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1 **Chemical Characterization and Source Apportionment of Carbonaceous**
2 **Aerosols during Post-Monsoon Biomass Burning and Diwali at an**
3 **Upwind Site of Delhi**

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10 **Abstract**

11 Sonipat, located ~40 km northwest of Delhi, lies along the principal transport corridor linking
12 post-monsoon agricultural burning regions of Punjab–Haryana with Delhi and serves as an
13 intermediate receptor for regional pollution. We conducted intensive high time resolution
14 measurements of composition-based PM_{2.5} (non-refractory PM_{2.5} plus black carbon) from 25
15 October to 15 November 2023 using a ToF-ACSM and an Aethalometer to characterize
16 carbonaceous aerosol sources during the biomass-burning period. Two severe haze episodes
17 occurred, with PM_{2.5} exceeding 300 µg/m³. Organic aerosol dominated the submicron mass
18 (~65% of non-refractory PM_{2.5}), with daily mean concentrations peaking near 140 µg/m³ during
19 the first haze episode. Positive Matrix Factorization resolved five components including
20 hydrocarbon-like, biomass-burning, and solid-fuel combustion organic aerosol, and two
21 oxygenated fractions representing semi-volatile and low-volatility aged aerosol. Secondary
22 organic aerosol accounted for ~57-60% of organic aerosol mass, with low-volatility oxygenated
23 organic aerosol reaching 42.8 µg/m³ during peak haze, indicating substantial regional aging and
24 accumulation. Biomass-derived black carbon contributed ~78% of total black carbon (mean 10.9
25 µg/m³), far exceeding fossil-fuel contributions (~3.1 µg/m³). Trajectory and wind analyses
26 consistently identified northwestern agricultural regions as dominant sources with minor traffic
27 influence, indicating that extreme carbonaceous aerosol over Delhi-NCR largely forms outside
28 the urban core through regional transport biomass and solid-fuel combustion emissions
29 combined with sustained secondary processing, highlighting the need for coordinated airshed-
30 scale emission reductions across the Indo-Gangetic Plain.



31 **1. Introduction**

32 The Indo-Gangetic Plain (IGP) is widely recognized as one of the most polluted regions globally in
33 terms of fine particulate matter (PM_{2.5}) loading and its associated atmospheric impacts (Dey and
34 Di Girolamo, 2010; Ganguly et al., 2009; Ghude et al., 2016). This region is densely populated and
35 a rapidly developing airshed is influenced by a complex mixture of emission sources, including
36 vehicular traffic, industrial activities, residential biofuel consumption, and seasonal agricultural
37 residue burning (Biswal et al., 2023; Ganguly et al., 2009). Within the IGP, the Delhi-National
38 Capital Region (Delhi-NCR) frequently experiences severe haze episodes, particularly during the
39 post-monsoon and winter seasons (Gani et al., 2019), when PM_{2.5} concentrations can exceed
40 World Health Organization (WHO) guidelines by an order of magnitude or more (Rathore et al.,
41 2025). These extreme pollution episodes have profound public health consequences. Air
42 pollution in India has been linked to 1.67 million premature deaths in 2019, with approximately
43 1 million deaths attributed specifically to ambient PM_{2.5} exposure (Pandey et al., 2021). Given its
44 significance in critical scientific and policy priority, it is necessary to understand the sources and
45 processes that drive these episodes over the study region.

46 Post-monsoon haze events over northern India are strongly influenced by extensive
47 agricultural residue burning in the northwestern IGP, particularly across Punjab and Haryana, and
48 are further reinforced by unfavorable meteorological conditions (Sarkar et al., 2018). Cold and
49 dry northwesterly winds coincide with the post-harvest agricultural cycle, transporting biomass-
50 burning emissions toward Delhi-NCR. At the same time, shallow planetary boundary layers
51 (PBLs), weak ventilation, and stagnant conditions limit dispersion and enhance pollutant
52 accumulation near the surface (Nair et al., 2007). Although the dominant drivers of these
53 episodes are broadly understood, robust quantification of individual source contributions and
54 their evolution during regional transport remains limited. This limitation is particularly
55 pronounced at upwind locations outside the urban core, where transported air mass begins to
56 interact with local emissions and undergoes complex chemical processing before reaching
57 megacity environments (Lalchandani et al., 2022).



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58 Reliable source apportionment is essential for designing effective mitigation strategies.
59 Policy frameworks such as the Graded Response Action Plan (GRAP) depend on accurate
60 identification of dominant sources, yet separating regional background pollution from episodic
61 enhancements associated with biomass burning and festival-related emissions remains
62 challenging (Ghude et al., 2024; Pant et al., 2016; Yadav et al., 2022). During the post-monsoon
63 period, intense agricultural burning and Diwali firework emissions can both elevate PM_{2.5}
64 concentrations (Satish et al., 2017; Thamban et al., 2019). Their chemical signatures often overlap
65 with those from domestic cooking and heating, complicating attribution (Tobler et al., 2020).
66 Distinguishing these contributions is critical because the relative importance of primary versus
67 secondary sources has direct implications for control strategies. For example, if secondary
68 aerosol formation dominates, mitigation efforts targeting only primary emissions may be
69 insufficient to significantly reduce peak pollution levels.

70 Most of the previous studies on source apportionment in the region have relied on filter-
71 based receptor modelling, which typically provides 24-hour averaged chemical composition (Jain
72 et al., 2019; Sharma et al., 2016). While such approaches yield valuable information on bulk
73 aerosol composition, they lack the temporal resolution required to capture rapid chemical
74 transformations and short-lived pollution peaks. Intense but transient emissions from firework
75 displays, moving fire plumes, or episodic industrial activities can be substantially smoothed in
76 daily averages. Moreover, measurements conducted within dense urban environments are often
77 dominated by local traffic and industrial emissions, which can mask the signatures of regionally
78 transported aerosols (Bhandari et al., 2020; Guttikunda and Calori, 2013; Singh et al., 2021).
79 These limitations highlight the need for high-time-resolution measurements at strategically
80 located receptor sites to isolate regional contributions.

81 Advances in online aerosol instrumentation provide new opportunities to overcome
82 these challenges. Instruments such as the Time-of-Flight Aerosol Chemical Speciation Monitor
83 (ToF-ACSM) enable continuous, high-time-resolution measurements of non-refractory aerosol
84 species and can resolve rapid changes in chemical composition associated with evolving
85 atmospheric processes (Fröhlich et al., 2013; Ng et al., 2011). When combined with Positive



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86 Matrix Factorization (PMF), these datasets can be decomposed into sources related to organic
87 aerosol factors without requiring a priori source profiles (Canonaco et al., 2013). This approach
88 has been widely applied across complex source regions within the IGP to differentiate fresh
89 biomass burning organic aerosol (BBOA) from more aged, oxygenated organic aerosol (OOA),
90 thereby providing deeper insight into aerosol evolution beyond bulk chemical characterization
91 (Lakra et al., 2024; Shukla et al., 2025). Recent studies have identified solid-fuel combustion
92 organic aerosol (SFCOA) and oxygenated organic aerosol (OOA) as major contributors to organic
93 mass in Delhi (Lalchandani et al., 2022) and Faridabad (Tobler et al., 2020). However, most such
94 investigations have focused on urban receptor sites, leaving the transformation and mixing
95 processes of carbonaceous aerosols during regional transport poorly constrained .

96 In this context, Sonipat, located northwest of Delhi along the dominant pollution
97 transport corridor from Punjab and Haryana, represents a strategically important upwind
98 receptor site. Air masses arriving at this location intercept regional biomass-burning plumes
99 before they are substantially mixed with strong local emissions from the Delhi metropolitan area
100 (Singh et al., 2023). Recent chemical characterization at this site during the post-monsoon period
101 indicated that carbonaceous aerosols accounted for nearly 80% of composition-based $PM_{2.5}$
102 mass, with relatively minor contributions from secondary inorganic species (Rathore et al., 2025).
103 This composition differs from haze episodes in other regions such as the North China Plain, where
104 secondary sulfate and nitrate often dominate. Analysis of satellite-derived vegetation indices and
105 fire counts data reveals increasing agricultural burning activity in source regions upwind of
106 Sonipat (Jethva et al., 2019), supporting the observations of enhanced aged organic aerosol at
107 this site (Rathore et al., 2025). Although earlier work characterized bulk aerosol chemistry and
108 highlighted the importance of biomass burning emissions, substantial variability during the Diwali
109 festival and post-rainfall periods suggests the presence of multiple temporally varying sources
110 that cannot be resolved through bulk analysis alone. Periods of relatively clean air following
111 rainfall, followed by rapid pollution buildup during Diwali, provide a natural experiment for
112 examining source dynamics under changing meteorological and emission conditions.



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113 Based on the research literature, there are several questions yet to be addressed for
114 better understanding the role of the source contributions and evaluating the effectiveness of
115 emission control strategies. I have tried to answer some of those questions in this manuscript as
116 follow:

117 1) What are the relative contributions of transported biomass burning emissions and local
118 sources such as traffic and industry emissions during peak haze conditions at this upwind
119 location?

120 2) To what extent does primary organic aerosol dominate over secondary formation under
121 conditions of weak photochemistry and limited inorganic precursor availability?

122 Addressing these questions is essential for evaluating the effectiveness of emission
123 control strategies. A detailed, high-time-resolution source apportionment is therefore required
124 to disentangle the contributions of individual sources and to assess the balance between primary
125 and secondary carbonaceous aerosols during regional transport into Delhi.

126 In this study, we present a comprehensive source apportionment of PM_{2.5} at the upwind
127 Sonipat site during the post-monsoon biomass-burning and Diwali period, building upon our
128 earlier chemical characterization study (Rathore et al., 2025). Using high-time-resolution ToF-
129 ACSM and Aethalometer measurements in conjunction with receptor modeling and trajectory
130 analysis, we resolve carbonaceous aerosol into source-specific components and examine their
131 temporal variability in relation to meteorology and independent tracers. The primary objectives
132 are to (i) identify and quantify major organic aerosol sources during haze and non-haze
133 conditions, (ii) assess the relative contributions of transported biomass burning emissions, local
134 traffic, industrial sources, and secondary formation, and (iii) contrast source dynamics between
135 prolonged biomass-burning episodes and the transient Diwali event. This work provides one of
136 the first high-resolution PM_{2.5} source apportionment analyses at a strategically located upwind
137 gateway to Delhi and offers new constraints on regional transport processes and source
138 contributions relevant for air quality management across the Delhi-NCR and the broader Indo-
139 Gangetic Plain.

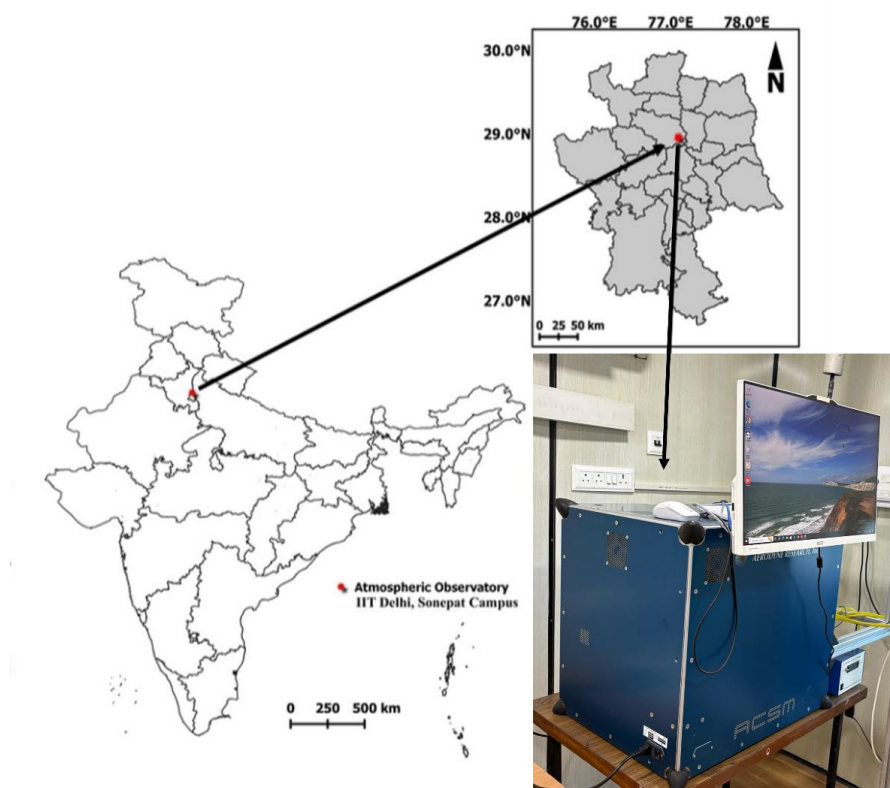


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140 2. Materials and Methods

141 2.1. Study Site Description

142 Measurements for this study were conducted at the Center for Atmospheric Sciences–
143 Atmospheric Observatory (CAS-AO) located on the IIT Delhi Sonipat campus (28.95°N, 77.10°E;
144 228 m above mean sea level). The observatory is situated approximately 14 km upwind of the
145 Delhi border in the state of Haryana and lies within the National Capital Region (NCR). The site is
146 positioned in a semi-urban environment within the Rajiv Gandhi Education City and is surrounded
147 by a heterogeneous land-use matrix comprising academic institutions, agricultural fields, small-
148 scale industrial activities, and major national highways. This setting makes the site well suited for
149 investigating the combined influence of regional and local emissions. The geographic location of
150 the observational site, and an image of the installed ToF-ACSM are presented in Figure1.



