



1 **Reactivity Over Abundance: Unveiling the True Kinetic Drivers of** 2 **Urban Ozone Using Process-Informed Machine Learning**

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13 **Abstract** Ground-level ozone remains a persistent challenge in East Asian urban centers, where concentrations
14 continue rising despite significant reductions in precursor emissions. Designing effective mitigation is complicated by
15 the non-linear relationship between precursor abundance and reactivity. Standard data-driven approaches often suffer
16 from "survivor bias," systematically undervaluing highly reactive precursors that are rapidly depleted. We introduce a
17 process-informed machine learning framework that uses net chemical consumption (ΔVOC) rather than ambient
18 concentrations to resolve this attribution failure. Applied to measurements from a NO_x -saturated roadside site in Seoul,
19 South Korea, the approach reveals a robust quantitative relationship between intrinsic hydroxyl radical reactivity ($k\text{OH}$)
20 and ozone formation sensitivity.

21 Across a comprehensive suite of precursors—including oxygenated VOCs (e.g., formaldehyde, acetone) and
22 aromatics—the framework identifies reactive aromatics (trimethylbenzenes, xylenes) and biogenics (isoprene,
23 monoterpenes) as the dominant kinetic drivers. Whereas static metrics such as OFP rank precursors by stoichiometric
24 capacity under idealized accumulated conditions, the process-informed framework shows that realized ozone
25 production in fresh urban plumes is governed by kinetic turnover rather than abundance, with this kinetic selectivity
26 further amplified during high-ozone episodes. These results indicate that mass-based VOC inventories and OFP-style
27 rankings, when applied without kinetic context, can systematically misallocate control priorities in NO_x -saturated
28 urban regimes. Because the framework requires only sub-hourly co-located VOC and ozone observations and no
29 prescribed mechanism, it offers a complementary empirical pathway in settings where explicit mechanism-based
30 modeling is constrained by incomplete VOC speciation or unmeasured radical precursors.

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34 1 Introduction

35 Ground-level ozone (O_3) remains a persistent environmental and health challenge in East Asian urban centers,
36 continuing to climb despite significant reductions in precursor emissions (Li et al., 2020; Wang et al., 2019). For
37 instance, between 2015 and 2021, South Korea reduced national emissions of nitrogen oxides (NO_x) by 26% and
38 VOCs by approximately 25%, yet urban O_3 levels increased over the same period (Colombi et al., 2023). This
39 divergence is characteristic of NO_x -saturated urban regimes, where non-linear chemistry renders strategies centered
40 on bulk emission mass reductions ineffective (Chang et al., 2020). Addressing this trend requires shifting regulatory
41 focus from aggregate reductions to prioritizing the control of specific, high-impact precursor classes.

42 However, the extreme chemical diversity of atmospheric precursors complicates this targeted approach. Ozone
43 formation efficiency varies dramatically between VOC classes—driven by the high reactivity of alkenes and aromatics
44 compared to less reactive alkanes—and is dependent on the specific VOC mixture in each environment (Atkinson and
45 Arey, 2003). Consequently, the central challenge for modern air quality management is accurately identifying which
46 specific VOCs drive local ozone production to prioritize control efforts effectively (Shi et al., 2024).

47 Current analytical frameworks struggle to bridge atmospheric complexity with policy applicability. Reactivity scales
48 like Maximum Incremental Reactivity (MIR) are useful for ranking but rely on static, idealized conditions that are not
49 inherently designed to capture the dynamics of urban atmospheres (Carter, 2009; Huang et al., 2025). Conversely,
50 explicit Observation-Based Models (OBMs) utilizing explicit mechanisms like the Master Chemical Mechanism
51 (MCM) provide detailed oxidation pathways; however, they are highly sensitive to unmeasured radical precursors and
52 intermediate reaction rates (e.g., RO_2 radical propagation) (Berndt et al., 2018; Huang et al., 2025). Recently, machine
53 learning (ML) has been applied to model the nonlinear ozone dynamics with high predictive accuracy, but its utility
54 for chemical interpretation remains controversial (Zhang et al., 2024). An inherent limitation in conventional ML
55 approaches trained on atmospheric data is their reliance on ambient concentrations as primary features. In atmospheric
56 chemistry, a critical "survivor bias" exists: the most reactive—and thus most important—precursors are rapidly
57 depleted, leaving low ambient concentrations that models statistically undervalue (Huang et al., 2025; Wang et al.,
58 2022). This leads to skewed interpretations where inert accumulators are misidentified as primary drivers while
59 reactive species are ignored, potentially misguiding control policies (Huang et al., 2025).

60 In this study, we introduce, to our knowledge for the first time, a process-informed machine learning framework that
61 uses net VOC consumption (ΔVOC) rather than ambient concentrations as the predictive feature. By embedding the
62 physical constraints of mass conservation and reaction kinetics directly into the learning process, the framework forces
63 the model to relate the extent of VOC processing—rather than standing abundance—to ozone production. As a
64 concrete illustration of the survivor-bias problem, we show that conventional concentration-based ML applied to our
65 Seoul roadside dataset assigns acetone a higher ozone-driver rank than toluene, despite an OBM-MCM radical budget
66 analysis indicating that toluene oxidation contributes an order of magnitude more peroxy radical flux. The ΔVOC
67 framework, in contrast, autonomously recovers from observations alone the kinetic hierarchy expected from
68 fundamental rate theory, including the mechanistic penalties of radical termination (organic nitrate formation,
69 suppressed OH recycling) and the boost from non-photolytic radical generation by monoterpene ozonolysis. Beyond



70 this methodological validation, we report that during high-ozone episodes the kinetic selectivity of the system
71 intensifies rather than the chemical regime shifting—providing a process-level explanation for why peak-ozone
72 reduction strategies based on bulk emission inventories may underperform in fresh-emission urban regimes. Because
73 the framework requires only sub-hourly co-located VOC and ozone observations and no prescribed mechanism, it
74 offers a complementary empirical pathway in environments where explicit mechanism-based modeling is constrained
75 by incomplete VOC speciation, unmeasured radical precursors (e.g., HONO, ClNO₂), or computational cost.

