



1 **Constraining boreal carbon allocation and turnover by assimilating** 2 **forest growth dynamics in a differentiable framework**

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17 **Abstract**

18 Terrestrial biosphere models (TBMs) exhibit substantial uncertainty in simulating the land carbon sink in particular at
19 interannual to decadal time scales, partly because parameters that govern carbon allocation and biomass turnover rates
20 are weakly constrained. Typical model-data fusion approaches, which rely heavily on high-frequency (minutes to days)
21 observations related to fluxes (e.g., gross primary production and leaf area index), often struggle to constrain the slow
22 turnover processes that govern long-term biomass accumulation over multiple years. Here, we employ DifferLand, a
23 JAX-based differentiable TBM, to jointly assimilate satellite-derived boreal forest biomass growth trajectories and
24 high-frequency observations. Calibrations using only high-frequency fluxes reproduce short-term dynamics but yield
25 large biases in mature forest biomass (RMSE = 138.7 Mg ha⁻¹), while incorporating a single-year biomass stock
26 constraint only partially reduces the error (RMSE = 87.5 Mg ha⁻¹) and provides limited constraints on the allocation
27 patterns. In contrast, incorporating our biomass growth curves reduce the biomass RMSE to 11.3 Mg ha⁻¹ (a 91.9%
28 reduction) without degrading fits to high-frequency fluxes. Our findings reveal that while a single-year biomass stock
29 data provides some constraints on biomass residence times, the full growth trajectory is essential to simultaneously
30 constrain carbon allocation and turnover. The retrieved parameters indicate that boreal forests sustain biomass primarily
31 through longer wood carbon residence times (74.1% lower wood turnover rates) rather than higher allocation to wood,
32 compared to calibrations using only high-frequency flux observations. Attribution analyses further show that climate
33 conditions are the dominant driver of wood turnover, with a sharp increase when the temperature of the coldest month
34 exceeds -20°C. Our study demonstrates the importance of assimilating slow ecological trajectories to improve long-
35 term predictions of carbon storage and highlights the potential acceleration of the sensitivity of the boreal carbon sink
36 to warming under future climate change.



37 1. Introduction

38 Terrestrial ecosystems play a pivotal role in regulating the global carbon cycle by storing a substantial fraction of
39 anthropogenic CO₂ emissions (Schimel et al., 2015). As a core component of Earth System Models (ESMs), terrestrial
40 biosphere models (TBMs) provide a fundamental tool to simulate terrestrial carbon and water fluxes and assess how
41 climate change influences land ecosystem functioning and the terrestrial carbon sink (Zachle, 2013). However, despite
42 increasing complexity in TBMs, their projections of future changes in the terrestrial carbon sinks still exhibit substantial
43 uncertainties (Famiglietti et al., 2021). These uncertainties primarily stem from differences in how models represent
44 key biogeochemical and biophysical processes, such as photosynthesis and respiration, carbon allocation, vegetation
45 dynamics, disturbance, and soil carbon turnover, and in how these processes respond to climate and CO₂ (Canadell et
46 al., 2021). The magnitude of resulting uncertainty in carbon sink projections arising from these process-level
47 differences is comparable to the uncertainty in physical climate sensitivity (e.g., equilibrium climate sensitivity, ECS),
48 and together these uncertainties represent a major barrier to reliable future climate projections (Kaufhold et al., 2025).

49 A primary uncertainty in TBMs lies in the post-photosynthetic processes, especially the parameterization of carbon
50 allocation and biomass turnover (i.e., carbon residence time) (Bloom et al., 2016). Uncertainty in carbon residence time
51 is a major contributor to the spread of TBM projections under changing climate and rising atmospheric CO₂, as reflected
52 by large inter-model differences in baseline forest biomass turnover times (Friend et al., 2014; Pugh et al., 2020). Such
53 divergence arises from contrasting assumptions about how photosynthates are allocated among tissues and how quickly
54 carbon is transferred and lost from biomass (Pugh et al., 2020). TBMs typically rely on Plant Functional Types (PFTs)
55 to prescribe these parameters, but this PFT-based classification exhibits a limited ability to account for the observed
56 spatial heterogeneity in carbon allocation and residence times (Bloom et al., 2016).

57 Using model-data fusion methods to constrain parameters faces the challenge of equifinality, where different parameter
58 combinations can fit the observational data equally well (Famiglietti et al., 2021). Most model-data fusion studies rely
59 on high-frequency observational data at hourly, daily, or monthly scales from satellites or eddy-covariance flux towers,
60 such as gross primary productivity (GPP), ecosystem respiration (RECO), and leaf area index (LAI) to constrain critical
61 parameters of the carbon cycle (Bloom et al., 2016; Kuppel et al., 2014; Meroni et al., 2019; Wang et al., 2019; Zheng
62 et al., 2020). These high-frequency observations capture diurnal and seasonal dynamics of canopy photosynthesis and
63 topsoil respiration, but they are insufficient for constraining model predictions at annual or longer timescales (Keenan
64 et al., 2012; Peylin et al., 2016). Indeed, signals originating from large, slow-turnover carbon pools (such as woody
65 biomass) are difficult to detect in high-frequency flux data (Braswell et al., 2005). Consequently, even if a model near
66 perfectly fits short-term GPP and RECO, its internal allocation and long-term carbon turnover parameters may still be
67 largely unconstrained, undermining the reliability of predictions at decadal to centennial time scales, such as for forest
68 biomass growth and long-term carbon storage (Bloom et al., 2016; Keenan et al., 2012; Pugh et al., 2020).

69 Forest biomass growth curves, which describe the trajectory of biomass accumulation with stand age following major
70 disturbances such as fires, represent a potentially powerful constraint for constraining carbon allocation and turnover
71 parameters in TBMs. While a single-year biomass stock data from lidar or vegetation optical depth can be assimilated
72 to characterize current carbon stocks (Dubayah et al., 2020; Fan et al., 2019), such single-point constraints fail to
73 effectively resolve the rate of biomass change or the underlying turnover times (Shchepashchenko & Ciais, 2026; Zhou
74 et al., 2015), both of which are critical for making reliable long-term projections. Unlike high-frequency flux data and
75 biomass snapshots, biomass growth curves capture the integrated outcomes of long-term carbon allocation and turnover



76 processes over decades (Xu et al., 2026; Zhou et al., 2015). A key challenge in TBM parameter calibration is equifinality,
77 where divergent parameter sets, such as high carbon allocation paired with rapid turnover versus low allocation with
78 slow turnover, produce identical biomass stocks. Single-year biomass stock constraints cannot distinguish between
79 these pathways because they provide no information on the actual rate of carbon accumulation. By assimilating age-
80 related biomass stock data derived from sources such as forest inventories (Thum et al., 2017), chronosequences (Zhou
81 et al., 2015), or remote sensing observations (Neigh et al., 2025), model parameters governing the accumulation of
82 slow-turnover carbon pools can be more directly constrained, thereby reducing equifinality in post-photosynthetic
83 parameter estimates. Such slow-process constraint is particularly valuable in boreal forests, where carbon storage is
84 dominated by long-lived pools and long-term biomass trajectories can be strongly influenced by both disturbance (e.g.,
85 fire) regimes and climate warming.

86 Boreal forests store approximately 30% of global terrestrial carbon (Hansen et al., 2023), most of which resides in
87 slow-turnover carbon pools such as woody biomass and carbon-rich permafrost soils (Schuur, 2019). Carbon dynamics
88 in this region are largely driven by disturbance-induced, non-equilibrium biomass accumulation following stand-
89 replacing events (Jonsson & Wardle, 2009), making it an ideal testbed for evaluating the effectiveness of using forest
90 biomass growth curves to constrain model parameters. Accordingly, this study focuses on boreal forests.

91 To achieve this, we employed DifferLand, an advanced, JAX-based differentiable TBM (Fang & Gentine, 2024). Using
92 the differentiable DALEC implementation provided within DifferLand, automatic differentiation computes gradients
93 of the multi-source mismatch with respect to model parameters, enabling efficient gradient-based calibration against
94 heterogeneous constraints that combine fast fluxes with slow carbon stocks/trajectories (Fang & Gentine, 2024; Lin,
95 2024; Shi et al., 2024).

96 The objectives of this study are threefold: (1) evaluate the efficacy of assimilating forest biomass growth curves to
97 constrain critical carbon allocation and turnover parameters; (2) characterize the spatial patterns of these parameters
98 across the entire boreal biome (Thurner et al., 2016); and (3) identify the key environmental drivers (e.g., climate, soil
99 properties, and vegetation types) governing these spatial patterns.

