





27 and  $R_{t-EOS}$ ) was quantified. We observed an urban-rural phenological disparity of  
28 12.06 days for Start of Season ( $\Delta$ SOS) and 9.86 days for End of Season ( $\Delta$ EOS)  
29 among the studied cities. Spatially, cities in high latitude regions and coastal areas  
30 exhibited pronounced negative  $\Delta$ SOS shifts and positive  $\Delta$ EOS shifts, positively  
31 correlating with  $R_{t-SOS}$  and  $R_{t-EOS}$ , respectively. Employing a continent-wide preseason  
32 temperature ( $T$ ), we observed a logistic decrease for SOS and an increase for EOS,  
33 illustrating the "saturated effect" of warming on plant phenology-patterns echoed  
34 within urban settings. First-order derivatives of those logistic curves identified a  
35 highest phenological sensitivity at  $T=4^{\circ}\text{C}$  and  $T=6^{\circ}\text{C}$ , as well as the warming benefit  
36 range of  $-3.5^{\circ}\text{C}$ - $10^{\circ}\text{C}$  and  $2^{\circ}\text{C}$ - $14^{\circ}\text{C}$  for SOS and EOS respectively. Substituting  $T$   
37 with LST, weaker  $\Delta$ SOS and  $\Delta$ EOS would be presented in warmer regions only when  
38 LST exceeded  $12.5^{\circ}\text{C}$  and  $4^{\circ}\text{C}$  for spring and autumn, respectively. Except for LST,  
39 AI exhibited a positive correlation with  $\Delta$ SOS and  $\Delta R_{t-SOS}$ , but a negative one with  
40  $\Delta$ EOS and  $\Delta R_{t-EOS}$ . Collectively, LST and AI explained 75.05% and 76.21% of the  
41 phenological variance across the continent for  $\Delta$ SOS and  $\Delta$ EOS, respectively. These  
42 findings lay the groundwork for predicting vegetation changes under global warming  
43 at large scales.

44 **Keywords:** urbanization; heat island effect; vegetation phenology sensitivity; aridity  
45 index; land surface temperature; "saturation effect"

## 46 1. Introduction

47 Vegetation phenology, a critical aspect of ecosystem dynamics, exhibits  
48 significant sensitivity to climate variations (Badeck et al., 2004). Recent investigations  
49 have highlighted the influence of urban warming phenomena in precipitating earlier  
50 onset of green-up phases (SOS) and postponing the transition to dormancy  
51 (EOS) (Jeonget et al., 2019; Meng et al., 2020; Qiu et al., 2020; Zhou et al., 2016).  
52 The cascading effects of these shifts extend beyond mere phenological alterations,  
53 contributing to a spectrum of ecological and environmental challenges (Zhou et al.,  
54 2016; Qiu et al., 2017; Tang et al., 2017).

55 However, the degree to which warming benefits phenological changes has been



56 widely debated(Chmielewski & Rötzer, 2001; Ding et al., 2020; Yao et al., 2017).  
57 Studies have argued that background temperature heterogeneity has the potential to  
58 disrupt phenology responses to urban warming(Meng et al., 2020). For example, in  
59 cold high-latitude regions of the Northern Hemisphere, more advanced SOS and  
60 delayed EOS has been reported in response to warming events(Meng et al., 2020; Qiu  
61 et al., 2020; Dallimer et al., 2016; Gazal et al., 2008; Li et al., 2021). While these  
62 studies offer empirical evidence, they do not sufficiently elucidate the underlying  
63 cause and its quantifiable relationship.

64 Differences in plant phenological sensitivity to background temperatures might  
65 contribute to the spatial heterogeneity of plant responses to urbanization(Meng et al.,  
66 2020). In the field of biology, “saturation effects” refer to the psychological  
67 phenomenon where the efficacy of a stimulus declines due to the continuous  
68 application of the same stimulus material. Similarly, in the field ecology, according to  
69 the theory of tolerance curves, an organism's physiological functions enhance with  
70 increasing environmental factors, reaching a point of physiological saturation, beyond  
71 which no further enhancement occurs and a decline may ensue(Ketola & Kristensen,  
72 2017). For example, many studies had argued that the plant physiological processes  
73 such as photosynthesis, respiration, transpiration etc. would not increase linearly with  
74 increased temperature(Sage & Kubien, 2007; Ge et al., 2021). On the contrary, they  
75 will gradually slow down, and even decreased when temperature surpass a  
76 threshold(Sage & Kubien, 2007; Mathur et al., 2021). A meta-analysis of the response  
77 of above-ground plant growth to experimental ecosystem warming had proved a  
78 greater positive response to warming in colder ecosystems(Rustad et al., 2001). Since  
79 the warmer regions have higher mean temperature, thus we expected that their urban  
80 warming related benefits for phenology (i.e., advanced SOS and delayed EOS) will  
81 also gradually approaching the saturated point. Extra warming due to urban warming  
82 effect will not promoted the advancement of SOS and delay of EOS, i.e., the  
83 “saturation effect”. However, these theories are still need to verified.

84 It is noteworthy that except for background temperature, drought conditions can  
85 passively disrupt phenological responses linked to ambient temperature(Peng et al.,



86 2019; Yuan et al., 2020). For spring phenology, drought induces the accumulation of  
87 abscisic acid (ABA)(Mcadam, 2013), leading to delayed spring germination(Eidt &  
88 Little, 1968). A pronounced delayed SOS was documented on the Yungui Plateau  
89 preceding drought occurrences(Peng et al., 2019). Likewise, delayed SOS resulting  
90 from drought was also observed in grassland ecosystems in Northeastern China(Yuan  
91 et al., 2020). Laboratory investigations have demonstrated that one-year-old saplings  
92 of Pedunculate oak exhibit delayed bud burst following a one-year drought  
93 treatment(Čehulić et al., 2019). Concerning autumn phenology, drought conditions  
94 prior to the season can trigger protein degradation and impair chloroplast function(Ge  
95 et al., 2021; Shi et al., 2014), potentially leading to widespread crown  
96 defoliation(Pollastrini et al., 2017). Consequently, under drought stress conditions,  
97 early dormancy (advanced EOS) is anticipated due to heightened water stress(Wang et  
98 al., 2020). Therefore, it is reasonable to anticipate that the sensitivity of phenology  
99 will be compromised in drought-affected regions, thereby diminishing the positive  
100 impact of warming on phenological responses.

101 As a developing country with diverse climates and rapid urbanization across its  
102 vast territory, China is confronted with significant challenges stemming from the  
103 urban heat island effect(He et al., 2017). The intensity of this effect averages at  $0.9 \pm$   
104  $1.1$  °C, although it exhibits considerable variation among different regions(Li et al.,  
105 2021). Thus, an investigation into the phenological response of plants to urbanization  
106 in China would enhance our understanding of their ability to acclimate to climate  
107 changes across various environments.

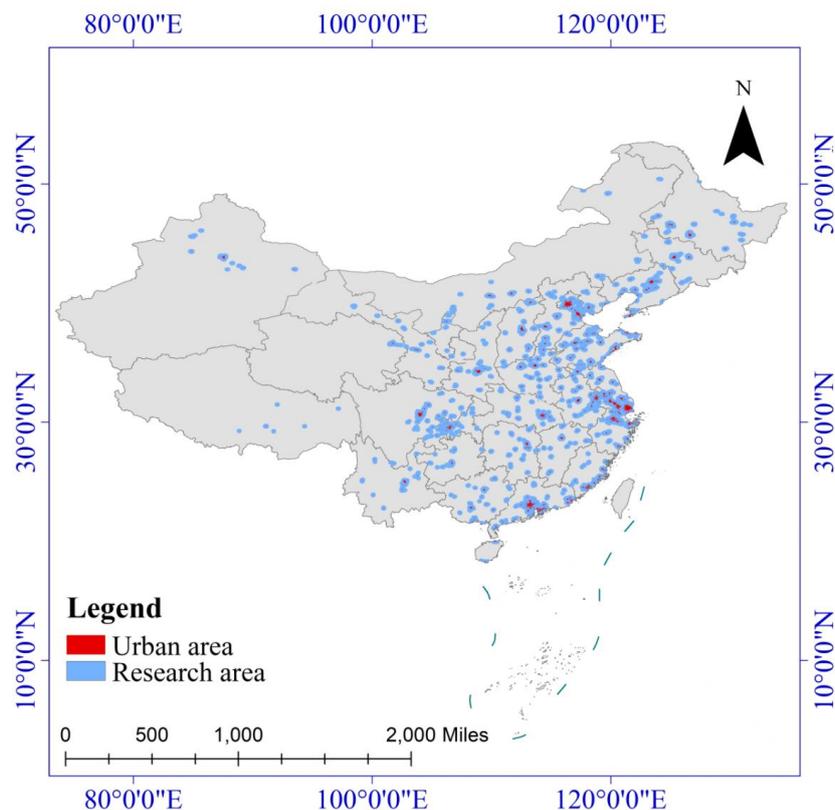
108 The primary objectives of this study encompassed the following: (1) exploring  
109 the spatial distribution of plant phenological responses to the urban heat island effect  
110 throughout mainland China; (2) quantify the saturation effect caused by the spatial  
111 variability of phenological responses correlated with background temperature across  
112 the cold-hot gradient; and (3) assessing whether the benefits of urban warming would  
113 be counteracted in drought-affected areas.



