

1 **Quantifying new versus old aerosol deposition in forest canopies: throughfall mass balance with**
2 **fallout radionuclide chronometry**

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6
7 **Abstract**

8 Net throughfall (NTF) measurements of the fallout radionuclides (FRNs) ⁷Be and ²¹⁰Pb confirm that the
9 atmosphere is a strong net source of particulate matter (PM) and trace metals (TMs) to a temperate
10 forest canopy, which retains nearly 60% of total wet and dry annual atmospheric flux of the FRNs (four
11 trees, three species, two sites, $n=159$). Estimation of dry deposition using a multiple regression
12 technique and predictors of precipitation depth and duration of the antecedent dry period agrees well
13 with ecosystem mass balance, with about 25% of both ⁷Be and ²¹⁰Pb annual flux deposited by dry
14 processes, and total FRN fluxes in reasonable agreement with regional soil inventories. In contrast to the
15 FRNs, other TMs including Pb and Hg show large enrichments in throughfall which derive from processes
16 of internal ecosystem recycling such as PM resuspension, leaching from tree metabolic pathways, and
17 physicochemical weathering of non-foliar biological tissues of the tree canopy (collectively
18 'phyllosphere'). To estimate the contributions to net throughfall from these internal pathways, which
19 we term a change storage (ΔS), a new FRN canopy mass balance is derived based on the different half-
20 lives of ⁷Be and ²¹⁰Pb. Estimated ΔS for selected elements are: $SO_4=7\%$; ²¹⁰Pb = 29%; As = 42%; ⁹Be =
21 45%; Cd = 60%; Hg = 60%; Pb = 63%; Fe = 79%; Al = 79%; P = 91%. The balance of throughfall ($1-\Delta S$)
22 represents new ecosystem input. Change in storage for all elements was strongly correlated with export
23 of particulate carbon (FPOM) and dissolved organic carbon (DOC) from the canopy, indicating that
24 physicochemical transformation during residence within the canopy facilitates the release of metals
25 from storage. ΔS thus represents an emergent ecosystem property through which metal, carbon, and
26 hydrologic cycles may converge to determine the fate, reactivity, and timing of metal delivery to
27 underlying soils. The forest canopy represents a substantial reservoir of decade-aged PM and cannot be
28 assumed at steady-state with respect to ongoing atmospheric deposition, especially in the context of
29 changing atmospheric composition, e.g., declining industrial emissions of Pb and Hg.

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31 **Plain Language Summary**

32 Particulate matter (PM) in the atmosphere contains nutrients and toxins that impact the health of both
33 humans and ecosystems. Understanding how PM is deposited to land from the atmosphere is
34 challenging, however, due to its very small size and complex composition. Here we develop a new
35 method using natural radioactive elements to better measure how much PM is deposited, well as the
36 timescales over which it recirculates between the atmosphere and land.

1. Introduction

Forest vegetation covers over 30% of the earth's landmass and thus mediates the impacts of particulate matter (PM) deposition on climate, ecosystem function, and human health (Bortolazzi et al., 2021; Emerson et al., 2020; Farmer et al., 2021; Hosker and Lindberg, 1982; IPCC, 2021; Jiskra et al., 2018; Johnson and Siccama, 1983; Lindberg et al., 1982; Lovett and Lindberg, 1984; McDowell et al., 2020; Van Stan and Pypker, 2015; WHO, 2016, 2021). Despite regulating the composition of both the atmosphere and terrestrial ecosystems, however, the interaction of vegetation with PM and other atmospheric trace metals (TMs) remains poorly understood. Even a basic understanding of mechanisms that regulate PM uptake by vegetation remain elusive (Luo et al., 2019; Shahid et al., 2017; Zhou et al., 2021). As a result, in part, current estimates for PM dry deposition in global chemical transport models vary across orders of magnitude, underpredict direct observations by as much as 200%, omit fundamental processes such as resuspension, and lack verification by mass balance (Emerson et al., 2020; Farmer et al., 2021; Hicks et al., 2016; Pryor et al., 2017; Saylor et al., 2019). Moreover, long residence times of PM in recirculation at the earth surface prolong impacts of critical pollutants such as Hg and Pb, and their ultimate fate remains unclear (Resongles et al., 2021; Wang et al., 2020). These uncertainties underscore a critical need for an improved understanding of atmosphere-biosphere interactions.

