



# Interpolating station quantile biases for tropospheric ozone MDA8 bias correction

Jan Peiker<sup>1,2</sup>, Jan Karlický<sup>1</sup>, and Peter Huszár<sup>1</sup>

<sup>1</sup>Department of Atmospheric Physics, Faculty of Mathematics and Physics, Charles University, V Holešovičkách 2, 180 00, Prague 8, Czech Republic

<sup>2</sup>Modeling and Expertise Pool, Czech Hydrometeorological Institute, Na Šabatce 17, 143 00, Prague 4, Czech Republic

**Correspondence:** Jan Peiker (jan.peiker@matfyz.cuni.cz)

**Abstract.** Chemistry transport models (CTMs) consistently exhibit systematic errors in ozone concentrations, which can be partly compensated by bias correction. There are several bias correction strategies suitable for using station data, but they are likely to introduce statistical artifacts when applied at high resolution. We propose a new bias correction strategy based on parametric interpolation of quantile biases (PIQB) suitable for high resolution simulations, which is designed to avoid such artifacts. In this study, we evaluate and compare the performance of our strategy with other older strategies with a focus on ambient maximum daily 8-h average ozone concentrations (MDA8). Our experimental setup consisted of two simulations from the CTMs WRF-Chem and CAMx at horizontal resolution of 9 km within the time period of 2007–2016 and 165 ground-based stations in central Europe. Our results show that each strategy brings the simulated MDA8 closer to observations, but PIQB performs the best in terms of mitigating systematic errors while retaining the modeled fine resolution structure of spatial variability. We conclude that out of the considered strategies, PIQB is the most suitable one for bias correction at high resolution, suggesting its possible applications for correcting climate projections of ozone MDA8.

## 1 Introduction

Tropospheric ozone belongs among important gas-phase pollutants, having negative impacts on health of both fauna and flora (Da et al., 2022; Gu et al., 2023; Pozzer et al., 2023). Because of that, it is desired by both the scientific community and policy-makers to study the long-term evolution of its concentrations, including the interactions among air quality and the changing climate (Racherla and Adams, 2008; Doherty et al., 2017; Archibald et al., 2020). The chemical production of ozone is initiated by photochemistry, and so the dominant influence upon its concentrations is the budget of its precursors, i.e., oxides of nitrogen ( $\text{NO}_x$ ) and volatile organic compounds (VOCs) as well as the amount of sunlight, temperature and humidity (Seinfeld and Pandis, 1998). That is mostly why projecting ozone concentrations into the future is burdened with uncertainty, as mitigating the anthropogenic  $\text{NO}_x$  budget may not be enough to compensate for the influence of the rising temperature and humidity (Colette et al., 2015).

To model the complex interactions among chemical species and other components of the climate system, global chemistry models can be used (e.g., Emmons et al., 2020; Sukhodolov et al., 2021). However, their relatively coarse resolution limits their performance in assessing regional ozone burdens simply due to the inability to fully represent subgrid-scale effects, such



25 as local sources of pollution or fine orographical features. Such limitations often result in biases or other types of systematic errors when compared to station observations (Rieder et al., 2015; Revell et al., 2018; Liu et al., 2022). To obtain information at finer resolutions, regional chemistry-transport models (CTMs) nested into global models are widely used, as they are capable of taking into account the transport of many different species from their sources along with their chemical interactions within the atmosphere (Grell et al., 2005; ENVIRON, 2018). Despite including features unresolved by global models and providing  
30 outputs at resolutions of units of kilometers, CTMs are still not perfect and exhibit systematic errors as well (Mar et al., 2016; Karlický et al., 2017; Karlický et al., 2024; Prieto Perez et al., 2025). Such model errors compared to measurements (which generally serve as the ground truth) are therefore an inherent feature of air quality simulations.

35 Luckily, the inevitability of errors can be at least partly compensated by postprocessing, namely by statistical procedures known as bias correction (e.g., Rieder et al., 2015; Rahmani et al., 2017; Abdelmoaty et al., 2025; Im et al., 2025), which adjust the model data to better fit the corresponding observations. In its general form, bias correction requires three inputs, which are observations, modeled data corresponding to the observations and modeled data to be corrected. Its main motivation is the application on projections through the assumption that the systematic errors arise from the model itself and are therefore identical in the projection and a historic run using input data generated by the same methodology. This assumption yields another source of uncertainty, because it depends on the definition of the systematic error, which may be assumed different  
40 across various bias correction strategies (Ugolotti et al., 2023; Staehle et al., 2024), on top of which, there is evidence from Liu et al. (2022) that biases differ with chemical environment and hence will potentially be different between a projection and a present-day simulation. Additionally, different sources of observations can be used for bias correction, and so the result and performance of such corrections may depend on the choice of reference data as well (Schnell et al., 2014; Staehle et al., 2024). In this study, we explore using ground based station measurements, as they represent the actual ground-level immissions.

45 As there are mostly studies applying bias correction of ozone concentrations using station data to model grids at coarse resolution (e.g., Schnell et al., 2014; Rieder et al., 2015), we aim to address their use for finer resolution climate simulations to demonstrate the possibility and to contribute to this research area. On the other hand, operational air quality forecasts use some form of bias correction on a routine basis (e.g., Neal et al., 2014). And so, regarding postprocessing of a regular model grid using station data, the task must be approached from the following perspectives: (i) the choice of statistical procedure to  
50 correct a single model time series (i.e., correcting corresponding datasets) and (ii) the strategy for extrapolating the outreach of such a procedure into the regular grid (i.e., correcting non-corresponding datasets).

55 To address (i), based on a recent comparison performed by Staehle et al. (2024), a procedure called quantile mapping (QM) detailed by Rieder et al. (2015) performs satisfactorily well compared to other methods (and is commonly used for other variables as well, e.g, precipitation; Abdelmoaty et al., 2025), and thus shall be considered in this study. The key principle of QM dwells in unifying each inter-quantile range (within a fixed set of quantiles) and its relative position of the modeled and observed data separately, therefore taking into account the possibility of having different biases in different parts of the modeled probability density function (PDF). This is especially crucial in the case of correcting ozone concentrations, as the statistical systematic errors may emerge naturally from the discrepancies between the modeled and observed PDFs (Karlický et al., 2024).



60 From the perspective of (ii), the task is to utilize the local information contained within the station sites in other parts of the model domain. In other words, we wish to somehow represent or to generalize the station measurement outside of the few model grid-cells to obtain a seamless dataset which would be directly comparable to the modeled data in each grid-box. It should not be surprising that in such a task many problems emerge. Most importantly, if done with an insufficient care, there is a risk of corrupting the simulation physics with statistical artifacts, as also demonstrated in this paper.

65 This study proposes a strategy of bias correction suitable for fine resolution simulations and usage of ground-based station data, which avoids introducing statistical artifacts and returns a seamless correction. Furthermore, both a qualitative and a quantitative comparison will show the differences between our proposed strategy and others found in literature, as well as the impact of the choice of a strategy upon the spatial distribution of a policy-relevant metric.

## 2 Data and methods

70 Further follows the description of three different strategies how to represent station data in the whole domain for the purposes of bias correction and the description of our experimental setup in terms of used data and methods. Each of the bias correction strategies has certain advantages as well as limitations. The strategies considered in this study are (i) statistically adjoint PDFs, (ii) direct interpolation of station data and (iii) parametric interpolation of quantile biases (PIQB).

### 2.1 Adjoint PDFs

75 The following strategy is the simplest one to implement and has been shown to be efficient in the conditions of coarse-resolution simulations (Rieder et al., 2015). From the statistical point of view, the idea is to treat all observations contained within the domain as random realizations of an experiment, which serve as the observational input for QM. Then, only those grid-cells containing stations are used as the corresponding modeled data.

Then, these two inputs form two PDFs which are compared to each other, which serves as the training process for QM. The 80 relationship between the PDFs is then extrapolated to the PDF of the rest of the domain, i.e., the full extent of modeled data within the domain. This step assumes representativeness of the limited number of stations, as well as of the corresponding biases, for the whole domain.

The main advantage of this strategy is its computational efficiency and that there is no need for any kind of interpolation. It 85 works very well at coarse resolutions where most model grid-cells contain at least one station and in the cases where the model correctly resolves large-scale spatial variability. However, in the cases of a thin station coverage, fine resolution or poorly resolved spatial distributions by the model, the results may contain statistical artifacts.

To explain such artifacts, let us consider an example where the observations do not cover areas with extreme values of the modeled data, which is typical for fine-resolution simulations. If the comparison of PDFs at station sites finds that the model 90 underestimates the width of its PDF, this strategy would then stretch the domain-wide modeled PDF, making minima lower and maxima higher regardless of their physical plausibility. Similarly, any errors in the large-scale spatial variability would be even more highlighted, as the strategy does not consider the possibility of having different biases throughout the domain.



## 2.2 Interpolation of observations

A different approach would be to truly rely on station data and their representativeness for the nearest neighboring grid-boxes. Such a strategy has been used to generate gridded datasets of observations (at a correspondingly coarse resolution) to correct  
95 the uncertainties of global models by Schnell et al. (2014).

The interpolation of station data into a regular model grid can be conducted using a weighted average with an appropriate weighting function, such as inverse distance weighting (IDW), i.e.,

$$\tilde{w}_k(d_k, \beta) = \frac{1}{d_k^\beta}, \quad (1)$$

where  $d_k$  is the distance to the  $k$ -th station and  $\beta$  is the interpolation parameter. For a total number of stations  $N$ , these weights  
100 are then normalized to have a unit sum using the following formula

$$w_k(d, \beta) = \frac{\tilde{w}_k(d_k, \beta)}{\sum_{i=1}^N \tilde{w}_i(d_i, \beta)} \quad (2)$$

and are used to obtain the resultant interpolated observation  $O$  in an arbitrary point  $p$  as

$$O^p = \sum_{k=1}^N w_k O^k. \quad (3)$$

The parameter  $\beta$  can be optimized based on various criteria, e.g., minimizing some residual validation statistic in cross-  
105 validation (Schnell et al., 2014).

The strategy would then be to interpolate the measurements into the model grid and perform bias correction in each grid-box individually. This way, the spatial distribution of observations is reflected in the corrected simulations, which adds the information about the large-scale spatial variability from observations compared to the previous strategy.

On the other hand, while the large-scale features may be resolved well, applying this strategy onto fine resolution simulations  
110 may result in suppressing some fine resolution model features in model grid-cells neighboring to the stations. This means that an area with locally lower ozone concentrations resolved by the model surrounded by stations measuring generally higher concentrations will be lost after the interpolation.

