



1 **Interpretable Machine Learning Quantifies Composition and Size**
2 **Controls on Aerosol Spectral Absorption**

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4 **Wenfang Wang¹, Pengfei Tian^{1,2*}, Shuhua Zeng¹, Yifei Zhang¹, Zeren Yu¹, Chen**
5 **Cui¹, Yunfei Wu³, Min Chen^{1*}, Lei Zhang^{1,2}**

6 ¹ Key Laboratory for Semi-Arid Climate Change of the Ministry of Education, College
7 of Atmospheric Sciences, Lanzhou University, Lanzhou 730000, China.

8 ² Collaborative Innovation Center for Western Ecological Safety, Lanzhou University,
9 Lanzhou 730000, China.

10 ³ State Key Laboratory of Atmospheric Environment and Extreme Meteorology,
11 Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing 100029,
12 China.

13 Corresponding author: P. Tian (tianpf@lzu.edu.cn) and M. Chen (chenmin@lzu.edu.cn)

14



15 **Abstract**

16 The spectral dependence of aerosol absorption, characterized by the absorption
17 Ångström exponent (AAE), strongly influences radiative effects, yet the relative
18 importance of controlling factors remains poorly quantified. We integrate multisource
19 observations with an interpretable machine-learning framework (Shapley Additive
20 Explanations, SHAP) to disentangle the roles of chemical composition and particle size
21 in shaping AAE and to evaluate radiative impacts. Field observation in Beijing reveal
22 that near-surface AAE is predominantly influenced by higher fine mineral dust and
23 water-soluble inorganic ions fractions. Multi-year columnar data identify dust loading
24 as the dominant factor, followed by carbonaceous aerosols. The fine-mode radius
25 accounts for 29% of size parameters cumulative importance and ranks closely with
26 black carbon. SHAP diagnostics highlight that columnar AAE contributes to radiative
27 forcing at the top of the atmosphere (TOA) comparably to single scattering albedo
28 (SSA), while its impact is clearly weaker at the bottom of the atmosphere and in the
29 atmosphere. These findings help clarify AAE determinants and reduce uncertainties in
30 aerosol radiative effect assessments.

31



1 Introduction

Light-absorbing aerosols (LAAs), primarily black carbon (BC), brown carbon (BrC), and mineral dust, significantly influence regional and global climate by absorbing solar radiation (Bahadur et al., 2012; Cappa et al., 2016; Kok et al., 2017; Nishant et al., 2019). For instance, BC contributed a net positive effective radiative forcing of $0.11 \text{ W} \cdot \text{m}^{-2}$ during 1750–2019, with a wide uncertainty range from -0.20 to $+0.42 \text{ W} \cdot \text{m}^{-2}$ (Intergovernmental Panel On Climate Change (Ipcc), 2023), reflecting limited observational constraints on aerosol optical/microphysical properties and their inaccurate representation in models (Gliß et al., 2021; Lee et al., 2016). A practical diagnostic for the spectral shape of absorption is the Absorption Ångström Exponent (AAE) (Ångström, 1929; Lewis et al., 2008). For pure BC, AAE is theoretically close to 1.0, but observations show a range of 0.6–1.6 (Kirchstetter et al., 2004; Lack and Cappa, 2010; Gyawali et al., 2012; Chakrabarty et al., 2013; Wang et al., 2021). BrC and mineral dust exhibit relatively stronger absorption in the ultraviolet and visible spectral wavelengths, typically yielding AAE values greater than 2.0 (Russell et al., 2010; Park et al., 2018; Zhang et al., 2020; Cuesta-Mosquera et al., 2024). Because AAE encodes source and process information that governs aerosol absorption from ultraviolet to the near-infrared wavelength range, tighter constraints on AAE can help reduce uncertainties in aerosol radiative effects (Cazorla et al., 2013; Lack and Langridge, 2013; Sand et al., 2021).



52 AAE varies with particle size distribution, chemical composition, and mixing state
53 (Scarnato et al., 2013; Li et al., 2016; Schuster et al., 2016a; Sotiropoulou et al., 2025).
54 For instance, AAE may decrease as BC cores grow or as aggregates become more
55 compact during aging processes (Liu et al., 2018). Recent numerical simulation further
56 indicates that secondary organic coatings can increase AAE, with sensitivity to coating
57 thickness (Zhang et al., 2025). In contrast, photochemical bleaching lowers BrC
58 ultraviolet absorption and AAE (Wang et al., 2019; Li et al., 2025). Heterogeneous
59 aging of long-range-transported dust may enhance absorption, also affecting AAE (Tian
60 et al., 2018). The magnitudes and signs of these effects depend on location, season, and
61 processing history, complicating both measurements and modeling and propagating to
62 radiative forcing uncertainty (Sand et al., 2021; Li et al., 2022; Ponczek et al., 2022).

63 Studying the impact of individual factors on AAE is relatively straightforward.
64 Previous studies have already examined the effects of particle size, chemical
65 composition, and mixing state on AAE in isolation (Wu et al., 2015; Schuster et al.,
66 2016b; Li et al., 2024). However, quantitatively attributing the relative contributions of
67 particle size and chemical composition to AAE remains a challenging task due to
68 nonlinearity and collinearity among predictors. For example, observations show that
69 composition appears dominant when the shape of size distribution is quasi-stationary
70 (Utry et al., 2014), whereas Mie-theory studies highlight the role of the imaginary
71 refractive index of organics over size in explaining absorption changes (Yang et al.,
72 2025). Although these studies effectively highlight the roles of particle size and



73 chemical composition, they lack quantitative assessments of their relative importance.

74 The Shapley Additive exPlanations (SHAP) method offers a principled framework
75 for feature attribution in machine learning predictions and has been widely adopted in
76 atmospheric sciences, such as boundary-layer height inversion, ozone formation and
77 cloud-condensation-nuclei studies (Peng et al., 2023; Tao et al., 2024; Wang et al.,
78 2025a). SHAP analysis has also been applied to aerosol absorption studies to precisely
79 quantify the relative contribution of various chromophores to BrC absorption,
80 providing a mechanistic understanding of its key drivers (Wang et al., 2024). Its
81 potential to predict AAE and quantify the relative contributions of individual factors
82 remains unexplored. Addressing this gap would offer an interpretable, data-driven
83 perspective on aerosol spectral absorption.

84 This study aims to quantify the relative importance of chemical composition and
85 particle size influencing AAE and to elucidate the critical role of AAE in radiative
86 effects. We applied machine learning methods for columnar AAE (AAE_{col}), and
87 assessed how AAE_{col} influences aerosol direct radiative forcing (ADRF) and aerosol
88 radiative forcing efficiency (ARFE). During the modeling process, an ensemble of
89 models was initially trained, after which the optimal model was selected to predict
90 AAE_{col} , ADRF, and ARFE, with the SHAP algorithm used for interpretative analysis.

