

Indoor Burning of Arabian Incense Generates Abundant Ultrafine Particles with Strong Oxidative Potential

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Abstract. Arabian incense (Bakhoor) burning is a widely practiced fragancing and ceremonial activity, yet how the Bakhoor composition controls particle emissions and oxidative potential remains poorly constrained, especially under repeated use in low-ventilation settings. Here we characterized emissions from Bakhoor burning in a controlled chamber using a charcoal-assisted heating configuration representative of common practice and quantified aerosol oxidative potential using complementary acellular dithiothreitol (DTT) activity and a macrophage-based intracellular oxidative-stress response, with smoldering sidestream cigarette smoke as a protocol-matched indoor combustion reference. Normalized by the initial Bakhoor mass per burn, Bakhoor burning produced particle mass and number emission rates of 670-1690 $\mu\text{g min}^{-1} \text{g}^{-1}$ and $(6-7) \times 10^{11}$ particles $\text{min}^{-1} \text{g}^{-1}$, respectively. Ultrafine particles contributed 70-75% of the total particle number, and their emission rates substantially exceeded those from sidestream cigarette smoke. Across Bakhoor materials, emission magnitude followed a nonlinear power-law relationship with the loading of the hexane-soluble fraction, indicating that this fraction is an important control on particle production. In the acellular assay, the water-soluble particle mass-normalized DTT consumption rate ($OP_{DTT,m}^{WS}$) was approximately 32 $\text{pmol min}^{-1} \mu\text{g}^{-1}$, modestly lower than that of cigarette smoke particles, whereas Bakhoor burning particles elicited stronger intracellular oxidative-stress responses. Ozone aging increased oxidative potential for both sources, and the acellular and cellular responses remained evident after aging equivalent to days of indoor exposure. Overall, Bakhoor burning represents a previously underrecognized source of ultrafine aerosol with substantial oxidative potential.

33 1. Introduction

34 Global exposure assessments indicate that people spend approximately 60-90% of their time indoors (Sadoun et al., 2016;
35 Pillarisetti et al., 2022), making indoor environments an important setting for air pollutant exposure. In arid regions such as
36 the Middle East, this percentage can approach 100% (Sadoun et al., 2016), driven by extended periods of extreme heat and
37 limited outdoor comfort. Despite this high exposure potential, indoor air quality has historically received less scientific and
38 regulatory attention than outdoor pollution (Cincinelli and Martellini, 2017). Indoor air is shaped by both outdoor infiltration
39 and indoor activities, with combustion-related processes often contributing substantially to particle exposure (Gligorovski and
40 Abbatt, 2018). While major indoor combustion sources such as tobacco smoke and household solid-fuel use have been widely
41 studied (Wang et al., 2018a), other routine combustion-based fragancing and ceremonial practices remain comparatively
42 under-characterized as particle sources.

43 A prominent example is the burning of Arabian incense (Bakhoor), a widely practiced fragancing and ceremonial
44 activity in the Middle East and other regions, in which a mixture of perfumed wood chips soaked with additives is heated on
45 charcoal. In addition to the aromatic wood chips, Bakhoor materials often contain scented oils, resins, and botanical additives
46 such as herbs and flowers (Wahab and Mostafa, 2007), forming a chemically complex mixture that can generate condensable
47 precursors upon thermal degradation. Because Bakhoor is typically heated on smoldering charcoal, both the Bakhoor material
48 and the ignition source can contribute to particle formation. Bakhoor burning commonly occurs indoors, often in enclosed
49 spaces with limited ventilation, and is embedded in daily routines as well as religious and social gatherings. In Jazan, Saudi
50 Arabia, 98% of households report regularly burning Bakhoor indoors (Jareebi et al., 2024). Frequent incense use has also been
51 reported in Gulf regions (Yeatts et al., 2012), while household incense burning is likewise prevalent in parts of Asia (Pan et
52 al., 2014; Zhao et al., 2025). This widespread use indicates that combustion-based fragancing and ceremonial practices are
53 not only regionally important, but also relevant to a broader class of indoor combustion sources. For such sources, quantitative
54 descriptors such as source strength, particle size distribution, and particle reactivity are important for understanding emission
55 characteristics and how they vary with source composition and burning conditions (Kuye and Kumar, 2023). Ultrafine particles
56 are often of particular interest because of their importance to particle number emissions and their potential sensitivity to source
57 and combustion conditions (Shen et al., 2017).

58 Previous studies have provided evidence that Bakhoor and related incense smoke can have adverse respiratory and
59 toxicological effects. Epidemiological work has linked domestic bakhoor use with respiratory symptoms. For example, Al-
60 Rawas et al. (2009) reported that Arabian incense burning was a common trigger of wheezing among asthmatic children in
61 Oman. Cohen et al. (2013) characterized particles and gases emitted from United Arab Emirates (UAE) incense burning under
62 indoor chamber conditions and reported inflammatory responses in exposed human lung cells, providing an important prior
63 hazard-assessment study of Arabian incense smoke. The chemical complexity of Bakhoor further complicates source
64 interpretation. Elsayed et al. (2016) reported that raw, unburned Bakhoor materials contain a chemically complex mixture of
65 organic constituents, including nitrogen-containing and other additive-related species. However, the composition of particles
66 emitted during Bakhoor burning can differ substantially from the raw-material profile because thermochemical processing

67 during heating and pyrolysis can transform the formulation and generate new condensable products that partition to the particle
68 phase. Dalibalta et al. (2015) provided preliminary indications of potential concern in Bakhoor smoke by noting the presence
69 of compounds classified as carcinogenic, toxic, and/or respiratory irritants; however, the evidence was not linked to size-
70 resolved emissions or dose-relevant particle metrics. Alarifi et al. (2004) further reported structural changes in pneumocytes
71 following repeated animal exposure to Bakhoor smoke. Together, these studies establish a health-relevant basis for concern.
72 However, the particle-level basis for assessing exposure from practical Bakhoor burning remains insufficiently constrained,
73 including how emissions vary with source configuration and product composition, and how the emitted particles express
74 oxidative potential before and after indoor-relevant aging. These gaps motivate a quantitative particle-based evaluation of
75 Bakhoor burning.

76 Here, we characterize particle emissions and oxidative responses of Bakhoor burning in a controlled chamber, using
77 smoldering sidestream cigarette smoke as a protocol-matched indoor combustion reference (Chen and Zhao, 2024). Real-time
78 particle number and mass concentrations and size distributions were measured using a scanning mobility particle sizer, and
79 size-resolved emission factors were quantified. Particular attention was given to ultrafine particles (UFPs, <100 nm) because
80 of their high respiratory deposition efficiency and potential relevance to systemic health effects (Oberdörster et al., 2005).
81 Control experiments were performed to quantify the respective contributions of the Bakhoor material and the ignition charcoal
82 to emitted particle size distributions and emission factors. Oxidative responses were evaluated using two complementary
83 approaches commonly applied in atmospheric and exposure research (Seo et al., 2025). These included acellular oxidative
84 potential quantified by the dithiothreitol (DTT) assay and cellular responses assessed via intracellular oxidative-stress response
85 and cell viability in alveolar macrophages. To probe potential chemical drivers, particle extracts were analyzed offline using
86 high-performance liquid chromatography coupled with high-resolution mass spectrometry, and compositional metrics were
87 examined in relation to DTT activity. Changes in chemical composition and associated oxidative responses of Bakhoor burning
88 particles were also evaluated following ozone exposure as an additional probe of post-emission processing. This work aims to
89 better constrain the size-resolved emissions, source contributions, and oxidative responses of particles emitted from Bakhoor
90 burning.

91

92 **2. Methods**

93 **2.1 Generation and collection of particles from Bakhoor burning and sidestream cigarette smoke**

94 All burning experiments were conducted inside a custom-built acrylic chamber (50 × 50 × 50 cm). Three commercially
95 available Bakhoor products were purchased from local markets in Saudi Arabia, selected based on consumer popularity. For
96 each experiment, 1 g of Bakhoor was placed onto a smoldering charcoal briquette, consistent with common household practice.
97 The charcoal was a commercially available quick-lighting type typically used for Bakhoor burning. Charcoal-only burns were
98 conducted under the same chamber configuration to quantify emissions attributable to the ignition source. As a protocol-
99 matched indoor combustion reference, smoldering sidestream cigarette smoke was generated in the same chamber to provide
100 a stable and continuous particle stream. Although this approach differs from puffing-based protocols, it improves

101 reproducibility and has been used in prior emission studies (Ott et al., 2021; Wu et al., 2012). Once Bakhoor was placed on
 102 the smoldering charcoal, visible smoke generation began immediately and typically persisted for approximately 20 min. The
 103 chamber was operated under a steady-state flow-through mode with purified air supplied at 20 L min⁻¹. Size-resolved particle
 104 number concentrations were continuously measured using a Scanning Electrical Mobility Spectrometer (SEMS; Model 2100,
 105 Brechtel Manufacturing Inc.) over a mobility diameter range of 10–500 nm. Larger particles (500–2500 nm) were quantified
 106 using an optical particle counter (OPC; Brechtel Manufacturing Inc.). Prior to each experiment, background particle
 107 concentrations in the chamber were reduced to <1 particle cm⁻³. Particle number size distributions from the SEMS and OPC
 108 were converted to mass concentrations assuming an effective particle density of 1 g cm⁻³. This value has been used in previous
 109 indoor particle and combustion-emission studies to estimate particle mass from size distributions (Wallace, 2006;
 110 Klosterk  ther et al., 2021) and is close to the reported density of incense smoke particles, approximately 1.1 g cm⁻³ (Ji et al.,
 111 2010). Fresh combustion aerosols may have source- and size-dependent effective densities because of differences in
 112 morphology and composition. In addition to real-time measurements, aerosol samples were collected on 47 mm quartz and
 113 PTFE filters, each at approximately 6 L min⁻¹ for offline chemical analysis. Quartz filters were pre-baked at 550  C for 12 h
 114 to minimize background organic contamination.