100 **2. Materials and Methods**

101 **2.1 Calibration Framework**

102 We used DifferLand, a JAX-based differentiable data assimilation framework (Bradbury et al., 2018; Fang & Gentine,
103 2024), to calibrate ecosystem carbon-cycle parameters against multi-source observational data streams. JAX is a high-
104 performance numerical computing library that combines NumPy-like syntax with automatic differentiation and XLA-
105 accelerated execution (Bradbury et al., 2018). By leveraging automatic differentiation to compute exact gradients with
106 respect to model parameters and initial conditions through the entire simulation period, DifferLand enables the accurate
107 joint calibration of the parameters, along with data assimilation (using observation operators between model variables
108 and observations). Calibrating ecological processes across long timescales is particularly challenging, as error signals
109 from slow-turnover pools are often difficult to attribute to the fast physiological drivers that govern them. DifferLand
110 addresses this difficulty by allowing error signals from slow processes (e.g., biomass turnover) to be backpropagated
111 to constrain fast processes (e.g., photosynthesis), effectively bridging the timescale gap. This integrated approach
112 resolves the trade-offs between diverse datasets within a physics-constrained framework, thereby more effectively
113 constraining carbon allocation and potentially carbon residence times.



114 The physical core of the framework is the Data Assimilation Linked Ecosystem Carbon (DALEC) model (Bloom &
115 Williams, 2015). We used the " β -JS" (Jarvis-Stewart) configuration, which couples the process-based carbon balance
116 model with a semi-empirical soil moisture stress function (Fang & Gentine, 2024). The model represents the ecosystem
117 state using eight pools: six carbon pools (labile, foliar, wood, fine roots, litter, and soil organic matter) and two water
118 pools (plant available and unavailable water).

119 Carbon enters the ecosystem through gross primary productivity (GPP), which is simulated by the Aggregate Canopy
120 Model (ACM) based on leaf area index (LAI), meteorological drivers, and atmospheric CO₂ concentration (Williams
121 et al., 1997). In this configuration (" β -JS"), GPP is further regulated by a water stress factor (β) derived from the plant
122 available water using a piecewise-linear Jarvis-Stewart function. Coupled with the carbon cycle, the model
123 simultaneously simulates evapotranspiration (ET) as a function of meteorological drivers and canopy state, which
124 determines the depletion of the plant available water pool. Net primary productivity (NPP) is calculated as the fraction
125 of GPP remaining after accounting for autotrophic respiration. The resulting NPP is partitioned into the four well-
126 mixed live biomass pools based on time-invariant allocation fractions, assuming no vegetation demography (i.e., tree-
127 size or age classes are not explicitly resolved). Leaf phenology follows the Combined Deciduous-Evergreen Analytical
128 (CEDA) scheme, while the turnover of live biomass feeds into litter and soil organic matter pools (Bloom et al., 2016).
129 In total, the model comprises 26 tunable physical parameters (Table S1) governing these ecohydrological processes.

130 For the subsequent analysis, the "allocation to leaf" is calculated as the sum of allocation fractions to the labile and
131 foliar pools since they both support canopy development. Additionally, the "leaf lifespan" reported in this study
132 represents the effective mean residence time derived from the integrated annual turnover rate in the CEDA phenology
133 scheme, rather than the raw shape parameter used in the model equations (see Text S1, Fig. S1 for details).

134 **2.2 Data Source and Preprocessing**

135 **2.2.1 Meteorological Forcing Data**

136 We used the CRU-JRA v2.2 reanalysis dataset (Harris, 2021) as the climate forcing to drive the DALEC model. This
137 dataset provides meteorological fields at a spatial resolution of $0.5^\circ \times 0.5^\circ$ and a 6-hourly temporal resolution. We
138 aggregated the 6-hourly 2 m air temperature, precipitation, and downward shortwave solar radiation to a daily time
139 step to match the model simulation step. Daily maximum and minimum temperatures (T_{\max} , T_{\min}) were extracted from
140 the 6-hourly records. Atmospheric CO₂ concentrations were prescribed using annual global mean values from the
141 TRENDY protocol (Friedlingstein et al., 2025), assuming a uniform distribution across the globe within each year.
142 Since the vapor pressure deficit (VPD) was not directly provided in the CRU-JRA dataset, we derived it from air
143 temperature and specific humidity (Table S2).

144 **2.2.2 Observational Constraints for Fast Processes**

145 We assimilated multiple earth observation data to constrain the fast-turnover carbon and water fluxes. For GPP, we
146 used the FluxSat v2.0 daily GPP product (Joiner et al., 2018; Joiner & Yoshida, 2020, 2021). This dataset upscales eddy
147 covariance measurements from FLUXNET 2015 using Moderate-Resolution Imaging Spectroradiometer (MODIS)
148 nadir bidirectional reflectance distribution function adjusted reflectance (NBAR) data and a neural network approach,
149 explicitly accounting for light use efficiency and meteorological drivers. Daily evapotranspiration (ET) estimates were
150 obtained from the Global Land Evaporation Amsterdam Model (GLEAM) dataset (Miralles et al., 2025). GLEAM4



151 employs a hybrid modelling framework that combines physical principles (Penman-Monteith equation) with deep
152 learning-based evaporative stress derived from eddy-covariance and sap flow data, while assimilating satellite soil
153 moisture and vegetation optical depth (VOD) (Miralles et al., 2025), as a proxy for vegetation water stress. For
154 ecosystem respiration (RECO), we used a data-driven upscaled product provided as 10-day means (Zeng et al., 2020).
155 This product was generated using a random forest algorithm trained on FLUXNET 2015 observations. Vegetation state
156 was constrained using the Global Land Surface Satellite (GLASS) 8-day leaf area index (LAI) v6 product (Ma & Liang,
157 2022). This version uses a bidirectional long short-term memory (Bi-LSTM) deep learning model to generate
158 spatiotemporally continuous LAI profiles from MODIS surface reflectance.

159 All observational datasets were aggregated to $0.5^\circ \times 0.5^\circ$ resolution. To derive constraints representative of mature
160 forests, we utilized the ESA CCI dataset of forest aboveground biomass (AGB) for the year 2020 (Santoro & Cartus,
161 2023), which was first resampled to a spatial resolution of 0.05° . Using this map, we identified mature forest pixels
162 within each 0.5° grid cell as those with a AGB density exceeding both the 90th percentile of the grid cell and a minimum
163 threshold of 50 Mg ha^{-1} . The observational data for GPP, ET, RECO, and LAI were then aggregated to the 0.5°
164 resolution by averaging only the values corresponding to these identified mature forest pixels.

165 **2.2.3 Forest Biomass Constraints for Slow Processes**

166 To constrain the long-term carbon turnover and allocation, we used a recently developed dataset of post-fire forest
167 biomass growth curves specifically for boreal forests (Xu et al., 2026). These growth curves were derived using a
168 space-for-time substitution approach, combining 36-year Landsat-based fire history records (Long et al., 2019) with a
169 high-resolution (100 m) woody biomass map (Santoro & Cartus, 2023). For each 0.5° grid cell, the relationship between
170 AGB and stand age (i.e., years after the previous fire) was fitted using the Richard-Chapman growth function. These
171 spatially explicit growth trajectories provide a critical constraint on the biomass accumulation potential (AGB_{max}) and
172 growth rate across the boreal biome (Fig. S2). As a reference for comparison, we also used the ESA CCI dataset of
173 forest AGB for the year 2020 (Santoro & Cartus, 2023) to provide a static constraint for mature forest states.

174 While the satellite-derived curves and the single-year AGB represent only the aboveground component, the DALEC
175 model simulates carbon dynamics through distinct pools where the wood pool aggregates both stems and belowground
176 coarse roots. To ensure structural consistency between the observational constraint and the model's state variables, we
177 converted both the satellite AGB trajectories and the 2020 AGB into total biomass curves (sum of above- and
178 belowground biomass). This conversion was performed using spatially explicit Root-to-Shoot (R:S) ratios derived from
179 a global map of root biomass, which was generated by a random forest model trained on extensive field measurements
180 (Huang et al., 2021). The resulting total biomass curves provide a comprehensive constraint for the sum of all living
181 carbon pools in the model.

182 **2.2.4 Data for Parameter Priors**

183 We incorporated trait-based datasets to define priors and ecological constraints for specific model parameters. Spatial
184 reference maps for leaf carbon mass per area (LCMA) were derived from a global specific leaf area (SLA) product
185 (Butler et al., 2017), converted to carbon mass units assuming a carbon fraction of 0.5. Additionally, to represent the
186 known physiological trade-off between leaf lifespan and construction cost (Wright et al., 2004), we fitted a functional
187 relationship between leaf lifespan and LCMA ($R^2 = 0.42$, $p < 0.001$, Fig. S3). This relationship was derived using a
188 compiled global leaf trait dataset (Wang et al., 2023), which integrates the Glopnet database with the China Plant Trait



189 Database, providing a robust global representation of trait coordination. Furthermore, we utilized a global reference
190 map of ecosystem carbon use efficiency (CUE), which was generated using a random forest model trained on CUE
191 derived from eddy covariance observations (Luo et al., 2025).

