



Potential of common extraction procedures to evaluate organic carbon association with different species of metal(loid)s in soils

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Abstract

A strong association of organic carbon (OC) with metal oxides, especially iron (Fe) and aluminium (Al) oxides has been reported in several studies. However, the role of various species of Fe, Al, manganese (Mn) and silicon (Si) in the stabilisation of OC in Fe-rich and phosphorus (P) deficient soils, common in tropical and sub-tropical regions, remains poorly understood. To address this, we collected topsoil (0-20 cm) and subsoil (20-40 cm) samples from 37 sites across agricultural regions of New South Wales, Australia. Soil samples were subjected to three separate chemical extractions, i.e., sodium pyrophosphate (PP) to extract organo-metal complexes, ammonium oxalate (OX) to extract poorly crystalline and short-range order (SRO) minerals and dithionite citrate bicarbonate (DCB) to extract total metal oxides, to determine OC associated with each of the extractions. Soil organic carbon (SOC) was extracted in the sequence: $C_{PP} > C_{DCB} > C_{OX}$, with a mean of 62 ± 0.16 %, 41 ± 0.19 % and 28 ± 0.21 %, respectively, of the total C extracted from soils. The extraction sequence for Fe was: $Fe_{DCB} > Fe_{OX} > Fe_{PP}$, with a mean of 49 ± 0.18 %, 9 ± 0.05 % and 3 ± 0.03 %, respectively, of the total Fe extracted from soils. Aluminium was extracted in the same sequence as Fe, with a mean value of 4.0 ± 0.02 % for Al_{DCB} , 3.9 ± 0.02 % for Al_{OX} and 2.0 ± 0.01 % for Al_{PP} of the total Al. Manganese extraction occurred in the sequence - $Mn_{OX} > Mn_{DCB} > Mn_{PP}$, with an average of 78 ± 1.58 %, 65 ± 1.10 % and 43 ± 1.23 %, respectively, of the total Mn extracted. All extractants dissolved < 1 % of the total Si. All extractable forms of Fe and Al showed significant positive correlations with extractable C, which suggested their potential role in the preservation of SOC. The large fraction of OC extracted by PP suggested that organic-Fe/Al complexes constituted a large fraction of the total SOC in the studied soils. The significant amount of OC associated with OX extraction also indicated that a substantial portion of the SOC was associated with SRO and poorly crystalline Fe oxides, while a relatively small proportion of the total OC was extracted by DCB suggested a limited role of crystalline Fe/Al oxides in the OC stabilisation.

1 Introduction

A huge amount (~ 2200 Pg) of organic carbon (OC) in the terrestrial ecosystems is stored in the top 100 cm soil, which is more than the combined together in the atmosphere and vegetation (Batjes, 2014). Soil OC carbon plays a key role in providing multiple ecosystem services including plant production, nutrient cycling, water storage and purification, and climate regulation (Ontl and Schulte, 2012; Paul, 2016; Jackson et al., 2017). Soil OC is an important and significant component of the global C biogeochemical cycle of C and a small change in the SOC pool can substantially increase or decrease atmospheric CO_2 concentration, thus has a



46 potential to accelerate or mitigate global warming (Lal et al., 1999; Lal, 2004; Crowther et al.,
47 2016). Managing soils to boost OC stocks has been suggested as a strategy to reduce the rise
48 in atmospheric CO₂ concentration (Mikutta et al., 2006). The association of OC with minerals
49 has received significant attention in recent decades, because it is recognised as a key
50 mechanism in the stabilisation and long-term storage of OC in soils (Sollins et al., 1996; Torn
51 et al., 1997; Lützow et al., 2006; Kögel-Knabner et al., 2008; Kleber et al., 2015; Lehmann and
52 Kleber, 2015). Several mechanisms, including surface complexation, cation bridging, weak
53 chemical interactions and co-precipitation, have been postulated in the formation of mineral-
54 associated organic matter (MAOM) in soils (Kögel-Knabner et al., 2008; von Lützow et al.,
55 2008; Kleber et al., 2015). Chemical interactions of OC with minerals decrease its susceptibility
56 towards oxidative attack and microbial mineralisation, which subsequently results in the
57 stabilisation and accumulation of OC in soils (Kaiser and Guggenberger, 2003; Kalbitz et al.,
58 2005; Kögel-Knabner et al., 2008). The nature of association and stability of the MAOM in
59 different soil types are still not well understood.

60 Iron (Fe) and aluminium (Al) are among the most prevalent elements in soils (Cornell and
61 Schwertmann, 2003). Iron and Al oxides (the term 'oxides' used for brevity, it includes
62 hydroxides, oxyhydroxides and oxides) are formed in soils from the weathering and dissolution
63 of primary minerals (Chadwick and Chorover, 2001). These minerals are effective sorbents for
64 dissolved OC because of their high specific surface area (SSA) and abundant reactive surfaces
65 (Kaiser and Guggenberger, 2000). Iron bound C has been estimated to contribute 15 to 78 %
66 of the total OC in sediments and soils (Cornell and Schwertmann, 2003; Wagai and Mayer,
67 2007; Zhao et al., 2016; Kramer and Chadwick, 2018). The concentration of Fe and Al oxides
68 is generally high in intensively weathered soils of tropical and subtropical regions (Singh and
69 Gilkes, 1992; Mikutta et al., 2009). The surface hydroxyl groups of Fe oxides are capable of
70 interacting with carboxyl or hydroxamate groups of SOC via ligand exchange, thus stabilise and
71 protect SOC against microbial mineralisation (Gu et al., 1994; Mikutta et al., 2007; Kleber et
72 al., 2015; Chen et al., 2020). Chemical reactions between Fe oxides and organic functional
73 groups have been well studied and it has been found that Fe oxides preferentially bind with
74 aromatic and carboxylic functional groups of OC (Gu et al., 1994; Kaiser, 2003; Chen et al.,
75 2014; Yeasmin et al., 2014). Co-precipitation of dissolved OC with polyvalent cations, such as
76 Fe and Al, in organo-metal complexes also stabilises SOC (Baldock and Skjemstad, 2000; Scheel
77 et al., 2007; Eusterhues et al., 2011; Mikutta et al., 2014). Co-precipitation may occur due to
78 changes in soil pH or redox conditions (Wagai and Mayer, 2007; Fritzsche et al., 2015). For
79 example, in temporarily waterlogged paddy soils, upon aeration, dissolved Fe (II) is rapidly
80 oxidised to relatively insoluble Fe (III) and large amount of OC co-precipitate with Fe (Chen et
81 al., 2014; Song et al., 2022). Some studies have reported that co-precipitation (organo-metal
82 complexes) was more effective in sequestering dissolved OC than adsorption (Mikutta et al.,
83 2011; Chen et al., 2014; Mikutta et al., 2014). Despite the wide-spread occurrence of Fe oxides
84 in Australian soils, the association of OC with Fe oxides remains poorly understood in these
85 soils (Davey et al., 1975; Singh and Gilkes, 1992; Viscarra Rossel et al., 2010).

86 Manganese (Mn) is a redox-active element that can strongly interact with organic functional
87 groups, thus influencing the stabilisation of SOC (Li et al., 2021). Relative to its bulk soil
88 concentration, Mn oxides could be strongly enriched in OC (Rennert et al., 2014). The poorly
89 crystalline Mn oxides in particular possess relatively large SSA and high adsorption capacity
90 for organics (Estes et al., 2017; Stuckey et al., 2018; Li et al., 2021), thereby stabilising and
91 protecting OC against decomposition (Li et al., 2021). Mn oxides have several properties
92 similar to Fe oxides (Li et al., 2021), except the point of zero charge, but the association of OC



93 with different forms of Mn in natural soils is rarely studied. Silicon (Si) plays a significant role
94 in regulating the biogeochemical C cycle and stabilization of SOC. In the terrestrial ecosystem,
95 the behaviour of Fe and Al may be regulated by biogeochemical cycling of Si through the
96 formation of stable Fe-Si and Al-Si complexes (Pokrovski et al., 2003; Matus et al., 2014).
97 Consequently, Si may play a significant role in stabilising SOC (Song et al., 2018), either directly
98 by stabilisation of SOC by chemical interaction with amorphous Si (phytolith) or indirectly by
99 MAOM stabilisation and soil aggregation (Pokrovski et al., 2003; Matus et al., 2014; Zhang et
100 al., 2016).

101 The abundance of various forms of Fe, Al, Mn, and Si and reactive mineral phases in soils is
102 frequently estimated by chemical extraction procedures. Extraction procedures using sodium
103 pyrophosphate (PP), ammonium oxalate (OX), and dithionite-citrate-bicarbonate (DCB)
104 reagents target certain metal phases (Mehra and Jackson, 1960; McKeague, 1967). Extraction
105 methods are feasible, especially with large number of samples and can be easily adopted by
106 commercial laboratories. In addition, these procedures are less-time consuming economical
107 than spectroscopic and other methods (Rennert, 2019) and do not require specialised
108 equipment. However, there are some potential pitfalls in the extraction methods, such as
109 exclusive selectivity, interpretation of extracted phases and extraction efficiency, which may
110 vary with soil properties (Parfitt and Childs, 1988; Kaiser and Zech, 1996; Pansu and
111 Gautheyrou, 2006; Rennert, 2019; Voelz et al., 2019; Rennert and Lenhardt, 2025). Despite
112 these drawbacks, extraction procedures offer the best option for the broad scale assessment
113 of OC stabilisation with different forms of Fe, Al, Mn, and Si in soils. Other techniques, such
114 as, X-ray diffraction and vibrational spectroscopy, are unable to distinguish the association of
115 OC with different species of Fe and Al oxides (Coward et al., 2017). Numerous studies have
116 used extraction procedure data to understand the stabilisation of SOC with Fe and Al oxides
117 in relatively young soils from temperate soils (Kaiser and Guggenberger, 2000; Kleber et al.,
118 2005; Wiseman and Püttmann, 2006; Kaiser et al., 2007; Wagai and Mayer, 2007; Rasmussen
119 et al., 2018; Ashida et al., 2021), and in highly weathered soils (Coward et al., 2017; Wagai et
120 al., 2020; Ye et al., 2022). Significant positive relationships between SOC and OX extractable
121 Fe and Al highlight the important role of poorly crystalline Fe/Al oxides and/or short range
122 order (SRO) Fe/Al oxides in the formation of MAOM in both surface (Rasmussen et al., 2006)
123 and sub-surface soils (Kaiser and Guggenberger, 2000; Kleber et al., 2005; von Fromm et al.,
124 2025). Likewise, crystalline Fe oxides content has been reported to be the major driver of
125 MAOM abundance in soils and close relationships between DCB extractable Fe/Al and SOC
126 content have been reported in soils from different regions (Mu et al., 2016; Zhao et al., 2016;
127 Yu et al., 2019; Zhao et al., 2023). Many other studies have reported a significant relationship
128 between PP extractable Fe/Al and SOC content in a wide range of soils, suggesting the
129 relevance of organo-metal complexes in the stabilisation of SOC (Percival et al., 2000; Wagai
130 and Mayer, 2007; Wagai et al., 2013; Lawrence et al., 2015; Hall and Thompson, 2022; Yu et
131 al., 2023). Most of the published studies have evaluated relationships between different forms
132 of extractable metals and SOC contents but did not actually quantify the OC content
133 associated with different forms of extractable metal(loid)s. Wagai and Mayer (2007)
134 determined OC associated with Fe oxides using dithionite-HCl extraction in several soils and
135 estimated between 2 % and 25% of the total SOM was bound with Fe oxides. Heckmann et al
136 (2018) used a selective and sequential procedure to evaluate OC association with different
137 forms of Al and Fe in four soils and found the order: Na-pyrophosphate < hydroxylamine <
138 dithionite-HCl. Thus, the largest pool of stabilized OC was found in organo-metal complexes
139 that was extracted by Na-pyrophosphate. In this study, we have attempted to determine the



140 role of different species of Fe, Al, Mn and Si in the stabilisation of OC of P deficient and Fe-rich
141 soils that occupy much of the tropic and subtropics and southern hemisphere including
142 Australia. We have employed the most common chemical extraction procedures,
143 acknowledging and accounting for possible errors involved in these, to evaluate the role of
144 different metal(loid)s in OC stabilisation.

145 We used PP (pH 10), OX (pH 3) and DCB to extract organo-metal complexes, SRO and/or poorly
146 crystalline metal phases and crystalline or free metal oxides, respectively, from a range of soil
147 types with a wide range of Fe concentrations. Also, we used both top and sub-soils for this
148 study. Soil OC associated with each of the extractable form was determined by analysing the
149 residues after each extraction. The main aim was to quantitatively estimate the role of
150 different fractions (or species) of metal(loid)s in stabilising OC in top- and sub-soils of
151 agricultural regions of New South Wales, Australia. We hypothesised that poorly crystalline
152 and SRO minerals of Fe, Al, Mn and Si contribute more to the stabilisation of OC in both top-
153 and sub-soils. We further hypothesised that the concentration of Fe will be highest in soils and
154 thus Fe will have the most prominent role in OC stabilisation amongst the metal(loid)s in both
155 top- and sub-soils.

156

157 2. Materials and methods

158

159 2.1 Soil samples and handling

160 Topsoil (0 – 20 cm) and subsoil (20 – 40 cm) samples were collected from 37 locations across
161 New South Wales, Australia (Fig. 1). The location names, coordinates, and Australian Soil
162 Classification (ASC) of the soil samples are presented in Table S1. Soil samples were air-dried,
163 crushed, passed through a 2 mm sieve and stored at room temperature for laboratory analysis.
164 The dataset of the extractable metal(loid)s and C used for this study is presented in Table A1
165 and A2, respectively.