In other words, if a large grid-box contains multiple stations, its mean climate corresponds well to the mean climate of the stations. However, in the opposite situation when multiple fine grid-boxes are supposed to be represented by one station, this  
115 strategy fails, because the model provides more information about the spatial variability within the domain.

## 2.3 Parametric interpolation of quantile biases

Lastly, we propose a new strategy to address the above mentioned problems, which is fairly similar to the previous one, but we perceive it as a hybrid of both the strategy of adjoint PDFs and the interpolation of observations. The key assumption is that the physics and the chemistry of the atmospheric processes in terms of spatial distribution are well represented within the  
120 simulations and that the only issue is with the overall bias and mismatch in PDF shapes, i.e., the maxima and minima may not be representative enough within the model.



We can expand the issues with bias and variability by fusing them into one which we shall call quantile bias. Consistently with the mechanism of QM, we can assume that each quantile is biased differently. Additionally, we may even assume that the spatial distribution of quantile biases is suitable for interpolation from stations.

125 Considering the above mentioned assumptions, let us define quantile bias  $B_q^k$  of a quantile  $q$  at a station  $k$  as follows,

$$B_q^k = M_q^k - O_q^k, \quad (4)$$

where  $O_q^k$  and  $M_q^k$  are the  $q$ -th quantile of the  $k$ -th station of observations and its corresponding model value respectively. We shall interpolate this quantity from stations to an arbitrary point  $p$  using a weighted average with weighting functions  $w_k$ , i.e.,

$$B_q^p = \sum_{k=1}^N w_k B_q^k. \quad (5)$$

130 By expanding  $B_q^k$  in Eq. (5) by their definition in Eq. (4) and rearranging the symbols, we obtain

$$O_q^p = M_q^p + \sum_{k=1}^N w_k (O_q^k - M_q^k), \quad (6)$$

which shall serve as the interpolation formula to obtain pseudo-observations  $O_q^p$  in the regular model grid.

From Eq. (6) it should be somehow apparent why we claim this strategy to be a hybrid of the two previously presented ones. On one hand, it takes into consideration the spatial variability of the observed data, but on the other hand uses the model data 135 as a support to fill in the blank space in between the stations.

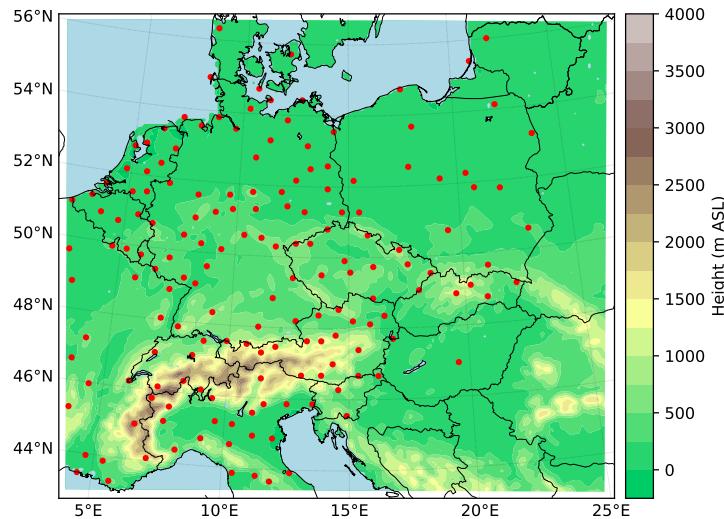
A similar idea has already been tested with a certain degree of success by Rahmani et al. (2017) in terms of absolute bias interpolation. The difference is that their work was focused on the biases in forecast models and the definition of quantile biases would be complicated with the limited amount of data. For this reason, their setup was not suitable for expanding to bias correction of climate projections, simply because longer simulated periods would be necessary.

140 The procedure then continues in the same way as in the previous strategy, which is the correction of each grid-box individually using QM. This also implies that only the pseudo-observed quantiles needed for QM are necessary to obtain from the interpolation by Eq. (6), which significantly reduces its computational demand.

The weights  $w_k$  can once again be chosen as IDW in Eq. (1), but that is not a choice we would recommend. Although IDW 145 is an exact interpolator, it is prone to errors (e.g., Achilleos, 2008; Gentile et al., 2012; Ohlert et al., 2023). Instead, we propose a smoother option on a similar basis, which is a gaussian weighting function, i.e.,

$$\tilde{w}_k(d_k, \alpha) = e^{-\alpha d_k^2}, \quad (7)$$

where  $\alpha$  is an interpolation parameter and the weights should be once again normalized to yield unit sum in an analogy to Eq. (2). A parameter figures in both presented weighting options, which allows for an optimal setting of the interpolation. Therefore, we shall name this overall strategy the Parametric Interpolation of Quantile Biases (PIQB).



**Figure 1.** The model domain of Central Europe with topography. Red dots mark all of the 165 stations considered in this study.

## 150 2.4 Model data and observations

Following the description of the statistical procedures, we shall now introduce the data used in this study. Starting with model data, we consider outputs of two CTMs: (i) the Weather Research Forecasting model with online coupling to chemistry (WRF-Chem, v. 4.0.3; Grell et al., 2005) and (ii) Comprehensive Air Quality Model with Extensions (CAMx, v. 6.50; ENVIRON, 2018). CAMx was offline coupled to WRF meteorology using the wrfcamx meteorological preprocessor, i.e., without considering feedbacks on model radiation (Huszar et al., 2012). The simulations were conducted in the period of 2007–2016 with horizontal resolution of 9 km in the domain of Central Europe with a spatial extend depicted in Fig. 1, consisting of  $190 \times 166$  grid-boxes in the meridional and zonal direction respectively. Our analysis is based exclusively on maximal diurnal averages of 8 hours of ozone concentrations (MDA8), which required a preprocessing of the original hourly output. The overall setup is identical to the one used in the study by Karlický et al. (2024), to which the reader is kindly referred to for further information.

160 From the simulations used by Karlický et al. (2024), we consider those using meteorological boundary conditions (BCs) taken from the ERA-interim reanalysis (Dee et al., 2011) and chemical BCs taken from the global model SOlar Climate Ozone Links (SOCOL; Sukhodolov et al., 2021). This choice is motivated by the possibility of using those simulations to calibrate corrections of climate projections, as this is the main purpose of bias correction. The possibility stems from the fact showed by Karlický et al. (2024) that the meteorological BCs only influence correlation, but not the overall bias, while chemical BCs only impact biases, and so this combination would correct the systematic errors in future climate simulations which rely on global model data.

Moving to the reference data, station measurements from 165 rural background stations of the European Environmental Agency (EEA) station network were taken and the data are freely available at <https://eeadmz1-downloads-webapp.azurewebsites.net>.



net/. The stations were chosen by an algorithm described by Karlický et al. (2024): first, only those stations with measurements 170 covering at least half of the simulation period were used. Then, a basic quality-control check was made for each station by counting the number of measured hourly concentrations lower than the seasonal mean decreased by 2.5-times the standard deviation of the hourly data. Several stations spiked with an unprecedented number of such values (in the order of thousands within two consecutive years) and were thus removed from the list. Lastly, an iterative elimination algorithm was used to homogenize the density of the stations based on their relative position, removing a station with the lowest mean distance to its 175 2 closest neighbors in each step. This way, 165 stations remained on the list and are marked in Fig. 1 by red dots. The spatial homogeneity of the station network is important for both the objectivity of the validation of the model performance and for the prevention of overfitting statistical models.

## 2.5 Methods of validation

Let us now proceed to the description of the optimization and evaluation process. For the optimization, we adopted cross-validation and its purpose is two-fold, as it allows us to optimize the interpolation parameters as well as to quantify the ability 180 of each strategy to represent the station information in the model grid. Here, the optimal value for the parameter  $\alpha$  in Eq. (7) was chosen for each month separately. To prevent overfitting, the data were divided into twelve 3-months long moving seasons in the full extent of the 10 years always with the middle month being the target period.

For each moving season, a brute-force 1-out cross-validation algorithm was adopted, which we shall now explain. First, a 185 station is chosen out of the  $N = 165$  stations considered, and removed momentarily. Then, the statistical correction within the chosen moving season is performed using the remaining  $N - 1$  stations with a chosen value of the parameter  $\alpha$ . The results are validated on that one missing station only in the middle month to obtain a residual value of a certain validation statistic which we describe later.

This process is then repeated with the fixed value of  $\alpha$  for all of the  $N$  stations. After that, the mean value of the chosen validation 190 statistic is taken. Repeating this procedure for different values of  $\alpha$  will return the empirical function of the dependence of the averaged residual validation statistic on the parameter  $\alpha$ . The optimal value for  $\alpha$  in each month is the minimizer of such a function for that given month.

As should be apparent, a suitable validation statistic for this task should be one that considers the overall match of the observed PDF  $f_O(x)$  and its corresponding statistical reconstruction  $f_M(x, \alpha)$ . Therefore, we chose the Hellinger distance 195 ( $H_D$ ; Jeffreys, 1946) describing the mismatch of two PDFs in the following form,

$$H_D(\alpha) = \frac{1}{2} \int_{-\infty}^{\infty} \left( \sqrt{f_O(x)} - \sqrt{f_M(x, \alpha)} \right)^2 dx, \quad (8)$$

where, in our case, the values of  $x$  would correspond to MDA8 of ozone, and  $H_D(\alpha)$  is the empirical function to be minimized.

The existence of a global minimum of  $H_D(\alpha)$  using normalized gaussian weights from Eq. (7) can be shown heuristically. Let us consider the extreme cases, i.e., (i)  $\alpha = 0$  and (ii)  $\alpha \rightarrow \infty$ . In the case (i), all stations would have the same normalized



200 weights, i.e.,

$$\forall k : \tilde{w}_k(d_k, \alpha = 0) = e^0 = 1 \Rightarrow \forall k : w_k(d_k, 0) = \frac{e^0}{\sum_{i=1}^N e^0} = \frac{1}{N}, \quad (9)$$

where the implied values of  $w_k$  follow from Eq. (2).

