011), 2019.
- 594 Li, X., He, X., Ren, K., Dong, H., and Lian B.: Mechanisms of carbonate precipitation
595 induced by two model bacteria. *Chem. Geol.* 628, 121461,
596 <https://doi.org/10.1016/j.chemgeo.2023.121461>, 2023a.
- 597 Li, S., and Lian, B.: Application of calcium carbonate as a controlled release carrier
598 for therapeutic drugs. *Minerals-Basel*, 13(9), 1136,
599 <https://doi.org/10.3390/min13091136>, 2023b.
- 600 Li, J., Feng, Y., Qiu, Y., Chen, D., Liang, D., Zhou, J., and Liu, G.: Recovery of
601 electron and carbon source from agricultural waste corncob by microbial
602 electrochemical system to enhance wastewater denitrification. *Sci. Total Environ.*,
603 878, 162926, <https://doi.org/10.1016/j.scitotenv.2023.162926>, 2023c.
- 604 Lyu, J., Ma, F., Li, F., Zhang, C., and Chen, J.: Vaterite induced by *Lysinibacillus* sp.
605 GW-2 strain and its stability. *J. Struct. Biol.*, 200 (2), 97-105,
606 [10.1016/j.jsb.2017.09.008](https://doi.org/10.1016/j.jsb.2017.09.008), 2017.
- 607 Lyu, J., Li, F., Zhang, C., Gower, L., Wasman, S., Sun, J., Yang, G., Chen, J., Gu, L.,
608 Tang, X., and Scheiffele, G. From the inside out: Elemental compositions and
609 mineral phases provide insights into bacterial calcification. *Chem. Geol.*, 559,
610 119974, <https://doi.org/10.1016/j.chemgeo.2020.119974>, 2021.
- 611 Liu, R., Huang, S., Zhang, X., Song, Y., He, G., Wang, Z., and Lian, B.:
612 Bio-mineralisation, characterization, and stability of calcium carbonate
613 containing organic matter. *Rsc Adv.*, 11(24), 14415-14425,
614 <https://doi.org/10.1039/D1RA00615K>, 2021.
- 615 Liu, R., and Lian, B. Immobilisation of Cd(II) on biogenic and abiotic calcium
616 carbonate. *J. Hazard. Mater.*, 378, 120707-120718,
617 <https://doi.org/10.1016/j.jhazmat.2019.05.100>, 2019a.



- 618 Liu, R., and Lian, B.: Non-competitive and competitive adsorption of Cd^{2+} , Ni^{2+} , and
619 Cu^{2+} by biogenic vaterite. *Sci. Total Environ.*, 659, 122-130,
620 <https://doi.org/10.1016/j.scitotenv.2018.12.199>, 2019b.
- 621 Liu, H., Huang H., Nie W., Sun M., Li, X., and Lian, B.: Facilitated mechanism of
622 biological vaterite stability mediated by *Bacillus velezensis* and its secretions.
623 *ACS Earth Space Chem.*, 7(10), 2019-2030.
624 <https://doi.org/10.1021/acsearthspacechem.3c00135>, 2023.
- 625 Park, J.H., Ok, Y.S., Kim, S.H., Cho, J.S., Heo, J.S., Delaune, R.D., and Seo, D.C.:
626 Competitive adsorption of heavy metals onto sesame straw biochar in aqueous
627 solutions. *Chemosphere*, 142, 77-83,
628 <https://doi.org/10.1016/j.chemosphere.2015.05.093>, 2016.
- 629 Rathod, M., Haldar, S., and Basha, S.: Nanocrystalline cellulose for removal of
630 tetracycline hydrochloride from water via biosorption: Equilibrium, kinetic and
631 thermodynamic studies. *Ecological Engineering*, 84, 240-249,
632 <https://doi.org/10.1016/j.ecoleng.2015.09.031>, 2015
- 633 Sarkar, A., and Mahapatra, S.: Synthesis of all crystalline phases of anhydrous
634 calcium carbonate. *Cryst. Growth Des.*, 10 (5), 2129-2135,
635 <https://doi.org/10.1021/cg9012813>, 2010.
- 636 Sheng, G., Wang, S., Hu, J., Lu, Y., Li, J., Dong, Y., and Wang, X.: Adsorption of
637 $\text{Pb}(\text{II})$ on diatomite as affected via aqueous solution chemistry and temperature.
638 *Colloids Surfaces A*, 339(1), 159-166, [10.1016/j.colsurfa.2009.02.016](https://doi.org/10.1016/j.colsurfa.2009.02.016), 2009.
- 639 Tan, P., Sun, J., and Hu, Y.: Adsorption of Cu^{2+} , Cd^{2+} and Ni^{2+} from aqueous single
640 metal solutions on graphene oxide membranes. *J. Hazard. Mater.*, 297, 251-260,
641 <https://doi.org/10.1016/j.jhazmat.2015.04.068>, 2015.
- 642 Wang, X., Kong, R., Pan, X., Xu, H., Xia, D., Shan, H., Lu, J.: Role of Ovalbumin in
643 the stabilization of metastable vaterite in calcium carbonate biomineralization. *J.*
644 *Phys. Chem. B*, 113 (26), 8975-8982, <https://doi.org/10.1021/jp810281f>, 2009.
- 645 Wang X., Yang L., Zhang J., Wang, C., and Li, Q.: Preparation and characterization of
646 chitosan - poly(vinyl alcohol)/bentonite nanocomposites for adsorption of $\text{Hg}(\text{II})$
647 ions. *Chem. Eng. J.*, 251(9), 404-412, [10.1016/j.cej.2014.04.089](https://doi.org/10.1016/j.cej.2014.04.089), 2014.