166

167 2.2 Soil physical and chemical properties

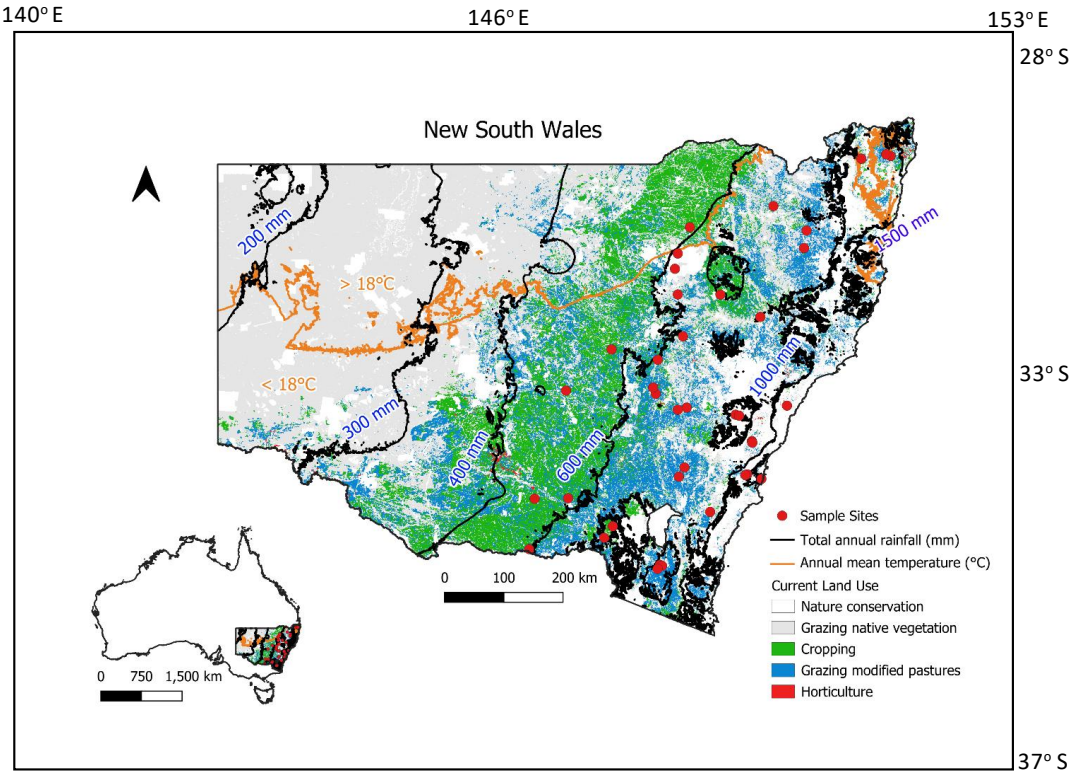
168 Soil pH and electrical conductivity (EC) were measured in 1:5 of soil and water extracts using
169 a combined glass electrode pH meter and an EC meter, respectively (Rayment and Lyons,
170 2011). Particle size analysis (PSA) was determined by the hydrometer method (Gee and Or,
171 2002). Cation exchange capacity (CEC) was determined by the silver thiourea method
172 (Rayment and Lyons, 2011). Total C and nitrogen (TN) were analysed using vario MACRO cube
173 Elementar CHN analyser. Total Fe, Al, Mn, and Si were analysed using a PANalytical Minipal 4
174 X-Ray Florescence (XRF) spectrometer.

175

176



177



178 **Figure 1. Map of New South Wales showing sampling (red dots) locations of soil samples,**
179 **current land use, total annual rainfall and annual mean temperature.**

180

181 **2.3 Selective extraction procedure and analyses**

182 Soil samples were subjected to three separate chemical extractions to dissolve different Fe
183 phases (Table 1). Aluminium, Mn and Si were also measured in these extractions, however,
184 the extractants are known to be most specific to Fe phases. All soil samples (in duplicates)
185 were separately extracted by sodium pyrophosphate (PP, pH 10), ammonium oxalate (OX, pH
186 3), and dithionite-citrate-bicarbonate (DCB). To validate the quantification of extracted C, a
187 subset of 10 samples (with a wide range in Fe content) were also extracted by sodium
188 pyrophosphate (PP, pH 7.5), dithionite-hydrochloric acid (DH) and hydrochloric acid-
189 hydroxylamine (HH), as described in more detail in Section 2.4. The concentrations of Fe, Al,
190 Si and Mn in supernatants from all extractions were analysed using PerkinElmer Nexion 300x
191 inductively coupled plasma-mass spectrometer (ICP-MS). The residues after each extraction
192 were analysed for total C to determine OC extracted by each reagent.

193
194



195 **Table 1. A summary of chemical extraction procedures and targeted phases.**

Solvent	Concentration	Solvent pH	Target phases	Number of soil samples
Sodium pyrophosphate (PP)	0.1 M	10	Organic matter- complexed metals	74
Sodium pyrophosphate (PP)	0.1 M	7.5	Organic matter- complexed metals	10
Ammonium oxalate (OX)	0.2 M	3	Non-crystalline, amorphous, short-range-ordered (SRO) minerals and poorly crystalline minerals	74
Dithionite-citrate-bicarbonate (DCB)	0.3M, 1 M	8	Total 'free' oxides or pedogenic or crystalline metal oxides	74
Dithionite-Hydrochloric acid (DH)	0.05 M, 0.05 M	2.3	Total 'free' oxides or pedogenic phases or crystalline metal oxides	10
Hydrochloric acid-hydroxylamine (HH)	0.25 M 0.25 M	0.8	Non-crystalline, amorphous, short-range-ordered (SRO) minerals and poorly crystalline minerals	10

196

197

198 2.3.1 Sodium pyrophosphate (PP) at pH 10

199 Pyrophosphate extraction provides an estimate organic complexed or bound Fe and Al in soils
 200 (Loveland and Digby, 1984). We used the procedure described by McKeague (1967) to extract
 201 organic C complexed with Fe and Al, where 500 mg of soil was mixed with 50 ml of 0.1 M PP
 202 solution (pH 10) in 55 ml centrifuge tubes. The samples were shaken for 16 hours and
 203 centrifuged at $3280 \times g$ and the supernatants filtered through a $0.2 \mu\text{m}$ PES membrane syringe
 204 filter (Merck Millipore, Millex-GP, 33 mm diameter) for ICP-MS analysis.

205

206 2.3.2 Acid ammonium oxalate (OX)

207 Extraction with 0.2 M ammonium oxalate (pH 3.0) in the dark dissolves SRO and poorly
 208 crystalline minerals (Schwertmann, 1964). For this extraction, 500 mg of soil was mixed with
 209 30 ml of 0.2 M ammonium oxalate (pH 3.0) in 55 ml centrifuge tubes, and the mixture was
 210 shaken for 4 hours in the dark (McKeague and Day, 1966). The suspensions were then
 211 centrifuged at $3280 \times g$ for 10 minutes. The supernatants were filtered through a $0.2 \mu\text{m}$ PES
 212 membrane syringe filter for ICP-MS analysis.

213

214 2.3.3 Dithionite-citrate-bicarbonate (DCB)

215 Dithionite-citrate-bicarbonate extraction was used to extract pedogenic Fe and Al phases in
 216 soils (Mehra and Jackson, 1960). Briefly, 1 g of soil was mixed with 40 ml of 0.3 M sodium (Na)-
 217 citrate and 5 ml of 1 M Na-bicarbonate, in 55 ml centrifuge tubes. The tubes were heated in a
 218 water bath to 80°C for 10 min, then 1 g of sodium dithionite was added and stirred
 219 immediately for 1 min and then occasionally 5 min. After the digestion, 10 mL of saturated



sodium chloride (NaCl) solution was added to promote flocculation. The samples were centrifuged at $3280 \times g$, and the supernatants were filtered through Whatman 42 filter paper for the analysis. The extraction procedure was repeated twice for some of the soil samples, where brown or red colouration was observed.

2.4 Extractable C methodology validation

Two of the extraction procedures involve C containing reagents (i.e., oxalate in OX, and citrate and bicarbonate in DCB), hence we compared the residual total C in soils after OX and DCB extractions with comparable procedures, i.e. HH and DH, respectively, which do not employ C compounds in the extraction procedure (Heckman et al., 2018).

2.4.1 Sodium pyrophosphate (PP) at pH 7.5

Selective samples (with a range of total Fe content) were extracted with PP at pH 7.5 in duplicate to determine the peptization of Fe oxides at pH 10 (Parfitt and Childs, 1988; Rennert, 2019). The procedure of McKeague (1967) as described earlier was followed, except the PP solution pH was adjusted to 7.5 with hydrochloric acid (HCl). Selective samples containing a range of PP extractable Fe were used in the extraction.

2.4.2 Hydrochloric acid-hydroxylamine (HH)

Inorganic HH extraction targets SRO and/or amorphous Fe and Al oxides (Chao and Zhou, 1983; Ross et al., 1985). Acidified hydroxylamine is capable of extracting Fe phases (Chao and Zhou, 1983) that closely align with SRO minerals extraction procedure using acid ammonium oxalate (McKeague and Day, 1966). Acidified hydroxylamine is a C-free alternative to OX and the residual soil C analysis after this extraction was used to compare with the OX extractant soil residues. For this extraction, we weighed 500 mg soil into a 55 mL centrifuge tube, mixed with 30 ml of HH solution, the mixture was shaken for 16 hours and then centrifuged at $3280 \times g$. The supernatant was filtered through a $0.2 \mu\text{m}$ PES membrane syringe filter for ICP-MS analysis.

2.4.3 Dithionite-hydrochloric acid (DH)

Dithionite-hydrochloric acid is a C-free alternative to DCB to extract pedogenic Fe and Al phases just like DCB (Wagai and Mayer, 2007; Wagai et al., 2013). This extraction was performed on selective samples to compare the DH extractable C values to DCB extractable C values. 500 mg of soil was mixed with 30 ml of 57.4 mM sodium dithionite, shaken for 16 hours and centrifuged at $3280 \times g$ for 40 mins. The supernatants were filtered through a $0.2 \mu\text{m}$ PES membrane syringe filter into another 50 ml centrifuge tubes. Soil residues were added with 10 ml of 0.05 M HCl, shaken for 1 hour and then centrifuged again. The supernatants were filtered as before and combined with Na dithionite extracts.

2.5 Post-extraction procedure and analysis

The soil residues after extractions were washed twice with ultrapure water to remove entrained solutions, centrifuged, dried overnight at 60°C , lightly ground manually and weighed for total C analysis. Total C concentrations of the soil samples before and after extraction were analysed in duplicates using a vario MACRO cube Elementar CHN analyser. Total C concentration represents OC concentration because most of studied soils had acidic



pH values. The presence of carbonate was also tested by adding hydrochloric acid to soil samples with near neutral pH values.

2.6 Data Analysis

Paired t-test was used to determine differences in the concentrations of extractable metal(loid)s (Fe, Al, Mn and Si) and extractable C between samples from the two depths. Correlations between physical and chemical properties of soils were performed using Spearman's rho correlation. The generalised linear mixed model (GLMM) regression was used to determine relationships between extractable metal concentrations and other soil properties. In GLMM regression, we used extractable metal concentrations and soil properties as the fixed effects and soil types as the random effects. After running the model with all soil properties first, the non-significant variables were removed from the fixed effects, and the model was re-run to obtain variables that were significant in the model. For data analysis, we used JMP Pro software (Version 17.0, SAS Institute Incorporation, Cary, NC, USA).

3 Results

3.1. Soil physical and chemical characteristics

A summary of soil physical and chemical properties is presented in Table 2. Soil pH was strongly acidic to near neutral, with values ranging from 4.7 to 7.6. The electrical conductivity (EC) values of soils were small (i.e. non-saline) for all samples ranging from 1.33 mS m⁻¹ to 42.6 mS m⁻¹ for the topsoils and 1.24 mS m⁻¹ to 13.7 mS m⁻¹ for the subsoils. Cation exchange capacity (CEC) of soils was mostly small, ranging between 69.8 mmol_c kg⁻¹ and 99.7 mmol_c kg⁻¹ for the topsoils and between 47.4 mmol_c kg⁻¹ and 99.7 mmol_c kg⁻¹ for the subsoils. The clay content in soils increased with depth and it ranged from 3 to 54 % in the topsoils and 6 to 59 % in the subsoils. Total Al increased significantly with depth, while total Fe, Mn and Si concentration remained similar.

The concentrations of TC and TN were significantly ($p < 0.05$) greater in the topsoils (TC: 25.0 ± 2.3 g kg⁻¹ and TN: 1.2 ± 0.2 g kg⁻¹) than the subsoils (TC: 17.5 ± 1.6 and TN: 0.7 ± 0.1 g kg⁻¹) as shown in Fig. S1. A linear relationship was observed between TC and TN in the topsoils ($R^2 = 0.60$, $p < 0.001$), subsoils ($R^2 = 0.78$, $p < 0.001$), and all samples combined (topsoils and subsoils) ($R^2 = 0.75$, $p < 0.001$) (Fig. S2). The C:N ratio significantly ($p < 0.05$) increased with soil depth, from 23.3 ± 1.3 in the topsoils to 28.7 ± 1.8 in the subsoils.

The correlation coefficients between soil physical and chemical properties are shown in Table 3. The CEC showed a significant positive correlation with the clay content ($r = 0.70$, $p < 0.01$) and a significant negative correlation with sand content ($r = -0.64$, $p < 0.01$). Soil pH showed significant positive correlations with CEC, exchangeable Ca, exchangeable Mg, exchangeable Na, and clay content. Soil EC had significant positive relationships with TC, TN, exchangeable Ca, exchangeable K, and CEC.

Total Fe and Al showed significant positive correlations with clay and CEC. Total Mn had significant positive relationships with pH, EC, TC, TN, exchangeable Ca, exchangeable K, CEC, and total Fe. Total Si showed a significant positive relationship with sand content ($r = 0.71$, $p < 0.01$).