## 114 **2. Materials and Methods**

### 115 **2.1 Study Area**

116 In our study, we conducted an analysis covering a total of 293 prefecture-level  
117 cities in China (excluding Taiwan, Autonomous Prefecture, and regions), taking into  
118 account the data from the year 2020 (<http://xzqh.mca.gov.cn/map>) (Figure 1). It was  
119 hypothesized that the urban warming effect is most pronounced in urban centers and  
120 gradually diminishes towards rural areas(Zhang et al., 2004). To determine the  
121 urbanization intensity of each city, nighttime light data from NPP-VIIRS (Table 1)  
122 was utilized, employing the reference comparison method. The optimal segmentation  
123 threshold was established based on the statistical yearbook(Fan et al., 2019).  
124 Moreover, urban built-up areas were identified using the comparison method and  
125 annual average data derived from NPP-VIIRS night lights between 2013 and 2020.  
126 All data processing procedures were performed using ArcGIS 10.2. A total of ten  
127 buffer zones were created within a radius of 20 kilometers from the city center, with  
128 each buffer zone set at a distance of 2 kilometers from the center (Figure 1).



129

130 **Figure 1.** City centers (red area) and buffer zones (blue area) of the 293  
131 prefecture-level cities in China. The map, generated using ArcGIS – version 10.2  
132 (<https://www.esri.com/about/newsroom/arcwatch/the-best-of-arcgis-10-2/>), features  
133 administrative boundaries sourced from the Ministry of Civil Affairs of the People's  
134 Republic of China (<http://xzqh.mca.gov.cn/map>). Source: Esri, TomTom, FAO, USGS,  
135 and the GIS User Community; Ministry of Civil Affairs of the People's Republic of  
136 China | Powered by Esri.

## 137 2.2 Vegetation phenology extraction

138 Phenology data spanning the years 2010 to 2020 was derived from MODIS  
139 Enhanced Vegetation Index (EVI) 16-day composite data with a spatial resolution of  
140 250m (MOD13Q1) and processed utilizing TIMESAT. To mitigate challenges such as  
141 cloud cover and snow, the raw EVI time series dataset underwent smoothing through  
142 an S-G filter method(Cai et al., 2017). Subsequently, a dynamic threshold technique



143 was employed to extract the phenology data, utilizing a seasonal amplitude of 20%, a  
144 parameter proven effective for phenology extraction in urban  
145 environments<sup>5,31-33</sup>(Zhou et al., 2016; Buyantuyev & Wu, 2012; Cong et al., 2012;  
146 White et al., 2009). The datasets underwent preprocessing to eliminate pixels with  
147 minimal vegetation EVI, such as bodies of water and fully paved surfaces. To ensure  
148 the exclusion of unrealistic values for the start of season (SOS) and end of season  
149 (EOS), a range of 30 to 180 days and 210 to 360 days was specified. Contingent upon  
150 previous studies(Zhou et al., 2016; Zhang et al., 2006). Efforts were made to reduce  
151 the impact of outlier years by averaging the phenology data for each pixel across the  
152 timeframe from 2010 to 2020.

153 Furthermore, we employed land cover data from the MCD12Q1 dataset, which  
154 boasts a 500-meter resolution, to mitigate the impact of agricultural practices on the  
155 mainland during our analysis. The MCD12Q1 dataset categorizes land into 17 distinct  
156 classes. These include various types of forested areas such as Evergreen Coniferous  
157 Forests, Deciduous Needle-leaf Forests, Evergreen Broad-leaf Forests, Deciduous  
158 Broad-leaf Forests, Mixed Forests, as well as Closed Shrub lands, Open Shrub lands,  
159 Savannas, woody savannas, Grasslands, and Croplands.

160 Given the frequent human-induced disturbances in Croplands in the rural regions  
161 and the regular manual pruning of Grasslands in the urban center, Savannas, woody  
162 savannas, Grasslands, and Croplands were excluded from our study. The remaining  
163 land types were consolidated into broader categories: Broad-leaf Forests (BF), which  
164 encompasses both Deciduous and Evergreen Broad-leaf Forests; Coniferous Forests  
165 (CF), comprising both Deciduous and Evergreen Coniferous Forests; Mixed Forests;  
166 and Shrubs, which includes both Closed and Open Shrublands. This classification  
167 approach allowed us to focus our analysis on land types that are less influenced by  
168 human activities.

169 Table1 Data inventory used in this study

Data type	Resolution	Layer	Data sources	Year
MOD13Q1	250m/16d	EVI	<a href="https://search.earthdata.nasa.gov">https://search.earthdata.nasa.gov</a>	2010-2020



MOD11A2	1000m/8d	LST_Day LST_Night	<a href="https://search.earthdata.nasa.gov">https://search.earthdata.nasa.gov</a>	2010-2020
MOD16A2	500m/8d	PET	<a href="https://search.earthdata.nasa.gov">https://search.earthdata.nasa.gov</a>	2010-2020
MCD12Q1	500m	Land Cover	<a href="https://search.earthdata.nasa.gov">https://search.earthdata.nasa.gov</a>	2010-2020
VIIRS	500m	yearly average	<a href="https://eogdata.mines.edu">https://eogdata.mines.edu</a>	2013-2020
Precipitation data	1000m	monthly average	<a href="http://www.geodata.cn">http://www.geodata.cn</a>	2010-2020
China's urban zoning data			<a href="https://www.resdc.cn">https://www.resdc.cn</a>	2020
Station climates		monthly average	( <a href="http://data.cma.cn/en">http://data.cma.cn/en</a> )	2010-2020

### 171 2.3 Pre-season temperature and background climates

172 Building upon the methodology of a previous study(Piao et al., 2006), the  
 173 pre-season temperature (T) was characterized as the mean temperature for the three  
 174 months preceding the onset of the phenological season. For alignment with the  
 175 phenology information, the average temperature dataset was resampled to a 250m  
 176 resolution using the nearest neighbor interpolation technique. To evaluate the  
 177 performance of the downscaling procedure, the observed long-term monthly  
 178 Temperature and Precipitations across China were obtained from the National  
 179 Meteorological Information Center of China (<http://data.cma.cn/en>). This dataset  
 180 included observations from 496 national weather stations during 2010–2020. Upon  
 181 conducting a fitting analysis with station data, we observed that the coefficient of  
 182 determination ( $R^2$ ) for monthly mean temperature and remote sensing data reached  
 183 0.93, while that for monthly precipitation and remote sensing data achieved 0.85.  
 184 These results indicated that the downscaled meteorological data sufficiently meet the  
 185 requirements of this study.

186 The background climates were evaluated utilizing land surface temperature (LST)  
 187 and aridity index (AI) as indicators. The LST and AI data for each pixel were  
 188 averaged over the temporal span from 2010 to 2020. Specifically, MODIS LST  
 189 datasets exhibiting an absolute bias of less than 1K were selected as optimal for our



190 investigation, in line with previous studies(Gow et al., 2016; Pablos et al., 2016; Qiao  
191 et al., 2013; Wan et al., 2008). Version-6 LST data sourced from Aqua MODIS  
192 (MOD11A2, 8-day composite, 1km spatial resolution, covering the period from 2010  
193 to 2020) were employed in this analysis. Additionally, daily temperature (T) was  
194 calculated as the mean of daytime and nighttime temperatures, following the  
195 methodology outlined by Kang(Kang et al., 2016). AI was expressed as:

$$AI = PET/P \quad (1)$$

196 where, PET denotes the annual potential vapor transpiration, while P represents the  
197 annual precipitation. The PET values spanning from 2010 to 2020 were extracted  
198 from the MOD16A2 monthly synthetic product, characterized by an 8-day temporal  
199 resolution and a spatial resolution of 500 meters, as delineated in Table 1. The annual  
200 precipitation (P) dataset was derived from the monthly precipitation records of China  
201 encompassing the timeframe from 1901 to 2020, possessing a spatial resolution of 1  
202 km. Subsequently, both PET and P datasets were resampled to a spatial resolution of  
203 250 meters to facilitate the ensuing spatial analysis(Peng, 2020).

#### 204 **2.4 Urbanization effect on phenology**

205 The phenology date of each vegetation was extracted in each buffer zone. Finally,  
206 the mean phenology in this buffer zone was weighted by the phenology of each  
207 vegetation types. To ensure comparability in phenological analyses across buffer  
208 zones, we first identify the common vegetation types present in all zones and exclude  
209 the non-shared ones. We then calculate the phenological means for these common  
210 types within each buffer zone and average these values to determine the overall mean  
211 vegetation phenology for the entire buffer region, as showed below (equation 2, 3).

$$212 \quad SOS_i = \sum SOS_{ij}/n \quad (2)$$

$$213 \quad EOS_i = \sum EOS_{ij}/n \quad (3)$$

214  $SOS_i$  and  $EOS_i$  is the mean SOS and EOS of  $i$ -th buffer zone,  $SOS_{ij}$  and  $EOS_{ij}$  is the  
215 mean SOS and EOS of the  $j$ -th vegetation type in the  $i$ -th buffer zone.  $n$  is the  
216 number of the shared vegetation types in all the buffer zone.

217 To investigate the impact of urbanization on phenology, mean values of start of



218 season (SOS) and end of season (EOS) in each buffer zone were analyzed the for the  
219 period between 2010 to 2020, and fitted with respect to their distance from the urban  
220 center. In this study, the outermost buffer zone and the urban center were defined as  
221 the rural region and the urban region, respectively, since SOS and EOS gradually  
222 decreases and approaches 0 at 15 km and beyond (Figure S1). Similar pattern was  
223 also observed for all the four vegetation types (Figure S2). As per prior investigations,  
224 the phenological sensitivity ( $R_t$ ) to temperature was quantified by calculating the  
225 partial correlation coefficient between T and the phenology dates ( $R_{t-SOS}$  and  
226  $R_{t-EOS}$ )(Meng et al., 2020).