192 **2.2.5 Ancillary Data for Spatial Analysis**

193 We incorporated additional time-invariant datasets to interpret the spatial heterogeneity of the calibrated parameters.
194 Data on soil physical and chemical properties, specifically bulk density, clay content, and soil H₂O pH, were obtained
195 from the Global Soil Dataset for Earth System Modeling (GSDE) at an original spatial resolution of 30 arc-seconds
196 (Shangguan et al., 2014). Additionally, the soil carbon-to-nitrogen ratio was derived from the organic carbon and total
197 nitrogen layers within the same dataset. The spatial distribution of vegetation composition was characterized using the
198 plant functional type (PFT) maps based on the CCI land cover products, provided at 0.25° resolution (Li et al., 2018).
199 We specifically extracted the 1985–2020 mean areal fractions of boreal needleleaf evergreen (ENF), boreal broadleaf
200 summergreen (DBF), and boreal needleleaf deciduous (DNF). All ancillary datasets were subsequently aggregated to
201 the 0.5° model grid.

202 **2.2.6 Independent Validation Data**

203 To independently validate the calibrated carbon allocation parameters, we utilized a dataset of forest biomass allocation
204 derived from field measurements (Xia et al., 2019). We aggregated the plot-level measurements to the 0.5° grid
205 resolution to match our model simulations. Further details on the processing of this dataset are provided in
206 Supplementary Text S3.

207 **2.3 Experimental Design and Model Calibration**

208 **2.3.1 Simulation Protocol**

209 In the framework of DifferLand, the initial sizes of the eight ecological pools (labile, foliar, wood, root, litter, soil
210 organic matter, plant available water, plant unavailable water) are treated as trainable parameters, retrieved
211 simultaneously with the 26 physical parameters from the observational data. In other words, within the differentiable
212 framework, initial conditions can be treated as parameters and we can compute the gradients/sensitivity of the model
213 to those initial conditions.

214 To capture the long-term legacy effects of carbon turnover and accumulation, we designed a 107-year continuous
215 simulation experiment (growth-curve-constrained experiment) for each grid cell. The meteorological drivers from 1985
216 to 2020 (36 years) were cycled three times, with the final year (2020) excluded from the last cycle to align the
217 simulation end date (2019) with the extent of the observational record. This protocol simulates the forest growth
218 trajectory from a stand-replacing disturbance event with a biomass of nearly zero to a mature state.

219 To quantify the specific impact of assimilating forest biomass growth dynamics, we conducted two additional
220 experiments, a baseline experiment and a single-year biomass-constrained experiment. Both simulations spanned only
221 19 years (2001–2019), driven directly by the corresponding meteorological forcing without cycling. The baseline setup
222 closely follows the original DifferLand configuration (Fang & Gentine, 2024), where the model is constrained only by
223 high-frequency data (GPP, ET, RECO, LAI) without the information provided by the long-term forest biomass growth
224 curves. While the baseline experiment excludes long-term biomass constraints entirely, the single-year biomass-



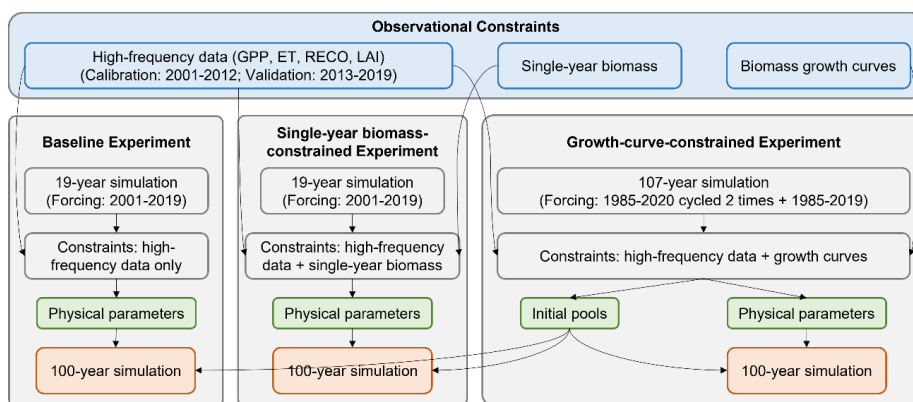
225 constrained experiment additionally assimilates the 2020 biomass to constrain the biomass at the end of the simulation
226 and evaluate the efficacy of this constraint in reducing parameter equifinality.

227 For all experimental setups, the calibration process utilized observational data from the period 2001–2012
228 (corresponding to the simulation years 89–100 for the 107-year run, and years 1–12 for the 19-year run), while data
229 from 2013–2019 were reserved for independent validation. By aligning the calibration period with simulation years
230 89–100, we assume that after approximately 90 years of recovery following the stand-replacing disturbance, the
231 simulated ecosystem has reached a mature state consistent with the present-day observational constraints (GPP, RECO,
232 ET and LAI), which were specifically filtered to represent mature forest pixels. To ensure consistency between model
233 outputs and observations, daily model fluxes were temporally processed to match the assimilation datasets: daily
234 outputs were averaged to match the 10-day means of RECO, while state variables were sampled to match the 8-day
235 interval of LAI.

236 During the calibration of all three experimental setups, the initial values of the physical parameters and ecological
237 pools for each grid cell were subject to five independent random initializations. Initiating multiple independent
238 optimization sequences from overdispersed starting values is a widely recognized standard in iterative simulation and
239 parameter estimation to ensure robust posterior inference (Gelman & Rubin, 1992). For the subsequent analysis of
240 parameter spatial patterns, the representative estimate for each parameter was determined by taking the median across
241 this 5-member ensemble. To ensure the reliability of the retrieved parameters, we assessed the robustness of the 5-
242 member ensemble for each grid cell using a robust coefficient of variation (Fig. S4), and excluded parameters within a
243 grid cell from subsequent spatial and attribution analyses if their robust coefficient of variation exceeded a threshold
244 of 0.3 to minimize the influence of non-convergent or unstable optimizations (Leys et al., 2013; Rousseeuw & Croux,
245 1993).

246 Finally, to evaluate the impact of the retrieved optimized parameters on long-term carbon accumulation, we performed
247 a 100-year forward simulation using the initial state derived from the growth-curve-constrained experiment. By
248 applying the physical parameters retrieved from the baseline and the single-year biomass-constrained experiments to
249 this same initial state from the growth-curve-constrained experiment and comparing the result against the growth-
250 curve-constrained experiment, we isolated the divergence in 100-year biomass accumulations caused solely by the
251 differences in physical parameters. The detailed workflow of these experimental designs is illustrated in Fig. 1.

252 To rigorously propagate parameter uncertainty into the prognostic outputs, these forward simulations under all three
253 experimental setups were executed individually for each of the five ensemble members, rather than running a single
254 simulation forced by the aggregated median parameters. Specifically, by applying the physical parameters retrieved
255 from the baseline and single-year biomass-constrained experiments to the corresponding initial states derived from the
256 growth-curve-constrained experiment, we simulated a 5-member ensemble of 100-year biomass accumulation
257 trajectories. We then calculated the median of these prognostic trajectories and compared it against the ensemble
258 median of the growth-curve-constrained experiment. This approach allowed us to robustly isolate the divergence in
259 long-term biomass accumulations caused solely by the differences in physical parameters.



260

261 **Fig. 1 Schematic illustration of the experimental design.** The framework consists of two calibration experiments: a
 262 baseline experiment constrained solely by high-frequency observations (GPP, ET, RECO, and LAI) over a 19-year
 263 period; a single-year biomass-constrained experiment that additionally incorporates a single-year mature forest biomass
 264 over the same 19-year period; and a growth-curve-constrained experiment that additionally assimilates long-term forest
 265 biomass growth curves over a 107-year simulation. To isolate the impact of physical parameters on biomass storage,
 266 the 100-year forward simulations for all three experiments use the initial state retrieved from the growth-curve-
 267 constrained experiment.

268 **2.3.2 Pre-calibration of Canopy Efficiency**

269 Prior to the model calibration, we performed an independent pre-calibration step to determine the prior value for the
 270 canopy efficiency (ce) parameter. This parameter governs the maximum light use efficiency in the ACM module and
 271 represents the intrinsic photosynthetic capacity of the vegetation under non-water-stressed conditions. Simultaneous
 272 optimization of ce and water stress parameters often leads to an equifinality problem, where an underestimated capacity
 273 is compensated by an underestimated stress. To isolate the optimized physiological capacity from soil moisture
 274 limitations, we identified "well-watered" periods using the evaporative fraction (EF). Theoretically, EF is defined as
 275 the ratio of latent heat flux (LE) to the available energy (net radiation R_n minus ground heat flux G). In our study, this
 276 ratio is approximated by the ratio of actual evaporation to potential evaporation (E / E_p) (Miralles et al., 2011;
 277 Seneviratne et al., 2010). We calculated this proxy using daily E and E_p estimates from the GLEAM v4.2a dataset. For
 278 each grid cell, we selected days where the EF exceeded the 80th percentile of its temporal distribution. During these
 279 periods, we assumed that soil water stress was negligible (i.e., the water stress factor $\beta \approx 1$). We then optimized the ce
 280 parameter by minimizing the mean squared error (MSE) between the ACM-simulated GPP (driven by observed LAI
 281 and meteorology) and the observed FluxSat GPP. The retrieved optimal value (ce_{opt}) was subsequently incorporated
 282 into the full model calibration as a prior constraint.

