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1 **Stoichiometric deviation and regulatory mechanisms of AOU-nutrient ratio in**
2 **the oligotrophic Northwest Pacific Ocean**

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13 **Abstract**

14 In oligotrophic oceans, the stoichiometric ratios of apparent oxygen utilization
15 (AOU) to nutrients often deviate from the classical Redfield ratio, yet the mechanisms
16 driving these deviations remain poorly constrained. Contrary to the commonly held
17 view that ratios of AOU to nutrients are typically elevated, our study found that the
18 mean ratios of AOU to dissolved inorganic nitrogen (DIN) and AOU to dissolved
19 inorganic phosphorus (DIP) in the upper 2000 m of the oligotrophic Northwest Pacific
20 are substantially lower than the classical Redfield ratios (8.6 and 138, respectively),
21 measuring only 6.28 and 86.79, respectively. Physical mixing alone cannot explain
22 these low ratios, as the region is strongly stratified. This persistent vertical isolation
23 drives chronic nutrient limitation in surface waters, promoting phytoplankton to
24 produce carbon-rich transparent exopolymer particles (TEPs) with high C:N ratios.
25 Meanwhile, the microbial community, dominated by *Pelagibacter* and *Alteromonas*,
26 exhibits functional partitioning. *Pelagibacter* efficiently recycles small organic
27 molecules, while *Alteromonas* degrades complex polymers and actively releases



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28 phosphate. This selective processing enhances nutrient regeneration relative to carbon
29 oxidation, depressing the AOU/nutrient ratios. These findings suggest that
30 biogeochemical models should account for such biological feedbacks to improve
31 predictions of ocean carbon export and nutrient cycling under future climate scenarios.

32 **1. Introduction**

33 The ocean is the largest active reservoir of carbon and nutrients on Earth, and its
34 internal biogeochemical cycles are crucial for sustaining the productivity of marine
35 ecosystems (Friedlingstein et al., 2025). The supply of nutrients such as nitrogen and
36 phosphorus directly affects marine primary productivity, and the regeneration of these
37 nutrients is closely coupled with the production, export, and burial of organic carbon,
38 thereby collectively regulating the intensity of the marine biological pump and the
39 global carbon cycle (Longhurst and Glen Harrison, 1989; Smith et al., 2008). Apparent
40 oxygen utilization (AOU) serves as a key link between surface photosynthesis and
41 deep-water heterotrophic respiration and reflects the amount of dissolved oxygen (DO)
42 consumed by respiration during the transport and sedimentation of organic matter in
43 water masses (Ito et al., 2004; Sulpis et al., 2023). Classical Redfield stoichiometry
44 establishes a proportional relationship between AOU, nutrient regeneration, and the
45 accumulation of dissolved inorganic carbon (DIC), which has led to the widespread use
46 of AOU/nutrient ratios, such as AOU/dissolved inorganic nitrogen (DIN) and
47 AOU/dissolved inorganic phosphorus (DIP) ratios, to infer the elemental composition
48 of degraded organic matter (Tanioka and Matsumoto, 2020). However, the relationships
49 between AOU and nutrients in marine environments often deviate from classical
50 stoichiometry, highlighting the presence of complex physical and biogeochemical
51 regulatory mechanisms.

52 Numerous observations indicate that the AOU/nutrient ratios in many oceanic
53 regions generally exceed the classical Redfield ratio (i.e., $\text{AOU/DIN} = 8.6$ and
54 $\text{AOU/DIP} = 138$) (Anderson and Sarmiento, 1994; Delaigue et al., 2024). This pattern
55 is particularly evident in highly stratified oligotrophic waters, where the degradation
56 rates of suspended and dissolved organic matter are slow, and spatiotemporal



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57 decoupling between production and remineralization occurs (Dai et al., 2023). In these
58 areas, some organic matter may migrate horizontally before settling or undergo multiple
59 cycles within surface waters. Simultaneously, the C:N ratio of nitrogen sources supplied
60 by nitrogen-fixing organisms is generally higher than the Redfield value, resulting in
61 greater oxygen consumption per unit of nitrogen regenerated during
62 degradation (Körtzinger et al., 2001; Singh et al., 2015). Furthermore, DIC generated
63 by shallow respiration as a result of organic matter degradation by heterotrophic
64 organisms may precede nutrient regeneration, and potential non-photosynthetic
65 oxidation processes also contribute to elevated AOU/nutrient ratios in oligotrophic
66 regions (Calleja et al., 2019). In contrast, in well-mixed marginal shelf seas where
67 organic matter sources are complex, strong physical agitation disrupts the coupling
68 between AOU and nutrients. This leads to highly scattered ratios that are systematically
69 lower than the Redfield values, reflecting the influence of physical processes on
70 biogeochemical signals (Zhu et al., 2025). To distinguish physical from biological
71 influences, chemical tracers such as preformed nitrate (PreNO₃) and preformed
72 phosphate (PrePO₄) are widely used, as they effectively separate nutrients pre-existing
73 in water masses from those regenerated locally (Letscher and Villareal, 2018).
74 Applying such tracers to strongly stratified oligotrophic oceans, where physical mixing
75 is weak, facilitates the isolation of biogeochemical signals, thereby revealing the
76 dominant role of internal processes in shaping the observed AOU-nutrient relationships.

77 The Tropical Northwest Pacific (TNWP) is a key region for global ocean
78 circulation and biogeochemical cycling, featuring the Western Pacific Warm Pool
79 (WPWP), the world's largest warm-water body, where high sea surface temperatures
80 and strong stratification create an extensive oligotrophic zone (Radenac et al., 2013).
81 Complex ocean circulation systems converge here, forming vertically layered water
82 masses that significantly influence nutrient supply, organic matter production, transport,
83 and transformation (Hu, D et al., 2015; Jian et al., 2022; Sun, C et al., 2008). Numerous
84 studies have characterized the macroscopic distribution of nutrients in this region (Ma
85 et al., 2021a) as well as the vertical degradation patterns of particulate organic carbon



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86 (POC) (Tian et al., 2025). However, POC is a complex mixture comprising components
87 with diverse chemical compositions and sources, and its various components differ
88 significantly in terms of bioavailability and their roles in the carbon cycle (Kharbush et
89 al., 2020). Therefore, the roles of specific functional particulate components within
90 POC in regional organic matter transport and transformation have gradually become a
91 focus of research in recent years (Digernes et al., 2025). Among these components,
92 transparent exopolymer particles (TEPs) is a class of polysaccharide colloid primarily
93 secreted by phytoplankton, characterized by a high C:N ratio (Engel and Passow, 2001).
94 They accumulate in large quantities in oligotrophic surface waters and may alter the
95 kinetics and degradation pathways of particulate matter (Kodama et al., 2014).
96 Meanwhile, regional differences in microbial community structure and function are
97 considered to directly affect organic matter degradation and nutrient regeneration
98 processes, thereby regulating the stoichiometry of deep nutrient regeneration and
99 efficiency of the carbon cycle (Guo et al., 2023; Saavedra et al., 2025). However,
100 although numerous studies have revealed several biogeochemical characteristics of this
101 region from different perspectives, our overall understanding remains relatively
102 fragmented. In particular, the stoichiometric characteristics of organic matter
103 degradation in water masses of varying sources and ages, as well as the mechanisms
104 driving deviations from the Redfield ratios, remain unclear. Additionally, there is a lack
105 of systematic field evidence regarding how microbial metabolic strategies influence the
106 coupled cycles of carbon, nitrogen, and phosphorus.