227 The difference of  $SOS_i$  and  $EOS_i$  between the urban and rural regions for the  
228 phenology dates were quantified. Since the change rates was approaching 0 at 15 km  
229 and beyond (Figure S2), The outermost buffer zone (20km away from the urban  
230 center) and the urban center were defined as the rural region and the urban region,  
231 respectively. The phenology difference of between urban and rural regions ( $\Delta SOS$  and  
232  $\Delta EOS$ ) was calculated as below:

$$233 \Delta SOS = SOS_{(0)} - SOS_{(10)} \quad (4)$$

$$234 \Delta EOS = EOS_{(0)} - EOS_{(10)} \quad (5)$$

235 Among which,  $SOS_0$  and  $SOS_{10}$  is the mean SOS of the urban center and the 10-th  
236 buffer zone (20km away),  $EOS_0$  and  $EOS_{10}$  is the mean EOS of the urban center and  
237 the 10-th buffer zone (20km away).

238 The  $R_t$  of SOS and EOS ( $R_{t-SOS}$  and  $R_{t-EOS}$ ) difference between the urban and  
239 rural regions was defined as below:

$$240 \Delta R_{t-SOS} = R_{t-SOS(0)} - R_{t-SOS(20)} \quad (6)$$

$$241 \Delta R_{t-EOS} = R_{t-EOS(0)} - R_{t-EOS(20)} \quad (7)$$

242 Among which,  $R_{t-SOS(0)}$  and  $R_{t-SOS(20)}$  is the mean  $R_{t-SOS}$  of the urban center and the  
243 10-th buffer zone (20km away),  $R_{t-EOS(0)}$  and  $R_{t-EOS(20)}$  is the mean  $R_{t-EOS}$  of the urban  
244 center and the 10-th buffer zone (20km away).

## 245 **2.5 Background climates determined phenology response**

246 In order to elucidate the distinct impacts of LST and AI on urban and rural



247 components, a partial correlation analysis was executed. Specifically, the average LST  
248 or AI values recorded from 2010 to 2020 for each of the 293 cities were correlated  
249 with their respective  $\Delta$ SOS,  $\Delta$ EOS,  $\Delta R_{t-SOS}$ , and  $\Delta R_{t-EOS}$  values, while controlling for  
250 the influence of the other climatic factor.

251 This study employed path analysis (PCA) as a methodological approach to  
252 examine both the direct and indirect effects of background climate variables (LST and  
253 AI) on phenological responses (SOS and EOS), as well as their corresponding  
254 discrepancies ( $\Delta$ SOS and  $\Delta$ EOS) and rates of change ( $\Delta R_{t-SOS}$  and  $\Delta R_{t-EOS}$ ).

255 In the path analysis, we initially ascertained the model's congruence with  
256 empirical data by evaluating fit indices, including the chi-square ( $\chi^2$ ) statistic,  
257 Comparative Fit Index (CFI), and Root Mean Square Error of Approximation  
258 (RMSEA). Subsequently, path coefficients was derived for both the direct and indirect  
259 effects of each independent variable on the dependent variable from the model. The  
260 squared path coefficients for the direct and indirect paths estimated the respective  
261 direct and indirect contribution degrees of each independent variable to the dependent  
262 variable. The aggregate of these direct and indirect contribution degrees constitutes  
263 the total contribution degree, or relative contribution.

## 264 **2.6 Validation of phenology “Saturation effect”**

265 To determine if there was a temperature saturation effect on vegetation  
266 phenology, we initially assumed a logistic growth relationship between phenology and  
267 temperature. We then extracted phenological information from all vegetation pixels  
268 across the country and plotted scatter diagrams of phenology versus pre-season  
269 temperature. Next, we fitted logistic growth curves to these scatter plots and evaluated  
270 the assumption using the goodness-of-fit metrics p-value and  $R^2$ . Subsequently, we  
271 applied the same logistic curve to constrain the phenology of urban and suburban  
272 vegetation in 293 different cities against their pre-season temperatures, testing  
273 whether the phenological response to temperature in urban areas also follows a  
274 logistic growth pattern. Subsequently, we replaced the pre-season temperature with



275 the background temperature LST to determine how the phenological saturation effect  
276 is regulated by the background temperature.

### 277 **3. Results**

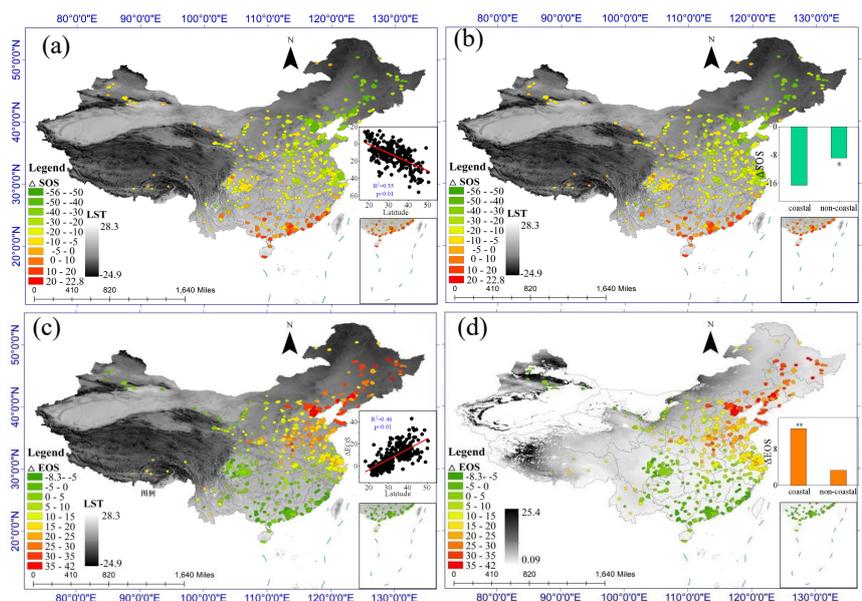
#### 278 **3.1 Spatial distribution of urban-rural phenological differences**

279 Among the 293 cities analyzed, 84% (N=246) exhibited an advanced SOS in  
280 urban areas compared to rural regions, while 82% (N=241) experienced a delayed  
281 EOS (Figure S3). Spatially, cities located in high-latitude regions generally  
282 demonstrated more negative  $\Delta$ SOS values than those in low-latitude regions (Figure  
283 2), indicating a stronger impact of urban warming on SOS in North and West China  
284 (Figure 2a, b). This pattern aligned with the conventional understanding of earlier  
285 SOS in colder regions. In addition, it seems that the cities in the coastal provinces  
286 have higher  $\Delta$ SOS than the other regions (Figure 2).

287 A strong positive correlation was observed between  $\Delta$ SOS and rural rate of SOS  
288 change ( $R_{t-SOS}$ ), suggesting that cities with more negative  $R_{t-SOS}$  values exhibit larger  
289 urban-rural disparities (Figure 3). It is noteworthy that cities with lower temperatures  
290 tended to have more negative  $R_{t-SOS}$  and  $\Delta$ SOS distributions.

291  $\Delta$ EOS exhibited an increasing trend with latitude gradient (Figure 2), with  
292 northern cities displaying more positive  $\Delta$ EOS values, indicating a delayed autumn  
293 phenology (Figure 2c, d). In contrast, the majority of cities in southern regions  
294 demonstrated negative  $\Delta$ EOS values, signifying an advancement in autumn phenology.  
295 Besides, cities in the coastal provinces also had higher  $\Delta$ EOS.

296 Similarly, when  $\Delta$ EOS was analyzed in relation to rural rate of EOS change  
297 ( $R_{t-EOS}$ ), a positive correlation was observed. Positive and negative  $\Delta$ EOS values were  
298 predominantly found in cold and warm regions, respectively (Figure 3). Moreover, as  
299  $\Delta$ EOS > 0, both the absolute value of  $\Delta$ EOS and  $\Delta R_{t-EOS}$  increased with temperature,  
300 while for  $\Delta$ EOS < 0, the absolute value of  $\Delta$ EOS and  $\Delta R_{t-EOS}$  decreased with  
301 temperature.



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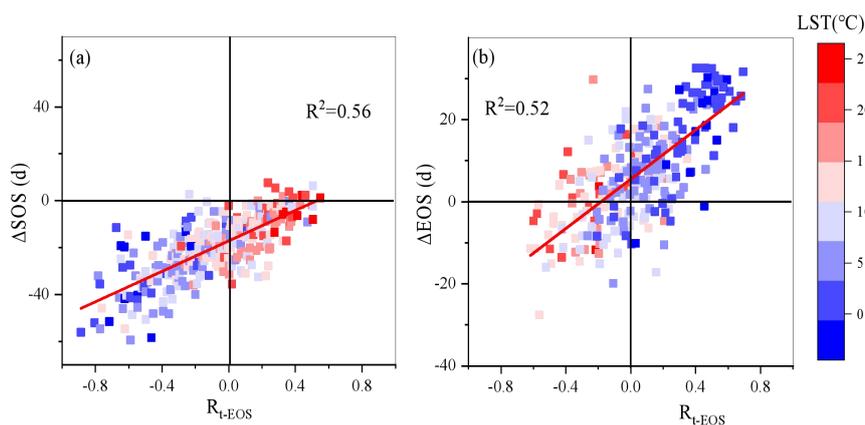
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**Figure 2.** The spatial distribution of urban-rural disparities for SOS ( $\Delta$ SOS) (a and b), EOS ( $\Delta$ EOS) (c and d) was analyzed across mainland China. T (a, c) and AI (b, d) were represented in black-and-white range for the entire continent, and the data in these figures reflected the 10-year mean from 2010 to 2020 for each city. The inset figure shown in (a, c) depicts the latitude and longitude corresponding to  $\Delta$ SOS and  $\Delta$ EOS, while the inset figure shown in (b, d) compares  $\Delta$ SOS and  $\Delta$ EOS in coastal provinces with those in non-coastal provinces. The map, generated using ArcGIS – version 10.2 (<https://www.esri.com/about/newsroom/arcwatch/the-best-of-arcgis-10-2/>), features administrative boundaries sourced from the Ministry of Civil Affairs of the People's Republic of China (<http://xzqh.mca.gov.cn/map>). Source: Esri, TomTom, FAO, USGS, and the GIS User Community; Ministry of Civil Affairs of the People's Republic of China | Powered by Esri.

