



1 **Macro constraints and local eutrophication shape sediment nitrogen**

2 **removal in the Eastern China Plain lakes**

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11 **Abstract:**

12 Eutrophication poses a profound threat to the provisioning of ecosystem services in lakes worldwide.  
13 Understanding how eutrophication, alongside other biotic and abiotic factors, drives sediment  
14 nitrogen removal is essential for maintaining the intrinsic purification capacity of ecosystems and  
15 safeguarding water security. To address this, we conducted a systematic investigation of 17  
16 representative lakes across the Eastern China Plain using  $^{15}\text{N}$  stable isotope tracing, 16S rRNA  
17 amplicon sequencing, and metagenomics profiling. This study aims to quantify the contribution of  
18 the anammox to nitrogen removal, elucidate the impacts of eutrophication on overall nitrogen  
19 removal, and identify the critical biotic and abiotic factors driving these processes. Results revealed  
20 that while denitrification was the dominant process across all lakes, anammox made substantial  
21 contributions of up to 34.3%, exhibiting a distinct environmental dependency. Eutrophication  
22 significantly amplified both the overall nitrogen removal capacity and the relative contribution of  
23 anammox, with local physicochemical properties and microbial communities demonstrating a  
24 pronounced joint driving effect on these processes. Notably, the observed decoupling among  
25 microbial community structure, functional gene abundance, and nitrogen removal rates highlights a  
26 significant discrepancy between genetic potential and active expression. This discrepancy is further  
27 manifested by the superior environmental adaptability of functionally redundant denitrifying  
28 bacteria compared to narrow ecological niche anammox bacteria. Furthermore, we demonstrate that  
29 macro scale factors, including spatial, climatic, and anthropogenic, indirectly drive nitrogen removal  
30 by reshaping local water-sediment physicochemical properties, rather than acting as direct  
31 determinants of process rates. Collectively, this study enriches our fundamental understanding of  
32 nitrogen removal, particularly the anammox process, and its multidimensional drivers in the Eastern  
33 China Plain lakes. We advocate that future lake management strategies must rigorously account for  
34 the overarching constraints jointly imposed by trophic status, alongside spatial and climatic  
35 gradients, to ensure the sustainability of nitrogen purification ecosystem services.

36 **Key Words:** Denitrification; Anammox; Functional redundancy; Eutrophication; Shallow lakes



## 37 **1 Introduction**

38 Lakes (including reservoirs) constitute critical components of the Earth's surface system, playing  
39 irreplaceable roles in regulating global biogeochemical cycles and energy fluxes, as well as  
40 maintaining regional ecological security (Mendonça et al., 2017; Woolway et al., 2025). However,  
41 since the Industrial Revolution, the dual stressors of climate change and intense anthropogenic  
42 activities have severely disrupted the homeostasis of lake ecosystems, precipitating a series of global  
43 lake environmental crises (Woolway et al., 2020; IPCC, 2022). Statistics indicate that algal blooms  
44 occur in over 20,000 lakes worldwide, covering approximately 2.57% of the total global lake surface  
45 area, with a continuing upward trend in both frequency and intensity (Ho et al., 2019; Hou et al.,  
46 2022). The synergistic between climate warming and anthropogenic nitrogen loading is recognized  
47 as the core mechanism driving the exacerbation of algal blooms (Huisman et al., 2018; Wang et al.,  
48 2025). This not only undermines ecosystem services but also poses severe challenges to global  
49 public health security (Dai et al., 2023).

50 Nitrogen, an essential element for the synthesis of proteins and nucleic acids, undergoes  
51 biogeochemical cycling that directly governs the regulation of primary productivity and the  
52 maintenance of homeostasis in aquatic ecosystems (Stüeken et al., 2024). When exogenous nitrogen  
53 inputs exceed the lake's carrying capacity, disrupted nitrogen cycling can induce ecological disasters  
54 such as algal blooms (Ho et al., 2019; Wang et al., 2025). As a primary limiting factor for algal  
55 blooms, the spatiotemporal distribution of nitrogen directly determines phytoplankton growth  
56 potential (Hutchins and Tagliabue, 2024). The removal of excess reactive nitrogen via microbially  
57 mediated pathways is a critical mechanism for maintaining lake health and self-purification capacity  
58 (Wang et al., 2024). Sediment nitrogen removal processes—primarily denitrification and anaerobic  
59 ammonium oxidation (anammox)—can eliminate 30% to 50% of nitrogen from the water column  
60 (Kuypers et al., 2018). Therefore, elucidating the spatiotemporal evolution and driving mechanisms  
61 of lake nitrogen removal is scientifically critical for controlling eutrophication and mitigating algal  
62 blooms (Helmond et al., 2020; Wu et al., 2025).

63 In China, driven by rapid socioeconomic development, approximately 69.5% of lakes face  
64 eutrophication, a phenomenon particularly pronounced in the densely populated Eastern Plain  
65 Region (Zhou et al., 2022). Ecological degradation resulting from eutrophication causes annual



66 economic losses exceeding \$10 billion, severely threatening the health of China's lake ecosystems  
67 and the realization of national water security strategies (Wang et al., 2025). However, previous  
68 studies in this region have predominantly focused on denitrification or dissimilatory nitrate  
69 reduction to ammonium (DNRA) (Jiang et al., 2023), often overlooking the potential contribution  
70 of anammox. Recent evidence suggests that under specific environmental, the contribution of  
71 anammox to nitrogen removal is non-negligible and may even become dominant (Deng et al., 2024).  
72 Consequently, the lack of systematic assessment of anammox limits our comprehensive  
73 understanding of the nitrogen budget and self-purification mechanisms in the Eastern China Plain  
74 lakes.