In the case (ii), only the station with the minimal distance to the interpolation point  $d_m = \min_k d_k$  would have a non-zero weight, which follows from the following limit:

$$205 \quad \forall k : w_k(d_k, \alpha \rightarrow \infty) = \lim_{\alpha \rightarrow \infty} \frac{e^{-\alpha d_k^2}}{\sum_{i=1}^N e^{-\alpha d_i^2}} = \lim_{\alpha \rightarrow \infty} \frac{e^{-\alpha(d_k^2 - d_m^2)}}{\sum_{i=1}^N e^{-\alpha(d_i^2 - d_m^2)}} = \begin{cases} 1 & \text{if } d_k = d_m \\ 0 & \text{otherwise} \end{cases} \quad (10)$$

As should be apparent, these extreme cases are not suitable for non-trivial spatial variability of the interpolated quantity, or in the case of PIQB, station quantile biases with respect to the model. Therefore, values of  $\alpha$  between those extreme cases consider spatial differences and should return lower values of  $H_D(\alpha)$ . Since the weights are continuous and smooth in  $\alpha$ , the function  $H_D(\alpha)$  should be as well. Its smooth character can in theory only be perturbed by short intermittent discontinuous oscillations 210 caused by QM due to the movement of values through quantile boundaries when different thresholds are interpolated using close values of  $\alpha$ , but such an effect should be negligible. Similar conclusions can be made for IDW from Eq. (1). Since the distances  $d$  were treated as euclidean distances between model grid-boxes, they were proportionally large and the optimal values of  $\alpha$  were then expected to be roughly of the order of  $10^{-3}$ . This estimate stems from a simple calculation where, 215 for  $\alpha = 10^{-3}$ , the diameter of the 1-sigma interval of the gaussian (coinciding with its inflection point) would be roughly 90 grid-cells, which is approximately half of the size of the domain. Greater values of  $\alpha$  reduce this diameter.

To summarize the strategies considered and cross-validated in this study, all of them are briefly described in Tab. 1. Please note that Adjoint PDFs does not consider any interpolation, as it functions on correcting the PDF in the entire domain as a whole. Let us further remind the reader that the cross-validation itself quantifies and compares the performance of each strategy to extrapolate data outside of the station sites through the minimal value of mean residual  $H_D$ .

220 Lastly, besides  $H_D$ , we shall further evaluate the performance of each strategy at station sites using several validation statistics. We shall quantify the model bias by normalized mean bias (NMB; Im et al., 2015) which describes the position of the modeled PDF relative to the observed PDF. Then, the mismatch in variability shall be quantified by a simple ratio of the modeled and observed standard deviation  $\sigma_M/\sigma_O$ . Similarly, the slope of the linear regression through the scatterplot 225 of modeled and observed data (Slope; Karlický et al., 2017) describes the mismatch in variability with higher sensitivity to outliers, most useful for the validation of policy-relevant metrics. Finally, to compare the spatio-temporal correlation, the Pearson correlation coefficient is used.

### 3 Results and discussion

Let us now proceed to the evaluation of the performance of the CTMs, the optimization of the parameter  $\beta$  in Eq. (1) and  $\alpha$  in Eq. (7) and finally the validation of the postprocessed simulations using different strategies. The validation was performed



**Table 1.** The list of the four different bias correction strategies evaluated in this study highlighting the differences in interpolated quantities and weighting functions.

Strategy	Interpolation	Weights
Adjoint PDFs	–	–
Obs. IDW	observations	IDW
PIQB IDW	quantile biases	IDW
PIQB gauss	quantile biases	gaussian

230 in terms of climatological seasons and focused on the qualitative shapes of the PDFs and modeled variability. As mentioned before, the analysis was performed on MDA8 of ozone, which were obtained from hourly data.

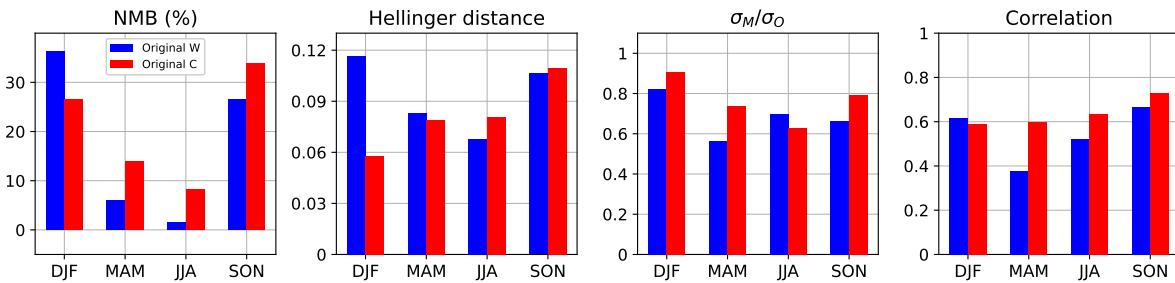
### 3.1 Validation of original simulations

Prior to postprocessing, a brief validation of the original simulations follows. Fig. 2 shows 4 validation statistics for the original simulations performed with models WRF-Chem and CAMx. The validation shows a strong seasonal dependency of NMB for 235 both models with higher NMB in DJF and SON and lower NMB in MAM and JJA.

Consistently with Revell et al. (2018), the global chemistry model SOCOL overestimates tropospheric ozone concentrations in our latitudes, which may then translate into the nested domain, as reported by Karlický et al. (2024). This means that the model bias in this case most probably originates from the BCs. As stated before, ozone bias is a common problem even in other studies (e.g., Grell et al., 2005; Racherla and Adams, 2008; Mar et al., 2016; Karlický et al., 2017; Karlický et al., 2024) and 240 cannot be simply prevented in CTMs.

The variability of the modeled data is weaker in all seasons when compared to the observations, as seen in  $\sigma_M/\sigma_O$ . This shows that simple bias removal may not fully correct all the shortcomings of the modeled data. Consequently, despite being theoretically bound by 1, the Hellinger distance of 0.06–0.12 is still relatively high given that the simulations are supposed to approximate the atmospheric processes in a detailed way. The origin of its values can be partially seen in both NMB by 245 shifting the modeled PDFs and the weaker variability by underestimating the widths of the PDFs, resulting in poor overlap of the modeled and observed PDFs. Additionally, the mismatch in higher moments of the PDFs may contribute as well, but they are not quantified here.

Finally, the correlation of modeled data and observations varies more notably for WRF-Chem than for CAMx, but is overall similar for both simulations. Correlation of ozone concentrations within 0.4–0.7 is well expected for CTMs of such resolutions 250 (Grell et al., 2005; Mar et al., 2016; Karlický et al., 2017; Karlický et al., 2024). A more detailed validation can be found in the original work of Karlický et al. (2024).



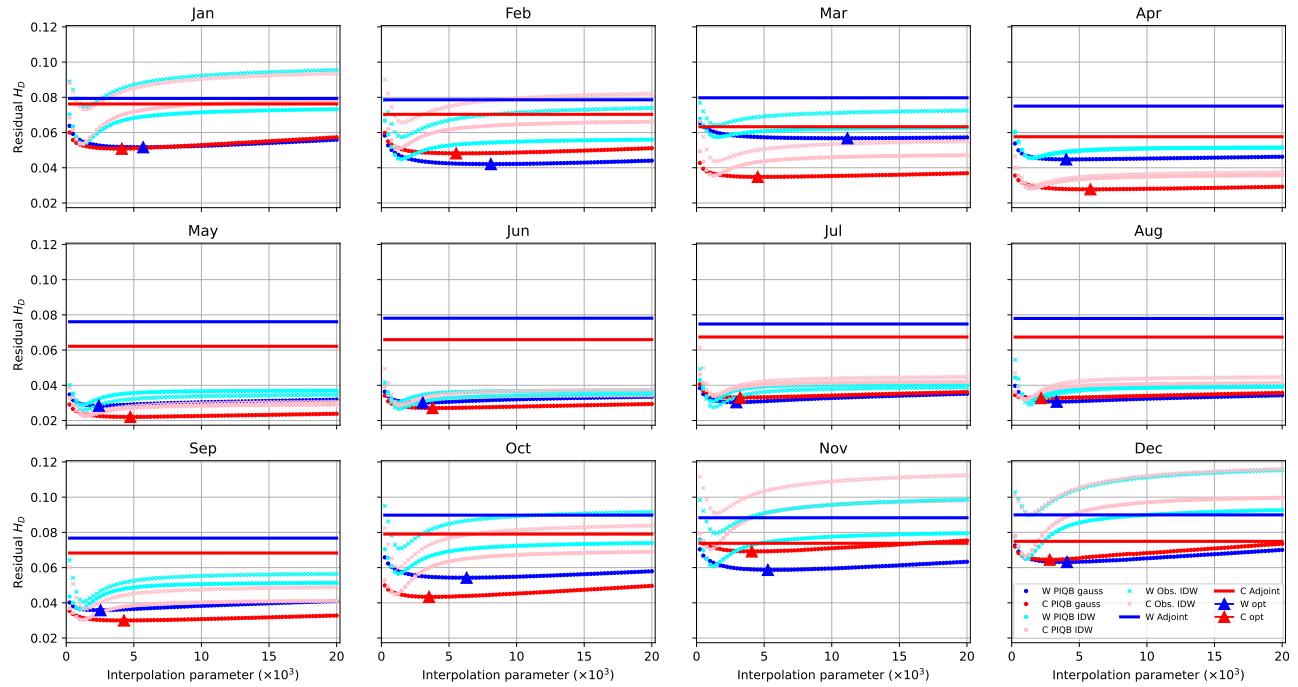
**Figure 2.** Seasonal validation of the original simulations using WRF-Chem (W; blue) and CAMx (C; red) against station data, from left to right: normalized mean bias (NMB), Hellinger distance, ratio of standard deviations of modeled and observed data respectively ( $\sigma_M/\sigma_O$ ) and Pearson correlation.

### 3.2 Calibration of free parameters and cross-validation

After confirming the presence of systematic errors, we proceeded with the cross-validation of each of the bias correction strategies considered using the methodology explained above. This was done to obtain the optimal values for the parameters  $\beta$  and  $\alpha$  in Eq. (1) and Eq. (7) respectively, but also to evaluate the ability of the strategies to reconstruct pseudo-observed PDFs at station sites which were omitted during calibration.

The empirical functions of the mean residual  $H_D$  were constructed with a step of  $0.25 \times 10^{-3}$  of the interpolation parameter for the range between  $0.25$ – $20 \times 10^{-3}$ . The minima of functions corresponding to the strategies were found in each month by fitting nearby data points with a 3rd-degree polynomial and finding its minimum. Results for each of the strategies and for both simulations are shown in Fig. 3. Few important observations can instantly be made, beginning with the fact that all empirical functions are truly smooth and continuous as has been assumed, with the interpolating strategies having always exactly one minimum. The optimal values of the interpolation parameter for the gaussian weights were chosen as marked in Fig. 3. For the strategies PIQB IDW and Obs. IDW, the parameter  $\beta$  was optimized in the same way, but is not displayed in Fig. 3 to preserve comprehensiveness. The parameter  $\beta$  was ranging between  $1$ – $1.5 \times 10^{-3}$ .