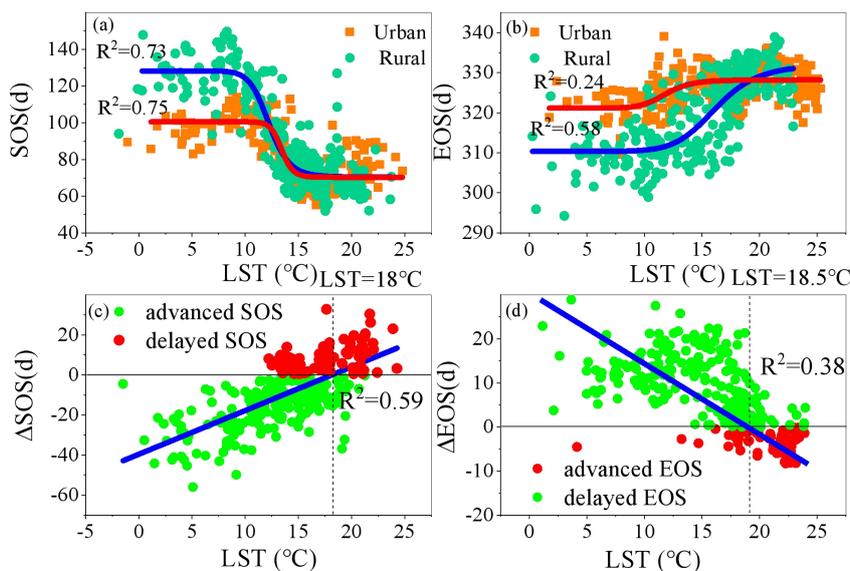


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317 Figure 3. The phenology sensitivity was fitted to its urban-rural difference. The LST  
 318 of each city was also presented by the red-blue ribbon.

### 319 3.2 phenology response to LST

320 As we compared urban and rural regions, logistic trends was observed for both  
 321 SOS and EOS in relation to LST(Figure 4). In urban areas where LST was below  
 322 12°C, the SOS occurred notably earlier compared to rural regions, while urban areas  
 323 exhibited a delayed EOS when LST was below 18.5°C. As LST decreased, the SOS  
 324 stabilized at 101.3 days for urban regions and 129.7 days for rural regions, while the  
 325 EOS stabilized at 321.3 days for urban areas and 310.8 days for rural areas.  
 326 Conversely, when LST exceeded 12°C or 18.5°C, the differences in SOS and EOS  
 327 between urban and rural areas were not significant and sometimes even reversed  
 328 (Figure 4a,b). As LST increased, the urban and rural SOS stabilized at 71.7 days  
 329 (Figure 4a), and the urban and rural EOS stabilized at 327.4 days and 332.4 days,  
 330 respectively. Fitting curve of  $\Delta$ SOS and  $\Delta$ EOS with LST revealed a positive and  
 331 negative relationship, respectively, intersecting at LST=18°C and 18.5°C, indicating a  
 332 shift in SOS and EOS dynamics with increasing temperatures beyond these  
 333 thresholds.



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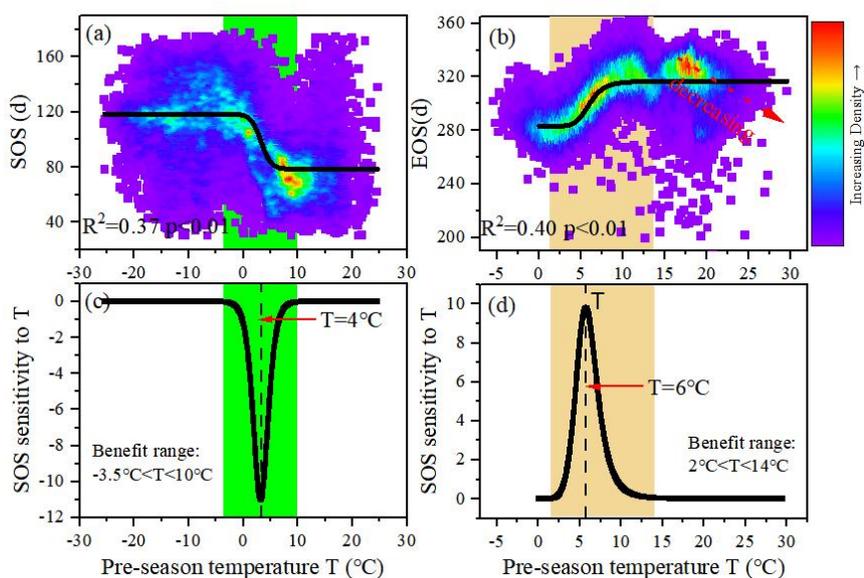
Figure 4. Relationship between background temperature (LST) and phenology (SOS

336

and EOS) across the urban-rural regions (a, b) from 2010 to 2020, and the relation

337

between  $\Delta$ SOS and  $\Delta$ EOS with the LST (c,d).



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Figure 5. The relationship between pre-season temperature (T) and SOS/EOS (a, b) as

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well as their First-order derivatives (implying the phenology sensitivity to T) of the

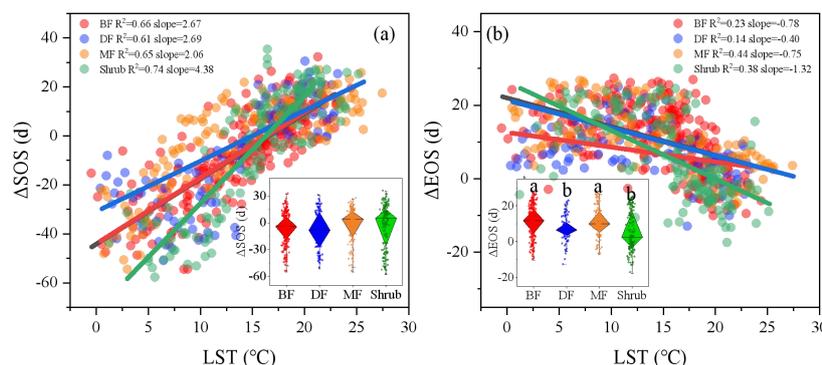
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fitting curve.



342 To determine whether the temperature sensitivity influenced the logistic trends of  
 343 phenology with LST, we conducted a nationwide fitting of phenology to T and also  
 344 derived logistic curves for SOS and EOS (Figure 5). The first-order derivatives of the  
 345 fitting curve that represented the phenology sensitivity to T revealed warming benefits  
 346 on SOS and EOS, ranging from  $-3.5^{\circ}\text{C}$  to  $10^{\circ}\text{C}$  and  $2.0$  to  $13.5^{\circ}\text{C}$ , respectively. The  
 347 highest sensitivity was observed at  $T=4^{\circ}\text{C}$  for SOS and  $T=6^{\circ}\text{C}$  for EOS. Notably, a  
 348 decreasing trend for EOS was observed when  $T>15^{\circ}\text{C}$ , corresponding to the advanced  
 349 EOS when  $T > 18.5^{\circ}\text{C}$  (Figure 5d).

350 To assess the temperature sensitivity of phenological variations between urban  
 351 and rural settings across vegetation types, we established regression models for  $\Delta\text{SOS}$ ,  
 352  $\Delta\text{EOS}$ , and LST for each type (Figure 6). Findings revealed that shrublands had the  
 353 highest temperature sensitivity for both metrics, whereas CF displayed the lowest.  
 354 The temperature response of MF and BF showed no significant variation.



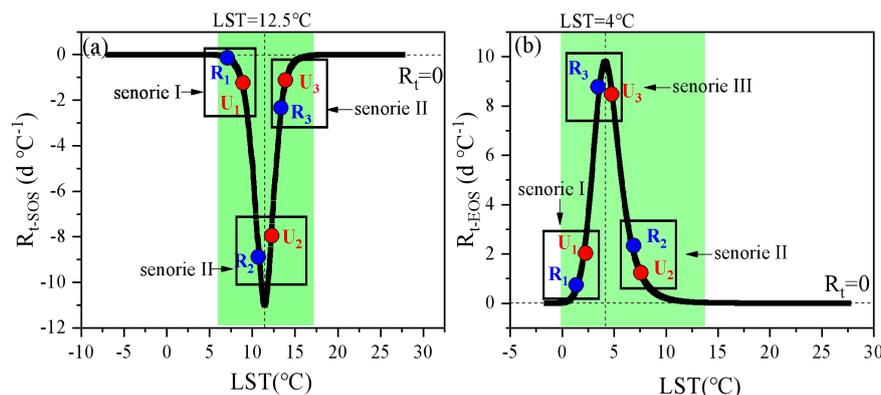
355  
 356 Figure 6. The relation between  $\Delta\text{SOS}$  and  $\Delta\text{EOS}$  with the LST across the 293 cities  
 357 for different vegetation types, insert figures presented the comparison of averaged  
 358 phenology difference across the 293 cities. BF: Broad-leaf forest, CF: Coniferous  
 359 forest, MF: Mixed forest, Shrubs: Bush wood.

### 360 3.3 Quantify the LST determined phenology sensitivity

361 In order to obtain LST determined phenology sensitivity, the T in the first-order  
 362 derivatives curves showed in Figure 5c, d was translated to LST considering the  
 363 strong correlation between T and LST (Figure S4). Based on the derived curves, three



364 possible scenarios for SOS and EOS sensitivity during urbanization across the  
 365 mainland was observed (Figure 7).