Two approaches are used for measuring PM and TM deposition to forest canopies, each with characteristic advantages and limitations (e.g., Draaijers et al., 1996). Micrometeorological approaches infer rates of deposition based on concentration gradients in air, or covariance of airborne concentrations with turbulence (e.g., eddy flux; Emerson et al., 2018; Laguionie et al., 2014; Obrist et al., 2021). These methods are critical insofar as they estimate deposition over large areas with in some cases, resolution of PM by size. However, these results are dependent on assumptions regarding micrometeorological physics, do not provide information on PM composition, and ultimately lack verification by mass balance ("where did it go?"). The logistical demands of the micrometeorological methods are challenging to satisfy, as well, requiring long term deployment of instrumentation to dedicated research sites, and dependence on strong vertical concentration gradients (in the eddy gradient approach). For the eddy covariance method, the need for fast and selective detectors restricts the chemical species that can be measured (Farmer et al., 2021).

A complementary approach exploits canopy chemical mass balance by comparing precipitation collected under vegetated canopies (throughfall or TF) with paired bulk deposition measurements under open sky (openfall or OF). Positive net throughfall (NTF = TF - OF) results for most elements. The excess is conventionally attributed to some combination of washoff of dry deposition that has accumulated in the canopy during the preceding dry period and leaching/weathering from biological tissues of the canopy itself. No accommodation is made for long-term PM or TM storage in ecosystem components, however, and quantifying the contribution of dry deposition thus requires uncertain assumptions for distinguishing between these competing processes ("where did it come from?"; Lovett & Lindberg, 1984; Ulrich, 1983). Distinguishing these processes is critical for understanding ecosystem trace metal budgets, since they can represent either a novel external source via dry deposition of PM, or internal recycling via either resuspended dust or leaching of biological materials (Avila and Rodrigo, 2004; Lindberg et al., 1982).

One critical source of error in throughfall mass balance is the assumption that precipitation is a simple and efficient process of 'washoff' that fully removes dry deposition that has accumulated during dry periods preceding the rain event (Bishop et al., 2020; Gandois et al., 2010; IAEA, 2009; Lovett and Lindberg, 1984; USEPA, 2005). The canopy is thus assumed at steady-state, and no accommodation is made for accumulation of PM or TMs over time, or for any change in storage within the canopy. Support for precipitation as an efficient washoff process derives from the large enrichments of most elements

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that are observed in throughfall; experimental dry applications of PM to vegetation which show residence times on the order of just days to weeks (Chamberlain, 1970; Miller and Hoffman, 1983), and washing of natural vegetation which shows exponential declines in rates of metal extraction (Avila and Rodrigo, 2004; Shanley, 1989). This conceptualization of washoff also aligns with the 'Lotus effect', which is the efficient self-cleaning of leaves by raindrops due to their hydrophobic waxy cuticle (Barthlott and Neinhuis, 1997).

On the other hand, it is well documented that a range of atmospheric metals and metalloids accumulate continuously in vegetation during exposure to the atmosphere. These include the natural fallout radionuclides (FRNs) ^7Be and ^{210}Pb (Landis et al., 2014; Sumerling, 1984), anthropogenic fission products including ^{90}Sr (Yamagata et al., 1969), ^{131}I (Yamagata, 1963), ^{137}Cs and others (Kato et al., 2012; Russell et al., 1981), TMs including Al, Pb, and Hg (Landis, 2024a; Rea et al., 2001, 2002; Wyttenbach and Tobler, 1988), and oxoanions AsO_3 , SeO_3 , and CrO_4 (Landre et al., 2009). Moreover, this accumulation continues unabated in senescent vegetation, indicating that PM and TM uptake occurs independent of leaf physiology and is thus unambiguously atmospheric in origin. The foliar metal residence times of Hg, ^7Be , ^{210}Pb , and ^{137}Cs have been shown by various means to exceed 700 days, which is generally longer than the lifetime of the foliage itself. This demonstrates strong coupling of atmospheric metals to the carbon cycle (Graydon et al., 2009; Kato et al., 2012; Landis et al., 2014).