115 **2.2 Emission rate calculations**

116 Particle mass emission rates (E , $\mu\text{g min}^{-1}$) were quantified using a size-resolved material-balance approach (Wang and Chan,
 117 2023; Jiang et al., 2021), where particle losses were treated separately for each size bin. The chamber was continuously mixed
 118 using two internal fans, and emission rates were therefore derived using a well-mixed flow-through chamber mass-balance
 119 framework. For analysis, particles were grouped into the following diameter bins: 10-100, 100-200, 200-300, 300-400, 400-
 120 500, 500-1000, 1000-1500, 1500-2000, and 2000-2500 nm.

121 For each size bin i , the time-dependent mass balance in the well-mixed chamber is

$$122 \quad V \frac{dC_i}{dt} = -F \times C_i + E_i - k_i \times V \times C_i, \quad (\text{Eq. 1})$$

123 where V is the volume of the chamber (m^3), C_i is the mass concentration of particles in size bin i ($\mu\text{g m}^{-3}$), F is the outlet airflow
 124 rate ($\text{m}^3 \text{min}^{-1}$), E_i is the particle emission rate in bin i ($\mu\text{g min}^{-1}$), and k_i is the loss rate constant for particles in bin i (min^{-1}).

125 Particle losses were assumed to follow first-order decay, consistent with previous chamber-based emission studies (Wang and
 126 Chan, 2023). Following each Bakhoor burning event, particle concentrations exhibited near-exponential decay, which was
 127 used to estimate k_i by fitting the decay curves, following the approach described by Jiang et al. (2021). Emission rates were
 128 then integrated over the full burning period and summed across all size bins to obtain the total particulate emission rate.

129 Particle formation, growth, and early size evolution were interpreted in the context of the controlled flow-through chamber
 130 configuration. The chamber volume was 125 L and the outlet flow rate was 20 L min⁻¹, giving a dilution rate of 0.16 min⁻¹,
 131 equivalent to an air exchange rate of 9.6 h⁻¹, and a flow-through residence time of 6.25 min. Because this configuration
 132 combines a high source-to-volume ratio with flow-through dilution, the measured size distributions should not be interpreted
 133 as direct room-averaged particle concentrations or room-scale aerosol distributions. Instead, they are best viewed as controlled

134 near-source source-term measurements from the practical Bakhoor-on-charcoal burning configuration. Under this source-term
135 interpretation, the measured size distributions include the outcome of near-source particle formation and early growth. For
136 room-scale applications, these measured emission rates and size distributions can be used as source inputs, whereas subsequent
137 evolution will depend on room volume, ventilation, mixing, deposition, surface losses, coagulation, and possible semivolatile
138 evaporation or gas-particle repartitioning.

139 **2.3 Ozone aging of particle samples**

140 To simulate oxidative aging, PM-loaded filters were placed in an in-line filter holder and exposed to a continuous flow of
141 ozone diluted in zero air. Ozone was generated by passing zero air through a mercury lamp ozone generator (Model 610,
142 Jelight Inc., USA), and its concentration was continuously monitored at the outlet of the filter holder using an ozone analyzer
143 (Model 106-L, 2B Technologies Inc.). Two ozone levels, approximately 200 ppb and 1500 ppb, were applied for the same 30
144 min duration to vary the cumulative O₃ exposure while keeping filter handling and exposure time consistent. These exposures
145 correspond to approximately 20 h and 150 h of equivalent indoor ozone exposure, respectively, assuming a representative
146 indoor ozone level of 5 ppb (Nazaroff and Weschler, 2022). This dose equivalence was used to approximate extended indoor
147 O₃ exposure, while recognizing that concentration-dependent kinetics and phase partitioning may differ between accelerated
148 and lower-concentration aging conditions. The lower exposure was selected to approximate airborne residence-scale O₃
149 processing, consistent with reported residential air-exchange-based particle residence times on the order of hours to about one
150 day (Zhao and Liu, 2020; Salthammer, 2011). The higher exposure was selected to probe more extended O₃ exposure relevant
151 to deposited indoor particle reservoirs, which can remain on indoor surfaces and undergo heterogeneous or multiphase
152 chemical processing after removal from the airborne phase (Abbatt and Wang, 2020; Ault et al., 2020). This accelerated
153 protocol was selected for practicality and reproducibility and to isolate ozone-driven processing under controlled conditions.
154 Although filter-bound PM could, in principle, be exposed to lower ozone concentrations for longer periods, multi-day
155 treatments increase the likelihood of uncontrolled changes that are not directly attributable to ozone, including slow
156 volatilization of semi-volatile constituents, continued dark aging, and time-dependent particle-filter interactions. Using
157 elevated ozone over short durations therefore enables consistent, well-defined oxidant doses while minimizing these
158 confounding processes. Similar filter-based O₃ exposure of particle-loaded samples has been used previously to evaluate
159 changes in aerosol oxidative potential, including for biomass-burning organic aerosol (Wong et al., 2019). This filter-bound
160 O₃ treatment provides a practical approach to probe the O₃ sensitivity of Bakhoor-burning particles. O₃ was selected because
161 it is widely regarded as an important indoor oxidant, particularly for heterogeneous and surface chemistry. It is commonly
162 introduced indoors through air exchange, and can be delivered as a stable and controllable oxidant dose for reproducible
163 particle-aging experiments (Waring and Wells, 2015; Weschler and Carslaw, 2018). Indoor particles may experience O₃
164 exposure both while airborne and after deposition onto indoor surfaces; deposited particles can persist longer and remain
165 exposed to oxidants. Accordingly, this experiment should be interpreted as a controlled test of the post-emission O₃
166 responsiveness of Bakhoor-burning particle-phase material, rather than as a full simulation of suspended-particle aging in
167 indoor air. Compared with suspended particles, filter-bound particles no longer evolve within the same gas-particle partitioning

environment, and their exposure to O₃ may be influenced by particle loading and contact with the filter substrate. In addition, indoor OH and NO₃ chemistry can occur but is more condition-dependent than the controlled O₃ exposure used here. Indoor OH can be generated through ozonolysis and photolysis-related processes, whereas NO₃ chemistry becomes more relevant under NO₂/O₃-rich conditions (Waring and Wells, 2015; Abbatt and Wang, 2020; Zannoni et al., 2022; Dewald et al., 2023). Therefore, OH- and NO₃-driven processing should be regarded as additional indoor aging pathways that were not simulated by the present filter-bound O₃ exposure and should be evaluated in future suspended-particle or mixed-oxidant studies.

2.4 Molecular characterization of organics in Bakhoor-burning particles

To analyze the molecular composition of organic species in Bakhoor-burning particles, filter samples were extracted in acetonitrile via 30-minute sonication in an ice-water bath to minimize sonication-induced heating and potential thermal degradation, followed by filtration through a 0.22 µm PTFE syringe filter. Acetonitrile was selected over methanol to minimize solvent-induced artifacts, such as the methanolysis of conjugated carbonyl species (Chen et al., 2022). The extracts were analyzed using ultra-high-performance liquid chromatography coupled with high-resolution Orbitrap mass spectrometry (UHPLC-HRMS; Orbitrap ID-X, Thermo Fisher Scientific Inc.), equipped with an electrospray ionization (ESI) source operated in positive ion mode. Chromatographic separation was performed on an Acquity HSS T3 column (1.8 µm, 2.1 mm × 100 mm; Waters Corp.) under conditions optimized for semi-polar organic compounds (Wang et al., 2021; Go et al., 2022). The liquid chromatography gradient and flow parameters were configured to improve separation efficiency and minimize matrix effects, following established protocols for analyzing dissolved organic matter and aerosol extracts (Go et al., 2022; Liang et al., 2024; Patriarca et al., 2018). Data acquisition and molecular formula assignment were conducted using Compound Discoverer software (Thermo Fisher Scientific Inc.), with elemental compositions determined from exact m/z values and isotopic patterns. This analytical workflow enabled untargeted molecular profiling of the complex organic mixtures emitted from Bakhoor burning.