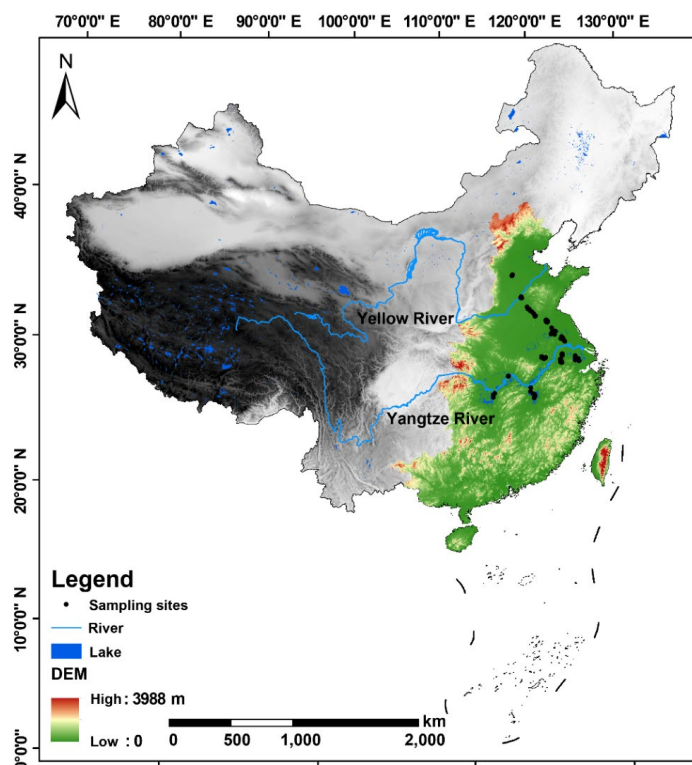
75 Therefore, we hypothesized that the lakes across the Eastern China Plain could serve as hotspots for  
76 the anammox process, and that eutrophication might induce significant shifts in anammox rates and  
77 their relative contribution to nitrogen removal. To test this hypothesis, we investigated 17  
78 representative lakes across the Eastern China Plain, covering both mesotrophic and eutrophic lakes.  
79 By integrating  $^{15}\text{N}$  stable isotope tracing, 16S rRNA amplicon sequencing, and metagenomics, this  
80 study aims to address the following questions: (1) What is the relative contribution of the anammox  
81 process to overall nitrogen removal in these lakes? (2) Does eutrophication significantly alter the  
82 dynamics of denitrification and anammox processes? (3) What are the relative roles of biotic and  
83 abiotic factors in driving these lacustrine nitrogen removal processes?



## 84 2 Materials and Methods

### 85 2.1 Study area and sample collection

86 The Eastern China Plain lakes region (ECPLR), one of China's five primary lake regions,  
87 encompasses the middle and lower reaches of the Yangtze River, the lower Yellow River, the Huai  
88 River, and the Eastern Route of the South-to-North Water Diversion Project. With a total surface  
89 area exceeding 20,000 km<sup>2</sup>, the ECPLR accounts for approximately 25.9% of China's total lake area  
90 (Ma et al., 2010). Seventeen representative lakes within this region were selected for this study (Fig.  
91 1, Tab. S1). These lakes provide critical ecosystem services and possess significant economic value,  
92 including potable water supply, navigation, flood control, fisheries, aquaculture, and biodiversity  
93 conservation, among others. Considering the morphological characteristics, hydrological  
94 connectivity, and historical monitoring data of each lake, a total of 81 sampling sites were  
95 established. Further details regarding the study area are provided in Supplementary Method S1.



96  
97 **Figure 1** Overview of the Eastern China Plain lakes region and distribution of sampling sites

98 Field surveys were conducted from August to October 2024. At each site, 2 L of composite water



99 samples were collected for physicochemical analysis and slurry incubation experiments.  
100 Concurrently, surface sediment (0–5 cm) was collected for physicochemical analysis and slurry  
101 incubation (~500 g), and microbial sequencing (~5 g). All samples were immediately stored at 4°C  
102 in darkness and transported to the laboratory.

## 103 **2.2 Physicochemical analysis, climate data acquisition, and human activity intensity** 104 **quantification**

105 Water temperature (WT), electrical conductivity (EC), dissolved oxygen (DO), and pH were  
106 measured on-site (YSI Professional Plus, USA). Water depth (WD) and secchi depth (SD) were  
107 determined using a Speedtech SM-5 depth sounder and Secchi disk, respectively. Detailed analytical  
108 parameters and methods for water and sediment samples are provided in Tab. S2. Trophic status of  
109 the lake water was evaluated using the Comprehensive Trophic Level Index (TLI) (Supplementary  
110 Method S2).

111 Climate data were obtained from the CRU TS dataset (Harris et al., 2020). Using the "ncdf4"  
112 package in R (v4.5.2), the monthly mean precipitation (MP) and mean temperature (MT) were  
113 extracted for the sampling months. Additionally, annual-scale climate indices (MAP, MAT, etc., Tab.  
114 S3) were calculated. Human Activity Intensity (HAI) was quantified using construction land  
115 equivalent conversion coefficients (Tab. S4) (Xu et al., 2015). Detailed procedures regarding land  
116 use data sources, preprocessing, and calculations are provided in Supplementary Method S3.

## 117 **2.4 Nitrogen removal rate determination**

118 Sediment nitrogen removal processes were quantified using <sup>15</sup>N isotope tracing slurry incubations  
119 (Supplementary Method S4). Dissolved <sup>28</sup>N<sub>2</sub>, <sup>29</sup>N<sub>2</sub>, and <sup>30</sup>N<sub>2</sub> concentrations were measured using  
120 membrane inlet mass spectrometer (MIMS-200, Bay Instruments, USA). Denitrification ( $V_{den}$ ) and  
121 anammox rates ( $V_{ana}$ ), and the relative contribution of anammox to nitrogen removal ( $Ra$ ), were  
122 calculated following Thamdrup and Dalsgaard (2002).

## 123 **2.5 DNA extraction, sequencing, and bioinformatic analyses**

124 Total DNA was extracted using the FastDNA® Spin Kit for Soil (MP Biomedicals, USA). The  
125 integrity of genomic DNA was assessed by agarose gel electrophoresis, while the concentration and  
126 purity were measured using a Nanodrop 2000 and Qubit 3.0 Spectrophotometer. The 16S rRNA  
127 gene V3–V4 region was amplified (primers 338F/806R) (Dai et al., 2022), and sequenced on the



128 Illumina MiSeq platform (Guangdong Magigene Biotechnology Co., Ltd., Guangzhou, China).  
129 Bioinformatic processing included quality trimming, paired-end read merging, and chimera removal.  
130 Amplicon Sequence Variants (ASVs) were resolved based on 100% sequence identity and assigned  
131 taxonomy using the SILVA database (v138.2).  
132 Metagenomic libraries were constructed using the ALFA-SEQ DNA Library Prep Kit and sequenced  
133 on the Illumina NovaSeq 6000 platform (PE150 mode; Guangdong Magigene Biotechnology Co.,  
134 Ltd., Guangzhou, China). Raw reads were quality-filtered and adapter-trimmed using fastp and  
135 Trimmomatic, yielding approximately 12 Gbp of high-quality clean data per sample. Subsequent  
136 bioinformatic analyses encompassed de novo assembly, Open Reading Frame (ORF) prediction,  
137 construction of a non-redundant gene catalog (unigenes), and gene abundance quantification  
138 (Steinegger and Söding, 2018). Nitrogen-cycling genes and their taxonomic affiliations were  
139 annotated using Diamond (v0.9.32) by aligning sequences against the NCycDB (Tu et al., 2019)  
140 and the NCBI non-redundant (NR) protein database, respectively. Taxonomic classification was  
141 determined using the Lowest Common Ancestor (LCA) algorithm to generate gene abundance  
142 profiles. Functional annotation was filtered based on an alignment coverage threshold of > 40%.