In the case of Hg, although its occurrence in the atmosphere is regulated by gaseous elemental mercury (GEM) rather than by PM, recent measurements nonetheless show that long-lived non-foliar materials of the forest canopy including bark, lichen, moss, and foliage (collectively "phyllosphere") collectively retain many times the annual Hg flux on an areal basis (Wang et al., 2020). This shows that GEM becomes particle-bound during its residence in the phyllosphere, and also that surfaces other than foliage therefore play a dominant role in canopy and ecosystem mass balances of Hg and other TMs (Obrist et al., 2021). Stable isotopes of both Hg and Pb reveal that legacy anthropogenic emissions continue to dominate contemporary atmospheric deposition, for example, despite large reductions in both emissions and measured rates of deposition (Farmer et al., 2010; Resongles et al., 2021; Taylor et al., 2022; Yang and Appleby, 2016). These legacy TM sources remain enigmatic but might include resuspended dust from surrounding soils and roads, or release from storage in living and dead surfaces of the phyllosphere. Resolving multiple processes that regulate accumulation, storage, and re-suspension of diverse TMs in terrestrial ecosystems, and their contributions to depositional budgets, remains a challenge to biogeochemical research. This is especially important in the context of changing global emissions of, e.g., Hg and Pb, since the forest cannot be assumed at steady-state with respect to contemporary deposition.

The natural fallout radionuclide (FRN) tracers ^7Be and ^{210}Pb provide a novel and emerging perspective on PM metal cycling. The FRNs have been shown to accumulate in vegetation primarily by wet deposition rather than dry as widely presumed (Landis, 2024a; Landis et al., 2014). This distinction is important because approximately 90% of PM removal from the atmosphere occurs by wet processes (in continental climates), and the accumulation of FRNs in forest canopies might thus reflect a quantitative view of PM cycling from the atmosphere. The FRNs ^7Be and ^{210}Pb are secondary aerosols produced in the atmosphere from gaseous precursors by cosmogenic spallation of N_2 and O_2 or radiogenic decay of ^{222}Rn , respectively, with identical PM activity-size distributions (mean aerodynamic diameters) of ca. 0.5 μm (Gründel and Porstendörfer, 2004; Winkler et al., 1998). FRNs are widely exploited as tracers of other PM components of broad interest including PM_{2.5}, black carbon, sulfate, and anthropogenic trace metals in long-range transport (Koch et al., 1996; Lamborg et al., 2013; Landis et al., 2021a; Liu et al., 2016).

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The FRNs hold special power as PM tracers through two consequences of their radioactive decay. First, their accumulation in the terrestrial environment is limited such that their occurrence can be unambiguously traced directly to an atmospheric origin. Second, their characteristic rates of decay can be exploited to explicitly measure timescales of their deposition and redistribution. For example, ^7Be and ^{210}Pb are produced in the atmosphere with a typical activity ratio on the order of 10. In vegetation exposed to ongoing atmospheric deposition, both FRNs accumulate with time but the $^7\text{Be}:^{210}\text{Pb}$ ratio decreases due to their differing half-lives. After one year of exposure foliage has ratios of 2-3. With its short half-life of 54 days, ^7Be records the PM-vegetation interaction over event-based timescales, whereas ^{210}Pb records longer-term storage, resuspension, and recycling of PM deposition over subannual to decadal timescales (Landis et al., 2021a; Landis et al., 2021b).

Here we use FRN chronometry in a throughfall mass balance approach to resolve processes and timescales that govern the interactions of PM and atmospheric TMs with forest canopies. We first review contemporary approaches to throughfall mass balance and then derive a new expression using FRNs to distinguish the initial absorption of PM during deposition from later release from canopy storage. We confirm that ^7Be and ^{210}Pb are not systematically fractionated and that the $^7\text{Be}:^{210}\text{Pb}$ ratio in throughfall is primarily a measure of PM age. We then extend our throughfall mass balance approach to link the FRNs and other major and trace elements (MTEs) to canopy export of dissolved organic carbon (DOC) and fine particulate organic matter (FPOM).