Table 2. Summary of important chemical and physical properties of the studied soils (n = 74)

Properties	Topsoil			Subsoil		
	Range	Mean $\pm$ SE	Median	Range	Mean	Median
pH (1:5 water)	4.70 – 7.38	5.92 $\pm$ 0.11a	5.85	4.87 – 7.60	6.15 $\pm$ 0.13a	6.14
EC _{1:5} (mS m ⁻¹)	1.33 – 42.6	6.73 $\pm$ 1.13a	4.99	1.24 – 13.7	4.18 $\pm$ 0.44b	3.43
CEC (mmol _c kg ⁻¹)	69.8 – 99.7	94.0 $\pm$ 1.3a	97.9	47.4 – 99.7	92.5 $\pm$ 2.1a	98.3
Exchangeable Ca (mmol _c kg ⁻¹)	1.3 – 37.9	18.0 $\pm$ 1.9a	20.9	0.6 – 39.8	15.7 $\pm$ 1.9a	13.4
Exchangeable Mg (mmol _c kg ⁻¹)	2.3 – 34.7	16.0 $\pm$ 1.4a	16.4	1.8 – 49.9	18.3 $\pm$ 1.8a	15.4
Exchangeable Na (mmol _c kg ⁻¹)	0.1 – 13.4	1.5 $\pm$ 0.4a	1	0.1 – 23	3.8 $\pm$ 0.9b	1.3
Exchangeable K (mmol _c kg ⁻¹)	0.4 – 7.6	2.5 $\pm$ 0.3a	2.2	0.1 – 6.1	1.5 $\pm$ 0.2b	1.1
Sand (%)	36 – 95	61.7 $\pm$ 2.9a	64	26 – 90	56.9 $\pm$ 2.9a	59
Silt (%)	2 – 30	13.2 $\pm$ 1.0a	14	2 – 30	11.9 $\pm$ 1.1a	11
Clay (%)	3 – 54	25.1 $\pm$ 2.6a	20	6 – 59	31.2 $\pm$ 2.7a	30
Total Fe (g kg ⁻¹)	9.34 – 161.97	50.12 $\pm$ 5.91a	35.12	11.26 – 136.96	53.29 $\pm$ 5.17a	40.83
Total Al (g kg ⁻¹)	18.08 – 125.79	76.54 $\pm$ 3.99a	77.72	40.18 – 132.91	87.31 $\pm$ 3.34b	88.16
Total Mn (g kg ⁻¹)	0 – 3.1	0.78 $\pm$ 0.13a	0.5	0 – 3.6	0.60 $\pm$ 0.13a	0.3
Total Si (g kg ⁻¹)	148.54 – 384.73	251.42 $\pm$ 9.35a	258.32	153.39 – 356.31	237.33 $\pm$ 8.26a	235.71

Similar letters in the same row indicate non-significant differences between mean values ($P < 0.05$) while different letters indicate significant differences between mean values ($P < 0.05$) of soil properties.

313

Table 3: Spearman's rho correlation coefficients for relationships between soil physical and chemical properties.

	pH	EC	TC	TN	Ca	Mg	Na	K
EC	0.25*							
TC	-0.06	0.40**						
TN	<0.1	0.58**	0.84**					
Ca	0.58**	0.51**	0.18	0.39**				
Mg	0.38**	0.13	-0.04	-0.01	0.14			
Na	0.24*	-0.03	-0.06	-0.11	-0.08	0.56**		
K	0.05	0.61**	0.36**	0.54**	0.47**	-0.02	-0.13	
CEC	0.57**	0.23*	0.02	0.11	0.51**	0.73**	0.44**	0.04
Sand	-0.22	0.05	0.01	-0.05	-0.22	-0.55**	-0.27*	0.22
Silt	<-0.1	0.22	0.24*	0.29*	0.20	0.13	0.14	0.23*
Clay	0.27*	-0.08	-0.06	<-0.1	0.25*	0.60**	0.30**	-0.24*
Total Fe	0.32**	-0.03	0.07	0.12	0.29*	0.56**	0.18	-0.28*
Total Al	0.14	-0.31**	-0.06	-0.14	-0.04	0.47**	0.19	-0.40**
Total Mn	0.49**	0.33**	0.27*	0.53**	0.66**	0.08	-0.16	0.34**
Total Si	-0.25*	0.07	-0.14	-0.11	-0.22	-0.48**	-0.14	0.30**

**Correlation significant at $P < 0.01$.

*Correlation significant at $P < 0.05$.

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319



Table 3 continued:

	CEC	Sand	Silt	Clay	Total Fe	Total Al	Total Mn
Sand	-0.64**						
Silt	0.22	-0.50**					
Clay	0.70**	-0.94**	0.22				
Total Fe	0.69**	-0.76**	0.16	0.81**			
Total Al	0.42**	-0.73**	0.03	0.81**	0.75**		
Total Mn	0.30**	-0.12	0.23	0.08	0.25*	-0.13	
Total Si	-0.60**	0.71**	-0.07	-0.78**	-0.89**	-0.84**	-0.14

**Correlation significant at $P < 0.01$.

*Correlation significant at $P < 0.05$.

3.2. Extractable Fe, Al, Mn and Si concentrations

The concentration of Fe extracted for the three extractants was in the sequence – $Fe_{DCB} > Fe_{OX} > Fe_{PP}$ (Fig. 2a). The Fe_{PP} concentration in all soil samples ranged from 0.01 g kg^{-1} to 5.80 g kg^{-1} , with a significantly ($p < 0.001$) greater mean concentration in the topsoils ($1.10 \pm 0.11 \text{ g kg}^{-1}$) than the subsoils ($0.85 \pm 0.12 \text{ g kg}^{-1}$). The Fe_{OX} concentration varied between 0.68 g kg^{-1} and 12.81 g kg^{-1} , and the topsoils ($4.15 \pm 0.33 \text{ g kg}^{-1}$) and subsoils ($3.99 \pm 0.33 \text{ g kg}^{-1}$) having a similar mean concentration. Fe_{DCB} concentration in soil samples ranged from 2.92 g kg^{-1} to 79.87 g kg^{-1} , and like Fe_{OX} , the mean Fe_{DCB} concentration was similar in the topsoils ($23.46 \pm 1.87 \text{ g kg}^{-1}$) and the subsoils ($25.76 \pm 1.94 \text{ g kg}^{-1}$) (Fig. 2a). On an average, PP, OX and DCB extracted $3 \pm 0.03 \%$, $9 \pm 0.05 \%$ and $49 \pm 0.18 \%$ of the total Fe, respectively, across both soil depths.

The concentration of Al dissolved by the three extractants was in the order: $Al_{DCB} > Al_{OX} > Al_{PP}$ (Fig. 2b). The Al_{PP} concentration ranged between 0.27 g kg^{-1} and 6.47 g kg^{-1} , with similar average concentrations in the topsoils ($1.41 \pm 0.11 \text{ g kg}^{-1}$) and the subsoils ($1.39 \pm 0.13 \text{ g kg}^{-1}$). Al_{OX} concentration varied from 0.34 g kg^{-1} to 8.0 g kg^{-1} , with a significantly ($p = 0.003$) greater concentration in the subsoils ($3.31 \pm 0.21 \text{ g kg}^{-1}$) than the topsoils ($3.00 \pm 0.20 \text{ g kg}^{-1}$). The Al_{DCB} concentration in soils ranged from 0.28 g kg^{-1} to 10.26 g kg^{-1} , with a mean of $2.94 \pm 0.2 \text{ g kg}^{-1}$ in the topsoils and $3.50 \pm 0.25 \text{ g kg}^{-1}$ in the subsoils. The subsoils had a significantly ($p = 0.008$) greater mean concentration of Al_{DCB} than the topsoils. PP, OX and DCB extracted Al constituted on an average $2 \pm 0.01 \%$, $3.9 \pm 0.02 \%$, and $4 \pm 0.02 \%$, respectively, of the total Al in all samples. Unlike Fe and Al, more Mn was extracted by OX than DCB, with the order being – $Mn_{OX} > Mn_{DCB} > Mn_{PP}$ (Fig. 2c). The concentration of Mn in all three extractions was generally very low, with values $< 1.1 \text{ g kg}^{-1}$ in all samples. The Mn_{PP} concentration ranged from $< 0.01 \text{ g kg}^{-1}$ to 1.07 g kg^{-1} and the topsoils ($0.25 \pm 0.03 \text{ g kg}^{-1}$) having a significantly ($p < 0.001$) greater concentration than the subsoils ($0.11 \pm 0.01 \text{ g kg}^{-1}$). The concentrations of Mn_{OX} in soil samples varied from 0.001 to 2.93 g kg^{-1} and it was also significantly ($p < 0.001$) greater in the topsoils (mean = $0.60 \pm 0.08 \text{ g kg}^{-1}$) than the mean concentration in the subsoils ($0.45 \pm 0.08 \text{ g kg}^{-1}$). DCB extracted Mn concentration ranged from 0.01 g kg^{-1} to 3.62 g kg^{-1} , and like Mn_{PP} and Mn_{OX} , the topsoils ($0.47 \pm 0.07 \text{ g kg}^{-1}$) had a significantly ($p = 0.04$) greater mean concentration of Mn_{DCB} than the subsoils ($0.36 \pm 0.07 \text{ g kg}^{-1}$). PP, OX and DCB extracted on an average $43 \pm 1.23 \%$, $78 \pm 1.58 \%$ and $65 \pm 1.10 \%$ of the total Mn, respectively, across both soil depths.

The concentration of Si dissolved by the three extractants occurred in the sequence – $Si_{DCB} > Si_{OX} > Si_{PP}$ (Fig. 2d). The concentration of Si_{PP} ranged from 0.19 g kg^{-1} to 2.45 g kg^{-1} , with similar concentrations in the topsoils ($0.69 \pm 0.03 \text{ g kg}^{-1}$) and subsoils ($0.66 \pm 0.05 \text{ g kg}^{-1}$). Si_{OX}



concentrations varied between 0.05 g kg⁻¹ and 3.03 g kg⁻¹, with significantly ($p=0.006$) greater concentration in the subsoils (0.81 ± 0.07 g kg⁻¹) than the topsoils (0.70 ± 0.06 g kg⁻¹). The concentration of Si_{DCB} ranged from 0.28 g kg⁻¹ to 4.35 g kg⁻¹, and the subsoils (1.76 ± 0.11 g kg⁻¹) had a significantly ($p=0.07$) greater mean concentration of Si_{DCB} than the topsoils (1.52 ± 0.08 g kg⁻¹). PP extracted 0.3 ± 0.001 % (mean) of the total Si, OX extracted 0.4 ± 0.003 % of the total Si, and DCB extracted 0.7 ± 0.01 % of the total Si across both depths.

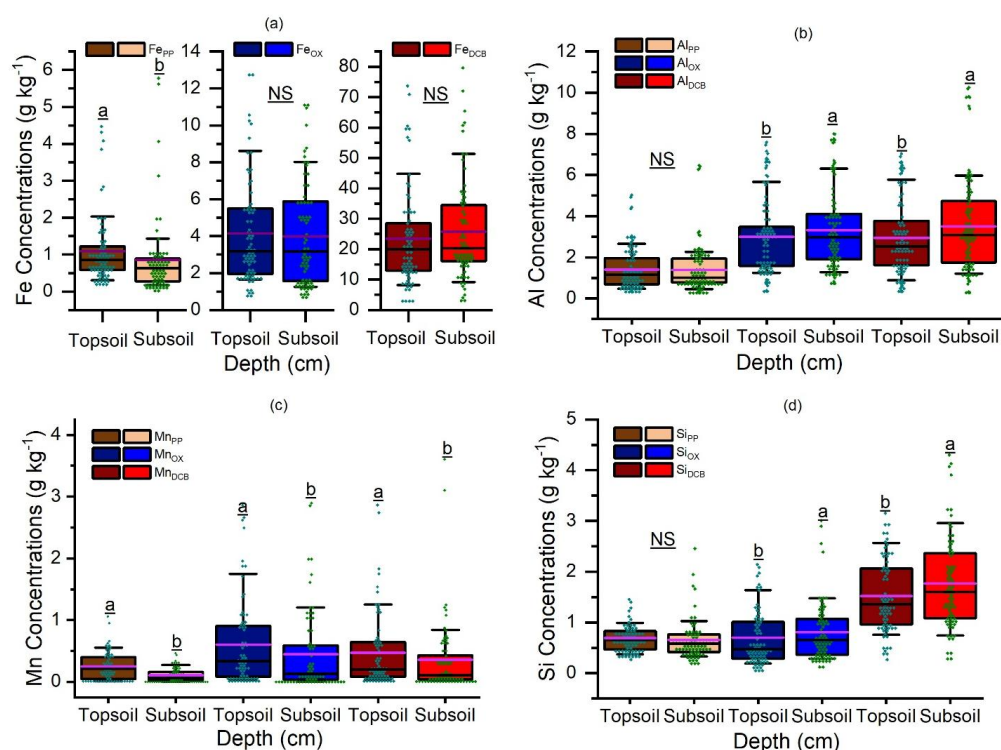


Figure 2. Concentrations of iron (Fe), aluminium (Al), manganese (Mn) and silicon (Si) in topsoils and subsoils extracted by sodium pyrophosphate (PP), ammonium oxalate (OX), and dithionite-citrate-bicarbonate (DCB) procedures. Boxes comprise of the 25th and 75th percentiles, whiskers extend from the 10th to 90th percentiles. Purple and black bars signify the mean and the median values, respectively. Dark cyan and green diamonds indicate data for individual samples; the symbols are displaced horizontally to avoid overlaps. Different letters indicate significant differences in the mean values between samples from different soil depths based on paired T-test ($p < 0.05$). NS means non-significant difference.



3.3. Relationship between extractable metal(loid)s

The Pearson correlation coefficients between extractable metals are shown in Table 4. Fe_{OX} showed a significant positive correlation with all other extractable forms of Fe, Al, Si, and Mn. Fe_{DCB} was positively and significantly correlated with Al_{OX}, Al_{DCB}, Mn_{OX}, Mn_{DCB}, Mn_{PP}, Si_{OX}, Si_{DCB}, and negatively and significantly correlated with Si_{PP}. Fe_{PP} showed significant positive correlations with Al_{OX} and Al_{PP}.

Al_{OX} had significant positive relationships with Al_{DCB}, Al_{PP}, and Si_{OX}. Al_{DCB} showed significant positive relationships with Al_{PP}, Si_{OX}, Si_{DCB} and a significant negative relationship with Si_{PP}. Al_{PP} showed a significant negative correlation with Mn_{OX}, Mn_{DCB}, and Si_{DCB}. Mn_{OX} had a significant positive correlation with Mn_{DCB}, Mn_{PP}, Si_{OX}, and Si_{DCB}, while Mn_{DCB} had a significant positive correlation with Mn_{PP}. Si_{OX} and Si_{PP} showed a significant positive relationship with Si_{DCB}.

