Vertical Structure and Driving Mechanism of $PM_{2.5}$ and PM_{10} Aerosols in Hefei Based on LiDAR Observations (2021–2023)

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Abstract:

Aerosol pollution remains a significant environmental concern in China. However, the vertical structure and evolution of particulate matter are poorly understood due to the lack of long-term, high-resolution observations. In Hefei, the aerosols during the study period were dominated by a mixture of fine particulate matter ($PM_{2.5}$) and coarse particulate matter (PM_{10}), mainly originating from urban traffic emissions, industrial activities, and regional transport, with significant contributions from secondary inorganic aerosols and occasional dust events. To address the knowledge gap in aerosol vertical distribution during different pollution episodes, this study employed an aerosol LiDAR system with 532 nm band to investigate the vertical profile characteristics of aerosols, with a focus on comparing the stratification differences of optical properties between $PM_{2.5}$ and PM_{10} pollution events over Hefei ~~across different periods and altitudes~~. The seasonal and diurnal variations of aerosol profiles were investigated~~, and vertical structures were compared~~ on polluted and clean days. The relationship between near-surface particulate matter concentrations and aerosol vertical properties ~~aerosol stratification~~ was analyzed at different heights, alongside the dynamic evolution of aerosol layers during typical pollution events. Our results demonstrated that the extinction coefficient (532 nm) of $PM_{2.5}$-polluted days below 0.6 km was approximately three times that of PM_{10}-polluted days.

In contrast, the depolarization ratio of PM_{10}-polluted episodes remains consistently higher than that of $PM_{2.5}$-polluted cases throughout the entire observed altitude range. The differences in extinction between polluted and clean days for $PM_{2.5}$ were most pronounced below 0.9 km and subsequently decreased as altitude increased, whereas the differences in PM_{10} remained significant below 1.2 km. For $PM_{2.5}$, the strongest enhancement appeared between 7:00 and 14:00 (Beijing time, BJT). A subtle lifting with height was observed around midday. PM_{10}-polluted days were characterized by a greater vertical extension of high aerosol extinction (~~reaching~~ up to ~1.2–1.4 km) but a shorter duration of strong extinction~~, . in~~ In contrast, ~~to~~ $PM_{2.5}$-polluted days~~, which~~ exhibited a ~~more~~ persistent but vertically confined aerosol layer~~. PM_{10} pollutant tended to accumulate within the altitude range of 0.4–1.2 km on polluted days~~. The vertical wind shear (VWS) was weaker on $PM_{2.5}$-polluted days compared to clean days. On PM_{10}-polluted days, the VWS in the near-surface layer (1000–900 hPa) was significantly stronger than that on clean days, especially during the early morning and evening periods. The $PM_{2.5}$ pollution in Hefei was mostly contributed by temperature inversion and high relative humidity, while PM_{10} pollution was driven by long-range transport of aerosol particles under the cold front system and dry conditions. These findings reveal ~~highlight the~~ complex interactions between aerosol optical properties, boundary-layer dynamics, and synoptic ~~scale meteorology~~ conditions, providing new insights into the vertical ~~processes governing~~ air quality processes in eastern China.

Keywords: Aerosol LiDAR, Vertical profile, Heavy pollution episodes, Vertical wind shear, Fine and coarse particulate matter.

1. Introduction

Air pollution remains one of the most pressing environmental challenges globally, with fine particulate matter and inhalable coarse particles posing serious risks to public health and atmospheric visibility (Chen et al., 2023; Deng et al., 2023). Air quality is closely related to the concentration of pollutants suspended in the atmosphere (Wang et al., 2024). Solid and liquid phase pollutants, known as aerosols, not only impact air quality and visibility through multiple mechanisms but also affect the climate by altering Earth's radiation budget and water cycle processes (Chen et al., 2016; Miao et al., 2018). In China, substantial improvements in air quality have been achieved in recent years through stringent emission control policies. However, severe particle pollution events still occur frequently, especially in winter haze episodes and spring dust storms across the Yangtze River Delta (YRD) region (Han and Cao 2022; Wang and Wang 2021). To date, most studies have focused on surface-level air ~~pollution data~~quality~~-~~, with particular attention to individual $PM_{2.5}$ and PM_{10} event in eastern China and urban agglomerations~~obtained~~ from ground monitoring networks (He et al., 2021). These data provide critical insights into near-surface concentrations. However, they lack information on the vertical distribution and formation mechanisms of pollutants, which is essential for

understanding aerosol processes in the troposphere (Mehta et al., 2021; Mishra and Shibata 2012; Wang et al., 2018).

While extensive ground-based monitoring has provided insights into surface-level pollution, the understanding of the vertical distribution and evolution of aerosols remains limited due to a lack of long-term and high-resolution vertical observations (He et al., 2022; Ou et al., 2021; Shen et al., 2022). The atmospheric boundary layer (ABL) plays a crucial role in regulating aerosol dynamics through physical mechanisms such as turbulence, convection, and mixing (Gao et al., 2011; Garratt 1994; Tombrou et al., 2007). As the interface between the surface and the free troposphere, the ABL governs the vertical exchange of energy, moisture, and pollutants. Numerous studies have demonstrated that lower ABL heights could trap pollutants near the surface, leading to elevated $PM_{2.5}$ levels. These lower heights are often associated with thermal inversions and stagnant synoptic conditions (Liu et al., 2020; Sun et al., 2024a). Under such conditions, limited mixing suppresses the vertical dispersion of aerosols, allowing surface emissions to accumulate rapidly, particularly in urban areas with high anthropogenic activity. In contrast, strong vertical mixing and higher ABL heights enhance dispersion and dilution of pollutants, which frequently results in improved surface air quality (Jin et al., 2021). Moreover, interactions between the lower ABL and the overlying free troposphere, including vertical wind shear, subsidence, and entrainment processes, also significantly influence aerosol layering and transboundary transport (Deng et al., 2023; Li et al., 2022). These interactions are significant during transition periods such as the morning boundary layer growth phase or evening collapse, which strongly affect aerosol vertical distribution (Li et al., 2018). Despite these known mechanisms, detailed observations of the vertical structure of aerosols and their relation to meteorological dynamics remain limited under varying pollution conditions and across different seasons (Yang et al., 2025). Comprehensive long-term observations are still needed to characterize how meteorology influences aerosol stratification and transformation over time.

Active remote sensing techniques, such as aerosol Light Detection and Ranging (LiDAR), have emerged as powerful tools to fill this observational gap (Ansmann et al., 2013; Chen et al., 2024a). Unlike passive satellite-based instruments, LiDAR systems provide high-resolution vertical profiles of aerosol optical properties at fine temporal scales and under both day and night conditions, capable of capturing sub-kilometer vertical gradients and diurnal variability that are critical for understanding boundary-layer dynamics (Wang et al., 2024; Zhang et al., 2020b). It enables continuous monitoring of aerosol structure, boundary layer formation, and pollutant layering in response to atmospheric dynamics (Fan et al., 2024; Fang et al., 2024; Li et al., 2024). Specifically, polarization-sensitive aerosol LiDAR can simultaneously retrieve extinction coefficients and depolarization ratios, which serve as indications of aerosol concentration and morphology (Cairo et al., 2024; Chen et al., 2024a; Kumar et al., 2024). The extinction coefficient quantifies the total attenuation of light caused by scattering and absorption by particles, and it is directly

related to aerosol optical depth and visibility (Chen et al., 2024b; Sun et al., 2024b). The depolarization ratio distinguishes the spherical and non-spherical particles, providing insights into aerosol types such as secondary inorganic aerosols, biomass burning smoke, and desert dust. These parameters are crucial for understanding not only the optical but also the physical properties of aerosols (Chen et al., 2024b; Gebauer et al., 2024). However, most previous studies using such measurements have been limited to short-term campaigns or case studies of individual pollution episodes (Chen et al., 2022; Zhong et al., 2018). These attempts often lack the long-term temporal continuity and seasonal representativeness necessary~~spatial and temporal representativeness necessary~~ for climatological or process-level understanding, particularly in regions subject to strong seasonal and meteorological variability.

Long-term monitoring of air pollutants is essential for the comprehensive understanding of their temporal characteristics (Fan et al., 2021; Xiang et al., 2021). Unlike ground-based air quality monitoring networks, vertically resolved data are required to evaluate the entire process of pollution events, vertical mixing, and potential long-range transport (Wang et al., 2024). Passive satellite remote sensing techniques have proven useful in bridging spatial gaps in air quality data, but are limited in their ability to resolve vertical structures with high temporal accuracy due to their coarse resolution and fixed overpass times (Chen et al., 2023). Chemical transport models (CTMs) have been widely used to simulate the spatiotemporal behavior of aerosols and their interactions with meteorology (Wang et al., 2025; Xiong et al., 2023). However, their accuracy is constrained by uncertainties in emissions inventories, boundary conditions, meteorological input fields, and the representation of aerosol microphysics. In addition, model performance varies significantly by region and pollution type (Zhan et al., 2024). LiDAR technology provides high temporal resolution and fine vertical resolution measurements~~spatiotemporal resolution vertical measurements~~ of aerosol and meteorological variables, including aerosol extinction, backscattering, and wind profiles (Chen et al., 2023; Chouza et al., 2015; Zhang et al., 2020b). This capability makes LiDAR particularly suitable for investigating aerosol evolution and transport within the ABL and lower troposphere. Nonetheless, long-term time series analyses based on continuous LiDAR measurements remain limited in the literature.

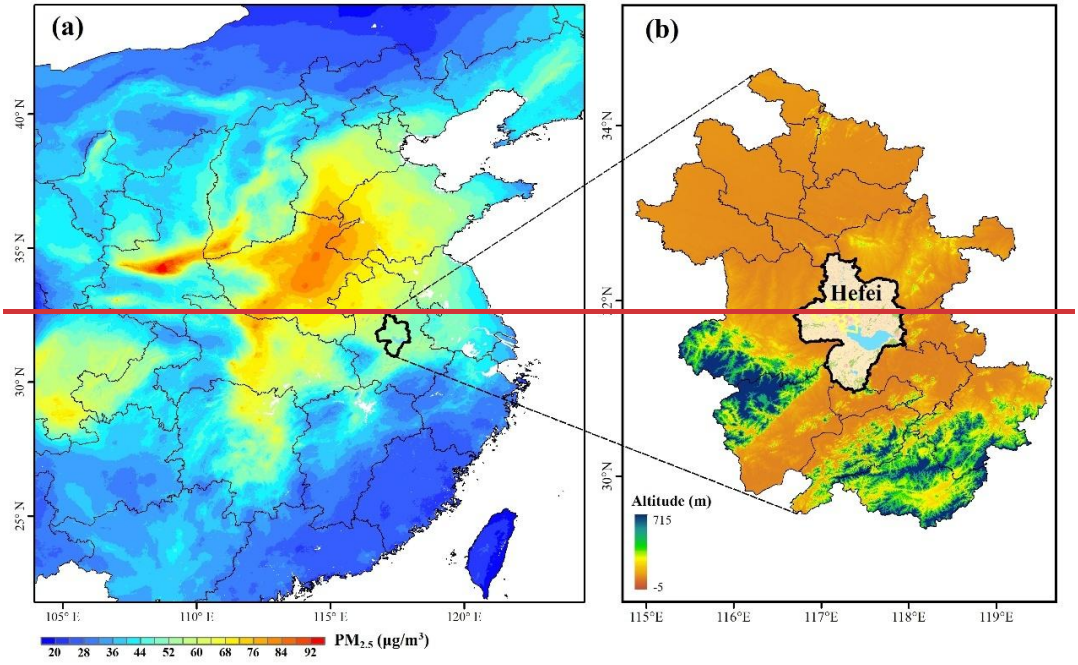
To address these limitations, we conducted multi-year observations using an aerosol LiDAR system in Hefei. The deployed LiDAR system provides real-time vertical profiles of aerosol extinction and depolarization ratios, allowing for a detailed assessment of aerosol structure under different pollution levels. This study fills the observational gap in the long-term vertical characteristics of both fine and coarse particulate pollution over this region. We characterized the vertical and temporal evolution of aerosol properties and linked them with pollution episodes and boundary-layer processes by integrating LiDAR data with surface $PM_{2.5}$ and PM_{10} concentrations, meteorological observations, and reanalysis products. Therefore, the objectives of this study were (1) to analyze the

seasonal and diurnal variations of vertical profiles with aerosol extinction and depolarization data; (2) to compare the vertical structures on polluted and clean days to clarify the differences in aerosol optical properties between these two conditions; (3) to assess the relationship between aerosol layering and near-surface pollutant concentrations at different heights; and (4) to investigate the spatiotemporal evolution of aerosols during specific pollution episodes and explore the mechanisms of their vertical structure, meteorological conditions, and accumulation process. The findings are expected to enhance our understanding of aerosol vertical structure in eastern China and provide scientific support for air quality management and modeling efforts.