1.1. Throughfall mass balance

The conventional throughfall mass balance includes two pathways by which PM enters and leaves the canopy. Inputs to the canopy by bulk wet deposition (W) or dry deposition (D) must be balanced by exports via wet throughfall (T) and any exchange with the canopy (C) via leaching (+) or absorption (-):

$$W + D = T \pm C \quad \text{Eq. 1}$$

Bulk deposition (W) is measured under open sky adjacent to the forest where it may also be called openfall (OF). Net throughfall (NTF) is the arithmetic difference between deposition under the canopy and open sky:

$$NTF = T - W = D \pm C \quad \text{Eq. 2}$$

W and T are easily measured. The challenge in estimating D by mass balance then remains to distinguish contributions of C . Two approaches have been proposed. The more common is the 'filtering' approach (Staelens et al., 2008; Ulrich, 1983; Yoshida and Ichikuni, 1989) where an index element is used to apportion dry deposition of the target element, D_i . The index element, typically Na or Al, is assumed to have identical rates of dry deposition as the target element (Eq. 3) but no metabolic contribution from the canopy. This assumption implicitly presumes similar PM size distributions, and thus sources with respect to secondary or primary emission, insofar as PM size controls rates of dry deposition (Jaenicke, 1980).

$$\frac{D_i}{W_i} = \frac{D_{Na}}{W_{Na}} \quad \text{Eq. 3}$$

With $C_{Na}=0$, dry deposition of an element of interest can be calculated as follows:

$$D_i = W_i \cdot \frac{T_{Na} - W_{Na}}{W_{Na}} = W_i \cdot (EF_{Na} - 1) \quad \text{Eq. 4}$$

Here, EF_{Na} represents the Na enrichment factor $=T/W$. Following the calculation of D_i by Eq. 4, C_i can be solved from Eq. 2. By this approach any EF_i greater than EF_{Na} yields a positive C_i , which represents

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leaching from the canopy. Any EF_i less than EF_{No} yields negative C_i , which represents a net loss to the canopy by absorption.

A second approach to describing the canopy interaction uses multiple regression of individual storm events to estimate rate constants for specific canopy processes (Lovett and Lindberg, 1984; Rea et al., 2001; Wu et al., 1992):

$$NTF = T - W = D + C = \beta_1 A + \beta_2 P \quad \text{Eq. 5}$$

The coefficient β_1 represents the rate of dry deposition to the canopy [$\text{Bq m}^{-2} \text{d}^{-1}$ or $\mu\text{g m}^{-2} \text{d}^{-1}$] and A is the duration of the antecedent dry period preceding the storm event [days]. The coefficient β_2 represents either canopy leaching or absorption [Bq or $\mu\text{g m}^{-2} \text{cm}^{-1}$] with P the precipitation total [cm]. Biotic leaching and abiotic exchange (desorption, dissolution) can be distinguished only by *a priori* assumptions based on elemental chemistry, and that one must dominate the other, for example that base cations K, Ca, Mg, etc., are strongly cycled through leaf tissue and are thus derived primarily from leaching, whereas trace metals, e.g. Pb, are not appreciably cycled and are thus dominated by canopy absorption (Lovett and Lindberg, 1984).

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An important advantage to this approach is that, provided sufficient measurements, additional explanators might be added to the multiple regression to understand effects from other environmental factors such as seasonality, tree species, precipitation pH, dissolved organic carbon (DOC), or fine particulate organic matter (FPOM).

1.2. FRN mass balance: change in canopy storage

The observed accumulation of FRNs and TMs in foliage and long-lived tissues of the phyllosphere requires that canopy exchange must accommodate a change in storage, if previously deposited PM is later susceptible to weathering processes and export from the canopy in subsequent rainfall over some characteristic timeframe. We can distinguish this change in storage as a separate term in mass balance as follows based on different characteristic timescales:

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$$T = W + D \pm (C + \Delta S) \quad \text{Eq. 6}$$

Analogous to the use of an index element for dry deposition in the filtering approach, we propose to use ^7Be as an index of canopy storage. Beryllium is highly reactive to natural surfaces but due to its short half-life, ^7Be cannot record long-term storage. In contrast, the longer-lived ^{210}Pb persists to record PM fate over annual to decadal timescales. The similar occurrence and behaviors of ^7Be and ^{210}Pb , but different half-lives, thus allow us to distinguish instantaneous canopy exchange (absorption or leaching) from a subsequent change in storage. Storage is thus defined as occurring over timescales that are long relative to the ^7Be half-life (54 days). We assume that the ^7Be : ^{210}Pb flux ratio is constant through time, but acknowledge that short-term variability and long-term decadal trends in ^{210}Pb deposition may require future attention (Landis et al., 2021a; Winkler and Rosner, 2000).