91 **2 Methods**

92 **2.1 Field Campaign and Data Processing**



93 An intensive observation campaign focusing on aerosol properties was conducted
94 in urban Beijing, China, from 16 December 2023 to 15 January 2024. Online and offline
95 instruments were deployed on the rooftop of the Institute of Atmospheric Physics,
96 Chinese Academy of Sciences (Building #3; 39.98°N, 116.39°E), approximately 45 m
97 above ground level. All online instruments were housed in a temperature-controlled
98 room maintained at ~20 °C to ensure measurement stability, and sampling lines were
99 equipped with Nafion dryers to minimize the influence of ambient humidity.

100 **2.1.1 In Situ Online Aerosol Observations**

101 Aerosol absorption coefficients ($b_{\text{abs},\lambda}$) at 375, 532, and 870 nm were measured
102 using photoacoustic extinctions (PAX, DMT Inc., USA). The PAX measures
103 aerosol light absorption using the photoacoustic technique, in which absorbed laser
104 energy is converted into periodic heating of the surrounding gas, generating an acoustic
105 pressure wave in an acoustic resonator that is detected by a sensitive microphone (Truex
106 and Anderson, 1979). The light absorption ($b_{\text{abs},\text{pax}}$) can be calculated as:

$$107 \quad b_{\text{abs},\text{pax}} = \frac{P_{\text{mic}} \times A_{\text{res}} \times \pi^2 \times f_{\text{res}}}{P_L \times (\gamma - 1) \times Q} \quad (1)$$

108 where P_{mic} and P_L are the pressure of the microphone and the laser power,
109 respectively; A_{res} , f_{res} and Q indicate the cross-sectional area, resonance frequency,
110 and quality factor of the resonator; γ is the isobaric and isosteric specific heat ratio.
111 Then the $b_{\text{abs},\lambda}$ is obtained by subtracting the background absorption measured with
112 particle-free air from $b_{\text{abs},\text{pax}}$. In addition, the PAX measures aerosol scattering with
113 an integrated wide-angle reciprocal nephelometer.



114 Prior to deployment, each PAX was calibrated following the procedure described
115 in Wu et al., 2015: (i) the scattering channel was calibrated using high-concentration
116 ammonium sulfate aerosol by regressing the extinction coefficient (b_{ext}) derived from
117 laser power attenuation against the instrument-recorded scattering coefficient (b_{sca}).
118 The scattering calibration factor was then adjusted by applying the regression slope as
119 a multiplicative correction; (ii) the absorption channel was calibrated using high-
120 concentration Aquadag aerosol by regressing ($b_{\text{ext}} - b_{\text{sca}}$) against the instrument-
121 recorded photoacoustic absorption. The absorption calibration factor was then updated
122 by dividing it by the regression slope.

123 The near-surface aerosol absorption Ångström exponent (AAE_{sfc}) was calculated
124 as:

$$125 \quad AAE_{\text{sfc}} = -\frac{\log(b_{\text{abs},\lambda_1}) - \log(b_{\text{abs},\lambda_2})}{\log(\lambda_1) - \log(\lambda_2)} \quad (2)$$

126 where $\lambda_1 = 375$ nm, $\lambda_2 = 870$ nm. Hourly $PM_{2.5}$ (particle matters with an aerodynamic
127 diameter ≤ 2.5 μm) mass concentrations were obtained from the China National
128 Environmental Monitoring Network for the Beijing urban site. These datasets were
129 used to calculate mass absorption efficiency (MAE) of $PM_{2.5}$:

$$130 \quad MAE_{\lambda} = \frac{b_{\text{abs},\lambda}}{PM_{2.5}} \quad (3)$$

131 Size distributions were measured with a scanning mobility particle sizer (SMPS,
132 Model 3082, TSI Inc., 8.8–310.6 nm, Stokes diameter) and an aerodynamic particle
133 sizer (APS, Model 3321, TSI Inc., 0.54–19.8 μm , aerodynamic diameter), with SMPS
134 data converted to aerodynamic diameter (Text S1) (Shang et al., 2018). To ensure



135 measurement accuracy, the flow systems of the SMPS and APS were periodically
136 checked and calibrated by the manufacturer (TSI Inc.) through regular return-service
137 calibration. In addition, hourly meteorological parameters (wind speed and direction,
138 temperature, and relative humidity) were obtained from the 47 m meteorological tower
139 at the Institute of Atmospheric Physics.

140 **2.1.2 Offline Aerosol Sampling**

141 Offline $PM_{2.5}$ samples were collected on quartz-fiber filters (90 mm diameter;
142 Whatman 1855-090). Prior to sampling, quartz-fiber filters were pre-cleaned to
143 minimize filter background. Briefly, filter cassettes were rinsed with absolute ethanol
144 and air-dried, and aluminum-foil liners cut to the filter size were pre-baked at 550 °C
145 for 3 h to remove residual carbon. The quartz filters were sequentially soaked in
146 ultrapure water (5 min \times 3 cycles, followed by 2 h \times 2 cycles), oven-dried at 150 °C for
147 1 h, and then prebaked at 550 °C for 5 h to reduce the influence of adsorbed organic
148 and inorganic materials. After cooling, the filters were wrapped in prebaked aluminum
149 foil and conditioned for 48 h in a constant temperature and humidity environment prior
150 to gravimetric determination. Then the pre-sampling filter mass was measured using an
151 electronic microbalance (BSA124S-CW, Sartorius; readability ± 0.1 mg). Sampling was
152 conducted using a medium-volume air sampler (Model 2030, Laoshan Electronic
153 Instrument Co., Ltd.) operated at 100 L \cdot min⁻¹ and equipped with a $PM_{2.5}$ inlet. The inlet
154 was installed at approximately 2 m above ground level. Daytime samples were collected
155 from 09:00 to 20:30, and nighttime samples from 21:00 to 08:30 the following day.



156 After sampling, all filters were analyzed for major chemical compositions,
157 including water-soluble inorganic ions (Na^+ , K^+ , NH_4^+ , Ca^{2+} , Mg^{2+} , Cl^- , NO_3^- , and SO_4^{2-})
158 measured by ion chromatography (881 Compact IC Pro, Metrohm; and ICS-1500,
159 Dionex Inc.), metallic elements (Al, Ca, Mg, Fe, and Ti) determined by inductively
160 coupled plasma–atomic emission spectrometry (ICP-AES; iCAP 7400, Thermo), and
161 organic carbon (OC), and elemental carbon (EC) quantified using the thermal/optical
162 carbon analyzer (DRI Model 2015, USA) based on the thermal/optical reflectance
163 (TOR) method (Chow et al., 2007). The chemical analysis process is described in detail
164 in Supplementary Text S2.