Another interesting remark is better visible in Fig. 4 where the optimal residual  $H_D$  is displayed for each month.  $H_D$  of optimized interpolating strategies exhibit an annual cycle with lower values in the warmer part of the year and the opposite in the colder part, while  $H_D$  of Adjoint PDFs does not. In this sense, it is not surprising that Adjoint PDFs performs the worst out of all of the four considered strategies with a weak annual cycle throughout the year, except for November and December for the CAMx simulation, as it forcefully increases the variability of the modeled data in the whole domain regardless of their spatial variability, which yields similar residual errors in each month. Furthermore, larger differences between interpolating strategies can be seen in the colder part of the year, while the opposite is true for the warmer part. This is because ozone concentrations correspond less to orography in summer than in winter, not only making corresponding quantities easier to interpolate, but also highlighting the importance of orographic features for a successful interpolation. All these remarks conclude that the spatial



**Figure 3.** Empirical monthly functions of mean residual Hellinger distance  $H_D$  between the observed PDFs and their statistical reconstructions for each cross-validated station site, both WRF-Chem (W; blue) and CAMx (C; red) simulations and each of the 4 strategies: PIQB gauss (dark dots), PIQB IDW (light dots), Obs. IDW (light crosses) and Adjoint PDFs (dark lines). The optimal values of the parameter  $\alpha$  for PIQB gauss are marked by triangles.

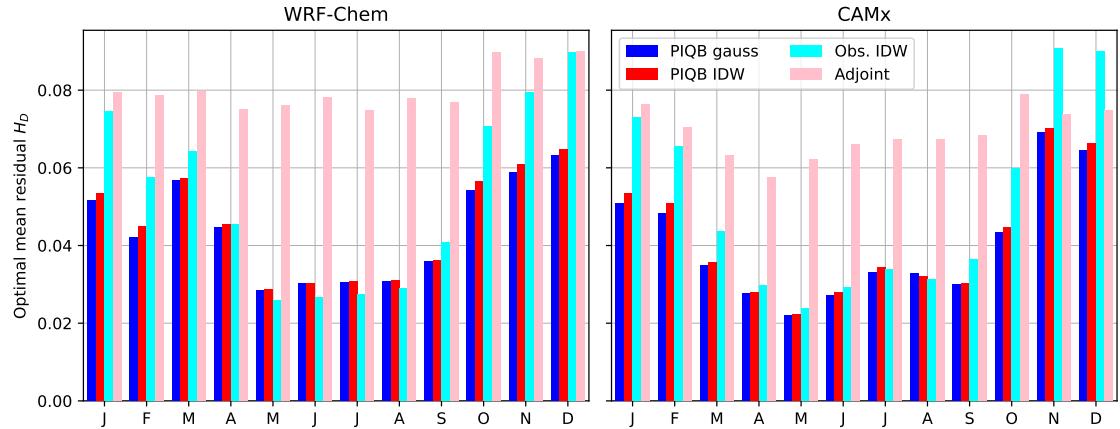
variability of the model errors plays a crucial role for bias correction. This is an important result to consider, because it shows

275 the ability of PIQB to reliably reconstruct station PDFs at places with no stations.

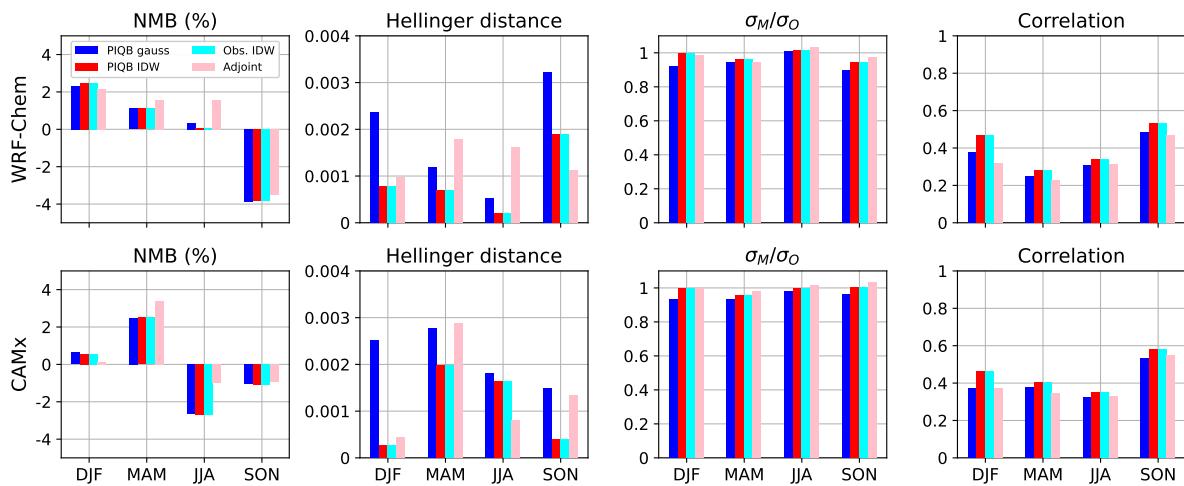
### 3.3 Validation of postprocessed simulations

Let us now shift our attention to the performance of the bias correction strategies listed in Tab. 1 using the optimal parameters for their calibration. A basic seasonal validation against the data at all 165 stations is depicted in Fig. 5, as an analogy to Fig. 2 (note the different scales for NMB and Hellinger distance).

280 Each of the bias correction strategies managed to improve NMB, Hellinger distance and the ratio of the standard deviations in each season for both simulations. The values of NMB decreased from the range of 1.5–36 % to –3.8–3.3 % and its seasonal dependency changed as well. Additionally,  $H_D$  decreased in terms of orders from  $6–11 \times 10^{-2}$  to  $2–32 \times 10^{-4}$ , owing to the mechanism of QM which well approximates the observed PDFs. On the other hand, the correlation decreased, because neighboring months were used to correct each month as well, thus permuting the order of the data in model time series.



**Figure 4.** Mean residual Hellinger distance in the empirical minima in Fig. 3 in each month for both the WRF-Chem (W; left) and CAMx (C; right) simulations and for each considered strategy, i.e., PIQB gauss (blue), PIQB IDW (red), Obs. IDW (cyan) and Adjoint PDFs (pink).



**Figure 5.** Seasonal validation of the postprocessed simulations using optimized strategies of bias correction against all station data: PIQB with gaussian (blue) and IDW (red) weights, Obs. IDW (cyan) and Adjoint PDFs (pink). Top row shows experiments performed on the WRF-Chem simulation, bottom row on CAMx. When comparing to Fig. 2, note the differences in scales for NMB and Hellinger distance.



285 To address the qualitative performance of each strategy, both strategies using IDW return the same validation statistic in each sub-figure of Fig. 5 despite using a different interpolation formula. This is because IDW is an exact interpolator and so the residual errors for both PIQB IDW and Obs. IDW could be practically perceived as the residual errors of QM. Therefore, if a strategy performs better than those using IDW, it is out of sheer luck. The reason for the errors to be non-zero, e.g., in the 290 case of NMB, is the way the simulations were corrected – using the neighboring months as well for calibration, but only the middle month in the result to prevent overfitting. This limitation is a shared feature for each strategy.

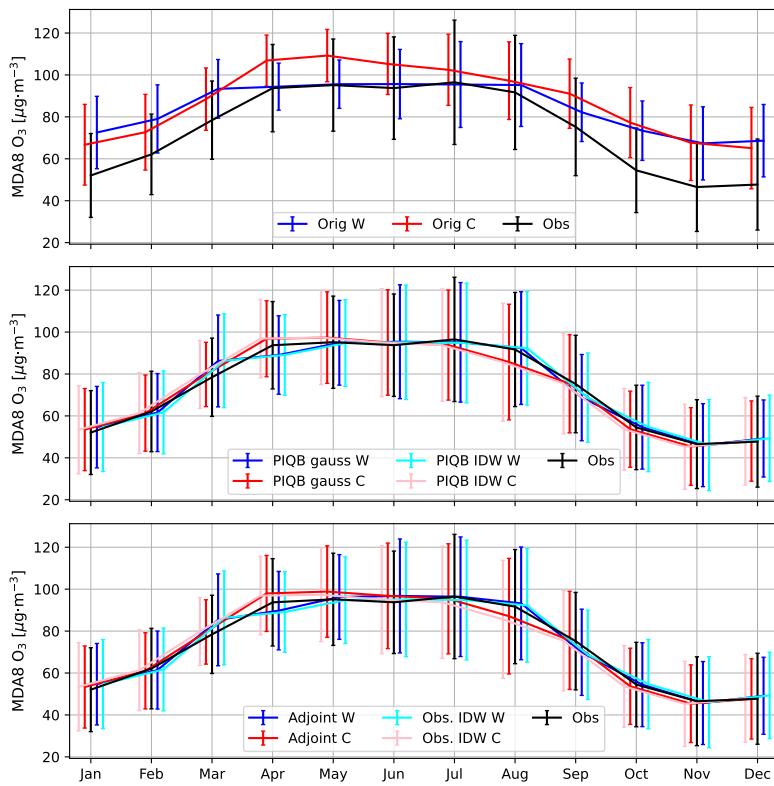
While the validation results for PIQB IDW and Obs. IDW in Fig. 5 are theoretically indifferent to the choice of the optimal interpolation parameter, the opposite is true for PIQB gauss because it does not use an exact interpolator. It removes NMB comparably to the strategies using IDW with a slightly worse performance and similarly it still underestimates  $\sigma_M/\sigma_O$ . The residual errors of PIQB gauss have a similar seasonal dependency as PIQB IDW and Obs. IDW, suggesting that a portion of 295 the residual errors may stem from the interpolation itself.

The strategy Adjoint PDFs is the only strategy not using interpolation, and so its seasonal dependency is different as well. This is only logical, as the error could potentially be random, because modeled and observed data from the whole domain are considered within the correction at each station and all with the same weight. Still, it performs relatively well in terms of NMB and  $\sigma_M/\sigma_O$  compared to other strategies.

