



Drone-based multispectral differentiation of subalpine vegetation at the treeline in the Southern Alps of New Zealand

Fabian Döweler^{1*} and Martin Karl-Friedrich Bader^{2*}

¹ Dragonfly Data Science, Wellington, New Zealand

² System Earth Science, Maastricht University, Brightlands Campus Greenport Venlo, The Netherlands

*Correspondence:

fab.doe@gmx.de, martin.bader@maastrichtuniversity.nl

Keywords

Vegetation index, unoccupied aerial vehicle (UAV), machine learning classifier, mountain biodiversity

Abstract

Subalpine ecosystems are highly dynamic environments that are particularly vulnerable to environmental change, yet their remote and rugged nature poses challenges for long-term monitoring. Unoccupied aerial vehicles (UAVs) equipped with multispectral sensors offer a scalable solution for high-resolution vegetation mapping in these landscapes. In this study, we integrated UAV-derived spectral data with machine learning (ML) classifiers to assess the effectiveness of different vegetation indices (VIs) in distinguishing subalpine plant communities. Principal component analysis (PCA) revealed that NDVI, SIPI2, MCARI, and CHL were highly correlated and strongly influenced the primary variance in the dataset, while NDRE and LCI contributed more evenly across principal components, and GNDVI was largely independent. Among the ML classifiers tested, extreme gradient boosting (XGBoost) achieved the greatest overall accuracy (81.3%) and Kappa (0.75), outperforming support vector machines (SVM) and random forest (RF). Classification confidence was highest for *Chionochloa* tussock (64.6–69.7%) and *Dracophyllum* scrub (70.6%), suggesting moderate reliability for these dominant vegetation types. Scrub and prostrate mat-forming communities exhibited lower classification accuracy, likely due to their heterogeneous canopy structure and greater spectral variability. The ecological boundaries of the subalpine zone, formed by *Fuscospora* forest and scree, were classified with high confidence, but the vegetation is dominated by tussock and shrubland. Feature importance analysis ranked NDVI, SIPI2, CHL, and MCARI highly in SVM and RF models, whereas LCI prevailed in XGBoost, underscoring how different algorithms leverage spectral information in classification tasks. These results emphasize the role of vegetation structure in classification accuracy, with dense, spectrally homogeneous vegetation types more reliably distinguished than mixed-species communities. Our study highlights UAV-based classification as a valuable tool for landscape-scale monitoring of subalpine vegetation. As UAV applications and ML workflows continue to evolve, optimizing classification approaches will enhance our ability to track ecological changes in subalpine and alpine regions worldwide.



49 1. Introduction

50

51 Globally, subalpine shrublands play a crucial role as biodiversity hotspots, supporting a wide
52 range of endemic plant species and serving as critical habitats for various alpine-adapted
53 fauna. Functioning as ecological transition zones, they facilitate interactions between
54 species from lower and higher elevations, leading to unique assemblages. These
55 ecosystems are also vital for carbon sequestration (Day et al., 2023), water regulation
56 (Nicholls, 2023; Nicholls & Carey, 2021), and plant-soil nutrient balance (Urbina et al., 2020),
57 exhibiting key ecosystem services in mountainous landscapes. Additionally, subalpine
58 vegetation plays a fundamental role in global mountain ecosystems by acting as a crucial
59 buffer against climate-driven changes (Hou, 2024) and microhabitat facilitation (Harsch et
60 al., 2009). These environments are highly dynamic, shaped by the interplay of climate,
61 topography, and ecological processes. However, ongoing climate change is reshaping
62 subalpine ecosystems worldwide, altering species distributions, ecosystem functions, and
63 landscape stability (Reid et al., 2022).

64

65 The vulnerability of alpine and subalpine vegetation to changing abiotic drivers is particularly
66 concerning, as many species possess limited dispersal capacities and may be limited in their
67 capability to respond to changes in their range limitations (Camac et al., 2021). Moreover,
68 research on climate-driven treeline shifts (Körner, 2014) and microclimatic variability
69 underscores the complexity and potential of subalpine ecosystems to modulate large scale
70 abiotic drivers (Döweler et al., 2021, 2024) and biotic effects (e.g. control invasive species
71 expansion; (Padalia et al., 2023). While some species may benefit from a warming climate,
72 others, particularly alpine specialists, may not be able to compete with generalist species
73 expanding their range from lower elevations (Thomas et al., 2023). This ecological
74 reshuffling has profound implications for biodiversity, carbon storage, and ecosystem
75 resilience, but often happens gradual and can only be thoroughly studied at the landscape
76 scale, where large scale assessments of change of the subalpine lacks temporal and spatial
77 resolution to adequately reflect these changes (Döweler et al., 2024).

78

79 In New Zealand, subalpine vegetation is characterized by a mosaic of tussock grasslands
80 and low-stature shrubs, forming ecologically significant communities that influence
81 ecosystem resilience and carbon storage (Mark, 2013; Day, 2023). The high cover of
82 *Chionochloa* tussocks (Fig. 1), along with species such as *Dracophyllum uniflorum*,
83 *Podocarpus nivalis*, and *Acrothamnus colensoi*, creates microclimates that buffer
84 temperature extremes and support treeline regeneration of *Leuphazonia menziesii* and
85 *Fuscospora cliffortioides* (Hook.f.) Oerst. (Döweler, 2021; Scherrer & Körner, 2010). These
86 ecosystems are not only important for biodiversity but also provide a range of ecological
87 functions, influencing water retention (van Galen et al., 2023) and soil carbon dynamics (Day
88 et al., 2023).

89



90

91 **Figure 1** A typical subalpine belt vegetation composition in New Zealand (1365 m a above
 92 sea level, Craigieburn Valley, Arthurs Pass)

93 Despite their ecological significance, New Zealand's subalpine landscapes are experiencing
 94 complex transformations, yet our understanding of these changes remains limited. The lack
 95 of landscape-scale detailed vegetation mapping in the subalpine belt limits insights to
 96 capture these gradual but potentially significant shifts in over time (Day et al., 2023) in these
 97 often inaccessible regions. While treelines have shown limited upslope movement in
 98 response to warming (Harsch et al., 2012), subalpine vegetation composition is shifting due
 99 to factors such as woody encroachment and thickening, increased drought stress, and the
 100 potential expansion of invasive species (Chardon et al., 2024; Singh et al., 2024).
 101 Understanding how these communities are responding to climate change is crucial to further
 102 elucidate their role in stabilising these ecosystems to help inform conservation strategies
 103 (Reid, 2022; De Toma, 2025).