165 $\text{PM}_{2.5}$ was reconstructed as the sum of organic matter ($\text{OM}=1.6\times\text{OC}$) (Guinot et
166 al., 2007), EC, non-dust water-soluble ions (nd-WSII, and fine mineral dust (FMD)
167 derived from crustal elements (Malm et al., 1994; Tian et al., 2023), showing good
168 agreement with measured $\text{PM}_{2.5}$ ($r = 0.82$). Here, nd-WSII was defined as the sum of
169 K^+ , NH_4^+ , NO_3^- , and SO_4^{2-} , while Na^+ , Ca^{2+} , Mg^{2+} , and Cl^- were excluded. Ca^{2+} and
170 Mg^{2+} were treated as dust-related species, Na^+ was excluded due to generally elevated
171 blanks associated with quartz-fiber filters and glassware, and Cl^- was excluded given
172 its strong association with Mg^{2+} ($r = 0.77$, Table S1). The FMD was defined as follows:

173
$$[\text{FMD}] = 2.20[\text{Al}] + 2.49[\text{Si}] + 1.63[\text{Ca}] + 2.42[\text{Fe}] + 1.94[\text{Ti}] \quad (4)$$

174 where $[\text{Si}] = 1.5 [\text{Al}]$.

175 2.2 Multiple Linear Regression

176 The influence of particle size and chemical composition on AAE_{sfc} was assessed



177 using a standardized multiple linear regression:

178
$$\widehat{AAE_{sfc}} = a + b \times \widehat{FMD} + c \times \widehat{nd-WSII} + d \times \widehat{D_{SMPS}} + e \times \widehat{D_{APS}} \quad (5)$$

179 where $\widehat{AAE_{sfc}}$ denotes the standardized AAE_{sfc} ; a represents the intercept term,
180 encompassing EC, OM, and other potential influencing factors not explicitly accounted
181 for; b , c , d , and e are regression coefficients; \widehat{FMD} , $\widehat{nd-WSII}$, $\widehat{D_{SMPS}}$, and $\widehat{D_{APS}}$
182 are standardized variables of FMD fraction, nd-WSII fraction, and mean diameters from
183 SMPS and APS, respectively. Offline chemical composition data were temporally
184 matched to the corresponding online measurements based on sampling periods. Due to
185 power outage on 27 December 2023 and 3 January 2024, daytime data for 27 December
186 and both daytime and nighttime data for 3 January were unavailable.

187 **2.3 AERONET Data**

188 We used data from the Beijing Aerosol Robotic Network (AERONET) site (39.98°
189 N, 116.38° E). Level 2.0 quality-assured Version 3 inversion products were selected.
190 Key parameters include aerosol absorption optical depth (AAOD) at 440, 675, 870, and
191 1020 nm to calculate columnar AAE (AAE_{col}):

192
$$AAOD_{\lambda} = k \times \lambda^{-AAE_{col}} \quad (6)$$

193 where k is a constant. In addition, aerosol size-related parameters were obtained,
194 including volume size distributions for radii in the 0.05–15.00 μm range, mean radii of
195 fine-mode and coarse-mode particles (R_{fine} , R_{coarse}), as well as volume concentrations
196 of fine-mode (0.05–0.60 μm , vol_{fine}) (Sinyuk et al., 2020) and total particles (vol_{total}).
197 The fine-mode fraction (FMF) was then calculated as:



198
$$FMF = \frac{vol_{fine}}{vol_{total}} \quad (7)$$

199 To investigate the influence of chemical composition on AAE_{col} , we used the
200 chemical composition dataset derived from AERONET inversions (Zhang et al., 2024),
201 including black carbon (BC), brown carbon (BrC), coarse-mode absorbing soluble
202 matter (CAI, representing coarse absorbing dust), coarse-mode non-absorbing soluble
203 matter (CNAI, representing coarse non-absorbing dust and aged carbonaceous aerosols),
204 and fine-mode non-absorbing soluble matter (FNAI, representing fine non-absorbing
205 dust and organic carbon) (Zhang et al., 2024).

206 To further assess the radiative impacts of aerosols, we also employed AERONET
207 shortwave instantaneous aerosol direct radiative forcing (ADRF) and aerosol radiative
208 forcing efficiency (ARFE) data, where ARFE is defined as ADRF per unit aerosol
209 optical depth (AOD) at 550 nm, reported at the top of the atmosphere (TOA), bottom
210 of the atmosphere (BOA), and in the atmosphere (ATM) (Holben et al., 1998).

211 **2.4 Shapley Additive Explanations (SHAP)**

212 To quantify how particle size and chemical composition control the AAE_{col} , we
213 trained and compared three ensemble tree-based regressors: Extreme Gradient
214 Boosting (XGBoost), Random Forest (RF), and Categorical Boosting (CatBoost). Each
215 model was trained using seven predictor variables, including five chemical
216 compositions (BrC, BC, CAI, CNAI, and FNAI), and four size parameters (R_{fine} , R_{coarse} ,
217 vol_{fine} , and volume concentrations of coarse-mode (vol_{coarse})). Model performance was
218 evaluated using a consistent training–testing split and quantified by the coefficient of



determination (R^2), root mean square error (RMSE), and mean absolute error (MAE). The CatBoost model in our case was subsequently adopted for further interpretation (Fig. S7). SHapley Additive exPlanations (SHAP) analysis was applied to decompose the model output into additive feature contributions, enabling quantitative assessment of the relative contribution and sensitivity of individual aerosol composition and size parameters in determining. Predicted-versus-observed diagnostics are shown in Supplementary Fig. S7.

Similarly, to evaluate aerosol radiative impacts, XGBoost, RF, and CatBoost models also were trained using distinct predictor sets for different radiative metrics. For aerosol direct radiative forcing (ADRF), five optical properties (AOD, single scattering albedo (SSA), asymmetry parameter (g), surface albedo (SA), and AAE_{col}) were used as inputs. For radiative forcing efficiency (ARFE), the target definition ($ARFE = ADRF/AOD$) was kept unchanged; however, AOD was included during model fitting together with SSA, g , SA, and AAE_{col} so that the models could learn any residual nonlinearity and interactions involving AOD. Performance was again quantified by R^2 , RMSE, and MAE on a consistent split. CatBoost in our case was retained as the best-performing model across TOA, BOA, and ATM. Predicted and observed comparisons are provided in Supplementary Fig. S8-S9.

To attribute ARFE variations while controlling for AOD, we employed SHAP with a scenario-based conditioning approach. Specifically, we recomputed SHAP values on the held-out test set after fixing AOD to four levels (25th, 50th, 75th percentiles, and



mean), computed from the training set to avoid information leakage, while leaving all other predictors unchanged. This yields SHAP attributions for SSA, g, SA, and AAE_{col} conditional on AOD at representative states. (The 50th percentile case is shown in the main text, others in Supplementary). For ADRF, SHAP was computed in the standard manner using all five predictors (AOD, SSA, g, SA, AAE) without conditioning.