76 **2 Methods**

77 **2.1 Measurement information**

78 The sampling campaign was conducted in Seoul, South Korea, at the Seoul Youth Hostel (37.56°N, 126.99°E).
79 Located in a dense urban center adjacent to a major roadside (Fig. S1), this site was selected to capture the complex
80 mixture of VOCs typical of dense metropolitan environments, ranging from vehicular to commercial sources.
81 Sampling occurred from April 1–30, 2022, a period selected to minimize the interference of biomass burning and
82 maximize the representativeness of urban anthropogenic sources.

83 VOC concentrations were quantified using a proton transfer reaction time-of-flight mass spectrometer (PTR-ToF-MS
84 1000, Ionicon Analytik, Austria), enabling real-time detection of a broad spectrum of oxygenated VOCs (OVOCs),
85 aromatics, and other high-proton-affinity compounds. Inorganic trace gases (NO_x, SO₂, O₃, CO) were monitored using
86 standard gas analyzers (Models 42i, 43i, 49i, and 48i, Thermo Fisher Scientific, USA). Meteorological parameters
87 (temperature, relative humidity, wind speed, and direction) were measured by a collocated automatic weather station
88 (AWS).

89 **2.2 Machine Learning Model**

90 **2.2.1. Curation of input VOC species**

91 To ensure chemical interpretability, input variables were selected from the high-dimensional PTR-MS data through a
92 knowledge-guided curation process. The input feature set comprised recognized ozone precursors based on the MIR
93 scale, including key oxygenated VOCs (OVOCs) such as formaldehyde (HCHO) and acetone, alongside primary
94 aromatics and biogenic species (Carter, 2009). We utilized these individual, curated species as inputs to maximize
95 granular insight. Alternative selection strategies, including statistical recursive feature elimination (RFE) (Guyon et
96 al., 2002) and physicochemical grouping based on reactivity or source, were evaluated and subsequently rejected due
97 to issues with chemical circularity and reduced predictive accuracy, respectively (details presented in Text S5). Finally,
98 it must be noted that certain key precursors, particularly light alkanes (e.g., ethane, propane), were not accurately
99 quantified by the specific PTR-MS configuration used and are absent from model inputs.

100 **2.2.2 ML Model Configuration and Temporal Integration**

101 We employed the XGBoost gradient boosting framework, selected for its handling of non-linear interactions and
102 compatibility with Shapley Additive Explanations (SHAP) (Chen and Guestrin, 2016; Lundberg et al., 2020). Two



103 ML model configurations were developed: a conventional model using ambient concentrations, and a process-
104 informed model using the net change in VOC concentration (ΔVOC). Input features included NO_x (representing the
105 titration sink), meteorological variables, and temporal markers to capture diurnal emission cycles (Tables S3 and S4).
106 To prevent overfitting and ensure robust model performance, we employed a 10-fold cross-validation approach. The
107 dataset was partitioned into ten distinct subsets, with each subset serving iteratively as an independent validation set
108 while the remaining data were used for training (detailed in Text S6).

109 To capture the relevant chemical timescale, data were resampled into time steps (Δt) of 18 minutes. The process-
110 informed variables (ΔX) were calculated as:

$$111 \quad \Delta X = \frac{X_t - X_{t-1}}{\Delta t} \quad (1)$$

112 where X represents the input (VOC, NO_x , and meteorological factors) or output (ozone) species. The 18-minute
113 integration window was rigorously determined via optimization tests (Fig. S7) to balance signal fidelity against
114 physical transport. This interval is sufficiently long to average out high-frequency turbulent noise (dominant at < 5
115 min) yet short enough to precede significant shifts in planetary boundary layer dynamics (dominant at > 30 min,
116 effectively isolating the local chemical signal) (Mauder et al., 2013; Metzger, 2018).

117 Time-shift analysis (introducing a lag between model inputs and the output target)—testing for potential delays
118 between VOC oxidation and ozone response—confirmed optimal model performance at zero lag. This immediate
119 response is consistent with the rapid kinetics of a NO_x -saturated roadside regime, where titration and alkene oxidation
120 drive ozone variability on timescales shorter than the integration window (Clapp and Jenkin, 2001). While ΔVOC
121 mathematically integrates emissions, transport, and chemistry, the study site offers a simplifying constraint: the strong
122 regularity of traffic emissions is captured by the 'hour of day' variable. Consequently, the remaining variance in ΔVOC
123 serves as a robust empirical proxy for the net chemical turnover driving local ozone production. It is important to
124 distinguish this from metrics like OHR and OFP, which are computed from current concentrations at the sampling site
125 and therefore represent an instantaneous propensity (a box model snapshot) rather than a realized response that
126 depends on prior air mass evolution. By design, ΔVOC summarizes the net change over Δt , so it is sensitive to the
127 recent history of the sampled air mass within the selected window.