## 143 **2.6 Statistical analyses**

144 All statistical analyses and data visualization were primarily performed using the R (v4.5.2) and  
145 ArcMap 10.4. Microbial diversity was calculated using the 'vegan' package based on a normalized  
146 sequencing depth of 34,235 reads per sample. Linear Discriminant Analysis Effect Size (LEfSe)  
147 was conducted using the 'microeco' package to identify discriminative taxa between groups (Liu et  
148 al., 2022). To quantify spatial structures, Principal Coordinates of Neighbor Matrices (PCNM)  
149 analysis was performed using the 'SoDA' package to extract spatial eigenvectors of the sampling  
150 sites (Borcard and Legendre, 2002). To disentangle the complex relationships driving nitrogen  
151 removal, a multi-tiered statistical approach was employed: (1) Spearman's correlations and Mantel  
152 tests were performed using the 'linkET' package to evaluate the interconnections among different  
153 factors. (2) Linear Mixed Models (LMMs) were constructed using the 'lme4' package to quantify  
154 the effects of geographic location and human activity intensity on nitrogen removal rates. (3)  
155 Random Forest (RF) regression models were built using the 'randomForest' package to identify key  
156 predictors. (4) Partial Least Squares Path Modeling (PLS-PM) was established using the 'plspm'

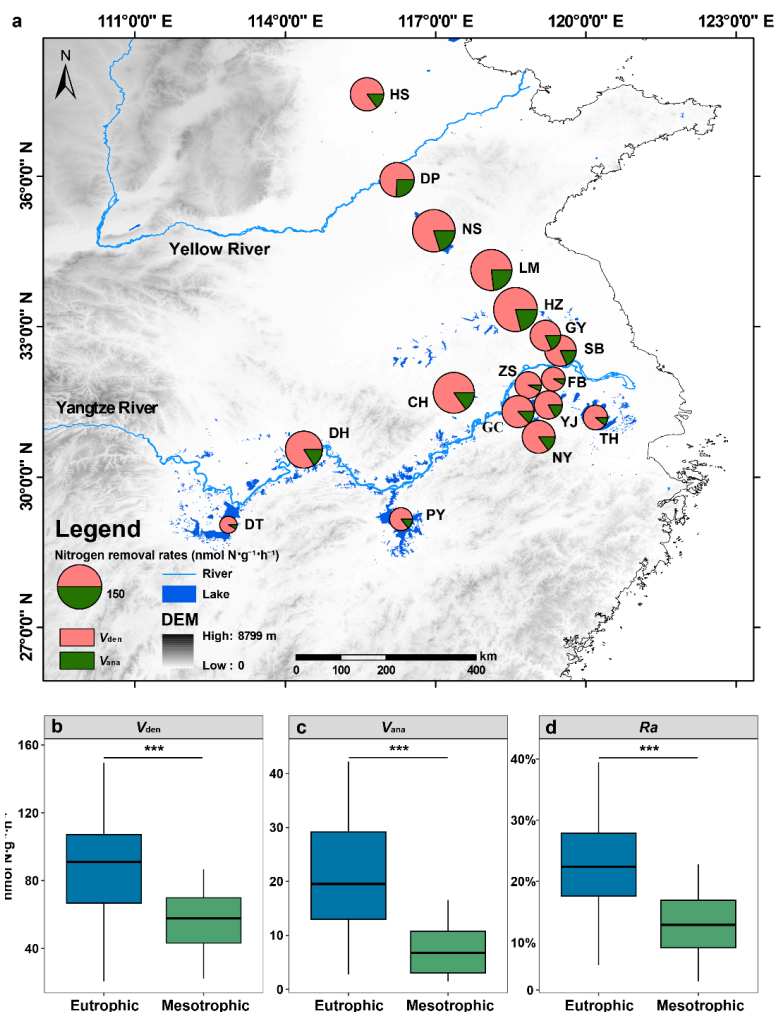


157 package to quantitatively disentangle the direct and indirect causal pathways linking environmental  
158 factors, microbial communities, and functional genes to nitrogen removal rates (Breiman, 2001;  
159 Yuan et al., 2021; Deng et al., 2024).

## 160 **3 Results**

### 161 **3.1 Spatial heterogeneity of nitrogen removal rates**

162 Measurements based on the  $^{15}\text{N}$  isotope pairing technique revealed substantial spatial heterogeneity  
163 in sediment nitrogen removal rates across the Eastern China Plain lakes ( $p < 0.05$ ; Fig. 2a, Tab. S5).  
164 Mean denitrification rates ( $V_{\text{den}}$ ) ranged from 26.58 to 117.54  $\text{nmol N}\cdot\text{g}^{-1}\cdot\text{h}^{-1}$ . Notably, lakes situated  
165 along the Eastern Route of the South-to-North Water Diversion Project (ER-SNWDP)—specifically  
166 Lakes Hongze, Chaohu, Nansi, and Luoma—exhibited elevated denitrification activities, with mean  
167  $V_{\text{den}}$  values consistently exceeding 100  $\text{nmol N g}$ . In contrast, Lake Dongting ( $26.58 \pm 4.84$   $\text{nmol}$   
168  $\text{N}\cdot\text{g}^{-1}\cdot\text{h}^{-1}$ ) and certain reservoir-type lakes displayed significantly lower rates. The spatial  
169 distribution pattern of anammox rates ( $V_{\text{ana}}$ ) was highly congruent with that of denitrification, with  
170 mean values ranging from 2.50 to 34.54  $\text{nmol N}\cdot\text{g}^{-1}\cdot\text{h}^{-1}$ . Although  $V_{\text{den}}$  consistently exceeded  $V_{\text{ana}}$ ,  
171 indicating the predominance of denitrification across all studied lakes, the contribution of anammox  
172 was substantial in specific lakes such as Lake Dongping (Tab. S5), where the maximum relative  
173 contribution ( $Ra$ ) reached 34.34%. Sampling sites were categorized into eutrophic and mesotrophic  
174 groups based on the TLI assessment (Fig. S1o). Comparative analysis demonstrated that  $V_{\text{den}}$ ,  $V_{\text{ana}}$ ,  
175 and  $Ra$  were all significantly higher in eutrophic sites compared to mesotrophic ones ( $p < 0.001$ ;  
176 Fig. 2).