Table 4: Pearson correlation coefficients for relationships between different extractable metal(loid)s in soils

	Fe _{OX}	Fe _{DCB}	Fe _{PP}	Al _{OX}	Al _{DCB}	Al _{PP}	Mn _{OX}	Mn _{DCB}	Mn _{PP}	Si _{OX}	Si _{DCB}
Fe _{DCB}	0.61**										
Fe _{PP}	0.41**	0.04									
Al _{OX}	0.64**	0.50**	0.24*								
Al _{DCB}	0.48**	0.83**	0.16	0.69**							
Al _{PP}	0.26*	0.06	0.64**	0.65**	0.40**						
Mn _{OX}	0.46**	0.39**	0.01	-0.07	0.05	-0.30**					
Mn _{DCB}	0.42**	0.34**	0.06	-0.10	0.05	-0.27*	0.93**				
Mn _{PP}	0.36**	0.29*	0.22	-0.17	-0.05	-0.23	0.77*	0.63**			
Si _{OX}	0.71**	0.44**	-0.10	0.51**	0.30**	-0.11	0.29*	0.19	0.09		
Si _{DCB}	0.35**	0.45**	-0.21	0.20	0.26*	-0.23*	0.25*	0.21	0.05	0.55**	
Si _{PP}	<0.01	-0.24*	0.12	-0.17	-0.37**	-0.03	0.10	0.02	0.16	0.13	0.32**

** Correlation significant at the 0.01 probability level (2-tailed).

* Correlation significant at the 0.05 probability level (2-tailed).

3.4. Extractable C concentrations

Organic C was dissolved by the three extractants in the sequence – C_{PP} > C_{DCB} > C_{OX} (Fig. 3). The PP extracted C (C_{PP}) ranged from 2.8 to 31.0 g kg⁻¹ with a mean of 13.3±1.17 g kg⁻¹ in the topsoils and 11.9±1.14 g kg⁻¹ in the subsoils. The mean C_{PP} concentration was significantly greater (p=0.004) in the topsoils than the subsoils. Sodium pyrophosphate extracted on an average 62±0.2 % of the total soil C across both soil depths.

The concentration of DCB extractable C (C_{DCB}) ranged from 0.5 g kg⁻¹ to 28.4 g kg⁻¹ in all samples, with a significantly (p=0.025) greater mean concentration in the topsoils (9.8±1.25 g kg⁻¹) than in the subsoils (8.5±1.11 g kg⁻¹) (Fig. 3). The extractable C_{DCB} constituted 41±0.2 % (mean) of the total soil C across both depths.

The OX extractable C (C_{OX}) concentration varied between 0.5 g kg⁻¹ and 26.5 g kg⁻¹ across the two depths. The concentration was similar in the topsoils (6.4±1.08 g kg⁻¹) and subsoils (6.3±1.11 g kg⁻¹) (Fig. 3). The C_{OX} represented on an average 28±0.2 % of the total C in all samples.

The concentration of the DCB extractable C minus OX extractable C (C_{DCB} – C_{OX}), i.e., C associated only with crystalline forms of Fe oxides was similar in the topsoils (3.3±0.55 g kg⁻¹) and the subsoils (2.29±0.28 g kg⁻¹) (Fig. 3). The C_{DCB} – C_{OX} constituted on the average 13±0.2 % of the total soil C across both depths.

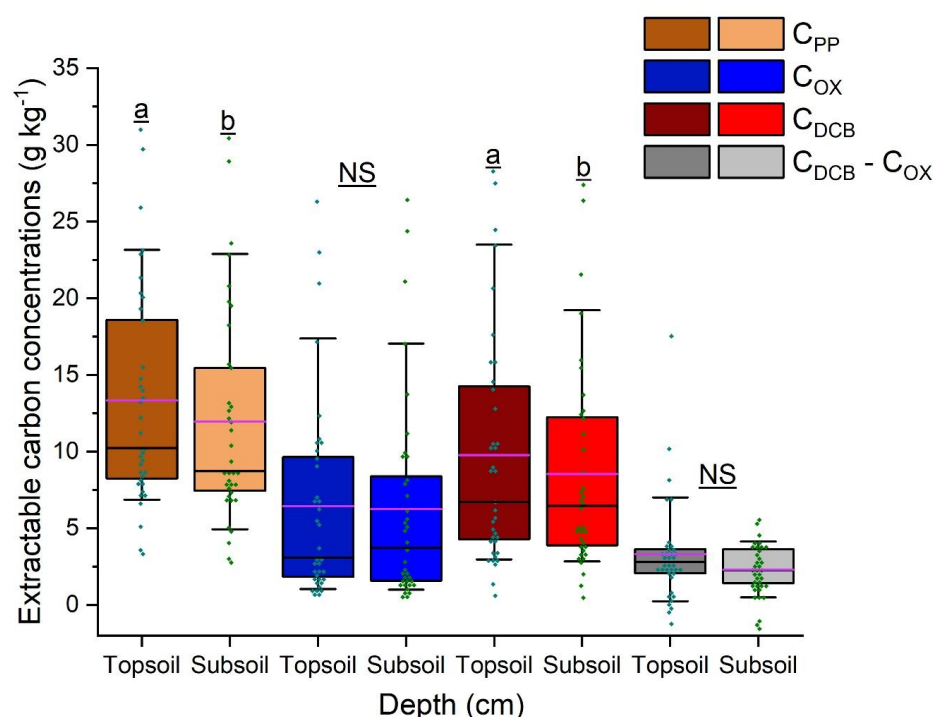


Figure 3. Concentrations of carbon (C) in topsoils and subsoils extracted by sodium pyrophosphate (PP), ammonium oxalate (OX), dithionite-citrate-bicarbonate (DCB) reagents, and concentration of carbon extracted by dithionite-citrate-bicarbonate (DCB) minus concentration of carbon extracted by ammonium oxalate (OX). Boxes comprise of the 25th and 75th percentiles, whiskers extend from the 10th to 90th percentiles. Purple and black bars signify the mean and the median values, respectively. Dark cyan and green diamonds indicate data for individual samples; the symbols are displaced horizontally to avoid overlaps. Different letters indicate significant differences in the mean values of samples from different soil depths based on paired T-test ($p < 0.05$). NS means non-significant difference.

3.5. Relationship between extractable metal phases, soil properties and extractable C

Relationships between extractable metals, extractable C and soil properties using the generalized linear mixed model (GLMM) regression are shown in Tables 5 and 6. When TC was included as a fixed effect term into the model (Table 5), different forms of extractable metals showed no significant relationship with the extractable forms of C in both top- and sub-soils except for significant negative relationships of Fe_{OX} (topsoil), Al_{OX} and Al_{DCB} (subsoil) with extractable C. However, the TC showed significant ($p < 0.001$) relationships with the different extractable forms of C in both topsoils and subsoils (Table 5). In addition, pH had a significant relationship with the three forms of extractable C in the topsoils. Clay content also showed a significant relationship with all three forms of extractable C in the subsoils, whereas silt showed a significant relationship with C_{OX} ($p < 0.001$) and C_{DCB} , ($p = 0.01$) in the topsoils. When TC was excluded from the generalized linear mixed model (GLMM), the three forms of extractable C showed positive relationships with pH, silt or clay and extractable Fe and Al



(Table 6). For example, Fe_{PP} , pH, and silt showed statistically significant ($p < 0.001$, $p < 0.001$, and $p = 0.0004$, respectively) relationships with C_{PP} in the topsoils. Fe_{OX} and pH had significant positive ($p < 0.001$ and $p = 0.026$, respectively) relationships with C_{OX} in the topsoils, and Fe_{OX} , pH, and silt ($p < 0.001$) had significant positive relationships with C_{OX} in the subsoils. Also, Al_{DCB} ($p < 0.001$) and pH ($p = 0.002$) showed significant positive relationships with C_{DCB} in the topsoils; and Fe_{DCB} ($p < 0.001$) and silt ($p = 0.0006$) showed significant positive relationships with C_{DCB} in the subsoils.

3.6. Extractable C: metal molar ratio

The molar ratio of C to metals (C: M (or Fe+Al)) extracted by each reagent in the samples from two depths are presented in Fig. 4. In the PP extracts, the C:M molar ratio for all samples varied between 1.2 and 40.9 and it was similar for the topsoils (9.4 ± 1.3) and subsoils (9.6 ± 1.4) (Fig. 4a). The C:M molar ratio in the OX extracts ranged from 0.1 to 10.7, with a similar mean value of 1.6 ± 0.3 and 1.4 ± 0.3 for the topsoils and the subsoils (Fig. 4b), respectively. The C:M molar ratio in the DCB extracts ranged from 0.03 to 8.5 in all samples and also had similar mean ratios in the topsoils (1.0 ± 0.2) and subsoils (0.7 ± 0.1) (Fig. 4c).

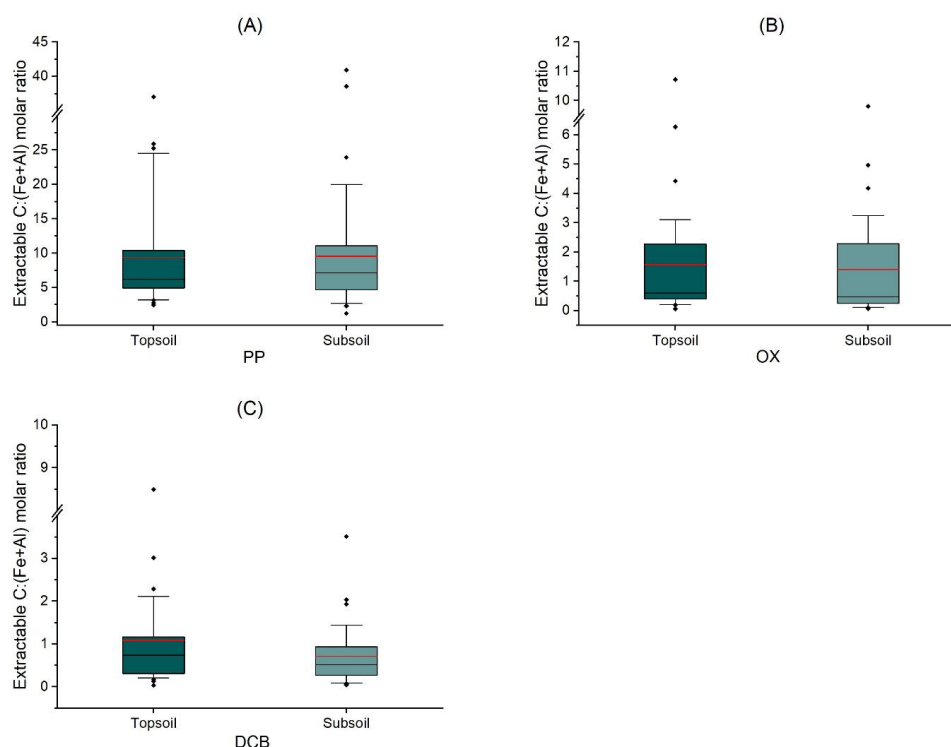


Figure 4. Boxplots of C:(Fe+Al) molar ratios in the extracts of sodium pyrophosphate (PP), ammonium oxalate (OX), and dithionite-citrate-bicarbonate (DCB) of topsoils and subsoils. Boxes comprise of the 25th and 75th percentiles, whiskers extend from the 10th to 90th percentiles, red bars signify mean, black bars signify median, and black dots show the outliers.



Table 5. Fixed effect parameter estimates in the topsoils and subsoils obtained from the generalized linear mixed model. The extractable metals and soil properties (including TC) were used as the fixed effect terms, soil types as the random effect and sodium pyrophosphate extractable C, ammonium oxalate extractable C and dithionite-citrate-bicarbonate extractable C as the dependent variables. The standard errors are given in parentheses.

	Depth	Fixed effect term	Regression coefficient	95 % Lower	95% Upper	F Ratio	Prob>F
C _{PP}	Topsoil	Fe _{PP}	0.06 (0.06)	-0.06	0.19	1.06	0.31
		pH	0.26 (0.08)	0.08	0.43	9.51	0.005
		TC	0.02 (0.003)	0.01	0.03	31.58	<0.001
		Intercept	0.002 (0.01)	-0.02	0.02		
	Subsoil	Fe _{PP}	0.05 (0.09)	-0.24	0.14	0.31	0.58
		Al _{PP}	0.08 (0.07)	-0.22	0.17	1.22	0.28
		Clay	0.01 (0.004)	8.08	0.003	0.02	0.007
		TC	0.05 (0.01)	0.04	0.07	48.57	<0.001
		Intercept	-0.001 (0.01)	-0.02	0.02		
C _{OX}	Topsoil	Fe _{OX}	-0.09 (0.04)	-0.16	-0.01	5.99	0.02
		pH	0.56 (0.130)	19.84	0.26	0.87	0.004
		Silt	0.06 (0.01)	0.04	0.09	25.80	<0.001
		TC	0.04 (0.01)	0.02	0.05	38.18	<0.001
		Intercept	0.004 (0.03)	-0.05	0.06		
	Subsoil	Fe _{OX}	-0.11 (0.04)	-0.19	-0.02	7.21	0.011
		Al _{OX}	-0.15 (0.06)	-0.26	-0.03	6.32	0.017
		Clay	0.02 (0.01)	0.01	0.03	8.40	0.007
		TC	0.09 (0.01)	0.07	0.11	86.31	<0.001
		Intercept	0.05 (0.06)	-0.07	0.16		
C _{DCB}	Topsoil	Fe _{DCB}	0.02 (0.01)	-0.01	0.02	0.52	0.47
		Al _{DCB}	-0.10 (0.09)	-0.29	0.09	1.13	0.30
		pH	0.26 (0.12)	0.01	0.52	4.78	0.04
		EC	-0.06 (0.01)	-0.08	-0.04	29.11	<0.001
		Silt	0.03 (0.01)	0.01	0.05	7.79	0.01
		TC	0.06 (0.01)	0.05	0.07	92.56	<0.001
		Intercept	0.01 (0.02)	-0.03	0.05		
	Subsoil	Fe _{DCB}	-0.001 (0.01)	-0.01	0.01	0.05	0.83
		Al _{DCB}	-0.10 (0.05)	-0.21	0.01	3.50	0.07
		Clay	0.01 (0.01)	0.001	0.02	5.26	0.03
		TC	0.06 (0.01)	0.05	0.07	77.69	<0.001
		Intercept	0.05 (0.05)	-0.05	0.15		



Table 6. Fixed effect parameter estimates in the topsoils and subsoils obtained from the generalized linear mixed model. The extractable metals and soil properties (excluding the TC) were used as the fixed effect terms, soil types as the random effect and sodium pyrophosphate extractable C, ammonium oxalate extractable C and dithionite-citrate-bicarbonate extractable C as the dependent variables. The standard errors are given in parentheses.