366  
 367 Figure 7. Schematic diagram for the three scenarios that likely occurred in the urban  
 368 (red dots) and rural regions (blue dots) for SOS (I:  $LST_{R1} < LST_{U1} < 12.5^{\circ}C$ ; II:  
 369  $12.5^{\circ}C < LST_{R2} < LST_{U2}$ ; III:  $LST_{R3} < 12.5^{\circ}C < LST_{U3}$ ) and EOS (I:  $LST_{R1} < LST_{U1} < 4^{\circ}C$ ; II:  
 370  $4^{\circ}C < LST_{R2} < LST_{U2}$ ; III:  $LST_{R3} < 4^{\circ}C < LST_{U3}$ ).

371 Scenario I: when LST was less than  $12.5^{\circ}C$  ( $4^{\circ}C$ ) for SOS (EOS), an increased  
 372 LST was associated with heightened temperature sensitivity, resulting in urban  
 373 regions ( $U_1$ ) exhibiting higher sensitivity than rural regions ( $R_1$ ) (Figure 5);

374 Scenario II: when LST exceeded  $12.5^{\circ}C$  ( $4^{\circ}C$ ) for SOS (EOS), there was a  
 375 weakened temperature sensitivity as LST increased, leading to urban regions ( $U_2$ )  
 376 demonstrating higher sensitivity than rural regions ( $R_2$ ) (Figure 5).

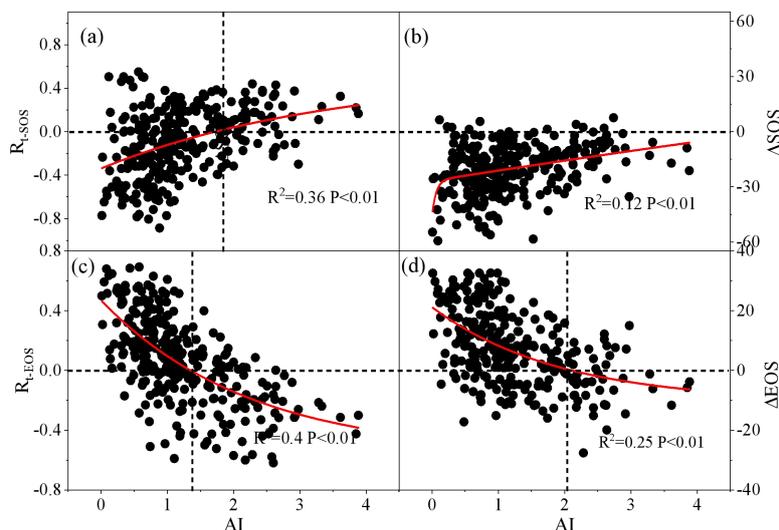
377 Scenario III: when urban LST surpassed  $12.5^{\circ}C$  ( $4^{\circ}C$ ) while rural LST remained  
 378 below  $12.5^{\circ}C$  ( $4^{\circ}C$ ), the difference in sensitivity between urban and rural regions was  
 379 inconclusive.

### 380 3.4 phenology response to aridity

381 When the phenological responses were fitted to the AI,  $R_{t-SOS}$  and  $\Delta SOS$  was  
 382 positive related to AI (Figure 8). It is noted that when  $AI < 1.7$ , there was a decrease in  
 383 absolute  $R_{t-SOS}$  with increasing AI, indicating that heightened aridity weakened the  
 384 temperature sensitivity of SOS. Conversely, when  $AI > 1.7$ , most urban areas exhibited  
 385 positive  $R_{t-SOS}$  (resulting in delayed SOS), with higher AI values corresponding to



386 greater  $R_{t-SOS}$ , suggesting that severe drought conditions passively delay the onset of  
 387 SOS.

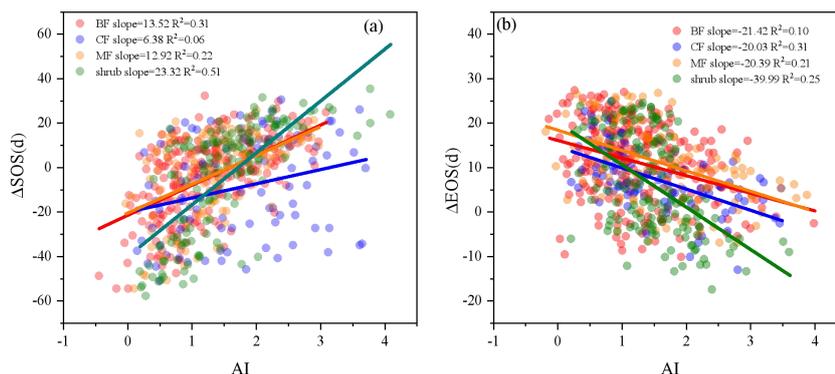


388

389 Figure 8. The AI determined phenology sensitivity and urban-rural difference

390 While  $R_{t-EOS}$  and  $\Delta EOS$  was fitted against AI, a negative relation was observed.  
 391 For  $R_{t-EOS}$ , when  $AI < 1.4$ , most of the  $R_{t-EOS}$  was positive, and decreased with AI, i.e.,  
 392 the enhanced drought will offset the benefit of urban warming on the delayed EOS.  
 393 When  $AI > 1.4$ , most cities shared a negative  $R_{t-EOS}$ , and the higher of AI, the more  
 394 negative of the  $R_{t-EOS}$ . I.e., extremely drought will lead to passive warming effects on  
 395 EOS. It is noted that the negative  $\Delta EOS$  tended to appeared when  $AI > 2$ .

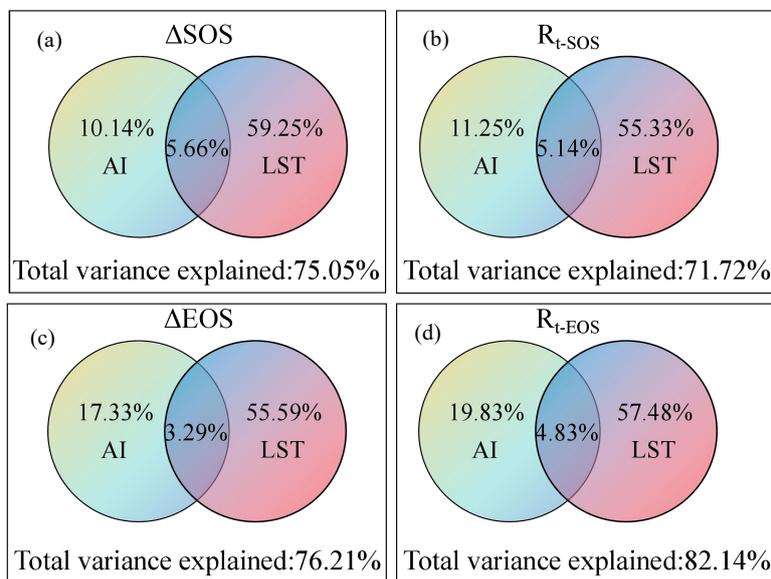
396 AI has been observed to advance  $\Delta SOS$  and hasten  $\Delta EOS$  across 293 cities,  
 397 with this pattern holding true for all four vegetation types (Figure 9). Notably,  $\Delta SOS$   
 398 in CF showed less sensitivity to AI influences than the other vegetation types. Both  
 399 BF and MF demonstrated lower sensitivities to AI for both  $\Delta SOS$  and  $\Delta EOS$ , as  
 400 compared to shrublands.



401

402 Figure 9. The AI determined phenology sensitivity and urban-rural difference for

403 different vegetation types.



404

405 Figure 10. Venn diagrams illustrating the relative contribution of AI and LST to  $\Delta$ SOS,

406  $\Delta$ EOS,  $R_{t-SOS}$ ,  $R_{t-EOS}$  across 293 cities in China. Each number in the panel indicates

407 that relative percentage explained by the combined factors.

408 The Venn diagram revealed that AI and LST explained 75.05%, 71.72%, 76.21%,

409 82.14% for  $\Delta$ SOS,  $\Delta$ EOS,  $R_{t-SOS}$ ,  $R_{t-EOS}$  respectively (Figure 10), and LST always had

410 more contributions for these phenology indicators. Besides, the interaction effect of

411 LST and AI ranged from 3.29%-5.66%.



412 **4. Discussion**

413 **4.1 Overall urbanization effects on plant phenology in China**

414 In this study, the impact of urbanization on plant phenology across 293 cities in  
415 China was assessed. Our findings demonstrated a consistent advancement of SOS and  
416 delay of EOS in urban areas compared to rural regions in most of the studied cities  
417 (Figure S1), as argued in the previous studies (Jeonget et al., 2019; Meng et al., 2020;  
418 Qiu et al., 2020; Zhou et al., 2016). Notably, the urban warming effect on plant  
419 phenology was predominantly observed within a 10-km radius from the city centers  
420 (Figure S3), validating our definition of rural areas as those located over 20-km from  
421 the city centers.

422 Spatially, variations in  $\Delta$ SOS and  $\Delta$ EOS progressed from the south towards the  
423 north of China, mirroring the land surface temperature (LST) distribution across the  
424 continent (Figure 2). This aligns with previous research suggesting that the northern  
425 regions experience greater shifts in plant phenology than southern regions (Meng et al.,  
426 2020; Jia et al., 2021). Additional analysis showed that cities in the coastal provinces  
427 enjoyed more pronounced warming benefits compared to inland cities (Figure 2).  
428 Moreover, a strong correlation was found between delta SOS ( $\Delta$ SOS), delta EOS  
429 ( $\Delta$ EOS), and temperature sensitivity indicators ( $R_{t-SOS}$  and  $R_{t-EOS}$ ), as well as LST and  
430 AI (Figure 3). Thus, the expected impact of the background temperature and and  
431 aridity on the phenology response across the 293 cites was confirmed.