104

105 Remote sensing is a powerful tool for monitoring subalpine environments, where limited
 106 accessibility and the need for large-scale landscape assessments pose significant
 107 challenges (Walsh et al., 2009). It enables the classification of vegetation and detection of
 108 ecological shifts, offering a comprehensive perspective on mountain biodiversity and
 109 ecosystem dynamics. As climate change increasingly affects alpine and subalpine
 110 ecosystems, the ability to remotely assess vegetation composition across vast and often
 111 inaccessible areas has become a powerful method to study these ecosystems (Garbarino et
 112 al., 2023). Advances in high-resolution satellite and UAV-based remote sensing, combined
 113 with machine learning, have significantly improved vegetation classification, enhancing
 114 mapping accuracy and long-term monitoring (Mashiane et al., 2024; Nguyen et al., 2022).
 115 Access to light-weight sensors which can readily be mounted on increasingly affordable
 116 unoccupied aerial vehicles (UAV's) enables us to monitor subalpine ecotone in
 117 unprecedented detail using optical, multispectral, thermal, and LiDAR sensors (Döweler et
 118 al., 2024). Remote sensing technologies provide a robust means of tracking vegetation
 119 dynamics at ecologically meaningful scales, with satellite and aerial imagery proving
 120 effective in mapping subalpine vegetation and detecting temporal changes (De Toma et al.,
 121 2025). In some field studies, UAV-based deep learning methods may outperform human
 122 observers in delineating complex patterns in subalpine shrub communities (Moritake et al.,
 123 2024), endorsing their use for larger mapping endeavours of the subalpine in an approach to
 124 more accurately study vegetation shifts in response to climate change. These technological
 125 advancements offer critical insights for conservation planning and land management,
 126 ensuring more effective strategies for protecting subalpine ecosystems (Padalia et al., 2023).

127



This study investigates the potential of remotely sensed optical and multispectral vegetation indices to differentiate vegetation composition in a complex subalpine shrubland ecotone in New Zealand. Building on our previous classification and segmentation research in the Craigieburn Range (Arthur's Pass, New Zealand, Döweler et al., 2024), we aimed to compare the performance of three widely used machine learning classifiers (support vector machine, random forest and extreme gradient boosting) and, by extension, to identify which vegetation indices are most effective for distinguishing vegetation classes. This study aims to offer recommendations for vegetation indices and ML classifiers for future remote sensing applications in subalpine ecosystems. We hypothesised that classification accuracy will be highest in vegetation types with distinct spectral reflectance signatures and relatively low spectral and structural variability (e.g. sparsely vegetated scree, subalpine forest, tussock) and lowest in various types of scrub and prostrate mats where increased species interspersation may cause greater spectral overlap.

2 Material and Methods

2.1 Study site

The Craigieburn Valley study site (-43.111, 171.713) is located at 1365 metres above sea level on a southeast to southwest aspect within the eastern slopes of the Southern Alps, New Zealand (Fig. 2). The site is characterized by a montane to subalpine climate, with frequent frost events throughout the year (approximately 135 frost days annually) and an annual rainfall of around 1300 mm. The subalpine belt is dominated by *Chionochloa* tussock grasslands, interspersed with species such as *Dracophyllum uniflorum* Hook.f., *Podocarpus nivalis* Hook., and *Acrothamnus colensoi* (Hook.f.) Quinn, alongside areas of exposed scree. The adjacent treeline is formed by *Fuscospora cliffortioides* (Hook.f.) Oerst., marking the transition to the alpine treeline.

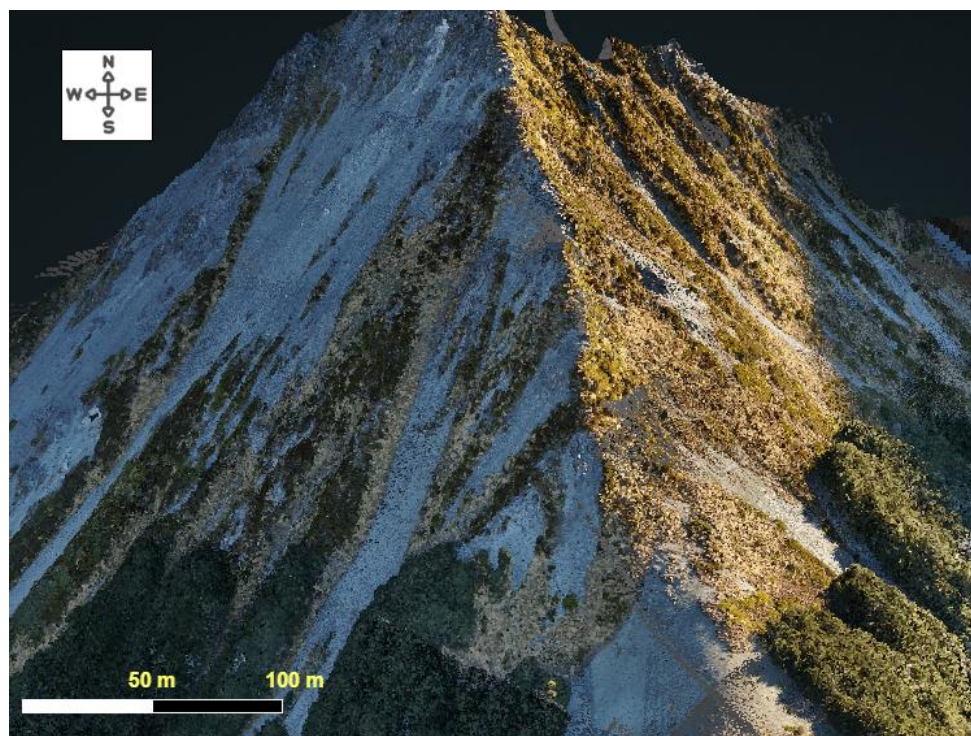


Figure 2 Craigieburn study site (~ 4 ha). The subalpine belt has been covered by the drone survey. The visualisation shows an RGB coloured point cloud derived from a 2019 (May, 9 am austral winter) point cloud acquisition.

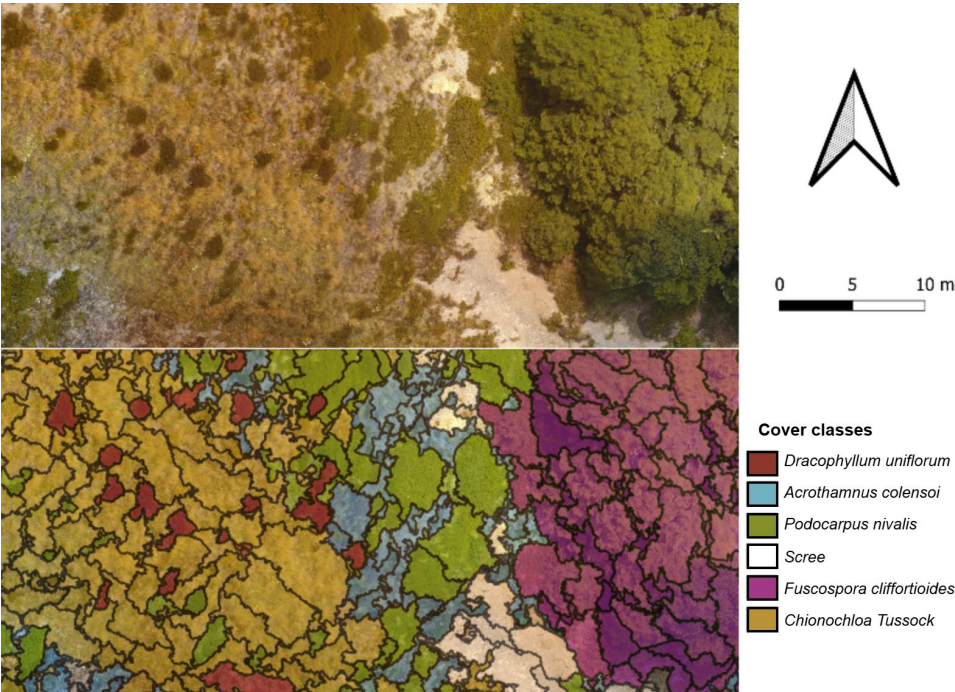
2.2 Subalpine vegetation classification and segmentation