107 In the present study, suspended particulate organic matter (POM) and water
108 samples were collected from six water columns within and around the core area of the
109 TNWP Warm Pool. The aim was to elucidate the coupled mechanism linking physical
110 stratification, organic matter transformation, nutrient regeneration, and microbial
111 regulation in the oligotrophic waters of the western Pacific by integrating multiple
112 biogeochemical and microbial parameters across distinct water masses.

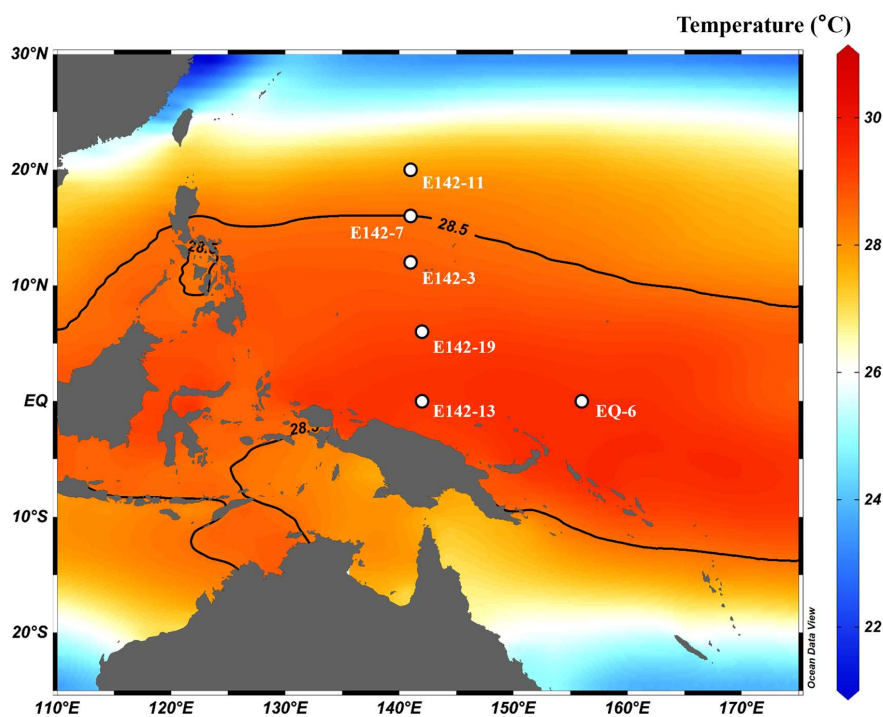
113 **2. Materials and Methods**

114 **2.1 Research area and sample collection:**



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115 From February to April 2022, a research cruise was conducted aboard the R/V
116 *Kexue* (voyage NORC2021-09) in the TNWP Warm Pool. During the survey, a total of
117 six sampling stations were set up: EQ-6, E142-3, E142-7, E142-11, E142-13, and E142-
118 19 (Fig. 1).



119

120 **Figure 1. Study area and sampling stations in the Tropical Northwest Pacific. The**
121 **sampling area is located in the Western Pacific Warm Pool, and the 28.5 °C isotherms**
122 **are shown as black lines.**

123 Seawater samples were collected from 16 depths (5, 50, 100, 150, 200, 300, 400,
124 500, 600, 800, 1,000, 1,200, 1,400, 1,600, 1,800, and 2,000 m) using 12 L Niskin bottles
125 (KC-Denmark, Denmark) mounted on a shipboard CTD rosette (Conductivity-
126 Temperature-Depth, Sea-bird SBE911, United States). After seawater collection, DO
127 samples were first aliquoted and fixed in 50 mL amber glass bottles. Then, the seawater
128 was filtered through pre-combusted and acid-washed GF/F glass fiber filters (0.7 µm
129 pore size, Whatman; 450 °C for 4 h, 0.5 M HCl for 24 h), followed by the collection of



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130 nutrient and DIC samples. Samples for nutrient determination were aliquoted into 250
131 mL high-density polyethylene bottles, fixed with chloroform, and stored at -20°C
132 before measurement. Samples for DIC analysis were collected in 50 mL airtight glass
133 bottles. After overflow, 1 mL of the water sample was withdrawn and the remaining
134 sample was poisoned with saturated mercuric chloride solution to suppress biological
135 activity. The samples were stored at 4°C until analysis.

136 Suspended particulate samples for POC analysis were collected by filtering 2–4
137 L of seawater through pre-combusted and acid-washed GF/F glass fiber filters (0.7 µm
138 pore size, Whatman; 450 °C for 4 h, 0.5 M HCl for 24 h). Samples were stored at -
139 20 °C until measurement. TEP samples were collected by filtering 1–4 L of seawater
140 through polycarbonate filters (0.45 µm pore size, Whatman) under a vacuum of < 170
141 mmHg. The filtered membrane was stained with 0.5–1 mL of a 400 mg L⁻¹ Alcian Blue
142 solution (Sigma-Aldrich, USA) for 5–10 s, rinsed with 5 mL of MilliQ water, and
143 stored at -20 °C until measurement (three replicate samples were collected at each
144 depth). Samples for chlorophyll-a (Chl-*a*) determination were collected from five
145 depths between 5 and 200 m. First, 2 L of seawater samples were filtered through a 200
146 µm mesh to remove zooplankton and then filtered onto a 0.7 µm GF/F glass fiber
147 membrane (Whatman; 450 °C for 4 h, 0.5 M HCl for 24 h). The filters were then stored
148 at -20 °C until analysis in the shore-based laboratory. Seawater samples (20 L) for
149 microbial community analysis were similarly pre-filtered through a 200 µm mesh and
150 then filtered through a 0.22 µm mixed cellulose ester (MCE) membrane. The filtered
151 samples were immediately transferred to cryovials, supplemented with RNA Later
152 stabiliser (Qiagen, Germany), and stored at -80 °C.

153 **2.2. Chemical analysis**

154 **2.2.1 Temperature, salinity, DO, and Chl-*a***

155 For each station, temperature, salinity, and DO were measured in situ using a
156 shipboard CTD. Simultaneously, discrete DO samples were fixed and titrated according
157 to the Winkler method with an accuracy of 0.22 µmol L⁻¹ (Bryan et al., 1976). These



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158 discrete DO samples were used to calibrate the DO concentration data obtained from
159 the CTD sensor. Chl-*a* concentrations were determined by acetone extraction. Filters
160 containing Chl-*a* were extracted with 90 % acetone for 14 h, and the Chl-*a*
161 concentration was then measured using a Turner fluorometer (Turner Designs, United
162 States) (Ma et al., 2020).

163 2.2.2 Nutrients

164 Phosphate (PO₄-P, as dissolved inorganic phosphorus, DIP), nitrate (NO₃-N),
165 nitrite (NO₂-N), ammonium (NH₄-N), and silicate (SiO₃-Si, as dissolved inorganic
166 silicon, DSi) concentrations were determined using a QuAAtro nutrient analyzer
167 (SEAL, Germany). The limits of detection for PO₄-P, NO₃-N, NO₂-N, and NH₄-N were
168 0.01, 0.02, 0.01, 0.01 and 0.01 μmol L⁻¹, respectively (Ma et al., 2020). The DIN
169 concentration was the sum of the concentrations of NO₃-N, NO₂-N, and NH₄-N.