432 It is noted that the pattern of urban warming's benefit on the vegetation  
433 phenology was consistent among the four vegetation types (Figure S2). This  
434 consistency could be attributed to the ecological convergence for plants coexist in the  
435 same habitation (Wang et al., 2016), and within-spring warming have potential to  
436 intensify this phenological synchrony among different species (Wang et al., 2016).



## 437 **4.2 Background temperature related phenological response variance**

438 When we quantified phenological responses to temperature, we expected a  
439 logistic growth model to fit both the SOS and EOS with rising temperatures (Figure 5).  
440 Interestingly, in urban areas, the growth rate for SOS/EOS began to plateau when  
441 temperatures exceeded 4°C/6°C respectively (Figure 5). Taking into account the  
442 significant correlation between LST and T (Figure S4), we could also anticipate a  
443 diminished sensitivity in phenology with higher LST (Figure 7). These findings lend  
444 credibility to the "saturation" effect hypothesis, suggesting that the accelerating  
445 influence of warmth on vegetation phenology reported in earlier studies has its limits  
446 (Badeck et al., 2004; Li et al., 2021). Indeed, plants in the southern region exhibit not  
447 just a dampened phenological response but also a decelerated growth rate when faced  
448 with increasing temperatures(Li et al., 2021; Buntgen et al., 2019). Collectively, these  
449 results underscore the importance of considering non-linear and threshold-based  
450 models when predicting phenological responses under scenarios of continuous  
451 environmental warming, especially in urban settings.

452 The observation of temperature saturation for the SOS in our study likely stems  
453 from the chilling requirements for vernalization(Kim et al., 2009; Hanninen et al.,  
454 2019). Typically, following a halt in growth during the early fall, vegetation needs to  
455 undergo a period of exposure to low chilling temperatures to exit dormancy. However,  
456 Nonetheless, when ambient temperatures are relatively high, these chilling  
457 requirements may not be met(Meng et al., 2020), thus diminished phenological  
458 sensitivity in warmer territories (Figure 2,3). Moreover, an earlier onset of greening  
459 increased the risk of damage from freezing events, posing a survival threat to the  
460 plants(Duan et al., 2011; Wang et al., 2015).This phenomenon had led to a delayed  
461 SOS in approximately 16% of the cities under study, predominantly located in lower  
462 latitude areas (most in the low latitude regions, Figure 4).

463 Regarding the EOS, a moderate increase in temperature could enhance the  
464 activities of photosynthetic enzymes(Shi et al., 2014), slow down the degradation of  
465 chlorophyll(Fracheboud et al., 2009), reduce the risk of frost exposure in autumn, and



466 increase the potential for growth and photosynthetic consumption, ultimately  
467 promoting the extension of the growing season. While the benefits of increased  
468 temperature can be offset by the negative impact of heat waves and droughts that  
469 result from higher LST. These passive feedback loops can cause widespread leaf  
470 senescence and even crown defoliation<sup>50-51</sup>(Estiarte & Penuelas, 2015; Pollastrini et  
471 al., 2017), potentially negating the benefits of an earlier onset of the growing season.  
472 In fact, when  $T > 15^{\circ}\text{C}$ , EOS tends to be reduced, with an increased sensitivity to  
473 temperature changes (Figure 5b).

474 While even the existence of “saturation” effects, the lower background  
475 temperatures was not necessarily correspond to higher phenological sensitivity  
476 (Figure 4b, c). For SOS, when  $T < 4^{\circ}\text{C}$ , the phenological sensitivity was approaching  
477 zero with decreased  $T$  (Figure 4c), which led to a stable SOS of 120d (Figure 5a,c). In  
478 fact, phenological events such as bud burst could not be triggered when high forcing  
479 temperatures threshold (cumulative temperature) was not fully provided<sup>46</sup>(Hanninen  
480 et al., 2019). For EOS, when  $T > 6^{\circ}\text{C}$ , the increased  $T$  sensitivity will produce more  
481 EOS advancement in the warmer regions. This could be related to the associated with  
482 the increase of abscisic acid synthesis induced by low temperature, which accelerated  
483 the progress of defoliation.

484 In our study, when we quantify the phenology sensitivity response to the LST, we  
485 found that the phenological sensitivity in response the LST was not unidirectional, but  
486 fluctuant, for both SOS and EOS. For SOS, the positive phenological response in the  
487 urban regions was validated  $6 < \text{LST} < 17.5^{\circ}\text{C}$  (Figure 5). However, when  $\text{LST} < 12.5^{\circ}\text{C}$ ,  
488 warming regions would have higher urban-rural difference of  $\Delta R_{t-\text{SOS}}$ , which  
489 contradicted the weakened urban-rural difference of  $\Delta R_{t-\text{SOS}}$  raised in the previous  
490 study(Meng et al., 2020). For EOS, the warming benefits only occurred when  
491  $0 < \text{LST} < 18^{\circ}\text{C}$ . Moreover, the weakened sensitivity only in the LST interval of  
492  $\text{LST} > 4^{\circ}\text{C}$ . The conclusions of weakened  $\Delta R_{t-\text{SOS}}$  and  $\Delta R_{t-\text{EOS}}$  in the USA could  
493 attributed to their LST ranges was mostly falling in the ranges of  $8^{\circ}\text{C}-22^{\circ}\text{C}$ (Meng et  
494 al., 2020).



#### 495 **4.3 AI related phenological response variance**

496 Except for the LST, the AI related spatial variation of  $\Delta$ SOS and  $\Delta$ EOS was also  
497 validated. Spatially, the coastal cities experienced more benefits from the urban  
498 warming(Figure 2). Statistically, both  $R_{t-SOS}$  and  $\Delta$ SOS increased with AI, while both  
499  $R_{t-SOS}$  and  $\Delta$ SOS decreased with AI. Consequently, AI explained 10.14% and 17.33%  
500 variation dependently for the spatial distribution of  $\Delta$ SOS and  $\Delta$ EOS. These results  
501 highlight that the warming benefits on the vegetation phenology will be weaken in  
502 those drought dominated regions. In fact, the delayed SOS was already proved in the  
503 previous studies(Ji et al., 2021; Yu et al., 2023).

504 It has been suggested that under the condition of water deficit, roots of plants  
505 would produce chemical signals such as abscisic acid and cytokinins, which may limit  
506 leaf growth(Knauer et al., 2017). Urbanization have potential to intensify the drought  
507 condition(Zhang et al., 2019), thus weaken and even reverse the warming benefits.  
508 Besides, in water-stressed region, temperatures are already close to the optimum for  
509 photosynthesis, thus less benefits will be expected(Penuelas et al., 2004). The  
510 weakened temperature sensitivity in drought cities will certainly led to less extension  
511 of growth season.

512 While it is noted that the influence of AI on the autumn phenology response was  
513 higher than that on the spring phenology response (Figure 10). During spring leaf  
514 development, plants may use water stored in their bodies to support the growth of new  
515 leaves(Shi et al., 2023). Evidence had been proved by the synchronous in plant water  
516 storage and leaf sprout for the boreal and temperate forests(Tian et al., 2018). In this  
517 case, the external water supply may have relatively little demand on the plant during  
518 the leaf development period, especially in some wetter areas. Comparatively, In  
519 summer and autumn, the transpiration of plants is larger, and the demand for water is  
520 strong(Wu et al., 2022), so the influence of drought degree on plant phenology will be  
521 more prominent.



#### 522 **4.4 Phenological response variation among different vegetation type**

523 In the response to variations in temperature and the Aridity Index (AI),  
524 significant differences in sensitivity are observed among different vegetation types.  
525 Shrub species exhibit heightened sensitivity, whereas coniferous tree species display  
526 relative insensitivity. These discrepancies reflect the influence of functional traits on  
527 the phenology of vegetation. Studies on forests in northern China indicate that  
528 broadleaf forests and mixed forests are the most sensitive to climatic factors, followed  
529 by coniferous forests (Zheng et al., 2022). Broadleaf tree species, with their higher  
530 demands for light and temperature (Wu et al., 2015), are more likely to benefit from  
531 warming temperatures, thereby promoting their growth (Malla et al., 2023). However,  
532 as AI increases, broadleaf species face more severe water limitations. Coniferous  
533 species, due to their drought-resistant characteristics such as needle morphology,  
534 stomatal recess, and weaker water transport capacity, exhibit lower sensitivity to  
535 changes in AI, which limits the potential positive impact of warming on their  
536 phenology. In contrast, the extensive resource utilization strategy of shrub species  
537 may be the primary reason for their high sensitivity to climate change (Xiong et al.,  
538 2024).

#### 539 **5. Conclusion**

540 Our study utilized remote sensing data to investigate the impact of urbanization  
541 on vegetation phenology in 293 cities from 2010 to 2020. The advanced SOS and  
542 delayed EOS was validated in most cities due to urban warming effects, however,  
543 with different sensitivity due to the spatial variance of LST and AI.

544 The assumption that the “saturation effect” of plant phenological response to  
545 warming can explain the gradually reduced urban SOS and EOS shift in warming  
546 regions when  $LST < 12.5\text{ }^{\circ}\text{C}$  and  $LST > 4\text{ }^{\circ}\text{C}$  respectively. Otherwise, warmer regions  
547 would experience much more benefits due to urban warming. The warming benefit  
548 would valid only when  $3.5\text{ }^{\circ}\text{C} < LST < 17.5\text{ }^{\circ}\text{C}$ , and  $0\text{ }^{\circ}\text{C} < LST < 18\text{ }^{\circ}\text{C}$ , otherwise, the



549 phenology sensitivity will approach 0, and moreover, extremely high temperatures has  
550 the potential to reverse the warming benefits, leading to weakly delayed or even  
551 advanced EOS in warming regions.