128 Effectively, the evaluation of the Δ derivative over this optimized 18-minute window acts as a dynamic kinetic high-
129 pass filter. It mathematically strips away the inert, accumulated background of long-lived species (where $\Delta\text{VOC} \approx$
130 0) and exclusively captures the instantaneous oxidative turnover of fresh local emissions.

131 **2.3 Chemical Indicators and Observation-Based Validation**

132 To benchmark the ML results against established chemical frameworks, we employed three complementary metrics.
133 First, theoretical ozone formation contributions were assessed using Ozone Formation Potential (OFP) based on MIR
134 factors (Carter, 2009) and OH reactivity (OHR), which quantifies the instantaneous kinetic driving force (Text S1)
135 (Sinha et al., 2012). Second, in situ ozone production dynamics were simulated using an Observation-Based Model
136 (OBM) driven by the Master Chemical Mechanism (MCM v3.3.1) via the AtChem2 framework (Text S2) (Sommariva



137 et al., 2020). The model yielded the Relative Incremental Reactivity (RIR) for each precursor, providing a direct metric
138 of effective control potential (Wang et al., 2025). Finally, to distinguish chemical processing from physical dilution,
139 photochemical age (time-integrated OH exposure) was calculated using a hydrocarbon clock based on the ratio of
140 benzene to trimethylbenzenes (TMB) (Text S3) (Borbon et al., 2013). Detailed model constraints, isomer allocations
141 derived from KORUS-AQ data, and the specific rate constants utilized for photochemical aging are provided in the
142 Supporting Information.

143 **3 Results and Discussion**

144 **3.1 Overview of measurements**

145 The sampling site exhibited a chemically active urban regime driven by the coexistence of fresh anthropogenic
146 emissions and regional background oxidation. While the bulk VOC composition was numerically dominated by
147 oxygenated species (OVOCs) such as methanol and acetone (Table S1), the substantial abundance of primary
148 aromatics—specifically toluene (4.51 ± 6.02 ppb) and C_8 aromatics (1.40 ± 1.10 ppb)—confirms the immediate
149 influence of local, continuous emissions rather than aged air masses. This precursor abundance fueled intense in situ
150 photochemistry: ground-level O_3 exhibited a sharp diurnal amplitude, rising from a morning minimum of 16.2 ppb to
151 an afternoon peak of 63.8 ppb (Fig. S2). Crucially, 83% of hourly measurements exceeded the WHO peak-season
152 guideline of 31 ppb (World Health Organization, 2021), highlighting the persistence of hazardous ozone levels in this
153 NO_x -saturated environment.

154 To verify the dominance of local primary sources, we employed a ternary diagram using benzene, toluene, and
155 ethylbenzene (Fig. 1) (Zhao et al., 2024). Because PTR-ToF-MS cannot chemically resolve C_8H_{10} isomers,
156 ethylbenzene was differentiated from xylenes using a fixed ratio (0.366) derived from speciated measurements during
157 the local KORUS-AQ campaign (Simpson et al., 2020). The resulting distribution clusters predominantly within the
158 traffic-specific wireframe, establishing vehicular exhaust as the persistent baseline source. However, a noticeable skew
159 toward the solvent/industrial sector indicates distinct episodic enhancements driven by toluene-rich commercial
160 activities.

161 This dual-source influence is further clarified by the diurnal stability of the toluene-to-benzene (T/B) ratio (Fig. 1b).
162 A distinct temporal divergence emerges between the two rush-hour periods. The morning rush (07:00–09:00) exhibits
163 a significantly elevated toluene-to-benzene (T/B) ratio. This enhancement reflects the combined impact of cold-start
164 traffic emissions (dominated by unburned fuel components rich in toluene due to inefficient catalytic conversion at
165 low temperatures) and the onset of daytime solvent usage (Zhang et al., 2022; Simpson et al., 2020). In contrast, the
166 afternoon rush (17:00–19:00) returns to a suppressed ratio indistinguishable from background levels. This suggests
167 that while traffic provides the fundamental source signature, high-intensity solvent emissions are temporally confined
168 to the morning operation, creating specific periods of heightened reactivity.



169 **3.2 Main drivers of ozone formation**

170 This section identifies the principal drivers of ozone formation by comparing the interpretability of a conventional
171 machine learning model against a process-informed framework. We systematically evaluated the sensitivity of ozone
172 production to inorganic precursors, meteorological parameters, and specific VOC classes. To verify the physical
173 consistency of the machine learning attribution, model outputs were cross-validated using observation-based chemical
174 mechanisms (MCM-OBM) and kinetic reactivity metrics (OFP, OHR). The following analysis aims to distinguish
175 between the environmental conditions that facilitate ozone accumulation and the specific chemical precursors that
176 drive rapid local production.

177 **3.2.1. Resolving instantaneous drivers with the process-informed framework**

178 By resolving the temporal dynamics of the system, the process-informed (Δ) framework decouples the accumulated
179 atmospheric state from its active drivers, revealing a regime governed by rapid titration and chemical limitation.