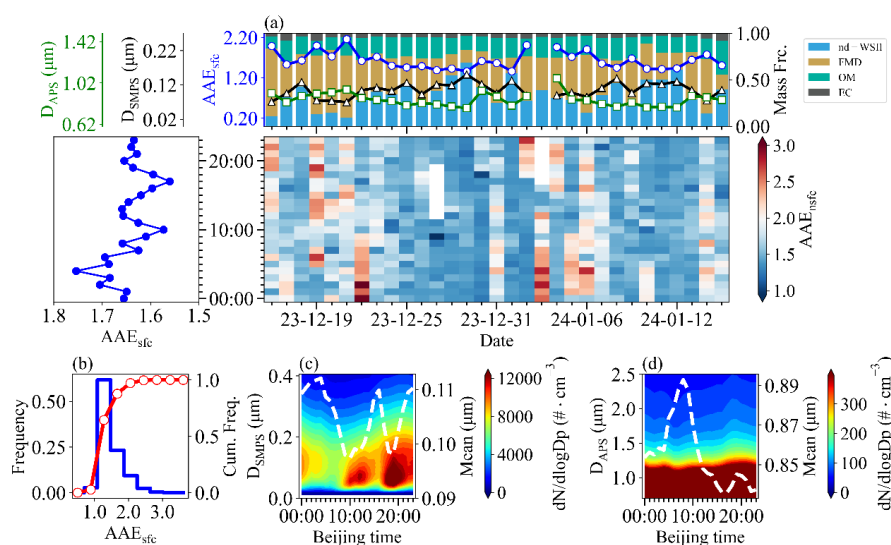
3 Results and Discussion

3.1 Aerosol Characteristics During the Field Campaign

During the campaign, aerosol absorption decreased systematically with wavelength (375, 532, and 870 nm) (Fig. S2d), demonstrating typical wavelength-dependent characteristics. The corresponding mass absorption efficiencies were relatively low (0.49 ± 0.24 , 0.21 ± 0.08 , and $0.12 \pm 0.04 \text{ m}^2 \cdot \text{g}^{-1}$), reflecting the dominance of nd-WSII, which accounted for 42.9% of $PM_{2.5}$ mass (Fig. 1a). Fig. 1a also showed that variations in near-surface AAE (AAE_{sfc}) were closely linked to changes in aerosol chemical composition mass fraction. Periods with elevated FMD fractions generally coincided with higher AAE_{sfc} , whereas intervals dominated by nd-WSII corresponded to lower values. The frequency distribution further indicates that AAE_{sfc} ranged from 0.90 to 3.0, and was most frequently 1.10–1.50 (mean 1.28 ± 0.39 ; Fig. 1b). Such elevated values likely resulted from winter heating emissions (Tian et al., 2019; Yan et al., 2017) and mineral dust contributions (Fig. 1a), both known to raise AAE (Liu et al., 2018).



Both absorption coefficients and AAE_{sfc} exhibited pronounced diurnal variability. Absorption coefficients were consistently higher at night and a peak around 23:00 (Fig. S2), driven by reduced tropospheric boundary layer height, lower afternoon temperatures and wind speeds (Fig. S3), and enhanced emissions from nighttime traffic and heating (Guo et al., 2016; Zhao et al., 2019). AAE_{sfc} showed a clear night-high and day-low pattern, consistent with the evolution of particle size distributions. Fine-mode number concentrations derived from scanning mobility particle sizer (SMPS) increased during the morning rush hours and nighttime residential activity (Fig. 1c). By contrast, coarse-mode diameters from aerodynamic particle sizer spectrometer (APS) were larger in the early morning and decreased during the day (Fig. 1d). These results demonstrate that AAE_{sfc} was co-regulated by both composition and size, providing the observational evidence for the subsequent machine-learning analysis to quantify their relative contributions and radiative implications.





274 **Figure 1. Aerosol absorption characteristics during the campaign.** (a) Time series
275 of mean particle diameters derived from SMPS and APS, mass fractions of organic
276 matter (OM), elemental carbon (EC), non-dust water-soluble inorganic ions (nd-WSII),
277 and fine mineral dust (FMD), together with daily averaged near-surface AAE (AAE_{sfc}).
278 (b) Frequency distribution of AAE_{sfc} . (c–d) Diurnal variations of aerosol particle
279 number size distributions from SMPS (c) and APS (d).

280 **3.2 Influence of Composition and Size on Near-surface AAE**

281 The EC mass fraction showed no correlation with AAE_{sfc} ($r = 0.09$, $p = 0.49$; Fig.
282 2a), which can be attributed to the weak wavelength dependence of EC absorption
283 (Samset et al., 2018) and its relatively small mass contribution ($\sim 4\%$) (Fig. 2a).
284 Similarly, the OM mass fraction is not significantly correlated with AAE_{sfc} ($r = -0.11$,
285 $p = 0.40$; Fig. 2a). In contrast to study dominated by biomass burning, where light-
286 absorbing organic carbon can account for $> 50\%$ of the mass fraction and strongly
287 enhance AAE (Wang et al., 2021). During the Beijing campaign, however, OM
288 contributed only $\sim 19\%$ of total $PM_{2.5}$ mass and BrC fractions were relatively low.
289 Although BrC exhibits intrinsically high AAE values (Laskin et al., 2015; Moosmüller
290 et al., 2011), its impact was diminished in the mixed aerosol matrix due to the influence
291 of other dominant compositions.

292 We observed a statistically significant negative correlation between AAE_{sfc} and
293 carbonaceous aerosol AAE (AAE_{CA}) (Fig. 2b), indicating that the non-carbonaceous
294 aerosol had a significantly stronger role in shaping the absorption spectral dependence



295 under complex pollution conditions. Due to nitrogen dioxide (NO_2) concentrations were
296 elevated at night (Fig. S4), which can interfere with PAX instruments, particularly at
297 shorter wavelengths (Arnott et al., 2000; Gyawali et al., 2012). Therefore, we restrict
298 the analysis here to daytime data (Fig S5). This pattern therefore cannot be ascribed
299 simply to inter-instrument discrepancies.

300 AAE_{sfc} exhibited a significant positive correlation with the mass fraction of FMD
301 ($r = 0.79$, $p < 0.01$) and a negative correlation with nd-WSII ($r = -0.78$, $p < 0.01$) (Fig.
302 2a). The AAE_{sfc} enhancement associated with FMD can be attributed to metal oxides
303 such as hematite and goethite, which strongly absorb in the UV wavelengths and
304 steepen the spectral dependence (Bi et al., 2016). By contrast, nd-WSII acted as non-
305 absorbing diluents, suppressing the overall wavelength dependence. Although the
306 "lensing effect" of nd-WSII can enhance the AAE_{CA} (Cappa et al., 2012; Zhang et al.,
307 2025), due to the very low contribution of carbonaceous components to total $\text{PM}_{2.5}$ mass,
308 preventing any observable positive correlation between AAE_{sfc} and the nd-WSII mass
309 fraction.