2. Data and methods

2.1 Study region

Hefei occupies a strategic location, bridging the Beijing-Tianjin-Hebei metropolitan cluster and the Yangtze River Delta economic hub. The study area lies in a transitional climatic zone, where temperature inversions occur frequently in winter. Such conditions often result in stagnant air that limits pollutant dispersion. In addition, the geography of Hefei plays a critical role in exacerbating air quality challenges. Hefei is surrounded by the Dabie Mountains in the west and the Huangshan ranges in the south. The open plains dominate its northern and eastern frontiers (**Fig. 1b**). These natural barriers hinder the southward dispersion of airborne pollutants while trapping those transported from industrialized northern regions within the basin. The unique combination of geographic constraints and atmospheric conditions has made Hefei a focal point for studying complex pollution mechanisms. Researchers here focus particularly on the synergistic effects of regional transportation, local emissions, and meteorological drivers on particulate matter formation (Fang et al., 2024; Huang et al., 2016; Liu et al., 2024).



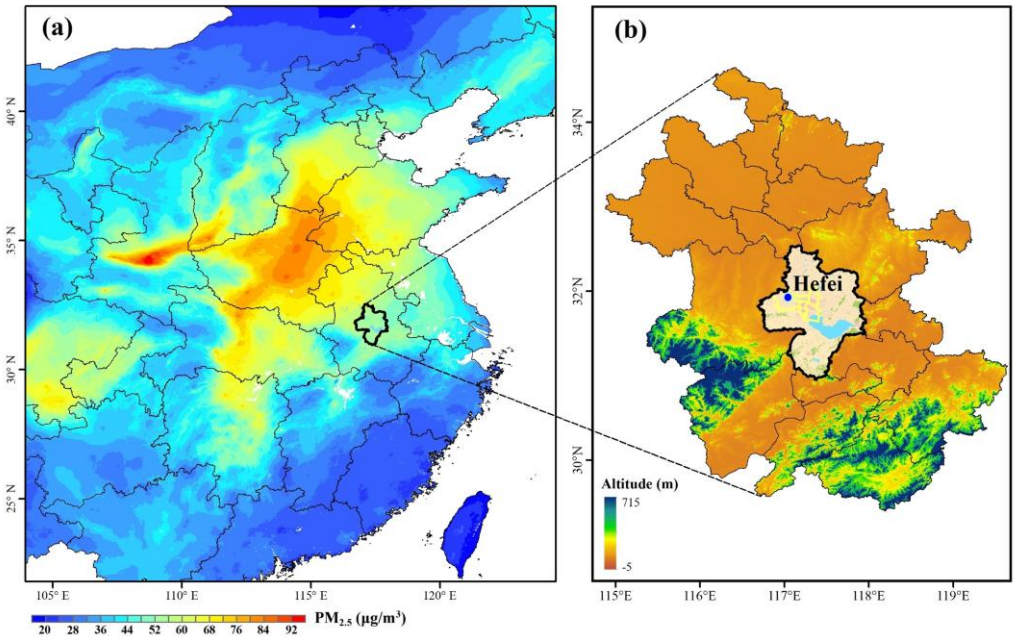


Figure 1. The spatial distribution of (a) the averaged $PM_{2.5}$ concentration during winter from 2021 to 2023(Ministry of Environmental Protection in China, https://air.cnemc.cn:18007/) and (b) the location of the study area (Hefei, China). The digital elevation model (DEM) data are derived from the NASA Shuttle Radar Topography Mission (SRTM) 30 m product (https://glovis.usgs.gov/), and the land use map of the study area is obtained from the 30 m resolution annual China Land Cover Dataset (CLCD) dataset. The monitoring station is marked by the blue dot.

2.2 Aerosol LiDAR data

The LGJ-05 aerosol LiDAR combines traditional radar technology and modern laser technology ~~with 355 nm,~~ with 532 nm~~, and 1064 nm~~ as detection light sources. Measurements at 532 nm were utilized exclusively for ~~the~~ ~~present~~this study. The LiDAR emits ~~532 nm~~ a laser pulse~~s~~ into the atmosphere to capture backscattered signals, thereby obtaining key aerosol optical characteristics. As a polarization-sensitive system, it can retrieve extinction coefficients and depolarization ratios to characterize aerosols in the atmosphere. During the transmission process, the laser pulse is scattered and extinguished by atmospheric aerosol particles. The intensity of backscattered light at different altitudes correlates with the scattering and extinction properties of aerosol and cloud particles at those altitudes. The backscattered light from spherical particles can maintain its polarization characteristics of the emitted pulse since the emitted laser pulse is linearly polarized. In contrast, non-spherical particles (such as dust particles and ice crystals in cirrus clouds) depolarize the backscattered light due to their irregular shapes and asymmetric interaction with linearly polarized laser pulses. The LiDAR could detect the echo signals of the parallel and vertical components in the backscattered light, enabling acquisition of the vertical profile of the depolarization ratio of atmospheric aerosol particles. It could enable three-dimensional monitoring of the atmosphere in real time through active remote sensing, with the capacity to invert spatiotemporal distribution information such as aerosol extinction, depolarization, and water vapor mixing ratio in the atmosphere. The aerosol extinction coefficient reflects the attenuation capacity of aerosol particles to incident light, while the depolarization ratio indicates the non-spherical nature of aerosol and cloud particles. A higher extinction coefficient typically indicates a larger concentration of aerosol particles or stronger light-scattering/absorbing properties. However, greater depolarization values denote a higher abundance of non-spherical aerosol particles (e.g., dust, rugged industrial aerosols) or mixed-phase cloud constituents. By contrast, lower depolarization ratios indicate a dominance of spherical particles or simpler particle structures.

~~The LGJ-05 aerosol LiDAR accomplishes all these operations using a laser as the light source. The laser system generates high-energy light beams at wavelengths of 532 nm, 355 nm, and 1064 nm. The three-wavelength laser beams are emitted into the atmosphere through a broadband mirror and undergo Mie scattering with aerosols.~~ Since the instrument is capable of receiving data below 30 km, the effective detection range of the aerosol LiDAR can cover 0.2-6 km and 0.2-15 km for daytime and nighttime, respectively. It offers a vertical spatial resolution of 7.5 meters and a temporal resolution of 10 minutes. The observations were conducted at the National Meteorological Observation Station in the northwest of Hefei (117.06°N, 31.96°E) from March 2021 to May 2023.

2.3 Ground-based observational data

Hourly concentrations of ~~major air pollutants (~~$PM_{2.5}$ and PM_{10}~~)~~ were obtained from ground-based monitoring stations in Hefei, which are operated and maintained by the China National Environmental Monitoring Centre (CNEMC) (Liu et al., 2017). These pollutant data were obtained from professional monitoring instruments, with the LGH-~~01B~~ 01E aerosol mass concentration monitor used for PM_{10} and the LGH-01B ~~aerosol mass concentration monitor used~~ for $PM_{2.5}$, both of which apply the beta attenuation method. All instruments are calibrated regularly according to national standards, and the data undergo strict quality control procedures, including hourly, daily, and annual audits, as described in the China Environmental Monitoring Quality Assurance and Quality Control Manual. It should be noted that we focused on $PM_{2.5}$ and PM_{10} concentrations during ~~the~~ winter and spring ~~seasons~~, respectively. All comparisons between $PM_{2.5}$ and PM_{10} pollution episodes were conducted across both seasons. The hourly resolution of the dataset allows for capturing diurnal variations in pollutant levels, while the multi-pollutant data facilitates analysis of the differences in meteorological responses of pollutants in Hefei.

The observed meteorological variables were obtained from the China Meteorological Administration (CMA) (http://data.cma.cn/en), measured by the Vaisala PTB210 Digital Barometer for atmospheric pressure, the Vaisala HMP155A Temperature and Relative Humidity Probe for air temperature (T) and relative humidity (RH), the EL15-2C Wind Direction Sensor for wind direction (WD), and the EL15-1C Wind Speed Sensor for wind speed (WS). All meteorological instruments are routinely maintained and calibrated to ensure the accuracy and reliability of observational data.

2.4 ECMWF reanalysis data

This study aims to investigate the impact of the synoptic system on $PM_{2.5}$ and PM_{10} pollution. The ERA5 reanalysis dataset, freely accessible via the Copernicus Climate Change Services platform (https://cds.climate.copernicus.eu/datasets), serves as the data source. Vertical atmospheric data within the ERA5 dataset are interpolated to 37 pressure levels. This interpolation yields comprehensive data ranging from the Earth's surface to the upper atmosphere. Given its high spatial and temporal resolution, the ERA5 dataset has been widely employed in extreme weather events, climate prediction, and air pollution studies (Fan et al., 2021; Zhang et al., 2020b).

This study collected hourly geopotential height, vertical wind velocity, temperature, and humidity data. The data correspond to multi-pressure levels (500 hPa, 700 hPa, 725 hPa, 750 hPa, 775 hPa, 800 hPa, 825 hPa, 850 hPa, 875 hPa, 900 hPa, 925 hPa, 950 hPa, 975 hPa, and 1000 hPa) ~~and cover the period from 2021 to 2023~~.~~.~~ The spatial

resolution of our dataset was 0.25°×0.25° (zonal and meridional) and supports a detailed and accurate exploration of the synoptic-particulate matter relationship. The data were extracted for winter and spring during 2021–2023, consistent with the periods of the surface pollutant and aerosol LiDAR observations.

2.5 Quality control

To explore the potential of aerosol LiDAR observations and improve data reliability, this study extracted the extinction coefficient and depolarization ratio data at 532 nm from the LiDAR system for the period from March 2021 to May 2023. In the quality control process, specific criteria were established to ensure the reliability of the retrieved parameters. Specifically, it covers three steps for outlier detection, removal of spurious points, and temporal consistency analysis. Outlier detection was conducted to eliminate all records outside the normal value range according to the parameter intervals provided by the instrument, with specific thresholds applied based on the signal-to-noise ratios (SNR) of the parallel (P) and perpendicular (S) channels. The extinction coefficient at 532 nm was calculated using data from the parallel (P) channel, which was considered reliable only when the corresponding signal-to-noiseSNR ratio (SNR) exceeded three. The depolarization ratio was derived from the P and perpendicular (S) channels, and data were accepted only when both channels had SNRs greater than three. The P channel detects linearly polarized light in the parallel direction, while the S channel detects linearly polarized light in the perpendicular direction. Based on the time series of SNR records, the individual extinction and depolarization data were cross-referenced. Temporal consistency analysis was conducted by identifying records that deviated by more than three standard deviations from the mean as outliers. The original LiDAR data have a 10-minute temporal resolution. After quality control, these data were averaged to hourly resolution to be consistent with the hourly ground-based pollutant observations. To further improve data reliability, this study calculated the credibility ratio of the extinction coefficient and depolarization ratio at different altitudes. Based on the criterion of a 60% valid data availability rate, the effective detection ranges for the 532 nm of the extinction coefficient and depolarization ratio were 0.2025~1.815 km.

2.6 Methodology

2.6.1 Definition of polluted and clean

In this study, the data from 2021-2023 are gathered and categorized into four seasons: spring (Mar.-May), summer (Jun.-Aug.), autumn (Sep.-Nov.), and winter (Dec.-Feb.). Given that $PM_{2.5}$ pollution is most severe in winter and sand-dust events are frequent in spring, the definitions of polluted and clean days in this study primarily focus

on these two seasons. Days in winter when the daily average concentration of $PM_{2.5}$ exceeds 75 μg/m³ are defined as $PM_{2.5}$-polluted days. Conversely, days in winter with a daily average $PM_{2.5}$ concentration below 50 μg/m³ are classified as $PM_{2.5}$-clean days. For PM_{10}, spring days with daily average concentrations above 150 μg/m³ are defined as polluted days, while those below 100 μg/m³ are considered clean days (Ministry of Environmental Protection of the People's Republic of China, Technical Regulation on Ambient Air Quality Index, 2012). To eliminate the wet deposition effect of precipitation, all observations in this study excluded precipitation daily data.