2.5 Acellular dithiothreitol (DTT) assay

The oxidative potential of particulate matter is a key indicator of its capacity to generate reactive oxygen species (ROS) and induce oxidative stress (Jiang et al., 2019). Among acellular assays, the DTT assay is widely used as a proxy for particle redox activity and quantification of the rate of DTT consumption by redox-active species that catalyze electron-transfer reactions (Cho et al., 2005). The experimental protocol followed established methods reported previously (Gao et al., 2017; Fang et al., 2017; Lyu et al., 2018; Cho et al., 2005). Briefly, PM-loaded PTFE filter samples were extracted in deionized water by 30 minutes of sonication in an ice-water bath to minimize heating during extraction. Given the potential for radical formation during sonication, a comparison between sonication and vortex shaking was conducted. The results showed no significant difference in DTT consumption, indicating that sonication did not introduce measurable artifacts. The PM mass used to calculate mass-normalized DTT oxidative potential was determined gravimetrically from the difference between the pre- and post-sampling masses of the PTFE filters. For the main Bakhoor-burning and cigarette-smoke samples used in the DTT comparison, the collected PM masses varied within about a factor of four before extraction. For the measurement of water-soluble oxidative potential ($OP_{DTT,m}^{WS}$), the extract was filtered and 0.7 mL of the filtrate was mixed with 0.1 mL of 1 mM DTT

202 solution and 0.2 mL of potassium phosphate buffer (0.5 M, pH 7.4). The mixture was incubated at 37 °C in a thermostatic
 203 shaker with continuous agitation. At designated time points (0, 5, 10, 15, 25, and 35 minutes), 0.1 mL aliquots were withdrawn
 204 and quenched by adding trichloroacetic acid (TCA). The residual DTT in each aliquot was then reacted with Tris buffer (0.4
 205 M containing 20 mM EDTA) and 5,5'-dithiobis-(2-nitrobenzoic acid) (DTNB). The reaction yields a yellow chromophore that
 206 absorbs at 412 nm, which was quantified using a microplate reader. The DTT consumption rate was determined from the linear
 207 decrease in residual DTT over the incubation period. To determine the total oxidative potential ($OP_{DTT,m}^{tot}$), the same procedure
 208 was used, except the extract was not filtered, and the filter itself remained in the reaction mixture. This ensured that DTT-
 209 active species associated with insoluble particles suspended in the extract or still attached to the filter surface were also included
 210 in the measurement (Fang et al., 2017; Gao et al., 2017). Blank PTFE filters processed using the same protocol were used for
 211 background correction. Because mass-normalized DTT activity can depend on particle loading and on the concentration-
 212 response behavior of the dominant DTT-active species (Charrier et al., 2016), mass-normalized DTT oxidative-potential
 213 comparisons were interpreted with this potential loading-related uncertainty in mind. However, the Chelex tests showed
 214 negligible changes in $OP_{DTT,m}^{WS}$ after metal chelation, suggesting that the specific Cu/Mn-driven mass-normalization bias
 215 reported by Charrier et al. (2016) is unlikely to dominate the source-level comparison in this study. To minimize photo-
 216 degradation of DTT and DTNB, all procedures were performed under low-light conditions by minimizing workspace lighting
 217 and covering vials with aluminum foil when not in use.

218 **2.6 Cellular macrophage assays**

219 Alveolar macrophages represent the first line of defense in the respiratory system and are widely used to assess oxidative and
 220 toxicological responses to PM exposure, particularly through intracellular ROS generation (Liu et al., 2023; Liu et al., 2020b;
 221 Liu et al., 2020a; Tuet et al., 2019). In this study, murine alveolar macrophages (MH-S, ATCC CRL-2019) were cultured in
 222 RPMI-1640 medium (ATCC) supplemented with 10% fetal bovine serum (FBS; VWR), 1% penicillin–streptomycin (Pen-
 223 Strep; VWR), and 50 μ M β -mercaptoethanol (BME; Sigma-Aldrich). Cells were maintained at 37 °C in a humidified 5% CO₂
 224 incubator. Prior to ROS measurements, 96-well plates were pre-coated with 10% FBS in phosphate-buffered saline (PBS) to
 225 promote cell adherence. MH-S cells were seeded at a density of 2×10^4 cells per well and incubated with the ROS-sensitive
 226 fluorescent probe carboxy-H₂DCFDA (Molecular Probes, C-400). To introduce particulate exposure, PTFE filters loaded with
 227 fresh or ozone-aged Bakhoor-burning particles or sidestream cigarette-smoke particles were directly immersed in the culture
 228 medium. A 3-minute vortex mixing step was applied to enhance dispersion of PM into the media. Cells were exposed for 24
 229 hours, after which the filters were removed and intracellular oxidative-stress responses were measured using a microplate
 230 reader (excitation/emission: 485/525 nm). All treatments were conducted in triplicate using independent filter samples. Blank
 231 controls (media only, no PM) were used to normalize fluorescence values. Blank PTFE filter controls processed under the
 232 same immersion, vortexing, and exposure protocol were used to normalize DCFH fluorescence values. The cellular exposure
 233 experiments were designed to compare oxidative-stress responses across particle sources and aging conditions, with sidestream
 234 cigarette smoke included as a protocol-matched indoor combustion reference; no separate chemical positive control for
 235 intracellular ROS generation was included in the same exposure batch.

236 A mitochondrial dehydrogenase (MTT) assay was conducted to evaluate cell viability following 24-hour exposure to
237 PM. After exposure, the culture medium in each well was replaced with 100 μL of fresh medium, and 10 μL of MTT solution
238 (5 mg mL^{-1}) was added. Cells were incubated for an additional 4 hours at 37°C to facilitate the formation of purple formazan
239 crystals. Following this, 100 μL of a solubilization solution consisting of 10% SDS in 0.01 M HCl was added to each well,
240 and incubation was continued for a further 15–18 hours under the same conditions. The absorbance at 570 nm was then
241 measured using a microplate reader to quantify formazan formation, which serves as an indicator of mitochondrial enzymatic
242 activity and, consequently, cell viability. A viability calibration curve was established by preparing mixtures of live and heat-
243 inactivated (autoclaved) cells at defined ratios, treating them with the same MTT procedure, and recording their absorbance at
244 570 nm to relate optical density to cell survival.

245

246 3. Results and discussion

247 3.1 Emission characteristics of Bakhoor burning particles

248 Figure 1A-B show the size distributions of particle number and mass for emissions from Bakhoor burning, with smoldering
249 sidestream cigarette smoke included as a protocol-matched indoor combustion reference. Because Bakhoor was heated on a
250 smoldering charcoal briquette, the measured distributions reflect combined contributions from the Bakhoor material and the
251 ignition source. A charcoal-only control is therefore included to isolate charcoal-derived emissions. In the number distribution
252 (Figure 1A), charcoal-only burning produced a weak nucleation-mode signal largely confined to the ultrafine range, whereas
253 Bakhoor burned on charcoal yielded substantially higher concentrations and a bimodal profile, with a dominant ultrafine mode
254 near 30 nm and a secondary mode around 160 nm. The ultrafine-mode peak was approximately 1.4 times the secondary peak,
255 indicating a large contribution of ultrafine particles (UFPs, $<100 \text{ nm}$) to the total particle number. Sidestream cigarette smoke,
256 by comparison, exhibited a unimodal number distribution centered at approximately 110 nm under the protocol used here, with
257 a substantially lower UFP contribution. In the mass distribution (Figure 1B), Bakhoor burning produced a unimodal profile
258 centered at approximately 350 nm, indicating that emitted mass was dominated by accumulation-mode particles. Time-
259 resolved measurements further show that particle concentrations rose within seconds after ignition, peaked at approximately 5
260 min, and then decayed (Figure 1C).

261 Figure 1D summarizes emission rates normalized to the initial Bakhoor mass per burn and to the mass of cigarette
262 tobacco filler consumed in the smoldering sidestream cigarette-smoke reference. On this basis, Bakhoor burning produced
263 particle mass emission rates of $670\text{--}1690 \mu\text{g min}^{-1} \text{ g}^{-1}$, substantially higher than the same-protocol smoldering sidestream
264 cigarette smoke reference in this study ($\sim 250 \mu\text{g min}^{-1} \text{ g}^{-1}$). The cigarette-smoke comparison was based on a controlled same-
265 protocol smoldering condition without active puffing. Previous measurements comparing sidestream particulate matter under
266 puffing and free-burn conditions showed that sidestream particulate matter under puffing conditions remained within
267 approximately 0.84–0.97 times the free-burn value (Browne et al., 1980). In the context of typical indoor sources, these mass
268 emission rates place Bakhoor burning toward the upper end of reported emission rates for activities including tobacco and
269 incense burning, moxa, candles, and non-combustion sources such as cooking and e-cigarettes ($117\text{--}3367 \mu\text{g min}^{-1}$), while

recognizing that reported rates can vary with experimental configuration and normalization (Hu et al., 2023; Chuang et al., 2013; Jetter et al., 2002; Tian et al., 2021). Unlike conventional incense burning, however, Bakhoor combines a charcoal ignition source with a chemically complex mixture of perfumed wood and additives. Number-based emission rates for Bakhoor reached $(6-7) \times 10^{11}$ particles $\text{min}^{-1} \text{g}^{-1}$, with UFPs accounting for 70–75% of total number, compared with ~47% for the smoldering cigarette-smoke reference. Taken together, these results identify Bakhoor burning as a high-intensity indoor particle source that is strongly enriched in UFPs by number while also producing substantial fine-particle mass emissions. Because Bakhoor burning generated high particle number concentrations in the chamber, particle-particle coagulation may contribute to the evolution of the measured size distribution. To evaluate its potential influence under our experimental conditions, representative same-size Brownian coagulation kernels were calculated for 20 and 100 nm particles using the Fuchs transition-regime formulation. The estimated kernels were 2.35×10^{-9} and $1.43 \times 10^{-9} \text{ cm}^3 \text{ s}^{-1}$, respectively (Hinds and Zhu, 2022; Seinfeld and Pandis, 2016). With the burning-period-averaged UFP number concentration of approximately $1.0 \times 10^6 \text{ cm}^{-3}$, these kernels correspond to coagulation timescales of approximately 7.1 and 11.6 min for 20 and 100 nm particles, respectively. These values are comparable to or longer than the chamber flow-through residence time of 6.25 min, suggesting that coagulation may influence the smaller end of the UFP mode but is unlikely to dominate the burning-period-averaged particle number estimate. Since coagulation reduces particle number and shifts particles toward larger diameters, the reported UFP number emission rates may represent conservative estimates.