- 648 Wang, Y., Konstantinou, C., Tang, S., and Chen, H., Applications of
649 microbial-induced carbonate precipitation: A state-of-the-art review.
650 Biogeotechnics, 1(1), 100008. <https://doi.org/10.1016/j.bgtech.2023.100008>,
651 2023.
- 652 Wang, H., Dong, Y., Yang Y, Toor, G.S., and Zhang, X.M.: Changes in heavy metal
653 contents in animal feeds and manures in an intensive animal production region of
654 China. J. Environ. Sci., 25(12), 2435-2442. 10.1016/S1001-0742(13)60473-8,
655 2013.
- 656 Wang, X., Yan, X., and Li, X.: Environmental safety risk for application of anaerobic
657 fermentation biogas slurry from livestock manure in agricultural land in China.
658 Scientia Agricultura Sinica, 54 (1) , 110-139, 2021.
- 659 Whitaker, J.M., Vanapalli, S., and Fortin, D.: Improving the strength of sandy soils via
660 ureolytic CaCO₃ solidification by *Sporosarcina ureae*. Biogeosciences, 15 (14),
661 4367-4380, <https://doi.org/10.5194/bg-15-4367-2018>, 2018.
- 662 Xiao, L., and Lian, B.: Heterologously expressed carbonic anhydrase from *Bacillus*
663 *mucilaginosus* promoting CaCO₃ formation by capturing atmospheric CO₂.
664 Carbonates and Evaporites, 31(1), 39-45. 10.1007/s13146-015-0239-4, 2016.
- 665 Xiao, M., Xu, S., Yang, B., Zeng, G., Qian, L., Huang, H., and Ren, S. L.:
666 Contamination, source apportionment, and health risk assessment of heavy
667 metals in farmland soils surrounding a typical copper tailings pond. Int. J. Env.
668 Res. Pub. He., 19, 14264, <https://doi.org/10.3390/ijerph192114264>, 2022.
- 669 Xu, Y., Zhu, L., Chen, S., Li, R., Li, J., Yuan, J., Wen, T., Xue, C., and Shen, Q.: Risk
670 assessment and dissemination mechanism of antibiotic resistance genes in
671 compost. Environ. Int., 178, 108126, 10.1016/j.envint.2023.108126, 2023.
- 672 Yu, L., Li, Q., Chen C., Lu Z., and Wang X.: The effects of two kinds of extracellular
673 monosaccharides on the mineralization of calcium carbonate. Acta Petrologica et
674 Mineralogica, 37(3), 497-504, 2018.
- 675 Zhang, C., Li, X., Lyu, J., and Li, F.: Comparison of carbonate precipitation induced
676 by *Curvibacter* sp. HJ-1 and *Arthrobacter* sp. MF-2: Further insight into the
677 biomineralization process. Journal of Structural Biology, 212(2), 107609,



- 678 10.1016/j.jsb.2020.107609, 2020.
- 679 Zhang, C., Li, F., Lyu, J., and Yao, Y.: Biomimetic mineralization of Ca-Mg
680 carbonates: relevance to microbial cells and extracellular polymeric substances,
681 *Microsc. Microanal.*, 29, (2), 665-674 <https://doi.org/10.1093/micmic/ozac044>,
682 2023.
- 683 Zhang, L., Sun, P., An, X., Wang, X., Li, S., and Lian, B., Vaterite synthesized by
684 waste liquid of extracting chitin from crab shells and the mineral loading for
685 doxorubicin hydrochloride. *Minerals-Basel*, 12(12), 1608
686 <https://doi.org/10.3390/min12121608>, 2022.
- 687 Zhuang, D., Yan, H., Tucker, M., Zhao, H., Han, Z., Zhao, Y., Sun, B., Li, D., Pan, J.,
688 Zhao, Y., Meng, R., Shan, G., Zhang, X., and Tang, R.: Calcite precipitation
689 induced by *Bacillus cereus* MRR2 cultured at different Ca²⁺ concentrations:
690 Further insights into biotic and abiotic calcite. *Chem. Geol.*, 500, 64-87,
691 <https://doi.org/10.1016/j.chemgeo.2018.09.018>, 2018
- 692 Zhu, C., Chen, T., and Zhao, L.: Magnesium partitioning into vaterite and its potential
693 role as a precursor phase in foraminiferal Mg/Ca thermometer. *Earth and
694 Planetary Science Letters*, 567, 116989,
695 <https://doi.org/10.1016/j.epsl.2021.116989>, 2021.
- 696