170 2.2.3 POC and DIC

171 The particulate filters were freeze-dried at -50 °C for 24 h. After drying, the filters
172 were acid-fumed with concentrated hydrochloric acid for 12 h to remove inorganic
173 carbonates and then dried at < 60 °C. The dried membrane samples were wrapped in a
174 tin boat for subsequent analysis. The pretreated samples were analyzed for POC
175 concentration using an elemental analyzer (Thermo Fisher Scientific Flash EA 1112,
176 United States), with an analytical precision of ± 0.8%. POC measurements were
177 calibrated using standard reference materials, including USGS64 (C % = 31.97 %,
178 Indiana University), USGS40 (C % = 40.8 %, U.S. Geological Survey), and Urea #2a
179 (C % = 20 %, Indiana University) (Ma et al., 2021b).

180 DIC concentrations were determined using an Apollo SciTech AS-C3 analyzer
181 (United States) with an analytical precision of ±0.1 % (Ma et al., 2020). Calibration
182 was performed using certified reference material (batch 144, 2031.53 ± 0.62 μmol L⁻¹)
183 provided by the Scripps Institution of Oceanography (University of California, San
184 Diego).

185 2.2.4 TEP



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186 In the shore-based laboratory, the TEP filters were transferred to a 10 mL amber
187 glass bottle, and 6 mL of 80 % (14.4 M) sulfuric acid solution was added for 10 h with
188 periodic gentle mixing. The absorbance was then measured at 787 nm using a
189 spectrophotometer (UV-1600, Shimadzu). The absorbance was converted to xanthan
190 gum equivalents (Xeq) weight using a calibration curve derived by Bittar et al. (2018),
191 with a detection limit of approximately 5.6 $\mu\text{g Xeq L}^{-1}$ (Bittar et al., 2018).

192 **2.2.5 Relative abundance of microorganisms**

193 DNA was extracted from the filters using the AllPrep DNA/RNA Mini Kit
194 (Qiagen, Germany). 16S rDNA was amplified using a universal bacterial primer set,
195 F515 (5'GTGCCAGCMGCCGCGG3') and R907 (5'CCGTCAATTCMTTTR
196 AGTTT3'). To reduce PCR bias, three replicate PCR reactions were performed for each
197 sample. The amplicon libraries were sequenced using the Illumina NovaSeq platform
198 (Novogene, China). During data processing, primers and low-quality reads were
199 removed using VSEARCH (v. 2.15.2), and sequences were denoised using USEARCH
200 (v. 10.0.240) to obtain an amplicon sequence variant (ASV) table. The amplicon
201 sequence variants (ASVs) were then analyzed using VSEARCH (v. 2.15.2) against a
202 reference database. The RDP database (v. 16) was used to annotate the ASV tables. The
203 ASV tables were standardized by resampling to ensure that the data from each sample
204 are at the same level of comparison. The number of sequences assigned to each ASV
205 in each sample was normalized to its proportion of total reads to calculate the relative
206 abundance of microbial taxa (Liu et al., 2024).

207 **2.3 Tracer computations**

208 To account for the non-Redfield stoichiometric characteristics of dissolved
209 organic matter (DOM), we applied an improved method for calculating residual PreNO₃
210 tracer (Letscher and Villareal, 2018):

$$211 \quad \text{residual PreNO}_3 = \text{NO}_{3\text{meas}} - (f_{\text{DOM}} \cdot \frac{\text{AOU}}{r_{\text{DOM}}}) - (f_{\text{POM}} \cdot \frac{\text{AOU}}{r_{\text{POM}}}) \quad (1),$$



212 where measured nitrate ($\text{NO}_{3\text{meas}}$) is defined as the sum of the concentrations of
 213 $\text{NO}_3\text{-N}$ and $\text{NO}_2\text{-N}$. AOU is given by:

$$214 \quad \text{AOU} = \text{DO}_{\text{sat}} - \text{DO}_{\text{meas}} \quad (2),$$

215 where DO_{sat} is the saturated DO concentration at a given temperature, salinity, and
 216 pressure (Garcia and Gordon, 1992), and DO_{meas} is the measured DO concentration.

217 f_{DOM} and f_{POM} represent the fractions of total oxygen consumption attributed to
 218 DOM and POM, respectively. r_{DOM} and r_{POM} denote the stoichiometric ratios of O_2
 219 consumed and regenerated nitrogen consumed during the remineralization of DOM and
 220 POM, respectively. The above parameters were selected based on water depth or neutral
 221 density (γ_n) of the water masses, following Letscher and Villareal (2018) (Table 1).

222 Similarly, the residual PrePO_4 tracer was computed as:

$$223 \quad \text{residual PrePO}_4 = \text{PO}_{4\text{meas}} - (f_{\text{DOM}} \cdot \frac{\text{AOU}}{r_{\text{DOM}}}) - (f_{\text{POM}} \cdot \frac{\text{AOU}}{r_{\text{POM}}}) \quad (3),$$

224 We convert nitrogen-based stoichiometric coefficients to phosphorus equivalents
 225 using a DON:DOP ratio of 16:1, following the stoichiometry of semi-labile DOM in
 226 North Pacific (Letscher and Villareal, 2018).

227 **Table 1. Parameters for the calculation of residual $\text{PreNO}_3/\text{PrePO}_4$ in the Pacific Ocean**

Depth (m)/neutral density (γ_n)	f_{DOM}	f_{POM}	r_{POM}	r_{DOM}
0–100 m/ γ_n : 24.2–24.7	0.5	0.5	6.9, 8.75, 10.6	18.1
$\gamma_n > 24.7$	0.5	0.5	6.9, 8.75, 10.6	18.9

228 2.4 Data Processing

229 Data processing was performed using OriginPro 2021 (v. 9.8.0.200). Depending
 230 on whether the variables follow a normal distribution, correlation analyses were
 231 performed using either Pearson’s or Spearman’s rho test (two-tailed, $\alpha = 0.05$), with p
 232 < 0.05 considered significant. The normality of the data was tested using the
 233 Kolmogorov-Smirnov test (two-tailed, $\alpha = 0.05$). The AOU/nutrient ratios were



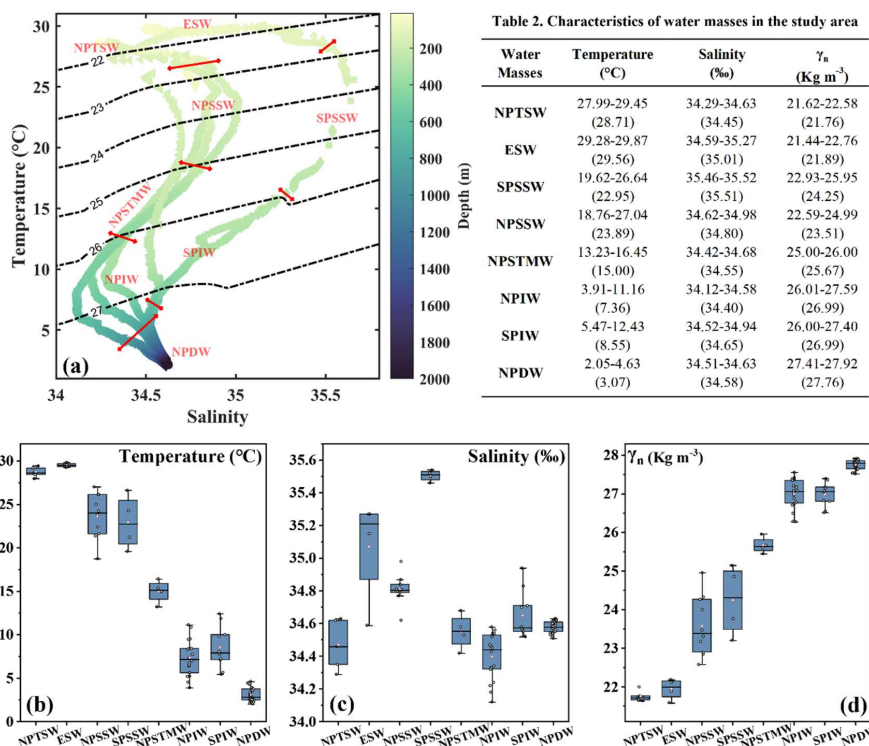
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234 estimated via simple linear regression using the ordinary least squares method, where
 235 AOU served as the independent variable and the nutrient (nitrate or dissolved inorganic
 236 phosphate) as the dependent variable. The slope of the resulting regression line
 237 represents the AOU per mole of nutrient regenerated (Álvarez-Salgado et al., 2014).