151



152 **Figure 1:** Location map of the study site (marked by a star symbol) and a photograph of the
153 installed ToF ACSM.

154 The regional climate is characterized by hot summers, a humid monsoon season, and cold
155 winters, broadly similar to prevailing meteorological conditions over Delhi. During the pre-
156 monsoon (March–May) and monsoon (June–September) seasons, the site is frequently affected
157 by dust transport from Thar Desert. In contrast, the post monsoon (October–November) and
158 winter (December–February) seasons are characterized by frequent haze and smog episodes that
159 substantially degrade visibility and air quality. During the post-monsoon period, northwesterly
160 winds often transport polluted air masses from the agricultural regions of Punjab and Haryana,
161 where large-scale crop-residue burning occurs(). In addition to these regional influences, the site
162 is also affected by local vehicular emissions, nearby industrial activities, and biomass-burning
163 sources. Owing to its proximity to Delhi and similarity in meteorological conditions, the site
164 provides a representative location for examining aerosol processes relevant to the Delhi-NCR
165 region. Additional details regarding the site characteristics, measurement, and long-term
166 observations are provided in Rathore et al. (2025).

167 **2.2. Instrumentation and Measurements**

168 This study utilizes measurements conducted between 25 October 2023 to 16 November
169 2023 at the CAS-AO site using multiple state-of-the-art instruments. Non-refractory submicron
170 particulate matter (NR-PM_{2.5}) was measured using a Time-of-Flight Aerosol Chemical Speciation
171 Monitor (ToF-ACSM; Aerodyne Research, Inc., Billerica, Massachusetts, USA) equipped with a
172 PM_{2.5} aerodynamic lens and a capture vaporizer. The instrument was housed in a temperature-
173 controlled room on the top floor(18m above ground) of I-TECH building at the observatory to
174 ensure stable operating conditions. The ToF-ACSM measures the mass concentration of non-
175 refractory aerosol species that flash vaporize at temperature above ~600 °C. Vaporized species
176 are ionized using 70eV electron-impact ionization, and the resulting ions are analyzed by a time-
177 of-flight mass spectrometer (Fröhlich et al., 2013; Nault et al., 2023). The inlet system was
178 specifically set up to enable the ToF-ACSM and the SMPS to run alongside each other at the same
179 time, sharing a common sampling line. The instrument is equipped with a 3.05m copper tube and



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180 a PM_{2.5} cyclone at beginning of the inlet ensuring only fine particles (smaller than 2.5 μm) are
181 analyzed from outside the window. The sampling line also has water trap and a Nafion membrane
182 diffusion dryer(Perma Pure MD) to remove moisture from the aerosol sample. The system
183 operates at a total flow rate of 2.5 L min⁻¹. This flow is split between the connected instruments-
184 the SMPS draws 1 L min⁻¹, the ToF-ACSM pulls approximately 0.1 L min⁻¹, and the leftover 1.4 L
185 min⁻¹ is handled by an in-line flow controller to maintain the overall balance. During the entire
186 sampling period, the vaporizer was maintained at a temperature below 600°C(approximately
187 525–530°C). For the data analysis, a composition-dependent collection efficiency (CDCE) value of
188 1 was applied to the capture vaporizer system, in line with the approach recommended by (Hu
189 et al., 2018). The instrument provides high-time-resolution measurements of organic aerosol,
190 sulfate, nitrate, ammonium, and chloride with a 40-s time resolution, with 20 seconds of each
191 cycle dedicated to background or filter sampling. Standard operating procedures, including
192 calibrations, air-beam corrections, application of fragmentation tables, and uncertainty
193 estimation, were performed using the TOFWARE 4.4.1 data analysis package prior to further
194 processing.

195 Black carbon (BC) mass concentrations were measured using a seven-wavelength
196 Aethalometer (model AE31; Magee Scientific, Berkeley, California, USA), which provides
197 wavelength-dependent aerosol light absorption coefficients. The AE31 Aethalometer had its
198 own dedicated sampling arrangement, independent of the ToF-ACSM inlet AE31 sampled the
199 ambient air at the rate of 2 L min⁻¹ through a 2.5m long conductive tube, with PM_{2.5} cyclone.
200 The vertical structure of the lower atmosphere, including atmospheric boundary layer (ABL)
201 height and depolarization ratios, was monitored using a Vaisala CL61 ceilometer.
202 Meteorological parameters, including temperature, relative humidity, wind speed, and wind
203 direction, were obtained from an automatic weather station (AWS) installed at the site.
204 Detailed descriptions of all the instrument configuration, inlet setup, calibration protocols, and
205 data acquisition procedures are discussed in Rathore et al. (2025).

206 **2.3. Source-apportionment Techniques**



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207 Receptor-based source apportionment techniques were applied to identify and quantify
208 major sources contributing to submicron aerosol mass at the study site. These methods resolve
209 the observed aerosol composition into a limited number of source-related factors based on
210 characteristic temporal and chemical signatures.

211 2.3.1. Positive Matrix Factorization (PMF)

212 Positive Matrix Factorization (PMF) was applied to the organic aerosol mass spectral
213 dataset to apportion sources of submicron particulate matter. PMF is a multivariate receptor-
214 modeling technique that decomposes the observed data matrix \mathbf{X} into a factor contribution
215 matrix \mathbf{G} and a factor profile matrix \mathbf{F} , along with a residual matrix \mathbf{E} , such that:

216

$$217 \quad \mathbf{X} = \mathbf{G} \cdot \mathbf{F} + \mathbf{E}$$

218 The optimal solution is obtained by minimizing the objective function:

$$219 \quad Q = \sum_{i=1}^n \sum_{j=1}^m \left(\frac{e_{ij}}{\sigma_{ij}} \right)^2$$

220 where e_{ij} represents residuals and σ_{ij} denotes measurement uncertainties. This weighting scheme
221 reduces the influence of variables with higher uncertainty on the solution (Paatero and Tapper,
222 1994). PMF has been widely used in aerosol mass spectrometry studies and is particularly well
223 suited for resolving organic aerosol sources from high-time-resolution datasets (Ulbrich et al.,
224 2009).

225 PMF analysis was performed using the Source Finder (SoFi, version 8) interface
226 (Datalystica Ltd.) implemented in Igor Pro. SoFi provides an interactive framework that enables
227 evaluation of rotational ambiguity through FPEAK exploration, as well as displacement and
228 bootstrap analyses to assess solution stability. The tool also allows direct comparison of factor
229 profiles with reference spectra from previous studies (Canonaco et al., 2021, 2013). The input
230 data consisted of organic mass spectra from the TOF-ACSM, preprocessed using standard



231 TOFWARE routines, including air beam correction, application of fragmentation tables, and
232 uncertainty estimation. Factor identification was guided by time series behavior, diurnal
233 patterns, correlations with external tracers, and comparison with established reference spectra.

234 **2.3.2 Aethalometer-Based Source Apportionment of Black Carbon**

235 The multi-wavelength AE31 Aethalometer data were used to apportion black carbon sources
236 using the Aethalometer Model proposed by Sandradewi et al. (2008). This model exploits the
237 wavelength dependence of aerosol light absorption to differentiate between fossil-fuel
238 combustion and biomass or wood combustion. Absorption coefficients at 470 nm and 950 nm
239 were used, as this wavelength pair provides strong contrast between traffic-related and biomass-
240 burning aerosols (Sandradewi et al., 2008; Zotter et al., 2017).

$$241 \quad b_{abs}(\lambda) = b_{abs,ff}(\lambda) + b_{abs,bb}(\lambda)$$

242 It is assumed that the total absorption coefficient at a certain wavelength is the sum of
243 contributions from fossil fuel (ff) and wood burning (bb) sources. The spectral dependence of
244 absorption follows a power-law relationship:

$$245 \quad b_{abs}(\lambda) \propto \lambda^{-\alpha}$$

246 where α is the Ångström exponent for absorption. We adopted fixed α values of 1.0 for
247 fossil fuel combustion and 2.0 for biomass burning (Figure S9), consistent with values reported
248 in previous studies conducted in Delhi-NCR and other urban regions of northern India (Tobler et
249 al., 2020). By solving the coupled equations described in Sandradewi et al. (2008), fractional
250 contributions of biomass-burning black carbon (eBC_{bb}) and fossil-fuel black carbon (eBC_{ff}) were
251 determined. Data screening and quality control followed procedures outlined by Segura et al.
252 (2014). The analysis methodology followed the approaches outlined by Harrison et al. (2013);
253 Kumar et al. (2020) and Titos et al. (2017). Following Petzold et al. 2013, the term equivalent
254 black carbon (eBC) is used to denote absorption-derived BC.

255 This method has been widely applied in Delhi and across the IGP to quantify contributions
256 from vehicular emissions and biomass combustion, particularly during winter and the post-



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257 monsoon periods (Bikkina et al., 2019; Dumka et al., 2018). Despite inherent assumptions
258 regarding fixed spectral exponents, the method provides a robust and widely accepted
259 framework for source attribution of black carbon in complex urban and semi-urban
260 environments.

261 **2.4. Trajectory and Wind Based Source Analysis**

262 To identify potential source regions influencing the receptor site, concentration-weighted
263 trajectory (CWT) analysis was conducted using the ZeFir toolkit by Petit et al. (2017). Five-day air
264 mass back trajectories were computed at hourly time resolution using the Hybrid Single-Particle
265 Lagrangian Integrated Trajectory (Hysplit v4.1) model (Draxler et al., 2018; Stein et al., 2015),
266 driven by Global Data Assimilation System (GDAS) meteorological fields at $1^\circ \times 1^\circ$ a spatial
267 resolution. These trajectories combined with time series of organic aerosol factors to compute
268 CWT fields according to:

269

$$270 \quad CWT_{ij} = \frac{1}{\tau_{ij}} \sum_{k=1}^N C_k \tau_{ijk}$$

271

272 where, C_k is the measured concentration associated with trajectory k , and τ_{ijk} is the
273 residence time of trajectory k in the grid cell i, j . This approach identifies geographic regions
274 associated with elevated concentrations at the receptor site. Organic aerosol factor time series
275 were averaged to 3-hour intervals to match trajectory resolution.

276 To examine the relationship between pollutant concentrations and local meteorology,
277 Non-Parametric Wind Regression (NWR) analysis was performed following (Henry et al., 2009).
278 NWR applies Gaussian kernel smoothing to estimate pollutant concentrations as a continuous
279 function of wind speed and direction, enabling separation of local and transported sources.
280 Pollution-rose analyses were generated by combining wind direction frequency with
281 concentration data to identify dominant source sectors. These complementary methods provide



282 a comprehensive assessment of local and regional source influences on observed aerosol
283 composition.

284 **3. Results and Discussions**

285 For consistency with our earlier chemical characterization study (Rathore et al., 2025), we
286 partition the measurement period into four phases representing distinct emission–meteorology
287 regimes: a pre-haze non-haze period (NH1; 25 October–1 November), a prolonged biomass-
288 burning–dominated haze episode (H1; 1–10 November), a rain-influenced non-haze period (NH2;
289 10–12 November), and a second haze episode coincident with Diwali emissions (H2; 12–15
290 November). Rathore et al. (2025) demonstrated that bulk $PM_{2.5}$ during these phases is dominated
291 by carbonaceous aerosols at Sonipat, indicating that this upwind site frequently experiences
292 severe particulate loading even before air masses reach the Delhi urban core. Here, we extend
293 that framework by resolving source-specific contributions to carbonaceous PM (organic aerosol
294 and black carbon) and by examining how primary emissions, secondary processing, and
295 meteorology jointly control the timing, magnitude, and composition of haze and non-haze
296 conditions along the northwest transport corridor into Delhi-NCR. This approach is particularly
297 relevant for mitigation because it distinguishes periods dominated by direct combustion
298 emissions from periods where a large fraction of the mass resides in aged, oxygenated material,
299 implying different leverage points for control strategies.