310 Particle size also played a critical role. AAE_{sfc} was negatively associated with the
311 fine-mode mean diameter from SMPS (D_{SMPS} , $r = -0.58$; Fig. 2c) and positively
312 associated with the coarse-mode mean diameter from APS (D_{APS} , $r = 0.58$; Fig. 2d).
313 Standardized multiple regression corroborated these effects, yielding coefficients of –
314 0.02 for smaller particles and 0.44 for larger particles. Regression coefficients for
315 composition were likewise consistent with the bivariate analysis: FMD had a positive



effect on AAE_{sfc} (coefficient is 0.35), whereas nd-WSII had a negative effect (coefficient is -0.17). Particle size covaried with composition: D_{APS} was positively correlated with the FMD mass fraction ($r = 0.64$) and negatively with nd-WSII ($r = -0.64$), while D_{SMPS} was strongly positively correlated with nd-WSII ($r = 0.89$) and negatively with FMD ($r = -0.87$) (Fig. S6). Together, these relationships indicate that larger particles coincide with higher FMD and lower nd-WSII, enhancing AAE_{sfc} , whereas smaller particles favor inorganic salts and reduced dust contributions, lowering AAE_{sfc} . Overall, AAE_{sfc} was not governed by carbonaceous aerosols alone but rather by the coupled influence of mineral dust, inorganic matter, and particle size distributions.

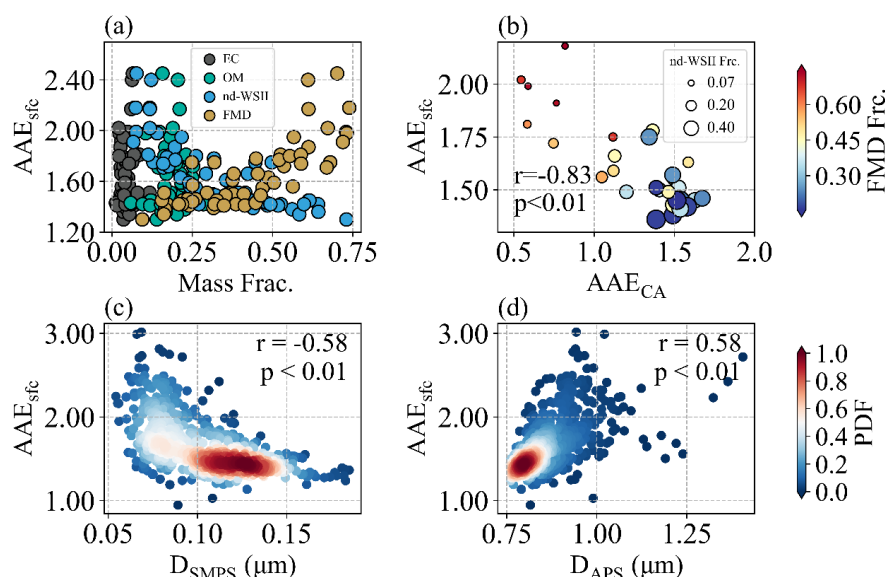


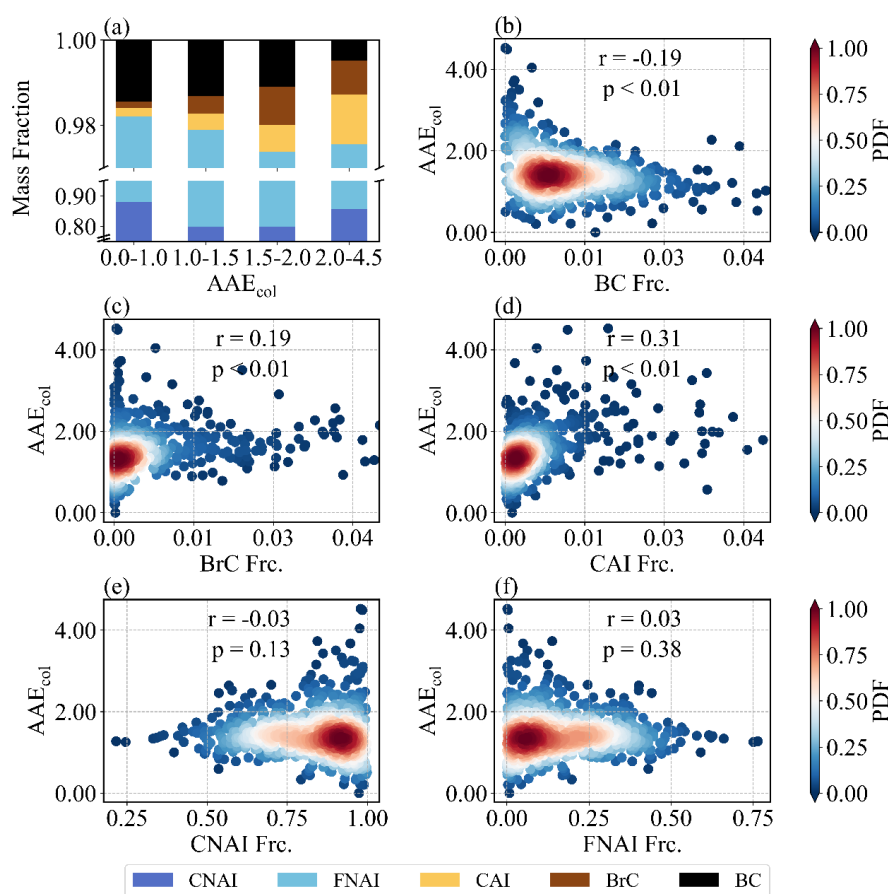
Figure 2. Relationships of near-surface absorption Ångström exponent (AAE_{sfc}) with chemical composition and particle size. (a) Scatter plots of chemical mass concentration fraction and total aerosol AAE_{sfc} . (b) Relationship between AAE_{sfc} and



330 carbonaceous aerosol AAE (AAE_{CA}), with symbol size representing the non-dust
331 water-soluble ions (nd-WSII) fraction and color denoting the fine mineral dust (FMD)
332 fraction. (c–d) Correlations of AAE_{sfc} with mean particle diameters derived from SMPS
333 (c) and APS (d), respectively. Shaded colors indicate probability density.

334 **3.3 Quantitative Contributions of Composition and Size to Columnar AAE**

335 The relationships between mass fractions of chemical composition and AAE_{col}
336 were studied using AERONET data (Fig. 3). The AAE_{col} (1.47 ± 0.56) was also
337 suggested to be greater than that derived from the surface field campaign (Fig. 1),
338 partially due to the vertical variation in aerosol absorption (Guan et al., 2024). The
339 CNAI and FNAI mass fractions varied little and exhibited no significant correlations
340 with AAE_{col} ($r = -0.03$ and 0.03 , $p > 0.1$), indicating a negligible role in setting the
341 absorption spectral dependence. In contrast, BC, BrC, and CAI displayed clear
342 associations. Higher AAE_{col} (>1.5) were associated with marked increases in BrC and
343 CAI, whereas lower AAE_{col} (<1.5) corresponded to relatively higher BC contributions.
344 Correlation analysis is consistent with these patterns: AAE_{col} was negatively correlated
345 with BC ($r = -0.19$, $p < 0.01$), in line with its weak wavelength dependence, but
346 positively correlated with BrC ($r = 0.19$, $p < 0.01$) and CAI ($r = 0.31$, $p < 0.01$),
347 underscoring the strong wavelength dependence of BrC and dust. These findings were
348 similarly to our surface campaign, particularly regarding dust's amplifying effect on
349 AAE_{col} .