283 **2.3.3 Optimization and Cost Function**

284 To retrieve the 26 physical parameters and initial states of the model, simultaneously, we employed the Adam optimizer
 285 within the JAX framework to minimize a composite cost function (J). This function integrates observational errors
 286 (J_{obs}), parameter priors (J_{prior}), and ecological constraints (J_{EDC}) into a single scalar metric. The observational loss (J_{obs})
 287 quantifies the mismatch between model simulations and data using the negative Normalized Nash-Sutcliffe Efficiency
 288 (NNSE). The NNSE transforms the standard error metric onto a bounded range of (0, 1], preventing variables with



289 large absolute magnitudes from dominating the optimization gradient (Text S2). Accordingly, J_{obs} is calculated as the
290 negative sum of NNSE values for the long-term total biomass trajectory and the high-frequency variables (GPP, ET,
291 RECO, and LAI) during the calibration period.

292 To constrain the parameter search space and prevent ecologically unrealistic solutions, we included prior cost terms
293 that penalize deviations of the calibrated parameters from independent estimates. These priors specifically target the
294 canopy efficiency (ce), LCMA and the fraction of GPP allocated to autotrophic respiration (f_{auto}). The model is
295 penalized for deviations from the pre-calibrated canopy efficiency (ce_{opt}), the satellite-derived LCMA reference maps,
296 and the CUE-inferred f_{auto} ($1 - \text{CUE}$) maps, ensuring that the retrieved parameters remain anchored to observational
297 evidence.

298 Finally, we imposed ecological and dynamical constraints (EDCs) to ensure biologically realistic process rates. We
299 adopted five standard EDCs (EDC1–5) from the CARDAMOM framework (Bloom et al., 2016; Fang & Gentine, 2024)
300 to enforce hierarchical relationships in carbon pool turnover and allocation. EDC1: Turnover rate of the litter pool
301 should be greater than the turnover rate of soil organic matter; EDC2: The rate of decomposition of litter into soil
302 organic matter should be faster than the release of carbon from the soil organic matter pool into the atmosphere; EDC3:
303 Foliar carbon turnover should be faster than the wood carbon turnover; EDC4: Root turnover faster than soil organic
304 matter turnover at mean temperature; and EDC5: Allocation to canopy is comparable to allocation to roots. Beyond the
305 five standard constraints, we introduced specific modifications for this study. A modified constraint (EDC6) was
306 applied to enforce a quasi-steady-state assumption for solely the soil organic matter pool, allowing vegetation carbon
307 pools to accumulate mass over time to simulate forest biomass growth. Furthermore, we incorporated the leaf
308 economics spectrum as a distinct ecological constraint (EDC7). This constraint enforces the fundamental physiological
309 trade-off between leaf longevity and construction cost by penalizing parameter combinations that deviate from the
310 global log-linear relationship between leaf lifespan and LCMA fitted in Section 2.2.4 (Fig. S3) (Wang et al., 2023;
311 Wright et al., 2004).

312 **2.4 Attribution Analysis using Machine Learning**

313 To attribute the spatial heterogeneity of the calibrated parameters to environmental drivers, we employed an attribution
314 analysis framework based on a gradient boosting regression algorithm (Friedman, 2001; Pedregosa et al., 2011). We
315 developed separate machine learning models to predict the calibrated values of key process parameters, specifically
316 leaf lifespan, carbon turnover rates for wood and fine roots, and NPP allocation fractions to leaf, wood, and fine roots.
317 The predictor variables for these gradient boosting models include climatic variables derived from the CRU-JRA
318 dataset (e.g., mean annual temperature, precipitation seasonality, and solar radiation), alongside with the soil properties
319 and vegetation fractions (Table S4).

320 To evaluate the model performance in predicting the spatial variation of the parameters, we conducted a five-fold cross
321 validation, where the results dataset of each parameter was randomly divided into five subsets, and the gradient
322 boosting models were trained and tested iteratively. The goodness of fit was quantified using the coefficient of
323 determination (R^2) and root mean square error (RMSE) obtained from these validation sets. We then interpreted their
324 internal decision-making using the SHAP (SHapley Additive exPlanations) tool. SHAP quantifies the relative
325 importance of each predictor and estimates its marginal effect on the output parameter, allowing us to disentangle the
326 contributions of climate, soil, and vegetation structure in driving the observed spatial distribution of carbon allocation
327 and turnover strategies across the boreal forest biome.



328 3. Results

329 3.1 Reducing Long-term Biomass Bias with Slow-process Constraints

330 The baseline experiment, which constrained the DALEC model solely using high-frequency flux and LAI observations,
331 failed to reproduce the magnitude and spatial variability of carbon accumulation in boreal forests (Fig. 2). Despite
332 successfully fitting the daily dynamics of GPP and RECO (Fig. S5), the baseline experiment exhibited a general
333 underestimation simulating mature forest biomass stocks throughout the boreal biome (Fig. 2a). The model tended to
334 underestimate biomass in high-latitude regions and overestimate biomass in the southern ecotones of the boreal biome.
335 Consequently, the baseline simulation yielded a poor fit with the 100-year biomass derived from the forest growth
336 curves ($R^2 = 0.02$), with a large RMSE of 138.7 Mg ha^{-1} . These discrepancies highlight that accurate short-term flux
337 simulations do not guarantee reliable long-term carbon storage estimates, because turnover processes remain weakly
338 constrained and allow compensating parameter combinations.

339 The single-year biomass-constrained experiment partially mitigated these biases but exhibited a distinct spatial pattern
340 of underestimating biomass in high-latitude regions and overestimating it in the southern ecotones (Fig. 2b). While it
341 improved the overall fit ($R^2 = 0.35$ and $\text{RMSE} = 87.5 \text{ Mg ha}^{-1}$), significant systematic errors remained. Notably, this
342 improvement was achieved primarily through a 64.8% average reduction in wood turnover rates compared to the
343 baseline (Fig. 2g), while the fraction of NPP allocated to wood remained nearly unchanged (Fig. 2d). This suggests
344 that a static biomass snapshot can constrain the residence time to some extent but lacks the information necessary to
345 adjust internal carbon allocation patterns.

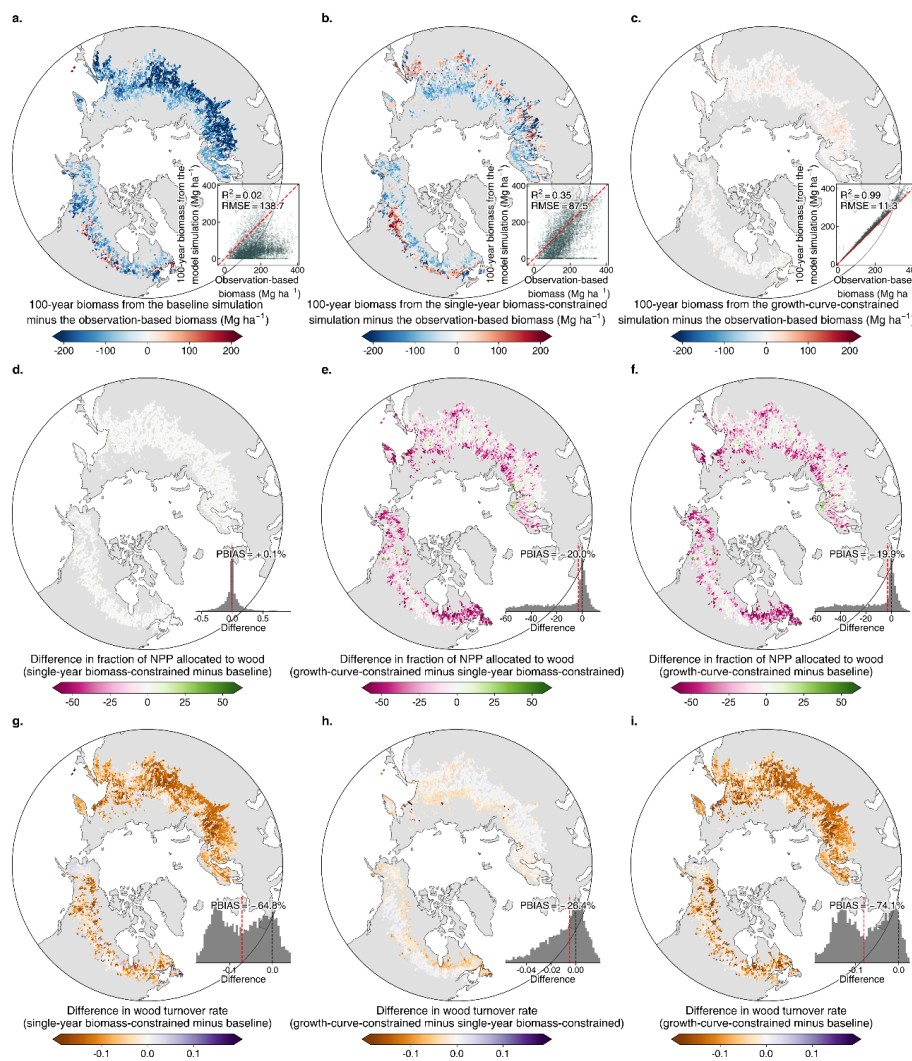
346 Assimilating the forest biomass growth curves effectively resolved these spatial biases. In the experiment with growth
347 curve constraints, the simulated 100-year biomass aligned closely with the observational constraints along the 1:1 line
348 (Fig. 2c). This constraint largely improved the spatial performance of the model simulation, increasing the R^2 to 0.99
349 and reducing the RMSE to 11.3 Mg ha^{-1} . Notably, this improvement in simulating slow carbon pools was achieved
350 without compromising the accuracy for fast processes (Fig. S5). The spatial patterns and magnitudes of the NNSE for
351 GPP, RECO, and ET remained similar across all experimental setups (Fig. S5), demonstrating the capacity of the
352 framework to reconcile multi-timescale observational constraints.