300 To compare the strategies further, seemingly contradictory to the conclusions drawn from Fig. 4, PIQB gauss leaves relatively high residual errors compared to the other strategies. Naturally, this is because it does not use an exact interpolator and Fig. 5 does not provide any information about the performance outside of the station sites. In other words, Fig. 5 (showing the residual errors at places with stations) displays fundamentally different measures than Fig. 4 (showing the performances outside the station sites). Nevertheless, the errors of PIQB gauss in terms of  $H_D$  are still comparable or even lower in Fig. 5 than in Fig. 4, 305 but the decisive figure for the performance of utilizing the information from station data should be Fig. 4

As the correction was made per months, we can compare the discrepancies of the original and corrected model data from the 310 observations in terms of the annual cycle of MDA8. Fig. 6 shows the annual cycles of the observations and the original simulations as well as their corrections. After the correction, all strategies perform very similarly, reducing the bias and increasing the monthly variability, thus adjusting the annual amplitudes as well. It is interesting to note that the qualitative shapes of the annual cycles did not change for either of the simulations. This can be observed on the fact that throughout all bias correction strategies, the corrected CAMx simulation still exhibits residual bias most notable in Apr and Aug, while the WRF-Chem simulation retained a sharp increase of MDA8 from Feb to Mar as well as a subtle increase from Mar to Apr. These residual errors are slight and very practical. They once again prove that the statistical procedures do not overfit, therefore suggesting a possibility to use them for correcting climate projections while retaining the qualitative character of the modeled data.

315 Fig. 6 also visualizes the main flaws in the original simulations, which helps to understand the key relationships between the validation statistics displayed in Fig. 2 and Fig. 5. Both simulations exhibit a systematic bias in the annual cycle, as well as shortage of monthly variability (i.e., the lengths of error bars). And so joining both types of discrepancies into a common problem in the shapes of the PDFs leads us to the motivation behind minimizing  $H_D$  in the calibration phase of the bias correction.

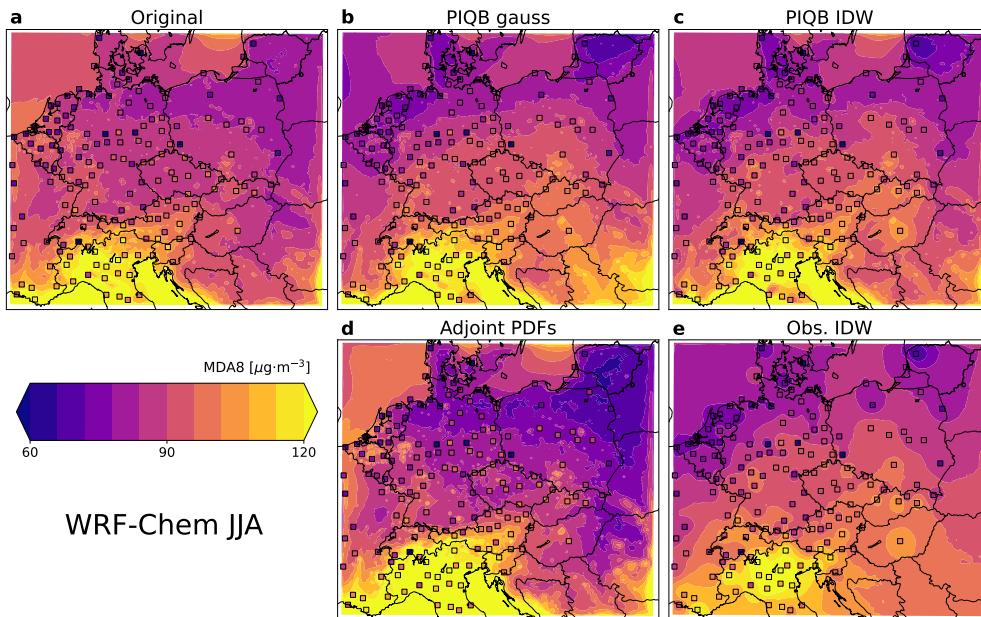


**Figure 6.** The mean and the standard deviations of monthly MDA8 of the observations (black) and original and processed simulations corresponding to WRF-Chem (W; blue) and CAMx (C; red). The top figure shows the original data, the middle figure shows PIQB gauss (dark) and PIQB IDW (light) and the bottom figure shows the correction by Adjoint PDFs (dark) and Obs. IDW (light).

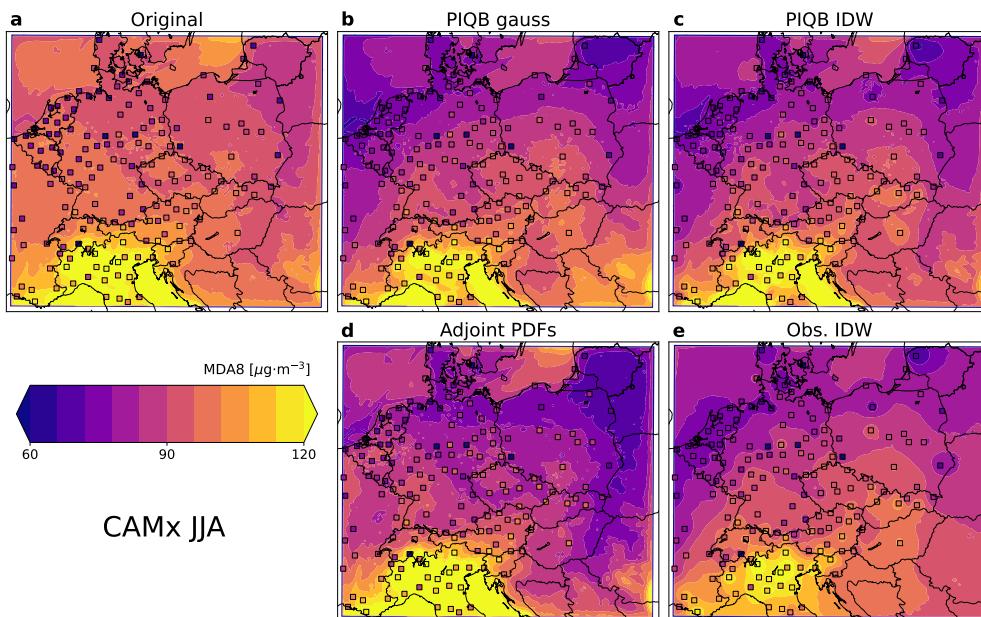
320 To fully understand the differences between the considered strategies of bias correction outside the station sites, one must look at the resolved spatial distributions after correction by each of them. Fig. 7 shows the spatial distribution of mean summertime MDA8 in the WRF-Chem simulation before and after correction with marked stations, while Fig. 8 shows the same for the CAMx simulation.

325 The following claims apply to both simulations and their conclusions are apparent in both Fig. 7 and Fig. 8. The original simulations do not take into account for the real observed spatial variability, showing a relatively similar meridional gradient of mean MDA8, but exactly opposite zonal gradient than the observations. Once the strategy of Adjoint PDFs is applied, this difference in gradients only becomes highlighted. That is because the Adjoint PDFs is only calibrated on the validation displayed in Fig. 2, i.e., the calibration tries to increase the variability of the model data to fit the observations, which results in a stronger, yet still incorrect, zonal gradient.

330 The strategy Obs. IDW results in fitting the observations almost perfectly, since IDW is an exact interpolator. Therefore, the zonal gradient fits better than in the case of Adjoint PDFs, showing the lowest MDA8 in the nort-west corner of the domain



**Figure 7.** Mean summer-time ozone MDA8 in the model domain as simulated by the WRF-Chem simulation (a) and their corresponding corrections using the strategies of PIQB gauss (b), PIQB IDW (c), Adjoint PDFs (d) and Obs. IDW (e). Colored squares mark measured values.



**Figure 8.** Same as Fig. 7, but for the CAMx simulation.



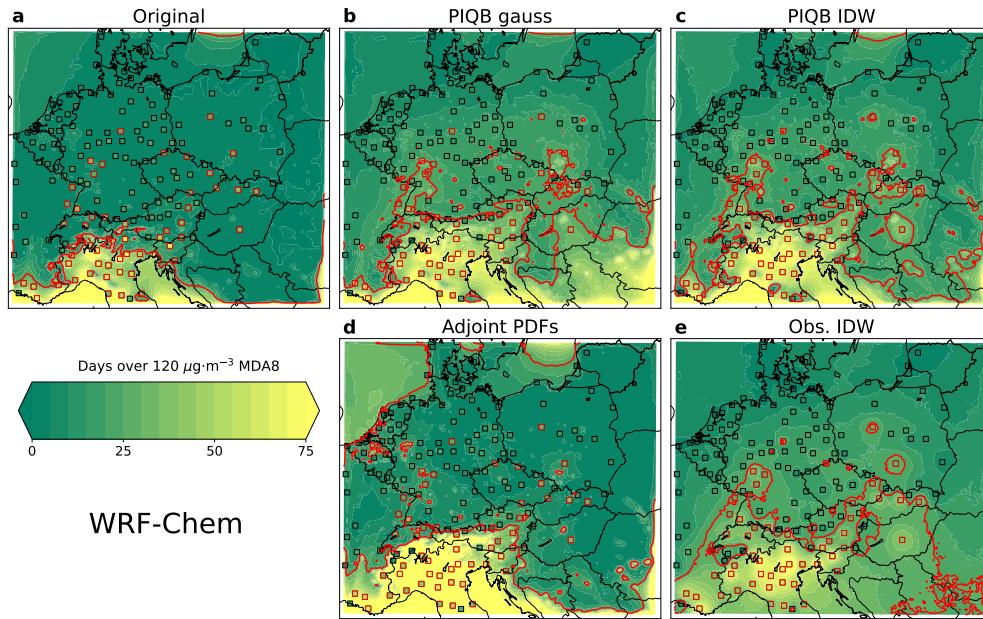
instead of the north-east. However, this strategy smoothed out the mean MDA8 field completely, omitting for the spatial variability described by the model, most notably seen in the south-east corner of the domain with a limited number of stations. Please note that the plots of Obs. IDW in Fig. 7e and Fig. 8e look different for both simulations, despite both describing the 335 result of interpolating measurements. That is because after the interpolation of station data (using different parameters), QM was applied. Therefore, the differences between these two figures can be attributed to the differences between the interpolation parameters  $\beta$ , the differences between the model simulations and the precision limits of QM.

Moving on to the interpolation of quantile biases, PIQB IDW not only corrects the zonal gradient of mean MDA8 in both Fig. 7c and Fig. 8c, but also preserves the qualitative spatial variability as modeled by the WRF-Chem and CAMx simulations, 340 which is one of the characteristics of PIQB in general. To compare it with Obs. IDW, despite having the same residual errors in Fig. 5, they perform very differently in terms of cross-validation in Fig. 4, which is apparent in the spatial distributions as well. The key differences between PIQB IDW and Obs. IDW are better visible in cases where the spatial distribution of mean MDA8 is less smooth, e.g., winter periods in Fig. A1 and Fig. A2, where considering the modeled spatial variability results 345 in resolved orographical effects, such as higher concentrations in the Carpathian mountains. This also demonstrates why the residual cross-validation errors in Fig. 4 are much higher in the winter period for Obs. IDW compared to PIQB IDW or PIQB gauss.