	Depth	Fixed effect term	Regression coefficient	95 % Lower	95% Upper	F Ratio	Prob>F
C _{PP}	Topsoil	Fe _{PP}	0.23 (0.03)	0.17	0.30	53.96	<0.001
		pH	0.27 (0.06)	0.16	0.38	22.56	<0.001
		Silt	0.02 (0.01)	0.01	0.03	14.12	0.0004
		Intercept	0.37 (0.37)	-0.37	1.11		
	Subsoil	Fe _{PP}	0.30 (0.04)	0.21	0.38	44.45	<0.001
		Al _{PP}	0.14 (0.05)	0.23	0.44	8.67	0.004
		Clay	0.01 (0.003)	0.002	0.01	7.34	0.009
		Intercept	2.17 (0.11)	1.92	2.41		
C _{OX}	Topsoil	Fe _{OX}	0.08 (0.02)	0.04	0.12	18.03	<0.001
		pH	0.18 (0.08)	0.02	0.34	5.16	0.026
		Intercept	0.30 (0.51)	-0.72	1.33		
	Subsoil	Fe _{OX}	0.14 (0.02)	0.09	0.18	33.09	<0.001
		Si _{OX}	-0.52 (0.19)	-0.77	-0.27	17.76	<0.001
		pH	0.35 (0.08)	0.19	0.52	18.44	<0.001
		Silt	0.04 (0.01)	0.02	0.06	29.84	<0.001
		Intercept	-1.15 (0.53)	-2.22	-0.09		
C _{DCB}	Topsoil	Al _{DCB}	0.22 (0.04)	0.15	0.29	36.47	<0.001
		pH	0.24 (0.07)	0.09	0.39	10.22	0.002
		EC	0.03 (0.01)	0.01	0.04	20.98	<0.001
		Clay	-0.01 (0.003)	-0.02	-0.002	5.92	0.018
		Intercept	-0.16 (0.58)	-1.33	1.02		
	Subsoil	Fe _{DCB}	0.02 (0.003)	0.01	0.02	23.48	<0.001
		Silt	0.02 (0.01)	0.01	0.04	13.08	0.0006
		Intercept	1.33 (0.24)	0.78	1.87		



474 4 Discussion

475

476 4.1 Soil physicochemical properties in relation to OC

477 The association of OC with Fe and Al oxides can be affected by soil pH, which influences
478 mineral phases of Fe and Al oxides, surface charge characteristics of these oxides, and
479 chemical interactions between OC and metal oxides (Zhao et al., 2016; Ye et al., 2022). Most
480 of the soils had acidic pH and a significant positive relationship was observed between soil pH
481 and extractable C (Tables 5 and 6). Both Fe and Al oxides carry a net positive charge in the
482 acidic to near neutral pH range, as the point of zero charge of these minerals range from 7.8
483 to 10 (Sparks et al., 2024). Also, a net negative charge on soil organic matter would be
484 expected under these pH conditions, which increases with increasing pH due to deprotonation
485 of acidic functional groups of OC, such as carboxyls, quinones and enols (Sparks et al., 2024).
486 Other studies have observed a significant correlation between soil pH and Fe/Al associated
487 OC, which implies that soil pH regulates the propensity of Fe and Al oxides to associate with
488 OC (Wang et al., 2017; Zhao et al., 2019; Ye et al., 2022).

489 The significant positive relationships between extractable C, clay and silt (Table 5 and 6)
490 indicates the important role of the fine soil fractions in stabilising OC in soils, which is
491 consistent with several previous studies (Six et al., 2002; Jindaluang et al., 2013; Yu et al.,
492 2019). Soil clay fraction is important in stabilising OC, with its ability to protect OC via physical
493 and chemical interactions and other indirect factors. Oades (1988) suggested that finer
494 textured soils have more cation bridges (Ca^{2+} in neutral, Fe^{3+} and Al^{3+} in acidic soils) to bind
495 organic molecules to negatively charged soil clay particles. While Christensen (1992)
496 supported the idea of increased binding of organic molecules via cation bridges in clayey soils,
497 he proposed that the principal effect of higher clay (or clay + silt) contents occurred via
498 stabilisation of microbially processed organic residues, that are readily lost in sandy soils.
499 In contrast, Feller et al. (2020) concluded that greater biomass inputs or aggregation in finer
500 textured soils caused correlations between SOC and clay concentrations in tropical soils. It has
501 been reported that there is a greater amount of MAOM in soils with increased clay and silt
502 contents, and hence increased OM storage through binding to mineral surfaces in soils
503 (Stewart et al., 2007; Laganière et al., 2010; Vos et al., 2018). Others have suggested increased
504 occlusion of particulate OM in soil aggregates rich in clay-size minerals that restricts access for
505 microbial degradation (McCarthy et al., 2008; Steffens et al., 2017; Schweizer et al., 2021).
506 Jindaluang et al. (2013) reported that organic compounds with aromatic and amide functional
507 groups are selectively preserved in fine textured soils, while recalcitrant aliphatic-C existed in
508 the coarse textured soils. Hassink (1997) proposed that C associated with organo-mineral
509 complexes are protected chemically and the level of protection increased with increased silt
510 fraction in the soil. It is possible that all or many of these mechanisms occur simultaneously,
511 but the adsorption of organic compounds to clay surfaces appeared to be a major factor in
512 the preservation of SOM.

513

514 4.2 Extractable metal(loid)s

515 The dissolution of metal(loid)s during PP, OX and DCB extractions provides an approximation
516 of their specific phases in soils, acknowledging that these extractions are not completely
517 selective and often some overlaps exist in the extracted fractions of metal(loid)s (Parfitt, 2009;
518 Rennert, 2019). In addition, re-adsorption of some extracted OC can also occur in the



519 extraction procedure. Nevertheless, chemical extraction procedures offer the best option for
520 routine evaluation of OC association with different chemical and mineralogical phases of
521 metal(loid)s in large number of soil samples. The extracted concentration of Fe, Al, and Si
522 followed the same sequence (i.e., DCB > OX > PP), while Mn was extracted in the sequence –
523 OX > DCB > PP.

524 The DCB procedure is known to extract pedogenic or free Fe oxides (and Al substituted in the
525 structure of goethite and hematite) as well as poorly crystalline and SRO Fe/Al oxides
526 (McKeague and Day, 1966; Shang and Zelazny, 2008). The large proportion of the total Fe
527 extracted by the DCB procedure reflected high concentration of secondary Fe oxides in the
528 studied soils. This partly support our hypothesis that concentration of extractable Fe will be
529 the most prominent element in the soils. In addition to the targeted phases, the DCB
530 procedure may also extract a fraction of organic complexed Fe from soils (Bascomb, 1968),
531 while some Fe oxides (such as magnetite and maghemite) and Fe sulfides are only partially or
532 not extracted in the DCB extraction procedure (Pansu and Gautheyrou, 2006; Rennert and
533 Lenhardt, 2025). In the oxalate extraction procedure, Fe present in non-crystalline, poorly
534 ordered and short-range ordered (such as ferrihydrite) was dissolved and a fraction in the
535 organic complexes may also be dissolved (Schwertmann, 1964; McKeague and Day, 1966;
536 Bascomb, 1968; Loeppert and Inskeep, 1996; Mikutta et al., 2006). Oxalate extracted Fe had
537 often been used to estimate ferrihydrite, a short range ordered Fe oxide phase, concentrations
538 in soils (Carlson and Schwertmann, 1981; Schwertmann et al., 1982). The small ratio of
539 $Fe_{OX}:Fe_{DCB}$, referred to as active Fe ratio, indicate the degree of crystallinity of Fe oxides in soils
540 (McKeague and Day, 1966). In topsoils and subsoils, the mean value of $Fe_{OX}:Fe_{DCB}$ ratio was
541 0.18 and 0.16 (< 0.2), respectively, which suggest that most of the Fe was present in crystalline
542 Fe oxides (i.e., goethite and hematite) in soils and the content of poorly crystalline Fe phases
543 was small irrespective of soil depth (Mikutta et al., 2006). The significant correlation between
544 Fe_{DCB} and Fe_{OX} (Table 4), may be due to some overlap in the phases extracted by the two
545 procedures. The DCB method is known to extract both crystalline and SRO (and poorly
546 crystalline) Fe oxides from soils (McKeague and Day, 1966; Shang and Zelazny, 2008). In
547 addition to SRO Fe oxides, oxalate is known to dissolve crystalline Fe oxides, such as magnetite,
548 maghemite, lepidocrocite, and poorly crystalline and nano-crystalline goethite and hematite
549 (Childs and Wilson, 1983; Walker, 1983; Mansfeldt et al., 2012). The close relationship
550 between Fe_{DCB} and Fe_{OX} may also suggest that both Fe forms were function of similar
551 pedogenic processes (Rezapour et al., 2015).

552 Pyrophosphate (at pH 10) has been considered as a good extractant for organic Fe and Al
553 complexes and 'active' amorphous inorganic forms of Fe from soils (Bascomb, 1968; Loveland
554 and Digby, 1984; Pansu and Gautheyrou, 2006). Fe_{PP} concentration in soils was generally small
555 and decreased with soil depth (Fig. 2a), a trend similar to the OC concentration in soils (Fig.
556 S1). These results are consistent with the published data for non-podzolic soils (Bascomb,
557 1968; Daly, 1982; Parfitt and Childs, 1988). Large variations in the Fe_{PP} concentration with
558 different centrifugation speeds and/or use of flocculating agents have been observed in
559 several past studies (Jeanroy and Guillet, 1981; McKeague and Schuppli, 1982; Schuppli et al.,
560 1983; Loveland and Digby, 1984; Skjemstad et al., 1990). The variability in the Fe_{PP}
561 concentration had also been attributed to peptization of micro-crystalline goethite (including
562 Al-substituted goethite) and ferrihydrite on which organic matter had been adsorbed
563 (Borggaard, 1988; Parfitt and Childs, 1988; Rennert, 2019). There was no clear evidence for
564 the presence of any solid phase in the supernatant in the procedure with medium-speed
565 centrifugation ($3280 \times g$) followed by micro-filtration ($0.2 \mu m$). On the contrary, Ruiz et al.



(2026) reported visual evident for the presence of suspended fine particles during filtration procedure, indicating colloidal dispersion. Most of the soils used in this study had acidic pH and moderate amount of Al. The presence of Al, and the use of medium-speed centrifugation and micro-filtration may have mitigated the dispersion of colloidal materials in these soils. The Fe_{PP} concentration data were consistent with the results obtained in other studies that used high-speed centrifugation and/or ultrafiltration (Schuppli et al., 1983; Loveland and Digby, 1984). The significant correlation between Fe_{PP} and Fe_{OX} (Table 4) suggests some overlapping extractions of amorphous and SRO phases and Fe-organic complexes by the two reagents (Bascomb, 1968; Higashi et al., 1981; Jeanroy and Guillet, 1981).

The differences between the mean Al_{PP} , Al_{OX} and Al_{DCB} concentrations are much smaller compared with the mean Fe concentrations extracted by the three extractants (Fig. 2a and b) and similar results have been obtained in previous studies (Skjemstad et al., 1990; Singh and Gilkes, 1991; Wiriyaakittateekul et al., 2005). Al_{DCB} and Al_{OX} concentrations are similar, with the mean value of $\text{Al}_{\text{OX}}:\text{Al}_{\text{DCB}}$ ratio in the topsoils and subsoils being 1.02 and 0.95, respectively. These results are consistent with a large North American soil dataset (Hall and Thompson, 2022). Crystalline Al oxides are generally not dissolved during the DCB extraction and Al in DCB extracts results from various sources, including Al-organic complexes, Al adsorbed on Fe oxides, and Al-substituted in the structure of Fe oxides (Rennert, 2019). A significant positive correlation was found between Al_{DCB} and Al_{OX} fractions (Table 4), which reflects that the two reagents might have extracted some similar Al-phases, such as, organo-Al complexes (Evans and Smillie, 1976; Parfitt and Childs, 1988; Dahlgren, 1994). The Al_{DCB} concentration showed a very strong correlation with Fe_{DCB} (Table 4), which was possibly due to a concurrent release of Al and Fe from the dissolution of crystalline Fe oxides (i.e., goethite and hematite) during the DCB extraction. Such observations have been made in other studies and Al substitution in goethite (up to 35 mol%) and hematite (up to 23 mol%) is well known in acidic and highly weathered soils (Bigham et al., 1978; Fitzpatrick and Schwertmann, 1982; Childs and Wilson, 1983; Parfitt and Childs, 1988; Singh and Gilkes, 1992). The significant relationship between Al_{PP} and Al_{OX} (Table 4), implies that oxalate may have dissolved Al from organic complexes (McKeague and Day, 1966; Parfitt and Henmi, 1982; Farmer et al., 1983).