3.2 The potential drivers of particle formation during Bakhoor burning

Figure 2A shows that charcoal-only burning produces a short-lived nucleation-mode signal immediately after ignition that decays rapidly, whereas Bakhoor burning sustains elevated particle concentrations and extends to larger particle diameters. This side-by-side comparison was conducted under the same chamber flow-through configuration and indicates that the sustained particle formation observed during Bakhoor burning cannot be explained by the charcoal ignition source alone. Instead, it is consistent with additional particle-forming material released from Bakhoor. To identify the Bakhoor components responsible for this behavior, Bakhoor materials were operationally separated into the hexane-soluble fraction (HSF) and the wood residue. The HSF is a solvent-defined operational fraction containing hexane-soluble organic material, which may include fragrance oils, resins, and related additives. The residue represents the nonextractable wood matrix. Burning the corresponding wood residues yielded nearly identical particle mass emission rates across the three Bakhoor types, whereas the original Bakhoor materials varied by approximately 2.5-fold in mass emission rates (Figure 2B). This contrast indicates that the observed differences among commercial products are primarily driven by the HSF rather than by the base wood substrate under the conditions used here. Because solvent extraction could, in principle, alter the physical structure of the remaining wood residue and independently influence combustion behavior, we conducted a reconstitution experiment as a functional control. In this experiment, the extracted HSF was recombined with the corresponding wood residue at the appropriate mass fractions and burned under the same conditions. The reconstituted material reproduced particle mass emission rates similar to those of the original samples (Figure S3), suggesting that extraction-induced changes to the residual wood matrix did not

dominate the observed emission differences. We therefore interpret the HSF as an operationally defined fraction that plays a major role in particle formation during Bakhoor burning.

Across the original Bakhoor samples, the HSF loading (9-22% by mass) was strongly associated with particle mass emissions (Figure 2B), with Type 2 exhibiting both the highest HSF loading and the highest mass emission rate. Perturbation experiments that partially reduced the HSF (tissue wiping) and increased it by recombining the HSF with the wood residue collapsed onto a common relationship between emission rate and HSF loading across all Bakhoor types. This relationship supports HSF loading as the primary control on emission magnitude, even though the composition of the HSF varied slightly among samples (Figure S4). This dependence was well described by a power-law relationship ($\beta \approx 2.1$; Figure 2B), demonstrating a nonlinear response in which emissions increase more steeply as the HSF increases. Adding HSF to reach approximately 10 wt % of the reconstructed material increased mass emission rates by ~ 1.7 -fold, whereas increasing the HSF content to approximately 20 wt % increased emissions by ~ 4.3 -fold. This pattern suggests a threshold-like increase in effective particle formation efficiency, where incremental increases in available precursors result in disproportionate increases in particle production (Zhang et al., 2010).

Thermogravimetric analysis (TGA) was used to assess whether the observed HSF dependence could be primarily explained by simple evaporation of intact additives. Across the three Bakhoor types, thermograms were similar and showed only minor mass loss (approximately 5%) below 100-120 °C (Figure S5), whereas most mass loss occurred between approximately 150 and 350 °C. Thermocouple measurements showed that the charcoal temperature reached approximately 300 °C within 15 s of sustained ignition in our setup (Figure S6), suggesting that Bakhoor materials are quickly driven into a higher-temperature regime rather than remaining in a low-temperature evaporation window. The reported onset of wood pyrolysis (approximately 180 °C) refers primarily to the lignocellulosic wood substrate (Escalante et al., 2022). In contrast, the HSF consists of nonstructural, solvent-extractable organics introduced during product formulation, which may volatilize and thermally transform over overlapping temperature windows during heating on charcoal. Taken together, these results suggest that HSF-linked particle formation is not solely explained by low-temperature evaporation and likely involves thermochemical processing during heating that generates condensable products, as illustrated in Figure 2C. Because TGA does not provide a mass balance for emitted particles, it is used here to bracket plausible precursor-generation regimes rather than to quantify particle yields.

3.3 Oxidative potential of particles from Bakhoor burning

To evaluate the oxidative potential of particles emitted from Bakhoor burning, we applied the dithiothreitol (DTT) assay. We report the particle mass-normalized DTT oxidative potential for both filtered aqueous extracts ($OP_{DTT,m}^{WS}$, $\text{pmol min}^{-1} \mu\text{g}^{-1}$) and total unfiltered aqueous suspensions ($OP_{DTT,m}^{tot}$, $\text{pmol min}^{-1} \mu\text{g}^{-1}$) to gauge the relative contributions of water-soluble versus suspension-associated components. Fresh Bakhoor water extracts exhibited an average $OP_{DTT,m}^{WS}$ of approximately 32 $\text{pmol min}^{-1} \mu\text{g}^{-1}$ (Figure 3A). This value was not corrected by subtracting the charcoal-only burn and therefore represents the exposure-relevant oxidative potential of the full particle mixture emitted under the common real-use scenario in which Bakhoor

is heated on smoldering charcoal. This magnitude lies within reported ranges for laboratory-generated particles from indoor combustion sources (8-95 pmol min⁻¹ μg⁻¹; Hu et al., 2023; Zhang et al., 2024) and for raw wood/biomass materials such as pine (35-117 pmol min⁻¹ μg⁻¹; Zhang et al., 2024), and is comparable to values reported for ambient PM and biomass-burning organic aerosol fractions. Notably, the $OP_{DTT,m}^{WS}$ of Bakhoor burning particles exceeds the typical range for wildfire-influenced PM (2-16 pmol min⁻¹ μg⁻¹; Fang et al., 2023), reaching values near the upper end of the wildfire range (around 32 pmol min⁻¹ μg⁻¹; Isenor et al., 2025). Under the same assay conditions, cigarette smoke particles showed a higher $OP_{DTT,m}^{WS}$ of approximately 50 pmol min⁻¹ μg⁻¹, about 40% higher than Bakhoor burning particles. To assess the contribution of insoluble components, $OP_{DTT,m}^{tot}$ was also measured for total particle suspensions (Fang et al., 2023; Charrier and Anastasio, 2012; Brehmer et al., 2019). Only a modest enhancement (approximately 10%) was observed relative to filtered extracts (Figure 3A), suggesting that DTT-reactive species are captured predominantly in the aqueous phase under the extraction conditions used here.

Oxidative aging can further shape exposure-relevant redox properties because indoor oxidants can transform particle composition and, in turn, DTT consumption (Wong et al., 2019; Wang et al., 2023). Following the O₃-aging protocol described in Section 2.3, we evaluate here how controlled O₃ exposure alters the oxidative potential of Bakhoor-burning and cigarette-smoke particles. Residential air exchange rates typically range from 0.05 to 1 h⁻¹ (Zhao and Liu, 2020; Salthammer, 2011), corresponding to ventilation-based particle residence times of approximately 1-20 h. Moreover, the high indoor surface-to-volume ratio promotes rapid particle deposition to surfaces (Abbatt and Wang, 2020), creating surface reservoirs that are not removed by ventilation and can persist for days. These deposited materials can continue to undergo chemical processing on indoor surfaces and may remain relevant to exposure if later disturbed or transferred by human activity, air movement, or cleaning. After an ozone dose equivalent to approximately 150 h at an indoor ozone level of 5 ppb (Figure 3B), $OP_{DTT,m}^{WS}$ increased by approximately 15% for both Bakhoor and cigarette smoke particles. This increase suggests the net formation of more DTT-active oxygenated products during ozonolysis, consistent with a rise in the carbon oxidation state (OSc) derived from mass spectrometric analysis (Figures S7 and S8). Across fresh and O₃-aged samples, the median $OP_{DTT,m}^{WS}$ of Bakhoor water extracts fell within the broad range reported for laboratory combustion-derived PM and overlapped values reported for field PM (Figure 3C), although many field studies report total $OP_{DTT,m}^{tot}$, which is typically higher than the water-soluble fraction. Prior work by Wong et al. (2019) reported an initial increase in DTT activity (~20%) during the early stages of ozonolysis of biomass-burning particles, followed by a decrease toward baseline after extended aging (up to 70 h), suggesting that some DTT-active products can be transient. Quinone-like redox-cycling species, which catalyze ROS formation, have been proposed as plausible contributors to the observed changes in oxidative potential (Xiong et al., 2017). These observations motivate identifying the chemical drivers that govern DTT consumption in Bakhoor smoke.