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We now write two equations to solve for one unknown, ΔS :

$$T_{Be} = W_{Be} + D_{Be} \pm C_{Be} \quad \text{Eq. 7}$$

$$T_{Pb} = W_{Pb} + D_{Pb} \pm (C_{Pb} + \Delta S) \quad \text{Eq. 8}$$

Eq. 7 ignores radioactive decay, with an assumption that the residence time of dry deposition in the canopy is very short (days) relative to the ^7Be half-life. This assumption is re-evaluated in *Sect. 3.4*. Based on their congruent interactions with natural vegetation and particulate matter (Landis, 2023; Landis et

al., 2014, 2016, 2021b, 2024a), we next assume that ^7Be and ^{210}Pb have similar canopy interactions and that for both C is an equivalent fraction of total wet+dry deposition (Eq. 9).

$$\frac{C_{Be}}{W_{Be}EF_{Na}} = \frac{C_{Pb}}{W_{Pb}EF_{Na}} \quad \text{Eq. 9}$$

This allows us to combine Eq. 7 and Eq. 8, with both D and C terms cancelling. We then solve for ΔS as follows, by which we attribute any throughfall excess in ^{210}Pb relative to ^7Be to a change in storage:

$$\Delta S = T_{Pb} - T_{Be} \cdot \frac{W_{Pb}}{W_{Be}} \quad \text{Eq. 10}$$

For ~~species~~ with a gaseous phase such as Hg or SO_4 , we cannot assume a constant D/W ratio as in Eq. 3. In these cases, we use an alternative mass balance by directly estimating dry deposition via multiple regression (δ_1) for both ^7Be and, e.g., Hg, as follows:

$$\Delta S_{Hg} = Hg_{NTF*} - Be_{NTF*} \cdot \frac{W_{Hg} + D_{Hg}}{W_{Be} + D_{Be}} \quad \text{Eq. 11}$$

The term Hg_{NTF*} represents NTF without contribution from C , i.e., $Hg_{NTF*} = T_{Hg} - W_{Hg} \cdot D_{Hg}$.

Finally, calculating a change in storage acknowledges that a process of exchange occurs within the canopy. Some fraction of new wet and dry deposition is retained by the canopy, while some fraction of previous deposition is released from storage. The proposed mass balance thus provides a new perspective on ~~PM and TM~~ deposition, new versus old. ΔS quantifies the old, and the fraction new follows from mass balance:

$$Hg_{new} = T_{Hg} - \Delta S_{Hg} \quad \text{Eq. 12}$$

2. Methods

2.1. Site and tree characteristics

We measured 156 samples of throughfall under mature canopies of two tree species at each of two sites. The Beaver Meadow (BM) site in Sharon, Vermont (104 km², population =1500) sits at an elevation of 502 m in mixed forest of sugar maple, red oak, American beech, white pine, and poplar, with understory of striped maple and hornbeam. Annual precipitation averages 117 cm y⁻¹ for the past 30 years (PRISM 2014). The site is underlain by green schists of the Waits River series and thin glacial till which leads to development of well-drained soils typically classified as Inceptisol. Throughfall was collected under red oak (*Quercus rubra*; diameter at breast height, dbh =28 cm) and white pine (*Pinus strobus*; dbh = 59 cm) in the forest interior, and openfall was collected within 50 m in an adjacent forest gap. The Shattuck Observatory (SO) site is located 18 km distant in Hanover, New Hampshire (130 km², population =8500), in the semi-urban forest of the Dartmouth College Park at an elevation of 165 m. Annual precipitation averages 101 cm y⁻¹ (PRISM 2014). The SO site is underlain by mafic volcanics of the Ammonoosuc series with frequent outcrops along this ridgeline site (Schumacher, 1988). Throughfall was collected under red oak (dbh = 53 cm) and Norway spruce (*Picea abies*; dbh = 60 cm) at the forest edge, with openfall collected within 50 m in the adjacent clearing where we have previously reported a long-term timeseries of atmospheric deposition (Landis et al., 2021a; Landis et al., 2021b).