2.6.2 Calculation of Vertical Wind Shear

The vertical wind shear (VWS) is a crucial factor in characterizing the change in wind velocity with height. It is defined ~~based on~~as the variation~~magnitude~~ of wind vector~~the difference in the bulk wind vector~~ across atmospheric layers~~between two vertical levels~~, quantified by the magnitude of the difference between upper and lower levels~~normalized by the height difference~~. The VWS can be calculated using the following formula:

$$\mathrm{VWS} = \frac{\sqrt{(u_t - u_l)^2 + (v_t - v_l)^2}}{(z_t - z_l)} \times 1000 \quad (1)$$

Where VWS is measured in units of m/s·km^{-1}. u_t and u_l denote the zonal wind components at the upper z_t and lower z_l levels, respectively, while v_t and v_l represent the corresponding meridional wind components. This calculation aids in quantifying the vertical dynamic forces that may impact the dispersion and transportation of pollutants.

3. Results and discussion

3.1 Variations of inter-annual $PM_{2.5}$ and PM_{10}

Figure 2 shows the variations in the concentrations of $PM_{2.5}$ and PM_{10} pollutants from 2021 to 2023. Both $PM_{2.5}$ and PM_{10} pollutants exhibited a clear seasonal pattern, with concentrations higher in winter and lower in summer (Liu et al., 2021; Wang et al., 2024). In addition, PM_{10} concentrations were relatively high during the spring months of 2022 and 2023, likely driven by increased dust events such as sand-dust storms that commonly occurred during this season. High $PM_{2.5}/PM_{10}$ ratios ($\geqslant$0.5) during winter suggest a dominant contribution from fine particulate matter, whereas lower ratios ($\leqslant$0.5) in spring indicate an increased presence of coarse particles (Liu et al., 2015). The episodic PM_{10} peaks observed in spring 2022–2023 coincided with a pronounced decrease in the $PM_{2.5}/PM_{10}$ ratio in Hefei. This phenomenon reflected intensified pollution of coarse-mode aerosols. Although winter months also suffered elevated PM_{10} pollution, the concurrent high $PM_{2.5}$ concentrations complicated the separation of dust-related contributions. In contrast, the lower $PM_{2.5}/PM_{10}$ ratios in spring facilitated the distinct observation of dust events due to the reduced interference from fine particulate matter.

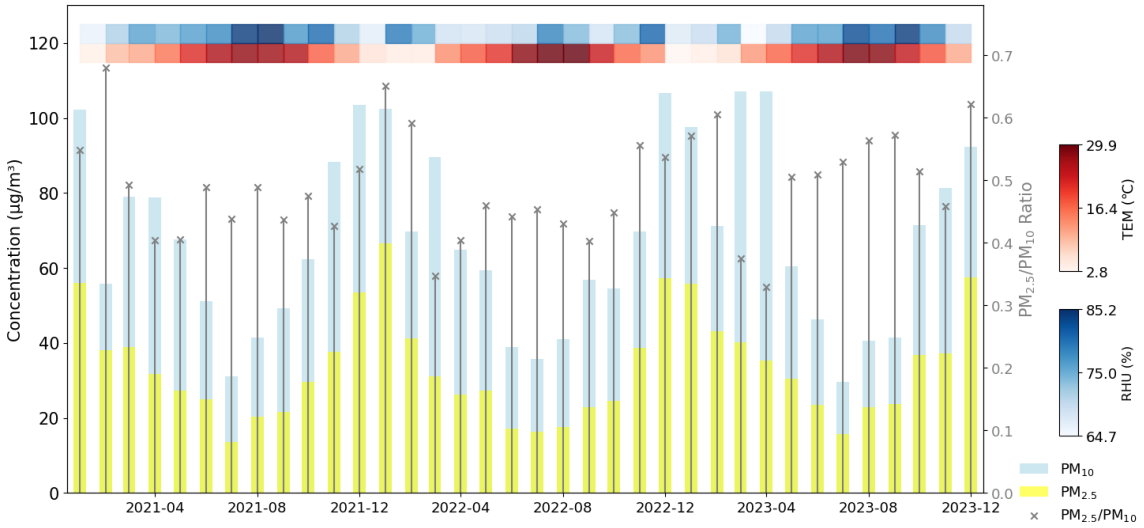


Figure 2. Monthly variation of $PM_{2.5}$ (marked as the yellow bars) and PM_{10} (marked as the blue bars) concentrations, $PM_{2.5}/PM_{10}$ ratio (the grey lines with markers on top), temperature (TEM), and relative humidity (RH) during 2021-2023 in Hefei (Beijing time, BJT = UTC + 8 h).

The hourly $PM_{2.5}$ and PM_{10} concentrations exhibited clear diurnal variation on both polluted and clean days (**Fig. 3**). Polluted days showed significantly higher $PM_{2.5}$ levels, with hourly averages ranging from 100 to 125 μg/m³, compared to the stable 35–55 μg/m³ range observed on clean days. The $PM_{2.5}$ concentrations on polluted days displayed a bimodal pattern. The concentrations remained consistently high throughout the morning period from 04:00 to 11:00 (Beijing time, BJT), with the peak at 10:00–11:00 (BJT) (Dai et al., 2020). Valleys in $PM_{2.5}$ concentrations occurred at 16:00 (BJT), possibly associated with enhanced atmospheric dispersion from midday heating that facilitates the dilution of particulate matter. Concentrations initiated an upward trend from 17:00 (BJT), with a gradual increase culminating in a peak between 21:00 and 22:00 (BJT). In contrast, clean days exhibited a subtle morning and overnight peaks. For PM_{10}, polluted days presented pronounced peaks at 10:00–12:00 and 21:00–22:00 (BJT), coinciding with increased dust resuspension or construction activities (Yu et al., 2020). However, valleys occurred at 07:00 and 19:00 (BJT). These may be due to reduced surface disturbances during dawn and enhanced vertical mixing in the afternoon. Clean days maintained relatively low hourly mean PM_{10} concentrations (50–70 μg/m³) with minor fluctuations~~, mirroring the $PM_{2.5}$ pattern~~. The time series curves of pollutants had different peak times and amplitudes, indicating distinct origins. $PM_{2.5}$ was more influenced by continuous anthropogenic emissions, while PM_{10} was sensitive to episodic coarse-particle events (Deng et al., 2023; Wang et al., 2024). These hourly dynamics highlighted the role of diurnal meteorological cycles in modulating particulate matter distribution, with polluted days amplifying both primary emissions and secondary aerosol formation (Liu et al., 2021).

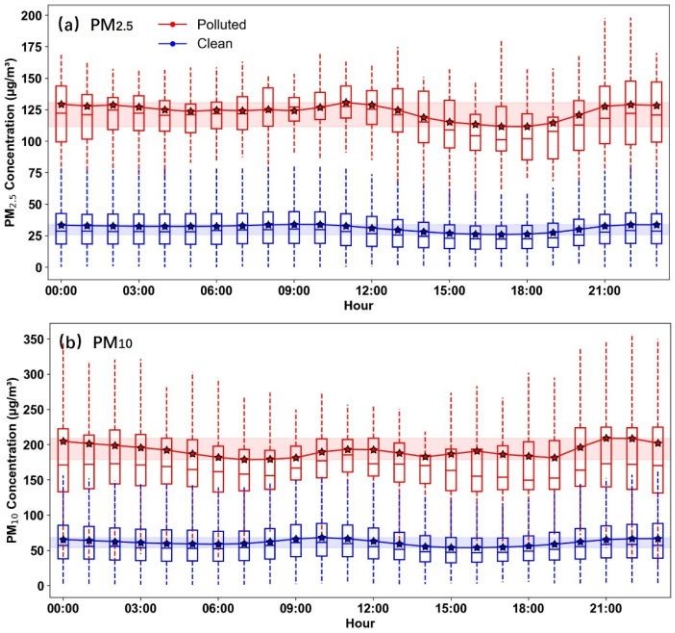


Figure 3. Hourly mean (a) $PM_{2.5}$ and (b) PM_{10} concentrations on polluted and clean days (Beijing time, BJT = UTC + 8 h). The blue and red shaded areas represent the ranges of mean concentration values for polluted and clean days, respectively. The whiskers and boxes represent the 95th, 75th, 50th, 25th, and 5th percentiles, respectively. The stars represent the mean values of the pollutant concentrations. The solid lines connect the hourly mean values to show the diurnal variation trends.

3.2 Temporal variation of vertical distribution structure

In addition to the impacts exerted by surface-level pollution, the investigation of vertical distribution is also of paramount significance for a comprehensive understanding of air quality dynamics. The extinction coefficient at 532 nm shows seasonal and vertical variation. Vertically, extinction coefficients were highest near the surface and gradually decreased with altitude in all seasons. This vertical decline arises from rapid reductions in aerosol loading due to vertical mixing, gravitational sedimentation, and dilution with increasing height, as well as the dominance of weaker molecular scattering above the planetary boundary layer (Wang et al., 2021). Winter presented the highest extinction values among the four seasons (**Fig. 4a**) (Chen et al., 2023; Wang et al., 2024). This was likely related to enhanced emissions from domestic heating and stagnant meteorological conditions that restrict vertical mixing (Zhong et al., 2018). In contrast, summer exhibited the lowest extinction coefficients at 532 nm across the profile, suggesting the effective dispersion of aerosols due to strong convection and a high planetary boundary layer (Li et al., 2015). Spring and autumn showed intermediate levels. The most significant differences between seasons appeared below 1 km, especially in the 0.2–0.6 km range (Zhong et al., 2018). Above 1.0 km, the extinction values in all

seasons converged toward high altitude, indicating a reduced aerosol presence in the upper layer (Liu et al., 2024).

The depolarization ratio also exhibited seasonal differences (**Fig. 4b**). With increasing altitude, the contribution of non-spherical particles increases, leading to a steady rise in depolarization ratio. This is attributed to the fact that non-spherical particles exhibit stronger anisotropic light scattering than nearly spherical fine-mode aerosols that dominate near the surface (Wang et al., 2020b). In the lower layer below 0.6 km, spring showed the highest depolarization ratios, which were attributed to a greater proportion of non-spherical particles such as mineral dust or internally mixed soot (Biuki et al., 2022). This may result from springtime dust resuspension and aged industrial emissions. In contrast, summer had the lowest depolarization ratios, implying that aerosols were dominated by more spherical particles, such as those formed through secondary processes (e.g., sulfate or organic aerosols) (Sun et al., 2013). In autumn and winter, values were moderate and stayed between the two extremes. However, the seasonal pattern changed above 0.6 km. Winter showed the highest depolarization ratio, while spring became the lowest. The vertical profile revealed that the change in particle shape with height strongly depends on seasons (Wang et al., 2004). In winter, the low ambient temperatures in the upper atmosphere promote the formation of non-spherical particles (e.g., ice crystals and irregular aerosols), which substantially enhance the depolarization ratio due to their anisotropic light-scattering properties (Haarig et al., 2017; Wang et al., 2021; Yu et al., 2021). In addition, elevated aerosol layers were dominated by long-range transported dust with pronounced nonspherical properties. However, aerosols in the upper layer were more spherical and less influenced by dust in spring (Wang et al., 2020b).

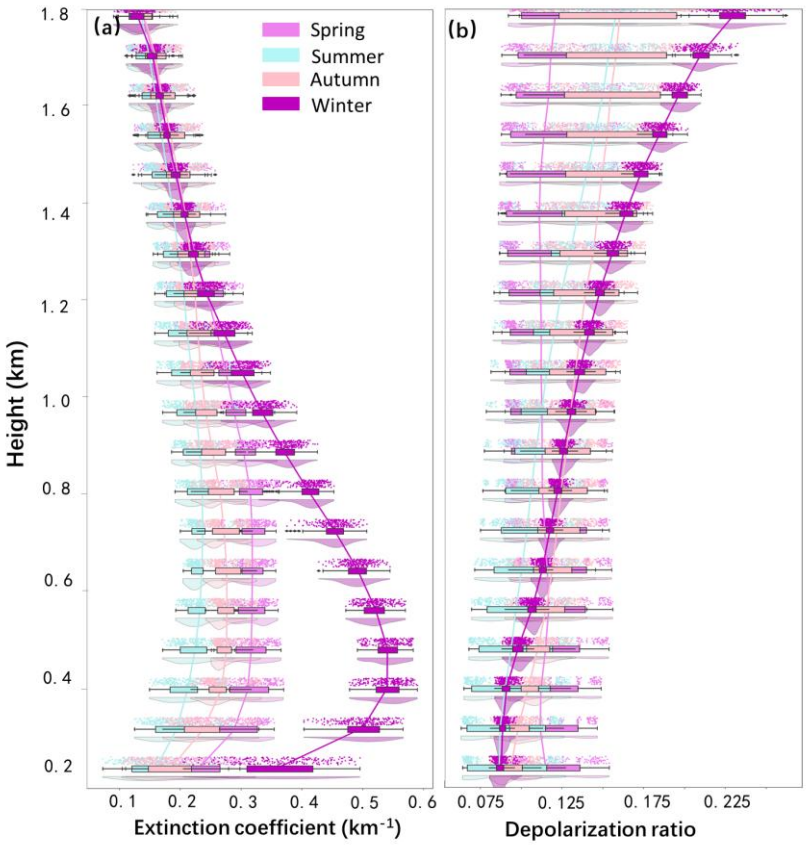


Figure 4. Seasonal variation of vertical aerosol (a) extinction coefficient and (b) depolarization profiles at 532 nm from LiDAR measurements. The whiskers represent the 90th and 10th percentiles, respectively. The shaded areas represent the 25th-75th percentiles. The connected lines depict the trend of mean values across different heights.