300 **3.1. Identification and Interpretation of OA Factors**

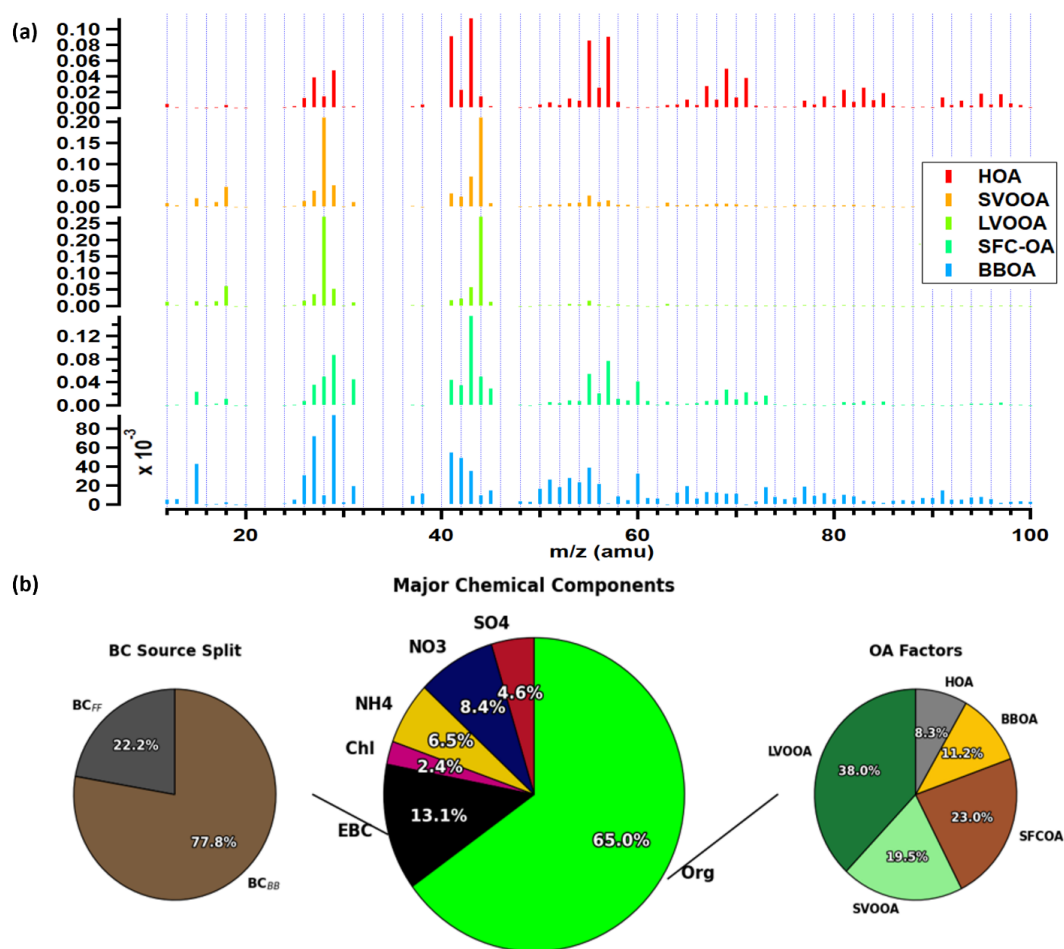
301 The PMF analysis applied to the ToF-ACSM organic mass spectra resolved five distinct
302 factors: hydrocarbon-like organic aerosol (HOA), biomass-burning organic aerosol (BBOA), solid-
303 fuel combustion organic aerosol (SFCOA), semi-volatile oxygenated organic aerosol (SVOOA), and
304 low-volatile oxygenated organic aerosol (LVOOA), providing a physically interpretable separation
305 between primary organic aerosol (POA) and secondary organic aerosol (SOA) (Figure 2a). The
306 corresponding factor mass spectra and elemental ratios are shown in Figure 2a and summarized
307 in Table S1 (Supporting Information), while bulk C- $PM_{2.5}$ (composition-based $PM_{2.5} = NR-PM_{2.5} +$
308 BC) and OA-factor contributions are shown in Figure 2b. The emergence of two oxygenated



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309 factors alongside multiple combustion-related factors is consistent with a receptor site that
 310 intercepts both fresh regional plumes and their progressively aged products. This structure is
 311 mechanistically meaningful for the IGP post-monsoon period because haze evolution often
 312 reflects not a single source type, but the superposition of transported combustion emissions and
 313 subsequent oxidation/partitioning under stagnant boundary-layer conditions.

314



315 **Figure 2:** a) Mass spectra of the OA factors (HOA, SVOOA, LVOOA, SFCOA and BBOA) resolved
 316 by PMF and, b) Species contribution to C-PM_{2.5} (NH₄⁺, SO₄²⁻, NO₃⁻, Org, Cl⁻ and BC) (center pie



317 *chart), Relative contribution of OA factors to total OA (right pie chart) and Relative contribution*
318 *of BC factors(left pie chart).*

319 **3.1.1. Primary OA Factors (HOA, BBOA, SFCOA)**

320 HOA is characterized by prominent aliphatic hydrocarbon fragments at m/z 41, 43, 55,
321 and 57 and low oxygenation ($O:C = 0.10$; Table S1 in Supporting Information), consistent with
322 fresh traffic-related emissions. Although the five-factor solution was obtained using an
323 unconstrained PMF framework, we constrained the HOA profile to reduce the mixing between
324 HOA and SFCOA, which can be substantial in environments influenced by both traffic and solid-
325 fuel combustion (Tobler et al., 2020). Specifically, we used the HOA reference profile (Canonaco
326 et al., 2013; Crippa et al., 2013) with a randomized α -value between 0 and 0.5 to allow limited
327 profile flexibility while stabilizing the HOA/SFCOA separation. This choice is important at Sonipat
328 because traffic emissions are present but do not dominate the regional plume, and without this
329 constraint, PMF can allocate part of the combustion-related hydrocarbon signal to HOA,
330 artificially inflating traffic contributions and obscuring the relative importance of solid-fuel and
331 biomass burning sources that are more actionable from an airshed-management perspective
332 during post-monsoon haze.

333 BBOA exhibits enhanced signals at levoglucosan-related fragments m/z 60 and 73,
334 characteristic of biomass-burning emissions(Figure 2a), and shows low oxygenation ($O:C = 0.09$;
335 Table S1 in Supporting Information), indicative of fresh crop-residue burning and agricultural
336 smoke. BBOA was a major dominant contributor during the post-monsoon haze over this site as
337 reported in the Rathore et al. (2025). The relatively low $O:C$ for BBOA (slightly lower than HOA)
338 should not be interpreted as a physical inversion of oxidation state. Capture vaporizer
339 fragmentation in ACSM-type instruments is known to underestimate $O:C$ for certain primary and
340 multifunctional OA by ~ 20 -46%, compressing expected separations among primary factors
341 (Canagaratna et al., 2015). In addition, the f_{44} - f_{60} triangle shown previously (see Figure 8 of
342 Rathore et al. (2025)) indicates strong dominance of fresh BBOA during the biomass-burning
343 period, consistent with the interception of relatively unprocessed smoke at this peri-urban
344 receptor in the IGP. From a process standpoint, the identification of a strong fresh biomass-



345 burning factor at Sonipat indicates that a substantial fraction of primary carbonaceous aerosol
346 can reach this upwind receptor directly, without requiring significant atmospheric processing
347 within Delhi. This finding highlights that peak PM levels across the wider NCR are strongly
348 influenced by episodic regional open burning, and therefore cannot be effectively mitigated by
349 city-level measures alone.

350 SFCOA accounts for a substantial fraction of OA (23%)(Figure 2b) and is characterized by
351 elevated aromatic fragments (e.g., m/z 55), along with m/z 60 and m/z 73 peaks(Figure 2a) that
352 are lower than in BBOA (Tobler et al., 2020). It also exhibits moderately elevated oxygenation $O:C$
353 = 0.21(Table S1 in Supporting Information), suggesting a mixture of residential and small-scale
354 industrial solid fuel combustion. SFCOA is moderately oxygenated ($O:C = 0.21$; Table S1 in
355 Supporting Information), which is consistent with a mixed source class combining residential
356 solid-fuel use (e.g., wood/dung/crop burning residues) and small-scale industrial combustion
357 typical of peri-urban northern India. The partial overlap in tracer ions between BBOA and SFCOA
358 is expected, but the resolved profiles and their temporal behavior (Section 3.1.6) suggest that
359 the two factors capture distinct emission mixtures and, potentially, distinct combustion practices
360 and fuel types. This distinction is valuable for mitigation because agricultural open burning and
361 household/industrial solid-fuel combustion respond to different regulatory and socio-economic
362 interventions. The prominence of SFCOA at an upwind NCR site further indicates that strategies
363 focused exclusively on urban traffic controls will have limited impact on the dominant
364 carbonaceous components during post-monsoon haze periods when the regional plume is the
365 controlling influence.

366 **1.1.2. Secondary OA Factors (SVOOA, LVOOA)**

367 SVOOA and LVOOA components separated primarily by oxidation state and volatility (Guo
368 et al., 2020), together constitute approximately 57.5% of the total OA mass (Figure 2b). The
369 higher contribution of SVOOA and LVOOA, indicates that secondary processing is a dominant
370 contributor to carbonaceous aerosol even at this upwind receptor. LVOOA is the largest single
371 OA component (38%) and is characterized by strong m/z 44 (CO_2^+) and high oxygenation ($O:C =$
372 0.91; Table S1), indicative of highly aged, regionally transported aerosol(Figure 2a). The stronger



373 m/z 44 signal in SOA relative to POA is associated with thermal decomposition of highly
374 oxygenated organic acids (Ng et al., 2010). LVOOA is also distinguished from SVOOA by a lower
375 f_{43}/f_{44} ratio (0.22 for LVOOA versus 0.33 for SVOOA), consistent with more extensive oxidation
376 and functionalization. SVOOA is also oxygenated but comparatively less aged ($O:C = 0.72$; Table
377 S1), consistent with a fresher SOA fraction formed through rapid oxidation and/or partitioning of
378 semi-volatile products during transport. As expected, SVOOA also shows a lower $O:C$ value of
379 0.72 than LVOOA, which represents a comparatively fresher secondary fraction formed through
380 rapid oxidation of volatile organic precursors.

381 The persistence of a large oxygenated fraction at Sonipat is mechanistically important
382 because it suggests that secondary mass is not merely a downwind feature of Delhi's urban
383 atmosphere but is already established in the upwind, before mixing with the local urban
384 emissions. This implies that reducing peak $PM_{2.5}$ concentrations during post-monsoon haze is
385 unlikely to be achieved solely by controlling urban emissions and instead requires coordinated
386 control of regional emission sources that supply both primary organic mass and the chemical
387 feedstock for continued oxidation and aging. In addition, the dominance of LVOOA indicates that
388 haze periods can be sustained by a regional aged background even when short-lived primary
389 spikes subside, meaning that mitigation measures must address both episodic emission and the
390 underlying transported burden that sets the baseline on which those pulses are superimposed.

391 **3.1.3. Black carbon source apportionment and linkage to OA sources**

392 Aethalometer based BC source apportionment reveals a pronounced dominance of
393 biomass/wood burning related equivalent black carbon (eBC_{bb}), which accounts for
394 approximately 77.8% of total BC, compared with fossil-fuel related black carbon (eBC_{ff} , 22.2%)
395 (Figure 2b). This strong biomass-burning signature is consistent with the geographical setting of
396 the Sonipat site that is along the principal post-monsoon transport pathway connecting the
397 agricultural burning regions of Punjab and Haryana to the Delhi-NCR. In addition to open-field
398 residue burning, widespread use of solid fuels (wood, dung, and crop residues) across peri-urban
399 and rural northern India likely contributes to the observed biomass dominated BC. The relatively
400 modest fossil-fuel contribution reflects the reduced influence of dense urban traffic at this



401 upwind receptor during haze episodes that are largely governed by regional smoke transport
402 under shallow and stagnant boundary-layer conditions. The BC apportionment is also internally
403 consistent with the behavior of the PMF resolved organic aerosol factors (Figure S1; Figure S4),
404 which show enhanced contributions from biomass-burning and solid-fuel combustion OA during
405 polluted periods, thereby supporting a coherent combustion-driven interpretation of the
406 carbonaceous aerosol mixture.