238 3. Results

239 3.1 Hydrological characteristics of the study area

240 Temperature-salinity (T-S) diagram and the vertical distribution of temperature,
 241 salinity, and γ_n indicate pronounced stratification in the water column, with eight
 242 distinct water masses identified (Fig. 2a, Table 2) (Sun et al., 2008). γ_n increases with
 243 depth from approximately 21.6 to 27.7 kg m⁻³ (Fig. 2d), marking a clear transition from
 244 low-density tropical surface waters to high-density deep waters.



245

246 **Figure 2. Physical properties of water masses in the research area. (a) Temperature-**
 247 **salinity diagram of six stations, and the distribution characteristics of (b) temperature,**



248 **(c) salinity, (d) neutral density of each water mass.**

249 In the upper 200-meter water column, tropical surface waters and subsurface
250 waters dominate. The North Pacific Tropical Surface Water (NPTSW) and Equatorial
251 Surface Water (ESW) lie at the warmest end of the T-S diagram, with temperatures
252 ranging from approximately 28–30 °C, characteristics of the WPWP. Below the mixed
253 layer, the North Pacific Subsurface Water (NPSSW) and South Pacific Subsurface
254 Water (SPSSW) exhibit pronounced salinity maxima along the 23–25 kg m⁻³
255 isopycnals. Both water masses have temperatures lower than the overlying surface
256 waters (approximately 19–27 °C). The salinity of the SPSSW is the highest in the upper
257 ocean (mean 35.51), while the NPSSW is slightly fresher (mean 34.80). These high-
258 salinity tongue-like intrusions indicate active subduction and lateral transport processes
259 within the thermocline. The North Pacific Subtropical Mode Water (NPSTMW)
260 constitutes the transition layer between the thermocline and the intermediate waters,
261 with physicochemical properties tightly constrained within a narrow range and
262 demonstrating strong vertical homogeneity (Figs. 2b-d). Below approximately
263 300–400 m, the water column is dominated by intermediate waters originating from
264 both hemispheres. The North Pacific Intermediate Water (NPIW) is characterized by
265 low salinity (mean 34.12), exhibiting a salinity minimum in the T-S diagram (Fig. 2a).
266 The South Pacific Intermediate Water (SPIW) has slightly higher temperature and
267 salinity relative to NPIW. Its γ_n value largely overlaps with those of the NPIW,
268 suggesting substantial mixing between the two within the study area. Below
269 approximately 1000 m, the North Pacific Deep Water (NPDW) becomes predominant,
270 exhibiting low temperatures, relatively uniform salinity, and high density. Its
271 hydrographic properties cluster tightly in box plots, reflecting high stability (Figs. 2b-
272 d). In summary, the water masses in the study area exhibit a precise vertical sequence:
273 warm, low-density surface waters overlie a high-salinity thermocline/modal-water layer,
274 which in turn overlies low-salinity intermediate waters and homogeneous deep waters.
275 This stratified structure creates a vertically isolated physical regime that strongly
276 suppresses vertical exchange of heat, dissolved constituents, and particulate matter.

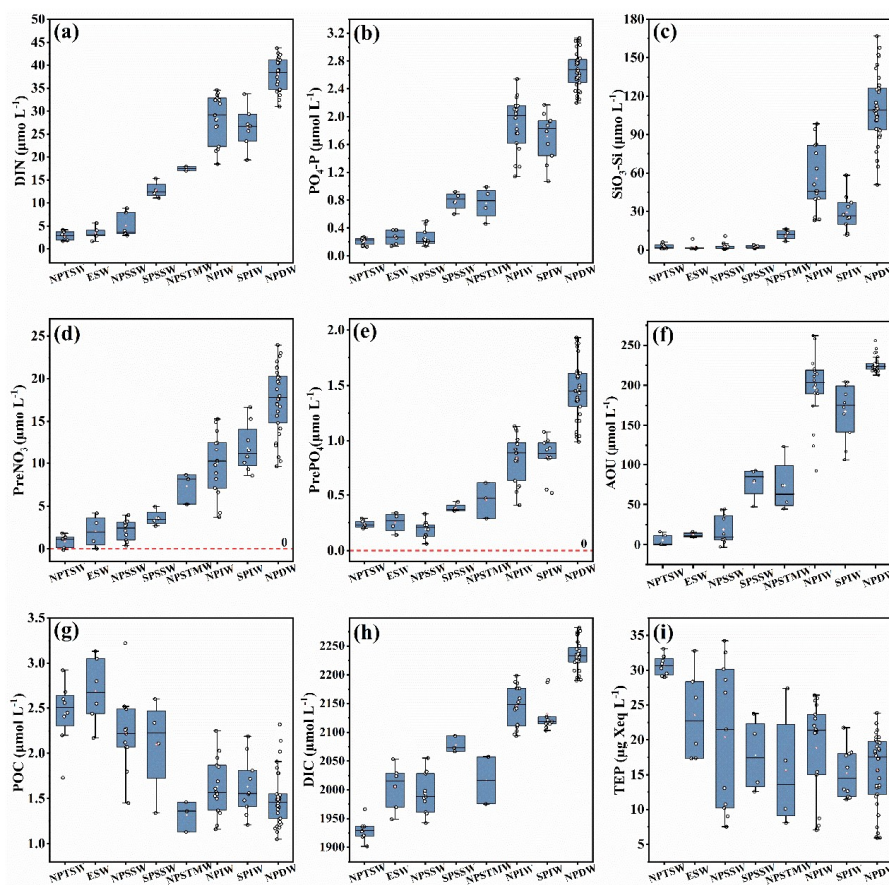


277 **3.2 Concentration and distribution of chemical parameters of each water mass**

278 Nitrate, phosphate, and silicate concentrations were lowest in warm, low-density
279 surface water masses (NPTSW, ESW) and gradually increased in the subsurface high-
280 salinity water masses (SPSSW, NPSSW, NPSTMW). Local highs were reached in the
281 intermediate water masses (NPIW, SPIW), indicating that these water masses are
282 important reservoirs for regenerated nutrients. The deep water masses (NPDW)
283 exhibited the highest nutrient concentrations, reflecting the accumulation of organic
284 matter degradation products over long time scales (Figs. 3a-c). Calculations of PreNO_3
285 and PrePO_4 based on empirical relationships between DO and nutrients further resolve
286 the proportions of nutrients supplied by physical transport versus local regeneration in
287 each water mass (Figs. 3d, e). Both PreNO_3 and PrePO_4 increased from the surface to
288 the intermediate and deep waters. Notably, PrePO_4 and PreNO_3 in surface waters should
289 approach zero theoretically, yet positive anomalies were observed in NPTSW and ESW.
290 With increasing depth, PreNO_3 and PrePO_4 concentrations in the intermediate waters
291 increase and reach their maxima in the deep waters, where their spatial distributions are
292 homogeneous, further indicating that this layer acts as an accumulation pool for long-
293 term products of biogeochemical processes.