180 The analysis identifies active titration as the primary high-frequency control governing local ozone variability within
181 the 18-minute integration window. While the conventional model (Fig. 2a) broadly characterizes the environment as
182 a static NO_x -saturated state, the process-informed model (Fig. 2a; blue scatter points) captures the instantaneous
183 kinetic reality: fresh NO emissions are associated with a clear negative response in ozone production ($\Delta\text{NO}_x > 0 \rightarrow$
184 $\Delta\text{O}_3 < 0$). This sharp negative response is not merely a statistical correlation but a mechanistic signature of the rapid
185 titration reaction ($\text{NO} + \text{O}_3 \rightarrow \text{NO}_2 + \text{O}_2$). Since this reaction operates on timescales of roughly one minute in urban
186 plumes (Liu et al., 2008; Sillman, 1999; Zhang et al., 2022), its full effect is completely captured within the 18-minute
187 integration window. By isolating this rapid suppression, the framework confirms that titration functions as a dynamic
188 modulator of ozone on sub-hourly scales rather than a static background condition, distinct from the slower evolution
189 of the underlying oxidation regime.

190 The dynamic model further corrects a critical misinterpretation regarding the stability of the oxidation regime. The
191 conventional model (Fig. 2e; red scatter points) displays a precipitous "titration cliff" across the range of HCHO/NO_x
192 ratios. While this plot accurately captures the diurnal association—where low morning ratios correlate with low ozone
193 and high afternoon ratios with high ozone—the static framework conflates this long-term evolution with instantaneous
194 sensitivity. This pattern could be misread as an instantaneous hypersensitivity, as if a small, immediate perturbation in
195 the ratio would trigger a disproportionate surge in ozone production. The process-informed model (Fig. 2e) resolves
196 this ambiguity by analyzing the rate of change (Δ). By showing negligible dependence on short-term ratio fluctuations
197 ($\Delta\text{Ratio} \rightarrow \Delta\text{O}_3$), it reveals that the oxidation regime functions as a stable thermodynamic boundary condition. The
198 HCHO/NO_x ratio evolves gradually over diurnal timescales—driven by changing solar angles and cumulative
199 emissions—rather than shifting rapidly enough to drive high-frequency ozone variance within the 18-minute window.
200 Consequently, the delta model confirms that the "cliff" observed in static analysis is a statistical artifact of time-
201 aggregated data, whereas the dynamic framework recovers the physical reality of a buffered, slowly evolving radical
202 cycle (Pusede and Cohen, 2012).



203 Finally, the framework clarifies that peak ozone production is chemically limited, rather than meteorologically driven.
204 Conventional models often conflate high ozone with its environmental prerequisites (high temperature, low humidity).
205 However, the process-informed model (Fig. 2; blue scatter points) shows negligible sensitivity to short-term
206 fluctuations in these parameters. This decoupling confirms that once the meteorological "stage" is set, the rate of ozone
207 production is controlled strictly by the supply and turnover of reactive precursors, aligning with box-model studies
208 demonstrating that in stagnation events, chemical throughput, not micrometeorological variance, dictates the
209 production rate (Tan et al., 2019).

210 3.2.2 Uncovering VOC Drivers via Net Chemical Consumption

211 The stability of the underlying oxidation regime (established in Section 3.2.1) implies that high-frequency ozone
212 fluctuations are driven by the specific reactive turnover of precursors processed within that regime. However, correctly
213 identifying these precursors requires filtering out the "survivor bias" that obscures the true chemical drivers in
214 concentration-based analyses.

215 Conventional ML analysis yields a chemically counterintuitive attribution: abundant, low-reactivity OVOCs (e.g.,
216 acetone) are identified as primary drivers, while highly reactive aromatics (toluene, TMB) are assigned negative or
217 negligible ozone formation potentials (Fig. 3; red scatter points). The OBM-MCM radical budget analysis (Fig. S5)
218 refutes this, demonstrating that toluene oxidation is a net source of peroxy radicals (RO_2), contributing an order of
219 magnitude more to the propagation cycle than acetone. Consequently, the conventional model's attribution is
220 confirmed as a statistical artifact arising from diurnal anti-correlation.

221 The process-informed framework resolves this causality gap by utilizing net chemical consumption (ΔVOC) as the
222 predictor. The resulting SHAP dependence plots (Fig. 3; blue scatter points) reveal a chemically coherent mechanism:
223 ozone production is strictly coupled to the reactive loss of primary precursors.

224 Specifically, the model recovers a strong inverse relationship for reactive aromatics ($\Delta\text{VOC} < 0 \rightarrow$ High SHAP Value),
225 indicating that the rate of ozone formation accelerates in direct proportion to the consumption rate of toluene, xylenes,
226 and TMB. Importantly, this relationship is absent for accumulated low reactivity OVOCs. This distinction provides a
227 critical atmospheric insight: despite the numerical dominance of oxygenated species in the ambient mix, the
228 instantaneous production of ozone in this NO_x -saturated environment is kinetically limited by the supply of fresh,
229 high-reactivity aromatics. In this sense, realized ozone production at the site appears tied more to the fast turnover of
230 reactive aromatics than to the standing abundance of oxygenated species.

231 3.3 Kinetic Control of Ozone Production and Radical Termination Pathways

232 The application of the process-informed machine learning framework revealed that local ozone production in the
233 Seoul Metropolitan Area is governed by a strict kinetic hierarchy, where the intrinsic reactivity of a precursor
234 outweighs its ambient abundance. We quantified this relationship by examining the net consumption of each VOC and
235 its corresponding SHAP response slope. This was achieved by calculating a "SHAP slope" for each species, defined
236 as the slope of the linear regression between the input ΔVOC and its corresponding SHAP value (contribution to ΔO_3).