407 In absolute terms, biomass-burning BC (eBC_{bb}) averaged $10.9 \mu\text{g m}^{-3}$ over the study
408 period, substantially exceeding fossil-fuel BC (eBC_{ff}), which averaged $3.1 \mu\text{g m}^{-3}$ (Table S2). This
409 pattern contrasts sharply with observations from urban Delhi, where fossil-fuel BC typically
410 dominates during winter and post-monsoon periods. For example, Dumka et al. (2018) reported
411 that eBC_{ff} contributed about 72% of total BC ($17.6 \mu\text{g m}^{-3}$) compared with $6.8 \mu\text{g m}^{-3}$ from
412 biomass burning during December 2015–February 2016, while (Goel et al., 2021) found fossil-
413 fuel BC contributions of 68.85% ($7.90 \mu\text{g m}^{-3}$) versus 31.15% ($4.73 \mu\text{g m}^{-3}$) from biomass burning
414 during August–November 2020. In contrast, the Sonipat observations show a reversed
415 distribution, with biomass-burning sources accounting for nearly four-fifths of total BC,
416 reinforcing the interpretation of this location as a regional receptor for transboundary
417 combustion emissions rather than a traffic-dominated urban environment.

418 Beyond source attribution, this predominance of biomass-derived BC has important
419 implications for both air quality and regional climate. BC emitted alongside biomass-burning
420 organic species can substantially modify aerosol optical properties, enhance light absorption, and
421 contribute to atmospheric heating, which in turn may influence boundary-layer stability and
422 pollutant accumulation during haze events (Sahu et al., 2012; Shiraiwa et al., 2008). Although
423 detailed quantification of such radiative and dynamical feedback is beyond the scope of the
424 present analysis, the strong co-variation between biomass related BC and combustion-derived
425 OA factors indicates that emission-control strategies targeting post-monsoon biomass and solid-
426 fuel combustion would simultaneously reduce light-absorbing carbon and the bulk organic
427 aerosol mass that drives elevated $\text{PM}_{2.5}$ at this upwind receptor.

428 **3.1.4. Inorganic Fraction and Implications for Haze Chemistry**



429 Although secondary inorganic aerosols (NO_3^- , SO_4^{2-} , NH_4^+ , and Cl^-) account for only ~20%
430 of bulk $\text{PM}_{2.5}$ mass (Figure 2b), the dominance of oxygenated OA within the organic fraction
431 implies that secondary mass growth at Sonipat is more strongly governed by organic
432 oxidation/aging processes than by inorganic neutralization and salt formation. This secondary
433 organic heavy character is consistent with a receptor that intercepts biomass-burning plumes
434 and their processed products before the air mass mixes with stronger urban NO_x and ammonia
435 emission environments that can promote substantial nitrate formation. The result is a haze
436 chemical regime in which aged carbonaceous aerosol, rather than sulfate/nitrate, plays an
437 outsized role in determining PM mass and optical properties at this upwind gateway to Delhi-
438 NCR.

439 This contrast is important for understanding why Delhi-centric mitigation strategies that
440 prioritize NO_x -, NH_3 -, or sulfate (SO_4^{2-}) driven inorganic aerosol control may not fully address the
441 severity of upwind haze during the post-monsoon period. At Sonipat, the dominant mass resides
442 in carbonaceous aerosol, much of it oxygenated, implying that regional combustion controls and
443 interventions that reduce primary carbonaceous emissions and their reactive precursors are
444 likely to provide stronger leverage for reducing the intensity of upwind haze events that
445 subsequently influence Delhi-NCR. At the same time, the presence of non-negligible nitrate and
446 chloride indicates that thermodynamic partitioning and multiphase processes still contribute to
447 the aerosol burden, particularly during cool, humid and stagnant conditions that favor
448 condensation and heterogeneous processing, setting up conditions conducive to sustained haze
449 for longer duration.

450 **3.1.5. Comparison with prior OA factorization studies over Delhi and the IGP**

451 A significant amount of work over Delhi and the wider IGP has reported the dominant role
452 of SOA contributions, particularly LVOOA, though with strong seasonality and event dependence
453 (Bhandari et al., 2020; Bhowmik et al., 2022; Goel et al., 2024; Lalchandani et al., 2022). For
454 example, Shukla et al., 2021 applied PMF to ACSM measurements during summer season in Delhi
455 and found strong LVOOA dominance, consistent with extensive photochemical aging under high
456 insolation and oxidant availability. Wintertime and post-monsoon studies more complex coupling



20

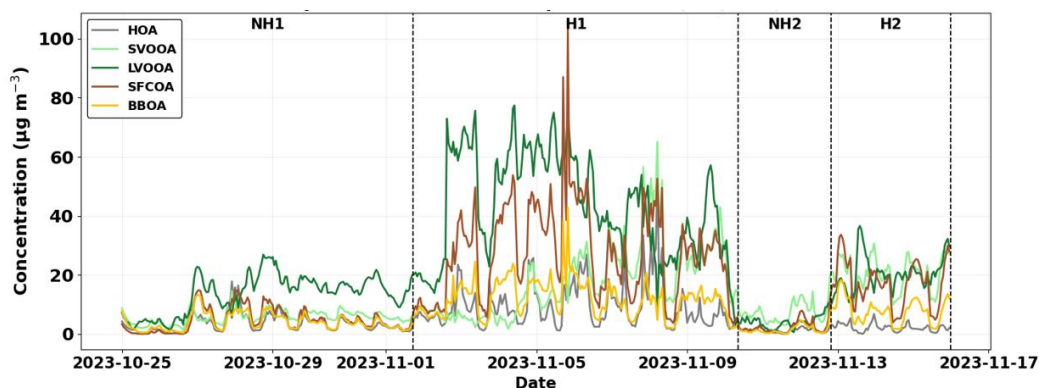
457 between primary emissions and secondary processing. Thamban et al. (2021), demonstrated that
458 SOA enhances not only OA but also the effective mass loading of primary OA families and
459 wintertime inorganic species (e.g., nitrate and chloride), pointing to multiphase processing and
460 nocturnal chemistry as critical amplifiers during haze (Haslett et al., 2023; Mathai et al., 2026).

461 More recent urban investigations further support this framework. Lalchandani et al.
462 (2021) resolved traffic, biomass-burning, and oxygenated OA factors over Delhi and highlighted
463 strong coupling between biomass-burning emissions and secondary OA formation during post-
464 monsoon and winter periods. Similarly, Reyes-Villegas et al, (2021) reported oxygenated OA
465 dominance during polluted conditions, consistent with enhanced oxidation states reflecting
466 regional transport and aging. Across the IGP, (Panda et al., 2025) reported substantial LVOOA
467 during prolonged haze, reinforcing the role of sustained secondary processing under stagnant
468 meteorological conditions. (Cash et al., 2023) emphasized that SOA formation is strongly
469 modulated by meteorology and episodic primary emissions, producing large temporal variability
470 in OA composition even within the same season. Within this broader context, the prominent
471 LVOOA fraction observed at Sonipat is consistent with the regional picture of strong secondary
472 processing, while the magnitude and timing of primary combustion enhancements (especially
473 SFCOA/BBOA) indicate that the proximity to upwind sources and shorter aging times can shift
474 the balance toward primary emissions during certain periods, particularly short-lived events such
475 as Diwali.

476 A key added value of the Sonipat perspective is that it enables a clearer separation of
477 regional aerosol evolution occurring prior to arrival in Delhi from the intense urban mixing and
478 source complexity within the city. The relatively low HOA and the dominant biomass-related BC
479 fraction indicate that the upwind plume arriving at this intermediate receptor is compositionally
480 distinct from Delhi's urban aerosol, and that the commonly reported large HOA contributions in
481 Delhi represent additional urban increments rather than the nature of the regional background
482 plume. This distinction matters for air-quality planning because it implies that a large fraction of
483 the haze burden affecting Delhi-NCR can be pre-formed outside the city, requiring coordinated
484 regional actions rather than city-only responses.



485 **3.1.6. Temporal Evolution of OA Factors Across Haze and non-Haze Periods**



486 **Figure 3:** Hourly time series of OA source components ($\mu\text{g m}^{-3}$) during the sampling period. Five
 487 factors were resolved via PMF: HOA, SVOOA, LVOOA, SFCOA, and BBOA. Dashed lines demarcate
 488 non-haze (NH) and haze (H) periods based on the classification criteria established by Rathore et
 489 al. (2025).

490 Figure 3 shows the hourly time series of the five deconvolved OA factors during the study
 491 period. The time series reveals sharp regime shifts in both mass loading and source contributions,
 492 reflecting the combined effects of changing emissions intensity in the source regions, evolving
 493 transport pathways, and the local boundary-layer state at the receptor.

494 During the initial non-haze phase (NH1), OA levels were relatively modest and the background
 495 was dominated by oxygenated OA, indicating a persistent regional secondary baseline even prior
 496 to the onset of severe haze. This baseline is important because it implies that even “non-haze”
 497 conditions can carry substantial aged material, so the transition to haze represents not a clean-
 498 to-dirty switch but an amplification of an already elevated carbonaceous background. As the
 499 system transitioned into H1, all OA factors increased markedly, with especially large
 500 enhancements in SFCOA and BBOA. SFCOA increased by roughly a factor of six relative to NH
 501 conditions and exceeded $100 \mu\text{g m}^{-3}$ during 5-6 November, while BBOA increased by roughly a
 502 factor of four. The timing is consistent with intensified upwind burning activity, also reflected by
 503 elevated fire counts (Figure S12). Such concurrent enhancements of primary factors are typical
 504 of IGP haze formation when stagnant meteorology and shallow boundary layers limit dispersion



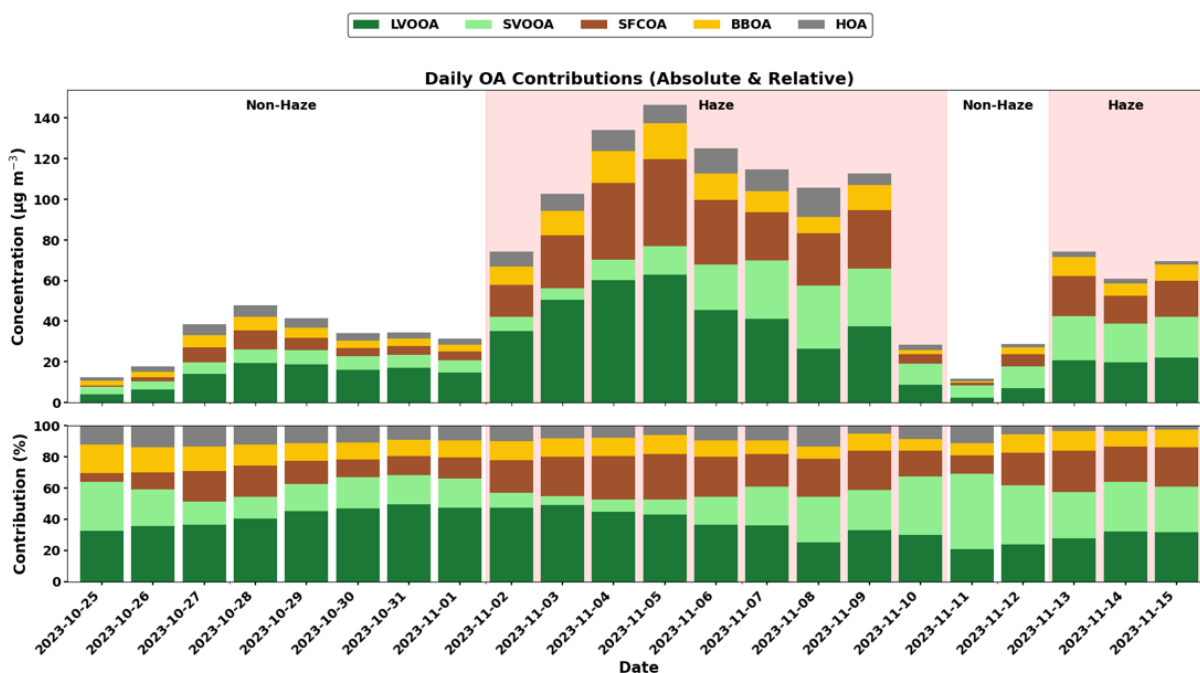
505 and promote the accumulation of regionally transported combustion emissions (Lalchandani et
506 al., 2022; Rathore et al., 2025). Importantly, LVOOA remained persistently elevated throughout
507 H1, indicating that regional processing/aging continues to contribute substantial mass even
508 during periods dominated by primary emission spikes. This sustained LVOOA suggests that multi-
509 day accumulation and aging within the IGP transport corridor can maintain high oxygenated OA
510 even when emission pulses fluctuate, implying that short-term episodic controls may need to be
511 complemented by sustained reductions in regional combustion emissions to lower the baseline
512 that supports prolonged haze.