353 The elimination of the biomass bias was primarily driven by substantial changes in carbon allocation and turnover
354 parameters. To support higher biomass accumulation in mature stands, the optimization algorithm adjusted the
355 parameter towards longer residence times. Specifically, the wood carbon turnover rate was reduced by an average of
356 74.1% compared to the baseline (Fig. 2i), implying more than a doubling of the woody biomass residence time. This
357 reduction in turnover allowed for biomass accumulation despite a 19.9% decrease in the fraction of NPP allocated to
358 wood (Fig. 2f). The model allocated more carbon to fine roots (Fig. S6i), potentially reflecting the high nutrient
359 acquisition demands in boreal soils. Meanwhile, allocation to leaves decreased moderately (Fig. S6c), compensated by
360 a 114.6% extension in leaf lifespan (Fig. S6f) to maintain the LAI required to match satellite observations, a shift
361 consistent with the dominance of evergreen conifers in these high-latitude ecosystems.

362 Comparisons with field measurements of NPP allocation (Xia et al., 2019) provided independent evidence that these
363 parameter changes moved the model towards a more ecologically realistic state (Text S3, Fig. S7). While the baseline
364 and single-year biomass-constrained estimates for wood allocation showed poor agreement with field data ($r < 0.01$
365 and bias $> 25\%$; Fig. S7b, e), the growth-curve-constrained experiment showed a marked improvement in the



366 consistency between the retrieved optimized wood allocation and the field observations, increasing the correlation to r
 367 $= 0.57$ and a reducing the bias to 5.86% (Fig. S7h). This suggests that only by assimilating forest biomass growth
 368 trajectories can the model effectively constrain internal carbon flows to be more consistent with observed ecological
 369 allocation patterns.



370

371 **Fig. 2. Comparison of model simulated long-term biomass with that from the reference forest biomass growth**
 372 **curves (observation-based), and changes in wood carbon turnover and allocation. (a-c) Spatial distribution of the**
 373 **difference in the 100-year total biomass between model simulations and the observation-based estimates (a, baseline**
 374 **experiment; b, single-year biomass-constrained experiment; c, growth-curve-constrained experiment). The baseline**
 375 **experiment is constrained by GPP, RECO, LAI, and ET. The single-year biomass-constrained experiment additionally**
 376 **assimilates present-day biomass stock, whereas the growth-curve-constrained experiment assimilates forest biomass**
 377 **growth curves. (d-f) Spatial difference maps showing the changes in the fraction of NPP allocated to wood across the**
 378 **three experiments (d, single-year biomass-constrained minus baseline; e, growth-curve-constrained minus single-year**
 379 **biomass-constrained; f, growth-curve-constrained minus baseline). (g-i) Spatial difference maps showing the changes**



380 in wood turnover rate (**g**, single-year biomass-constrained minus baseline; **h**, growth-curve-constrained minus single-
381 year biomass-constrained; **i**, growth-curve-constrained minus baseline). PBIAS in the insets denotes the percent bias,
382 indicating the average relative difference between the corresponding experiments.

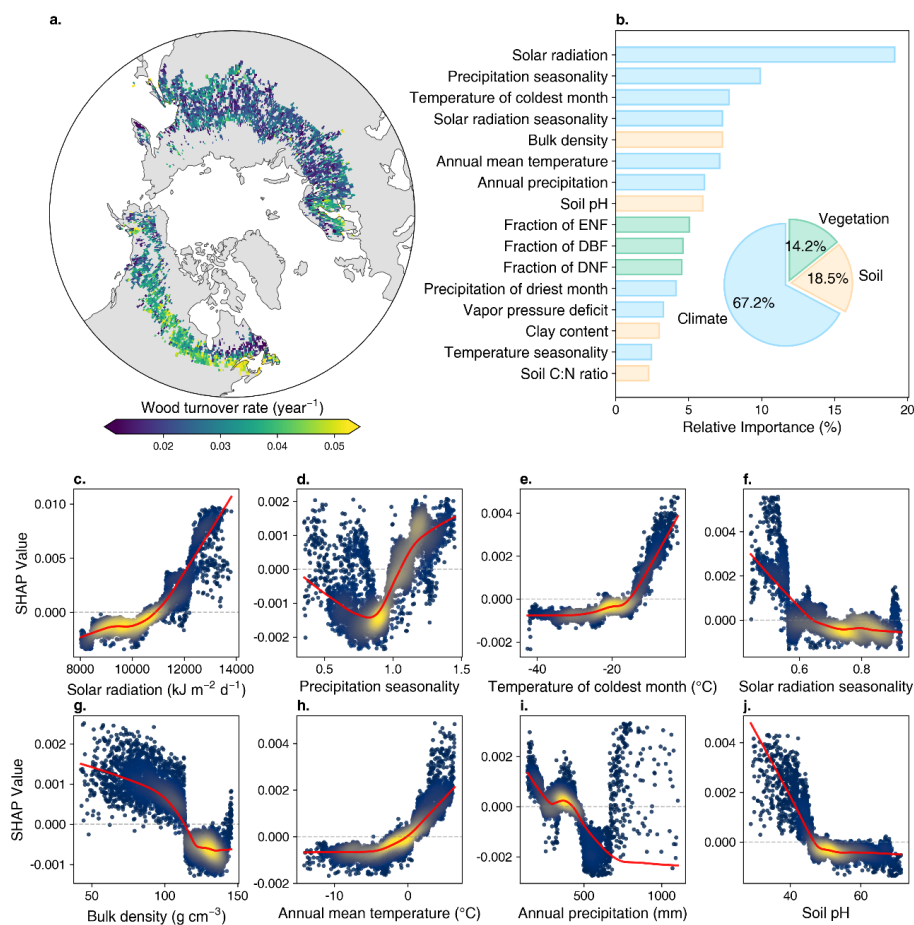
383 **3.2 Environmental Drivers Controlling Carbon Turnover and Allocation**

384 To disentangle the complex environmental controls on the calibrated carbon turnover and allocation parameters, we
385 employed a gradient boosting regression framework. The gradient boosting models captured ecologically consistent
386 signals from environmental drivers, explaining a moderate proportion of the spatial variance with R^2 values ranging
387 from 0.28 to 0.76 (Fig. S8). Predictive performance was highest for canopy-related traits, specifically the fraction of
388 NPP allocated to leaves ($R^2 = 0.60$) and leaf lifespan ($R^2 = 0.76$). In contrast, parameters associated with slower or
389 belowground processes, such as wood and fine root dynamics, showed lower explainability (R^2 between 0.26 and 0.34).
390 This remaining unexplained variance likely reflects unobserved local heterogeneity (e.g., micro-site soil conditions)
391 and stochastic non-climatic mortality agents (e.g., windthrow, insect outbreaks) that are not fully captured by the
392 coarse-resolution environmental predictors.

393 **3.2.1 Environmental Drivers Controlling Carbon Turnover**

394 For the spatial variability of our optimized wood turnover rate ($R^2 = 0.44$), climatic variables were the most important
395 predictors, accounting for 67.2% of the relative importance (Fig. 3b). Solar radiation was found to be the most important
396 predictor and exhibited a strong positive correlation with the wood turnover rate (Fig. 3c). This suggests that forests in
397 light-limited environments tend to adopt a strategy of extended biomass residence times to maintain standing stocks.
398 Additionally, the response to the temperature of the coldest month revealed a non-linear threshold (Fig. 3e). Wood
399 turnover rates remained stable in regions with extreme winter cold (i.e., the mean temperature of the coldest month
400 below -25°C) but increased sharply as the temperature of the coldest month rose above -20°C . This implies that extreme
401 cold may currently act as a stabilizing mechanism for woody biomass, probably by limiting biotic disturbance agents.
402 Annual mean temperature also exhibited a positive relationship with turnover rate (Fig. 3h). Furthermore, wood
403 turnover rates showed a sharp decline in more alkaline soils (Fig. 3j), likely reflecting the impacts of soil nutrient
404 availability and pH-mediated stress on tree longevity.