3.4 Potential chemical contributors to DTT activity

To examine which particle-phase constituents may underlie the observed DTT response, we analyzed extracts of Bakhoor burning and cigarette smoke particles using UHPLC-HRMS. Both sources showed broad distributions of detected features

spanning $m/z \approx 100$ –400 (Figure S9), consistent with LC/ESI-MS characterizations of biomass-burning organic aerosol and combustion-derived organic aerosol (Smith et al., 2009). Despite comparable m/z envelopes, the assigned elemental-composition classes differed markedly between the two sources. Bakhooor burning particles were dominated by CHO formulas (~65% of assigned species), whereas cigarette smoke particles were relatively less oxygenated and strongly enriched in nitrogen-containing formulas (CHN+CHON $\approx$ 84%) (Figure S9). These comparisons were based on relative ESI+ feature intensities, calculated as the signal contribution of each assigned feature normalized to the total assigned ESI+ signal within the same sample. This normalization emphasizes differences in the distribution of detected molecular features between Bakhooor-burning and cigarette-smoke particles, rather than differences in total extracted signal or absolute ion abundance. Because all samples were analyzed using the same extraction, chromatographic, ionization, and data-processing workflow, the observed contrast provides an internally consistent ESI+-resolved source fingerprint comparison. Electrospray ionization polarity is analyte-dependent, and positive and negative modes provide complementary molecular coverage (Liigand et al., 2017; Patriarca et al., 2018). Therefore, the stronger CHN/CHON contribution detected in cigarette smoke may reflect both real source differences and favorable ESI+ response of nitrogen-containing compounds, while acidic or highly oxygenated CHO species may be less efficiently represented under ESI+ conditions. Accordingly, these molecular patterns should be interpreted as ESI+-resolved source fingerprints rather than a complete quantitative inventory of particle-phase organics. Since elemental class alone provides limited insight into structural motifs relevant to redox cycling, we also computed the modified aromaticity index (AI_{mod} , see Text S1 for details) as a proxy for unsaturation and aromatic character and classified formulas as aromatic ($AI_{mod} \geq 0.5$), olefinic ($0 < AI_{mod} < 0.5$), or aliphatic ($AI_{mod} = 0$) (Koch and Dittmar, 2006; Song et al., 2021; Song et al., 2019).

We then related $OP_{DTT,m}^{WS}$ to molecular-formula-derived metrics (AI_{mod} , O:C, N:C) across fresh and ozone-aged samples from both sources (Figure 4). Across the full dataset, aromatic-like formulas ($AI_{mod} \geq 0.5$) exhibited the strongest positive association with $OP_{DTT,m}^{WS}$ ($R^2 > 0.5$). The strongest correlations were observed for moderately oxygenated formulas (O:C $\approx$ 0.05–0.15), indicating that the DTT response likely tracks most closely with unsaturated/aromatic-like formulas and is consistent with the O:C range often associated with redox-active aromatic products, including quinone-like structures (Charrier and Anastasio, 2012). In contrast, highly oxidized formulas (e.g., acid-rich CHO species at higher O:C) and aliphatic-like formulas ($AI_{mod} = 0$) showed weaker associations with $OP_{DTT,m}^{WS}$ (Shultz et al., 2011; Park et al., 2006). These interpretations are based on molecular formulas and bulk metrics. Definitive assignment to specific compound classes (e.g., individual quinones) requires targeted identification and standards.

Nitrogen-containing aromatic-like formulas also exhibited positive associations with $OP_{DTT,m}^{WS}$, with the strongest relationships observed at N:C $\approx$ 0.10–0.25. Because the strongest correlations overall occurred for aromatic-like formulas with high AI_{mod} , this pattern suggests that nitrogen substitution within unsaturated/aromatic-like structures may mark an additional set of formulas associated with DTT activity at the molecular-formula level. The intrinsic DTT reactivity of individual nitrogen-containing organics remains less well constrained than that of quinone-type species, but prior work has linked

402 nitroaromatics and N-heterocycles to enhanced DTT responses in complex mixtures (Lai et al., 2025). Nitrogen functionalities
403 may also influence redox cycling indirectly, for example by modulating electron density or stabilizing redox intermediates in
404 mixed organic matrices (Dou et al., 2015). Consistent with these ESI+ resolved source fingerprints, Bakhoor-burning particles
405 were dominated by CHO formulas and only a minority met the aromatic-like criterion ($AI_{mod} \geq 0.5$; Figure S11), whereas a
406 substantially larger fraction of cigarette-derived CHN/CHON formulas were aromatic-like (Figure S11). Within the ESI+
407 molecular feature space, this pattern is consistent with the higher $OP_{DTT,m}^{WS}$ observed for cigarette-smoke particles and with the
408 positive association between $OP_{DTT,m}^{WS}$ and ESI+ detected nitrogen-containing aromatic-like formulas. However, these
409 correlations are interpreted as empirical structure–activity associations within the measured ESI+ molecular feature space,
410 rather than as quantitative apportionment of the full particle-phase organic mixture or as evidence for the absence of redox-
411 active species outside this analytical window.

412 Finally, since dissolved transition metals can contribute to DTT activity directly and/or via interactions with organics
413 (Jiang et al., 2019; Wang et al., 2018b), we conducted metal-chelation tests using Chelex resin. $OP_{DTT,m}^{WS}$ changed negligibly
414 after chelation (Figure S12), indicating a minimal contribution from dissolved metals and suggesting that $OP_{DTT,m}^{WS}$ in both
415 sources is dominated by organic species.

416 3.5 Cellular oxidative-stress responses induced by Bakhoor burning particles

417 To further assess cellular oxidative-stress responses to Bakhoor burning particles, we quantified intracellular fluorescence
418 signals in alveolar macrophages using a DCFH-based assay and reported responses as fold changes relative to untreated
419 controls. This cellular endpoint complements the acellular DTT assay by capturing integrated intracellular oxidant signaling
420 under particle exposure. As described in the Methods, five particle dose levels were tested for each sample, dose-response
421 relationships were fitted using the Hill equation, and the area under the fitted dose-response curve (AUC) was used as an
422 integrative metric to quantify the overall oxidative response (Figure 5A) (Tuet et al., 2016).

423 Bakhoor burning particles elicited stronger DCFH fluorescence responses than sidestream cigarette smoke particles
424 across the tested dose range (Figure 5B). The AUC metric ranged from 0.51 to 0.62 for Bakhoor samples, higher than 0.41 for
425 cigarette smoke. Naphthalene-derived SOA was used as an external cellular-response reference because it is a well-studied
426 anthropogenic aromatic SOA system that has been shown to induce strong macrophage DCFH responses under a comparable
427 assay framework (Tuet et al., 2017). The magnitude of the response induced by Bakhoor burning particles was comparable to
428 that reported for naphthalene-derived secondary organic aerosol. At higher particle doses, the fluorescence response
429 approached a plateau (Figure 5A), consistent with saturation of the assay response and/or the onset of cytotoxicity, as supported
430 by the accompanying MTT viability results (Figure S13). Because high particle doses can reduce cell viability, DCFH
431 fluorescence at the upper end of the dose-response curve may include contributions from cytotoxicity-associated processes.
432 We therefore used the MTT assay to define a non-cytotoxic low-dose range, with cell viability greater than 80% of the untreated
433 control as a conservative criterion. This threshold is stricter than the 70% viability cutoff commonly used in ISO 10993-5-
434 based cytotoxicity assessment (International Organization for Standardization, 2009). The three lowest loading levels, $0.0001\times$,

0.001×, and 0.01×, were selected for the low-dose AUC analysis. The DCFH AUC over this range was calculated using baseline-subtracted DCFH fold changes. Because this range included three dose points, trapezoidal integration over log10-transformed particle loading was used. For fresh total-particle samples, the low-dose AUC values for the three Bakhoor samples were 0.249, 0.241, and 0.215, compared with 0.140 for the protocol-matched sidestream cigarette-smoke reference. These results indicate that Bakhoor-burning particles induced measurable intracellular oxidative-stress responses under low-dose exposure conditions before substantial viability loss. The higher-dose responses were retained as part of the full dose-response characterization but were interpreted in light of the MTT viability results. MTT viability further showed a dose-dependent reduction in cellular metabolic activity at higher particle mass loadings (Figure S13). Cell viability was largely maintained at the lower loadings used for the low-dose DCFH AUC analysis, whereas higher loadings led to a marked reduction in viability. This result supports the interpretation that DCFH responses in the low-dose range occurred before substantial viability loss, whereas responses at higher loadings may be influenced by reduced cellular metabolic activity. Such combined interpretation of acellular oxidative potential, intracellular oxidative-stress response, and MTT-based cytotoxicity is consistent with prior PM toxicology studies using complementary chemical and cellular endpoints (Lionetto et al., 2021). An additional consideration for the macrophage assay is the potential contribution of endotoxin, because lipopolysaccharide can activate macrophage inflammatory and ROS-related responses (Sanlioglu et al., 2001; Maitra et al., 2009). Endotoxin-specific quantification, such as the Limulus amoebocyte lysate assay or recombinant Factor C analysis (Lee et al., 2004), was not included in this study. However, the cells were exposed to particles emitted from Bakhoor heated on smoldering charcoal, rather than to raw wood or botanical materials directly suspended in culture medium. Dry-heat treatment at 170–250 °C has been shown to reduce lipopolysaccharide activity (Tsuji and Harrison, 1978). In our setup, the charcoal temperature reached approximately 300 °C (Figure S6), and this high-temperature processing may have reduced potential contributions from raw-material-associated endotoxin. Residual particle-associated endotoxin nevertheless cannot be fully excluded without endotoxin-specific quantification. If such material remained associated with the emitted particles, it would be part of the particle mixture generated under the real-use Bakhoor-burning condition, rather than an external contaminant introduced during sample handling. The absence of endotoxin-specific quantification mainly limits detailed attribution of the macrophage response drivers, rather than the observation that Bakhoor-burning particles induced a DCFH response under the exposure protocol used here.