237 In this framework, a steeper negative slope indicates that the net depletion of a precursor (negative ΔVOC) is
238 associated with a higher contribution to ozone production (positive ΔO_3).

239 3.3.1 The Primacy of Kinetic Reactivity (kOH)

240 As shown in Fig. 4a, there is a distinct non-linear relationship between this data-driven SHAP response slope and the
241 hydroxyl radical reaction rate constant (kOH). VOCs with low kinetic reactivity, such as toluene ($k\text{OH}=5.7 \times 10^{-12}$
242 $\text{cm}^3 \text{molecule}^{-1} \text{s}^{-1}$) and benzene ($k\text{OH}=1.2 \times 10^{-12} \text{cm}^3 \text{molecule}^{-1} \text{s}^{-1}$), exhibit SHAP slopes near zero. Despite their
243 high dominance in the ambient mixing ratio (Table S1), their slow oxidation rates prevent them from driving the rapid
244 radical cycling required to sustain elevated ozone levels in this NO_x -saturated regime.

245 Conversely, the sensitivity curve steepens dramatically for highly reactive compounds, including isoprene
246 ($k\text{OH}=100 \times 10^{-12} \text{cm}^3 \text{molecule}^{-1} \text{s}^{-1}$) and trimethylbenzenes ($k\text{OH}=36.0 \times 10^{-12} \text{cm}^3 \text{molecule}^{-1} \text{s}^{-1}$; average of
247 three isomers). This trend confirms that the rate of oxidative turnover is the primary predictor of ozone variability.

248 Isoprene presents a unique challenge in this analysis. Typically, its strong daytime emissions, originating from both
249 adjacent urban vegetation (e.g., Namsan Mountain) and vehicular traffic, mask its chemical consumption, often
250 resulting in positive ΔVOC values even during active oxidation. However, the model assigns isoprene a high
251 sensitivity index. This suggests that the model successfully detects deviations from the expected accumulation rate—
252 identifying moments where rapid chemical loss competes with strong emissions. This high sensitivity aligns with the
253 premise that in a NO_x -saturated regime, the system is strictly limited by the supply of peroxy radicals, which high-
254 kOH species like isoprene generate most efficiently. These results extend established chemical frameworks by
255 demonstrating that rapid depletion dynamics are the strongest statistical predictor of ozone formation potential (de
256 Gouw et al., 2017; Pusede and Cohen, 2012). Crucially, this result demonstrates a major methodological breakthrough:
257 without being explicitly programmed with physical or chemical mechanisms, the ΔVOC -informed ML framework
258 autonomously recovered the fundamental kinetic rate laws of atmospheric chemistry purely from observational data.

259 Finally, to verify that the ΔVOC term represents chemical oxidation rather than physical dilution, we performed an
260 independent observational verification detailed in Text S4. By analyzing VOC depletion relative to photochemical age
261 (Fig. S4), we confirmed that dilution-dominated species (e.g., toluene) follow CO decay profiles, while oxidation-
262 driven drivers (e.g., TMB) exhibit rapid depletion relative to CO. This confirmed that the derived sensitivities isolated
263 the chemical consumption specifically responsible for local ozone production.

264 3.3.2 Mechanistic Penalties: Radical Termination and Outliers

265 While the process-informed model broadly aligns SHAP response slope with kOH, specific outliers reveal the model's
266 capacity to distinguish between simple consumption speed and the ultimate efficiency of radical recycling (Fig. 4a).

267 Specifically, heavy aromatics like styrene and TMB exhibit lower sensitivities than their reactivity would suggest
268 (falling below the trendline). For TMB, this suppressed sensitivity is consistent with the competition between chain
269 propagation and termination via organic nitrate formation. Upon oxidation, large alkyl-substituted aromatics exhibit
270 significantly higher organic nitrate yields (RONO_2) compared to lighter species (Perring et al., 2013). This branching



271 pathway effectively sequesters both NO_x and radicals, reducing the net ozone yield per oxidized TMB molecule
272 relative to less-substituted aromatics. The model correctly detects that a substantial fraction of the oxidized TMB mass
273 is diverted into these termination reservoirs rather than fueling the catalytic cycle. Similarly, styrene shows a
274 suppressed sensitivity consistent with inefficient radical regeneration. Unlike light alkenes, the ozonolysis of styrene
275 is characterized by a negligible yield of recycled OH radicals (Tuazon et al., 1993). Consequently, styrene acts partially
276 as a radical sink, consuming oxidants without regenerating the oxidative capacity required to sustain the
277 photochemical chain.

278 In contrast, monoterpenes (measured as $\Sigma\text{C}_{10}\text{H}_{16}$) exhibit higher-than-expected sensitivities, placing them above the
279 general reactivity trend. Theoretically, monoterpenes also possess high organic nitrate yields, which should penalize
280 their production. However, our results suggest this penalty is outweighed by efficient non-photolytic radical generation.
281 The ozonolysis of monoterpenes—particularly α -pinene and limonene—is a potent source of non-photolytic OH
282 radicals with a yield of approximately 0.85 (Chew and Atkinson, 1996). This strong radical boosting effect could
283 compensate for the loss of NO_x to nitrates, allowing these species to sustain rapid oxidation chains even under NO_x -
284 saturated conditions.

285 It is important to acknowledge that the PTR-MS technique measures the sum of monoterpene isomers, preventing the
286 distinct separation of species with varying yields (e.g., α -pinene vs. β -pinene). Since the efficiency of OH recycling
287 varies among isomers, the elevated sensitivity observed here likely reflects a composite signal dominated by high-
288 yield species in the local biogenic mix. Nevertheless, the model's differentiation of these species from aromatics
289 underscores its capacity to capture the nuance of radical stoichiometry beyond simple reaction kinetics.