552 Drought also has potential to weaken the urban warming benefit on SOS and  
553 EOS, which leading to higher  $\Delta$ SOS and  $\Delta$ EOS in the coastal cities. When  $AI < 1.4$  or  
554  $AI > 2.0$ , increased AI will weaken the warming benefits for SOS and EOS, while  
555 when  $1.4 < AI < 2.0$ , warming benefits was even reversed for SOS and EOS.

## 556 **6. Author contributions: CRediT**

557 Cui Shufen: Conceptualization, Methodology; Zhang Zhenzhen: Data curation,  
558 Writing- Original draft preparation; Zhang Yongqi: Visualization, Investigation; Sun  
559 Liheng, Wu Chaofan and Lin Xingwen: Supervision; Zhang Zhaoyang, Chen  
560 Yuanjian and Wen Qingqing: Software, Validation.

## 561 **7. Data availability statement**

562 The data that support the findings of this study are available in those public domain  
563 resources that listed in the table 1.

## 564 **8. Conflict of Interest**

565 The authors declared that the research was conducted in the absence of any  
566 commercial or financial relationships that could be construed as a potential conflict of  
567 interest.

## 568 **Founding**

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## 572      **References**

- 573 Badeck, F.-W., Bondeau, A., Böttcher, K., Doktor, D., Lucht, W., Schaber, J. and Sitch, S.  
574      (2004). Responses of spring phenology to climate change. *New Phytologist*. **162**,  
575      295-309.
- 576 Buyantuyev, A., & Wu, J. (2012). Urbanization diversifies land surface phenology in arid  
577      environments: Interactions among vegetation, climatic variation, and land use pattern  
578      in the Phoenix metropolitan region, USA. *Landscape and Urban Planning*. **105**,  
579      149-159.
- 580 Büntgen, U., Krusic, P.J., Piermattei, A., Coomes, D.A., Esper, J., Myglan, V.S.,  
581      Kirilyanov, A.V., Camarero, J.J., Crivellaro, A., Körner, C. (2019). Limited capacity  
582      of tree growth to mitigate the global greenhouse effect under predicted warming.  
583      *Nature Communication*. **10**, 2171.
- 584 Cai, Z., Jönsson, P., Jin, H., & Eklundh, L. (2017). Performance of smoothing methods for  
585      reconstructing NDVI time-Series and estimating vegetation phenology from MODIS  
586      Data. *Remote Sensing*. **9**, 1271.
- 587 Čehulić, I., Sever, K., Katičić Bogdan, I., Jazbec, A., Škvorec, Ž., & Bogdan, S. (2019).  
588      Drought Impact on Leaf Phenology and Spring Frost Susceptibility in a *Quercus*  
589      *robur* L. Provenance Trial. *Forests*. **10**, 50.
- 590 Chmielewski, F.-M., & Rötzer, T. (2001). Response of tree phenology to climate change  
591      across Europe. *Agricultural and Forest Meteorology*. **108**, 101-112.
- 592 Cong, N., Piao, S., Chen, A., Wang, X., Lin, X., Chen, S., Han, S., Zhou, G., Zhang, X.  
593      (2012). Spring vegetation green-up date in China inferred from SPOT NDVI data: A  
594      multiple model analysis. *Agricultural and Forest Meteorology*. **165**, 104-113.
- 595 Dallimer, M., Tang, Z., Gaston, K. J., & Davies, Z. G. (2016). The extent of shifts in  
596      vegetation phenology between rural and urban areas within a human-dominated  
597      region. *Ecology and Evolution*. **6**, 1942-1953.
- 598 Ding, H., Xu, L., Elmore, A. J., & Shi, Y. (2020). Vegetation phenology influenced by  
599      rapid urbanization of The Yangtze Delta region. *Remote Sensing*. **12**, 1783.
- 600 Duan, J., Zhang, Q.-B., Lv, L., & Zhang, C. (2011). Regional-scale winter-spring



- 601 temperature variability and chilling damage dynamics over the past two centuries in  
602 southeastern China. *Climate Dynamics*. **39**, 919-928.
- 603 Eidt, D. C., & Little, C. J. C. E. (1968). Insect control by artificially prolonging plant  
604 dormancy — a new approach. *The Canadian Entomologist*. **100**, 1278-1279.
- 605 Estiarte, M., & Penuelas, J. (2015). Alteration of the phenology of leaf senescence and fall  
606 in winter deciduous species by climate change: effects on nutrient proficiency. *Global  
607 Change Biology*. **21**, 1005-1017.
- 608 Fan, J., He, H., Hu, T., Zhang, P., Yu, X., & Zhou, Y. (2019). Estimation of landscape  
609 pattern changes in BRICS from 1992 to 2013 using DMSP-OLS NTL images.  
610 *Journal of the Indian Society of Remote Sensing*. **47**, 725-735.
- 611 Fracheboud, Y., Luquez, V., Bjorken, L., Sjodin, A., Tuominen, H., & Jansson, S. (2009).  
612 The control of autumn senescence in European aspen. *Plant Physiology*. **149**,  
613 1982-1991.
- 614 Gazal, R., White, M. A., Gillies, R., Rodemaker, E. L. I., Sparrow, E., & Gordon, L.  
615 (2008). GLOBE students, teachers, and scientists demonstrate variable differences  
616 between urban and rural leaf phenology. *Global Change Biology*. **14**, 1568-1580 .
- 617 Ge, W., Han, J., Zhang, D., & Wang, F. (2021). Divergent impacts of droughts on  
618 vegetation phenology and productivity in the Yungui Plateau, southwest China.  
619 *Ecological Indicators*. **127**, 107743.
- 620 Gow, L. J., Barrett, D. J., Renzullo, L. J., Phinn, S. R., & O'Grady, A. P. (2016).  
621 Characterising groundwater use by vegetation using a surface energy balance model  
622 and satellite observations of land surface temperature. *Environmental Modelling &  
623 Software*. **80**, 66-82.
- 624 Hanninen, H., Kramer, K., Tanino, K., Zhang, R., Wu, J., & Fu, Y. H. (2019). Experiments  
625 are necessary in process-based tree phenology modelling. *Trends in Plant Science*. **24**,  
626 199-209.
- 627 He, C., Gao, B., Huang, Q., Ma, Q., & Dou, Y. (2017). Environmental degradation in the  
628 urban areas of China: Evidence from multi-source remote sensing data. *Remote  
629 Sensing of Environment*. **193**, 65-75.
- 630 Jeong, S. J., Park, H., Ho, C. H., & Kim, J. (2019). Impact of urbanization on spring and



- 631 autumn phenology of deciduous trees in the Seoul Capital Area, South Korea.  
632 *International Journal of Biometeorology*. **63**, 627-637.
- 633 Ji, S., Ren, S., Li, Y., Dong, J., Wang, L., Quan, Q., & Liu, J. (2021). Diverse responses of  
634 spring phenology to preseason drought and warming under different biomes in the  
635 North China Plain. *Science of the Total Environment*. **766**, 144437.
- 636 Jia, W., Zhao, S., Zhang, X., Liu, S., Henebry, G. M., & Liu, L. (2021). Urbanization  
637 imprint on land surface phenology: The urban-rural gradient analysis for Chinese  
638 cities. *Global Chang Biology*. **27**, 2895-2904.
- 639 Kang, X., Hao, Y., Cui, X., Chen, H., Huang, S., Du, Y., Li, W., Kardol, P., Xiao, X., Cui,  
640 L. (2016). Variability and changes in climate, phenology, and gross primary  
641 production of an Alpine wetland ecosystem. *Remote Sensing*. **8**, 391.
- 642 Ketola, T., & Kristensen, T. N. (2017). Experimental approaches for testing if tolerance  
643 curves are useful for predicting fitness in fluctuating environments. *Frontiers in*  
644 *Ecology and Evolution*. **5**, 129.
- 645 Kim, D. H., Doyle, M. R., Sung, S., & Amasino, R. M. (2009). Vernalization: winter and  
646 the timing of flowering in plants. *Annual Review of Cell and Developmental Biology*.  
647 **25**, 277-299.
- 648 Knauer, J., Zaehle, S., Reichstein, M., Medlyn, B. E., Forkel, M., Hagemann, S., & Werner,  
649 C. (2017). The response of ecosystem water-use efficiency to rising atmospheric CO  
650 2 concentrations: sensitivity and large-scale biogeochemical implications. *New*  
651 *Phytologist*. **213**, 1654-1666.
- 652 Li, K., Wang, C., Sun, Q., Rong, G., Tong, Z., Liu, X., & Zhang, J. (2021). Spring  
653 phenological sensitivity to climate Change in the northern hemisphere:  
654 comprehensive evaluation and driving force analysis. *Remote Sensing*. **13**, 1972.
- 655 Malla, R., Neupane, P. R., & Köhl, M. (2023). Climate change impacts: Vegetation shift of  
656 broad-leaved and coniferous forests. *Trees, Forests and People*. **14**, 100457 .
- 657 Mathur S., Agrawal D., Jajoo A. (2014). Photosynthesis: response to high temperature  
658 stress. *Journal of Photochemistry and Photobiology B: Biology*. **137**, 116-126.
- 659 Mcadam, B. S. A. M. J. P. P. (2013). Abscisic acid mediates a divergence in the drought  
660 response of two conifers. *Plant physiology*. **162**, 1370-1377.