177

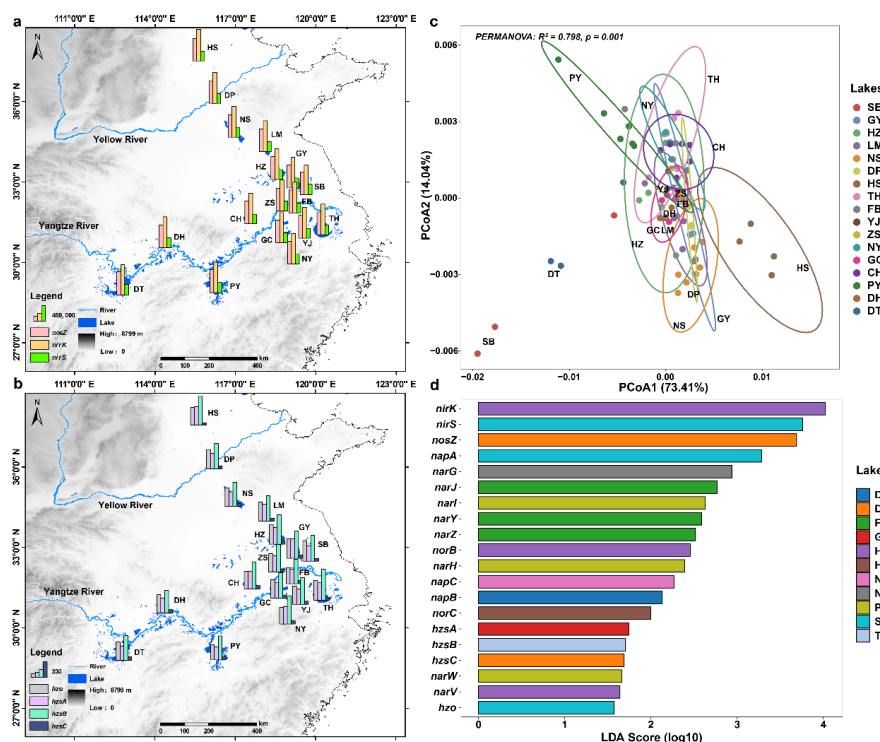
178 **Figure 2** Spatial patterns of nitrogen removal rates in the Eastern China Plain lakes (a) and their  
 179 variations across sites with different trophic states (b–d). Note: The size of the pie charts is  
 180 proportional to the total nitrogen removal potential (the sum of mean denitrification and anammox  
 181 rates). Statistical significance was determined using the Mann-Whitney U test with Benjamini-  
 182 Hochberg correction, indicated by asterisks (\*\*\*:  $p < 0.001$ ).

### 183 3.2 Spatial patterns of nitrogen removal functional gene abundance and diversity

184 Metagenomic analysis revealed that the key functional genes mediating denitrification and  
 185 anammox were ubiquitous across all studied lakes, yet their abundances varied substantially (16–  
 186 910,554 reads; Fig. S2, Tab. S6). Denitrification genes overwhelmingly dominated the functional



187 gene pool, with a mean abundance of  $1.98 \times 10^6$  reads, approximately three orders of magnitude  
 188 higher than anammox genes ( $1.33 \times 10^3$  reads). Spatially, the major denitrification genes maintained  
 189 a consistent abundance hierarchy (*nirK* > *nosZ* > *nirS*) across all lakes (Fig. 3a). For major anammox  
 190 genes, *hzsB* exhibited the highest abundance (Fig. 3b). No significant inter-lake differences were  
 191 found for *hzo*, *hzsB*, and *hzsC* ( $p > 0.05$ ). Conversely, *hzsA* displayed distinct spatial heterogeneity,  
 192 with significantly higher abundance in Lake Gaoyou compared to Lakes Poyang, Dongting, and the  
 193 reservoir-type lakes ( $p < 0.05$ ).



194  
 195 **Figure 3** Spatial distribution of functional genes associated with sediment denitrification (a) and  
 196 anammox (b) in the Eastern China Plain lakes. (c) Beta-diversity of functional gene profiles  
 197 visualized via Principal Coordinates Analysis (PCoA). (d) Linear Discriminant Analysis Effect Size  
 198 (LEfSe) identifying discriminative functional genes between groups.  
 199 Beta-diversity elucidated the spatial patterns of the nitrogen removal functional gene (Fig. 3c).  
 200 Sampling sites overlapped extensively in the two-dimensional ordination space, with the first  
 201 principal coordinate axis (PCoA1) explaining the majority of the variance (73.41%). However,



202 PERMANOVA tests ( $R^2 = 0.798$ ,  $p = 0.001$ ) indicated that significant differences in functional gene  
203 profiles did exist among lakes. Combined with similarity matrices and hierarchical clustering (Fig.  
204 S3), these differences were primarily driven by specific sites within Lakes Shaobo, Dongting, and  
205 Hengshui. The generally low clustering coefficient ( $< 0.03$ ) suggest that, with the exception of a  
206 few specific habitats, the functional gene composition of sediments was highly similar across most  
207 lakes. LEfSe analysis further identified denitrification genes as the decisive discriminative features  
208 driving the variation in functional gene profiles among lakes (Fig. 3d). These genes consistently  
209 showed LDA scores ( $\log_{10}$ ) exceeding 3 and were significantly enriched in Lakes Hengshui,  
210 Shaobo, and Dongting ( $p < 0.01$ ). In comparison, although anammox genes were enriched in Lakes  
211 Shaobo, Gaoyou, and Taihu, their LDA scores were generally below 2, indicating a minor  
212 contribution to inter-group differences compared to denitrification genes.