The extinction coefficient is significantly higher on $PM_{2.5}$-polluted days than on PM_{10} polluted days. This is attributed to the optimal size match between fine-mode particles ($\leqslant$ 2.5 μm) and visible light (0.4–0.7 μm), as well as their chemical components, which are more efficient at scattering and absorbing light (Wang et al., 2021; Chen et al., 2023). The extinction coefficients at 535 nm exhibited clear differences between polluted and clean days for both $PM_{2.5}$ and PM_{10}, with strong vertical and diurnal variation (**Fig. 5**). For $PM_{2.5}$, polluted days were characterized by significantly higher extinction values below 0.9 km, especially during the nighttime and the period from early morning to near noon (Wang et al., 2024). This enhancement was mainly attributed to the stable boundary layer and limited vertical mixing during the night, which favored the accumulation of aerosols near the surface. Strong solar radiation enhances photochemical reactions, leading to substantial secondary aerosol formation at noon and a lift in altitude (**Fig. 5a**). However, extinction values near the surface (below 0.4 km) were relatively low during 12:00-15:00 (BJT), with high values concentrated mainly around 0.6 km. This midday vertical uplift of the aerosol layer corresponds to the decrease in near-surface $PM_{2.5}$ shown in **Fig. 3**, as enhanced convective mixing lifts fine particles

upward and weakens near-surface extinction.~~, which was consistent with the low ground-based $PM_{2.5}$ concentrations during this period shown in **Fig. 3**.~~ Hence, strong daytime turbulence promotes the vertical transport of aerosols, resulting in an elevated and vertically expanded aerosol layer~~The lagged effect of boundary layer development indicated that pollutants accumulated in the morning were not yet fully dispersed~~ (**Fig. 5a**) (Wang et al., 2024). During this period (12:00–15:00 BJT), the enhanced extinction was concentrated around 0.6 km, extending vertically from the surface to this altitude, rather than peaking near the ground~~extinction values rose not only near the surface but also extended vertically up to around 0.6 km~~. The upward expansion of aerosols was driven by boundary layer development and persistent emission sources (Dai et al., 2020). In contrast, clean days showed low extinction values throughout the day, with little vertical variation and no significant peak. For PM_{10}, extinction coefficients at 532 nm were also increased on polluted days (**Fig. 5b**). However, the enhancement of extinction values extended to a higher altitude compared to that of $PM_{2.5}$, reaching up to 1.2 km. The primary increase occurred during the night and early morning, with a minor secondary rise observed around midday.(Li et al., 2020). Clean-day extinction remained uniformly low with minimal temporal variation (**Fig. 5d**).

For the depolarization ratio ~~(532 nm)~~, distinct patterns also emerged between polluted and clean days for both $PM_{2.5}$ and PM_{10} (**Fig. 6**). For $PM_{2.5}$, depolarization ratios near the surface (below ~0.6 km) were relatively low on polluted days, particularly during nighttime and early morning (**Fig. 6a**). This can be attributed to the accumulation of fine, spherical particles in the stable boundary layer, which tend to have lower depolarization ratios. As altitude increased, depolarization ratios rose. This indicates the presence of more irregularly shaped or coarser particles aloft, possibly from transported dust or aged aerosols. In contrast, clean days showed uniformly low depolarization ratios across most heights, consistent with the dominance of fine, spherical aerosols under favorable dispersion conditions (**Fig. 6c**). For PM_{10}, depolarization ratios at 532 nm on polluted days exhibited a different behavior (**Fig. 6b**). Starting around 8:00 (BJT), elevated depolarization values were observed across a broad vertical layer, extending up to ~1.2 km throughout the day. This suggests that coarser particles, which typically have higher depolarization ratios, were more effectively mixed vertically and persisted at higher altitudes compared to $PM_{2.5}$. The broader vertical distribution of PM_{10} depolarization signals reflects the larger size and potentially different sources (e.g., resuspended dust, industrial emissions) of these particles, which are less confined by boundary layer dynamics than finer $PM_{2.5}$. The increased depolarization ratio at noon during PM_{10}-polluted days is attributed to midday boundary layer growth and frontal dust transport, elevating non-spherical coarse particle concentrations. ~~The broader vertical distribution of PM_{10} depolarization signals reflects the larger size and potentially different sources (e.g., resuspended dust, industrial emissions) of these particles, which are less confined by boundary layer dynamics than finer $PM_{2.5}$.~~

The difference between polluted and clean days highlighted a strong increase in near-surface extinction during early morning on polluted days, indicating suppressed vertical mixing and increased emission accumulation (**Fig. 5e&f**). For $PM_{2.5}$, the differences were concentrated near the ground, especially below 0.9 km. The strongest enhancement appeared between 7:00 and 14:00 (BJT). A subtle lifting with height was observed around midday. Noon solar radiation maximizes surface heating, which causes severe thermal convection. The lower atmosphere becomes unstable as a result of the hot Earth transferring energy upward. It could break the stable stratification and drive upward air motion. Pollutants trapped near the surface are lifted, reducing extinction in the lower layer (<0.4 km) while increasing it aloft as vertical mixing intensifies (Wang et al., 2024). For PM_{10}, the enhancement tended to be localized and less vertically extended. The coarse particles may be lifted in upstream source regions, but their vertical transport weakens significantly along the trajectory. As a result, the transport altitude of PM_{10} reaching Hefei was largely confined below 1.2 km (Shim et al., 2022). Compared to $PM_{2.5}$, the high-value region for PM_{10} on polluted days has a higher vertical extension, and the period of strong extinction is shorter. The particles of PM_{10} are relatively large and heavy, which leads to rapid sedimentation and thus a short duration of heavy pollution. Pollutants tend to accumulate in the middle and low altitudes (0.4-1.2 km) on polluted days.

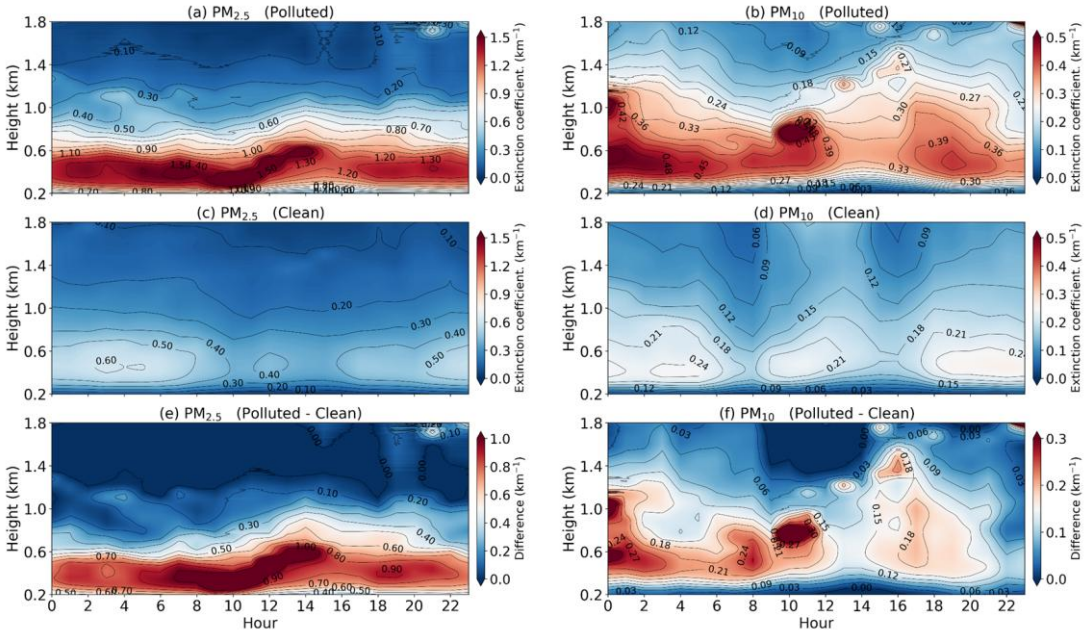


Figure 5. Diurnal variation in the vertical distribution of aerosol extinction coefficient (Winter days for $PM_{2.5}$: First column; Spring days for PM_{10}: Second column) under days with (a, b) polluted, (c, d) clean, and (~~E~~e, ~~F~~f) the difference between the two. All data are presented in Beijing time (BJT = UTC + 8 h) for the 532 nm channel.

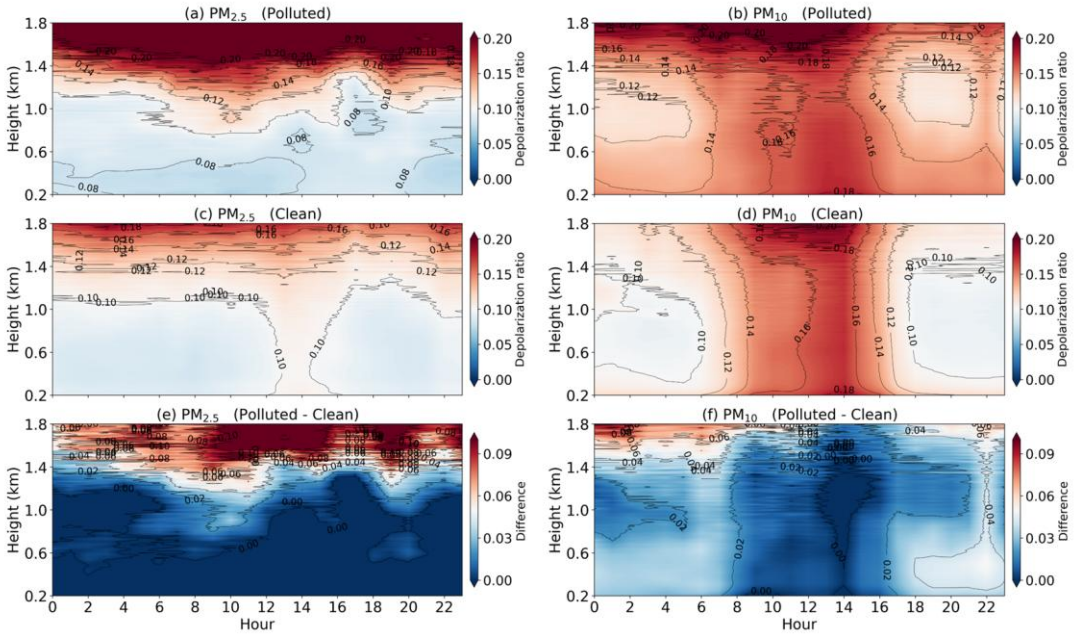


Figure 6. Same as Fig. 5, but for depolarization ratio.

To further analyze the vertical aerosol properties under different pollution levels, the mean extinction coefficient and depolarization ratio at 532 nm were calculated across three representative conditions (**Fig. 7**). Unlike the temporal variation that emphasized vertical profile distributions (**Fig. 5&6**), these curves highlight altitude-related trends (e.g., gradual changes, peak positions, and inflection points) for the comparison between polluted, clean, and difference conditions. For $PM_{2.5}$, the extinction coefficient decreased exponentially with height and intensified sharply below 0.4 km on polluted days (**Fig. 7a**) (Wang et al., 2020a). The near-surface value reached up to 1.4 km^{-1}at 0.4 km, which was nearly three times higher than that on clean days. The difference was most significant within the boundary layer, especially between 0.3-0.6 km. For PM_{10} (**Fig. 7b**), although the overall extinction values were lower than those of $PM_{2.5}$, the difference between polluted and clean days remained obvious. The polluted-clean difference for PM_{10} was more confined below 1.2 km. This discrepancy may be attributed to that the coarse-mode particles tend to concentrate closer to the surface due to their limited vertical transport.