For the classification and segmentation of the subalpine belt we we applied the same methodology as previously described in Döweler et al., 2024 producing a detailed map (0.05 m) of the subalpine vegetation, with low spatial offset (6.14 ± 4.03 cm; mean \pm standard error) over an area of approximately 4 ha covering the supaline belt. We provide a brief summary of the method below, for the full workflow please see the respective paper.

For georeferencing, we used a differential GPS rover-base setup (Emlid Reach RS+) to locate ground control points (GCPs) consisting of chessboard-patterned panels or high-visibility spray markers, which were distributed across the UAV flight path for photogrammetric orthorectification. The differential GPS was also used to validate classification results by geolocating vegetation patches within the subalpine, resulting in over 600 ground-truth vegetation identifications in Craigieburn Valley. UAV imagery was captured using a DJI Phantom 4 (0.01 m RGB resolution) and a Parrot Sequoia+ multispectral sensor (0.05 m resolution), mounted for aerial surveys during the 2018/19 austral growing season. Pre- and post-flight calibration followed the One-Point Calibration plus Sunshine Sensor method using the Parrot target plate. The UAV flight paths were planned using UgCS software (SPH Engineering, 2025), incorporating an 8 m resolution digital terrain model



182 (Geographx, 2016) for altitude control, with a 5 m/s flight speed and 80% along-track overlap
183 for photogrammetric processing.
184
185 Post-processing was performed in Pix4D 9.4 (Pix4D, 2025) for image orthorectification and
186 spatial alignment using mapped GCPs. RGB and multispectral datasets were aligned
187 manually and projected into the New Zealand Transverse Mercator (NZTM 2000) coordinate
188 system. Object-based image analysis was conducted using eCognition Developer 9.0
189 (Trimble, 2025), applying multiresolution segmentation (scale: 250, shape: 0.1,
190 compactness: 0.7) to generate spectrally homogeneous objects. A nearest neighbor
191 classifier, trained on geolocated field data, was used to categorize land cover into five
192 vegetation types and scree. Feature selection was optimized through the Feature Space
193 Optimization tool in eCognition, maximizing vegetation separability across spectral bands.
194 Classification accuracy was evaluated using an error confusion matrix based on the 600
195 ground-truth points, with Kappa statistics indicating an overall accuracy of 89.7% (Fig. 3,
196 Döweler et al., 2024)
197



198
199 **Figure 3** Results for the pre-existing subalpine vegetation segmentation and classification
200 for the Craigieburn valley established by Döweler et al. (2024)
201
202

203 **2.3 Vegetation indices used to discern cover classes**

204
205 Following the segmentation and classification of the vegetation cover classes, we derived a
206 suite of vegetation indices from the available multispectral bands to test their capability in
207 discerning cover classes. These indices capture plant functional traits that influence



208 productivity, stress responses, and spectral variability across different vegetation types. By
209 leveraging multispectral reflectance data, we aimed to improve vegetation classification
210 accuracy and investigate how vegetation properties vary across topographic gradients.

211
212 All UAV-derived spectral data was sampled to a uniform 0.01 m resolution using a bicubic
213 interpolation in GDAL (3.4.1; Rouault et al., 2025). Vegetation classification polygons were
214 processed in GeoPandas (1.0.1; Van den Bossche et al., 2024) by extracting their centroids,
215 which were then buffered by 0.2 m to create localized sampling zones. This buffering step
216 minimized the influence of mixed-pixel effects and ensured that spectral values extracted
217 within each zone were representative of potential spectral heterogeneity within a single
218 vegetation type. We used a spatial filtering GeoPandas, ensuring that each buffered centroid
219 remained entirely within its original vegetation class. After filtering, the final vegetation
220 sample counts were 4436 *Fuscospora cliffortioides*, 1809 scree, 1384 *Chionochloa* tussock
221 (-12%), 883 *Acrothamnus colensoi* (-13%), 676 *Podocarpus nivalis* (-10%), and 293
222 *Dracophyllum* spp. (-3%), with a total of 9476 vegetation samples retained for analysis. The
223 filtered dataset was used to compute zonal statistics for each vegetation index using the
224 rasterstats (0.15.0; Perry, 2025) package, extracting median, mean and standard deviation
225 values within each segmented vegetation type.

226
227 The vegetation indices used in this analysis (Table 1) are widely applied in remote sensing
228 analyses on vegetation health and growth. We included the Normalized Difference
229 Vegetation Index (NDVI), Structure Insensitive Pigment Index 2 (SIPI2), Modified Chlorophyll
230 Absorption Ratio Index (MCARI), Green Normalized Difference Vegetation Index (GNDVI),
231 Chlorophyll Index (CHL), Normalized Difference Red Edge Index (NDRE), and Leaf
232 Chlorophyll Index (LCI). NDVI was calculated to assess overall vegetation health, while
233 SIPI2 was included as an indicator of vegetation stress and pigment ratios. MCARI, GNDVI,
234 CHL, and LCI were shifting the focus on chlorophyll concentrations and photosynthetic
235 potential, while the NDRE has been selected to provide a focus on red-edge detectable
236 indication of early stress.

237
238 **Table 1** Overview of vegetation indices and respective band calculations used in the current
239 study, which represent unpublished data from a previous remote sensing study in the same
240 area (Döweler et al., 2024). Green (550 nm), Red (660 nm), Near-infrared (790 nm, NIR),
241 Red Edge (735 nm, RE) derived from the Parrot Sequoia, Blue band (450 nm) extracted
242 from the Phantom 4 RGB sensor.