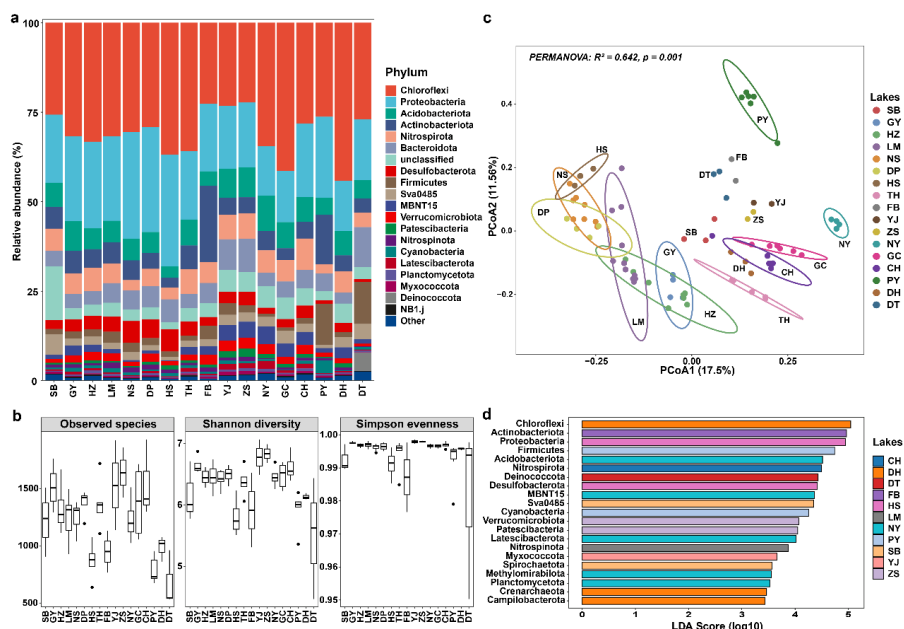
### 213 3.3 Characteristics of microbial community structure

214 16S rRNA sequencing revealed that sediment microbial communities in the study area were  
215 primarily dominated by the phyla Chloroflexi and Proteobacteria, each exceeding 20% in relative  
216 abundance (Fig. 4a). Although taxa such as Desulfobacterota and Firmicutes exhibited lower overall  
217 relative abundances, they showed significant enrichment in specific lakes, suggesting that regional  
218 environmental strongly influence the proliferation of specific taxa. Cluster analysis indicated that  
219 lakes along the ER-SNWD (e.g., Lakes Hongze and Gaoyou) grouped together, reflecting high  
220 community similarity (Fig. S4). In contrast, Lakes Poyang and Dongting formed distinct,  
221 independent branches, highlighting the profound influence of regional environmental.

222 Alpha-diversity (Fig. 4b) indicated that Lakes Gaoyou, Gucheng, and the reservoir-type lakes  
223 possessed high species richness (Observed species  $> 1000$ ) and Shannon diversity ( $> 6$ ). The  
224 Simpson evenness was generally above 0.95, indicating a high overall community evenness.  
225 Conversely, Lake Dongting exhibited relatively lower indices, suggesting a stronger dominance  
226 effect by a few specific taxa. This implies that strong environmental filtering within these lakes has  
227 selectively enriched taxa equipped with specific adaptive traits. Beta-diversity revealed highly  
228 significant differences in community structure among lakes ( $R^2 = 0.642$ ,  $p = 0.001$ ). Sampling sites  
229 appeared highly dispersed in the two-dimensional ordination space (Fig. 4c). Most lake sites formed  
230 independent branches with clustering coefficient generally exceeding 0.4 (Fig. S5), indicative of



231 substantial inter-lake variation. LEfSe analysis further identified specific microbial taxa  
 232 significantly enriched in different lakes, with marked variations in the number of discriminative taxa  
 233 and LDA scores (Fig. 4d). Collectively, these findings underscore the critical role of regional  
 234 environmental in driving microbial community.



235 **Figure 4** Sediment microbial community characteristics in the Eastern China Plain lakes. (a)  
 236 Taxonomic composition of the top 20 phyla. (b) Alpha-diversity indices. (c) Microbial beta-diversity  
 237 based on PCoA. (d) LEfSe analysis identifying discriminative taxa at the phylum level between  
 238 groups.  
 239

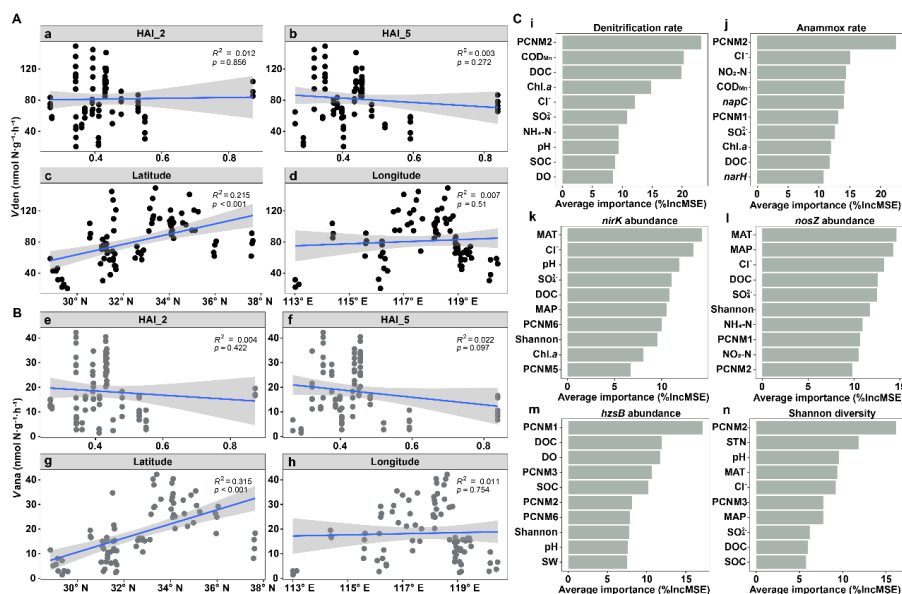
### 240 3.4 Relationships between biotic/abiotic factors and nitrogen removal rates

241 Spearman’s correlations and Mantel tests revealed complex interplay between biotic/abiotic factors  
 242 and nitrogen removal rates (Fig. S6). Both  $V_{den}$  and  $V_{ana}$  rates exhibited significant correlations with  
 243 multiple physicochemical parameters, spatial structural variables (PCNM), and functional genes  
 244 (e.g., *nirS*, *napC*). However,  $V_{ana}$  showed a stronger correlation with microbial community structure  
 245 (Shannon diversity), whereas  $V_{den}$  was more strongly associated with hydro-chemical parameters.  
 246 This divergence suggests that the regulatory mechanisms governing these two nitrogen removal  
 247 processes are not identical.