13



294

295 **Figure 3. Distribution of (a) dissolved inorganic nitrogen (DIN), (b) phosphate (PO₄-P, as**
 296 **as dissolved inorganic phosphorus, DIP), (c) silicate (SiO₃-Si, as dissolved inorganic silicon,**
 297 **DSi), (d) preformed nitrate (PreNO₃), (e) preformed phosphate (PrePO₄), (f) apparent**
 298 **oxygen utilization (AOU), (g) particulate organic carbon (POC), (h) dissolved inorganic**
 299 **carbon (DIC), (i) transparent exopolymer particle (TEP) at each water mass.**

300 The vertical distribution pattern of DIC is consistent with that of nutrient
 301 regeneration: DIC concentrations are lowest in the surface waters, increase markedly
 302 from the subsurface and mode waters, and attain high values in the intermediate and
 303 deep waters (Fig. 3h). In contrast, POC and TEP exhibit a distribution pattern that is
 304 nearly opposite to that of nutrients and DIC among the different water masses. The
 305 surface and subsurface water masses contain higher POC and TEP contents, with TEP
 306 reaching a maximum of 34.22 µg Xeq L⁻¹, reflecting the aggregation of surface plankton



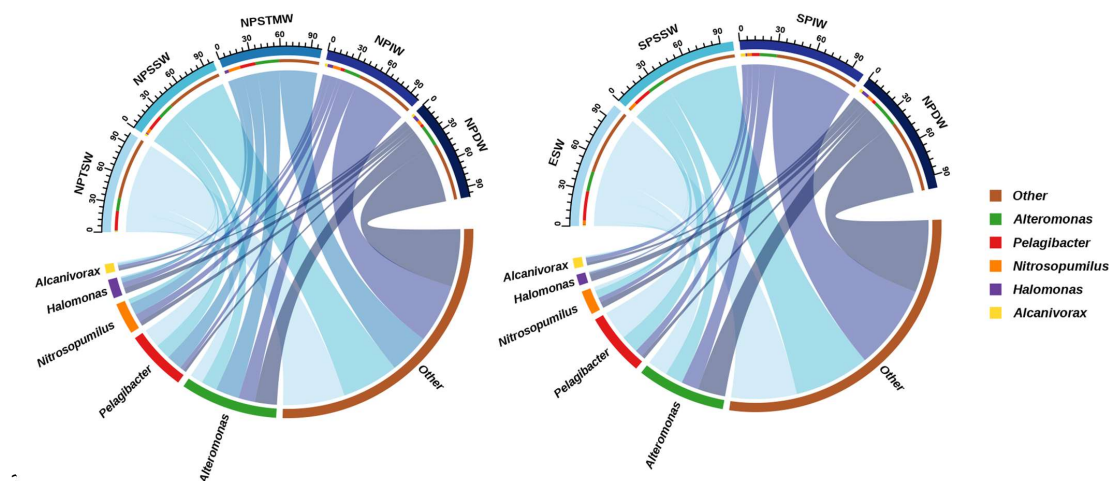
307 and bacterial extracellular secretions. POC concentrations decrease substantially in the
308 intermediate and deep waters, while the AOU continuously increased, indicating that
309 organic particles are transitioning from the production phase to the remineralization
310 phase (Figs. 3f, g). In contrast, TEP concentrations decrease more slowly and remain
311 within a measurable range even in the deep waters (Fig. 3i). The persistence of such
312 polymers in the deep waters influences the composition of organic matter and modify
313 pathways of carbon cycling.

314 **3.3 Microbial community structure**

315 Among the identified dominant genera, *Pelagibacter* and *Alteromonas* occur at
316 relatively high proportions in all water masses, followed by *Nitrosopumilus*, whereas
317 *Halomonas* and *Alcanivorax* remain minor taxa (Fig. 4). *Pelagibacter*, a typical
318 oligotrophic bacterium, has a highly streamlined genome and is efficient in utilizing
319 small organic molecules, enabling it to thrive in upper water column environments with
320 low nutrient concentrations but relatively stable dissolved organic matter fluxes.
321 Therefore, in both surface and subsurface water masses, *Pelagibacter* exhibits the
322 highest relative abundance of 15 %–24 %. In these water masses, *Alteromonas*
323 maintains a relatively stable relative abundance of 14 %–17 %, while *Nitrosopumilus*
324 accounts for only 0–6 %, and *Halomonas* and *Alcanivorax* each contribute < 2 %. In
325 the transitional water mass NPSTMW, the community structure undergoes significant
326 reorganization. The proportion of *Pelagibacter* decreases to ~16 %, while *Alteromonas*
327 increases to ~25 %, the highest among all water masses. *Nitrosopumilus* also rises to
328 13 %, and *Halomonas* increases to 4 %. This depth layer corresponds to the interface
329 between water masses, where nutrients concentrations increase sharply and POC
330 concentrations decrease markedly (Fig. 3f), and the microbial community shifts from
331 being dominated by oligotrophic groups represented by *Pelagibacter* to a structure
332 jointly dominated by opportunistic heterotrophic bacteria (e.g., *Alteromonas*) and
333 chemoautotrophic ammonia-oxidizing archaea (e.g., *Nitrosopumilus*). In intermediate
334 and deep water masses, the proportion of *Pelagibacter* further decreases to 3 %–6 %,
335 while *Alteromonas* remains at a relatively high level of 14 %–23 %, and the proportions



336 of *Halomonas* and *Alcanivorax* also increase. The high relative abundance of
 337 *Alteromonas* suggests that regenerative biogeochemical processes, characterized by the
 338 decomposition of heterotrophic organic matter and the release of nutrients, dominate in
 339 these water masses. Meanwhile, the increase of *Halomonas* and *Alcanivorax* in
 340 intermediate and deep water masses reflects the adaptation of opportunistic bacteria to
 341 high-salinity and recalcitrant organic matter environments, suggesting that these taxa
 342 may contribute to the further transformation of deep organic carbon.



344 **Figure 4. Relative abundance (%) and diversity of major microbial genera in each water**
 345 **mass**

346 **4. Discussion**

347 **4.1 Spatial variation and controlling factors of nutrient dynamics in TNWP water**
 348 **masses**

349 The AOU of the ocean water column is an important indicator of the intensity of
 350 organic matter remineralization and oxygen consumption (*Sulpis et al., 2023*). In the
 351 surface and subsurface water masses of the study area, AOU is slightly negative and
 352 close to zero, reflecting photosynthetic oxygen production and bubble injection (Fig.
 353 3f). From the surface to the intermediate water mass, AOU increases rapidly, indicating
 354 intensified organic-matter remineralization dominated by heterotrophic respiration (Fig.