513

514 Across the entire study period, LVOOA had the highest overall mean concentration ($25.0 \mu\text{g}/\text{m}^3$),
515 followed by SFCOA ($15.1 \mu\text{g}/\text{m}^3$) and SVOOA ($12.8 \mu\text{g}/\text{m}^3$) (Table S2). These concentrations are
516 comparable to, and in some cases higher than, values reported in previous PM_1 studies from
517 urban center Delhi, noting that our study collects the $\text{PM}_{2.5}$ that captures additional aged
518 material at larger sizes. For example, in Old Delhi during post-monsoon, total oxygenated OA
519 (LVOOA + SVOOA) reached $29.13 \mu\text{g}/\text{m}^3$ in PM_1 (Cash et al., 2021), whereas our $\text{PM}_{2.5}$
520 measurements show combined OOA (LVOOA + SVOOA) is $37.8 \mu\text{g}/\text{m}^3$, consistent with enhanced
521 inclusion of larger, more processed material. Mean HOA ($5.5 \mu\text{g m}^{-3}$) is broadly consistent with
522 reported HOA ranges of $\sim 3\text{-}5 \mu\text{g m}^{-3}$ in PM_1 in Delhi (Bhandari et al., 2020; Tobler et al., 2020),
523 reinforcing that Sonipat is not a traffic-dominated urban core receptor and that the haze burden



524 here is shaped primarily by regional combustion and subsequent aging rather than by intense
 525 local traffic increments.



526

527 **Figure 4:** Daily organic aerosol (OA) source contributions during haze and non-haze periods from
 528 25 October 2023 to 15 November 2023. The upper panel shows absolute concentrations ($\mu\text{g m}^{-3}$)
 529 of OA factors, including low-volatility oxygenated OA (LVOOA), semi-volatile oxygenated OA
 530 (SVOOA), solid fuel combustion OA (SFCOA), biomass burning OA (BBOA), and hydrocarbon-like
 531 OA (HOA). The lower panel presents the corresponding relative contributions (%) of each OA
 532 factor to total OA. Shaded regions indicate haze periods, while unshaded regions represent non-
 533 haze conditions.

534 Figure 4 presents daily absolute and fractional OA contributions. Secondary OA (SVOOA +
 535 LVOOA) remained dominant during haze periods, frequently contributing >55% of total OA,
 536 consistent with continued secondary processing under stagnant and humid conditions. At the
 537 same time, strong enhancements in SFCOA and BBOA during haze indicate that primary



538 combustion emissions are the immediate drivers of the most intense peaks, with secondary
539 formation contributing to large and persistent baseline on top of which primary spikes are
540 superimposed. This structure helps reconcile two seemingly competing narratives about IGP
541 haze: it is both combustion-driven and strongly secondary. Combustion emissions provide the
542 mass and precursors, while stagnant meteorology and multiphase processing sustain and amplify
543 oxygenated aerosol. Notably, during H1, LVOOA dominated early and remained persistently high,
544 whereas during H2 the SVOOA and LVOOA contributions were more comparable, indicating
545 differences in event character and/or processing pathways that are consistent with the shift from
546 a prolonged regional burning episode to a shorter event with strong local/regional pulse
547 emissions associated with Diwali.

548 **3.1.7. Episode-wise contrasts (H1 versus H2; NH periods)**

549 For the BC fraction, biomass burning emissions (eBC_{bb}) were the dominant contributor
550 with an overall mean of $10.9 \mu\text{g}/\text{m}^3$, significantly outweighing the average fossil fuel fraction
551 (eBC_{ff}), of $3.1 \mu\text{g}/\text{m}^3$ (Table S2). This contrasts markedly with Delhi urban winter studies where
552 eBC_{ff} dominated at 72% with mean eBC_{ff} of $17.6 \mu\text{g}/\text{m}^3$ versus eBC_{bb} of $6.8 \mu\text{g}/\text{m}^3$ during
553 December 2015-February 2016 (Dumka et al., 2018) and eBC_{ff} contributed 68.85% ($7.90 \mu\text{g}/\text{m}^3$)
554 compared to eBC_{bb} at 31.15% ($4.73 \mu\text{g}/\text{m}^3$) during August-November 2020 (Gupta et al., 2022).
555 However, measurements at this upwind Sonipat site shows an opposite pattern, with wood
556 burning accounting for approximately 78% of total BC, confirming the site's status as a regional
557 receptor for transboundary biomass burning emissions rather than a localized traffic hotspot.

558 The first haze episode (H1) exhibited the highest concentrations for most factors. LVOOA
559 reached $42.8 \mu\text{g}/\text{m}^3$ during H1 compared to $13.5 \mu\text{g}/\text{m}^3$ in the preceding non-haze period (NH1),
560 while SFCOA increased from $4.6 \mu\text{g m}^{-3}$ (NH1) to $27.5 \mu\text{g m}^{-3}$ (H1) and BBOA from 4.0 to $11.8 \mu\text{g}$
561 m^{-3} (Table S2). This magnitude is exceptional when compared to reported values from Delhi-NCR
562 and other global megacities. For instance, during severe winter haze episodes in Beijing, total OA
563 concentrations ranged from 43.8 to $87.9 \mu\text{g}/\text{m}^3$ with SOA contributing 46-66% (Zhao et al., 2019),
564 indicating that the aged oxygenated fraction at Sonipat during H1 approaches the lower end of
565 total OA seen in Chinese megacity haze, despite Sonipat being a semi-urban upwind receptor.



566 This comparison demonstrates that the extreme aerosol aging and accumulation can occur over
567 the IGP transport corridor, not only within major urban cores, and therefore regional emission
568 reductions can yield substantial benefits by reducing the mass that is already present before the
569 plume reaches large population centers.

570 Similarly, SFCOA concentrations rose dramatically from $4.6 \mu\text{g}/\text{m}^3$ in NH1 to $27.5 \mu\text{g}/\text{m}^3$
571 in H1, while BBOA increased from $4.0 \mu\text{g}/\text{m}^3$ to $11.8 \mu\text{g}/\text{m}^3$ (Table S2). These SFCOA levels during
572 H1 are notably higher than typical Delhi urban winter averages, where combined SFCOA factors
573 contributed approximately 20-35% of total OA in PM_{10} measurements across urban Delhi and
574 suburban Faridabad sites (Lalchandani et al., 2022; Tobler et al., 2020).

575 In contrast, the second haze episode (H2) showed a different SOA partitioning as SFCOA
576 remained elevated ($17.0 \mu\text{g m}^{-3}$), but SVOOA ($20.4 \mu\text{g m}^{-3}$) became comparable to LVOOA (20.2
577 $\mu\text{g m}^{-3}$)(Table S2). This shift points to a comparatively stronger contribution from fresh secondary
578 formation and/or rapid partitioning processes during H2, consistent with a shorter-duration
579 event in which newly emitted precursors and semi-volatile products can contribute strongly
580 before being fully processed into a low-volatility fraction. Such behavior is consistent with the
581 expectation that Diwali-period emissions can provide intense but transient injections of reactive
582 organics and primary particles, and that the resulting aerosol may show a relatively younger
583 oxygenated signature compared with a prolonged regional haze dominated by multi-day
584 accumulation. The rain influenced non-haze period (NH2) was the cleanest, with LVOOA dropping
585 to $4.1 \mu\text{g m}^{-3}$ and BBOA to $1.5 \mu\text{g m}^{-3}$ (Table S2), consistent with wet scavenging and enhanced
586 ventilation, comparable to clean-season behavior reported for Delhi (Gani et al., 2020). The sharp
587 reductions during NH2 also emphasize the effectiveness of meteorological cleansing in resetting
588 the regional aerosol background, and they highlight why sustained emission reductions are
589 needed to prevent rapid re-accumulation under the next stagnation event in H2.

590 During H1, eBC_{bb} peaked at $14.2 \mu\text{g m}^{-3}$ while eBC_{ff} averaged $3.8 \mu\text{g m}^{-3}$ (Table S2). During
591 H2, eBC_{bb} remained substantial ($8.3 \mu\text{g m}^{-3}$), while eBC_{ff} decreased strongly (reported here as
592 $0.03 \mu\text{g m}^{-3}$), highlighting the overwhelming dominance of biomass/solid-fuel sources during that
593 episode. These results are consistent with isotopic constraints indicating that open burning of



594 post-harvest crop residue and wood in nearby rural regions can contribute $\sim 42 \pm 17\%$ to severe
595 haze in Delhi during autumn and winter (Bikkina et al., 2019); the higher biomass fraction at
596 Sonipat is expected because of its closer proximity to upwind burning regions and weaker
597 masking by local fossil-fuel emissions. Taken together, the episode wise contrasts indicate that
598 the most severe events at Sonipat are driven by a combination of intensified regional combustion
599 influence and suppressed dispersion, while the relative balance between aged and fresh
600 oxygenated OA can shift depending on event duration, precipitation interruptions, and the
601 presence of festival-related emission pulses. This has practical relevance because it implies that
602 episodic interventions targeting a single source category may have limited efficacy unless they
603 are aligned with the dominant regime as prolonged regional haze requires sustained regional
604 combustion reductions, whereas short lived Diwali linked events may respond more to targeted
605 restrictions and timing specific controls.

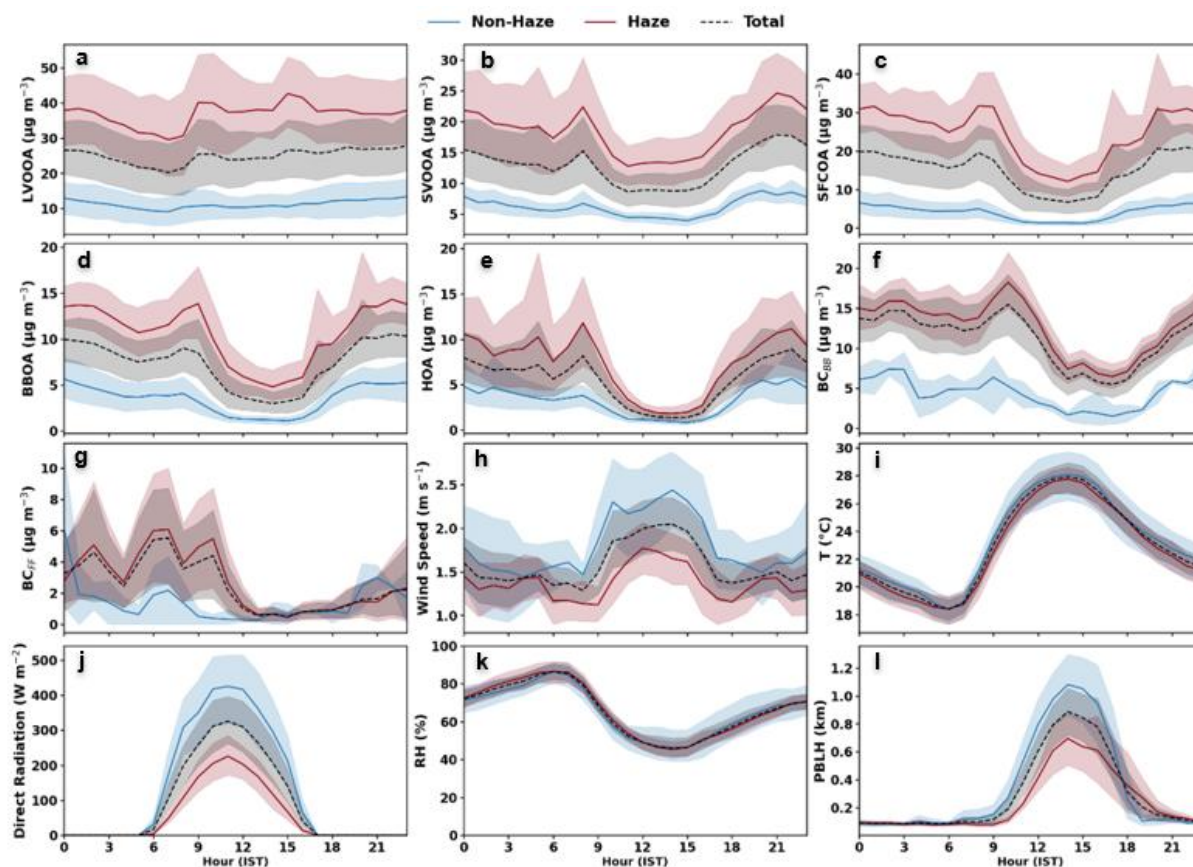
606 The temporal evolution of OA at Sonipat reveals pronounced shifts in both mass loading
607 and source dominance during the transition from non-haze to haze conditions. OA constituted
608 approximately 65% of the NR-PM_{2.5} mass and exhibited multi-day accumulation episodes, with
609 daily mean concentrations peaking near $140 \mu\text{g m}^{-3}$ during the first major haze event (H1). Under
610 non-haze conditions, the aerosol composition was dominated by regionally aged secondary
611 organic aerosol, particularly low-volatility oxygenated organic aerosol (LVOOA; mean $\sim 11.2 \mu\text{g}$
612 m^{-3}), reflecting the persistent oxidized background characteristic of the Indo-Gangetic Plain. With
613 the onset of haze phases (H1 and H2), there was a marked intensification of primary combustion-
614 related factors, notably solid-fuel combustion OA (SFCOA) and biomass-burning OA (BBOA),
615 which drove the extreme OA enhancements. This behavior is consistent with regional biomass-
616 burning influences documented for the Delhi–NCR upwind sector (e.g., Rathore et al., 2025) and
617 underscores the role of transported combustion emissions in modulating pollution levels at the
618 site. Overall, the observations demonstrate that OA at Sonipat is shaped by a sustained regional
619 oxygenated background onto which episodic influxes of primary combustion emissions are
620 superimposed, leading to severe carbonaceous aerosol loading even before air masses mix with
621 additional urban emissions from Delhi. These findings reinforce the necessity of airshed-scale