350

351 **Figure 3. Relationships between columnar absorption Ångström exponent (AAE_{col})**

352 **and major aerosol chemical compositions. (a)** Mass fractions of fine-mode non-

353 absorbing soluble matter (FNAI), coarse-mode non-absorbing soluble matter (CNAI),

354 coarse-mode absorbing soluble matter (CAI, representing dust), brown carbon (BrC),

355 and black carbon (BC) across different AAE_{col} bins. (b-f) Correlations between AAE_{col}

356 and the mass fractions of BC, BrC, CAI, CNAI, and FNAI, respectively; shaded colors

357 denote probability density.

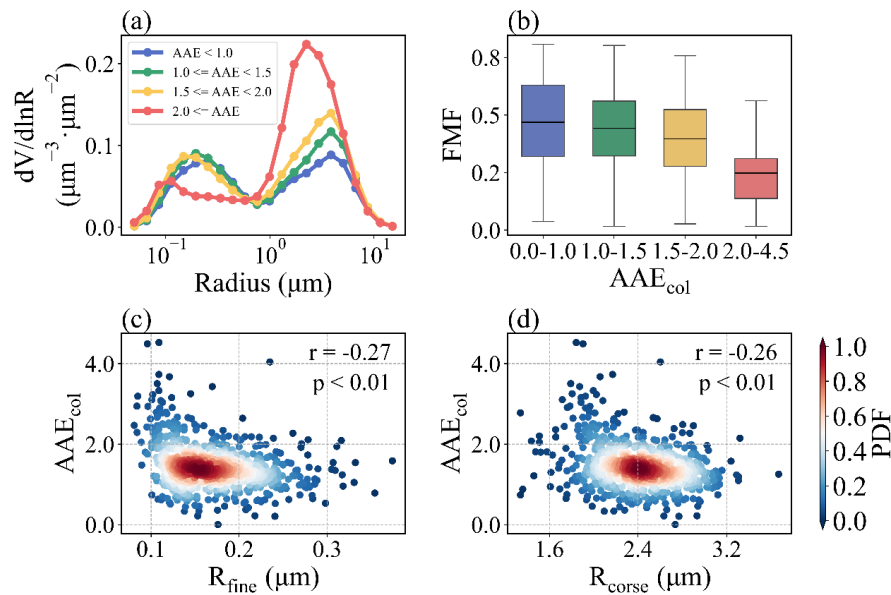
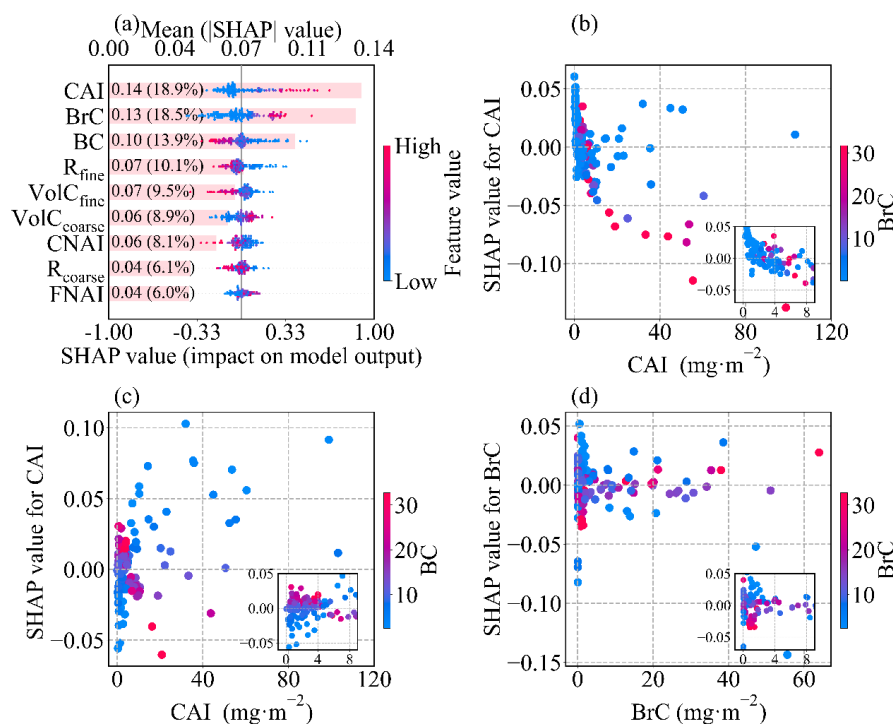


Figure 4. Relationships between columnar absorption Ångström exponent (AAE_{col}) and aerosol size distribution characteristics. (a) Volume size distributions grouped by AAE_{col} bins. (b) Fine-mode fraction (FMF) by AAE_{col} bins, with horizontal lines indicating means. (c–d) Correlations of AAE_{col} with effective radius of fine-mode (R_{fine}) and coarse-mode (R_{coarse}) particles, respectively.

The impacts of aerosol size distribution on AAE_{col} were clearly reflected (Fig. 4). With AAE_{col} increasing, the peaks of both fine and coarse modes shifted to smaller sizes, indicating an overall refinement of the size distribution. The R_{fine} and R_{coarse} were both negatively correlated with AAE_{col} ($r = -0.27$ and -0.26 , $p < 0.01$), demonstrating that reductions in particle size in both modes enhanced the spectral dependence. Interestingly, FMF decreased with increasing AAE_{col} , suggesting that coarse-mode particles retained a substantial volumetric contribution even under high AAE_{col} conditions.



372

373 **Figure 5. SHAP interpretation of the machine-learning model for columnar AAE.**

374 (a) Attribution of chemical compositions and particle size parameters to columnar AAE.

375 Each horizontal bar represents the mean absolute SHAP value of a feature, indicating

376 its overall impact on the model output; the color gradient shows the effect of feature

377 values on columnar AAE, with red indicating a positive influence and blue indicating a

378 negative influence. Features are ranked by importance. (b-c) SHAP dependence plots

379 corresponding to CAI. (d) SHAP dependence plots corresponding to BrC.