The sampled trees represent a range of foliar ages and characteristics that may influence FRN activities and MTE concentrations. Red oak is deciduous and sheds most leaves annually after ca. 6 months of growth (May-October); however, some remain on branches following abscission through the following winter (a phenomenon called marcescence). White pine retains needles for 1.5-2 years. Norway spruce retains needles for up to 5-8 years (Reich et al., 1996). Leaf area index (LAI) at each throughfall site was

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289 estimated over a 30° zenith angle using hemispheric photos and the software package Hemisfer
 290 (Schleppi et al., 2007; Thimonier et al., 2010). Measured LAI are as follows: BM oak (leaf off) =0.9 ±0.3,
 291 BM oak (leaf on) =3.1 ±0.5, BM pine =2.79 ±0.14, SO oak (leaf off) 1.13 ±0.4, SO oak (leaf on) =5.9 ±0.6,
 292 SO spruce =4.52 ±0.16.

293 2.2 Sample collection and filtration

294 Each openfall and throughfall precipitation sampler consisted of 4 tandem polyethylene open collectors
 295 (area =650 cm² each, depth=40 cm; cumulative projected area = 0.26 m²) mounted to a PVC post at a
 296 height of 1.5 m above ground. Total collector surface area to projected ground area (surface area index)
 297 =5.2 (e.g., Hicks, 1986). At each site samplers were located at the same aspect relative to the sampled
 298 tree stem, under the midpoint of the tree crown to avoid dripline or stem effects. Collectors were
 299 deployed prior to, and retrieved following, individual rain and snowstorms. Upon retrieval an aliquot
 300 was immediately removed for measurement of operationally-dissolved **major and trace elements**
 301 (MTEs) and DOC. This aliquot (*d*) was filtered with a 0.45 µm Nylon filter (Fisher Scientific), the first 1
 302 mL discarded, **the** next 5 mL used to rinse an acid-washed polyethylene vial and discarded, and a final 20
 303 mL retained and acidified to 2% HCl (Optima grade). Sample pH was then measured in the bulk sample
 304 with dual ROSS-type electrodes (Thermo Scientific), calibrated daily with fresh, low ionic strength buffer.
 305 Bulk samples were then passed through a 1 mm polypropylene screen or hand-picked to remove coarse
 306 debris and then filtered to tared 0.5 µm quartz fiber filters (QFF; Advantech QR-100) to remove an
 307 insoluble fine particulate organic fraction (FPOM). The filtrate was acidified to 2% HCl with concentrated
 308 acid (Trace Metal grade), and the acidified sample in turn was used to rinse and recover metals
 309 absorbed to the sample collection train and collector surfaces.

310 FPOM on QFF filters was extracted sequentially using 2% HCl and reverse aqua regia (9:3 HNO₃:HCl, plus
 311 0.2 mL BrCl) to estimate MTE solubility and **to** distinguish fluxes of dissolved (*d*.Pb, etc.), soluble, and
 312 total **TM** deposition. The acid-soluble fraction (*s*.Pb, etc.) is the sum of dissolved and 2% HCl extraction.
 313 The total fraction (*t*.Pb, etc.) is the sum of dissolved, 2% HCl, and aqua regia fractions. The acid soluble
 314 fraction is considered environmentally relevant (Lindberg and Harriss, 1981; Mahowald et al., 2018).

315 From dissolved and FPOM fractions we calculate **distribution coefficients (*K_D*)** as follows, where *V* is total
 316 sample volume (mL) and *M* is FPOM mass (g):

$$317 \quad K_D = \frac{d.Pb/V}{(t.Pb-d.Pb)/M} \quad \text{Eq. 13}$$

318 Enrichment factors of MTEs with respect to upper continental crust (*EF_{crustal}*) were calculated from
 319 dissolved fractions in OF collections to maintain comparability the prior works (Duce et al., 1975;
 320 Gandois et al., 2010; Taylor and McLennan, 1995):

$$321 \quad EF_{crust} = (M_i/Al)_{OF} / (M_i/Al)_{crust} \quad \text{Eq. 14}$$

322 2.3. Sample Analysis

323 FRNs were measured by gamma spectrometry following preconcentration by MnO₂ co-precipitation,
 324 with yields averaging 88 ±15% for Be and 88 ±16% for Pb (mean ±SD). **A workflow schematic is**
 325 **illustrated in Fig. S1, and details are given elsewhere** (Landis et al., 2012, 2021). MTEs were measured by
 326 inductively-coupled plasma optical emission spectrometry (ICPOES; Spectro ARCOS) operated in axial
 327 view. Low-level trace elements including As, Cd, Co, Cr, Cu, Fe, Pb, and V were measured by ICP triple-
 328 quadrupole mass spectrometry (ICPMS-QQQ; Agilent 8900) in the Dartmouth Trace Element Analysis
 329 Core Facility (TEA Core). Reference material NIST1643f was used for quality control, with recoveries
 330 within 95-105% for all analytes. Total Hg was measured by double amalgamation purge-and-trap atomic