243

Index	Name	Calculation	Reference
NDVI	Normalized Difference Vegetation Index	$\frac{NIR - Red}{NIR + Red}$	Rouse et al., 1973
SIPI2	Structure Insensitive Pigment Index 2	$\frac{NIR - Blue}{NIR - Red}$	Peñuelas et al., 1995
MCARI	Modified Chlorophyll Absorption Ratio Index	$[(RE - Red) - 0.2 \times (Red\ Edge - Green)] \times \frac{RE}{Red}$	Daughtry et al., 2000



Index	Name	Calculation	Reference
GNDVI	Green Normalized Difference Vegetation Index	$\frac{NIR - Green}{NIR + Green}$	Gitelson & Merzlyak, 1998
CHL	Chlorophyll Red-Edge Index	$\frac{NIR}{RE} - 1$	Gitelson et al., 2003
NDRE	Normalized Difference Red Edge Index	$\frac{NIR - RE}{NIR + RE}$	Barnes et al., 2000
LCI	Leaf Chlorophyll Index	$\frac{NIR - RE}{NIR + Red}$	Haboudane et al., 2002

2.4 Statistical analysis

All statistical computations and graphics were performed using the R software within the RStudio integrated development environment (R version 4.4.1, R Core Team, 2024, RStudio version 2024.09.0+375, Posit team, 2024). To provide a general overview, we performed a principal component analysis (PCA) based on the overall medians of the vegetation indices (i.e. the median of the sample medians of each index). The PCA results were visualised in a biplot to illustrate similarities among vegetation types and the contribution of VIs to the principal components (loading vectors) as well as their interrelationships.

We used three popular machine learning approaches to classify the six vegetation types based on the centroid buffered medians of seven features (vegetation indices): NDVI, GNDVI, CHL, LCI, MCARI, NDRE, SIPI2. We created a balanced 80/20 training-to-test split of our data, ensuring random sampling within each class to preserve the overall class distribution (balanced splits are obtained by providing a factor, i.e. the vegetation type labels, to the *createDataPartition* function, R package *caret*, Kuhn, 2008)

We trained a Support Vector Machine (SVM) classifier using a radial basis function (RBF) kernel to differentiate between vegetation types (R package *e1071*, Meyer et al., 2024). To address class imbalances, we assigned class weights inversely proportional to their frequencies (following the approach in the Python package *scikit-learn*, Pedregosa et al., 2011). Hyperparameter optimization was performed via 5-fold cross-validation, selecting the optimal values of C (regularization parameter) and γ (kernel coefficient, controlling the influence of data points) using a grid search. The final model featured an RBF kernel with C = 10 and $\gamma = 0.01$.

We also ran a random forest approach to classify the vegetation types (R package *randomForest*, Liaw & Wiener, 2002). The model algorithm was run with the default 500 number of trees and a hyperparameter tuning procedure suggested two randomly sampled features (predictors) at each split ($mtry = 2$). Stratified sampling was used to ensure that each tree was trained on a random sample containing observations from all vegetation types.

In addition, we applied extreme gradient boosting (function *xgb.train* in R package *xgboost*, Chen et al., 2024) (Chen et al., 2024) for vegetation type classification. A grid search was used for hyperparameter tuning (final hyperparameter settings: $\eta = 0.05$, $max_depth = 6$, $\gamma = 2$) and 5-fold cross-validation to determine the optimal number of iterations (100 cross-validation runs allowing a maximum of 500 iterations yielded a mean of 150 iterations).

For all three machine learning approaches, the model performance scores were derived from a confusion matrix contrasting true and predicted class labels (*confusionMatrix* function



in R package *caret*, Kuhn, 2008). Class-level sensitivity (recall) and specificity scores were each averaged to an overall score. In addition, the multi-class area under the receiver operating characteristic curve (AUC_{mc}) was calculated as the mean of the class-specific AUCs (function *multiclass.roc* in R package *pROC*, Robin et al., 2011). Finally, we applied a permutation-based feature importance analysis to all three classifiers. This feature importance procedure was run with 30 permutation rounds using a cross-entropy loss function to evaluate feature contribution (*feature_importance* function in R package *ingredients* relying on the cross-entropy loss function in R package *DALEX*, Biecek, 2018; Biecek et al., 2023).

3 Results

3.1 Principal component analysis (PCA)

The PCA showed that PC1 explained roughly 95 % and PC2 nearly 5 % of the variation in the aggregated data. In the PCA biplot, *Podocarpus* scrub and *Dracophyllum* scrub formed the only discernible cluster (which could perhaps include *Chionocholea* tussock), indicating similar profiles related to the VIs. By contrast, the remaining vegetation types showed distinct profiles (Fig. 4). The small angular distances between the loading vectors of NDVI, SIPI2, MCARI and CHL suggest that these indices were all positively correlated and their fairly horizontal alignment indicates a strong influence on PC1 (Fig. 4). The LCI and NDRE were also positively correlated and contributed roughly equally to both PCs. The GNDVI had a strong negative contribution to PC2, and was weakly or uncorrelated with the other VIs considering the large angular distances to the other loading vectors (Fig. 4).

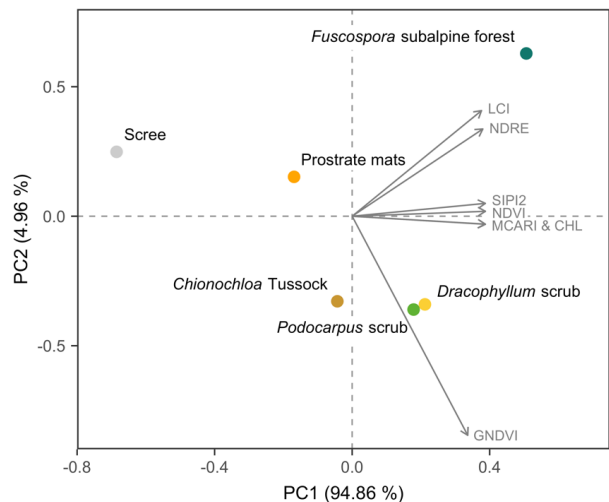


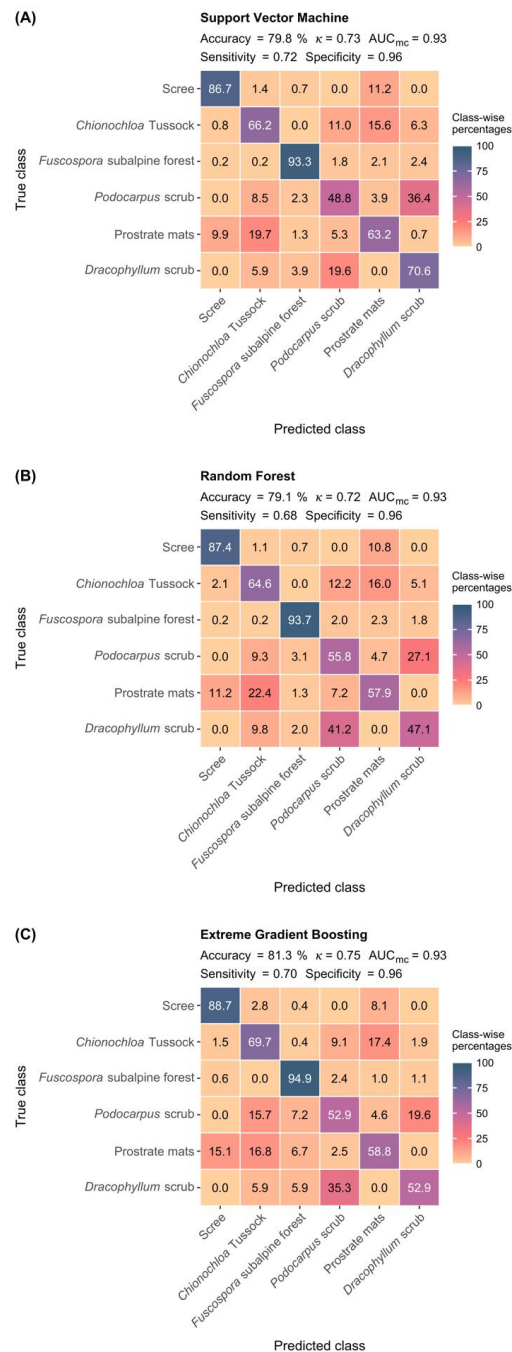
Figure 4 Principal component analysis (PCA) biplot displaying the scores associated with various subalpine vegetation types (filled circles) and the contribution of vegetation indices (arrows indicating loading vectors) to the principal components. The vegetation indices were scaled prior to PCA. Please note that the eigenvectors of the MCARI and CHL indices are virtually equal.