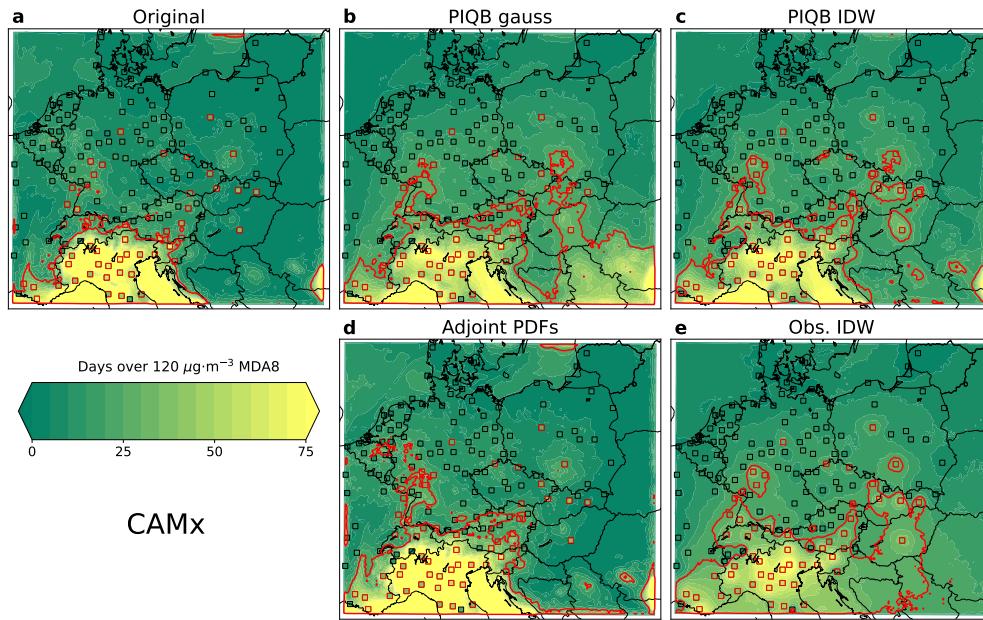
The last strategy is PIQB gauss in Fig. 7b and Fig. 8b, which uses the same interpolation formula as PIQB IDW, i.e., 350 interpolation of quantile biases, but a smoother weighting function. As a result, a sacrifice was made in terms of not fitting the observations perfectly, as seen in Fig. 5. In the JJA case, the mean interpolation error in the whole domain for PIQB gauss in terms of Hellinger distance within cross-validation is not reliably the lowest of the four considered strategies, but is comparable 355 (and consistently lower in other months) to those of Obs. IDW and PIQB IDW. This is because not using an exact interpolator in this case preserves more of the model structure, which is important for orographic effects more notable in the colder parts of the year seen by the models.

A general feature among the correction strategies is the correction of model variability and spatial distributions, which is 360 relevant for the spatial distribution of policy-relevant metrics as well. Fig. 9 shows the mean annual number of days with MDA8 exceeding  $120 \mu\text{g}\cdot\text{m}^{-3}$  in the WRF-Chem simulation before and after correction, while Fig. 10 shows the same for the CAMx simulation. The immission limit of ozone does not allow for the annual number of such days to be higher than 25 on average in 3 consecutive years (EUR-Lex, 2015). As we are presenting climate simulations, this average was taken in the 365 whole 10-year-long period for both the station and model data and the threshold exceedance is marked red.

In both model cases in Fig. 9a and Fig. 10a, it is apparent that the original simulations have very limited variabilities and the 370 area of exceeded immission limit spans only to the north of Italy, while in the case of observations, many of the stations with exceedances are located also further north and east in the center of the domain. Upon correcting the simulations using Adjoint PDFs, this area spreads further, mostly to the north and west, resulting in some exceedances even in Austria and Germany. In the case of the WRF-Chem simulation in Fig. 9d, Adjoint PDFs hints exceedances even over sea in the north-west corner of the 375 domain, showing how this strategy does not consider the spatial distribution of observations, but only increases the variability of the model data in the whole domain.



**Figure 9.** Mean annual number of days above  $120 \mu\text{g} \cdot \text{m}^{-3}$  MDA8 of ozone on average in 2007–2016 in the WRF-Chem simulation before (a) and after correction using the strategies of PIQB gauss (b), PIQB IDW (c), Adjoint PDFs (d) and Obs. IDW (e). The immission limit exceedance defined by the threshold of 25 days is marked red and took place at 53 station sites.



**Figure 10.** Same as Fig. 9, but for the CAMx simulation.



In the case of Obs. IDW, the red contour in Fig. 9e and Fig. 10e spreads further and envelops the stations with exceedances almost perfectly. However, there are some stations in Poland and Germany which cannot be attributed to this spread directly, despite being enveloped by the contour as well. This is the result of overfitting the correction with station data, disregarding  
370 the modeled spatial variability.

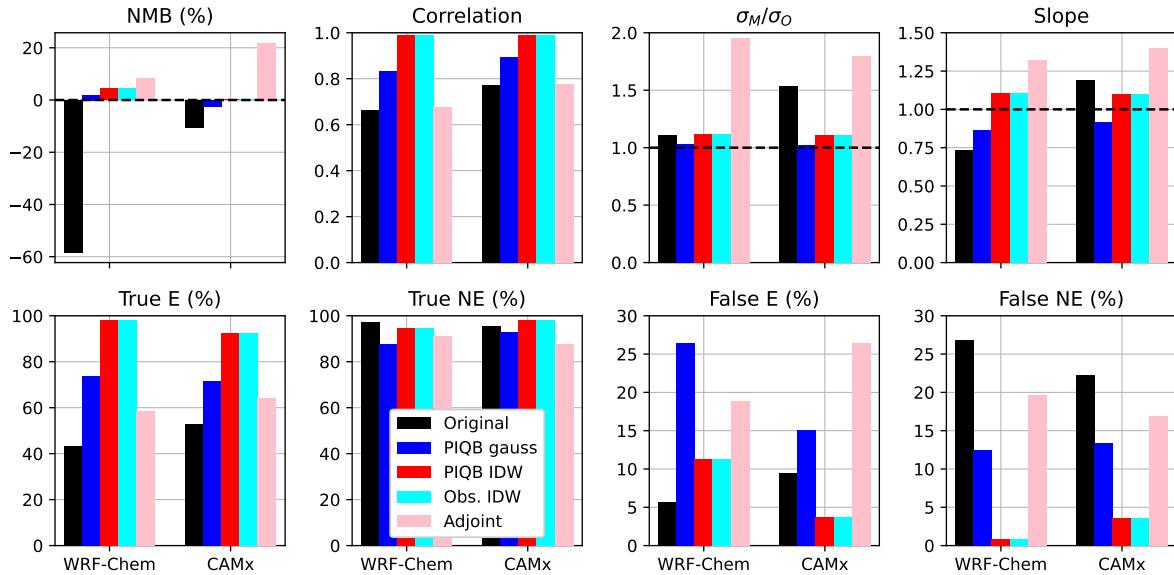
This shortcoming is partly solved by switching to a different interpolation formula, i.e., to PIQB IDW. The contoured borders in Fig. 9c and Fig. 10c look more natural and it is apparent that the number of days have sharper gradients within the contoured area, hinting a better use of the model spatial variability during the correction. Nevertheless, there are some singular contoured points in the eastern part of the domain which cannot be linked directly to the southern area of exceedance.

375 Lastly, PIQB gauss solves this problem, connecting the singular points without overfitting the correction with station data. In terms of the ability to qualitatively resolve the spatial distribution of the immission limit exceedances, PIQB gauss shows the most plausible results for both simulations. The key difference between PIQB gauss and PIQB IDW is in the south-eastern part of the domain with limited number of stations, where the gaussian interpolation assumes generally more days above the threshold than IDW.

380 To quantify the differences in the number of days exceeding the immission limit, one may treat this quantity as a variable which can be validated. Such a validation is displayed in Fig. 11 for both before and after correction. As is apparent, all strategies mitigate the bias in the number of days of exceedance, except for Adjoint PDFs in the CAMx simulation, showing how the unrealistically high mean concentrations over northern Italy in the original simulation transformed into the overestimation of exceedances in the rest of domain after correction. Contrary to Fig. 2 and Fig. 5, the correlations in Fig. 11 (spatial correlations) are expectedly increased by all considered strategies, most notably by both Obs. IDW and PIQB IDW as they use an exact interpolator.

390 Regarding variability, the ratio of standard deviations  $\sigma_M/\sigma_O$  gets closer to the perfect match of 1 in all cases except for Adjoint PDFs. The reason is that Adjoint PDFs forces the variability of MDA8 to be greater than the original model data in the whole domain, regardless of the station spatial distribution of MDA8. This results in an expansion of the threshold contour in the south of the domain in Fig. 9d and Fig. 10d while retaining relatively low values in the center of the domain, thus inflating the variability of immission limit exceedances. Discussing the variability further, the slope statistic of the original WRF-Chem and CAMx simulations have the exact opposite shortcoming relative to the perfect match of 1 and all strategies compensate it at least partially, except for Adjoint PDFs which results in its overestimation for similar reasons as in the case of  $\sigma_M/\sigma_O$ .

400 In the lower panel of Fig. 11, the success rates of predicting an immission limit exceedance (E) or non-exceedance (NE) are displayed, which is the percentage of the count of binary E/NE at station sites. It is clear that each strategy enhances the original results, increasing the true predicted E from 40–50 % to almost 60–100 %. On the other hand, predicting true NE is generally slightly worse after the corrections, which is complementarily seen in the overestimation of false E, i.e., mistakenly predicting an E at a station site without it. The overall predictions were better in terms of true NE and false E than in the case of true E and false NE, which could be due to the fact that the number of stations with a true E is 53 while without it is 112, and so the success rates were more sensitive to errors.



**Figure 11.** Validation of number of days above the  $120 \mu\text{g}\cdot\text{m}^{-3}$  threshold (upper panel) using NMB, Pearson correlation,  $\sigma_M/\sigma_O$  and the slope of linear regression in the scatterplot of observations and modeled values. Lower panel displays predictions of the immission limit exceedances in terms of the success rate of true/false predicted exceedance (E/NE; i.e., elements of the confusion matrix). The black dashed line marks the ideal match. Note the differences in scales in the lower panel between True and False.