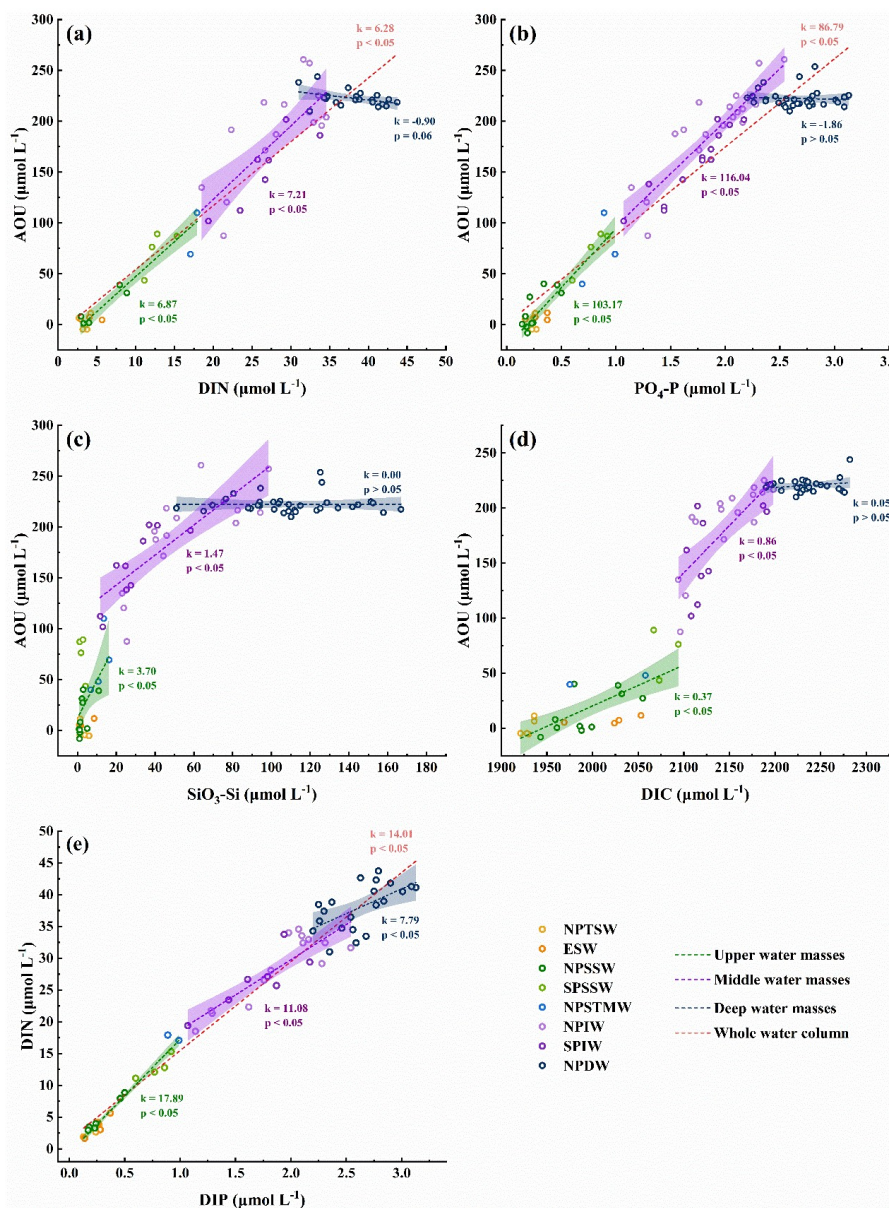


355 3f). In NPDW, AOU increases to a maximum of approximately $225 \mu\text{mol L}^{-1}$ and then
356 remains relatively stable, suggesting a dynamic balance between oxygen consumption
357 and supply (Fig. 5). In this study area, the AOU/DIN and AOU/DIP of the total water
358 column are 6.28 and 86.79, respectively, much lower than the ideal Redfield model (i.e.,
359 AOU/DIN = 8.6, AOU/DIP = 138) (Redfield, 1960) and also lower than the reported
360 values from the Atlantic and Indian Oceans (AOU/DIN = 10.6, AOU/DIP =
361 170) (Anderson and Sarmiento, 1994). Although previous studies have shown that in
362 strongly mixed ocean areas, enhanced water aeration and shorter water mass residence
363 time can systematically suppress linear regression slopes (AOU/DIN = 3.1, AOU/DIP
364 = 57.6) (Zhu et al., 2025), our study area is located in the WPWP, where strong
365 stratification inhibits vertical mixing and exchange (Hu, D et al., 2015). Hence, physical
366 “noise” is weak, and the low ratios observed here are more likely to originate from
367 biogeochemical processes operating within the water masses.

368 To examine how these biogeochemical processes vary across water masses, we
369 calculated AOU/nutrient regression slopes separately for the upper (0–300 m), middle
370 (300–1000 m), and deep (> 1000 m) water masses (Fig. 5). The AOU/DIN value is
371 lowest in the upper water mass (6.87), increases in the middle water mass (7.21), and
372 becomes statistically insignificant in the deep water mass ($p > 0.05$) (Fig. 5a). The
373 AOU/DIP slopes follow a consistent pattern: lowest in the upper water mass (103.17),
374 increasing in the middle water mass (116.04), and non-significant in the deep water
375 mass (Fig. 5b). The low AOU/nutrients values in the upper water mass reflect a
376 pronounced surplus of nitrogen and phosphorus relative to oxygen consumption. This
377 surplus is driven by the synthesis of carbon-rich transparent exopolymer particles (TEP).
378 Due to insufficient vertical mixing, surface nutrients are not replenished after
379 consumption, resulting in persistently low concentrations (Figs. 3a-c) and strong
380 oligotrophy (Chl-*a* in the deep chlorophyll maximum: $0.19\text{--}0.31 \mu\text{g L}^{-1}$). Under this
381 nutrient stress, phytoplankton synthesize TEP with a high C:N ratio (25–35) rather than
382 investing in biomass (Yamada et al., 2017; Zamanillo et al., 2019), as evidenced by high
383 TEP concentrations in surface and subsurface waters (Fig. 3i). TEP production sustains



384 oxygen evolution while consuming little nitrogen and phosphorus, generating positive
385 $\text{PreNO}_3/\text{PrePO}_4$ anomalies (Figs. 3d, e) (Curran et al., 2025; Smyth and Letscher, 2023).
386 In addition, since microorganisms preferentially degrade low-C:N substrates (e.g.,
387 proteins), the carbon-rich TEP is remineralized slowly, causing net nutrient
388 regeneration to exceed organic carbon oxidation, thereby depressing the linear
389 regression slopes of AOU/DIN and AOU/DIP in the study area (Figs. 5a, b) (Hwang et
390 al., 2006; Tian et al., 2025). In the deep water mass, the absence of significant
391 AOU/nutrients relationships is consistent with microbial processes that decouple
392 oxygen consumption from nutrient regeneration, including chemoautotrophy mediated
393 by ammonia-oxidizing archaea such as *Nitrosopumilus* (Martens-Habbena et al., 2009)
394 and alternative phosphorus cycling via alkaline phosphatase activity of
395 *Alteromonas* (Saavedra et al., 2025).