405 In terms of leaf lifespan, the fraction of ENF was found to be the most important predictor, positively correlated with
406 leaf lifespan (Fig. S9b, c). The fraction of DBF also emerged as an important driver, exhibiting a sharp negative
407 relationship with leaf lifespan (Fig. S9e). This relationship confirms that the data assimilation framework successfully
408 retrieved the distinct phenological traits of needleleaf and broadleaf species (reflecting multi-year leaf retention versus
409 annual turnover) directly from the constraints of satellite observations.



410

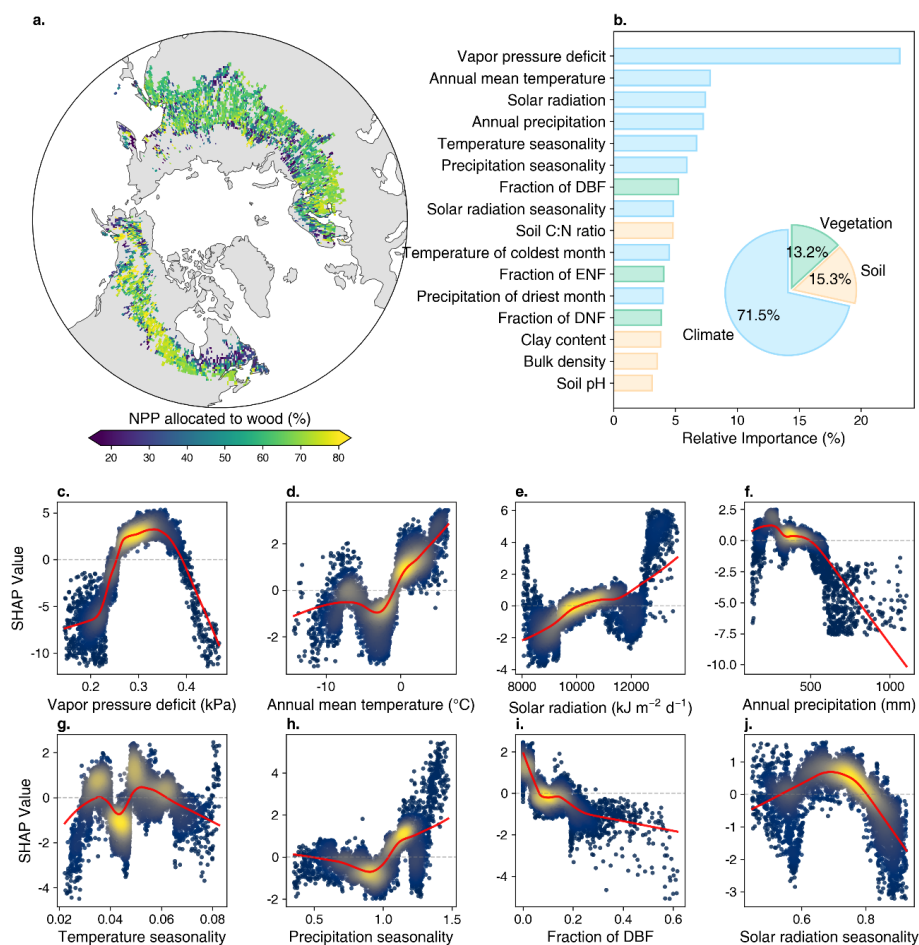
411 **Fig. 3 Spatial patterns and environmental drivers of wood turnover rates.** (a) Spatial distribution of the calibrated
 412 wood turnover rates across the boreal biome. (b) Relative importance of environmental predictors derived from the
 413 gradient boosting regression model. (c–j) SHAP dependence plots for the top eight most important predictors,
 414 illustrating the marginal effect of each variable on the wood turnover rate. Red solid lines indicate the locally weighted
 415 scatterplot smoothing (LOWESS) fit to highlight the trend. Note that ENF, DBF, DNF and C:N represent boreal
 416 needleleaf evergreen, boreal broadleaf summer green, boreal needleleaf deciduous, and carbon:nitrogen, respectively.

417 **3.2.2 Environmental Drivers Controlling Carbon Allocation**

418 For the fraction of NPP allocated to wood, VPD emerged as the most important predictor, followed by annual mean
 419 temperature and solar radiation (Fig. 4b). Climatic variables collectively accounted for 71.5% of the relative importance.
 420 Wood allocation exhibited a distinct non-linear threshold response to water stress, where investment initially increased
 421 with VPD but dropped abruptly when it exceeded about 0.35 kPa (Fig. 4c). This implies that vegetation in drier
 422 environments allocates a smaller fraction of NPP to wood. Furthermore, annual mean temperature acted as a
 423 fundamental constraint on structural investment, with wood allocation increasing as temperatures rose (Fig. 4d). In
 424 these cold environments, the reduction in wood investment was accompanied by a shift towards fine roots (Fig. S11e).
 425 This pattern suggests that vegetation in cold-limited biomes tends to prioritize belowground resource acquisition over



426 aboveground structural development (Reich et al., 2014). Environmental seasonality also remains a key driver, with
 427 wood investment generally declining under conditions of high temperature or solar radiation seasonality (Fig. 4g, j).



428

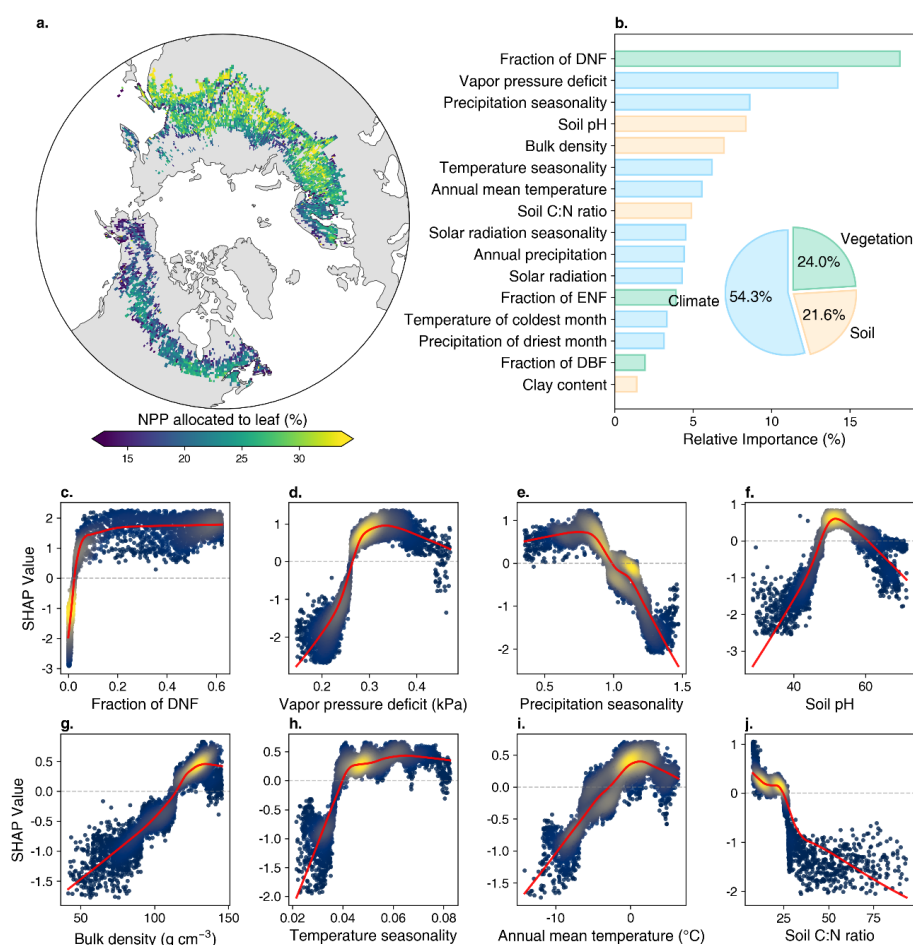
429 **Fig. 4 Spatial patterns and environmental drivers of the fraction of net primary productivity (NPP) allocated to**
 430 **wood.** (a) Spatial distribution of the fraction of NPP allocated to wood across the boreal biome. (b) Relative importance
 431 of environmental predictors derived from the gradient boosting regression model. (c–j) SHAP dependence plots for the
 432 top eight most important predictors, illustrating the marginal effect of each variable on the fraction of NPP allocated
 433 to wood. Red solid lines indicate the locally weighted scatterplot smoothing (LOWESS) fit to highlight the trend. Note
 434 that ENF, DBF, DNF and C:N represent boreal needleleaf evergreen, boreal broadleaf summer green, boreal needleleaf
 435 deciduous, and carbon:nitrogen, respectively.