3.2 Machine learning classifiers



320 Based on common performance metrics (accuracy, Cohen's Kappa statistic, AUC, sensitivity
321 and specificity), the support vector machine (SVM) and random forest (RF) classifiers
322 performed similarly well but were slightly outperformed by the extreme gradient boosting
323 algorithm (XGBoost) (Fig. 5). Accuracy was just below 80 % in the SVM and the RF, while
324 reaching 81.3 % in the XGBoost approach. Given our unbalanced data, the more robust
325 Kappa statistic seems more informative than overall accuracy. Cohen's Kappa was highest
326 in the XGBoost classifier at 0.75, compared to 0.73 in the SVM and 0.72 in the RF approach
327 (Fig. 5). The multi-class AUC-ROC and specificity were identical in all three algorithms
328 ($AUC_{mc} = 0.93$, specificity = 0.96), while sensitivity (recall) varied from 0.72 in the SVM and
329 0.70 in the XGBoost to 0.68 in the RF classifier (Fig. 5). As judged by the percentage of
330 correct classifications in the confusion matrices of the three classifiers, *Fuscospora*
331 subalpine forest (93.3 – 94.7 %) and scree (86.7 – 88.7 %) can be identified with high to very
332 high confidence, followed by moderate classification confidence for *Chionochloa* Tussock
333 (64.6 – 69.7 %). Classification confidence for the remaining vegetation types was mostly low
334 (< 60 %), except for the SVM's 70.6 % correct classifications of *Dracophyllum* scrub and its
335 63.2 % accuracy in classifying *Acrothamnus colensoi* (Fig. 5).

336
337



338
339
340 **Figure 5** Heatmaps of the confusion matrices of a support vector machine (A) a random forest model (B)
341 and an extreme gradient boosting machine (C). κ indicates Cohen's Kappa statistic, AUC_{mc} denotes the
342 multi-class area under the ROC curve. Accuracy, κ and AUC_{mc} are overall model statistics, while
343 sensitivity (recall) and specificity indicate averages of the class-specific metrics.

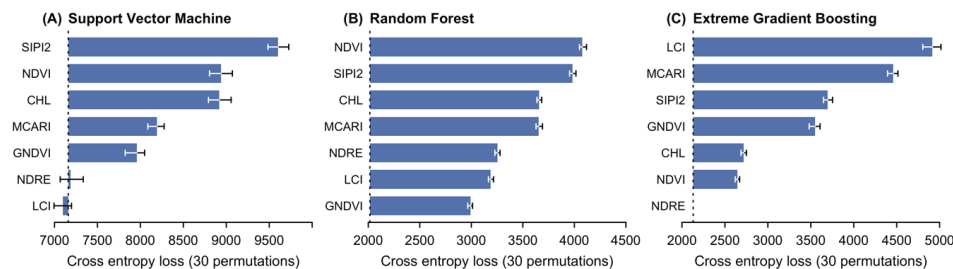


Figure 6 Permutation-based feature importance (variable importance) of a support vector machine (A), a random forest model (B), and an extreme gradient boosting model (C). The vertical dotted lines indicate the cross-entropy loss of the full model (no variables dropped). The error bars signify the 2.5th and 97.5th percentiles of the 30 permutations.

In the SVM algorithm, SIPI2 represented the most important variable, followed by the equally relevant NDVI and CHL, while MCARI and GNDVI were of moderate importance (Fig. 6 A). NDRE and LCI had little to no significance in the SVM. In the RF model, NDVI and SIPI2 were similarly important, followed by equal contributions from CHL and MCARI, and to a lesser but similar extent from NDRE and LCI, with GNDVI coming in with a slightly lower entropy loss value (Fig. 6 B). By contrast, the XGBoost algorithm relied most heavily on LCI, followed by MCARI and, on a similar level, SIPI2 and GNDIV (Fig. 6 C). CHL and NDVI were of lesser importance, while NDRE represented an entropy-irrelevant feature.

4. Discussion

Mountain ecosystems support a high level of plant diversity and endemism, especially at high elevation but coordinated monitoring efforts are scarce (Perrigo et al., 2020), often hampered by limited access. In the future, more affordable drone technology will bolster remote sensing based mapping and monitoring of these hard-to-reach ecosystems, and information that aids in optimising classification accuracy will facilitate progress in this field. Here, we compared the performance of three popular ML classifiers and assessed their feature importance to rank common VIs based on their impact on the classification algorithm.

Judged by the overall accuracy and Cohen's Kappa, the XGBoost algorithm differentiated the existing vegetation types best, but SVM and RF both performed nearly as well. Consistent with our findings, XGBoost outperformed SVM and RF in urban land use-land cover and forest classification tasks (Georganos et al., 2018; Ramdani & Furqon, 2022). In line with our hypothesis, spectrally more distinct vegetation types with lower spectral and structural variation showed the greatest proportion of correct classifications in the confusion matrices of the three classifiers, which is consistent with previous findings reported for diverse wetland ecosystems (Schmidt & Skidmore, 2003). In our study, these spectrally distinct vegetation types included *Fuscospora cliffortioides* subalpine forest and *Chionochloa* tussock, which are characterised by dense, uniform foliage and/or well-defined canopy structures, making them easier to differentiate in the ML classification process (Ollinger, 2011). The sparsely vegetated scree is characterised by a relatively uniform spectral signature, i.e. spectrally bland, which facilitates classification.

The importance of the used VIs was ranked similarly in the SVM and RF but differed greatly in the XGBoost algorithm, suggesting that the different ML approaches rely on distinct spectral properties for classification. The SVM and the RF model agreed closely in regard to



the four most important VIs (SIPI2, NDVI, CHL, MCARI). These VIs reflect pigment content and general vegetation vigor and are closely related to photosynthetic activity, which makes them suitable for distinguishing between broad vegetation types. By contrast, in the XGBoost approach LCI emerged as the most important feature, which had minimal leverage in the other two classifiers. This discrepancy likely reflects XGBoost's greedy search for splits that minimise the loss function, allowing a feature to gain importance even if it is of little relevance in SVM and RF algorithms (Kamdern and Fokoue, 2022). LCI's top ranking in XGBoost suggests it may hold critical information for distinguishing subtle variations in leaf structure or pigment content that were not prioritized in the RF and SVM algorithms. A recent review on plant and vegetation classification based on spectral signatures revealed that besides the biological properties also the methodological approaches, the scale at which the recordings are performed and not least the applied feature selection procedure itself may all have a strong influence on feature importance in ML classifiers (Hennessy et al., 2020). To eliminate the latter source of variation, we applied the same permutation-based feature importance analysis based on cross-entropy loss to all three classifiers (Biecek, 2018; Biecek and Baniecki, 2023).