290 3.3.3 Drivers of Peak Pollution Episodes

291 The practical implication of this kinetic control is clearest when diagnosing high-ozone episodes. Fig. 5 compares the
292 model sensitivity during baseline conditions versus high-ozone events. A critical evolution is observed in the NO_x
293 response: while the negative sensitivity (titration) persists during high-ozone events, the magnitude attenuates (slope
294 shifts from -0.45 to -0.28). This indicates that while the system remains strictly within the NO_x -saturated regime
295 (where increasing NO_x suppresses ozone), the efficiency of this suppression diminishes as the air mass evolves toward
296 peak concentrations.

297 Crucially, as the titration effect attenuates, the system exhibits an increase in response to reactive radical precursors.
298 During high-ozone episodes, the reactivity-response slope significantly steepens (linear fit coefficient shifts from -
299 0.34 to -0.44; Fig. S8). This steepening indicates that the ozone influence for high-kOH species is amplified: the
300 system becomes disproportionately sensitive to monoterpenes and reactive aromatics, while the response to lower-
301 kOH species like toluene remains negligible. This confirms that the transition to hazardous pollution levels is not
302 driven by a fundamental shift in the chemical regime (e.g., from VOC-limited to NO_x -limited), but by the intensified
303 processing of specific rate-limiting precursors.

304 These results show the distinct roles of different chemical metrics in diagnosing local pollution episodes. Conventional
305 metrics like OHR and OFP (Fig. 4b) measure the current ozone production potential at sampling time, which



306 effectively represents the theoretical yield if the air mass remained isolated in a static 0-D box model. In the real
307 atmosphere, ozone depends on air mass history, and the Δ VOC metric captures some of this dynamic evolution by
308 accounting for rapid chemical depletion. While OFP identifies the ultimate stoichiometric capacity over longer scales,
309 the process-informed framework reveals that in this fresh-emission environment, the rapid reaction kinetics (k_{OH})
310 dictate the realized ozone formation.

311 In this NO_x -saturated regime, ozone production is rate-limited by the instantaneous supply of peroxy radicals (RO_2)
312 required to convert NO to NO_2 . High- k_{OH} species provide this radical flux immediately upon emission, rapidly
313 initiating the oxidation cycle. By prioritizing this instantaneous turnover over integrated potential (OFP and OHR),
314 the process-informed framework effectively closes the gap between data-driven modeling and kinetic theory. The
315 model recovers the kinetic dominance of high- k_{OH} species—specifically aromatics and isoprene—consistent with
316 mechanistic box-modeling studies (Lu et al., 2019). This validates that by resolving short-term chemical variances
317 (e.g., the 18-minute window), data-driven approaches can successfully align with the fundamental kinetic principles
318 established in experimental atmospheric chemistry.

319 **4 Conclusions**

320 This study demonstrates that, for the Seoul roadside observations analyzed here, concentration-based machine-
321 learning interpretation can systematically misidentify the kinetic drivers of urban ozone—a ‘survivor bias’ to which
322 standard data-driven analyses are particularly susceptible in NO_x -saturated, fresh-emission regimes. Within the
323 process-informed (Δ VOC) framework, reactive aromatics (xylenes, TMB) and biogenic species (isoprene,
324 monoterpenes) emerge as the compounds most closely tied to short-timescale ozone production, whereas the more
325 abundant OVOCs primarily reflect accumulated background conditions rather than immediate chemical control.
326 Realized ozone production in fresh urban plumes is therefore governed by precursor turnover and radical-production
327 efficiency rather than by bulk abundance alone. The persistent negative response to NO_x is also consistent with a
328 continuing titration influence during the observation period, suggesting that, in this regime, NO_x reductions are
329 unlikely to translate into proportional peak-ozone reductions without accompanying VOC controls. Importantly,
330 during high-ozone episodes the system’s response to reactive precursors steepens while the titration efficiency
331 attenuates, indicating that hazardous events are driven not by a regime change but by intensified turnover of specific
332 rate-limiting species—a finding with direct implications for the marginal-benefit calculations underlying VOC control
333 strategies. More broadly, these results indicate that mass-based emission inventories and OFP-style rankings, when
334 applied without kinetic context, can misallocate control priorities in NO_x -saturated urban environments; reactivity-
335 informed and process-constrained metrics should therefore complement—and in such regimes potentially supersede—
336 conventional inventories when informing precursor control. Because the analysis is based on a single roadside site, a
337 one-month spring observation period, an 18-minute integration window, and a PTR-ToF-MS suite that does not fully
338 capture all precursor classes (notably light alkanes), further applications across seasons, source environments, and
339 chemical regimes will be needed to test generality. The framework itself, however, requires only sub-hourly co-located



340 VOC and ozone observations and is therefore broadly applicable wherever such data are available—including settings
341 where explicit mechanism-based modeling is constrained by incomplete VOC speciation or unmeasured radical
342 precursors.

343 **Data availability.**

344 The observational data and ML code used in this study are available from the corresponding author upon reasonable
345 request.

346 **Author contributions**

347 Conceptualization, Methodology, Writing—review & editing, H.K.; Methodology, Investigation, Modeling,
348 Writing—original draft, Q.H.; Discussion—J.G.; Measurement—S.K., Y.P., S.L.