The vertical distributions of the depolarization ratio under different pollution levels are presented in **Figs. 7 (c&d)** for $PM_{2.5}$ and PM_{10}, respectively. For $PM_{2.5}$, the depolarization ratio on clean days was higher than that on polluted days across the entire vertical column up to 1.8 km (**Fig. 7c**). The depolarization ratio of $PM_{2.5}$ increases with height. The most significant difference in depolarization ratios between polluted and clean days was observed around 1 km. Above this height, the difference narrowed but clean days still showed slightly higher values. Unlike near-surface $PM_{2.5}$, which is dominated by local pollution sources, $PM_{2.5}$ at higher altitudes is mainly associated with

long-range transport in the free troposphere, leading to higher depolarization ratios. Higher altitudes have more $PM_{2.5}$ particles from long-range transport in the free troposphere, which possess higher depolarization ratios (Vakkari et al., 2021; Wang et al., 2021). Even at higher altitudes, the residual influence of cleaner, non-spherical particle sources in the background air kept the clean-day depolarization ratio higher (Li et al., 2020). For PM_{10}, the depolarization ratio of polluted days was higher than that of clean days across the entire vertical range (**Fig. 7d**), reflecting a greater abundance of coarse-mode and irregularly shaped particles on polluted days. Below 1.2 km, the depolarization ratio of PM_{10} was distinctly higher than that of $PM_{2.5}$ on polluted days. This difference emphasized the dominant influence of coarse, non-spherical particles (e.g., dust and mechanically suspended material) in the PM_{10} fraction, particularly near the surface (Shim et al., 2022). On PM_{10}-polluted days in autumn, the depolarization ratio increased with height, which is mainly also affected by the long-distance transport of dust in the free troposphere.

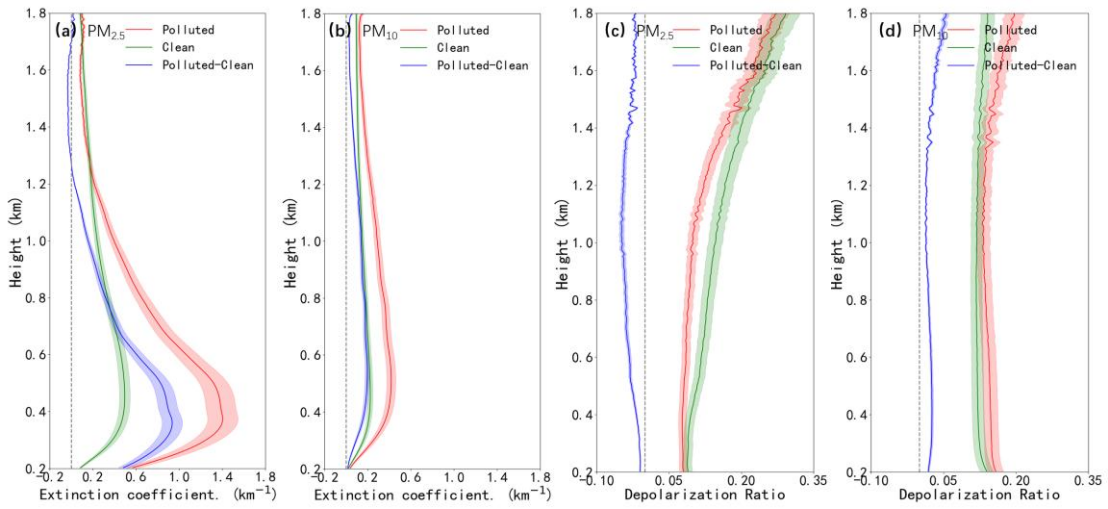


Figure 7. Vertical profiles of aerosol (a, b) extinction coefficient and (c, d) depolarization ratio at 532 nm for (a, c) $PM_{2.5}$ in winter and (b, d) PM_{10} in spring on polluted, clean days, and the difference between the two. The gray dashed line represents the zero line. Note: The extinction coefficient and depolarization ratio of individual polluted and clean days are positive physical quantities. Negative values in the "Polluted-Clean" panels indicate that the aerosol optical parameters (extinction coefficient or depolarization ratio) are higher on clean days than on polluted days. All data are presented in Beijing time (BJT = UTC + 8 h) for the 532 nm channel.

Figure 8 presents the 3-hourly variation of extinction coefficient and depolarization ratio at 532 nm on polluted and clean days across three altitude layers (0-0.5 km, 0.5-1 km, and 1-1.8 km). Below 0.5 km, the extinction coefficients were markedly higher on polluted days for $PM_{2.5}$ than that on clean days, with peak values occurring at 03:00-12:00 (BJT) (**Fig. 8a**), which was consistent with the variation in surface pollutant concentrations (**Fig. 3**). The

depolarization ratio for $PM_{2.5}$ was lower on polluted days in contrast with clean days. This is due to the dominance of spherical fine particles (e.g., sulfates, organic aerosols) from anthropogenic and secondary sources (**Fig. 8a&c**). These particles scatter light less directionally compared to the irregular coarse particles (e.g., dust) prevalent on clean days. In contrast, both the extinction coefficient and depolarization ratio for PM_{10} were high on polluted days (Zhang et al., 2020a). At the layer between 0.5 and 1 km, the extinction coefficient for $PM_{2.5}$ showed a significant peak around 12:00 to 15:00 (BJT) on polluted days, which was consistent with the upward transport of pollutants to around 0.6 km during the same period as shown in the diurnal vertical structure (**Fig. 5)**. The extinction values decreased significantly and remained at low levels~~showed limited diurnal variability~~ above 1 km (**Fig. 8e**). However, the depolarization ratio remained relatively high during polluted periods. The depolarization ratio for on days in winter was higher than that of spring, mainly because $PM_{2.5}$ pollution contains a higher proportion of irregular particles from coal-fired fly ash and industrial emissions (**Fig. 8f**). The low temperatures inhibit the diffusion and settlement of coarse particles. In contrast, days in spring for PM_{10} pollution were more affected by coarse particles such as windblown dust. Although these particles are irregular in shape, their large particle size makes them prone to settling. Moreover, the strong atmospheric fluidity in spring led to a slightly lower depolarization ratio compared to winter. Overall, the greatest contrast between polluted and clean days was observed in the lowest layer, where both particle concentration and shape varied most distinctly with pollution level and time of day (Zhong et al., 2018).

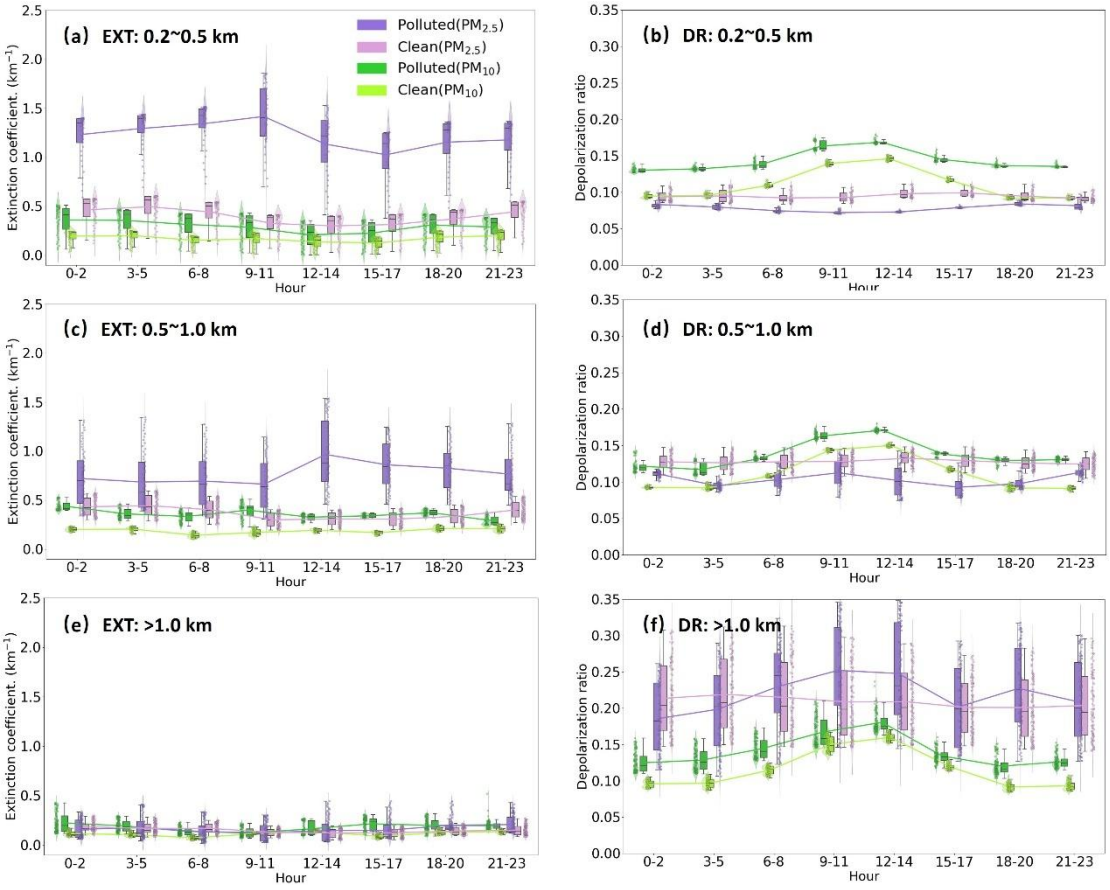


Figure 8. Diurnal variations of (EXT: a, c, e) extinction coefficient and (DR: b, d, f) depolarization ratio at 532 nm on polluted and clean days for $PM_{2.5}$ and PM_{10} at (a, b) 0.2-0.5 km, (c, d) 0.5-1.0km, and (e, f) above 1.0 km, respectively. The whiskers represent the 90th and 10th percentiles, respectively. The shaded areas represent the 25th-75th percentiles. The connected lines depict the trend of mean values across different heights. All data are presented in Beijing time (BJT = UTC + 8 h) for the 532 nm channel.

3.3 Role of aerosol extinction coefficient and depolarization ratio at different altitudes on ground-based $PM_{2.5}$ and PM_{10} concentrations

Figure 9 shows the vertical profiles of the relationships between surface pollutant concentrations and aerosol optical properties (extinction coefficient and depolarization ratio at 532 nm) on polluted and clean days. For $PM_{2.5}$, the extinction coefficients showed a clear positive correlation with surface $PM_{2.5}$ concentration throughout the profile below 0.9 km on polluted days, marking a key boundary for vertical variation in particle composition (**Fig. 9a**). The values on polluted days increase sharply near the surface, suggesting a local accumulation. A negative correlation was observed between the depolarization ratio and $PM_{2.5}$ concentration below 0.9 km on polluted days, as elevated $PM_{2.5}$ levels corresponded to a higher fraction of fine, spherical particles with low depolarization~~On polluted days~~

~~with $PM_{2.5}$, the depolarization ratio was negatively correlated with $PM_{2.5}$ concentration below 0.9 km, with fine particulate matter dominating. As the proportion of fine particulate matter increased, the proportion of spherical particles rose, leading to reduced depolarization~~ (Zhang et al., 2020a). In contrast, clean days showed a consistently positive correlation between depolarization ratio and surface pollutants throughout the vertical profile.

The increased depolarization levels are linked to greater proportions of non-spherical particles, such as dust or mechanically suspended matter (Vakkari et al., 2021). The depolarization ratio showed consistently positive correlations with surface PM_{10} concentration on polluted days, especially below 1.2 km (**Fig. 9b**). On clean days, the correlation between the depolarization ratio and PM_{10} concentration remained near zero, reinforcing that the depolarization ratio is a distinguishing feature of coarse-mode aerosols. These results revealed the contrasting aerosol properties between $PM_{2.5}$ and PM_{10} pollution. $PM_{2.5}$ events were dominated by fine, spherical particles that reduce depolarization, while PM_{10} events were closely associated with the presence of coarse, non-spherical particles that enhanced depolarization signals (Biuki et al., 2022).