- 661 Meng, L., Mao, J., Zhou, Y., Richardson, A.D., Lee, X., Thornton, P.E., Ricciuto, D.M., Li,  
662 X., Dai, Y., Shi, X., Jia, G. (2020). Urban warming advances spring phenology but  
663 reduces the response of phenology to temperature in the conterminous United States.  
664 Proceedings of the National Academy of Sciences of the United States of America.  
665 **117**, 4228-4233.
- 666 Pablos, M., Martínez-Fernández, J., Piles, M., Sánchez, N., Vall-llossera, M., & Camps, A.  
667 (2016). Multi-temporal evaluation of soil moisture and land surface temperature  
668 dynamics using in situ and satellite observations. Remote Sensing. **8**, 587.
- 669 Peng, J., Wu, C., Zhang, X., Wang, X., & Gonsamo, A. (2019). Satellite detection of  
670 cumulative and lagged effects of drought on autumn leaf senescence over the  
671 Northern Hemisphere. Global Change Biology. **25**, 2174-2188.
- 672 Peng, S. (2020). 1-km monthly precipitation dataset for China (1901-2021). Retrieved  
673 from: <https://dx.doi.org/10.5281/zenodo.3185722>.
- 674 Penuelas, J., Gordon, C., Llorens, L., Nielsen, T., Tietema, A., Beier, C., ... & Gorissen, A.  
675 (2004). Nonintrusive field experiments show different plant responses to warming  
676 and drought among sites, seasons, and species in a north–south European gradient.  
677 Ecosystems. **7**, 598-612.
- 678 Piao, S., Fang, J., Zhou, L., Ciais, P., & Zhu, B.(2006). Variations in satellite-derived  
679 phenology in China's temperate vegetation. Global Change Biology. **12**, 672-685.
- 680 Pollastrini, M., Nogales, A.G., Benavides, R., Bonal, D., Finer, L., Fotelli, M., Gessler, A.,  
681 Grossiord, C., Radoglou, K., Strasser, R.J., Bussotti, F. (2017). Tree diversity affects  
682 chlorophyll a fluorescence and other leaf traits of tree species in a boreal forest. Tree  
683 Physiology. **37**, 199-208.
- 684 Pollastrini, M., Puletti, N., Selvi, F., Iacopetti, G., & Bussotti, F. (2019). Widespread  
685 crown defoliation after a drought and heat wave in the forests of Tuscany (Central  
686 Italy) and their recovery—A case study from summer 2017. Frontiers in Forests and  
687 Global Change. **2**.
- 688 Qiao, Z., Tian, G., & Xiao, L. (2013). Diurnal and seasonal impacts of urbanization on the  
689 urban thermal environment: A case study of Beijing using MODIS data. ISPRS  
690 Journal of Photogrammetry and Remote Sensing. **85**, 93-101.



- 691 Qiu, T., Song, C., & Li, J. (2017). Impacts of urbanization on vegetation phenology over  
692 the past three decades in Shanghai, China. *Remote Sensing*, **9**, 970.
- 693 Qiu, T., Song, C., Zhang, Y., Liu, H., & Vose, J. M. (2020). Urbanization and climate  
694 change jointly shift land surface phenology in the northern mid-latitude large cities.  
695 *Remote Sensing of Environment*, **236**, 111477.
- 696 Rustad, L., Campbell, J., Marion, G., Norby, R., Mitchell, M., Hartley, A., Cornelissen, J.,  
697 Gurevitch, J., Gcte, N. (2001). A meta-analysis of the response of soil respiration, net  
698 nitrogen mineralization, and aboveground plant growth to experimental ecosystem  
699 warming. *Oecologia*, **126**, 543–562.
- 700 Sage, R.F. and Kubien, D.S. (2007). The temperature response of C3 and C4  
701 photosynthesis. *Plant, Cell & Environment*, **30**, 1086-1106.
- 702 Shi, C., Sun, G., Zhang, H., Xiao, B., Ze, B., Zhang, N., & Wu, N. (2014). Effects of  
703 warming on chlorophyll degradation and carbohydrate accumulation of Alpine  
704 herbaceous species during plant senescence on the Tibetan Plateau. *PLoS One*, **9**,  
705 e107874.
- 706 Shi, W., Li, J., Zhan, H., Yu, L., Wang, C., & Wang, S. (2023). Relation between water  
707 storage and photoassimilate accumulation of *neosinocalamus affinis* with phenology.  
708 *Forests*, **14**, 531.
- 709 Tang, J. W., Korner, C., Muraoka, H., Piao, S. L., Shen, M. G., Thackeray, S. J., & Yang,  
710 X. (2016). Emerging opportunities and challenges in phenology: a review. *Ecosphere*,  
711 **7**, e01436.
- 712 Tian, F., Wigneron, J.-P., Ciais, P., Chave, J., Ogée, J., Peñuelas, J., Ræbild, A., Domec,  
713 J.-C., Tong, X., Brandt, M., Mialon, A., Rodriguez-Fernandez, N., Tagesson, T.,  
714 Al-Yaari, A., Kerr, Y., Chen, C., Myneni, R.B., Zhang, W., Ardö, J., Fensholt, R.  
715 (2018). Coupling of ecosystem-scale plant water storage and leaf phenology observed  
716 by satellite. *Nature ecology & evolution*, **2**, 1428-1435.
- 717 Wan, Z. (2008). New refinements and validation of the MODIS Land-Surface  
718 Temperature/Emissivity products. *Remote Sensing of Environment*, **112**, 59-74 .
- 719 Wang, C., Cao, R., Chen, J., Rao, Y., & Tang, Y. (2015). Temperature sensitivity of spring  
720 vegetation phenology correlates to within-spring warming speed over the Northern



- 721 Hemisphere. *Ecological Indicators*. **50**, 62-68.
- 722 Wang, C., Tang, Y., & Chen, J. (2016). Plant phenological synchrony increases under rapid  
723 within-spring warming. *Scientific Reports*. **6**, 25460.
- 724 Wang, Z., Ma, R., Zhao, M., Wang, F., Zhang, N., & Si, H. (2020). NO and ABA  
725 interaction regulates tuber dormancy and sprouting in Potato. *Frontiers in Plant*  
726 *Science*. **11**, 311.
- 727 White, M.A., De Beurs, K.M., Didan, K., Inouye, D.W., Richardson, A.D., Jensen, O.P.,  
728 O'Keefe, J., Zhang, G., Nemani, R.R., Van Leeuwen, W.J.D., Brown, J.F., De Wit, A.,  
729 Schaepman, M., Lin, X., Dettinger, M., Bailey, A.S., Kimball, J., Schwartz, M.D.,  
730 Baldocchi, D.D., Lee, J.T., Lauenroth, W.K. (2009). Intercomparison, interpretation,  
731 and assessment of spring phenology in North America estimated from remote sensing  
732 for 1982-2006. *Global Change Biology*. **15**, 2335-2359 .
- 733 Wu, C., Peng, J., Ciais, P., Peñuelas, J., Wang, H., Beguería, S., ... & Ge, Q. (2022).  
734 Increased drought effects on the phenology of autumn leaf senescence. *Nature*  
735 *Climate Change*. **12**, 943-949.
- 736 Wu, D., Zhao, X., Liang, S., Zhou, T., Huang, K., Tang, B., & Zhao, W. (2015). Time-lag  
737 effects of global vegetation responses to climate change. *Global change biology*. **21**,  
738 3520-3531.
- 739 Xiong, X., Wu, H., Wei, X., & Jiang, M. (2024). Contrasting temperature and light  
740 sensitivities of spring leaf phenology between understory shrubs and canopy trees:  
741 Implications for phenological escape. *Agricultural and Forest Meteorology*. **355**,  
742 110144.
- 743 Yao, R., Wang, L., Huang, X., Guo, X., Niu, Z., & Liu, H. (2017). Investigation of  
744 Urbanization effects on land Surface phenology in Northeast China during  
745 2001–2015. *Remote Sensing*. **9**, 66.
- 746 Yu, F., Price, K. P., Ellis, J., & Shi, P. (2003). Response of seasonal vegetation  
747 development to climatic variations in eastern central Asia. *Remote sensing of*  
748 *environment*. **87**, 42-54.
- 749 Yuan, M., Zhao, L., Lin, A., Wang, L., Li, Q., She, D., & Qu, S. (2020). Impacts of  
750 pre-season drought on vegetation spring phenology across the Northeast China



- 751 Transect. *Science of the Total Environment*. **738**, 140297.
- 752 Zhang, X., Friedl, M. A., Schaaf, C. B., Strahler, A. H., & Schneider, A. (2004). The  
753 footprint of urban climates on vegetation phenology. *Geophysical Research Letters*.  
754 **31**, L12209.
- 755 Zhang, X., Friedl, M. A., & Schaaf, C. B. (2006). Global vegetation phenology from  
756 Moderate Resolution Imaging Spectro radiometer (MODIS): Evaluation of global  
757 patterns and comparison with in situ measurements. *Journal of Geophysical Research:*  
758 *Biogeosciences*. **111**.
- 759 Zhang, X., Chen, N., Sheng, H., Ip, C., Yang, L., Chen, Y., ... & Niyogi, D. (2019). Urban  
760 drought challenge to 2030 sustainable development goals. *Science of the Total*  
761 *Environment*. **693**, 133536.
- 762 Zheng, W., Liu, Y., Yang, X., & Fan, W. (2022). Spatiotemporal variations of forest  
763 vegetation phenology and its response to climate change in northeast China. *Remote*  
764 *Sensing*. **14**, 2909.
- 765 Zhou, D., Zhao, S., Zhang, L., & Liu, S. (2016). Remotely sensed assessment of  
766 urbanization effects on vegetation phenology in China's 32 major cities. *Remote*  
767 *Sensing of Environment*. **176**, 272-281.
- 768