27

622 mitigation strategies that simultaneously address regional biomass and solid-fuel combustion
 623 sources in conjunction with local urban controls.

624 **3.2 Diurnal Variation of Carbonaceous Aerosols and effect of Meteorology**



625

626 **Figure 5:** Diurnal variations of PMF factors: HOA, SVOOA, LVOOA, SFCOA and BBOA; 2 resolved
 627 BC factors: eBC_{bb} and eBC_{ff} , meteorological parameters: Wind Speed, Temperature, Direct
 628 Radiation, Relative Humidity, PBLH during Haze (H1+H2), Non-Haze (NH1+ NH2) and study period
 629 (Total) from 25 October 2023 to 15 November 2023.

630

Figure 5 shows mean diurnal cycles of PMF resolved OA factors (HOA, SVOOA, LVOOA,
 631 SFCOA, BBOA), the source apportioned BC components (eBC_{bb} and eBC_{ff}), and key
 632 meteorological variables for haze (H1+H2), non-haze (NH1+NH2), and the full periods. The



633 diurnal behavior reflects the combined influences of the variation in emission, boundary layer
634 dynamics, and secondary formation/partitioning. The features provide a process level evidence
635 for how daily meteorological cycling modulates both primary emission accumulation and the
636 partitioning/aging of oxygenated aerosol components. The correlation values between different
637 variables and their tracers can be found in supplementary material (Figure S3-S6)

638 **3.2.1. Traffic-related factors (HOA and eBC_{ff})**

639 The diurnal variation of both HOA and eBC_{ff} exhibit coherent diurnal variability
640 characterized by enhanced nighttime concentrations (21:00–00:00 IST) and a secondary peak
641 during the morning hours (06:00–09:00 IST; Figure 5e). The similarity in their temporal
642 evolution (Figure S1) and a moderate correlation between the two species ($R^2 = 0.48$; Figure 5g)
643 indicates their common origin in traffic-related emissions and with previous observations from
644 urban and peri-urban environments (Bhandari et al., 2020; Crippa et al., 2013; Lalchandani et al.,
645 2022). Elevated nighttime levels likely reflect increased heavy-duty vehicle activity following
646 relaxation of daytime traffic restrictions, while the morning peak corresponds to commuter
647 traffic. These patterns are evident under both haze and non-haze conditions; however, the
648 amplitude of the nocturnal diurnal contrast is substantially stronger during haze episodes when
649 shallow boundary layers suppress dilution and promote accumulation (Figure S7).

650 During non-haze periods, deeper daytime mixing, precipitation-driven scavenging, and
651 reduced combustion-related activity lead to lower concentrations and a weaker diurnal contrast,
652 in agreement with Rathore et al. (2025). Notably, the overall magnitude of HOA remains relatively
653 modest even during severe haze, indicating that traffic emissions are not the principal driver of
654 peak PM_{2.5} at this upwind receptor. Nevertheless, localized traffic contributions can still produce
655 short-term enhancements under nocturnal stagnation, particularly when combined with
656 regionally transported pollution.

657 **3.2.2. Biomass/solid-fuel factors (SFCOA, BBOA, and eBC_{bb})**

658 SFCOA, BBOA, and eBC_{bb} exhibit coherent diurnal behavior (Figure 5), with their
659 concentrations rising from late evening (~19:00 IST), remaining elevated through the night, and



660 declining after morning (~08:00 IST). The eBC_{bb} correlates strongly with SFCOA ($R^2 = 0.63$; Figure
661 5f), and BBOA exhibits an even stronger correlation with SFCOA ($R^2 = 0.95$; Figure 5d), indicating
662 that the nocturnal enhancement reflects widespread biomass/solid-fuel combustion influences.
663 Such nighttime peaks are consistent with residential combustion timing (cooking/heating) and
664 with the tendency for agricultural burning and regional smoke impacts to be most pronounced
665 under stable nighttime boundary layers (Singh et al., 2023; Tobler et al., 2020). A smaller
666 secondary enhancement around ~08:00–10:00 IST is consistent with morning household activity
667 and localized waste burning. As with traffic factors, haze-period boundary-layer suppression
668 amplifies these nocturnal enhancements.

669 The correlated temporal variations of SFCOA, BBOA, and eBC_{bb} also imply that at Sonipat,
670 combustion sources contributing to organic aerosol mass and light-absorbing carbon are coupled
671 in time and likely co-located within the regional influence footprint, rather than representing
672 independent local-only behaviors. This coupling strengthens the inference that a large fraction
673 of the carbonaceous aerosol burden during haze is regionally coherent and thus amenable to
674 coordinated regional interventions, including strategies aimed at reducing open burning and
675 promoting cleaner household and small-scale industrial fuels.

676 **3.2.3. Contrasting diurnal behavior of SVOOA and LVOOA**

677 Similar diurnal patterns, characterized by higher nighttime and lower daytime
678 concentrations, are observed for HOA and eBC_{ff} , during both haze and non-haze periods.
679 However, this contrast is substantially stronger during haze conditions. The weaker diurnal
680 contrast during non-haze periods may be associated with enhanced atmospheric mixing due to
681 higher boundary-layer heights, precipitation-driven scavenging, and reduced combustion-related
682 activities, as reported by Rathore et al. (2025).

683 SVOOA and LVOOA show clearly distinct diurnal signatures that reflect distinct formation
684 mechanisms and atmospheric processing pathways. SVOOA shows pronounced nocturnal
685 enhancement and covaries with particulate nitrate (NO_3^-) and chloride (Cl^-) (Figure S2),
686 consistent with thermodynamically driven gas–particle partitioning (Zhang et al., 2021). As

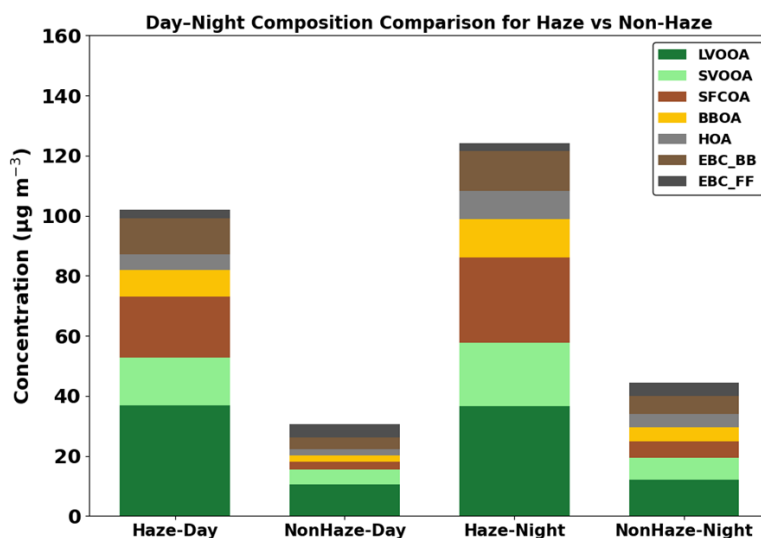


687 nighttime temperatures decrease and relative humidity increases, semi-volatile species,
688 including inorganic nitrate and freshly formed organic oxidation products, preferentially partition
689 to the particle phase (Lanz et al., 2007). This behavior supports the interpretation that SVOOA
690 represents a relatively fresh SOA fraction formed through rapid oxidation of local and regional
691 precursors followed by volatility-controlled condensation. Its diurnal variability therefore reflects
692 both chemical production rates and equilibrium partitioning processes. In contrast, LVOOA
693 closely tracks sulfate (SO_4^{2-}) (Figure 5a and Figure S1) and exhibits a strong correlation ($R^2 = 0.86$),
694 indicating a regionally aged component that covaries with slower-forming secondary aerosol
695 produced and transported over multi-day timescales. The comparatively flat diurnal profile of
696 LVOOA further supports a transported and highly processed origin. This interpretation is
697 consistent with previous studies over Delhi and the IGP, which attribute LVOOA to highly
698 oxidized, carboxylic-acid-rich aerosol formed via sustained regional processing (Rathore et al.,
699 2025; Shukla et al., 2023, 2021) and aligns with observations from other polluted megacities such
700 as Guangzhou, where LVOOA was predominantly associated with transported aged aerosol (Guo
701 et al., 2020). Across pollution episodes, LVOOA reaches its highest concentrations during H1,
702 followed by H2, consistent with stronger regional accumulation during the prolonged biomass-
703 burning haze, whereas SVOOA remains broadly comparable between H1 and H2, suggesting that
704 once favorable thermodynamic and chemical conditions are established, the semi-volatile
705 secondary fraction responds rapidly and approaches saturation.

706 The day-night contrasts summarized in Table S3 (Supporting Information) and illustrated
707 in Figure 6 further clarify these behaviors. Transitioning from non-haze (NH1, NH2) to haze
708 conditions (H1, H2), haze-night totals increase by approximately 3-4 times relative to NH-day
709 baselines. For example, LVOOA rises from $12.8 \mu\text{g m}^{-3}$ (NH1 day) to $42.9 \mu\text{g m}^{-3}$ (H1 night), while
710 SVOOA increases from $4.7 \mu\text{g m}^{-3}$ to $20.3 \mu\text{g m}^{-3}$ (H2 night). Primary combustion-related factors
711 show similarly strong nocturnal amplification, with SFCOA increasing from $3.1 \mu\text{g m}^{-3}$ (NH1 day)
712 to $31.2 \mu\text{g m}^{-3}$ (H1 night) and BBOA from 2.7 to $13.5 \mu\text{g m}^{-3}$, reflecting enhanced evening
713 combustion activity and boundary-layer contraction (Lakra et al., 2024). These findings are
714 consistent with the post-monsoon PM_{10} observations of Cash et al. (2021), where biomass-related
715 factors (SVBBOA and SFOA) peaked under shallow nighttime boundary layers. Figure 6



716 demonstrates that haze-night composition is dominated by oxygenated secondary components
 717 (SVOOA and LVOOA, ~60%), superimposed on elevated SFCOA, BBOA, and biomass-burning black
 718 carbon (eBC_{bb}). The increase in eBC_{bb} from 13.6 to 14.8 $\mu\text{g m}^{-3}$ between H1 day and night parallels
 719 the rise in SVOOA (e.g., 9.1 to 20.3 $\mu\text{g m}^{-3}$ during H2), linking biomass-derived precursors to
 720 enhanced secondary formation under weakly ventilated nocturnal conditions. Comparable
 721 wintertime nocturnal enhancement of SVOOA has been reported by (Tobler et al., 2020),
 722 although the stronger nighttime increase in HOA observed here (6.6 to 12 $\mu\text{g m}^{-3}$ during H1)
 723 suggests more pronounced trapping of traffic emissions under stagnation.