Mn oxides in soils often occur in small concentration and in poorly crystalline forms (Scheinost and Singh, 2023); our results were consistent with this assertion, with Mn mostly being present in poorly crystalline forms that were slightly more efficiently extracted by the oxalate procedure than the DCB procedure. Jarvis (1984) also observed greatest Mn extraction with oxalate from acidic soils, where several extractants including DCB and PP were used. It was speculated that during oxalate extraction Mn was released from the reduction of relatively unweathered, highly crystalline oxides, or from mixed forms of Mn-Fe oxides. Manganese has been found to be strongly associated (adsorbed or substituted in structures) with natural Fe oxides (Singh and Gilkes, 1992). Structural substitution of Mn in synthetic and natural goethite and hematite is well known (Singh and Gilkes, 1992; Manceau et al., 2000; Singh et al., 2000; Singh et al., 2002). The order of Mn concentration extracted by the three extractants in our study is similar to the order obtained for several soils from England (Jarvis, 1984). However, the trend with greater Mn concentrations in the topsoils than the sub soils in our study is opposite to the earlier study, which might be due to the poor drainage and periodic waterlogging conditions in English soils. A significant relationship was observed between Mn_{PP} and Mn_{OX} (Table 4) that is similar to an earlier study; however, the relationship of Mn_{OX} was stronger with Mn_{DCB} than Mn_{PP} (Jarvis, 1984).



612 Silicon concentrations were small in all three extracts (Fig. 2). These results indicate that
613 poorly crystalline Al-Si phases, such as allophane and imogolite, were absent and there was
614 little or no dissolution of well-crystalline aluminosilicates in the soils (Parfitt and Childs, 1988).
615 The sequence of Si extracted is consistent with the limited published data for the three
616 extractions (Paterson et al., 1993). The highest Si concentration was found in the DCB extracts,
617 which might have released from different sources including adsorbed Si on the surfaces of
618 SRO and crystalline Fe oxides, Si occluded in Fe oxides, and some Si dissolved from micro-
619 crystalline aluminosilicate clay minerals and quartz (Weaver et al., 1968; Gamble and Daniels,
620 1972; Parfitt and Childs, 1988). Hall and Thompson (2022) reported greater Si in DCB extract
621 than OX extract in soils from all soil orders except Andisols, where the trend was opposite.
622 Fe_{OX} showed a significant correlation with Si_{OX} (Table 4), indicating that Si_{OX} was largely
623 adsorbed onto the surfaces of SRO Fe phases and small amounts may also have been extracted
624 from SRO Fe silicates (Parfitt and Henmi, 1982; Colombo and Torrent, 1991). Non-crystalline
625 Si is not dissolved during the ammonium oxalate extraction (Wada, 1989).

627 4.3 Organic C association with different fractions of extractable metals

628 Sodium pyrophosphate is generally assumed to dissolve organo-metal complexes (Bascomb,
629 1968; Takahashi and Dahlgren, 2016). Pyrophosphate extracted the least amount of Fe, Al, Mn
630 and Si amongst the three extractants (Fig. 2) but dissolved the highest concentration of OC
631 (Fig. 3). Pyrophosphate extractable C constituted an average 62 % (range = 19 - 90 %) of the
632 total SOC content across all samples from both soil depths. Yu et al. (2023) reported 39 to 60
633 % contribution from C_{PP} to the total SOC, whereas 30 - 43 % contribution to the total SOC was
634 reported in other studies (Wagai et al., 2013; Heckman et al., 2018). A significant positive
635 relationship between Fe_{PP} and C_{PP} in both topsoils and subsoils (Table 6) suggests that organo-
636 Fe complexes were predominantly extracted in the PP procedure. Several other studies have
637 observed significant correlations of Fe_{PP} and Al_{PP} with SOC (Masiello et al., 2004; Wagai et al.,
638 2013; Lawrence et al., 2015; Takahashi and Dahlgren, 2016; Heckman et al., 2018). The
639 association of significant amount of SOC in Fe-organic complexes limits desorption, oxidative
640 degradation, and biodegradation of OC (Parfitt et al., 2002; Kaiser and Guggenberger, 2003;
641 Zimmerman et al., 2004; Rasmussen et al., 2006).

642 The interpretation of PP extraction data requires some caution due to the dissolution of alkali-
643 induced peptised organic matter (OM) from other fractions (Wagai et al., 2013; Coward et al.,
644 2017). A limited number of samples containing a range of total Fe were extracted with PP at
645 pH 7.5 to evaluate the dissolution of OC due to peptization of Fe/Al oxides adsorbed with
646 organic matter at pH 10 (Parfitt and Childs, 1988; Rennert, 2019). The C_{PP10} and C_{PP7.5} were
647 similar in the four replications of the analysed samples (regression slope = 1.02 ± 0.04 , $R^2 =$
648 0.97 ; Fig. S3a). However, the PP at pH 10 extracted nearly three times greater Fe as compared
649 to PP at pH 7.5 (regression slope = 2.88 ± 0.57 , $R^2 = 0.59$; Fig. S3b). The higher Fe extracted at
650 pH 10 than pH 7.5 in this study may be due to the dissolution of Fe from phases other than
651 Fe-organic complexes at higher pH. The other three elements showed variable behaviour, with
652 no consistent trend or relationship between Al_{PP7.5} and Al_{PP10} (Fig. S3c), while Mn_{PP10} and
653 Mn_{PP7.5} concentrations were almost similar (regression slope = 0.85 ± 0.18 , $R^2 = 0.55$; Fig. S3d),
654 and PP at pH 10 extracted about half of the Si concentration as compared to the Si
655 concentration in PP extract at pH 7.5 (regression slope = 0.48 ± 0.09 , $R^2 = 0.57$; Fig. S3e). These
656 results show dissolution of some non-target particles, particularly Fe oxides, at pH 10,
657 however, this did not release any additional OC from the dissolved phases. Ruiz et al. (2026)
658 reported an average of about 46 % greater Fe and Al concentrations extracted with PP at pH



10 than pH 7. PP at pH 10 extracted higher OC concentrations with an average of approximately 40 % compared to PP at pH 7. However, PP can extract poorly and nano crystalline Fe and Al oxides particularly OM complexed by cation bridging. Parfitt and Childs (1988) reported that dispersion typically occurs in clayey soils (> 40 % clay content) containing goethite and ferrihydrite; soils in our study generally had > 40 % clay content, with kaolinite and hydroxy-interlayered vermiculite as the dominant phyllosilicates in the clay fraction. Short-range-order Fe phases such as ferrihydrite have the capacity to adsorb high concentrations of SOC due to their extensive specific surface area and the presence of reactive sites on surfaces. The significant positive relationship between C_{OX} and Fe_{OX} in the soils (Table 6) suggests that SRO and poorly crystalline Fe oxides contributed substantially to the OC accumulation in the studied soils. About 28 % of the total SOC was solubilised during the OX extraction across both soil depths. Positive correlations between Fe_{OX} and C_{OX} have been reported in several other studies (Kaiser and Guggenberger, 2003; Kleber et al., 2005; Wiseman and Püttmann, 2006; Rasmussen et al., 2018; Yu et al., 2021; Hall and Thompson, 2022; Amenkhiennan et al., 2024). Oxalate extractable Si had a significant negative relationship with the C_{OX} in subsoils (Table 6), which suggests that soil samples were devoid of poorly-crystalline aluminosilicates (e.g., allophane and imogolite), which are common in Andosols and Spodosols (Dahlgren, 1994). Oxalate is a C based complexing agent, hence residual C in soils after OX extraction was compared with C in soils after HH extraction, which does not have C compounds (Heckman et al., 2018). Carbon concentrations extracted by the two reagents in selected samples were quite similar (regression slope between C_{OX} and C_{HH} = 1.18 ± 0.07 , $R^2 = 0.94$; Fig. S4a), which shows that there was little of no C contamination from OX in the OX extraction procedure. In addition, HH and OX extracted similar Fe (regression slope = 1.06 ± 0.07 , $R^2 = 0.94$; Fig. S4b), Al (regression slope = 0.99 ± 0.03 , $R^2 = 0.98$; Fig. S4c) and Mn (regression slope = 0.93 ± 0.02 , $R^2 = 0.99$; Fig. S4d) concentrations, while more Si was dissolved in the HH extraction (regression slope = 0.62 ± 0.05 , $R^2 = 0.88$; Fig. S4e). Hydroxylamine Hydrochloric acid and OX have been reported to have similar ability to extract SRO minerals (Chao and Zhou, 1983), which is consistent with our results. The C_{DCB} is expected to be C associated with secondary crystalline Fe oxides including SRO phases. In the DCB procedure, 41 ± 0.2 % (mean) of the total SOC was extracted from soil samples from both depths. Since DCB extracts both SRO and crystalline Fe and Al oxides, we calculated the C only in association with well crystalline phases by subtracting the C_{OX} content from the C_{DCB} content for soil samples individually. On an average only 13 ± 0.2 % of the total SOC was found to be associated with crystalline Fe and Al oxides, which shows a relatively small contribution of crystalline Fe oxides in the preservation of OC in the studied soils. The relatively small contribution of crystalline Fe oxides in the preservation of OC might be related to smaller SSA and reduced hydroxyl surface sites of crystalline Fe oxides than to SRO Fe and Al oxides (Mikutta et al., 2006; Heckman et al., 2013; Newcomb et al., 2017; Heckman et al., 2018). Consequently, despite the high concentration of DCB extractable Fe and other elements in the studied soils (Fig. 2a), the C_{DCB} portion was small. In the GLM model, the significant positive relationship between C_{DCB} and Al_{DCB} in topsoils (Table 6), was possibly due to Al substitution in the structure of crystalline Fe oxides (e.g. goethite and hematite) in soils, which dissolved during the DCB extraction. In the subsoils that have small OC concentration, a significant positive relationship existed between C_{DCB} and Fe_{DCB} . In both the top- and subsoils, Fe_{DCB} showed a similar significant positive relationship with Al_{DCB} , which suggests Al occurred in the structures of goethite and hematite and dissolved during the DCB extraction.



Significant positive correlations between OC and Fe_{DCB} , in soils dominated by crystalline Fe oxides have been observed in past studies (Mikutta et al., 2006; Yeasmin et al., 2017). The C_{DCB} and C_{DH} were similar in the selected samples (regression slope = 1.04 ± 0.05 , $R^2 = 0.97$; Fig. S5a), which showed that no additional C was introduced in soil residues during the DCB extraction. The DH extracted lot less Fe and Al as compared to DCB (Fig. S5b, S5c), which was expected because the DH procedure involved shaking for 16 hours at room temperature, whereas in the DCB method samples were extracted by heating for 15 minutes at 80°C . Mn_{DCB} and Mn_{DH} were similar (Fig. S5d), and the DCB method extracted nearly 1.6 times Si as compared to the DH extraction (Fig. S5e).

4.4 Extractable C: metals (Fe+Al) molar ratio and the potential binding mechanisms

The molar ratio of C:M had been used to describe binding mechanisms between C and Fe + Al metals. The C:M molar ratio provides information on the potential nature and structure of organo-metal associations in the dissolved phases (Kaiser et al., 2007; Huang et al., 2021; Li et al., 2025). Adsorption and co-precipitation experiments have demonstrated that organic functional groups preferentially bind to Fe and Al oxides (Eusterhues et al., 2014; Han et al., 2019; Jardine and Zelazny, 2020; Vance et al., 2020; Li et al., 2023). A C:M molar ratio of 1.0 ($\text{C:M} = 1.0$) indicated the maximum sorption capacity of metals oxides for natural OC (Wagai and Mayer, 2007; Lalonde et al., 2012; Zhao et al., 2016; Coward et al., 2017). Molar ratio values less than 1.0 ($\text{C:M} < 1.0$) suggested the formation of the Fe/Al-associated OC complex via inner-sphere adsorption mechanism (Coward et al., 2018), while C:M values > 1.0 , indicated both adsorption and co-precipitation processes co-existed in forming Fe/Al-associated OC complexes. Specifically, C:M values > 6.0 , indicated co-precipitation or chelation of organic ligands, generating low-density organic-rich structures forming Fe/Al-associated OC complexes (Higashi, 1983; Guggenberger and Kaiser, 2003; Wagai and Mayer, 2007; Zhao et al., 2016; Coward et al., 2017; Jeewani et al., 2021). Our extraction results suggested the presence of mixed mechanisms for association between C and (Fe+Al) phases in the studied soils. In the PP fractions, high C:M molar ratios indicated the formation of precipitates of Fe/Al-associated OC complexes in both the top- and sub-soils. The PP extraction liberated C from low-density organic dominated structures (Coward et al., 2017; Coward et al., 2018; Jeewani et al., 2021). The OX fractions had C:M molar ratio > 1 , suggesting the occurrence of both adsorption and co-precipitation (in organo-metal complexes) processes. In contrast, the relatively low C:M molar ratios in the DCB fraction suggested the adsorption of OC on to crystalline Fe/Al oxides, which possibly occurred via inner-sphere complexation formation (Coward et al., 2017; Jeewani et al., 2021; Yu et al., 2023). The actual mechanisms need to be verified using spectroscopy or other techniques.

5 Conclusions

We determined the association of OC with different fractions of Fe, Al, Mn and Si in top- and sub-soils using chemical extraction procedures. The largest portion ($62 \pm 0.2\%$) of the total OC was found in organo-metal complexes extracted by PP, while about $28 \pm 0.2\%$ of the total OC was associated with SRO phases extracted by OX. Crystalline Fe/Al oxides appeared to play a relatively small role in the stabilisation of OC in the studied soils, with only 13% of the total OC associated with crystalline Fe oxides.



751 The concentration of extractable Fe was the highest amongst the metal(loid)s in the both top-
752 and sub-soils. DCB and OX extracted almost similar Al concentrations, suggesting significant
753 Al present in SRO, while DCB extracted Al that was substituted for Fe in the structure of Fe
754 oxides. Crystalline Fe oxides (i.e., goethite and hematite) were dominant in both topsoils and
755 subsoils. Poorly crystalline and SRO alumino-silicate minerals were absent in the soils.
756 The C:M molar ratio > 6.0 in the PP extracts suggested the co-precipitation of OC with metals
757 by forming Fe/Al complexes, while C:M molar ratio of about 1.5 in the OX fraction indicated
758 both adsorption and co-precipitation processes in the OC association with SRO of Fe and Al.
759 In contrast, the C:M molar ratio ≤ 1 in the DCB fraction suggested that OC was adsorbed on to
760 crystalline Fe/Al oxides possibly forming inner-sphere complexes. Our results show that
761 despite of some non-selectivity and over-lapping dissolution, extraction procedures have
762 potential to routinely determine the role of different forms of Fe and Al in stabilising OC in
763 soils, particularly for large sample sets.