Another notable finding is the weak contribution of NDRE in all classifiers, indicating its limited role in distinguishing subalpine vegetation types for the subalpine ecotone at our study site, which calls for verification in other high-elevation transition zones. Unlike NDVI and MCARI, which are widely used in vegetation classification, NDRE is often associated with deeper canopy penetration and is particularly useful in detecting nitrogen stress and subtle variations in chlorophyll content (Boiarskii & Hasegawa, 2019). The lack of importance of NDRE in this study suggests that these characteristics were not primary drivers of spectral separability among the subalpine vegetation types analyzed here. Instead, indices related to general canopy structure and pigment concentration (such as NDVI, CHL, and MCARI) proved more effective.

The results of our study highlight the critical role of vegetation structure in classification accuracy, with dense, spectrally uniform vegetation types being more reliably identified than structurally diverse shrublands and mixed-species communities. As UAV technology becomes increasingly accessible, further refinement of vegetation index selection and classification methodologies is essential to capture the often subtle responses of subalpine vegetation to abiotic stressors, which are being exacerbated by climate change. The decline of certain species could lead to the loss of critical microhabitats and climatic niches, which serve as stepping stones for the recruitment of subalpine specialists and treeline forming species (Döweler et al., 2021; Frei et al., 2018; Harsch et al., 2009). For New Zealand, a landscape-scale classification of the subalpine can support monitoring the impact of invasive herbivores on these ecosystems, as their grazing pressure threatens both vegetation dynamics and the region's carbon sequestration potential (Lee et al., 2000). Advancements in remote sensing and machine learning offer novel pathways to improve monitoring efforts, enabling us to more clearly formulate and track conservation targets.

5. Conclusion

The effective integration of vegetation indices with modern ML classifiers presents a powerful tool for tracking ecological shifts, particularly in remote and rugged environments. As UAV operations become more affordable, their application in long-term monitoring will be invaluable for detecting and understanding vegetation changes in otherwise inaccessible regions. Expanding these efforts through global collaboration will provide deeper insights into the poorly understood dynamics of subalpine ecosystems under changing climatic conditions. Given the crucial role of subalpine grasslands and woody vegetation in carbon sequestration (Ward et al., 2014) and other ecosystem services such as maintaining biodiversity, erosion protection, runoff regulation and snow retention, their ecological



trajectories must be closely monitored to inform conservation strategies aimed at mitigating species loss and preserving ecosystem functions.

Data availability

All data for this publication can be requested from the corresponding author.

Author contributions

FD contributed to data acquisition and sample collection and data preparation. MB contributed to data analysis. Both authors contributed equally to the writing of this manuscript.

Competing interests

The contact author has declared that none of the authors has any competing interests.

Acknowledgements

Support for this research was provided by Auckland University of Technology (AUT, Auckland). We would like to thank the Department of Conservation (DOC) for access permissions, as well as Lincoln University (Canterbury, New Zealand) for providing accommodation.

References

- Barnes, E., Clarke, T., Richards, S., Colaizzi, P., Haberland, J., Kostrzewski, M., Waller, P., Choi, C., Riley, E., Thompson, T., & others. (2000). *Coincident detection of crop water stress, nitrogen status and canopy density using ground based multispectral data*. 1619, 6.
- Biecek, P. (2018). DALEX: Explainers for Complex Predictive Models in R. *Journal of Machine Learning Research*, 19(84), 1–5.
- Biecek, P., Baniecki, H., & Izdebski, A. (2023). *ingredients: Effects and Importances of Model Ingredients* (Version 2.3.0) [Computer software]. <https://cran.r-project.org/web/packages/ingredients/index.html>
- Boiarskii, B., & Hasegawa, H. (2019). Comparison of NDVI and NDRE indices to detect differences in vegetation and chlorophyll content. *J. Mech. Contin. Math. Sci*, 4, 20–29.



- 482 Camac, J. S., Umbers, K. D. L., Morgan, J. W., Geange, S. R., Hanea, A., Slatyer, R. A.,
483 McDougall, K. L., Venn, S. E., Vesk, P. A., Hoffmann, A. A., & Nicotra, A. B. (2021).
484 Predicting species and community responses to global change using structured
485 expert judgement: An Australian mountain ecosystems case study. *Global Change*
486 *Biology*, 27(18), 4420–4434. <https://doi.org/10.1111/gcb.15750>
- 487 Chen, T., He, T., Benesty, M., Khotilovich, V., Tang, Y., Cho, H., Chen, K., Mitchell, R.,
488 Cano, I., Zhou, T., Li, M., Xie, J., Lin, M., Geng, Y., Li, Y., Yuan, J., &
489 implementation), Xgb. contributors (base Xgb. (2024). *xgboost: Extreme Gradient*
490 *Boosting* (Version 1.7.8.1) [Computer software]. [https://cran.r-](https://cran.r-project.org/web/packages/xgboost/index.html)
491 [project.org/web/packages/xgboost/index.html](https://cran.r-project.org/web/packages/xgboost/index.html)
- 492 Daughtry, C. S. T., Walthall, C. L., Kim, M. S., de Colstoun, E. B., & McMurtrey, J. E. (2000).
493 Estimating Corn Leaf Chlorophyll Concentration from Leaf and Canopy Reflectance.
494 *Remote Sensing of Environment*, 74(2), 229–239. [https://doi.org/10.1016/S0034-](https://doi.org/10.1016/S0034-4257(00)00113-9)
495 [4257\(00\)00113-9](https://doi.org/10.1016/S0034-4257(00)00113-9)
- 496 Day, N., Lavorel, S., Barratt, B., Dickinson, K., Buckley, H., Christensen, B., Curran, T., &
497 Norton, D. (2023). Predicting ecological change in tussock grasslands of Aotearoa
498 New Zealand. *New Zealand Journal of Ecology*.
499 <https://doi.org/10.20417/nzj ecol.47.3549>
- 500 De Toma, A., Malavasi, M., Marzialetti, F., & Cutini, M. (2025). Unveiling spatial patterns and
501 trajectories of shrub dynamics in Mediterranean alpine ecosystems. *Plant Ecology*,
502 226(2), 149–160. <https://doi.org/10.1007/s11258-024-01479-6>
- 503 Döweler, F. (2021). *Causes of Recruitment Limitation at Abrupt Alpine Treelines* [Auckland
504 University of Technology]. <https://hdl.handle.net/10292/14038>
- 505 Döweler, F., Case, B. S., Buckley, H. L., & Bader, M. K.-F. (2021). High light-induced
506 photoinhibition is not limiting seedling establishment at abrupt treeline ecotones in
507 New Zealand. *Tree Physiology*, 41(11), 2034–2045.
508 <https://doi.org/10.1093/treephys/tpab061>
- 509 Döweler, F., Fransson, J. E. S., & Bader, M. K.-F. (2024). Linking High-Resolution UAV-