380 Machine learning analysis further quantified relative contributions, as illustrated

381 in Fig. 5a. It is found that showed that CAI had the strongest explanatory power,

382 accounting for ~19% of the model performance, confirming the dominant role of dust

383 in amplifying spectral absorption. BrC was second (18.5%) and BC was third (13.9%),



384 together with CAI explaining ~50% of model performance. Among the size-related
385 predictors, R_{fine} alone accounted for about one quarter (~29%) of the cumulative
386 importance of all size metrics, making it the most influential size parameter. Its
387 importance was also clearly higher than CNAI and FNAI, indicating that the mean
388 radius of fine-mode particles is the key size control on AAE_{col} . During the prediction
389 process, it is observed that higher values of BrC, CAI, and volume concentrations of
390 coarse-mode ($\text{vol}_{\text{coarse}}$) corresponded to higher SHAP values and higher values of other
391 predictors corresponded to smaller SHAP values. These responses are fully consistent
392 with the correlations between AAE_{col} and these parameters (Fig. 3b–d; Fig. 4c–d).
393 Collectively, these results demonstrated that AAE_{col} is not governed by BC or BrC alone;
394 it is primarily modulated by dust and secondarily by particle-size structure (size metrics
395 together ~35%), underscoring the need to account for both composition and size when
396 evaluating spectral absorption.

397 To further investigate the interaction effects of major parameters on AAE_{col}
398 prediction, we selected CAI, BrC, and BC, the three most influential predictors by
399 SHAP, to analyze their interactions (Fig. 5b–5f). When CAI loading was below 10
400 $\text{mg}\cdot\text{m}^{-2}$, BrC suppressed the positive impact of CAI loading and progressively drove it
401 toward a negative contribution. When CAI loading exceeded 10 $\text{mg}\cdot\text{m}^{-2}$, lower BrC is
402 more likely than higher BrC to sustain or enhance the positive marginal effect of CAI,
403 although the magnitude is weaker than in the CAI loading $< 10 \text{ mg}\cdot\text{m}^{-2}$ regime. For
404 CAI loading in the range 0–4, higher BC yielded a positive marginal effect of CAI on



AAE_{col}, whereas lowed BC yielded a negative one. Once CAI loading is greater than 4 mg·m⁻², this relationship reverses, with higher BC more likely to make further increases in CAI contribute negatively to AAE_{col}. As BrC increases, BC progressively reduces the positive and negative contributions of BrC to AAE_{col}. These interactions indicate that models of aerosol spectral absorption should explicitly represent the mutual constraints among CAI, BrC, and BC to better identify and quantify AAE drivers.

3.4 Radiative Forcing and Efficiency Responses to Columnar AAE in Beijing

Joint analysis of the boxplots and SHAP diagnostics revealed a robust, layer-dependent coupling between the AAE_{col} and ADRF. As AAE_{col} increased from 0–1 to 2–4.5, cooling at the TOA intensifies, atmospheric heating weakened, and cooling at the BOA was alleviated (Fig. 6a-6c), reflecting a transition from BC-dominated, low-SSA conditions to high-SSA regimes influenced by BrC and mineral dust. SHAP method confirmed that AAE_{col} is the third strongest driver (~16%) after AOD (~56%), and comparably to SSA (~18%) at TOA and consistently shifts ADRF toward more negative values (Fig. 6d). At BOA, AAE_{col} contributes only ~4%. BOA cooling is primarily regulated by AOD (~65.0%) and SSA (~16 %) (Fig. 6e). In the ATM, AOD and SSA jointly dominated, but AAE_{col} still contributed comparably to surface albedo (SA) (both ~12%) (Fig. 6f), underscoring its role in vertically redistributing radiative energy. Mechanistically, higher AAE_{col} is associated with BrC and dust, which exhibit higher SSA but lower mass absorption efficiencies (MAE), thereby enhancing backscattering and solar escape (more negative TOA forcing), reducing absorption



(weaker atmospheric heating), and producing a net transmission effect that mitigates
BOA cooling.

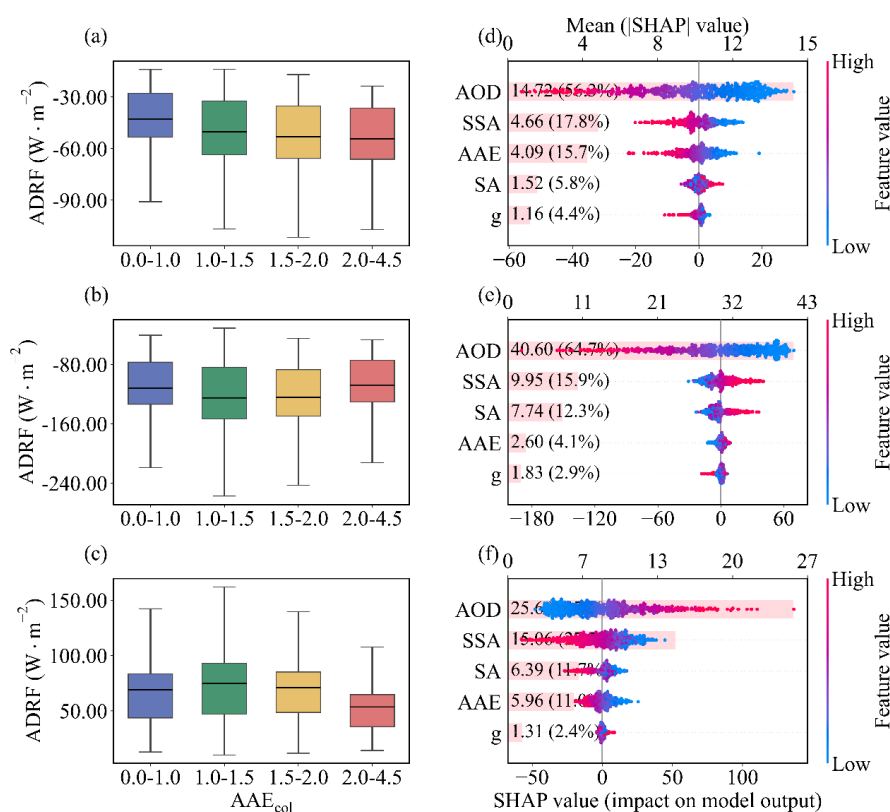
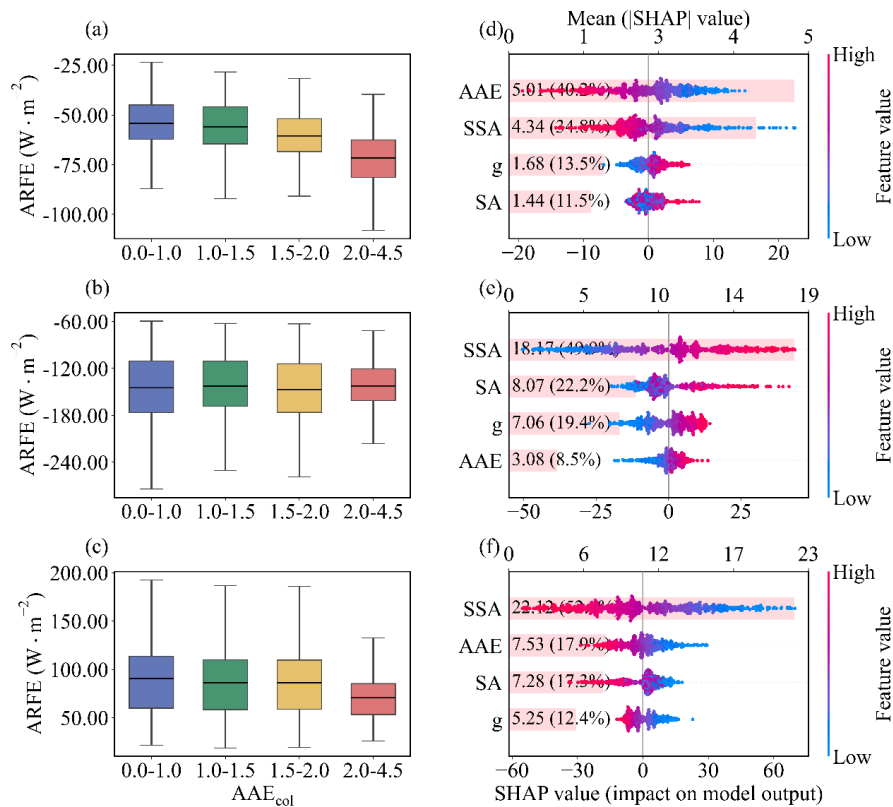