Notably, these cellular oxidative-stress responses differed from the acellular DTT results, for which cigarette smoke particles exhibited 47–100% higher $OP_{DTT,m}^{WS}$ than Bakhoor burning particles. Such divergence is expected because DTT and DCFH-based cellular readouts probe different chemical and biological processes and can be weakly correlated for complex mixtures (Kim et al., 2014; Tuet et al., 2016; Seo et al., 2025). Similar assay-dependent discrepancies have been reported previously; for example, Tuet et al. (2016) observed weak or absent correlations between DTT activity and DCFH-based cellular oxidative-stress responses for some wintertime ambient PM samples. Mechanistically, DTT primarily reflects the electron-transfer capacity of extractable species in a simplified buffer system, whereas the cellular assay integrates particle delivery and uptake, intracellular bioactivation, depletion of redox buffers, and activation of oxidant-generating pathways

469 (Kim et al., 2014). Collectively, these results underscore that DTT alone may not capture the full spectrum of oxidative-stress
470 potential for indoor combustion aerosols.

471 Although mechanistic attribution of cellular signaling to specific molecules is inherently challenging, the
472 compositional information provides plausible hypotheses for why Bakhoor burning particles produced stronger cellular
473 oxidative-stress responses. Bakhoor extracts were enriched in oxygenated CHO formulas (Figures S9 and S10), consistent
474 with a larger contribution from oxygenated organics, some of which may contain electrophilic carbonyl functionality.
475 Electrophilic aldehydes and related carbonyls are well known to form covalent adducts with cellular nucleophiles, particularly
476 cysteine thiols in glutathione and proteins, thereby depleting intracellular thiol buffering capacity and perturbing redox
477 homeostasis in ways that can amplify downstream oxidant signaling (Fritz and Petersen, 2013; Lopachin and Gavin, 2014).
478 AI_{mod} classification further places a substantial fraction of Bakhoor CHO formulas in the olefinic and non-aromatic unsaturated
479 regimes (Figure S11). These less-condensed CHO structures may be more conformationally flexible than condensed aromatic
480 frameworks and may interact differently with lipid phases, potentially affecting cellular delivery and subsequent bioactivation
481 (Liu et al., 2011). By contrast, cigarette smoke particles were enriched in nitrogen-containing formulas (CHN/CHON, Figure
482 S11) yet produced a lower net DCFH response under the exposure conditions used here, suggesting that the dominant chemical
483 drivers of $OP_{DTT,m}^{WS}$ and cellular oxidative-stress responses differ across sources.

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485 4. Atmospheric implications

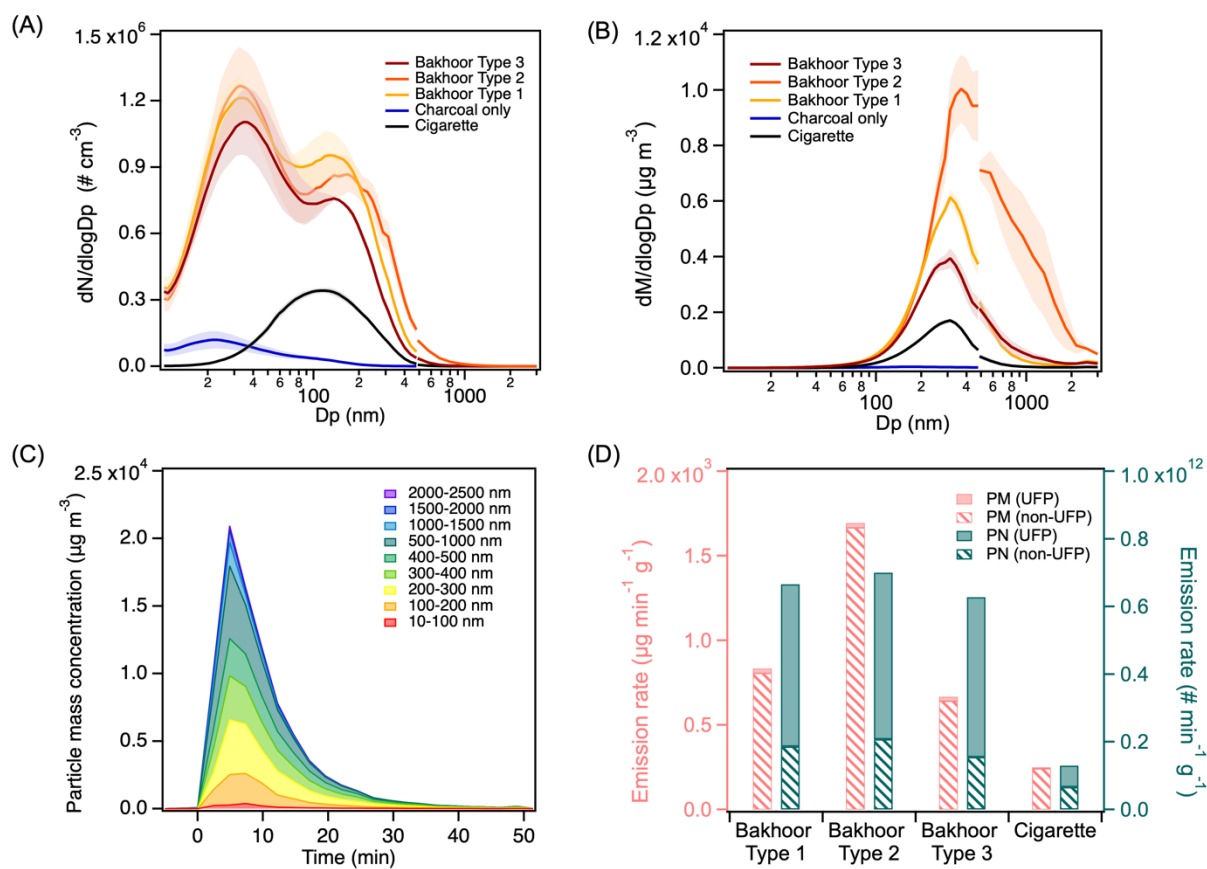
486 This study demonstrates that Bakhoor burning is a potent indoor source of fine particles, particularly with a substantial ultrafine
487 fraction. Across the three products tested here, Bakhoor burning produced particle mass and number emission rates of 670-
488 $1690 \mu\text{g min}^{-1} \text{g}^{-1}$ and $(6-7) \times 10^{11} \text{ particles min}^{-1} \text{g}^{-1}$, respectively, with ultrafine particles accounting for 70-75 % of total
489 particle number. On a mass-normalized basis, these emission rates were approximately 4.3-fold higher than those of the
490 protocol-matched smoldering sidestream cigarette-smoke reference. Together with the observed size distributions, these results
491 indicate that routine Bakhoor use can generate exposure-relevant particle burdens comparable to, and in some use scenarios
492 exceeding, those from more widely recognized indoor combustion sources. Because respiratory deposition varies strongly with
493 particle diameter and differs between the tracheobronchial (Tb) and alveolar (Al) regions (Figures 6A and 6B), lung-deposited
494 surface area (LDSA; see Text S2) was estimated as an inhalation-relevant dose metric that has been shown to track
495 inflammatory and cytotoxic responses across particle types more consistently than particle mass alone (Hofmann, 2011; Oh et
496 al., 2023). Using the mean emission factors measured here together with size-resolved regional deposition fractions from
497 Hofmann (2011), the Bakhoor-to-cigarette LDSA ratio exhibits strong size dependence and reaches its highest values in the
498 ultrafine range (Figure 6B). Because particles at intermediate diameters dominate the emitted size distribution and the
499 integrated surface-area dose (Figure S14), the LDSA ratio integrated over the measured size range is approximately 4 for both
500 the Tb and Al regions (Figure 6C). This LDSA calculation assumes spherical particles when converting measured size
501 distributions to particle surface area. It should be noted that nonspherical or agglomerated combustion particles may introduce
502 a morphology-dependent surface-area uncertainty (Decarlo et al., 2004; Ku and Maynard, 2005; Levin et al., 2016). For a

503 reported Bakhoor use rate of 0.4-2.9 g d⁻¹ (Bu-Olayan and Thomas, 2021), this scaling suggests that the daily inhalation-
504 relevant dose from Bakhoor burning may be comparable to that from multiple cigarettes of smoldering sidestream smoke under
505 the reference condition used here, although the real-world equivalence will vary with room volume, ventilation, and burning
506 practice.