396

397 **Figure 5. The correlation between apparent oxygen utilization (AOU) and (a) dissolved**
 398 **inorganic nitrogen (DIN), (b) phosphate ($\text{PO}_4\text{-P}$, as dissolved inorganic phosphorus,**
 399 **DIP), (c) silicate ($\text{SiO}_3\text{-Si}$, as dissolved inorganic silicon, DSi), (d) dissolved inorganic**
 400 **carbon (DIC), and (e) Relationship between DIN and DIP**

401 Unlike DIN and DIP, $\text{SiO}_3\text{-Si}$ and AOU do not exhibit a clear linear relationship
 402 in the whole water column, but instead show distinct vertical stratification. In the upper



403 waters, intense silicate uptake by phytoplankton such as diatoms drives concentrations
404 to near-zero values (Fig. 5c) (Brzezinski et al., 2024). In the middle waters, the variation
405 in silicate concentration is mainly influenced by diatom deposition and remineralization.
406 As AOU increases, silicates are rapidly released, resulting in an AOU/SiO₃-Si ratio
407 (1.47) lower in these waters than in the upper water masses (3.70) (Figure 5c). In the
408 deep waters, AOU tends to remain nearly constant, whereas SiO₃-Si concentration
409 continued to increase owing to the delayed dissolution of siliceous biogenic debris (Yu
410 et al., 2022). It is estimated that about two-thirds of biogenic silica dissolves at depths
411 shallower than 2,000 m, with the remainder released more slowly at greater
412 depths (Tréguer and De La Rocha, 2013). Consequently, SiO₃-Si levels at depth are
413 largely decoupled from organic matter supply and exhibits no significant linear
414 relationship with AOU (Figure 5c). A similar decoupling between biological
415 consumption and accumulation is observed in the distribution of DIC, which exhibits a
416 characteristic three-stage pattern from surface to deep waters (Fig. 5d). In upper waters,
417 photosynthesis consumes DIC; however, this signal is masked by intense air-sea CO₂
418 exchange, which leads to a net increase in DIC and consequently a low AOU/DIC ratio
419 of 0.37 (Ito and Reinhard, 2025). In middle waters, organic matter decomposition
420 dominates, and DIC increases rapidly with increasing AOU (AOU/DIC = 0.86, Fig.
421 5d) (Hu, X et al., 2016). In deep waters, AOU remains stable while DIC accelerates
422 again, due to the accumulation of respired CO₂ from above and enhanced calcium
423 carbonate dissolution with depth (Chen et al., 2022; Feely et al., 2004).

424 Furthermore, DIN and DIP in each water mass of the study area show a strong
425 linear correlation, and the slope decreases progressively from 17.89 in the upper waters
426 to 11.08 in the middle waters and further to 7.79 in the deep waters, with an overall
427 slope of 14, which is lower than the Redfield ratio (Fig. 5e). The overall low N:P ratio
428 (14) and its vertical decline reflect that phosphorus is regenerated more efficiently than
429 nitrogen in this region. Denitrification, which removes fixed nitrogen, can lower the
430 N:P ratio. However, classical water-column denitrification typically requires dissolved
431 oxygen concentrations below 5 μmol L⁻¹ (Deutsch et al., 2011). In the present study, the



432 minimum observed oxygen concentration exceeds $55 \mu\text{mol L}^{-1}$, making extensive
433 pelagic denitrification unlikely. Alternatively, enhanced phosphate release from
434 dissolved organic phosphorus (DOP) via alkaline phosphatase (APase) activity
435 represents one potential mechanism. Although direct APase measurements were not
436 made in this study, previous work has demonstrated that APase produced by
437 *Alteromonas* can maintain high expression and abundance under phosphorus-limited
438 conditions and even in deeper waters where inorganic phosphorus is not
439 scarce (Saavedra et al., 2025). Given that *Alteromonas* accounts for a high proportion
440 in all water masses (Fig. 4), it is plausible that this genus contributes to DOP hydrolysis
441 and thereby to the observed low N:P ratio.

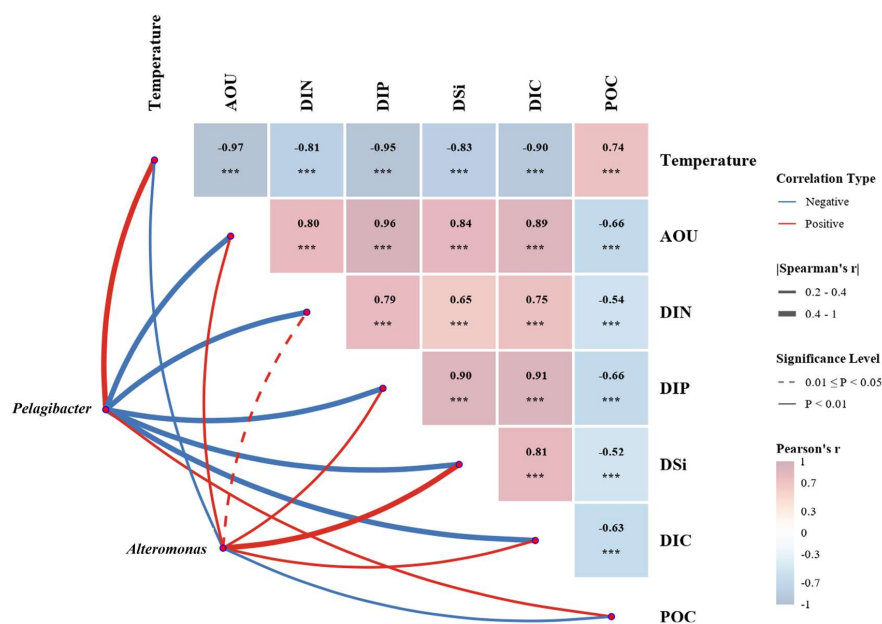
442 **4.2 Microbial regulation of non-Redfield stoichiometry in organic matter** 443 **degradation**

444 The low AOU/DIN (6.28) and AOU/DIP (86.79) ratios observed in the TNWP
445 result from the metabolic strategies of the resident microbial community. Within the
446 study area, *Pelagibacter* and *Alteromonas* emerge as the two most abundant bacterial
447 genera, their complementary metabolic capabilities directly modulate the coupling
448 between organic carbon oxidation and nutrient regeneration, thereby driving the
449 observed deviations from Redfield stoichiometry. *Pelagibacter* exhibits a significant
450 positive correlation with organic matter and a negative correlation with inorganic
451 nutrients (Fig. 6), consistent with its streamlined genome and oligotrophic adaptive
452 strategy (Carini et al., 2013; Giovannoni, 2017). This bacterium lacks the ability to
453 synthesize several essential metabolites and relies on exogenous supply, but its cell
454 surface is enriched with high-affinity transport proteins that enable efficient uptake of
455 small-molecule dissolved organic matter (Giovannoni, 2017). Critically, *Pelagibacter*
456 selectively removes nitrogen-rich organic substrates such as amino acids, proteins, and
457 glucose while leaving behind carbon-rich polymers (Malmstrom et al., 2005).

458 In contrast, *Alteromonas* displays the opposite environmental correlation pattern:
459 showing a significantly positive correlation with inorganic nutrients and a negative
460 correlation with POC (Fig. 6), reflecting its role as a particle-associated opportunist



461 with a versatile genome encoding diverse polysaccharide-degrading enzymes (Koch et
 462 al., 2019; Lopez-Perez et al., 2012). This genus is a major contributor to alkaline
 463 phosphatase gene expression (particularly *phoD* and *phoX*) and enzymatic
 464 activity (Saavedra et al., 2025). Under phosphorus-limited conditions, these enzymes
 465 hydrolyze dissolved organic phosphorus (DOP) to release bioavailable
 466 phosphate (Saavedra et al., 2025). *Alteromonas* was abundant in all water masses (Fig.
 467 4), resulting in faster phosphorus regeneration than nitrogen regeneration in the study
 468 area, leading to an observed DIN:DIP ratio of approximately 14:1, lower than the
 469 Redfield value of 16:1 (Fig. 5e). Beyond these two dominant genera, the ammonia-
 470 oxidizing archaeon *Nitrosopumilus*, which is abundant in the subsurface and mid-water
 471 (Fig. 4), can mediate high-affinity nitrification and consume oxygen even at extremely
 472 low ammonium concentrations ($< 10 \text{ nmol L}^{-1}$) (Martens-Habbena et al., 2009).