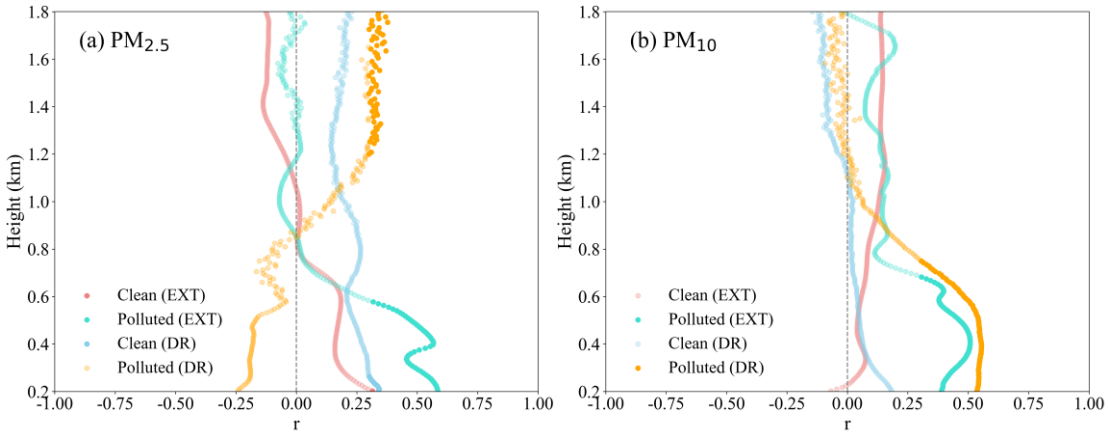


Figure 9. Vertical distributions of correlation (*r*) between aerosol optical properties (532 nm) and ground-based pollutants (a: $PM_{2.5}$ in winter, b: PM_{10} in spring) on clean and polluted days. EXT = extinction coefficient; DR = depolarization ratio. The gray dashed line represents the zero line.

3.4 Analysis of LiDAR observations during heavy pollution episodes

Temporal variations of the extinction coefficient and depolarization ratio at 532 nm exhibited a clear association with $PM_{2.5}$ and PM_{10} concentrations, effectively highlighting severe pollution periods (**Fig. 10**). Severe $PM_{2.5}$ pollution was evident at 10:00 ~~a.m.~~ on January 2, 2023 (BJT), with the maximum hourly concentration reaching 159.6 μg/m³. The peak of the extinction coefficient matched the actual maximum concentration of $PM_{2.5}$, with fine particulate matter exhibiting strong light-scattering properties at this time (**Fig. 10a**). However, the depolarization

ratio of $PM_{2.5}$-polluted days ~~showed reduced values~~ was markedly decreased during this peak period (**Fig. 10c**). Physically, the extinction coefficient scales with aerosol concentration and size, while the depolarization ratio is governed by particle morphology. Spherical fine particles such as hygroscopic sulfates and nitrates dominating this $PM_{2.5}$ event, leading to increased extinction and decreased depolarization. A dramatic peak in PM_{10} concentration was observed, with the highest hourly value of 1440 μg/m³ recorded at 22:00 p.m. (BJT) on April 11, 2023. The severe PM_{10} event was reflected by a pronounced increase in the extinction coefficient (**Fig. 10b**) (Sun et al., 2013). In addition, the depolarization ratio exhibited high values during this period, which aligned with the elevated extinction coefficient and was consistent with the irregular morphology of coarse particles that contributed to both stronger light scattering and greater depolarization (**Fig. 10d**). Overall, the profiles of the extinction coefficient and depolarization ratio correspond directly to the temporal patterns of $PM_{2.5}$ (January 1–8, 2023) and PM_{10} (April 11–13, 2023) concentrations (**Fig. 10**). Peaks in pollutant concentrations were reflected in enhanced extinction coefficients and characteristic depolarization ratio patterns (Zhong et al., 2018). This confirms that these optical metrics effectively identify severe pollution episodes.

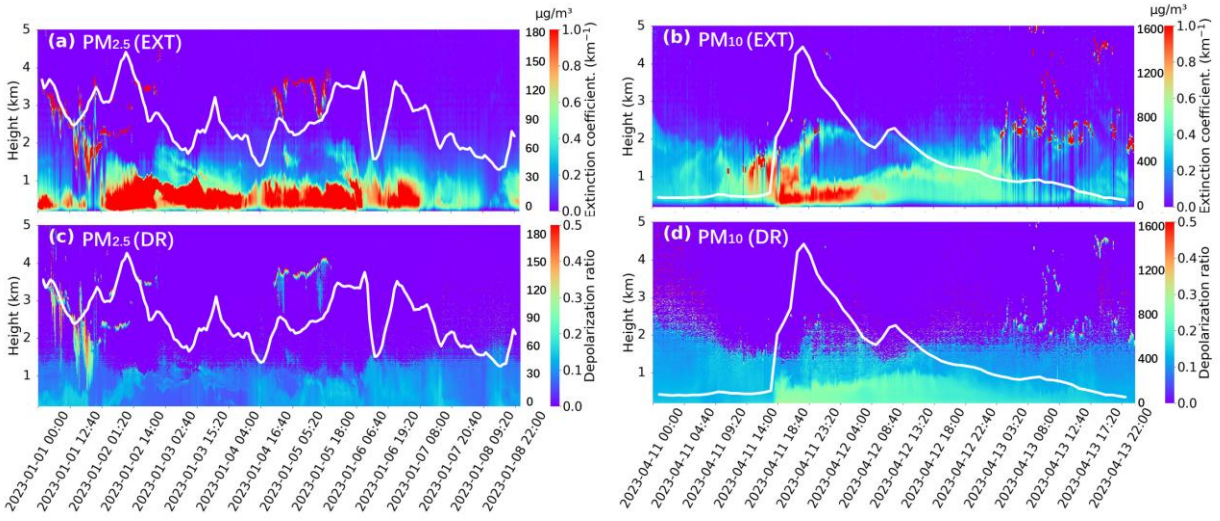


Figure 10. Pollution Episodes: Temporal dynamics of (a, b) extinction and (c, d) depolarization parameters for the 532 nm channel during severe (a, c) $PM_{2.5}$ (January 1–8, 2023) and (b, d) PM_{10} (April 11–13, 2023) periods in Hefei. The white lines represent the hourly (a) $PM_{2.5}$ and (b) PM_{10} concentrations, with their scale on the secondary y-axis. All data are presented in Beijing time (BJT = UTC + 8 h). EXT = extinction coefficient; DR = depolarization ratio.

3.5 Analysis of meteorological conditions and their driving role in air pollution

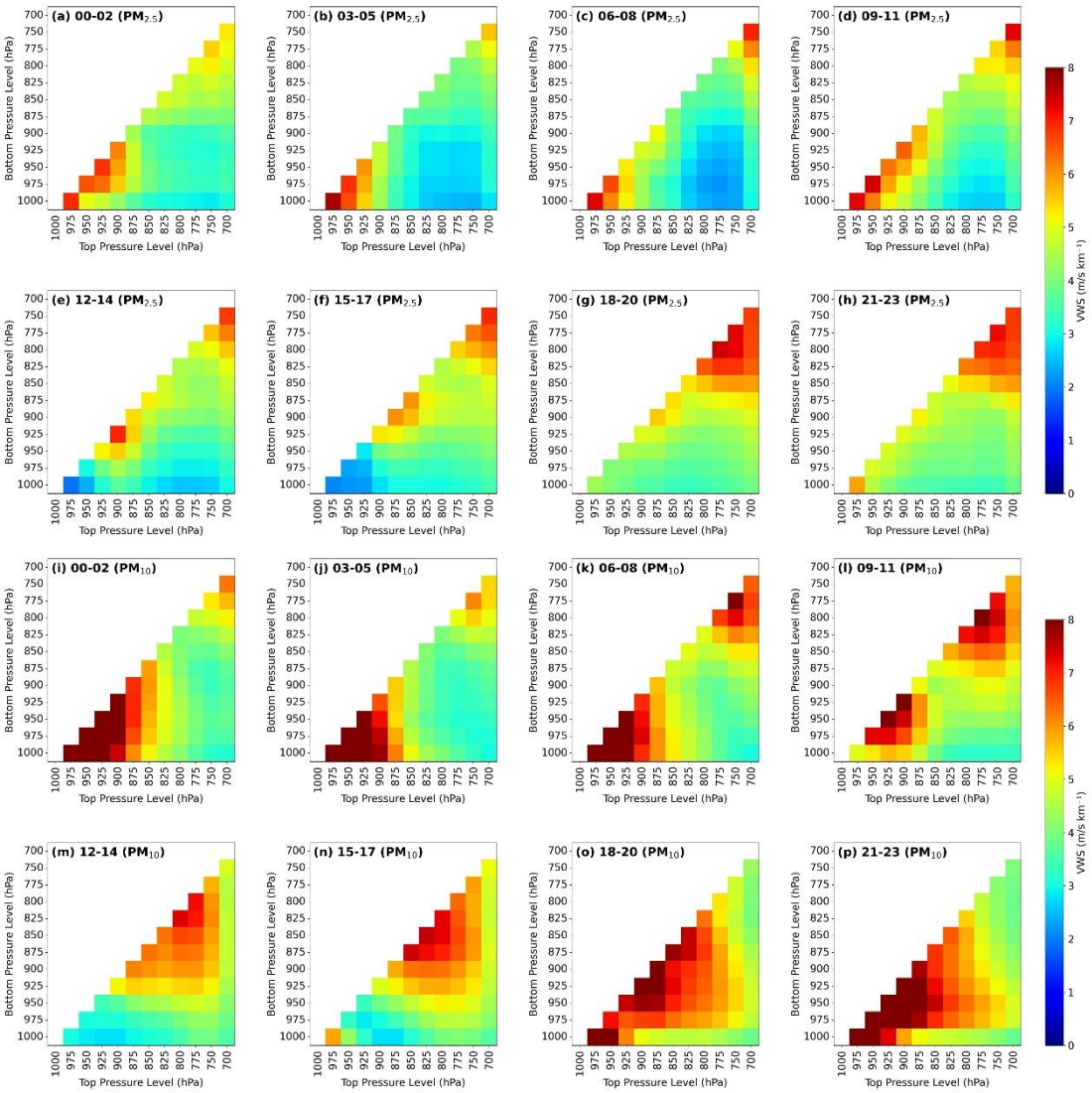
Figure S1 shows boxplots of the concentrations of $PM_{2.5}$ and PM_{10} in temperature (T), relative humidity (RH), surface pressure (PRS), and wind speed (WS) bins, respectively. Overall, the relationships of temperature with $PM_{2.5}$,

513 PM_{10} were not significant (**Fig. S1a&e**). The RH exerted distinct influences on PM_{10} and $PM_{2.5}$ concentrations (**Fig.**
514 **S1b&f**). Higher $PM_{2.5}$ concentrations were associated with increased RH, primarily due to elevated humidity
515 promoting the condensation of gaseous precursors (e.g., sulfur dioxide, nitrogen oxides) onto pre-existing $PM_{2.5}$
516 particles (**Fig. S1b**) (Wang et al., 2004). Additionally, it accelerated secondary aerosol formation via aqueous-phase
517 chemical reactions in the atmosphere, thereby elevating $PM_{2.5}$ levels (Yang et al., 2022). In contrast to $PM_{2.5}$, the
518 relationship between PM_{10} concentrations and RH was inverse (**Fig. S1f**). Higher RH promotes ~~drives coarse particles~~
519 ~~like PM_{10} to absorb~~ moisture uptake and growth of coarse PM_{10} particles , accelerating their sedimentation and wet
520 deposition, thereby reducing ambient concentrations~~grow in size, and then settle out of the atmosphere more easily.~~
521 ~~It may also enhance wet deposition processes such as rain or fog scavenging, reducing their airborne concentrations~~
522 (Gao et al., 2020; Ma et al., 2023). PRS tended to increase with rising concentrations of $PM_{2.5}$ and PM_{10} (**Fig. S1c&g**).
523 The average pressure values were lower on clean days ($PM_{2.5} \leqslant 75$ μg/m³, $PM_{10} \leqslant 150$ μg/m³) and became higher on
524 polluted days. The high-pressure systems promote atmospheric stability and suppress vertical mixing, thereby
525 facilitating the accumulation of particulate matter near the surface (Li et al., 2015). Wind speed exhibited a decreasing
526 trend with increasing $PM_{2.5}$ concentrations (**Fig. S1d&h**). Lower wind speeds during heavy pollution periods indicate
527 weak horizontal dispersion, which contributes to the persistence and accumulation of fine particles in the boundary
528 layer. In addition, the polar plot of $PM_{2.5}$ concentrations shows a directional pattern, with the highest values
529 predominantly associated with winds from the northwest (**Fig. S2a**). The cities of northern Anhui Province, Henan
530 Province, and even further north were characterized by dense industrial activities and frequent wintertime heating
531 emissions, which contribute to elevated $PM_{2.5}$ levels (Qian et al., 2024; Shi et al., 2018). For PM_{10}, the concentration
532 distribution was more spatially confined but still shows a dominant contribution from the northwest area (Huang et
533 al., 2016). The most intense PM_{10} concentrations were clustered in the northwest, appearing localized and patchy
534 (**Fig. S2b**).~~ This may reflect the influence of transport and mechanical resuspension processes, such as road dust and~~
535 ~~construction activities.~~

536 VWS is a key dynamic factor affecting pollutant dispersion, as it influences mechanical turbulence and vertical
537 mixing within the boundary layer (Deng et al., 2023). **Figures 11** and **12** illustrate the diurnal variation of VWS on
538 polluted and clean days, respectively. For $PM_{2.5}$, weak shear persisted throughout the boundary layer on polluted days
539 when temperature inversions and stable stratification typically dominate the lower atmosphere (**Fig. 11a-h**). Clean
540 days showed stronger shear during most periods, enhancing upward transport and dilution of fine particles (**Fig. 12**).
541 Notably, the difference in VWS between polluted and clean days ~~WAS~~ was more pronounced in the lower boundary
542 layer (1000–900 hPa) than in the upper boundary layer (875–700 hPa) (Zhang et al., 2020b). Hence, suppressed

vertical mixing near the surface contributed significantly to the accumulation of $PM_{2.5}$. Surface air quality tends to deteriorate significantly due to pollutant accumulation near the ground, especially in cases where shear is insufficient in the lower layer (Wang et al., 2024), which was consistent with the trend that the extinction coefficient at 532 nm near the surface is higher than that in the upper layer (**Figs. 7 & 8**).