507 Across the samples tested here, the combination of acellular DTT activity and macrophage DCFH responses shows
508 that Bakhoor burning particles exhibit substantial oxidative potential and cellular oxidative-stress responses. Relative to
509 sidestream cigarette smoke particles, these responses differ between assays on a per-mass basis and persist after ozone aging
510 over multi-day equivalent exposures. Fresh Bakhoor extracts showed an average $OP_{DTT,m}^{WS}$ of approximately 32 pmol min⁻¹ µg⁻¹,
511 ¹, ozone aging increased $OP_{DTT,m}^{WS}$ by about 15 % under the conditions used here, and the DCFH-based AUC values for Bakhoor
512 samples ranged from 0.51 to 0.62, compared with 0.41 for the sidestream cigarette-smoke reference. These results suggest that
513 the oxidative properties of Bakhoor-burning particles can remain relevant after post-emission aging, rather than being limited
514 to freshly emitted particles alone. This may be particularly important in indoor environments with limited ventilation and
515 repeated daily use, where particle accumulation and continued chemical processing by indoor oxidants and surface interactions
516 may extend the relevance of these responses beyond the point of emission. Such conditions are especially relevant in hot-arid
517 regions such as the Middle East, where indoor spaces are frequently cooled using air-conditioning systems and windows may
518 remain closed; in many settings with limited outdoor-air supply, air-exchange rates can be low (for example, <0.5 h⁻¹)
519 (Indraganti et al., 2016; Diapouli et al., 2013). The accompanying reduction in cell viability further suggests that these oxidative
520 responses may be associated with measurable biological effects under the in vitro exposure conditions tested here. Relating
521 these in vitro exposure conditions to human lung dose, however, will require dedicated dosimetry and exposure modeling and
522 should be addressed in future work.

523 The control and perturbation experiments further identify the hexane-soluble fraction (HSF) as an important leverage
524 point for particle formation during Bakhoor burning. Operational treatments that reduced the HSF, including solvent extraction
525 to generate wood residue and a simple tissue-wiping step to partially remove surface-associated HSF, lowered particle mass
526 emissions. However, these perturbations yielded slightly higher particle mass-normalized DTT oxidative potential relative to
527 untreated material (Figure S15), indicating that emission reductions do not necessarily lead to a reduction in oxidative potential
528 on a mass basis. Nevertheless, when mass-normalized DTT oxidative potential is combined with the corresponding emission
529 factors to estimate emission-weighted total DTT consumption, the net oxidative burden decreases for the HSF-reduced
530 treatments (Figure S15). These results indicate that the HSF strongly influences both emission magnitude and oxidative
531 properties. Because HSF loading varies across products, lowering HSF loading may provide a practical pathway to reduce
532 emitted particle mass and, consequently, exposure, although hazard-relevant endpoints should be evaluated in parallel rather
533 than inferred from emissions alone. At the user level, wiping Bakhoor pieces prior to burning may partially remove surface-
534 associated HSF and reduce emissions, although the extent of emission reduction will likely depend on product composition
535 and user practice and should be evaluated under realistic operating conditions. Electric Bakhoor burners may also represent a
536 useful alternative by eliminating charcoal-assisted combustion and potentially reducing combustion-related particle formation,

537 but their net effects on emissions and toxicity require evaluation under realistic use conditions before broad recommendations
538 can be made. More broadly, these results identify source composition and burning configuration as important controls on
539 particle emissions from Bakhoor burning, while highlighting the additional role of ventilation in shaping real-use exposure.
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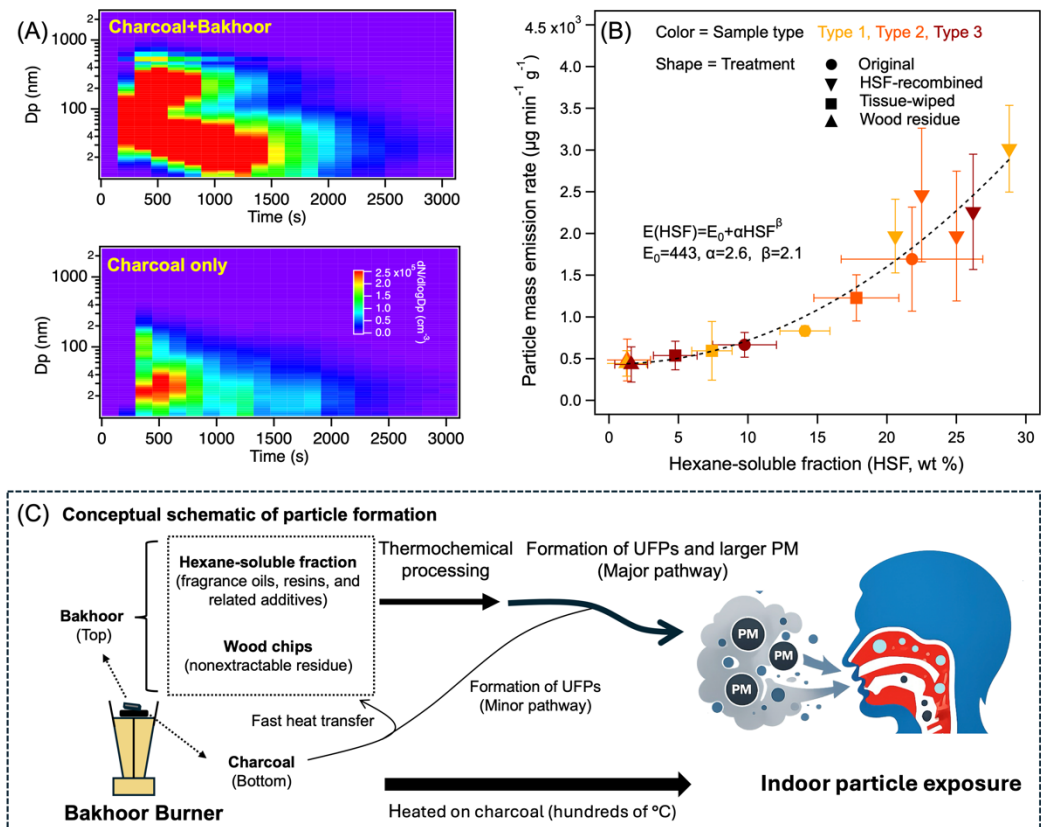
542 **Figure 1.** (A) Number and (B) mass size distributions measured in the chamber during Bakhoor burning on charcoal (three
543 products) and smoldering sidestream cigarette smoke, with a charcoal-only control; distributions are shown without size-
544 dependent loss correction. (C) Average particle mass concentration as a function of time following Bakhoor burning, resolved
545 by particle size ranges and shown without loss correction. (D) Particle emission rates, with mass (left axis) and number (right
546 axis) normalized to the initial Bakhoor mass per burn and to the mass of cigarette tobacco filler, respectively, reported
547 separately for ultrafine and non-ultrafine size fractions and corrected for size-dependent losses. The apparent discontinuity
548 near 500 nm in panels A and B reflects the use of two instruments: particles below 500 nm were measured by the SEMS,
549 whereas particles above 500 nm were measured by the OPC. This discontinuity is therefore not interpreted as a physical feature
550 of the emitted aerosol size distribution.