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336 fluorescence (Brooks Rand Merckx-T, Model III), where samples preserved in 2% HCl were oxidized by
337 reaction with 1% v/v BrCl at room temperature for 24 hours and then reduced to Hg₀ by addition of
338 SnCl₂ and hydroxylamine immediately prior to analysis. Reference material NIST1641e was used for Hg
339 quality control, with recovery of 10 pg averaging 97 ±8% (mean ±SD; *n*=48). Sample recovery spikes
340 averaged 97 ±6% (*n*=9). Dissolved organic carbon (DOC) was measured by automated combustion of
341 acidified samples (General Electric); see also Hou et al. (2005) and Gandois et al. (2010).

342 **2.4. Statistical analyses**

343 Statistical analyses were performed in JMP Pro 16.0. Data were transformed to achieve normal
344 distributions and equal variances. For NTF which includes positive and negative values, this required
345 transform of the form $\log_{10}(x+n)$ with *n* selected to scale all values >0. Our multiple regression approach
346 is described in more detail elsewhere (Landis et al., 2021a). We observe Bonferroni correction for
347 describing significance of multiple regression parameters but presented *p*-values do not include this
348 correction since the number of explanators are shown for each regression.

349 **2.5. Literature review**

350 Throughfall enrichment factors were compiled from the following literature sources (Ali et al., 2011;
351 Atteia and Dambrine, 1993; Avila and Rodrigo, 2004; Böhlke and Michel, 2009; Bringmark et al., 2013;
352 Demers et al., 2007; Eissalou et al., 2013; Forti et al., 2005; Fu et al., 2010; Gandois et al., 2010; Graydon
353 et al., 2008; Grigal et al., 2000; Henderson et al., 1977; Hou et al., 2005; Huang et al., 2011; Iverfeldt,
354 1991; Karwan et al., 2016, 2018; Kolka et al., 1999; Kopáček et al., 2009; Landre et al., 2009; Larssen et
355 al., 2008; Lawson and Mason, 2001; Lindberg and Harriss, 1981; St. Louis et al., 2001, 2019;
356 Mahendrappa, 1987; Malek and Astel, 2008; Matschullat et al., 2000; Michopoulos et al., 2005, 2018;
357 Oziegbe et al., 2011; Petty and Lindberg, 1990; Rea et al., 2001; Rehmus et al., 2017; Rodrigo et al.,
358 1999; Schwesig and Matzner, 2000; Skrivan et al., 1995; Sohrt et al., 2019; Stachurski and Zimka, 2000;
359 Tan et al., 2019; Ukonmaanaho et al., 2001; Wang et al., 2020; Wilcke et al., 2017; Zhang and Liang,
360 2012).