- 510 Based Remote Sensing Data to Long-Term Vegetation Sampling—A Novel Workflow
511 to Study Slow Ecotone Dynamics. *Remote Sensing*, 16(5), Article 5.
512 <https://doi.org/10.3390/rs16050840>
- 513 Frei, E. R., Bianchi, E., Bernareggi, G., Bebi, P., Dawes, M. A., Brown, C. D., Trant, A. J.,
514 Mamet, S. D., & Rixen, C. (2018). Biotic and abiotic drivers of tree seedling
515 recruitment across an alpine treeline ecotone. *Scientific Reports*, 8(1), 10894.
516 <https://doi.org/10.1038/s41598-018-28808-w>
- 517 Garbarino, M., Morresi, D., Anselmetto, N., & Weisberg, P. J. (2023). Treeline remote
518 sensing: From tracking treeline shifts to multi-dimensional monitoring of ecotonal
519 change. *Remote Sensing in Ecology and Conservation*, n/a(n/a).
520 <https://doi.org/10.1002/rse2.351>
- 521 Geographx. (2016). *LINZ NZ 8m Digital Elevation Model 2016*.
522 <https://data.linz.govt.nz/layer/51768-nz-8m-digital-elevation-model-2012/>
- 523 Georganos, S., Grippa, T., Vanhuyse, S., Lennert, M., Shimoni, M., & Wolff, E. (2018). Very
524 High Resolution Object-Based Land Use–Land Cover Urban Classification Using
525 Extreme Gradient Boosting. *IEEE Geoscience and Remote Sensing Letters*, 15(4),
526 607–611. *IEEE Geoscience and Remote Sensing Letters*.
527 <https://doi.org/10.1109/LGRS.2018.2803259>
- 528 Gitelson, A. A., Gritz †, Y., & Merzlyak, M. N. (2003). Relationships between leaf chlorophyll
529 content and spectral reflectance and algorithms for non-destructive chlorophyll
530 assessment in higher plant leaves. *Journal of Plant Physiology*, 160(3), 271–282.
531 <https://doi.org/10.1078/0176-1617-00887>
- 532 Gitelson, A. A., & Merzlyak, M. N. (1998). Remote sensing of chlorophyll concentration in
533 higher plant leaves. *Advances in Space Research*, 22(5), 689–692.
534 [https://doi.org/10.1016/S0273-1177\(97\)01133-2](https://doi.org/10.1016/S0273-1177(97)01133-2)
- 535 Haboudane, D., Miller, J. R., Tremblay, N., Zarco-Tejada, P. J., & Dextraze, L. (2002).
536 Integrated narrow-band vegetation indices for prediction of crop chlorophyll content
537 for application to precision agriculture. *Remote Sensing of Environment*, 81(2), 416–



- 538 426. [https://doi.org/10.1016/S0034-4257\(02\)00018-4](https://doi.org/10.1016/S0034-4257(02)00018-4)
- 539 Harsch, M. A., Hulme, P. E., McGlone, M. S., & Duncan, R. P. (2009). Are treelines
- 540 advancing? A global meta-analysis of treeline response to climate warming. *Ecology*
- 541 *Letters*, 12(10), 1040–1049. <https://doi.org/10.1111/j.1461-0248.2009.01355.x>
- 542 Körner, C. (2014). *Mountain ecosystems in a changing environment*. ResearchGate.
- 543 <http://dx.doi.org/10.1553/ecomont-6-1s71>
- 544 Kuhn, M. (2008). Building Predictive Models in R Using the caret Package. *Journal of*
- 545 *Statistical Software*, 28, 1–26. <https://doi.org/10.18637/jss.v028.i05>
- 546 Lee, W. g., Fenner, M., Loughnan, A., & Lloyd, K. m. (2000). Long-term effects of defoliation:
- 547 Incomplete recovery of a New Zealand alpine tussock grass, *Chionochloa pallens*,
- 548 after 20 years. *Journal of Applied Ecology*, 37(2), 348–355.
- 549 <https://doi.org/10.1046/j.1365-2664.2000.00498.x>
- 550 Liaw, A., & Wiener, M. (2002). *Classification and Regression by randomForest*. 2.
- 551 Mashiane, K., Ramoelo, A., & Adelabu, S. (2024). Prediction of species richness and
- 552 diversity in sub-alpine grasslands using satellite remote sensing and random forest
- 553 machine-learning algorithm. *Applied Vegetation Science*, 27(2), e12778.
- 554 <https://doi.org/10.1111/avsc.12778>
- 555 Meyer, D., Dimitriadou, E., Hornik, K., Weingessel, A., Leisch, F., Chih-Chung, C. (libsvm, &
- 556 Chih-Chen, L. (2024). *e1071: Misc Functions of the Department of Statistics,*
- 557 *Probability Theory Group (Formerly: E1071), TU Wien* (Version 1.7-16) [Computer
- 558 software]. <https://cran.r-project.org/web/packages/e1071/index.html>
- 559 Moritake, K., Cabezas, M., Nhung, T. T. C., Lopez Caceres, M. L., & Diez, Y. (2024). Sub-
- 560 alpine shrub classification using UAV images: Performance of human observers vs
- 561 DL classifiers. *Ecological Informatics*, 80(102462), 1–16.
- 562 <https://doi.org/10.1016/j.ecoinf.2024.102462>
- 563 Nguyen, T.-A., Kellenberger, B., & Tuia, D. (2022). Mapping forest in the Swiss Alps treeline
- 564 ecotone with explainable deep learning. *Remote Sensing of Environment*, 281,
- 565 113217. <https://doi.org/10.1016/j.rse.2022.113217>