436 For the fraction of NPP allocated to leaf, the fraction of deciduous needleleaf forests emerged as the most important
 437 predictor (Fig. 5b). Leaf investment showed a rapid increase as the fraction of deciduous needleleaf forests rose, before
 438 stabilizing at higher fraction (Fig. 5c). Atmospheric water demand also played an important role. It initially increased
 439 as VPD rose from 0.1 to 0.35 kPa (Fig. 5d), likely because a moderate increase in evaporative demand enhances the
 440 transpiration required for nutrient and water transport to the canopy, providing the necessary resources to support the
 441 construction and maintenance of a larger canopy (López et al., 2021; Massmann et al., 2019; Matimati et al., 2014).



442 However, beyond the threshold of about 0.35 kPa, NPP allocated to leaf declined (Fig. 5d), consistent with a shift
 443 toward stomatal closure to prevent hydraulic failure, which in turn limits carbon uptake and restricts further canopy
 444 expansion (Koehler et al., 2023). Moreover, leaf investment showed a strong positive correlation with bulk density
 445 (Fig. 5g). Leaf allocation also rose as the annual mean temperature increased from about -10°C to 0°C but plateaued
 446 thereafter (Fig. 5i), suggesting that extreme cold limits canopy expansion (Way & Oren, 2010).

447 For the fraction of NPP allocated to fine roots, it was upregulated under both high VPD conditions (indicative of
 448 hydraulic stress) and low VPD conditions typical of cold, humid areas (Fig. S11c). The NPP allocated to fine roots also
 449 increased when the soil C: N (carbon: nitrogen) ratio exceeded 25 (Fig. S11f). This to some extent confirms that the
 450 suppression of vegetation growth in high-latitude regions is driven by a functional shift towards nutrient acquisition in
 451 nitrogen-limited soils.



452

453 **Fig. 5 Spatial patterns and environmental drivers of the fraction of net primary productivity (NPP) allocated to**
 454 **leaf. (a) Spatial distribution of the fraction of NPP allocated to leaf across the boreal biome. (b) Relative importance**
 455 **of environmental predictors derived from the gradient boosting regression model. (c–j) SHAP dependence plots for the**
 456 **top eight most important predictors, illustrating the marginal effect of each variable on the fraction of NPP allocated**



457 to leaf. Red solid lines indicate the locally weighted scatterplot smoothing (LOWESS) fit to highlight the trend. Note
458 that ENF, DBF, DNF and C:N represent boreal needleleaf evergreen, boreal broadleaf summer green, boreal needleleaf
459 deciduous, and carbon:nitrogen, respectively.

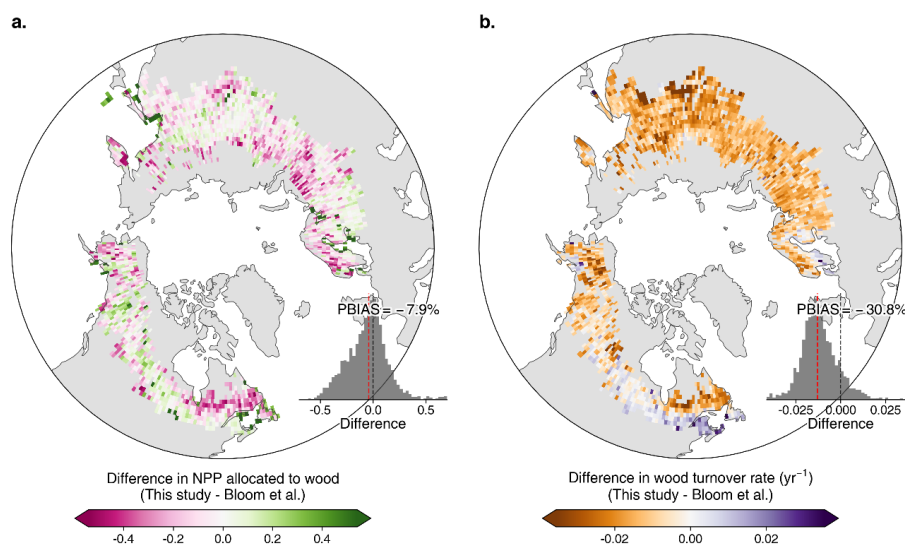
460 **3.3 Benchmarking against Global Carbon Cycle Retrievals**

461 We compared our calibrated parameters against the CARDAMOM product, a spatially explicit global carbon cycle
462 reanalysis derived by assimilating satellite observations into the DALEC model (Bloom et al., 2016). Although both
463 frameworks perform grid-level parameter retrieval, they differ critically in the observational constraints used for boreal
464 ecosystems. CARDAMOM relies primarily on LAI constraints in the boreal zone, whereas our framework assimilates
465 forest biomass growth curves. This methodological difference resulted in substantial variations in the estimated carbon
466 turnover and allocation parameters (Fig. 6 and Fig. S12).

467 For the wood carbon pool, our results indicate a more conservative turnover strategy compared to the CARDAMOM
468 benchmark. We found that the fraction of NPP allocated to wood was generally lower than CARDAMOM estimates
469 across the boreal biome, showing a mean percent bias (PBIAS) of -7.9% (Fig. 6a). The wood turnover rates were also
470 retrieved on average to be 30.8% lower than CARDAMOM estimates (Fig. 6b). These results suggest that when
471 constrained by long-term biomass growth trajectories rather than LAI alone, the model identifies extended biomass
472 residence times as the primary mechanism for maintaining boreal biomass carbon stocks, rather than high allocation
473 inputs.

474 Substantial differences also emerged in canopy parameters. Our framework retrieved longer leaf lifespan across the
475 boreal biome, higher than CARDAMOM estimates by an average of 123.1% (0.82 years vs 0.37 years; Fig. S12b).
476 This retrieval is consistent with the physiological traits of the evergreen conifers that dominate the region. In contrast
477 to previous estimates, our results show a slightly higher fraction of NPP allocated to the canopy, which was 3.8% higher
478 than the CARDAMOM baseline (Fig. S12a).

479 The reduction in allocation to wood and leaf in our retrievals was balanced by a higher allocation to fine roots compared
480 to the CARDAMOM baseline, pointing to a larger investment in nutrient acquisition relative to structural growth than
481 the benchmark (Fig. S12c). Our estimates of fine root turnover rates are also higher than those of CARDAMOM (Fig.
482 S12d), although this parameter remains the less constrained by the available observation dataset. These comparisons
483 indicate that assimilating forest biomass growth dynamics results in a functional shift in the model parameters towards
484 a longer-lived wood and leaves, supported by a higher proportional investment in fine root allocation.



485

486 **Fig. 6** Spatial differences in the calibrated (a) NPP allocated to wood and (b) the wood turnover rate between
 487 **this study and the estimates from the CARDAMOM product** (Bloom et al., 2016). PBIAS in the insets denotes the
 488 percent bias, indicating the average relative difference between the estimates in this study and the CARDAMOM
 489 product.

490 **4. Discussion**

491 A major uncertainty in terrestrial biosphere modeling arises from equifinality, where different parameter combinations
 492 yield similar fits to observational data (Bloom et al., 2016; Famiglietti et al., 2021). In our baseline experiment,
 493 constrained solely by high-frequency observations (GPP, RECO, ET, and LAI), the model reproduced seasonal flux
 494 dynamics well but showed large biases in long-term biomass stocks. It is mainly because high-frequency data primarily
 495 constrain fast physiological processes, such as photosynthesis, while leaving the processes and parameters governing
 496 long-term turnover and allocation less constrained (Friend et al., 2014; Pugh et al., 2020). Specifically, without explicit
 497 constraints of accumulated biomass, the optimization algorithm tends to exploit the equifinality between carbon
 498 allocation and turnover. In our baseline experiment, the model compensated for the lack of long-term constraints by
 499 converging toward a fast-turnover strategy. This allowed the model to maintain appropriate LAI and GPP levels to
 500 match satellite observations but prevented the accumulation of woody biomass, resulting in the systematic
 501 underestimation of carbon stocks in boreal regions. Our results further demonstrate that while incorporating a snapshot
 502 of mature biomass information (single-year biomass-constrained experiment) provides a valuable anchor for carbon
 503 residence times, it remains insufficient to resolve the equifinality of internal allocation patterns. Furthermore, potential
 504 inconsistencies between independent GPP and RECO products could introduce biases in the net carbon flux (NEE). In
 505 the absence of stock constraints, such possible flux biases might propagate into long-term biomass estimates. The
 506 assimilation of forest biomass growth dynamics can effectively mitigate this uncertainty by providing an independent
 507 constraint on the net cumulative carbon.