Regarding differences between correction strategies in Fig. 11, both Obs. IDW and PIQB IDW expectedly outperform all of the other strategies when validated on station data, as they use an exact interpolator. PIQB gauss in this sense performs worse than that, but still better than Adjoint PDFs in terms of predicting a true E. On the other hand, the true NE and false E seem to be more model dependent in terms of differences between strategies, showing worse performance of PIQB gauss compared to 405 Adjoint PDFs for the WRF-Chem simulation, while the opposite is true for the CAMx simulation. This difference in particular can be attributed to the different spatial variabilities predicted by both models.

#### 4 Conclusions

In this study, we presented a newly introduced strategy to use station data for bias correction of MDA8 of tropospheric ozone concentrations and compared the performance with other strategies applied on the outputs of two 10-year-long simulations 410 made by the WRF-Chem and CAMx models. Our strategy involved the interpolation of station quantile biases into the regular model grid and we demonstrated its flexibility using two different weighting functions for spatial interpolation, i.e., IDW and gaussians.

First, we tested the ability of each strategy to interpolate station data using methods of cross-validation. The results showed that PIQB generally leaves the least minimal residual Hellinger distance, which demonstrates that it performs the best in terms



415 of reconstructing the observed PDFs at places with no observations, while the overall better interpolation option for PIQB in the whole year was the gaussian interpolator.

Regarding residual errors after full correction at station sites, the errors and differences between strategies were substantially smaller compared to those resulting from cross-validation, while the best performing strategies were Obs. IDW and PIQB IDW. This was hardly surprising, as IDW is an exact interpolator and the corrected simulations were validated this way only at station 420 sites. Since the simulations corrected by both strategies yield identical residual errors, one could come to a wrong conclusion that both strategies perform the same. To avoid such conclusions, either a cross-validation or maps of mean MDA8 must be analyzed.

The spatial distributions of MDA8 using different strategies showed that Adjoint PDFs do not add any new information regarding spatial distribution of station data. In contrast to that, Obs. IDW resulted in a better agreement with stations, but 425 omitted modeled features. In this sense, both PIQB IDW and PIQB gauss represented a middle ground between the two, resolving modeled features while preserving the spatial variability of station data.

Evaluating the performance of each strategy on a derived quantity, such as number of days with MDA8 above the  $120 \mu\text{g}\cdot\text{m}^{-3}$  threshold, may be more convincing. Even in this case, both PIQB IDW and PIQB gauss mostly outperformed the other two strategies, i.e., Obs. IDW and Adjoint PDFs, in terms of resolving continuous contours of immission limit exceedances. In this 430 sense, the success rate of both simulations to predict an immision limit exceedance improved after the correction by all of the strategies, but most dramatically by PIQB IDW, which preserved modeled features as well.

Ultimately, our findings show that PIQB performs satisfactorily well at a relatively high resolution, making it a suitable strategy to use station data for bias correction of chemistry-climate simulations. Results in this study thus suggest the possibility of using ground based station data to reliably correct systematic errors in climate projections of ozone MDA8.

435 *Code and data availability.* The complete input ozone surface data selected from model outputs used in the study are available at Czech National Repository via <https://doi.org/10.48700/datst.ev7ej-gv255> (Karlický, 2022). The output ozone MDA8 data and the python scripts used to generate and visualize them are available at Zenodo via <https://doi.org/10.5281/zenodo.17881911> (Peiker, 2025). Both repositories possess the Creative Commons Attribution 4.0 International Licence.

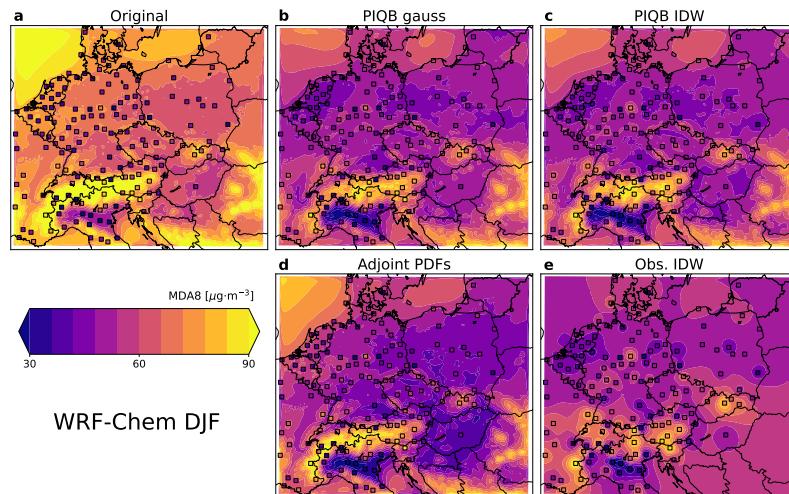
440 *Author contributions.* JP: methodology, formalization, statistical analysis, postprocessing, visualization and writing. JK: methodology, preparation of the WRF-Chem simulation, supervision, writing and reviewing. PH: preparation of the CAMx simulation and supervision.

*Competing interests.* The authors declare that they have no conflict of interest.

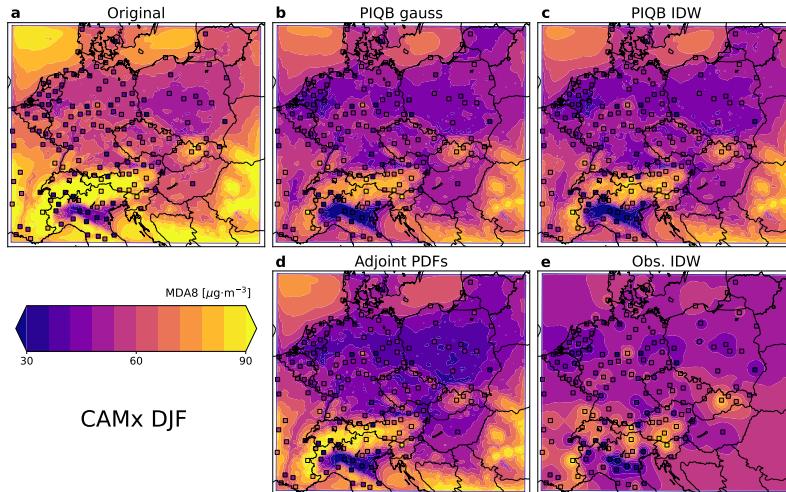


*Acknowledgements.* We acknowledge the European Environmental Agency for providing free air quality station data used in this study for the validation and postprocessing of used simulations (<https://eemdzm1-downloads-webapp.azurewebsites.net/>, last access 12 December 2025). The present work was supported by the Charles University Grant Agency (project no. 418625), by the Ministry of Education, Youth and Sport (grant no. CZ.02.01.01/00/22\_008/0004605) and by the Technology Agency of the Czech Republic (grant no. SS02030031).

## Appendix A



**Figure A1.** Mean winter-time ozone MDA8 in the model domain as simulated by the WRF-Chem simulation (a) and their corresponding corrections using the strategies of PIQB gauss (b), PIQB IDW (c), Adjoint PDFs (d) and Obs. IDW (e). Colored squares mark measured values.



**Figure A2.** Same as Fig. A1, but for the CAMx simulation

## References

Abdelmoaty, H. M., Rajulapati, C. R., Nerantzaki, S. D., and Papalexiou, S. M.: Bias-corrected high-resolution temperature and precipitation projections for Canada, *Scientific Data*, 12, 191, <https://doi.org/10.1038/s41597-025-04396-z>, 2025.

450 Achilleos, G.: Errors within the Inverse Distance Weighted (IDW) interpolation procedure, *Geocarto International*, 23, 429–449, <https://doi.org/10.1080/10106040801966704>, 2008.

Archibald, A. T., Neu, J. L., Elshorbany, Y. F., Cooper, O. R., Young, P. J., Akiyoshi, H., Cox, R. A., Coyle, M., Derwent, R. G., Deushi, M., Finco, A., Frost, G. J., Galbally, I. E., Gerosa, G., Granier, C., Griffiths, P. T., Hossaini, R., Hu, L., Jöckel, P., Josse, B., Lin, M. Y., Mertens, M., Morgenstern, O., Naja, M., Naik, V., Oltmans, S., Plummer, D. A., Revell, L. E., Saiz-Lopez, A., Saxena, P., Shin, Y. M., Shahid, I., 455 Shallcross, D., Tilmes, S., Trickl, T., Wallington, T. J., Wang, T., Worden, H. M., and Zeng, G.: Tropospheric Ozone Assessment Report: A critical review of changes in the tropospheric ozone burden and budget from 1850 to 2100, *Elementa: Science of the Anthropocene*, 8, 034, <https://doi.org/10.1525/elementa.2020.034>, 2020.

Colette, A., Andersson, C., Baklanov, A., Bessagnet, B., Brandt, J., Christensen, J. H., Doherty, R., Engardt, M., Geels, C., Giannakopoulos, C., Hedegaard, G. B., Katragkou, E., Langner, J., Lei, H., Manders, A., Melas, D., Meleux, F., Rouil, L., Sofiev, M., Soares, J., Stevenson, D. S., Tombrou-Tzella, M., Varotsos, K. V., and Young, P.: Is the ozone climate penalty robust in Europe?, *Environmental Research Letters*, 10, 084015, <https://doi.org/10.1088/1748-9326/10/8/084015>, 2015.

460 Da, Y., Xu, Y., and McCarl, B.: Effects of Surface Ozone and Climate on Historical (1980–2015) Crop Yields in the United States: Implication for Mid-21st Century Projection, *Environmental and Resource Economics*, 81, 355–378, <https://doi.org/10.1007/s10640-021-00629-y>, 2022.

Dee, D., Uppala, S., Simmons, A., Berrisford, P., Poli, P., Kobayashi, S., Andrae, U., Balmaseda, M., Balsamo, G., Bauer, P., Bechtold, P., Beljaars, A., van de Berg, L., Bidlot, J., Bormann, N., Delsol, C., Dragani, R., Fuentes, M., Geer, A., Haimberger, L., Healy, S., Hersbach, H., Hólm, E., Isaksen, L., Källberg, P., Köhler, M., Matricardi, M., McNally, A., Monge-Sanz, B., Morcrette, J.-J., Park, B.-K., Peubey, C., de Rosnay, P., Tavolato, C., Thépaut, J.-N., and Vitart, F.: The ERA-Interim reanalysis: configuration and performance of the data



assimilation system, *Quarterly Journal of the Royal Meteorological Society*, 137, 553–597, [https://doi.org/https://doi.org/10.1002/qj.828](https://doi.org/10.1002/qj.828), 470 2011.