- 566 Nicholls, E. (2023). *Resolving the role of subarctic vegetation on mountain water cycling in a*
567 *rapidly changing climate* [Thesis].
568 <https://macsphere.mcmaster.ca/handle/11375/28472>
- 569 Nicholls, E., & Carey, S. K. (2021). Evapotranspiration and energy partitioning across a
570 forest-shrub vegetation gradient in a subarctic, alpine catchment. *Journal of*
571 *Hydrology*, 602, 126790. <https://doi.org/10.1016/j.jhydrol.2021.126790>
- 572 Ollinger, S. V. (2011). Sources of variability in canopy reflectance and the convergent
573 properties of plants. *New Phytologist*, 189(2), 375–394.
- 574 Padalia, H., Rai, I. D., Pangtey, D., Rana, K., Khuroo, A. A., Nandy, S., Singh, G., Sekar, K.
575 C., Sharma, N., Uniyal, S. K., Talukdar, G., Saran, S., Chandra, N., Bushra, A.,
576 Ahmad, R., Thakar, A., Deepak, Stanzin, J., & Rawat, G. S. (2023). Fine-scale
577 classification and mapping of subalpine-alpine vegetation and their environmental
578 correlates in the Himalayan global biodiversity hotspot. *Biodiversity and*
579 *Conservation*, 32(13), 4387–4423. <https://doi.org/10.1007/s10531-023-02702-y>
- 580 Pedregosa, F., Varoquaux, G., Gramfort, A., Michel, V., Thirion, B., Grisel, O., Blondel, M.,
581 Prettenhofer, P., Weiss, R., Dubourg, V., Vanderplas, J., Passos, A., Cournapeau,
582 D., Brucher, M., Perrot, M., & Duchesnay, É. (2011). Scikit-learn: Machine Learning
583 in Python. *J. Mach. Learn. Res.*, 12(null), 2825–2830.
- 584 Peñuelas, J., Baret, F., & Filella, I. (1995). Semi-empirical indices to assess
585 carotenoids/chlorophyll a ratio from leaf spectral reflectance. *Photosynthetica*, 31(2),
586 221–230.
- 587 Perrigo, A., Hoorn, C., & Antonelli, A. (2020). Why mountains matter for biodiversity. *Journal*
588 *of Biogeography*, 47(2), 315–325. <https://doi.org/10.1111/jbi.13731>
- 589 Perry, M. (2025). *Perrygeo/python-rasterstats* [Python]. [https://github.com/perrygeo/python-](https://github.com/perrygeo/python-rasterstats)
590 [rasterstats](https://github.com/perrygeo/python-rasterstats) (Original work published 2013)
- 591 Pix4D. (2025). *Professional photogrammetry and drone mapping software*. Pix4D.
592 <https://www.pix4d.com/>
- 593 Ramdani, F., & Furqon, M. T. (2022). The simplicity of XGBoost algorithm versus the



- 594 complexity of Random Forest, Support Vector Machine, and Neural Networks
595 algorithms in urban forest classification. *F1000Research*, 11, 1069.
- 596 Reid, K. A., Reid, D. G., & Brown, C. D. (2022). Patterns of vegetation change in Yukon:
597 Recent findings and future research in dynamic subarctic ecosystems. *Environmental*
598 *Reviews*, 30(3), 380–401. <https://doi.org/10.1139/er-2021-0110>
- 599 Robin, X., Turck, N., Hainard, A., Tiberti, N., Lisacek, F., Sanchez, J.-C., & Müller, M. (2011).
600 pROC: An open-source package for R and S+ to analyze and compare ROC curves.
601 *BMC Bioinformatics*, 12(1), 77. <https://doi.org/10.1186/1471-2105-12-77>
- 602 Rouault, E., Warmerdam, F., Schwehr, K., Kiselev, A., Butler, H., Łoskot, M., Szekeres, T.,
603 Tourigny, E., Landa, M., Miara, I., Elliston, B., Chaitanya, K., Plesea, L., Morissette,
604 D., Jolma, A., Dawson, N., Baston, D., de Stigter, C., & Miura, H. (2025). *GDAL*
605 (Version v3.10.2) [Computer software]. Zenodo.
606 <https://doi.org/10.5281/ZENODO.5884351>
- 607 Rouse, J. W., Haas, R. H., Deering, D. W., & Schell, J. A. (1973). *Monitoring the vernal*
608 *advancement and retrogradation (green wave effect) of natural vegetation* (RSC-
609 1978-2). <https://ntrs.nasa.gov/citations/19740004927>
- 610 Scherrer, D., & Körner, C. (2010). Infra-red thermometry of alpine landscapes challenges
611 climatic warming projections. *Global Change Biology*, 16(9), 2602–2613.
612 <https://doi.org/10.1111/j.1365-2486.2009.02122.x>
- 613 Schmidt, K., & Skidmore, A. (2003). Spectral discrimination of vegetation types in a coastal
614 wetland. *Remote Sensing of Environment*, 85(1), 92–108.
- 615 SPH Engineering. (2025). *UgCS - Drone flight planning software*.
616 <https://www.sphengineering.com/flight-planning/ugcs>
- 617 Thomas, A., Meudt, H. M., Larcombe, M. J., Igea, J., Lee, W. G., Antonelli, A., & Tanentzap,
618 A. J. (2023). Multiple origins of mountain biodiversity in New Zealand's largest plant
619 radiation. *Journal of Biogeography*, 50(5), 947–960. <https://doi.org/10.1111/jbi.14589>
- 620 Trimble. (2025). *eCognition*. [https://geospatial.trimble.com/products/software/trimble-](https://geospatial.trimble.com/products/software/trimble-ecognition)
621 *ecognition*



- 622 Urbina, I., Grau, O., Sardans, J., Ninot, J. M., & Peñuelas, J. (2020). Encroachment of
623 shrubs into subalpine grasslands in the Pyrenees changes the plant-soil
624 stoichiometry spectrum. *Plant and Soil*, 448(1), 37–53.
625 <https://doi.org/10.1007/s11104-019-04420-3>
- 626 Van den Bossche, J., Jordahl, K., Fleischmann, M., Richards, M., McBride, J., Wasserman,
627 J., Garcia Badaracco, A., Snow, A. D., Ward, B., Tratner, J., Gerard, J., Perry, M.,
628 cjqf, Hjelle, G. A., Taves, M., ter Hoeven, E., Cochran, M., Bell, R., rraymondgh, ...
629 Gardiner, J. (2024). geopandas/geopandas: V1.0.1. *Zenodo*.
630 <https://doi.org/10.5281/zenodo.12625316>
- 631 van Galen, L. G., Lord, J. M., Orlovich, D. A., & Larcombe, M. J. (2023). The effect of
632 stratification and shade on germination of *Nothofagus cliffortioides* (Hook.f.) Oerst
633 seeds. *New Zealand Journal of Botany*, 61(4), 351–360.
634 <https://doi.org/10.1080/0028825X.2022.2156355>
- 635 Walsh, S. J., Brown, D. G., Geddes, C. A., Weiss, D. J., McKnight, S., Hammer, E. S., &
636 Tuttle, J. P. (2009). Chapter 2 Pattern–Process Relations in the Alpine and Subalpine
637 Environments: A Remote Sensing and GIScience Perspective. In D. R. Butler, G. P.
638 Malanson, S. J. Walsh, & D. B. Fagre (Eds.), *Developments in Earth Surface*
639 *Processes* (Vol. 12, pp. 11–34). Elsevier. [https://doi.org/10.1016/S0928-](https://doi.org/10.1016/S0928-2025(08)00202-2)
640 [2025\(08\)00202-2](https://doi.org/10.1016/S0928-2025(08)00202-2)
- 641 Ward, A., Dargusch, P., Thomas, S., Liu, Y., & Fulton, E. A. (2014). A global estimate of
642 carbon stored in the world's mountain grasslands and shrublands, and the
643 implications for climate policy. *Global Environmental Change*, 28, 14–24.
644 <https://doi.org/10.1016/j.gloenvcha.2014.05.008>