508 To constrain these slower processes, we assimilated forest biomass growth dynamics derived from high-resolution
 509 satellite observations (Xu et al., 2026). Unlike instantaneous flux measurements or static biomass snapshots, these
 510 biomass growth curves reflect the cumulative outcome of carbon allocation and turnover over decades (Zhou et al.,



511 2015). Incorporating this constraint significantly reduced the error in the simulated biomass stocks, with the RMSE
512 decreasing from 138.7 to 11.3 Mg ha⁻¹. Notably, this improvement in carbon stock estimation was achieved while
513 maintaining similar performance for high-frequency fluxes. These findings indicate that robust model predictions
514 require integrating data across multiple timescales. Jointly constraining fast fluxes and slow state variables helps reduce
515 equifinality by ruling out compensating combinations of carbon input and turnover that can reproduce short-term
516 dynamics while producing biased long-term carbon storage.

517 Our assimilation of forest biomass growth curves relies on the 0.5° grid-level space-for-time substitution approach,
518 which assumes that spatial biomass gradients across different stand ages accurately represent the temporal growth
519 trajectory of a single stand. However, in the context of rapid climate change, the conditions under which mature stands
520 established decades ago may differ from those influencing regeneration today (Likens, 1989; Walker et al., 2010).
521 Historical biomass growth curves might not fully capture future growth trajectories under climate change and
522 increasing atmospheric CO₂. Furthermore, converting satellite-derived aboveground biomass to total biomass may
523 introduce additional uncertainty. Although we employed spatially explicit root-to-shoot ratios (Huang et al., 2021),
524 these estimates may not fully capture stand-age-dependent changes in belowground allocation. Accordingly, the
525 belowground processes, especially the fine root turnover rates, still have great uncertainties due to the lack of effective
526 constraints from direct observations. Moreover, the accuracy of our parameter retrieval depends on the quality of the
527 assimilated data products. Systematic biases in satellite-derived GPP or biomass estimates would inevitably propagate
528 into the calibrated parameters (Keenan et al., 2011). Continuous validation against independent field measurements
529 and the integration of next-generation biomass products are therefore essential to further refine these constraints.

530 While the DALEC configuration assumes time-invariant allocation fractions, our assimilation of 107-year biomass
531 growth curves effectively captures the integrated outcome of these potentially dynamic processes. In real-world boreal
532 successions, allocation patterns may shift as stands mature (Gundersen et al., 2021; Xia et al., 2019). However, by
533 optimizing parameters against the full accumulation trajectory, DifferLand retrieves effective mean parameters that
534 robustly represent the long-term carbon balance of the ecosystem. This approach offers a significant advancement over
535 baseline calibrations constrained only by high-frequency fluxes, as it ensures that the retrieved traits are consistent with
536 observed decadal biomass stocks. Future iterations could incorporate age-dependent functions to further refine these
537 internal dynamics without compromising the parsimony of the current differentiable framework.

538 Additionally, the representation of non-structural carbohydrates (NSC) in DALEC is simplified into a single labile pool
539 that exclusively supplies foliar development. In contrast, real-world NSCs are distributed throughout various plant
540 tissues and serve multiple functions, including maintenance metabolism, osmoregulation, and xylem embolism repair
541 (Dietze et al., 2014; Hartmann & Trumbore, 2016; McDowell et al., 2008; Zhou et al., 2025). This structural
542 simplification may introduce uncertainties in the retrieved parameters, as the model may compensate for the absence
543 of a more complex NSC storage and remobilization system by adjusting carbon allocation or turnover rates to match
544 the observed biomass trajectories.

545 Beyond forest biomass growth curves, additional long-term constraints could further reduce equifinality, particularly
546 for slow soil carbon pools which were not directly constrained in this study. For example, observational constraints on
547 soil organic carbon (SOC) stocks with chronosequence-based SOC trajectories (Walker et al., 2010) or repeated soil
548 resampling from long-term monitoring networks (Bellamy et al., 2005; He et al., 2016) could be used to constrain litter
549 and soil turnover parameters that are otherwise weakly informed by high-frequency fluxes. Where available, isotope-



550 or radiocarbon-informed metrics of SOC mean age and residence time can provide complementary information on
551 long-term decomposition dynamics (He et al., 2016; Trumbore, 2000), helping to separate uncertainties in biomass
552 turnover from those in soil carbon processes. Integrating such slow-pool constraints would enable a more complete
553 constraint on long-term ecosystem carbon storage, because biomass and SOC jointly determine multi-decadal to
554 centennial carbon accumulation (Schimel et al., 2015).

555 Although this study focuses on the boreal biome, the method of jointly constraining fast fluxes with long-term trajectory
556 information should be transferable to other regions. In temperate forests, long-term forest inventories and management
557 records can provide stand-age-dependent biomass trajectories similar to growth-curve constraints (Pan et al., 2011;
558 Pugh et al., 2019). In the tropics, repeated biomass mapping (e.g., lidar- or radar-informed products) combined with
559 disturbance chronosequences may offer similar long-term constraints (Heinrich et al., 2021; Poorter et al., 2016),
560 although high species diversity and mixed disturbance types (e.g., selective logging, windthrow) can complicate
561 interpretation. In semi-arid and savanna systems, long-term constraints on woody cover and carbon stocks may be
562 particularly valuable because short-term fluxes are strongly driven by episodic rainfall and can be highly equifinal
563 (Brandt et al., 2016; Poulter et al., 2014). Extending this framework globally will therefore require region-specific
564 long-term datasets and disturbance histories and may motivate developing growth-curve constraints for multiple
565 disturbance agents rather than fire alone.

566 Our study emphasizes that boreal forest biomass stocks are sustained by slower carbon turnover rates than those
567 inferred from high-frequency flux constraints alone, highlighting the potential vulnerability of this carbon sink to
568 disturbances that shorten carbon residence time. Current warming trends pose multiple threats to this stabilizing effect
569 of long residence times. The increasing fire frequency (Magney & Pierrat, 2025) and drought-induced mortality (Peng
570 et al., 2011) are already reducing carbon residence times. Our analysis identifies a specific non-linear threshold for
571 turnover stability when the temperature of the coldest month rises above -20°C . This finding aligns with thermal
572 constraints on several disturbance agents. For example, winter minimum temperatures ranging from -16°C to -20°C
573 act as a lethal lower limit for major boreal pests, such as the southern pine beetle (*Dendroctonus frontalis*) (Lesk et al.,
574 2017; Ungerer et al., 1999). Crossing this thermal threshold effectively lifts the climatic barrier to pest survival and
575 northward expansion (De Grandpré et al., 2018). Furthermore, warm winters can disrupt tree cold-hardening processes,
576 increasing the risk of frost damage (Hänninen, 2006) or winter embolism (Mayr et al., 2020). This implies that future
577 ESM projections relying on fixed turnover rates may underestimate the risk of a rapid release of stored carbon as these
578 bioclimatic barriers are removed. Together, these results suggest that representing climate-sensitive turnover processes
579 with effective constraints from long-term biomass information may be critical for projecting the persistence of boreal
580 carbon storage under future climate change.

581 **5. Conclusion**

582 This study demonstrates that long-term forest biomass growth trajectories provide an effective constraint on carbon
583 allocation and turnover parameters in terrestrial biosphere models. Using a differentiable DALEC framework, we
584 showed that calibrations based only on high-frequency observations reproduced short-term GPP, RECO, ET, and LAI
585 dynamics but generated large biases in long-term biomass stocks. Adding a single-year biomass constraint partially
586 reduced these biases, whereas assimilating the full biomass growth curves reduced the biomass RMSE from 138.7 to
587 11.3 Mg ha^{-1} without degrading the simulation of fast processes. These results show that slow ecological trajectories
588 provide information that is not captured by flux observations or static biomass snapshots alone.



589 The growth-curve-constrained retrievals indicate that boreal forest biomass stocks are maintained primarily by long
590 woody carbon residence times rather than by high allocation to wood. Compared with the baseline calibration, the
591 optimized wood turnover rate decreased by 74.1%, while the fraction of NPP allocated to wood also declined.
592 Environmental attribution further suggests that boreal carbon residence times are sensitive to climate, with wood
593 turnover increasing sharply when the temperature of the coldest month exceeds approximately -20°C . More broadly,
594 our results highlight the need to assimilate observations that match the timescales of the processes being constrained.
595 Future work should integrate additional long-term constraints, such as soil carbon trajectories, repeated biomass
596 observations, and disturbance-specific recovery curves, to improve projections of ecosystem carbon residence time and
597 the stability of the land carbon sink under climate change.

598 **Code and Data Availability**

599 The code and data that support the findings of this study are openly available in Zenodo at
600 <https://zenodo.org/records/19135427> (Wu, 2026).

601 **Author Contributions**

602 Conceptualization: W.L. and P.C. Methodology: J.W. Investigation: J.W. Visualization: J.W. Writing—original draft:
603 J.W. Writing—review & editing: All authors.

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609 **Conflicts of Interest**

610 The authors declare no conflicts of interest.

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