248 First, linear regression analysis indicated no significant correlation between Human Activity



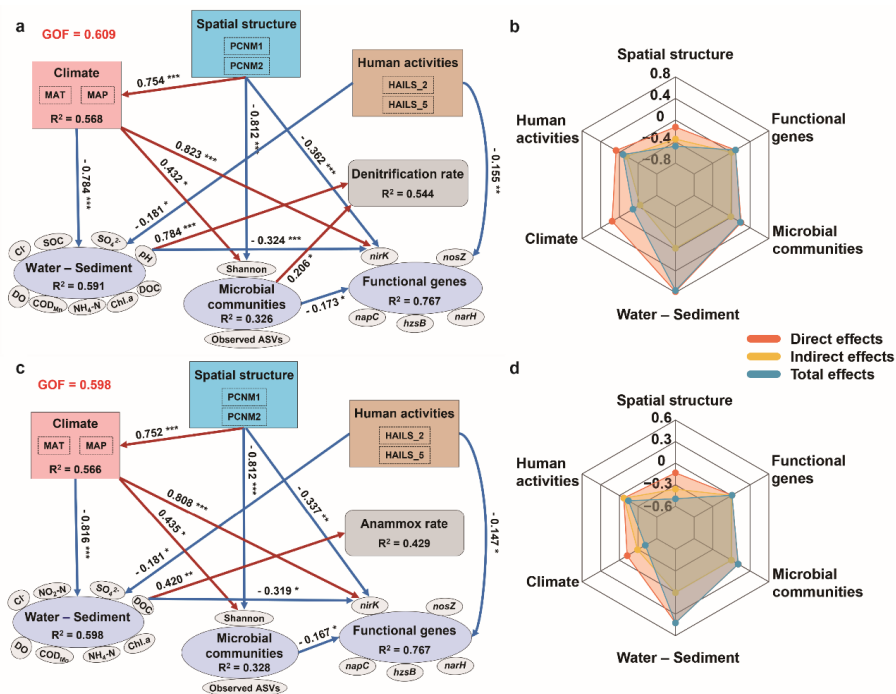
249 Intensity (HAI) and nitrogen removal rates across different spatial scales ( $p > 0.05$ ; Fig. 5A, B). In  
 250 contrast, latitude showed significant positive correlations with both  $V_{den}$  ( $R^2 = 0.215$ ) and  $V_{ana}$  ( $R^2 =$   
 251  $0.315$ ,  $p < 0.001$ ), and anammox may be subject to stricter latitudinal constraints. Random Forest  
 252 (RF) models further identified the spatial structural variable PCNM2 as the shared major predictor  
 253 for both  $V_{den}$  and  $V_{ana}$  (Fig. 5c). Organic matter indicators (COD<sub>Mn</sub> and DOC) were identified as key  
 254 drivers for  $V_{den}$  (Fig. 5i). In contrast,  $V_{ana}$  was more significantly influenced by ionic conditions (Cl<sup>-</sup>)  
 255 and substrate (NO<sub>2</sub>-N) (Fig. 5j). Regarding functional genes, major denitrification genes (*nirK* and  
 256 *nosZ*) were jointly regulated by climatic (MAT and MAP) and ionic conditions (Cl<sup>-</sup> and SO<sub>4</sub><sup>2-</sup>) (Fig.  
 257 5k, l). The key anammox gene (*hzsB*) was primarily driven by PCNM1, DOC, and redox conditions  
 258 (DO) (Fig. 5m). Additionally, variations in Shannon diversity were dominated by PCNM2, with  
 259 sediment trophic status (STN, Fig. S8) serving as another crucial predictor (Fig. 5n). Collectively,  
 260 these results underscore the pivotal role of environmental differences resulting from spatial  
 261 heterogeneity in shaping nitrogen removal processes.



262 **Figure 5** Drivers of nitrogen removal processes in the Eastern China Plain lakes. Linear regressions  
 263 of denitrification (A) and anammox (B) rates against Human Activity Intensity (HAI) and  
 264 geographic location. (C) Key predictors of nitrogen removal rates, functional gene abundance, and  
 265 microbial Shannon diversity identified via Random Forest regression models.  
 266 To disentangle the causal pathways of these multi-factor interactions, a Partial Least Squares Path  
 267



268 Model (PLS-PM) was constructed (Fig. 6). The models exhibited good predictive power, with  
 269 Goodness of Fit (GoF) values of 0.609 for denitrification and 0.598 for anammox, explaining 54.4%  
 270 and 42.9% of the variance in  $V_{den}$  and  $V_{ana}$ , respectively. The results indicated that water-sediment  
 271 physicochemical properties exerted significant direct driving effects on both  $V_{den}$  ( $r = 0.784$ ,  $p <$   
 272  $0.001$ ) and  $V_{ana}$  ( $r = 0.420$ ,  $p < 0.01$ ).  $V_{den}$  was also significantly directly driven by microbial  
 273 community structure ( $r = 0.206$ ,  $p < 0.05$ ). Crucially, effect decomposition highlighted that while  
 274 spatial structure had no significant direct effect, it exerted substantial influence via indirect pathways  
 275 mediated through climate, environmental conditions, and microbial communities (Fig. 6b, d).  
 276 Overall, the two nitrogen removal processes exhibited distinct sensitivities to different drivers,  
 277 highlights the complex and hierarchical regulatory patterns underlying nitrogen removal in the  
 278 Eastern China Plain lakes.



279 **Figure 6** Causal mechanisms driving nitrogen removal in the Eastern China Plain lakes. Partial  
 280 Least Squares Path Modeling (PLS-PM) disentangling the direct and indirect effects of multiple  
 281 drivers on denitrification (a) and anammox (c) rates, alongside their standardized total effects (b, d).  
 282 Note: Arrows represent significant unidirectional causal relationships. Red and blue arrows denote  
 283 positive and negative effects, respectively. Numbers adjacent to arrows indicate standardized path  
 284



285 coefficients (\*:  $p < 0.05$ , \*\*:  $p < 0.01$ , \*\*\*:  $p < 0.001$ ).  $R^2$  values represent the proportion of  
286 variance explained for each dependent variable. Goodness of Fit (GoF) is displayed at the top of  
287 each model.

## 288 **4 Discussion**

### 289 **4.1 Eutrophication enhances the contribution of anammox to nitrogen removal in the Eastern**

#### 290 **China Plain lakes**

291 This study systematically characterized the significant spatial heterogeneity of  $V_{\text{den}}$  and  $V_{\text{ana}}$  in the  
292 Eastern China Plain lakes (Fig. 2a, Tab. S5). Although both processes were ubiquitous across all  
293 surveyed lakes, distinct variations in rates were evident among individual lakes and lake zones,  
294 indicating that lacustrine nitrogen removal capacity is non-uniformly distributed at the regional  
295 scale (Jiang et al., 2023). Furthermore, previous studies have highlighted substantial variability in  
296 the relative contributions of denitrification and anammox (Deng et al., 2023; Stüeken et al., 2024),  
297 our results show that denitrification overwhelmingly dominated in all surveyed lakes (Tab. S5).  
298 However, the contribution of anammox was non-negligible in specific systems (Fig. 2),  
299 underscoring its distinct environmental dependency (Zhu et al., 2013). These findings highlight the  
300 critical regulating role of the local environmental in shaping nitrogen removal processes.