473

474 **Figure 6. Correlation between dominant microbes and environmental variables.**
 475 **Heatmaps were based on Pearson correlation analysis among environmental variables**
 476 **in the study area (***) represents $p < 0.01$), and Spearman correlation analysis between**
 477 ***Pelagibacter/Alteromonas* and environmental variables.**

478 The ecological division of labor between these functional groups creates a



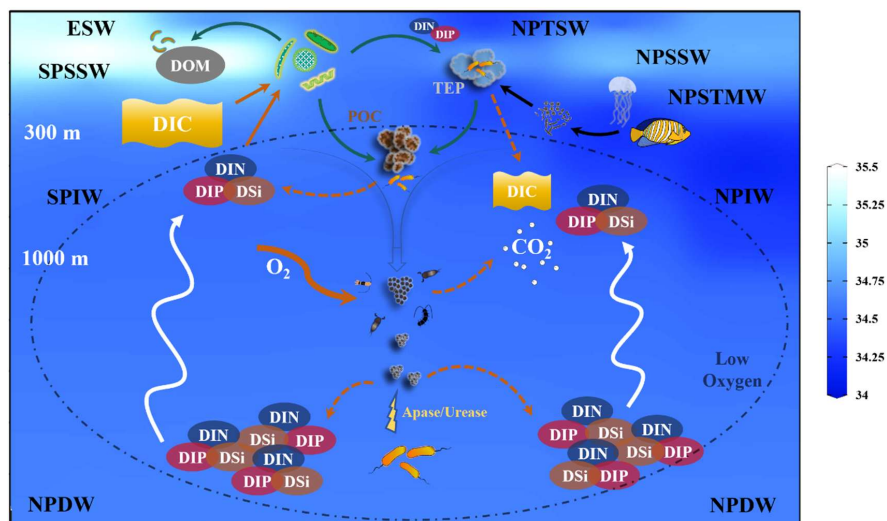
479 vertically structured processing network: *Pelagibacter* sustains the microbial loop and
480 small-molecule cycling in oligotrophic surface waters through high-affinity uptake,
481 while *Alteromonas* dominates the degradation of settling TEP and other complex
482 polymers, releasing inorganic nutrients in deeper waters (Robertson et al., 2024). This
483 functional differentiation enables the preferential degradation of low C:N substrates,
484 coupled with active phosphorus scavenging from organic pools, driving the system
485 toward enhanced nutrient regeneration relative to organic carbon oxidation, thereby
486 maintaining a dynamic equilibrium in its biogeochemical cycles and overall ecosystem
487 function.

488 **4.3 Coupled physical-biogeochemical-microbial controls on carbon and nutrient** 489 **cycling in the TNWP**

490 The strong vertical stratification of the TNWP creates a multi-water-mass structure.
491 Nutrients are primarily retained in deep reservoirs and remain chronically depleted in
492 surface waters. Nutrient regeneration in subsurface waters relies mainly on the release
493 during organic matter degradation (Hirose and Kamiya, 2003). In this nutrient-poor
494 surface environment, large amounts of TEP with a high C: N ratio accumulate (Fig. 3i).
495 The synthesis of this carbon-rich organic matter creates an imbalance between oxygen
496 production and nitrate consumption, leading to positive anomalies in the PrePO₄ and
497 PreNO₃ in surface waters (Figs. 3 d, e) (Curran et al., 2025). POM is gradually degraded
498 during settling, while intermediate and deep waters are active areas for heterotrophic
499 respiration. It is estimated that approximately 90 % of the organic matter exported to
500 this area is remineralized and releases CO₂ (Robinson et al., 2010). This process
501 consumes a large amount of DO, and the strong stratification hinders the timely
502 replenishment of DO in the intermediate and deep waters, thereby forming low oxygen
503 zones in these water masses (Fig. 7).



23



504

505 **Figure 7. Schematic diagram of coupled physical-biogeochemical-microbial controls**
 506 **on carbon and nutrient cycling in the study area with a salinity background.**

507 Simultaneously, nutrient release also mainly occurs in the intermediate and deep
 508 water masses, exhibiting a clear regeneration gradient from shallow to deep (Figs. 3a-
 509 c) (Hirose and Kamiya, 2003). The distribution patterns and functional traits of key
 510 bacteria in the water column are closely linked to their distinct ecological niches.
 511 *Pelagibacter* is abundant in oligotrophic surface waters (Fig. 4), and its streamlined
 512 genome and metabolic mechanisms enable it to drive the remineralization of small-
 513 molecule organic carbon and nutrient recycling through slow but continuous
 514 growth (Sun, J et al., 2011). *Alteromonas* relies on a diverse enzyme system to
 515 efficiently respond to organic substrates, thereby utilizing complex polymers (Koch et
 516 al., 2019). This functional differentiation within the microbial community maintains the
 517 slow cycling of dissolved organic matter under oligotrophic conditions and ensures the
 518 efficient degradation of complex polymeric organic matter when available, thereby
 519 regulating both the pathways and the intensity of carbon, nitrogen, and phosphorus
 520 cycling in the water column.

521 **5. Conclusions**

522 The oligotrophic waters of the tropical Northwest Pacific are characterized by



523 pronounced vertical stratification, with nutrients largely trapped at depth, while surface
524 waters remain chronically nutrient-poor. Phytoplankton adapt to chronic nutrient stress
525 by producing carbon-rich transparent exopolymer particles, whereas the microbial
526 community exhibits functional partitioning, in which oligotrophic *Pelagibacter*
527 recycles small organic molecules, and opportunistic *Alteromonas* degrades polymers,
528 releasing phosphorus through alkaline phosphatase activity. This selective processing
529 enhances nutrient regeneration relative to carbon oxidation, depresses DIN:DIP ratios
530 throughout the water column, and generates the observed low AOU/DIN (6.28) and
531 AOU/DIP (86.79) ratios. Collectively, our findings establish that organism-level
532 adaptations to nutrient stress can fundamentally restructure the stoichiometry of organic
533 matter degradation in oligotrophic oceans. This biologically mediated stoichiometric
534 imbalance challenges the core assumption of invariant Redfield ratios embedded in
535 current biogeochemical models, suggesting that incorporating such microbial-driven
536 flexibility in carbon-nitrogen-phosphorus coupling is essential for improving
537 predictions of carbon export efficiency and ocean carbon storage under future climate
538 scenarios.

539 **Data Availability:** The data used in this paper are available at Figshare (Tian, 2026).

540 **Competing interests:** The authors declare that they have no conflict of interest.

541 **Author contributions:** Detong Tian carried out the investigation, developed the
542 methodology, performed data curation and visualization, and prepared the original draft.
543 Xuegang Li and Yunping Xu acquired the funding, and reviewed and edited the
544 manuscript. Feng Zhao provided resources. Jun Ma, Shanshan Liu, and Muhammad
545 Inayat Ullah Khan reviewed and edited the manuscript. Weichao Wu conceptualized
546 the study, and reviewed and edited the manuscript.

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