349 **Declaration of competing interests**

350 The authors declare that they have no known competing financial interests or personal relationships that could have
351 appeared to influence the work reported in this paper.

352 **Supplemental Information**

353 S1. Measurement results of gas species, VOCs, and meteorology; S2. Details of the photochemical clock and MCM-
354 OBM model; S3. Validation, selection of input parameters, and results of the ML model.

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359

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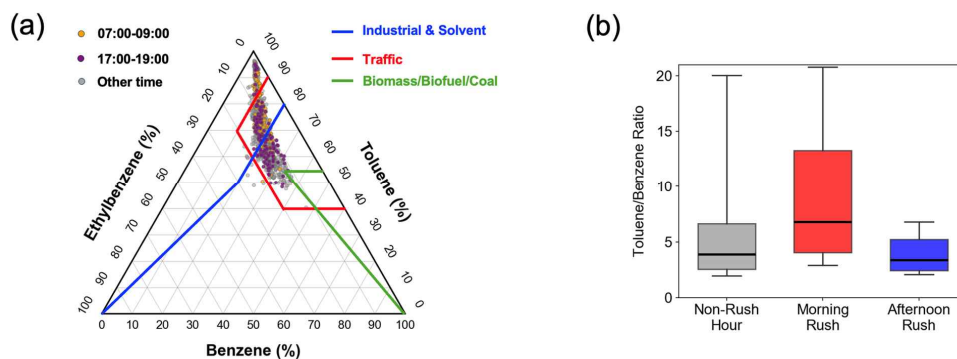
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468 **Figure Captions**

- 469 • Figure 1: (a) Ternary diagram of normalized benzene, toluene, and ethylbenzene mixing ratios. Data points are
470 stratified by sampling time. The wireframes indicate source profiles for industrial and solvent emissions, traffic,
471 and biomass/biofuel/coal burning. (b) Box plots of Toluene/Benzene ratios categorized by non-rush, morning rush,
472 and afternoon rush hours. The horizontal line within each box represents the median value.
- 473 • Figure 2: SHAP dependence plots comparing the concentration-based model (red scatter points, bottom/left axes)
474 and the delta model (blue scatter points, top/right axes) for (a) NO_x, (b) temperature, (c) RH, (d) solar radiation,
475 and (e) HCHO/NO_x ratio. The color of the markers represents the hour of day, as indicated by the respective blue
476 and red color bars.
- 477 • Figure 3: SHAP dependence plots comparing the concentration-based model (red markers, bottom/left axes) and
478 the delta model (blue markers, top/right axes) for: (a) C₁₀H₁₆, (b) C₉H₁₂, (c) C₇H₈, (d) C₃H₄O, and (e) C₃H₆O.
- 479 • Figure 4: Relationship between SHAP-derived response slope (slope of SHAP dependence for ΔVOC) and
480 chemical reactivity metrics: (a) OH reaction rate constant (kOH), and (b) OH reactivity (OHR, left axis) and ozone
481 formation potential (OFP, right axis) using the same x-axis VOC ordering as in (a). Analysis includes only VOCs
482 with negative SHAP slopes, where consumption correlates with ozone formation. kOH for isomeric groups are
483 weighted averages from KORUS-AQ: C₈H₁₀ (1.39×10^{-11}) for ethylbenzene and xylenes; C₉H₁₂ (3.60×10^{-11}) for
484 trimethylbenzenes; and C₁₀H₁₆ (6.63×10^{-11}) for a 1:1 α- to β-pinene ratio (Kenagy et al., 2020; Simpson et al.,
485 2020).
- 486 • Figure 5: SHAP dependence plots for (a) NO_x, (b) C₁₀H₁₆, (c) C₉H₁₂, and (d) C₇H₈. Red and green markers
487 represent high (O₃ > 45 ppb) and low (O₃ < 45 ppb) ozone periods, respectively. Dashed lines denote linear
488 regressions, where the fitted slope indicates the ozone response slope to the precursor.