301 Nitrogen and dissolved organic carbon (DOC) serve as critical metabolic substrates for nitrogen-  
302 removing microorganisms and act as key environmental factors of these processes (Liu et al., 2018;  
303 Pascal et al., 2025). Our results demonstrated that both  $V_{\text{den}}$  and  $V_{\text{ana}}$  were significantly elevated in  
304 eutrophic sites compared to mesotrophic ones (Fig. 2b, c). Subjected to prolonged exogenous  
305 nutrient loading, the Eastern China Plain lakes generally sustain high substrate availability (Fig. S1)  
306 (Zhou et al., 2022). This enrichment likely enhances microbial substrate utilization efficiency,  
307 thereby stimulating and accelerating the corresponding nitrogen cycling processes (Zhou et al., 2021;  
308 Jiang et al., 2023). Organic carbon and ionic conditions act as critical electron donors for  
309 denitrification (Yoon et al., 2015). The effective supply of electrons (Fig. S1c–f) directly fueled  
310 elevated  $V_{\text{den}}$  in lakes such as Nansi and Luoma (Fig. 2a, Tab. S5). However, high concentrations of  
311  $\text{SO}_4^{2-}$  may exert an inhibitory effect on denitrification through competitive inhibition for electron  
312 donors or sulfide toxicity (Jiang et al., 2020), which likely accounts for the relatively lower  $V_{\text{den}}$   
313 observed in Lake Hengshui (Fig. S1f).



314 Intriguingly, although denitrification remained dominant under eutrophic conditions, the relative  
315 contribution of anammox significantly increased (Fig. 2d). This shift may be attributed to the  
316 susceptibility of denitrification to factors such as electron donor (organic carbon) allocation and  
317 redox conditions (Carlson et al., 2013). Additionally, given that denitrifying communities possess a  
318 broader tolerance range for oxygen compared to anammox communities (Deng et al., 2020),  
319 processes such as DOC mineralization (Yang et al., 2025) and variations in sediment water content  
320 (Tab. S7) and oxygen permeability (Sun et al., 2023) can alter local oxygen concentrations. This  
321 result creates more intensely anoxic or anaerobic niches, which favor the survival of anammox  
322 bacteria (Yang et al., 2026). This process also stimulates heterotrophic denitrification, which in turn  
323 supplies crucial intermediate substrates  $\text{NO}_2\text{-N}$  for anammox, thereby boosting the contribution of  
324 anammox to lacustrine nitrogen removal (Tsementzi et al., 2016).

325 Collectively, these findings suggest that under specific substrate concentrations and redox  
326 conditions, anammox can form a synergistic or complementary relationship with denitrification.  
327 This complementarity ultimately enhances both the magnitude and stability of the nitrogen removal  
328 function in lake ecosystems (Kuypers et al., 2018), bearing profound ecological significance for  
329 facilitating nitrogen cycling and buffering against environmental changes. As "sinks" within  
330 watershed ecosystems, lakes integrate substantial material and energy inputs (Bergström and  
331 Karlsson, 2019). This results in a complex scenario where nitrogen cycling is simultaneously  
332 regulated by multiple environmental variables, leading to the pronounced spatial heterogeneity  
333 observed in lacustrine nitrogen removal capacity.

#### 334 **4.2 Pronounced decoupling among microbial community, functional gene abundance, and** 335 **nitrogen removal rates**

336 As quintessential microbe-mediated biogeochemical processes, the intensities of denitrification and  
337 anammox are fundamentally governed by microbial community structure and the distribution of  
338 functional genes (Jiang et al., 2020). Our study revealed a certain degree of coupling between  
339 metagenome-derived functional gene abundance/diversity and nitrogen removal rates. Specifically,  
340 key denitrification genes (e.g., *nirK*, *nosZ*) exhibited significant correlations with nitrogen removal  
341 rates (Fig. S6), aligning with previous assertions that gene abundance can serve as a proxy for  
342 nitrogen removal potential (Xiong et al., 2017).



343 However, it is crucial to note that functional gene abundance does not always scale linearly with  
344 nitrogen removal rates (Dandie et al., 2011; Ma et al., 2020). We observed a highly consistent  
345 functional gene composition across the studied lakes, where diversity variations were dictated by  
346 only a few outlier samples (Fig. 3c, S3). This indicates a high similarity in the potential nitrogen  
347 removal capacity among lakes, which stands in sharp contrast to the significant spatial heterogeneity  
348 observed in the measured rates. Furthermore, the substantial discrepancy—approximately three  
349 orders of magnitude—between denitrification and anammox gene abundances (Tab. S5) was not  
350 proportionally reflected in their respective rate differences. This non-linear association likely stems  
351 from the fact that the presence of genes at the DNA level represents only the genetic potential and  
352 does not guarantee their transcription and translation into active proteins (Ma et al., 2020). Path  
353 modeling (PLS-PM), which showed no significant direct effect of functional genes on rates, further  
354 corroborated this phenomenon (Fig. 6a, c). Consequently, assessing regional nitrogen removal  
355 capacity based solely on functional gene abundance may lead to misjudgments of actual rates.

356 Beyond the genetic level, microbial community structure is also a critical driver of spatial  
357 differentiation in nitrogen removal (Fig. 6). Variations in the dominance of functional taxa  
358 fundamentally determine the system's potential to execute complete nitrogen removal processes  
359 (Sveen et al., 2025). Our results showed that core nitrogen-cycling taxa, such as Proteobacteria and  
360 Nitrospirota, exceeded 20% relative abundance in all lakes (Fig. 4a), with specific dominant taxa  
361 significantly enriched in different lakes (Fig. 4d). This reflects a robust potential for nitrogen cycling  
362 and organic matter degradation across the Eastern China Plain lakes (Han et al., 2024). However,  
363 shifts in the proportions of these dominant taxa may alter community response thresholds to  
364 substrates and environmental conditions, thereby driving the spatial divergence of nitrogen removal  
365 rates (Cline and Zak, 2015; Egidi et al., 2023).