Contrary to the expectation that polluted days generally exhibited weak VWS, the VWS associated with PM_{10} on polluted days was not consistently lower than that on clean days (**Fig. 11 i-p, Fig. 12 i-p**). In particular, dust storms with high PM_{10} levels are typically accompanied by intensified VWS, which is conducive to dust uplift and transport under dynamic atmospheric conditions (Yang et al., 2019). The large VWS associated with PM_{10} would clear the air in the upper air at the beginning but later contributed to raising surface PM_{10} levels through regional transport at the near-surface and through downward transport of aerosol particles from the upper air. During the onset phase of dust intrusion, intensified shear facilitates both horizontal and vertical transport of dust particles (Biuki et al., 2022). From 00:00 ~~am~~ to 11:00 (BJT), the VWS in the upper layer (850 ~~hpa~~hPa-700 ~~hpa~~hPa) gradually increases on $PM_{2.5}$-polluted days. At night, the near-surface air cools, forming a stable temperature inversion layer that inhibits vertical mixing. Hence, the stable structure would lead to weak VWS in the upper layer. As the sun rises in the mornings, the inversion layer gradually dissipates, and vertical mixing shifts from being suppressed to active (**Fig. 11 i-l**). The VWS in the upper layer gradually increases due to the mixing effect. In addition, PM_{10} pollution often occurs during cold front passage. The vertical gradient of wind direction and speed changes suddenly near the frontal surface when cold and warm air masses converge. Once the front is located in a certain pressure range, the wind shear may be significantly higher than that in the surrounding layers. The weaker VWS between adjacent pressure layers (e.g., near 825–850 hPa) may correspond to the fault distribution (**Fig. 11 i-l**). The sinking of cold air also leads to an increase in VWS at the lower level.



564

565 **Figure 11.** Three-hourly variation of vertical wind shear for (a-h) $PM_{2.5}$ and (i-p) PM_{10} on polluted days in Hefei ~~,~~
566 ~~presented as 3-hourly intervals (~~at (a&i~~:~~) 00:00–02:00 BJT, (b&j~~:~~) 03:00–05:00 BJT, (c&k~~:~~) 06:00–08:00 BJT,
567 (d&l~~:~~) 09:00–11:00 BJT, (e&m~~:~~) 12:00–14:00 BJT, (f&n~~:~~) 15:00–17:00 BJT, (g&o~~:~~) 18:00–20:00 BJT, and (h&p~~:~~)
568 21:00–23:00 BJT~~)~~. Beijing Time (BJT) is UTC+8, the local standard time used in this study. The axes indicate the
569 vertical bounds (upper and lower pressure levels) used to compute VWS, where the main diagonal (from top-right to
570 bottom-left) corresponds to self-referential values for single-layer calculations. The color gradient represents the
571 magnitude of VWS between adjacent atmospheric layers.

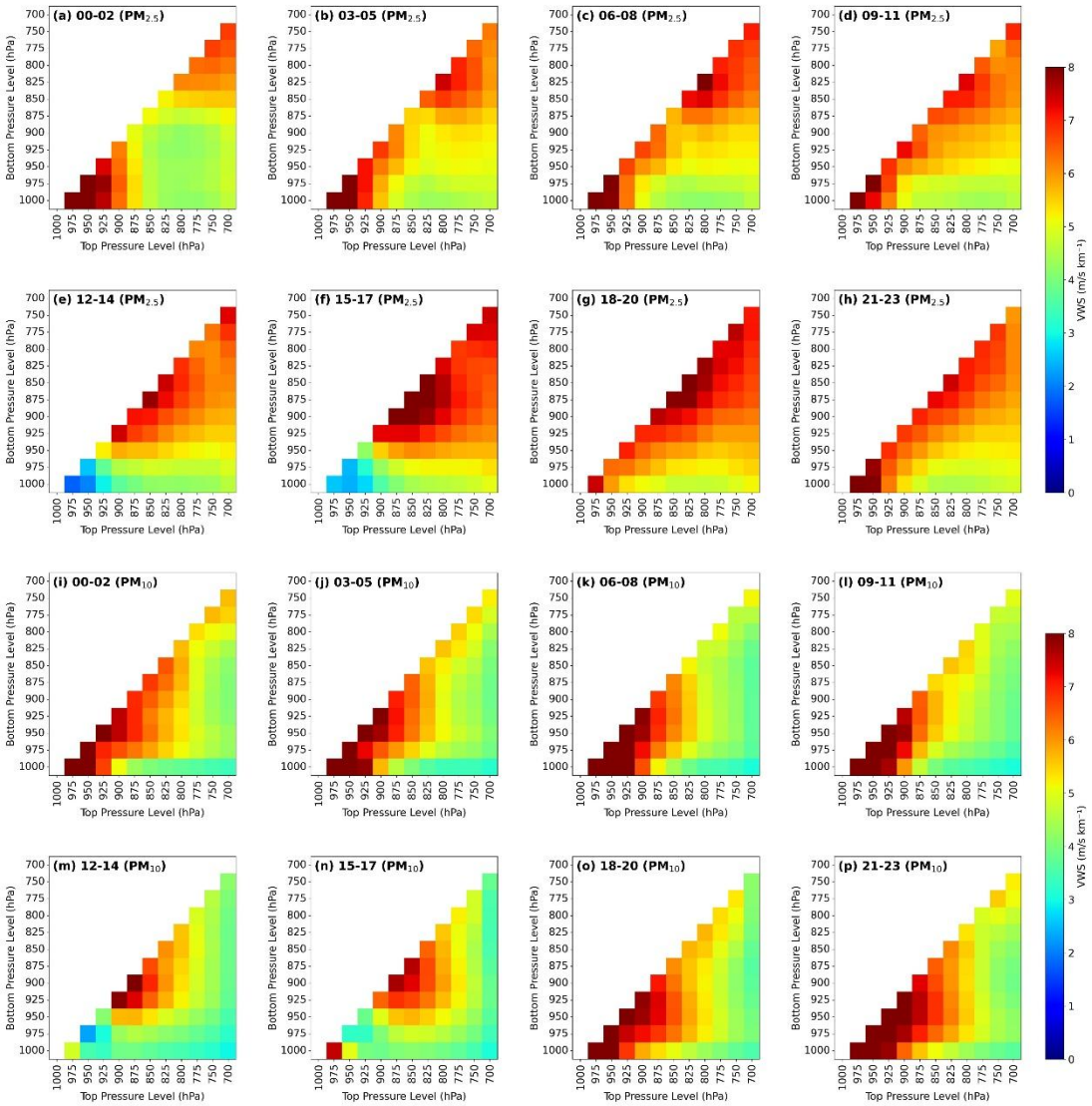


Figure 12. Same as Fig. 11, but under clean days.

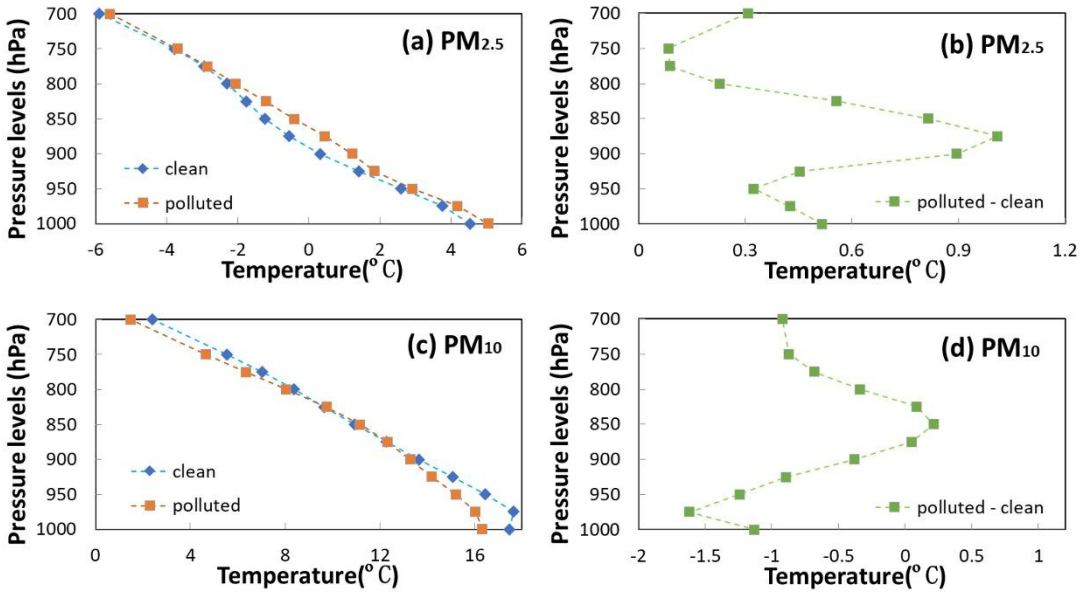
To clarify the synoptic influences on $PM_{2.5}$ and PM_{10} pollution, we analyzed the distributions of air temperature across different pressure levels from March 2021 to May 2023. During $PM_{2.5}$ pollution days, temperatures at all levels were consistently higher than those on clean days. The warming showed a notable pattern, with the most pronounced trend initially observed at 875 hPa, (**Fig. 13b**). The enhanced upper-level warming relative to the lower-level led to a thermally stable stratification, which inhibited vertical mixing and promoted the accumulation of $PM_{2.5}$ near the surface. In contrast, PM_{10}-polluted days exhibited negative temperature differences below 875 hPa (**Fig. 13d**), indicating the intrusion of cold air masses. These cold air masses, often accompanied by long-range dust transport, not only bring in PM_{10} particles from upwind regions but also suppress vertical mixing in the lower atmosphere due to their higher density. Notably, the altitude of the cooling (875 hPa) coincided with the effective aerosol layer height identified in our LiDAR-based analyses (**Figs. 8, 11, 12**), which supported the conclusion that dust particles were

primarily transported within this level.

To further investigate the vertical thermal and humidity differences between clean and polluted days, we focus on three key pressure levels (1000, 850, and 500 hPa). During $PM_{2.5}$-polluted periods, the warming at 850 hPa was greater than that near the surface at 1000 hPa in Hefei (**Fig. S3f&i**). This vertical temperature structure created a pronounced inversion, which inhibited the vertical mixing and enhanced the atmospheric stability. The lower troposphere also exhibited elevated relative humidity, particularly in Hefei and nearby regions. Elevated relative humidity suppresses turbulent exchange and reduces dilution capacity, and hence forms a shallow, moist, and stable boundary layer that favors pollutant persistence and contributes to poor air quality. In contrast, the clean days showed a lower humidity in the boundary layer and a weaker vertical temperature gradient, allowing enhanced upward motion and effective removal of surface aerosols (Deng et al., 2023). PM_{10} pollution episodes occurred under a markedly different meteorological background (**Fig. S4**). Specifically, the relative humidity at both 1000 hPa and 850 hPa was significantly lower on PM_{10}-polluted days than on clean days (**Fig. S4f&i**). Aloft cold air intrusions occurred on PM_{10} pollution days, creating favorable conditions for the large-scale transport of dust into Hefei. Under such dry and calm conditions, the accumulation of PM_{10}~~PM_{10} accumulation~~ is likely driven by mechanical resuspension or regional dust transport rather than secondary aerosol formation processes enhanced by moisture (Li et al., 2020). Overall, these findings reveal that $PM_{2.5}$ episodes over Hefei are typically governed by moist and thermodynamically stable boundary layers under subsiding air masses, while PM_{10} events are more influenced by dust-laden airflow and dry boundary-layer dynamics. The distinct thermal and humidity structures observed across the vertical profile emphasize the importance of differentiating pollution types when diagnosing meteorological drivers.