724 **Figure 6:** Daytime and nighttime OA and eBC factors during haze and non-haze periods. Bars
 725 represent mean contributions of organic factors: HOA, SVOOA, LVOOA, SFCOA, and BBOA and
 726 EBC factors: eBC_{ff} and eBC_{bb}. Data is for the full period from 25th October 2023 to 15th November
 727 2023.

728 Meteorological conditions (Figure 5 and Figure S7) provide the physical framework
 729 underlying these compositional shifts. During haze episodes, particularly H1, planetary boundary
 730 layer heights (PBLH) remain suppressed (<500 m during daytime and <200 m at night), wind
 731 speeds are weak (often <1.5–2.0 m s^{-1}), and direct solar radiation is reduced. Elevated relative
 732 humidity (frequently 60–80%) further favors heterogeneous and multiphase processing. These



733 conditions create an effective meteorological lid that traps primary emissions (SFCOA, BBOA,
734 HOA) while simultaneously promoting secondary mass growth and chemical aging. In contrast,
735 non-haze periods (NH1 and NH2) are characterized by deeper daytime PBL development (often
736 >1000 m), stronger ventilation, higher radiation, and, in NH2, rainfall-driven scavenging, leading
737 to lower aerosol concentrations and reduced diurnal amplitude. Notably, LVOOA maintains a
738 persistently elevated baseline during haze ($\sim 42.8 \mu\text{g m}^{-3}$ during both day and night in H1),
739 underscoring its regional persistence. This behavior indicates that even if local emissions
740 fluctuate, aged, oxygenated aerosols can remain elevated under stagnant conditions due to
741 sustained regional inflow and limited ventilation. Consequently, emission reductions alone may
742 not immediately translate into lower concentrations during strong stagnation events unless
743 regional transport is also mitigated, as shallow boundary-layer conditions facilitate multi-day
744 retention and recycling of aged secondary aerosol.

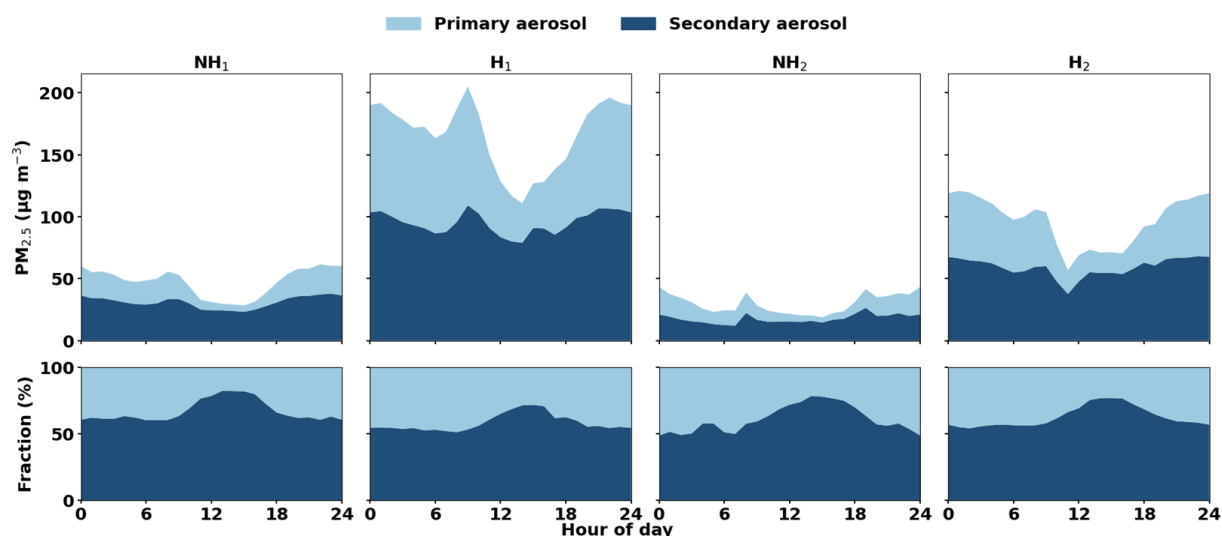
745

746 **3.3 Primary versus Secondary Carbonaceous Aerosol Dynamics**

747 To interpret haze growth and chemical evolution, we examine the interplay between
748 primary and secondary components of C-PM_{2.5} (Figure 7; Figure S13; Table S4-S5). Primary
749 components (SFCOA + BBOA + HOA + eBCff + eBCbb + Cl⁻) reflect direct combustion emissions
750 and show strong sensitivity to activity patterns and boundary-layer dynamics, with bimodal peaks
751 in the early morning and late evening during haze periods. As the boundary layer deepens during
752 daytime, primary concentrations typically decrease through dilution, unless regional transport
753 continues to supply fresh emissions. The prominence of SFCOA and BBOA in the primary budget
754 during haze indicates that a large fraction of the peak-driving mass is combustion-derived and



755 can respond to measures that reduce burning intensity and solid-fuel use, particularly when
 756 these measures are timed to coincide with the high-risk stagnation window.



757 **Figure 7:** Average diurnal variation of mass concentrations and mass fractions of primary and
 758 secondary C-PM_{2.5} by different haze periods. Data is for the full period from 25th October 2023 to
 759 15th November 2023.

760

761 Secondary components ($\text{SO}_4^{2-} + \text{NO}_3^- + \text{NH}_4^+ + \text{Cl}^- + \text{LVOOA} + \text{SVOOA}$) form through
 762 atmospheric oxidation and multiphase processes and can remain elevated for longer periods,
 763 especially under haze-favorable conditions. During peak haze, the secondary fraction (SOA +
 764 secondary inorganics) often dominates C-PM_{2.5} mass, frequently exceeding ~60%. While daytime
 765 photochemistry can contribute to secondary production, winter/post-monsoon haze in the IGP
 766 is also strongly shaped by nighttime partitioning and aqueous/multiphase reactions under high
 767 RH, allowing secondary mass to remain elevated even when direct radiation is reduced. This
 768 contributes to a positive feedback: stagnant meteorology traps primary emissions, which provide
 769 both condensational sink and reactive surfaces, promoting secondary growth and further
 770 reducing visibility. The secondary to primary ratio (Figure S13) captures this behavior: it tends to
 771 decrease during morning/evening emission peaks and increase during midday when dilution
 772 reduces primary components while secondary material persists or continues to form. In more



773 humid conditions, this ratio can remain elevated even at night, indicating continued secondary
774 processing.

775 From a mitigation perspective, the sustained dominance of secondary material during the
776 most polluted periods implies that controlling only direct primary emissions will not fully
777 suppress peak $PM_{2.5}$ unless precursor emissions and the regional aged background are also
778 reduced. Conversely, if secondary components dominate, mitigation efforts focused solely on
779 primary emissions may be insufficient to reduce peak pollution levels because secondary
780 formation and partitioning can maintain a large mass burden even when local primary increments
781 are reduced. The Sonipat observations therefore support integrated strategies that reduce both
782 regional combustion emissions (which supply primary mass and precursors) and the conditions
783 that enable rapid accumulation, including targeted actions during predicted stagnation episodes
784 when boundary-layer suppression maximizes the sensitivity of concentrations to emissions.

785 **3.4 Biomass Burning Influence on Haze Severity**

786 Fire counts (Figure S12) increase sharply during haze periods, indicating strong
787 coupling between regional biomass burning and extreme pollution at Sonipat. To quantify how
788 biomass-burning-related factors scale with pollution severity, we binned C- $PM_{2.5}$ into five
789 concentration ranges from low ($<50 \mu\text{g m}^{-3}$) to extreme ($>200 \mu\text{g m}^{-3}$) and evaluated the
790 fractional contributions of biomass burning (BBOA + SFCOA + eBC_{bb}), fossil fuel combustion (HOA
791 + eBC_{ff}), secondary inorganic aerosols ($SO_4^{2-} + NO_3^- + NH_4^+ + Cl^-$), and secondary organic aerosol
792 (LVOOA + SVOOA) (Figure S14). Biomass-burning contributions increase from ~21% under low
793 pollution to ~34% under extreme pollution ($>200 \mu\text{g m}^{-3}$), indicating that intensified
794 biomass/solid-fuel influence is a key driver of the highest PM loadings. Fossil-fuel contributions
795 remain comparatively small (~4–10%) across all bins. SOA remains persistently large (~32–44%)
796 across pollution levels, while secondary inorganic aerosol contributions vary less strongly.

797 The persistence of a large SOA fraction even at extreme pollution suggests that regional
798 combustion emissions supply abundant precursors and aging products (including dark aging and
799 transported oxidized OA), sustaining a substantial secondary burden (Lalchandani et al., 2022).
800 Notably, secondary aerosol (SOA + secondary inorganics) still accounts for ~56% of C- $PM_{2.5}$ even

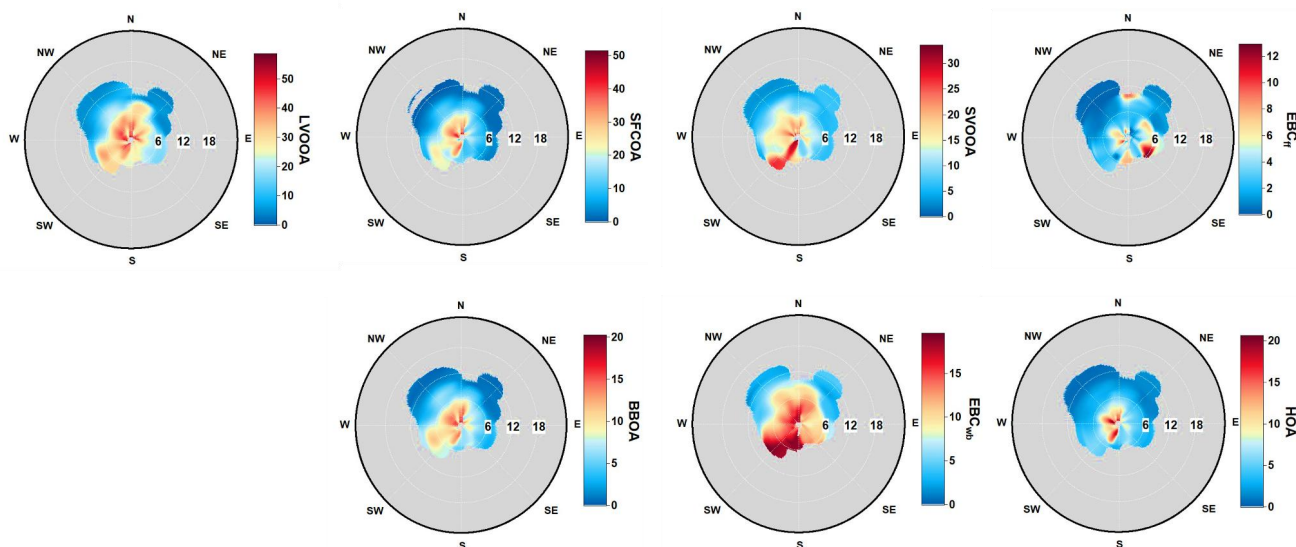


35

801 during extreme pollution, underscoring that haze severity is not solely a function of primary
 802 emissions but is strongly modulated by secondary mass growth and aging. This result has direct
 803 relevance for clean-air strategies because it implies that measures that only suppress immediate
 804 combustion activity may reduce the sharpest primary spikes but may leave a large secondary
 805 burden largely intact unless the broader regional emissions that feed secondary formation are
 806 also reduced. It also implies that actions aimed at reducing open burning can yield amplified
 807 benefits: reductions in biomass-burning emissions would simultaneously lower primary
 808 BBOA/eBC_{bb} and reduce the supply of reactive organic precursors that contribute to oxygenated
 809 OA during transport.

810

811 **3.5 Regional versus Local Influences: Trajectory and Wind-Based Source Diagnostics**



812 **Figure 8:** Estimation of source regions using NWR approach around the site using wind speed
 813 and wind direction for organics and black carbon factors during full period from 25th October
 814 2023 to 15th November 2023.