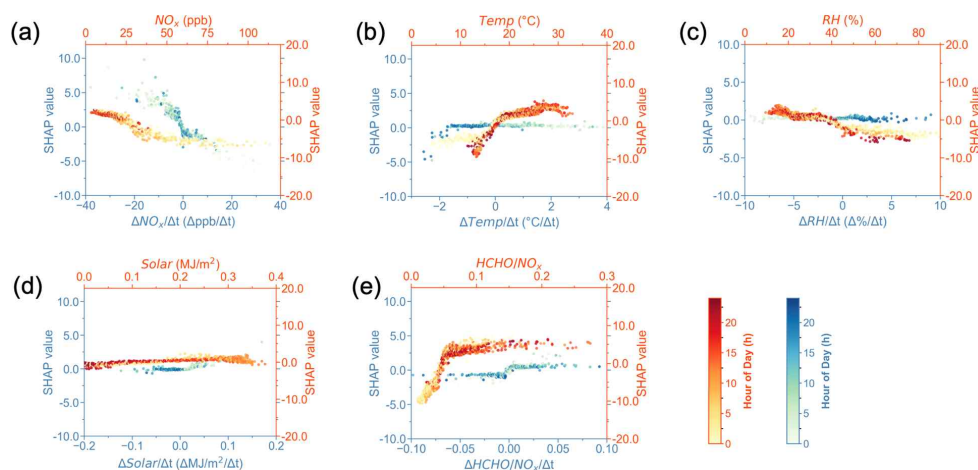


489 **Figures list**



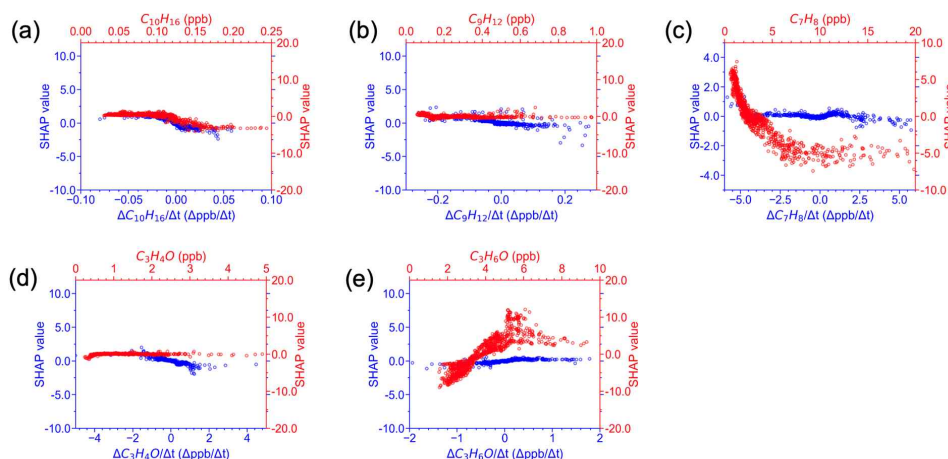
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 494 **rush, and afternoon rush hours. The horizontal line within each box represents the median value.**



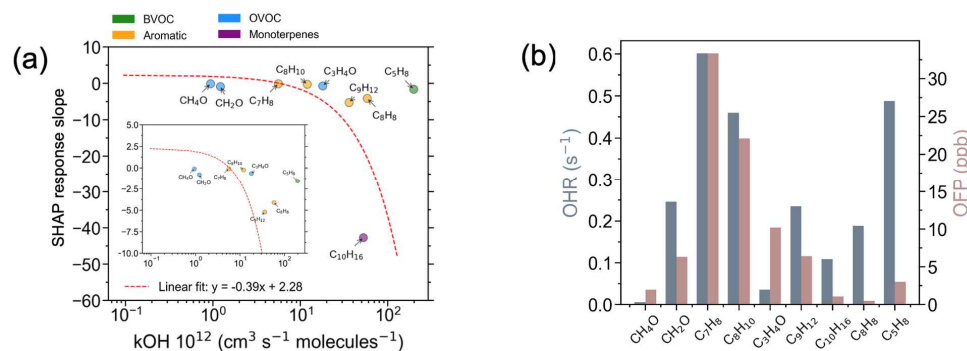
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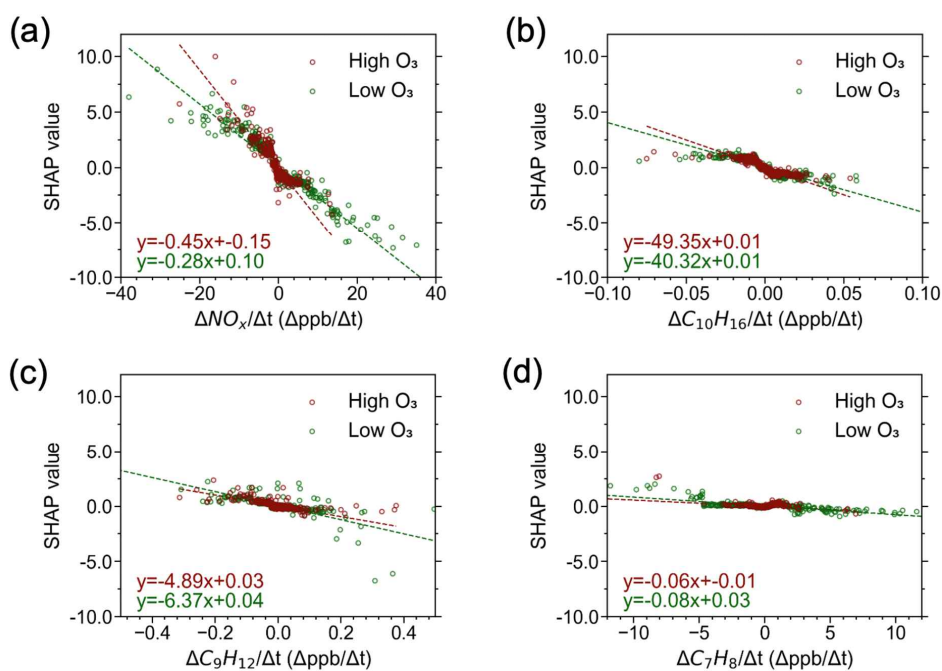
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501 **Figure 3: SHAP dependence plots comparing the concentration-based model (red markers, bottom/left axes)**
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503

504 **Figure 4: Relationship between SHAP-derived response slope (slope of SHAP dependence for ΔVOC) and**
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 509 **× 10⁻¹¹) for trimethylbenzenes; and C₁₀H₁₆ (6.63 × 10⁻¹¹) for a 1:1 α- to β-pinene ratio (Kenagy et al., 2020;**
 510 **Simpson et al., 2020).**



511

512 **Figure 5: SHAP dependence plots for (a) NO_x, (b) C₁₀H₁₆, (c) C₉H₁₂, and (d) C₇H₈. Red and green markers**
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