The vertical wind velocity was investigated at three pressure levels to reveal the dynamic mechanisms on $PM_{2.5}$ and PM_{10} pollution days (**Figs. 14 & 15**). To better elucidate the meteorological characteristics between polluted and clean conditions, we calculated the differences in vertical velocity between these two scenarios, as their broadly similar large-scale background circulation patterns obscure the subtle but critical dynamic anomalies that drive pollution formation. On $PM_{2.5}$-polluted days, the atmosphere around Hefei exhibits a clear subsidence trend at 850 and 1000 hPa (**Fig. 14**). The subsidence reflects the existence of high-pressure systems and stagnant synoptic conditions, which contribute to the accumulation of fine particles. This inhibition of the upward motion of vertical wind corresponds with a stable thermal structure, which can effectively restrict vertical exchange and compress the boundary layer. During PM_{10} pollution events, subsidence at 500 hPa is significantly strongergreater than ~~on clear days compared to~~ that on clean days (**Fig. 15**). However, ~~However,~~ Hefei is located in a unique region defined by upward motion at 850 hPa, with a comparable weak upward movement also observed at 1000 hPa. The subsidence

615 at 500 hPa stabilizes the atmosphere and suppresses the upward dispersion of particles. Meanwhile, upward motion
616 at 850 hPa and near the surface (1000 hPa) may be caused by local convergence or orographic lifting, allowing the
617 uplift and recirculation of dust within the lower troposphere. This dynamic difference could reflect more complex
618 interactions between local surface sources and large-scale synoptic. Upward motion at 850 hPa supports the
619 interpretation that PM_{10} events are influenced by weak convective movements or vertical recycling of particles within
620 a confined layer rather than stagnation alone. These findings reveal that $PM_{2.5}$ and PM_{10} pollution in Hefei are
621 governed by contrasting vertical dynamic regimes. $PM_{2.5}$ events are closely tied to a uniformly subsiding atmosphere,
622 which favors the trapping and accumulation of fine particles. In contrast, PM_{10} events are characterized by a layered
623 structure, where upper-level subsidence and lower-level ascent work in tandem to recirculate coarse particles, rather
624 than simply trapping them. This distinction underscores the need for targeted, pollutant-specific strategies in air
625 quality forecasting and management.



626

627 **Figure 13.** Vertical distributions of temperature for polluted, clean, and their difference in (a, b) $PM_{2.5}$ and (c, d)
628 PM_{10}.

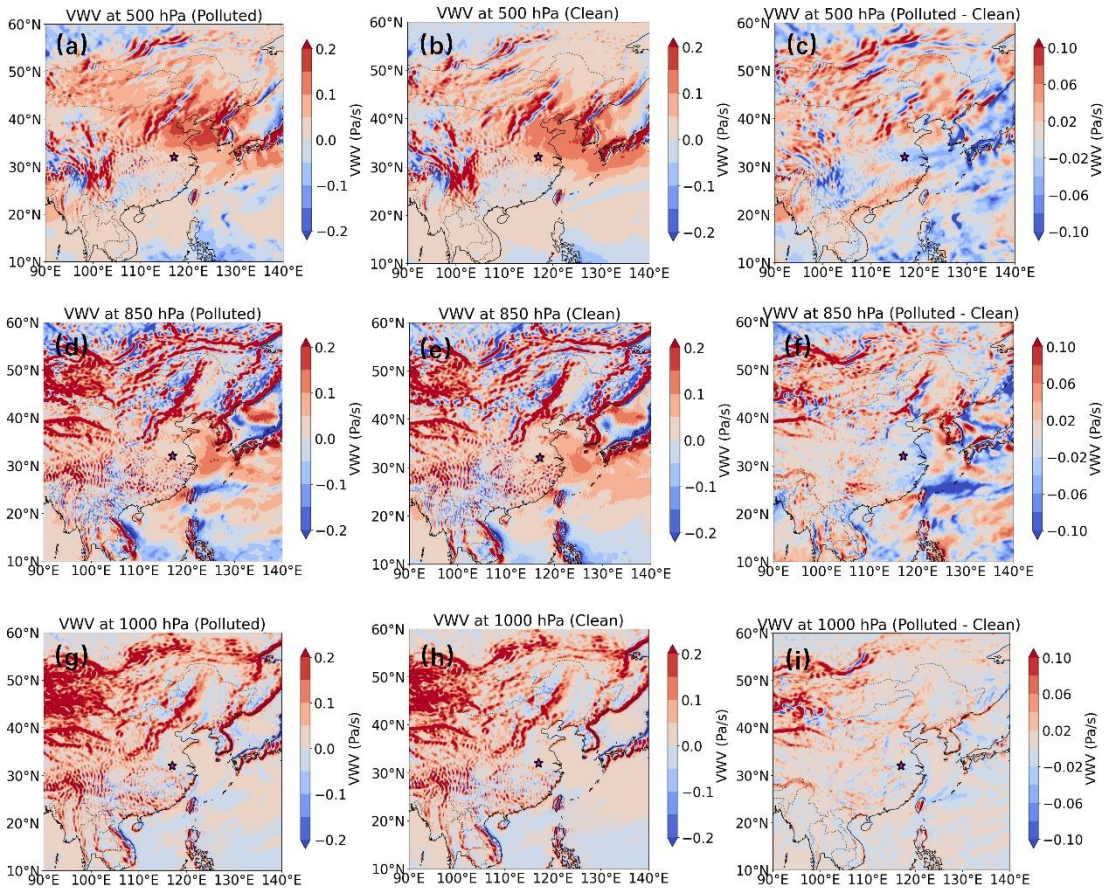


Figure 14. Spatial distribution of the vertical wind velocity (VWV) at three pressure levels (500 hPa: a–c; 850 hPa: d–f; 1000 hPa: g–i) on $PM_{2.5}$-polluted (left column), clean (middle column) days and the difference between them (right column), respectively. The areas highlighted with a red star represents the location of Hefei.

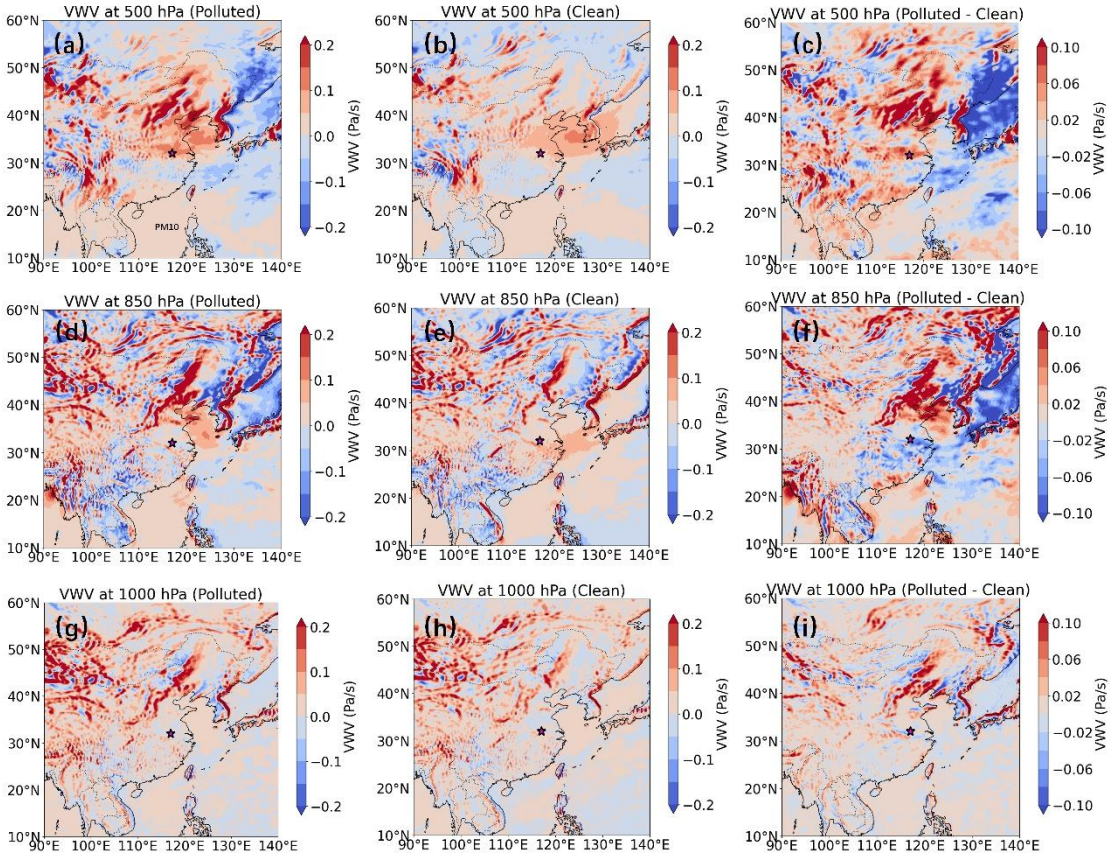


Figure 15. Same as Fig. 14, but on PM_{10}-polluted (left column), clean (middle column) days and the difference between them (right column), respectively.

4. Conclusions

This study comprehensively assesses the vertical structures of aerosols in Hefei by comparing fine ($PM_{2.5}$) and coarse (PM_{10}) particulate matter on clean and polluted days. We utilized long-term aerosol LiDAR measurements and reanalysis meteorological data to compare the vertical distribution characteristics of different particulate pollution types as well as the underlying thermodynamic and dynamic mechanisms that drive their vertical development. Aerosol extinction coefficients at 532 nm for $PM_{2.5}$ and PM_{10} were consistently higher on polluted days than on clean days. Since fine-mode particles predominate the aerosol population on $PM_{2.5}$ pollution days, the depolarization ratios fell within a range lower than those observed on clean days. In contrast, PM_{10}-polluted days had larger depolarization ratios than clean days, indicating significant contributions from non-spherical coarse particles.

Our findings reveal that the accumulation of $PM_{2.5}$ is closely linked to stable boundary layer structures characterized by high humidity, weak vertical shear, and a mid-level inversion temperature. These conditions collectively inhibit vertical exchange and confine fine particles near the surface. The vertical extinction and depolarization profiles support the dominance of aerosols during $PM_{2.5}$ pollution, particularly in the lower 0.9 km. In

652 contrast, PM_{10} pollution is typically associated with low humidity and mechanically driven transport of dust from
653 upwind sources. The vertical profiles show enhanced depolarization ratios, indicating the presence of non-spherical,
654 coarse particles such as dust or resuspended material. Distinct vertical motion patterns identify the two types of
655 pollution. $PM_{2.5}$ episodes showed upward mobility in the mid-troposphere and near-surface subsidence, which
656 reinforced stratification and pollution trapping. However, PM_{10} events were distinguished by upper-level subsidence
657 coupled with low-level ascent, which might facilitate vertical transport of coarse particles during regional dust
658 episodes. These dynamic differences highlight that $PM_{2.5}$ and PM_{10} are generated by fundamentally separate
659 atmospheric processes, and a unified treatment of particulate pollution risks oversimplifies their behavior.

660 Expanding such analyses across multiple sites would also aid in identifying geographical heterogeneity and
661 testing the generalizability of the discovered trends. Future research may be directed toward integrating long-term,
662 high-resolution vertical observations with advanced modeling approaches to better understand the multiscale
663 dynamics behind various forms of particle pollution and create tailored mitigation solutions.

664

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671

672 ## Data Availability

673 Hourly concentrations of major air pollutants were obtained from the China National Environmental Monitoring
674 Centre (CNEMC) (https://www.cnemc.cn/). Surface meteorological variables were obtained from the China
675 Meteorological Administration (CMA) (http://data.cma.cn/en). The ERA5 reanalysis dataset, used to investigate the
676 impact of synoptic systems on $PM_{2.5}$ and PM_{10} pollution, was freely accessible via the Copernicus Climate Change
677 Services platform (https://cds.climate.copernicus.eu/datasets).

678

679 ## Author Contribution

680 X.D. conceived the study; Y.Y. acquired and analyzed the data and drafted the manuscript; R.D. and Q.X.
681 performed data processing and investigation; Q.H., Y.L., and C.W. provided theoretical support; J.X. and Y.S.

682 discussed the results and revised the manuscript; Y.L. acquired funding; all co-authors reviewed and approved the
683 final version of the manuscript.
684
685 ## Competing interests
686 The authors declare that they have no conflict of interest.
687
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