764 Appendix

765 Table A1. Dataset of extractable metal(loid)s used for this study

Location	Depth	Fe _{ox} (g/kg)	Fe _{DCB} (g/kg)	Fe _{PP} (g/kg)	Al _{ox} (g/kg)	Al _{DCB} (g/kg)	Al _{PP} (g/kg)	Mn _{ox} (g/kg)	Mn _{DCB} (g/kg)	Mn _{PP} (g/kg)	Si _{ox} (g/kg)	Si _{DCB} (g/kg)	Si _{PP} (g/kg)
Westwood (Cobbitty)	0-20cm	6.50	38.00	4.37	1.65	3.33	1.30	0.87	1.15	0.68	0.29	1.27	0.48
Westwood (Cobbitty)	0-20cm	6.92	37.74	3.86	1.76	3.35	1.19	0.91	1.14	0.62	0.27	1.33	0.32
Westwood (Cobbitty)	20-40cm	5.17	51.37	3.13	2.72	5.28	1.70	0.22	0.37	0.11	0.31	1.26	0.19
Westwood (Cobbitty)	20-40cm	5.12	49.98	4.13	2.64	5.43	2.25	0.21	0.36	0.14	0.25	1.32	0.24
Landsdowne (Cobbitty)	0-20cm	5.32	25.89	1.00	1.52	2.50	0.57	2.64	2.89	0.36	0.44	0.91	0.27
Landsdowne (Cobbitty)	0-20cm	5.38	24.80	1.28	1.58	2.32	0.76	2.53	2.76	0.44	0.34	0.93	0.39
Landsdowne (Cobbitty)	20-40cm	6.76	35.47	0.82	2.39	3.58	0.89	2.89	3.11	0.31	0.68	1.67	0.59
Landsdowne (Cobbitty)	20-40cm	6.86	39.78	0.70	2.46	3.92	0.77	2.93	3.62	0.26	0.70	1.86	0.49
Narrabri (W2)	0-20cm	1.81	8.25	1.05	3.07	1.94	2.63	0.22	0.13	0.16	0.19	0.40	0.70
Narrabri (W2)	0-20cm	1.74	14.78	1.23	3.25	3.37	3.04	0.21	0.23	0.18	0.20	0.90	0.78
Narrabri (W2)	20-40cm	1.65	15.35	0.79	2.54	2.83	2.05	0.04	0.05	0.03	0.25	0.94	0.67
Narrabri (W2)	20-40cm	1.60	16.40	0.81	2.55	3.05	2.20	0.04	0.05	0.03	0.25	1.06	0.52
Nowley (L Paddock)	0-20cm	1.02	9.04	0.70	0.34	0.94	0.31	0.11	0.12	0.09	0.05	0.91	0.45
Nowley (L Paddock)	0-20cm	0.97	9.36	0.67	0.39	0.95	0.32	0.10	0.13	0.10	0.12	0.99	0.39
Nowley (L Paddock)	20-40cm	1.34	61.19	0.02	2.73	10.26	0.60	0.00	0.03	0.00	1.27	2.96	0.22
Nowley (L Paddock)	20-40cm	1.30	59.04	0.03	2.74	9.81	0.69	0.00	0.03	0.00	1.21	2.74	0.41
Kiama	0-20cm	10.19	24.43	2.78	3.34	2.78	1.03	0.65	0.72	0.40	1.64	2.38	0.84
Kiama	0-20cm	10.57	23.00	2.89	3.33	2.73	2.38	0.67	0.70	0.43	1.74	2.36	0.89
Kiama	20-40cm	11.17	29.55	2.03	3.28	3.32	0.85	0.56	0.68	0.24	1.48	2.65	0.70
Kiama	20-40cm	10.10	28.62	1.97	3.23	3.10	0.82	0.54	0.65	0.24	1.52	2.40	0.84
Braidwood	0-20cm	1.76	21.33	0.45	1.25	3.29	0.64	0.04	0.06	0.01	0.47	1.48	0.48
Braidwood	0-20cm	1.82	17.08	0.46	1.32	2.82	0.66	0.03	0.05	0.01	0.47	1.30	0.42
Braidwood	20-40cm	1.32	8.96	0.75	1.48	1.70	0.75	0.10	0.07	0.04	0.47	0.74	0.50
Braidwood	20-40cm	1.38	9.16	0.84	1.53	1.75	0.74	0.10	0.07	0.04	0.40	0.68	0.47
Cooma 1	0-20cm	4.02	26.73	0.68	3.96	3.77	1.18	0.23	0.23	0.13	1.08	1.89	0.65



Cooma 1	0-20cm	4.09	22.56	0.68	3.93	3.13	1.11	0.23	0.19	0.14	1.21	1.45	0.63
Cooma 1	20-40cm	3.14	17.80	0.40	4.04	3.48	1.30	0.02	0.02	0.01	1.02	1.36	0.48
Cooma 1	20-40cm	3.18	18.74	0.39	4.10	3.67	1.35	0.01	0.02	0.01	1.06	1.38	0.55
Cooma 2	0-20cm	7.54	19.32	0.63	3.03	1.62	0.88	0.60	0.39	0.24	1.97	1.27	0.55
Cooma 2	0-20cm	7.04	18.02	0.62	2.94	1.46	0.85	0.59	0.36	0.24	1.99	1.50	0.68
Cooma 2	20-40cm	8.33	17.37	0.50	4.14	1.67	0.94	0.62	0.35	0.13	2.56	1.44	0.55
Cooma 2	20-40cm	8.02	21.04	0.50	3.76	2.24	1.02	0.59	0.43	0.13	2.40	1.73	0.67
Coolringdon	0-20cm	4.19	21.00	0.90	1.55	2.47	0.69	1.40	1.84	0.43	0.40	2.78	0.75
Coolringdon	0-20cm	4.13	20.09	1.01	1.45	2.31	0.81	1.46	1.78	0.45	0.42	2.69	0.85
Coolringdon	20-40cm	4.99	21.04	1.09	3.32	3.08	0.96	0.46	0.32	0.07	0.98	1.56	0.39
Coolringdon	20-40cm	5.02	22.87	1.08	3.19	3.47	0.97	0.44	0.35	0.07	0.90	1.81	0.51
Batlow	0-20cm	3.35	11.29	2.04	3.02	2.56	1.97	0.35	0.27	0.31	0.45	0.76	0.99
Batlow	0-20cm	3.32	10.25	2.03	3.12	2.37	2.07	0.35	0.24	0.32	0.43	0.83	0.54
Batlow	20-40cm	4.04	16.76	1.06	4.84	3.73	2.25	0.14	0.11	0.08	0.76	1.08	0.61
Batlow	20-40cm	4.11	16.06	1.01	4.74	3.46	2.08	0.14	0.10	0.07	0.85	0.96	0.59
Tumbarumba	0-20cm	7.75	42.34	0.74	7.60	5.77	2.65	0.05	0.08	0.02	1.17	1.00	0.34
Tumbarumba	0-20cm	7.63	41.07	0.67	7.15	5.52	2.66	0.05	0.08	0.02	1.10	0.93	0.36
Tumbarumba	20-40cm	7.37	38.90	0.20	7.74	6.09	2.44	0.02	0.06	0.00	1.38	1.03	0.36
Tumbarumba	20-40cm	7.38	39.19	0.18	7.58	5.93	2.43	0.02	0.05	0.00	1.33	1.03	0.30
Wagga Wagga	0-20cm	1.71	12.61	0.43	1.58	1.26	0.88	0.30	0.20	0.22	0.41	1.17	0.78
Wagga Wagga	0-20cm	1.78	13.05	0.46	1.65	1.25	0.99	0.31	0.20	0.25	0.40	1.20	0.94
Wagga Wagga	20-40cm	1.57	17.32	0.20	1.89	1.51	0.77	0.19	0.14	0.12	0.59	1.41	0.97
Wagga Wagga	20-40cm	1.61	18.54	0.16	1.91	1.63	0.71	0.19	0.15	0.10	0.65	1.49	0.86
Sandigo (Narrandera)	0-20cm	2.19	13.59	0.67	2.14	1.49	1.35	0.16	0.12	0.15	0.49	1.40	0.96
Sandigo (Narrandera)	0-20cm	1.82	12.99	0.65	1.77	1.41	1.34	0.13	0.12	0.14	0.38	1.31	1.05
Sandigo (Narrandera)	20-40cm	2.82	19.01	0.82	3.70	3.05	2.27	0.03	0.04	0.03	0.56	1.09	0.79
Sandigo (Narrandera)	20-40cm	2.92	17.24	0.83	3.77	2.86	2.18	0.03	0.03	0.03	0.61	1.09	0.80
Boree Creek	0-20cm	1.96	7.53	1.01	1.38	0.81	1.15	0.43	0.25	0.26	0.29	1.22	1.42
Boree Creek	0-20cm	2.01	7.75	0.60	1.40	0.78	0.69	0.45	0.25	0.16	0.34	1.17	0.94
Boree Creek	20-40cm	1.25	18.55	0.48	1.15	1.49	0.70	0.29	0.32	0.09	0.65	4.07	1.73



Boree Creek	20-40cm	1.23	17.39	0.51	1.20	1.51	0.73	0.31	0.31	0.09	0.61	4.35	1.70
Grabben Gullen	0-20cm	2.85	29.77	0.91	3.06	4.67	1.85	0.08	0.08	0.05	0.46	2.09	0.86
Grabben Gullen	0-20cm	2.85	28.55	0.89	2.99	4.61	1.90	0.08	0.08	0.04	0.46	2.04	0.83
Grabben Gullen	20-40cm	2.61	31.90	0.89	3.18	4.92	2.09	0.06	0.06	0.03	0.38	1.89	0.61
Grabben Gullen	20-40cm	2.51	34.55	0.79	3.02	5.51	1.94	0.06	0.06	0.03	0.38	2.15	0.58
Crookwell	0-20cm	2.99	16.74	1.70	3.44	3.62	2.44	0.03	0.03	0.02	0.31	1.09	0.61
Crookwell	0-20cm	3.01	15.96	1.61	3.36	3.28	2.29	0.03	0.03	0.02	0.33	1.00	0.57
Crookwell	20-40cm	3.43	20.57	1.11	4.03	4.34	2.12	0.01	0.01	0.00	0.56	1.31	0.47
Crookwell	20-40cm	3.38	21.28	1.10	4.05	4.38	2.20	0.01	0.01	0.00	0.62	1.31	0.55
Bathurst 1	0-20cm	0.78	3.42	0.35	0.80	0.38	0.46	0.98	0.57	0.55	0.22	0.52	0.66
Bathurst 1	0-20cm	0.76	3.10	0.35	0.71	0.34	0.47	0.96	0.53	0.57	0.23	0.45	0.69
Bathurst 1	20-40cm	0.73	3.20	0.18	0.82	0.32	0.37	0.64	0.32	0.21	0.29	0.43	0.64
Bathurst 1	20-40cm	0.68	3.09	0.19	0.73	0.28	0.37	0.59	0.30	0.20	0.30	0.42	0.61
Bathurst 2	0-20cm	1.20	2.94	1.04	0.92	0.58	0.88	0.35	0.19	0.33	0.12	0.52	0.90
Bathurst 2	0-20cm	1.14	2.92	1.00	0.83	0.55	0.88	0.33	0.18	0.31	0.09	0.51	0.78
Bathurst 2	20-40cm	1.04	4.94	0.45	1.53	1.10	0.60	0.05	0.04	0.03	0.30	0.92	0.52
Bathurst 2	20-40cm	0.98	5.29	0.41	1.42	1.21	0.60	0.05	0.04	0.03	0.28	1.07	0.49
Molong 1	0-20cm	3.34	21.73	0.58	1.34	1.67	0.55	1.75	1.25	0.52	0.47	1.51	0.76
Molong 1	0-20cm	3.54	19.85	0.57	1.37	1.67	0.60	1.91	1.18	0.52	0.57	1.55	0.78
Molong 1	20-40cm	4.84	21.00	0.63	2.96	2.40	0.85	2.00	1.27	0.33	1.17	2.49	0.81
Molong 1	20-40cm	5.00	20.14	0.58	2.99	2.51	0.80	2.01	1.21	0.33	1.17	2.38	0.82
Molong 2	0-20cm	5.78	32.23	0.62	3.18	2.38	0.80	0.90	0.59	0.28	1.36	2.36	0.99
Molong 2	0-20cm	5.73	33.00	0.54	3.22	2.43	0.74	0.90	0.61	0.25	1.35	2.57	0.80
Molong 2	20-40cm	5.91	36.40	0.46	4.06	3.19	0.85	0.48	0.33	0.13	1.30	2.00	0.59
Molong 2	20-40cm	5.82	37.69	0.53	4.05	3.24	0.89	0.48	0.35	0.14	1.36	2.08	0.70
Wellington	0-20cm	2.72	24.79	0.20	3.20	1.71	0.65	0.60	0.37	0.24	1.74	2.37	1.33
Wellington	0-20cm	2.77	26.64	0.19	3.34	1.90	0.70	0.62	0.41	0.24	1.76	2.52	1.15
Wellington	20-40cm	2.39	25.43	0.11	1.83	1.47	0.40	1.04	0.63	0.25	1.00	1.89	0.80
Wellington	20-40cm	2.41	29.00	0.11	2.05	1.56	0.40	1.06	0.74	0.25	0.95	2.13	0.84
Dunedoo	0-20cm	3.06	33.07	0.23	2.35	2.60	0.42	1.13	0.87	0.20	0.99	2.97	0.72