Figure 6. Role of columnar absorption Ångström exponent (AAE_{col}) in regulating aerosol direct radiative forcing (ADRF). (a–c) Box plots of ADRF at the top of the atmosphere (a), bottom (b), and in the atmosphere (c) as a function of AAE_{col} . (d–f) SHAP analysis quantifies the relative contributions of aerosol optical depth (AOD), single scattering albedo (SSA), asymmetry parameter (g), surface albedo (SA) and AAE_{col} in driving ADRF variations at the top of the atmosphere (d), at the bottom of the atmosphere (e), and in the atmosphere (f). The mean absolute SHAP values



(numbers in parentheses) indicate the relative contribution of each predictor to the model output.



438

Figure 7. Role of columnar absorption Ångström exponent (AAE_{col}) in regulating aerosol radiative forcing efficiency (ARFE). (a–c) Box plots of ARFE at the top of the atmosphere (a), bottom (b), and in the atmosphere (c) as a function of AAE_{col} . (d–f) SHAP analysis with AOD fixed at its median (50th percentile) quantifies the relative contributions of single scattering albedo (SSA), asymmetry parameter (g), surface albedo (SA) and AAE_{col} in driving ARFE variations at the top of the atmosphere (d), bottom (e), and in the atmosphere (f). The mean absolute SHAP values (numbers in



parentheses) indicate the relative contribution of each predictor to the model output.

To better show columnar AAE's impact on ADRF, we introduced the ARFE, which removes the scaling by aerosol loading and highlights intrinsic optical controls. At TOA, AAE_{col} was the dominant driver of cooling efficiency based on mean $|SHAP|$ (~40.0%), exceeding the asymmetry factor (g), SSA, and SA even when AOD was conditioned at 25th (Fig. S10), 50th (Fig. 7), 75th percentiles (Fig. S11), or mean (Fig. S12). Larger AAE_{col} was associated with more negative TOA ARFE, consistent with stronger shortwave backscattering by BrC and dust (Fig. 7d). At BOA, ARFE was governed primarily by SSA (~50%), followed by g and SA, with AAE_{col} contributing more modestly (~8%), increases in SSA and g tended to make ARFE less negative (Fig. 7e), indicative of enhanced forward scattering and greater transmittance for a fixed AOD. In the ATM, SSA dominated the heating-efficiency (>50%), with AAE_{col} and SSA providing secondary control (both ~17%), while g played a minor role (Fig. 7f). Higher AAE_{col} is linked to lower atmospheric heating efficiency, reflecting a shift toward aerosol types with weaker mass absorption than BC, and higher SSA further suppresses in-column absorption.

4 Conclusions

LAAs exert a strong influence on the Earth's radiation budget, yet the spectral dependence of their absorption, commonly summarized by the AAE, remains poorly constrained in urban regions. Here we combined in situ observation with long-term



466 AERONET column data and an interpretable machine-learning framework to
467 disentangle the drivers of the AAE and to clarify its radiative implications.

468 Near the surface in wintertime Beijing, AAE variability co-varied primarily with
469 enhanced fractions of fine mineral dust and water-soluble inorganic ions, underscoring
470 that non-carbonaceous species can substantially modulate local absorption spectra in
471 addition to BC and BrC. At the column level, SHAP diagnostics identified CAI as the
472 leading driver of columnar AAE, followed by BrC and BC. Among particle size metrics,
473 the fine-mode effective radius emerged as the most influential variable and accounted
474 for about 29% of the cumulative importance of all size parameters, whereas non-
475 absorbing composition (coarse and fine non-absorbing dust and non-absorbing
476 carbonaceous aerosols) played only a minor role.

477 The radiative analysis shows that columnar AAE is a key regulator of ADRF at the
478 TOA. Columnar AAE explained about 16% of the variance in TOA ADRF, comparable
479 to the contribution from SSA, and became the leading predictor of TOA ARFE,
480 accounting for roughly 40% of its variability and systematically enhancing TOA
481 cooling efficiency. By contrast, the direct influence of AAE on ADRF and ARFE in the
482 ATM and at the BOA was clearly weaker, where SSA and loading-related quantities
483 remained dominant.

484 The findings of our study demonstrate the multifactorial control of AAE by
485 composition and size and highlight its pivotal role in partitioning radiative forcing
486 vertically among TOA, ATM, and BOA, especially at the TOA. Consequently,



487 accurately constraining AAE and its drivers is essential for a realistic representation of
488 aerosol radiation interactions in regional and global models.

489 **Data and code availability**

490 The data that support the findings of this study are available in the Zenodo data
491 repository (<https://doi.org/10.5281/zenodo.17852818>, Wang et al., 2025b). The
492 AERONET data is freely available on the AERNOET website
493 (<https://aeronet.gsfc.nasa.gov/>). The aerosol chemical composition derived from
494 AERONET inversion data is available from [https://doi.org/10.1175/BAMS-D-23-](https://doi.org/10.1175/BAMS-D-23-0260.1)
495 0260.1. The code scripts are also available in the Zenodo data repository
496 (<https://doi.org/10.5281/zenodo.17852818>, Wang et al., 2025b).

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508 **Author contribution**

509 PT designed the study. PT and WW received the funding. WW, PT, SZ, ZY and
510 MC analyzed the data and led the writing. WW, YZ, YW and PT conducted the field
511 campaign. WW, PT, SZ, YZ, ZY, CC, YW, MC and LZ contributed to discussion, review,
512 and edit the manuscript.

513 **Competing interests**

514 The authors declare that they have no conflict of interest.

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