Doherty, R., Heal, M., and O'Connor, F.: Climate change impacts on human health over Europe through its effect on air quality, *Environmental Health*, 16, 118, <https://doi.org/10.1186/s12940-017-0325-2>, 2017.

Emmons, L., Schwantes, R., Orlando, J., Tyndall, G., Kinnison, D., Lamarque, J.-F., Marsh, D., Mills, M., Tilmes, S., Bardeen, C., Buchholz, R., Conley, A., Gettelman, A., Garcia, R., Simpson, I., Blake, D., Meinardi, S., and Pétron, G.: The Chemistry Mechanism in 475 the Community Earth System Model Version 2 (CESM2), *Journal of Advances in Modeling Earth Systems*, 12, e2019MS001882, <https://doi.org/https://doi.org/10.1029/2019MS001882>, 2020.

ENVIRON: CAMx User's Guide, Comprehensive Air Quality model with Extensions, version 6.50, [code], Novato, California, <https://www.camx.com/download/>, last access: 11 Jun 2025, 2018.

EUR-Lex, 2015: Consolidated text: Directive 2008/50/EC of the European Parliament and of the Council of 21 May 2008 on ambient 480 air quality and cleaner air for Europe, <https://eur-lex.europa.eu/legal-content/EN/ALL/?uri=CELEX:02008L0050-20150918>, online, last access 29 Oct 2025, 2015.

Gentile, M., Courbin, F., and Meylan, G.: Interpolating point spread function anisotropy, *Astronomy and Astrophysics*, 549, <https://doi.org/10.1051/0004-6361/201219739>, 2012.

Grell, G. A., Peckham, S. E., Schmitz, R., McKeen, S. A., Frost, G., Skamarock, W. C., and Eder, B.: Fully coupled “online” chemistry 485 within the WRF model, *Atmospheric Environment*, 39, 6957–6975, <https://doi.org/https://doi.org/10.1016/j.atmosenv.2005.04.027>, 2005.

Gu, Y., Henze, D. K., Nawaz, M. O., and Wagner, U. J.: Response of the ozone-related health burden in Europe to changes in local anthropogenic emissions of ozone precursors, *Environmental Research Letters*, 18, 114 034, <https://doi.org/10.1088/1748-9326/ad0167>, 2023.

Huszar, P., Miksovsky, J., Pisoft, P., Belda, M., and Halenka, T.: Interactive coupling of a regional climate model and a chemical transport 490 model: evaluation and preliminary results on ozone and aerosol feedback, *Climate Research*, 51, 59–88, <https://doi.org/10.3354/cr01054>, 2012.

Im, U., Bianconi, R., Solazzo, E., Kioutsioukis, I., Badia, A., Balzarini, A., Baró, R., Bellasio, R., Brunner, D., Chemel, C., Curci, G., Flemming, J., Forkel, R., Giordano, L., Jiménez-Guerrero, P., Hirtl, M., Hodzic, A., Honzak, L., Jorba, O., Knote, C., Kuenen, J. J., Makar, P. A., Manders-Groot, A., Neal, L., Pérez, J. L., Pirovano, G., Pouliot, G., San Jose, R., Savage, N., Schroder, W., Sokhi, R. S., Syrakov, D., Torian, A., Tuccella, P., Werhahn, J., Wolke, R., Yahya, K., Zabkar, R., Zhang, Y., Zhang, J., Hogrefe, C., and Galmarini, S.: 495 Evaluation of operational on-line-coupled regional air quality models over Europe and North America in the context of AQMEII phase 2. Part I: Ozone, *Atmospheric Environment*, 115, 404–420, <https://doi.org/https://doi.org/10.1016/j.atmosenv.2014.09.042>, 2015.

Im, U., Ye, Z., Schuh, N., Chowdhury, S., Christensen, J. H., Geels, C., Hänninen, Risto and Hodnebrog, Ø., Marelle, L., Sofiev, M., Brandt, J., and Aunan, K.: Europe will struggle to meet the new WHO Air Quality Guidelines, *npj Clean Air*, 1, 13, <https://doi.org/10.1038/s44407-025-00013-w>, 2025.

Jeffreys, H.: An invariant form for the prior probability in estimation problems, *Proceedings of the Royal Society of London. Series A. Mathematical and Physical Sciences*, 186, 453–461, <https://doi.org/10.1098/rspa.1946.0056>, 1946.

Karlický, J., Rieder, H. E., Huszár, P., Peiker, J., and Sukhodolov, T.: A cautious note advocating the use of ensembles of models and driving data in modeling of regional ozone burdens, *Air Quality, Atmosphere & Health*, 17, 1415–1424, <https://doi.org/10.1007/s11869-024-01516-3>, 2024.

Karlický, J., Huszár, P., and Halenka, T.: Validation of gas phase chemistry in the WRF-Chem model over Europe, *Advances in Science and 505 Research*, 14, 181–186, <https://doi.org/10.5194/asr-14-181-2017>, 2017.



Karlický, 2022: WRF-Chem and CAMx ozone surface data from central-European simulations of period 2007-2016, <https://doi.org/10.48700/datst.ev7ej-gv255>, online, last access 12 Dec 2025, 2022.

Liu, Z., Doherty, R. M., Wild, O., O'Connor, F. M., and Turnock, S. T.: Correcting ozone biases in a global chemistry–climate model: implications for future ozone, *Atmospheric Chemistry and Physics*, 22, 12 543–12 557, <https://doi.org/10.5194/acp-22-12543-2022>, 2022.

510 Mar, K. A., Ojha, N., Pozzer, A., and Butler, T. M.: Ozone air quality simulations with WRF-Chem (v3.5.1) over Europe: model evaluation and chemical mechanism comparison, *Geoscientific Model Development*, 9, 3699–3728, <https://doi.org/10.5194/gmd-9-3699-2016>, 2016.

Neal, L., Agnew, P., Moseley, S., Ordóñez, C., Savage, N., and Tilbee, M.: Application of a statistical post-processing technique to a gridded, operational, air quality forecast, *Atmospheric Environment*, 98, 385–393, <https://doi.org/https://doi.org/10.1016/j.atmosenv.2014.09.004>, 515 2014.

Ohlert, P. L., Bach, M., and Breuer, L.: Accuracy assessment of inverse distance weighting interpolation of groundwater nitrate concentrations in Bavaria (Germany), *Environmental Science and Pollution Research*, 30, 9445–9455, <https://doi.org/10.1007/s11356-022-22670-0>, 2023.

Peiker, 2025: Bias corrected ozone MDA8 in central Europe using various bias correction strategies, <https://doi.org/10.5281/zenodo.17881911>, online, last access 12 Dec 2025, 2025.

520 Pozzer, A., Anenberg, S. C., Dey, S., Haines, A., Lelieveld, J., and Chowdhury, S.: Mortality Attributable to Ambient Air Pollution: A Review of Global Estimates, *GeoHealth*, 7, e2022GH000711, <https://doi.org/https://doi.org/10.1029/2022GH000711>, e2022GH000711 2022GH000711, 2023.

Prieto Perez, A. P., Huszár, P., and Karlický, J.: Validation of multi-model decadal simulations of present-day central European air-quality, 525 *Atmospheric Environment*, 349, 121 077, <https://doi.org/https://doi.org/10.1016/j.atmosenv.2025.121077>, 2025.

Racherla, P. N. and Adams, P. J.: The response of surface ozone to climate change over the Eastern United States, *Atmospheric Chemistry and Physics*, 8, 871–885, <https://doi.org/10.5194/acp-8-871-2008>, 2008.

Rahmani, M., Mohammadi, A., and Azadi, M.: Comparison of Spatial Interpolation Methods for Gridded Bias Removal in Surface Temperature Forecasts, *Journal of Meteorological Research*, 31, <https://doi.org/10.1007/s13351-017-6135-1>, 2017.

530 Revell, L. E., Stenke, A., Tummon, F., Feinberg, A., Rozanov, E., Peter, T., Abraham, N. L., Akiyoshi, H., Archibald, A. T., Butchart, N., Deushi, M., Jöckel, P., Kinnison, D., Michou, M., Morgenstern, O., O'Connor, F. M., Oman, L. D., Pitari, G., Plummer, D. A., Schofield, R., Stone, K., Tilmes, S., Visioni, D., Yamashita, Y., and Zeng, G.: Tropospheric ozone in CCM1 models and Gaussian process emulation to understand biases in the SOCOLv3 chemistry–climate model, *Atmospheric Chemistry and Physics*, 18, 16 155–16 172, <https://doi.org/10.5194/acp-18-16155-2018>, 2018.

535 Rieder, H. E., Fiore, A. M., Horowitz, L. W., and Naik, V.: Projecting policy-relevant metrics for high summertime ozone pollution events over the eastern United States due to climate and emission changes during the 21st century, *Journal of Geophysical Research: Atmospheres*, 120, 784–800, <https://doi.org/10.1002/2014JD022303>, 2015.

Schnell, J. L., Holmes, C. D., Jangam, A., and Prather, M. J.: Skill in forecasting extreme ozone pollution episodes with a global atmospheric chemistry model, *Atmospheric Chemistry and Physics*, 14, 7721–7739, <https://doi.org/10.5194/acp-14-7721-2014>, 2014.

540 Seinfeld, J. H. and Pandis, S. N.: *Atmospheric Chemistry and Physics: From Air Pollution to Climate Change*, Wiley, ISBN 9781118591505, 1998.

Staehle, C., Rieder, H. E., Fiore, A. M., and Schnell, J. L.: Technical note: An assessment of the performance of statistical bias correction techniques for global chemistry–climate model surface ozone fields, *Atmospheric Chemistry and Physics*, 24, 5953–5969, <https://doi.org/10.5194/acp-24-5953-2024>, 2024.



545 Sukhodolov, T., Egorova, T., Stenke, A., Ball, W. T., Brodowsky, C., Chiodo, G., Feinberg, A., Friedel, M., Karagodin-Doyennel, A., Pe-  
ter, T., Sedlacek, J., Vattioni, S., and Rozanov, E.: Atmosphere–ocean–aerosol–chemistry–climate model SOCOLv4.0: description and  
evaluation, *Geoscientific Model Development*, 14, 5525–5560, <https://doi.org/10.5194/gmd-14-5525-2021>, 2021.

550 Uglolotti, A., Anders, T., Lanssens, B., Hickler, T., François, L., and Tölle, M. H.: Impact of bias correction on climate  
change signals over central Europe and the Iberian Peninsula, *Frontiers in Environmental Science*, Volume 11 - 2023,  
<https://doi.org/10.3389/fenvs.2023.1116429>, 2023.