366 Notably,  $V_{\text{den}}$  responded more significantly to microbial community structure than  $V_{\text{ana}}$  (Fig. 6a, b).  
367 This distinction may be attributed to the modular nature of denitrification, a stepwise process  
368 performed by a diverse array of microorganisms (Kuypers et al., 2018; Lupon et al., 2026). The  
369 generally high microbial diversity observed in the study area (Fig. 4b, c) implies a high level of  
370 functional redundancy (Louca et al., 2018). This taxonomic diversity allows distinct taxa to perform  
371 identical or similar metabolic functions, enabling denitrifying communities to occupy broader  
372 ecological niches. In contrast, anammox is a specialized, narrow ecological niche process restricted



373 to a limited number of bacteria within the phylum Planctomycetes (Wang et al., 2019), making its  
374 rate strictly constrained by the population size and gene expression of key species. Consequently,  
375 the differential stability of these functional communities in the face of environmental fluctuations  
376 (Hernandez et al., 2023) results in the observed divergence in process rates.

#### 377 **4.3 Multiple driving effects of macro-environmental heterogeneity on nitrogen removal** 378 **processes**

379 Macro-ecological studies consistently demonstrate that nitrogen cycling is significantly correlated  
380 with geographic factors such as latitude, longitude, and altitude (Fuhrman et al., 2008; Tu et al.,  
381 2019). Our study reveals that the latitudinal gradient and spatial structural variables (PCNM2)  
382 constitute the primary macro-scale background regulating nitrogen removal processes in the Eastern  
383 China Plain lakes (Fig. 5, S6). Essentially, latitude serves as a proxy for climatic effects—including  
384 temperature, precipitation, and seasonal irradiance (Ding et al., 2025)—which aligns with the  
385 subtropical and temperate monsoon climate characteristics of the region (Method S1, Tab. S3).  
386 Climate reshapes microbial habitat quality by modulating hydraulic residence time, organic matter  
387 mineralization pathways, and the stratification stability of thermal regimes, dissolved oxygen, and  
388 nutrients (Manzoni et al., 2014; Hu et al., 2020). And subsequently exert profound impacts on  
389 functional gene abundance, highlighting the critical roles of dispersal limitation and spatial  
390 heterogeneity in shaping microbial biogeography (Du et al., 2023). Notably, the response of  $V_{\text{ana}}$  to  
391 latitude was significantly stronger than that of  $V_{\text{den}}$  (Fig. 5c, g). This discrepancy is likely attributable  
392 to the broader metabolic versatility and phylogenetic distribution of denitrifying communities. In  
393 contrast, anammox communities possess a narrower ecological niche specificity and higher  
394 sensitivity to macro-environmental fluctuations, subjecting their spatial distribution to stricter  
395 geographic constraints (Gao et al., 2018).

396 Despite the study area being located in eastern China, a region characterized by intense  
397 anthropogenic disturbance (Fig. S7), the Human Activity Intensity (HAI) across different buffer  
398 zones scales exhibited no significant direct effects on nitrogen removal rates (Fig. 5A, B). This  
399 influence likely underwent a process of environmental filtering followed by redistribution (Keck et  
400 al., 2025). Specifically, while anthropogenic activities drive the input of nitrogen, phosphorus, and  
401 organic matter, thereby altering the trophic state and sediment characteristics (Wei et al., 2024), it



402 is the local water-sediment physicochemical properties that constitute the immediate habitat for  
403 microbial metabolism. These local factors directly regulate functional genes and nitrogen removal  
404 rates through substrate and redox conditions, serving as the primary drivers explaining the spatial  
405 heterogeneity observed in this study (Fig. 6a, c).

406 It is noteworthy that anthropogenic activities and water-sediment physicochemical properties did  
407 not exert significant direct effects on microbial community structure, yet significantly shaped the  
408 distribution of functional genes (Fig. 6). This suggests that microbial communities in the Eastern  
409 China Plain lakes have achieved high structural stability following long-term environmental  
410 filtering. Given the generally high microbial diversity observed (Fig. 4b, c), the system likely  
411 possesses a high level of functional redundancy (Louca et al., 2018). When environmental  
412 physicochemical properties fluctuate due to anthropogenic or other factors, communities may  
413 prioritize regulating functional gene expression to maintain metabolic homeostasis rather than  
414 undergoing large-scale taxonomic restructuring. Consequently, the response of community structure  
415 to environmental changes exhibits a certain hysteresis or dampening effect (Martiny et al., 2023).  
416 Moreover, the robust buffering effects provided by microbial biofilm architectures and community  
417 dispersal processes (Yan et al., 2024) also partially explain why physicochemical fluctuations failed  
418 to translate directly and linearly into shifts in community structure (Kuypers et al., 2018).

## 419 **5 Conclusion**

420 Elucidating the driving mechanisms of sediment nitrogen removal is crucial for understanding the  
421 resilience of lacustrine ecosystems to eutrophication. This study reveals that while denitrification  
422 dominates nitrogen removal in the Eastern China Plain lakes, anammox serves as a critical, highly  
423 environmentally dependent complementary process. Notably, eutrophication significantly amplified  
424 both the overall nitrogen removal capacity and the relative contribution of anammox, with local  
425 physicochemical properties and microbial communities demonstrating a pronounced joint driving  
426 effect on these processes. The observed decoupling among microbial community, functional gene  
427 abundance, and nitrogen removal rates highlights the intricate complexity of microbial responses to  
428 environmental perturbations. This complexity is further evidenced by the stark contrast between the  
429 functional redundancy of denitrifying bacteria and the ecological niche specificity of anammox  
430 bacteria. Furthermore, we demonstrate that macro-scale factors (spatial, climatic, and anthropogenic)



431 indirectly drive nitrogen removal by reshaping local water-sediment physicochemical properties,  
432 rather than acting as direct determinants of process rates. These findings enrich our fundamental  
433 understanding of nitrogen removal, particularly the anammox process, and its multidimensional  
434 drivers in the Eastern China Plain lakes. Given the profound sensitivity of nitrogen removal  
435 processes to habitat conditions, we advocate that future lake management strategies must rigorously  
436 account for the overarching constraints jointly imposed by trophic status, alongside spatial and  
437 climatic gradients, to ensure the sustainability of nitrogen purification ecosystem services.

#### 438 **Data availability**

439 All data analyzed in this study are available in figshare  
440 (<https://doi.org/10.6084/m9.figshare.32005818>).

#### 441 **Author contributions**

442 JHW: conceptualization, sampling, data analysis, manuscript preparation. KY: analyzed the samples  
443 and original draft preparation. YRL: designed figures and tables, original draft preparation. YY:  
444 original draft preparation. LCW: design and guidance of experiments, draft revision. ELZ: draft  
445 revision, funding acquisition; ZJG: design and guidance of experiments, draft revision, funding  
446 acquisition; YJC: design and guidance of experiments, draft revision, funding acquisition.

#### 447 **Competing interests**

448 The authors declare that they have no known competing financial interests or personal relationships  
449 that could have appeared to influence the work reported in this paper.

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