815

816 To distinguish regional transport from local influences, we integrate non-parametric wind
 817 regression (NWR), pollution rose (PR) diagnostics, and concentration-weighted trajectory (CWT)



818 analysis (Figure 8; Figure S10; Figure S11). Together, these complementary methods provide a
819 consistent framework for identifying dominant transport pathways and likely source regions, and
820 they help interpret whether high concentrations arise predominantly from local stagnation or
821 from coherent regional inflow.

822 The NWR results show that LVOOA and SVOOA exhibit strong northwest influence with
823 peak concentrations ($40\text{-}50\ \mu\text{g m}^{-3}$), consistent with regional transport along the post-monsoon
824 corridor. PR plots (Figure S10) show that LVOOA is elevated primarily under northwest flow at
825 moderate-to-high wind-speed percentiles (40-60+), supporting an interpretation of transported
826 aged aerosol rather than purely local stagnation. CWT further localizes the likely source influence
827 on the Punjab-Haryana agricultural belt, with hotspots of $15\text{-}25\ \mu\text{g m}^{-3}$ for LVOOA (Figure S11).
828 This spatial alignment with satellite-derived fire activity strengthens the inference that
829 transported agricultural burning emissions, and their processed products contribute substantially
830 to aged OOA at Sonipat (Gunthe et al., 2021). SVOOA shows broadly similar northwest
831 preference, but with a wider spatial footprint in CWT, consistent with a combination of formation
832 during transport and regionally distributed precursor sources. The presence of a southwest
833 influence for SVOOA in NWR/PR suggests an additional contribution from relatively fresher
834 emissions nearer to the site compared to LVOOA, consistent with the idea that semi-volatile
835 oxygenated aerosol can respond rapidly to local-to-regional precursor inputs and thermodynamic
836 partitioning.

837 SFCOA exhibits one of the strongest northwest directional signatures, with high
838 concentrations in NWR ($25\text{-}30\ \mu\text{g m}^{-3}$) at moderate wind speeds and PR enhancements extending
839 into high wind-speed bins (60-80+ percentile). CWT identifies SFCOA hotspots ($12\text{-}16\ \mu\text{g m}^{-3}$)
840 concentrated in Punjab and western Haryana, consistent with widespread residential solid-fuel
841 use and small-scale combustion emissions across rural/agricultural districts. These transport
842 signatures are consistent with trajectory-based estimates indicating that northwesterly winds
843 can deliver stubble-burning aerosol to the Delhi region within $\sim 15\text{-}51$ hours depending on source
844 distance (185-420 km) (Bikkina et al., 2019; Cusworth et al., 2018). BBOA also shows northwest
845 influence, but with a somewhat broader spatial distribution, consistent with contributions from
846 both intense agricultural burning and more distributed residential biomass burning. The



847 coherence of SFCOA and BBOA regional signatures reinforces that what is often labeled as Delhi
848 haze is, in substantial part, the manifestation of a broader IGP combustion plume that is already
849 heavily polluted and chemically evolved before reaching the city (Sharma et al., 2020).

850 The BC apportionment reinforces this regional narrative. eBC_{bb} shows strong northwest
851 directionality in NWR ($15\text{-}20\ \mu\text{g m}^{-3}$), increasing concentrations under moderate-to-high wind
852 speeds in PR, and distinct CWT hotspots ($7\text{-}10\ \mu\text{g m}^{-3}$) co-located with SFCOA and BBOA source
853 regions. The tri-method agreement (NWR direction, PR wind-speed dependence, and CWT
854 geographic localization) provides robust evidence that biomass burning dominates BC at Sonipat
855 during this period. In contrast, eBC_{ff} is comparatively small ($5\text{-}10\ \mu\text{g m}^{-3}$), exhibits a centre-
856 weighted NWR pattern, is associated mainly with lower wind-speed bins in PR, and shows
857 weak/limited spatial structure in CWT ($<3\ \mu\text{g m}^{-3}$). This indicates predominately local fossil-fuel
858 influence with minimal regional transport signatures. The strong contrast between eBC_{bb} and
859 eBC_{ff} at Sonipat is consistent with its role as an intermediate receptor site, where transported
860 biomass-burning plumes are intercepted before dilution by Delhi's urban fossil fuel emissions,
861 and it suggests that reductions in regional biomass burning could produce immediate downwind
862 benefits by lowering the incoming BC burden that contributes to both exposure and aerosol
863 absorption.

864 HOA remains low ($5\text{-}10\ \mu\text{g m}^{-3}$) and displays center-weighted NWR behavior, broad
865 directional spread in PR at low wind speeds, and minimal spatial structure in CWT ($<3\ \mu\text{g m}^{-3}$),
866 indicating that HOA is dominated by local traffic with negligible long-range transport. This differs
867 from many Delhi urban studies where HOA is higher and transport contributions to $PM_{2.5}$ can be
868 substantial (Pant et al., 2015). The weak regional signature in HOA at Sonipat provides an
869 important contrast: the regional plume arriving at Sonipat is strongly biomass/solid-fuel
870 dominated, and traffic-related emissions become more important only after mixing within the
871 Delhi urban environment. This upwind-downwind contrast supports the interpretation that
872 policies targeting only urban traffic will not substantially reduce the regional carbonaceous plume
873 that sets the baseline for Delhi-NCR haze, whereas coordinated actions addressing agricultural
874 burning and solid-fuel combustion across the corridor are likely to be more effective for reducing
875 both upwind and downwind exposure.



876 Overall, the integrated NWR-PR-CWT evidence indicates that the Sonipat site experienced
877 overwhelming regional influence during the post-monsoon period, with Punjab-Haryana
878 agricultural regions consistently identified as dominant source areas for aged, oxygenated OA,
879 solid-fuel combustion OA, and biomass-derived BC. Secondary OA (LVOOA + SVOOA) frequently
880 reached combined levels approaching $80\text{--}100\ \mu\text{g m}^{-3}$ during intense haze, and biomass-related
881 combustion factors (SFCOA, BBOA, eBC_{bb}) exhibited coherent transport signatures and
882 geographically consistent hotspots. At the same time, HOA remained low and locally influenced.
883 These results establish Sonipat as a sensitive receptor for transboundary biomass-burning
884 emissions and their processed products before they mix with Delhi's urban emissions,
885 underscoring that effective mitigation of post-monsoon haze across Delhi-NCR requires
886 addressing regional combustion sources along the broader IGP transport corridor and aligning
887 short-term actions with meteorological forecasts that identify high risk stagnation periods.

888

889 **4. Summary and Conclusions**

890 This study presents the first comprehensive, high-time-resolution source apportionment
891 of carbonaceous PM_{2.5} at Sonipat, an upwind receptor site located ~40 km northwest of Delhi
892 along the principal post-monsoon transport corridor linking the Punjab-Haryana agricultural
893 burning regions with the Delh-NCR, providing detailed insights into the chemical evolution and
894 source dynamics that drive severe haze episodes across the Indo-Gangetic Plain. By integrating
895 ToF-ACSM measurements, Aethalometer-based black carbon apportionment, Positive Matrix
896 Factorization (PMF), and trajectory- and wind-based diagnostics, we quantified the relative
897 contributions of primary combustion emissions, secondary processing, and meteorological
898 controls in shaping haze evolution prior to mixing with urban emissions.

899 Two major haze episodes during 25 October–15 November 2023 were characterized by
900 composition-based PM_{2.5} exceeding $300\ \mu\text{g m}^{-3}$. Organic aerosol (OA) dominated the submicron
901 mass, accounting for ~65% of NR-PM_{2.5}, with daily mean OA peaking near $140\ \mu\text{g m}^{-3}$ during the
902 first haze episode. PMF resolved five factors, namely, hydrocarbon-like OA (HOA), biomass-
903 burning OA (BBOA), solid-fuel combustion OA (SFCOA), and two oxygenated components (SVOOA
904 and LVOOA). Secondary organic aerosol (SVOOA + LVOOA) constituted ~57-60% of OA mass



905 across the campaign, with LVOOA alone representing the largest fraction (mean $25 \mu\text{g m}^{-3}$) and
906 reaching $42.8 \mu\text{g m}^{-3}$ during the most intense haze period, indicating substantial regional aging
907 and multi-day accumulation. Primary combustion related components also increased sharply
908 during haze, with SFCOA and BBOA showing strong enhancements relative to non-haze
909 conditions. Black carbon apportionment revealed a pronounced biomass-burning dominance,
910 with biomass-derived BC (eBC_{bb}) contributing $\sim 78\%$ of total eBC (mean $10.9 \mu\text{g m}^{-3}$), substantially
911 exceeding fossil-fuel BC ($3.1 \mu\text{g m}^{-3}$). During the peak haze episode, eBC_{bb} reached $14.2 \mu\text{g m}^{-3}$.
912 This reversed BC partitioning relative to urban Delhi confirms that the upwind plume intercepted
913 at Sonipat is strongly influenced by regional biomass and solid-fuel combustion rather than
914 traffic-dominated emissions.

915 The transition from non-haze to haze conditions reflects amplification of a persistently
916 elevated regional carbonaceous background. Even during non-haze periods, oxidized OA (LVOOA
917 $\sim 11\text{-}13 \mu\text{g m}^{-3}$) remained substantial, indicating sustained regional processing across the Indo-
918 Gangetic Plain. During haze, sharp increases in SFCOA and BBOA were superimposed on this aged
919 baseline, driving extreme mass loadings. Notably, secondary components (SOA plus secondary
920 inorganics) continued to account for $\sim 56\text{-}60\%$ of C- $\text{PM}_{2.5}$ even under extreme pollution ($>200 \mu\text{g}$
921 m^{-3}), demonstrating that haze severity arises from the combined effects of direct combustion
922 emissions, abundant precursors, and meteorologically driven secondary growth. Distinct
923 differences between haze episodes further suggest that prolonged stagnation favors
924 accumulation of highly aged LVOOA, whereas shorter duration emission pulses (e.g., Diwali) are
925 associated with comparatively fresher oxygenated fractions.

926 Independent trajectory and wind-regression analyses consistently identified northwestern
927 transport from Punjab-Haryana agricultural districts as the dominant source influence during
928 polluted periods. LVOOA, SVOOA, SFCOA, and eBC_{bb} exhibited coherent northwest signatures
929 and concentration-weighted trajectory hotspots aligned with agricultural burning regions,
930 whereas HOA remained low and largely local in origin. These findings establish Sonipat as a
931 regional receptor site intercepting a combustion-dominated plume prior to mixing with Delhi's



40

932 urban emissions and demonstrate that extreme carbonaceous loading can develop outside the
933 urban core.

934 The results of this study show that post-monsoon haze over the Delhi-NCR is strongly shaped by
935 regionally transported biomass and solid-fuel combustion emissions and by sustained secondary
936 organic aerosol formation under stagnant meteorological conditions. A substantial fraction of the
937 pollution burden affecting Delhi is therefore already pre-formed before entering the city.
938 Effective mitigation of severe haze requires coordinated, airshed-scale emission reductions
939 targeting agricultural residue burning, rural solid-fuel combustion, and regional precursor gases.
940 Strategies focused solely on urban traffic or primary emissions are unlikely to sufficiently reduce
941 peak pollution if secondary formation remains dominant. Integrated regional controls, aligned
942 with meteorological forecasts of stagnation, are essential for achieving durable reductions in
943 PM_{2.5} exposure across the broader Indo-Gangetic Plain and the Delhi-NCR.

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949 **Data availability**

- 950 • The observational datasets used in this study and meteorological parameters from the
951 automatic weather station (AWS) installed at the Sonipat site (Rathore et al., 2025), are
952 available at <https://doi.org/10.5281/zenodo.15902605>.
- 953 • The fire counts data can be accessed from <https://firms.modaps.eosdis.nasa.gov/>.
- 954 • ZeFir is used for wind and trajectory-based
955 analysis(<https://sites.google.com/site/zefirproject>).
- 956 • All plots in this study were generated using Python
957 (<https://www.anaconda.com/download>).



41

958 **Author Contributions:**

959 **Conceptualization:** VS, DG, SD

960 **Data curation:** VS, JR

961 **Investigation, Methodology:** VS, DG,SD, JR

962 **Software, Visualization:** VS

963 **Writing – original draft:** VS

964 **Writing – review & editing:** DG, JR, SD, RKK, LS

965 **Resources-** DG, SG, SD, RKK, LS

966 **Competing Interest**

967 The authors declare no conflicts of interest relevant to this study.

968

969 **5. References**

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