Dunedoo	0-20cm	3.21	26.12	0.22	2.35	1.95	0.46	1.13	0.67	0.20	0.89	2.21	0.65
Dunedoo	20-40cm	3.22	24.71	0.16	2.44	2.14	0.42	1.20	0.67	0.16	1.12	2.14	0.52
Dunedoo	20-40cm	3.29	24.61	0.16	2.43	2.02	0.46	1.22	0.66	0.16	1.08	2.03	0.63
Mount Wilson	0-20cm	2.62	14.60	1.70	3.18	3.62	2.24	0.06	0.04	0.04	0.24	0.78	0.47
Mount Wilson	0-20cm	2.91	15.09	1.67	3.41	3.81	2.18	0.06	0.04	0.04	0.34	0.76	0.51
Mount Wilson	20-40cm	1.25	15.90	0.90	2.24	4.25	1.46	0.02	0.03	0.01	0.30	0.71	0.42
Mount Wilson	20-40cm	1.32	16.94	0.87	2.30	4.31	1.46	0.02	0.03	0.01	0.37	0.76	0.48
Mount Tomah	0-20cm	8.61	14.36	4.54	6.46	3.26	5.02	0.25	0.12	0.16	0.94	0.88	1.07
Mount Tomah	0-20cm	8.66	17.71	4.15	6.74	4.00	4.43	0.26	0.14	0.14	0.89	0.96	0.97
Mount Tomah	20-40cm	8.88	16.73	5.63	8.00	5.15	6.31	0.28	0.14	0.14	0.70	0.74	1.03
Mount Tomah	20-40cm	8.71	21.96	5.80	7.73	6.11	6.47	0.26	0.19	0.15	0.67	0.99	1.10
Condobolin	0-20cm	4.02	14.30	0.99	2.09	1.63	0.58	0.22	0.16	0.16	0.84	1.60	0.64
Condobolin	0-20cm	3.60	12.95	0.98	1.84	1.40	0.61	0.21	0.15	0.17	0.84	1.49	0.64
Condobolin	20-40cm	2.90	10.99	0.97	1.64	1.26	0.77	0.03	0.03	0.02	0.36	1.17	0.60
Condobolin	20-40cm	2.96	10.70	1.00	1.66	1.18	0.83	0.03	0.03	0.02	0.43	1.21	0.58
Narromine	0-20cm	2.22	8.58	1.29	1.29	0.89	1.17	0.36	0.19	0.27	0.31	1.38	1.50
Narromine	0-20cm	2.11	7.58	1.17	1.20	0.72	1.07	0.34	0.17	0.27	0.26	1.11	1.26
Narromine	20-40cm	1.67	7.66	1.69	1.15	0.83	1.40	0.12	0.07	0.08	0.32	2.36	2.45
Narromine	20-40cm	1.64	7.91	1.43	1.11	0.89	1.16	0.11	0.07	0.08	0.25	2.56	1.99
Coonabarabran	0-20cm	1.74	12.86	0.29	3.48	2.75	1.21	0.08	0.05	0.05	0.72	1.34	0.63
Coonabarabran	0-20cm	1.56	14.10	0.28	3.24	2.91	1.24	0.07	0.06	0.05	0.58	1.46	0.78
Coonabarabran	20-40cm	1.49	11.04	0.27	3.54	2.79	1.25	0.05	0.04	0.04	0.67	1.33	0.63
Coonabarabran	20-40cm	1.38	12.58	0.24	3.45	3.06	1.14	0.05	0.05	0.03	0.65	1.52	0.56
Baradine	0-20cm	1.96	20.18	0.42	2.65	3.13	1.32	0.39	0.32	0.24	0.29	1.31	0.35
Baradine	0-20cm	2.10	15.51	0.41	2.76	2.40	1.29	0.41	0.24	0.23	0.27	0.90	0.49
Baradine	20-40cm	1.47	20.53	0.17	2.02	2.66	0.81	0.10	0.10	0.06	0.29	1.18	0.36
Baradine	20-40cm	1.41	16.42	0.19	2.00	2.08	0.93	0.09	0.08	0.06	0.36	0.97	0.38
The Pilliga	0-20cm	3.04	15.25	0.95	1.76	1.58	1.10	1.10	0.65	0.60	0.19	1.27	0.63
The Pilliga	0-20cm	3.16	14.00	1.03	1.81	1.36	1.16	1.17	0.62	0.65	0.24	1.17	0.73
The Pilliga	20-40cm	3.75	18.40	1.04	1.24	1.13	0.79	0.84	0.51	0.36	0.43	1.60	1.01



The Pilliga	20-40cm	3.83	19.10	0.77	1.28	1.25	0.62	0.85	0.53	0.32	0.48	1.62	0.76
Inverell	0-20cm	12.81	56.63	0.85	5.54	5.72	0.95	1.91	1.25	0.47	2.14	3.17	0.77
Inverell	0-20cm	12.77	56.76	0.86	5.66	5.74	0.97	1.99	1.25	0.49	2.16	2.96	0.82
Inverell	20-40cm	11.13	47.40	0.56	6.34	5.28	0.92	1.76	1.10	0.18	2.92	3.92	1.14
Inverell	20-40cm	11.04	51.45	0.65	6.03	5.75	1.05	1.62	1.19	0.20	3.03	4.17	1.35
Tabulam	0-20cm	4.85	60.40	0.91	4.66	6.64	2.99	0.02	0.04	0.01	0.49	2.60	0.42
Tabulam	0-20cm	4.85	44.82	0.84	4.39	5.07	2.76	0.02	0.03	0.01	0.58	2.06	0.38
Tabulam	20-40cm	2.70	14.93	0.31	4.46	3.08	2.22	0.00	0.01	0.00	0.32	3.24	0.26
Tabulam	20-40cm	2.70	14.64	0.34	4.55	3.04	2.41	0.00	0.01	0.00	0.33	3.24	0.42
Clovass	0-20cm	5.53	25.60	0.77	6.82	5.12	1.35	0.09	0.08	0.01	1.32	2.76	0.58
Clovass	0-20cm	5.53	23.66	0.77	6.67	4.52	1.39	0.09	0.07	0.01	1.25	2.36	0.46
Clovass	20-40cm	5.96	35.42	0.59	6.54	4.86	1.31	0.10	0.10	0.01	1.50	2.91	0.60
Clovass	20-40cm	5.93	33.16	0.61	6.69	4.74	1.31	0.10	0.09	0.01	1.48	2.75	0.56
Casino	0-20cm	4.22	33.67	1.09	5.66	6.41	2.96	0.03	0.04	0.01	0.53	1.55	0.33
Casino	0-20cm	4.46	32.72	1.14	5.98	6.34	3.04	0.03	0.04	0.01	0.61	1.55	0.45
Casino	20-40cm	5.07	29.93	0.28	6.17	4.16	3.08	0.01	0.01	0.00	0.51	1.93	0.33
Casino	20-40cm	5.18	41.50	0.31	6.32	5.97	3.26	0.01	0.02	0.00	0.55	2.65	0.33
Guyra	0-20cm	10.39	71.71	1.99	4.64	6.60	1.69	2.90	1.48	1.07	1.06	2.08	0.71
Guyra	0-20cm	9.45	74.04	1.84	4.26	7.03	1.53	2.66	1.57	0.97	1.01	2.07	0.80
Guyra	20-40cm	8.00	79.87	0.73	5.39	9.80	1.14	1.14	0.91	0.26	1.07	2.04	0.35
Guyra	20-40cm	7.93	72.09	0.74	5.13	9.35	1.28	1.13	0.84	0.27	0.99	1.75	0.39
Armidale	0-20cm	5.51	59.70	0.65	4.79	6.08	1.26	0.95	0.64	0.44	1.02	1.99	0.61
Armidale	0-20cm	5.50	61.44	0.71	5.06	6.26	1.29	0.99	0.65	0.46	1.07	1.95	0.64
Armidale	20-40cm	5.87	62.08	0.72	3.78	5.78	1.06	1.04	0.71	0.47	0.63	1.61	0.42
Armidale	20-40cm	6.14	66.11	0.70	4.11	6.24	0.99	1.13	0.77	0.45	0.72	1.65	0.41
Murrurundi	0-20cm	2.68	20.00	0.30	1.78	1.97	0.38	0.08	0.08	0.03	0.57	1.90	0.39
Murrurundi	0-20cm	2.65	22.95	0.30	1.71	2.37	0.34	0.08	0.09	0.03	0.54	2.36	0.38
Murrurundi	20-40cm	2.25	22.58	0.17	1.60	2.28	0.27	0.13	0.16	0.03	0.79	3.13	0.59
Murrurundi	20-40cm	2.22	18.09	0.20	1.59	1.79	0.33	0.12	0.13	0.03	0.70	2.47	0.61
Somersby	0-20cm	1.90	11.52	1.41	3.05	3.13	2.28	0.02	0.01	0.01	0.08	0.41	0.46



Somersby	0-20cm	1.66	8.73	1.42	2.59	2.29	2.30	0.01	0.01	0.01	0.11	0.27	0.45
Somersby	20-40cm	0.91	11.55	0.90	1.98	3.02	1.87	0.00	0.01	0.00	0.15	0.28	0.32
Somersby	20-40cm	0.93	12.04	0.87	2.02	3.23	1.83	0.00	0.01	0.00	0.12	0.30	0.30



767 Table A2. Dataset of extractable carbon (C) used for this study

Location	Depth	C _{OX} (g/kg)	C _{DCB} (g/kg)	C _{PP} (g/kg)
Westwood (Cobbitty)	0-20cm	21.20	24.50	26.06
Westwood (Cobbitty)	20-40cm	26.48	27.46	30.52
Landsdowne (Cobbitty)	0-20cm	26.35	28.42	31.01
Landsdowne (Cobbitty)	20-40cm	24.52	26.62	29.01
Narrabri (W2)	0-20cm	10.57	20.87	21.44
Narrabri (W2)	20-40cm	6.17	8.74	13.08
Nowley (L Paddock)	0-20cm	2.26	2.85	3.32
Nowley (L Paddock)	20-40cm	4.94	3.38	7.67
Kiama	0-20cm	10.06	27.74	29.88
Kiama	20-40cm	13.82	19.23	23.66
Braidwood	0-20cm	23.19	23.50	23.05
Braidwood	20-40cm	21.28	21.78	22.89
Cooma 1	0-20cm	17.39	17.64	23.16
Cooma 1	20-40cm	17.04	16.11	19.77
Cooma 2	0-20cm	9.65	13.00	15.74
Cooma 2	20-40cm	7.25	11.32	13.43
Coolringdon	0-20cm	7.26	10.46	14.24
Coolringdon	20-40cm	5.39	7.63	12.71
Batlow	0-20cm	5.68	6.23	9.93
Batlow	20-40cm	2.10	4.99	6.99
Tumbarumba	0-20cm	6.25	10.48	14.18
Tumbarumba	20-40cm	9.83	13.81	15.90
Wagga Wagga	0-20cm	1.60	4.39	8.01
Wagga Wagga	20-40cm	1.66	5.30	7.27
Sandigo (Narrandera)	0-20cm	1.16	4.48	9.31
Sandigo (Narrandera)	20-40cm	1.48	2.84	7.89
Boree Creek	0-20cm	0.92	3.27	7.43
Boree Creek	20-40cm	5.67	7.27	10.50
Grabben Gullen	0-20cm	2.32	4.68	10.24
Grabben Gullen	20-40cm	2.29	3.93	8.72
Crookwell	0-20cm	1.83	5.78	9.82
Crookwell	20-40cm	2.41	4.33	6.94
Bathurst 1	0-20cm	0.66	1.46	3.63
Bathurst 1	20-40cm	0.82	1.46	3.11
Bathurst 2	0-20cm	2.81	5.62	8.86
Bathurst 2	20-40cm	5.15	6.62	7.89
Molong 1	0-20cm	1.18	3.56	6.86
Molong 1	20-40cm	0.71	3.25	7.99
Molong 2	0-20cm	1.05	2.96	8.24
Molong 2	20-40cm	1.87	7.55	12.11
Wellington	0-20cm	10.83	14.27	18.59
Wellington	20-40cm	7.95	12.64	18.27
Dunedoo	0-20cm	12.46	15.96	20.15
Dunedoo	20-40cm	11.35	15.48	19.84



Mount Wilson	0-20cm	6.75	9.04	13.48
Mount Wilson	20-40cm	9.73	10.36	12.22
Mount Tomah	0-20cm	9.04	15.96	20.42
Mount Tomah	20-40cm	8.39	12.24	20.98
Condobolin	0-20cm	10.66	14.57	19.40
Condobolin	20-40cm	10.05	12.81	15.47
Narromine	0-20cm	6.83	10.54	14.77
Narromine	20-40cm	4.12	6.46	9.60
Coonabarabran	0-20cm	1.48	4.28	8.10
Coonabarabran	20-40cm	1.91	3.38	6.92
Baradine	0-20cm	0.77	3.10	7.37
Baradine	20-40cm	1.52	3.89	5.09
The Pilliga	0-20cm	1.91	4.43	8.80
The Pilliga	20-40cm	1.37	3.18	4.94
Inverell	0-20cm	1.97	8.98	13.46
Inverell	20-40cm	3.00	6.89	11.64
Tabulam	0-20cm	5.32	8.80	11.43
Tabulam	20-40cm	3.72	5.15	8.74
Clovass	0-20cm	3.91	3.48	9.65
Clovass	20-40cm	1.01	4.94	8.65
Casino	0-20cm	2.73	5.03	8.50
Casino	20-40cm	1.79	4.05	7.46
Guyra	0-20cm	3.08	6.72	12.32
Guyra	20-40cm	1.43	4.84	8.67
Armidale	0-20cm	3.08	3.05	10.21
Armidale	20-40cm	1.18	2.18	8.31
Murrurundi	0-20cm	1.84	0.60	5.19
Murrurundi	20-40cm	1.59	0.48	4.26
Somersby	0-20cm	2.31	10.62	7.16
Somersby	20-40cm	0.52	3.56	2.76

768

769 Author contribution

770 BEA: Conceptualization, Data curation, Formal analysis, Investigation, Methodology, Software,
 771 Validation, Project administration, Visualization and Writing (original draft preparation). FAD and
 772 CW: Supervision, Funding acquisition and writing (review and editing). BS: Conceptualization,
 773 Funding acquisition, Methodology, Supervision and Writing (review and editing).

774

775 Data availability

776 The raw data used for this study is available in Tables A1 and A2 in the appendix.

777

778 Competing interests

779 The authors declare that they have no conflict of interest.



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