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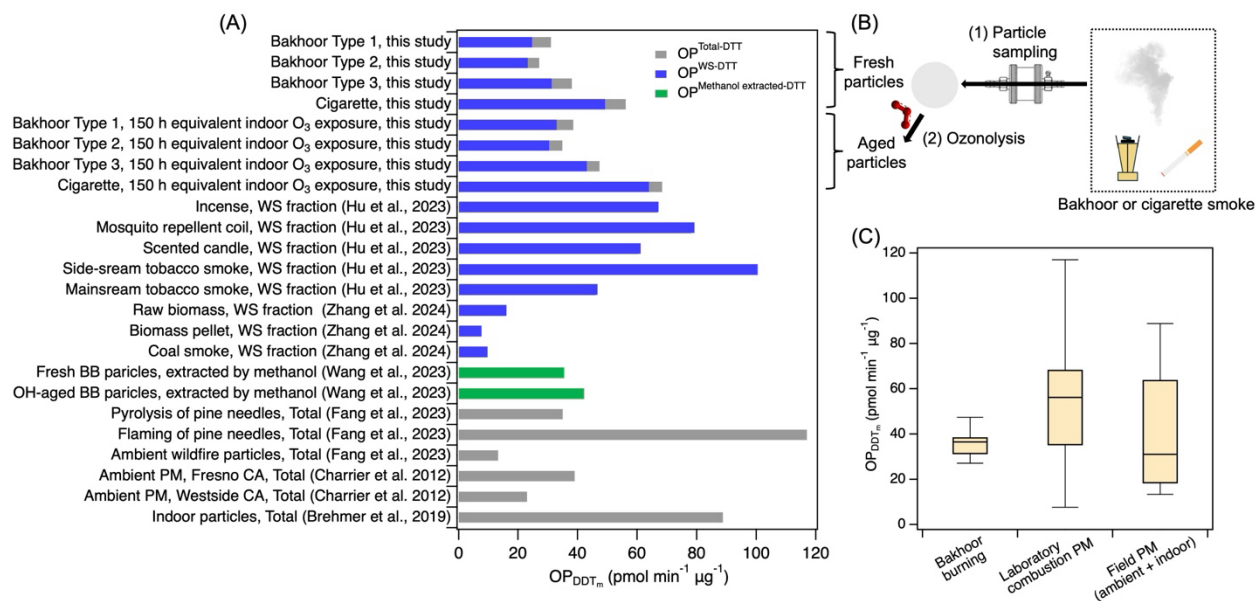
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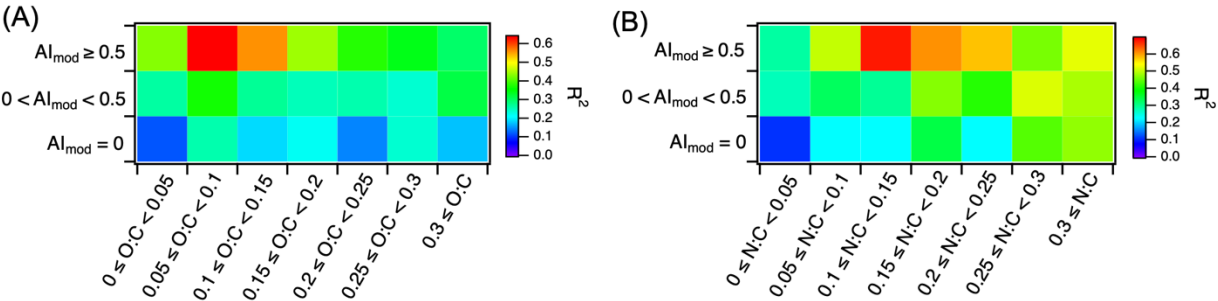
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556 **Figure 2.** (A) Time-resolved particle number size distributions ($dN/d\log D_p$) measured during Bakhoor burning (with charcoal
 557 used for ignition) and during a charcoal-only control, shown as contour maps of mobility diameter (D_p) versus time. (B) Particle
 558 mass emission rate as a function of the loading of the hexane-soluble fraction (HSF, wt %) in Bakhoor materials. Colors denote
 559 Bakhoor type, and symbols denote treatment (original, HSF-recombined, tissue-wiped, and wood residue). The dashed curve
 560 shows an empirical fit, $E(\text{HSF}) = E_0 + \alpha \text{HSF}^\beta$ (parameters shown). (C) Conceptual schematic linking Bakhoor composition
 561 and heating on charcoal to fine particle formation and indoor particle exposure.



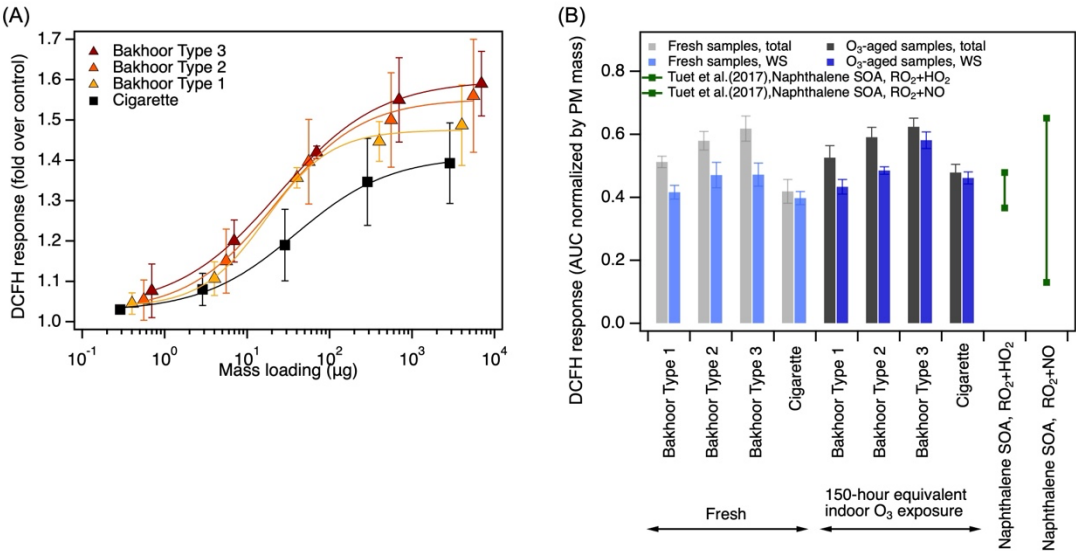
563 **Figure 3.** (A) The mass-normalized DTT consumption rate (OP_{DTTm}) for Bakhoor burning and sidestream cigarette smoke
564 particles in this study (fresh and ozone-aged), shown alongside literature values for other laboratory combustion aerosols and
565 field PM. (B) Schematic of particle collection on filters and subsequent ozonolysis of PM-loaded filters prior to DTT analysis.
566 (C) Box plot summarizing OP_{DTTm} grouped as Bakhoor burning, other laboratory combustion PM and field PM.

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577 **Figure 4.** Heat maps of the coefficient of determination (R^2) from linear regressions relating the water-soluble mass-
578 normalized DTT oxidative potential ($OP_{DTT,m}^{WS}$) to the summed UHPLC–HRMS signal intensity of molecular-formula bins
579 grouped by modified aromaticity index (AI_{mod}) and (A) oxygen-to-carbon (O:C) or (B) nitrogen-to-carbon (N:C) ratio, across
580 fresh and ozone-aged Bakhoor burning and cigarette smoke samples.

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592 **Figure 5.** Intracellular oxidative-stress responses (DCFH assay) induced by particles from three Bakhoor types and sidestream
593 cigarette smoke. (A) Dose–response curves of DCFH fluorescence (reported as fold change relative to untreated controls) as
594 a function of particle mass loading; solid lines denote Hill-equation fits. (B) Integrated responses quantified as the area under
595 the fitted dose–response curve (AUC), normalized by particle mass, for fresh and O_3 -aged samples and for total particle
596 suspensions and water-soluble (WS) extracts. Reference values for naphthalene SOA are from Tuet et al. (2017)

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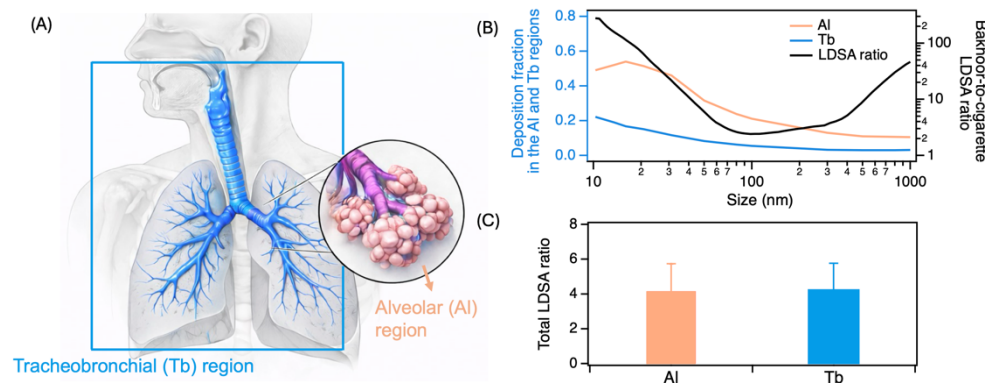
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608 **Figure 6.** (A) Schematic of the human respiratory tract highlighting the tracheobronchial (Tb) and alveolar (Al) regions used
609 for deposition calculations. (B) Size-resolved regional deposition fractions for the Tb and Al regions (left axis) and the
610 corresponding size-resolved ratio of lung-deposited surface area (LDSA) for Bakhoor burning particles relative to sidestream
611 cigarette smoke particles (right axis). (C) Bakhoor-to-cigarette LDSA ratio integrated over the measured particle size range
612 for the Tb and Al regions; error bars denote ± 1 standard deviation (1σ).

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619 **Data availability.** All datasets used in this study, as well as the data products generated during the analysis, are available from
 620 the corresponding author upon request. These data are not publicly archived because they are maintained as multiple processed
 621 and analysis-specific files rather than as a repository-ready dataset.

622 **Author contributions.** CKC and LZ designed the experiment; LZ and ZL conducted the experiments; LZ, ZL and WX
 623 performed the data interpretation; LZ, ZL, WX, RJH and CKC wrote the paper. All authors contributed to the paper with useful
 624 scientific discussions or comments.

625 **Competing interests.** The contact author has declared that neither they nor their co-authors have any competing interests.

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