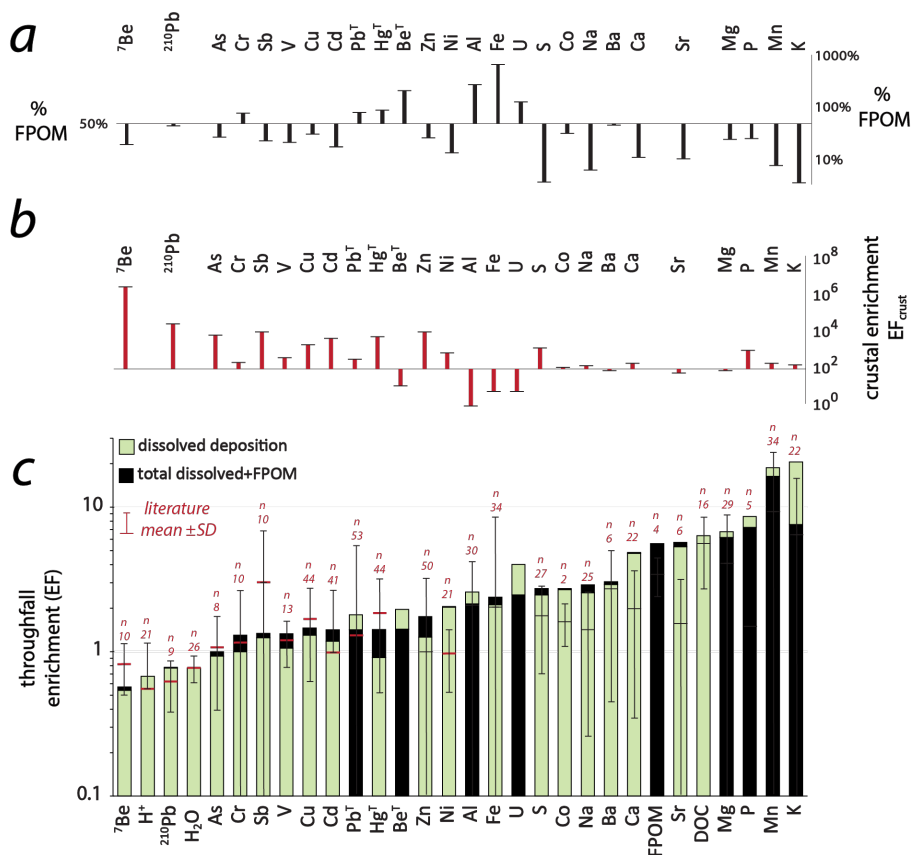


Figure 1. throughfall enrichment factors for FRNs and MTEs. (a) % increase in throughfall flux when including a fraction extracted from fine particulate organic matter (FPOM); (b) crustal enrichment factors, where $\text{EF}_{\text{crust}} = [\text{M}/\text{Al}]_{\text{OF}}/[\text{M}/\text{Al}]_{\text{crust}}$. (c) Median flux-based throughfall enrichment factors (EFs) for this study shown in bars. Note log₁₀ scaling. Line and whiskers show means and standard deviations compiled from literature sources, with number of experiments indicated as *n* (see Methods).

3. Results

3.1. FRN mass balance in the forest canopy

Throughfall enrichment factors and the partitioning of FRNs and MTEs to FPOM are shown in Fig. 1. Throughfall measurements including rainfall interception, FPOM, and DOC concentrations are shown in Fig. S2. The studied tree canopies typically intercept (retain) 20% of precipitation. DOC and FPOM are greatly increased through the canopy with TF concentrations in ranges of 1-10 $\mu\text{g mL}^{-1}$ and 10-100 μg

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376 mL⁻¹, respectively. Throughfall mean pH values for the seasons summer through winter were 5.29, 5.36,
377 4.98, and 4.86, respectively. Effect of the canopy on throughfall pH was variable (Fig. S3). The canopy
378 was a net sink for H⁺ in summer for each of spruce, pine, and oak [*p*<0.05]. On an annual basis, oak was
379 a net sink at both sites but both pine and spruce were net sources of H⁺.

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380 Flux-based throughfall enrichment factors (EFs) for ⁷Be and ²¹⁰Pb were 54.7 ±0.3% and 75.9 ±1%,
381 respectively (±SE), significantly lower than 1 and thereby demonstrating that the canopy is an
382 unambiguous net sink for FRNs in wet deposition. ⁷Be shows stronger absorption than either H⁺ or H₂O,
383 which demonstrates its strong preferential absorption by the canopy. Only the following showed EFs
384 less than 1: ⁷Be < H⁺ < ²¹⁰Pb < H₂O < As. All other MTEs showed a net gain through the canopy due to
385 some combination of dry deposition during the antecedent dry period, leaching from biological tissues,
386 or weathering from storage in the phyllosphere. Putative reference elements expected to have minor
387 leaching contributions such as Al, Fe, U and Na showed moderate gains from the canopy, with EFs in the
388 range of 2-3. The following trace metals showed EFs less than these reference elements, which
389 nominally indicates net sorption to the canopy: Cr, Sb, V, Cu, Cd, Pb, Hg, ⁹Be, Ni and Zn. Alkaline and
390 nutrient elements showed much larger EFs than reference elements, indicating strong net gains from
391 the canopy: Co, Ba, S, Ca, Sr, Mg, P, C, Mn and K, as well as DOC and FPOM.

392 Crustal enrichment factors (EF_{crust}) in the range of 10²-10³ indicate that **TF** metals are generally of
393 anthropogenic atmospheric origin, whereas values ~1 confirm a terrestrial dust source for Al, Fe, and U
394 (Duce et al., 1975; Fig. 1b). EF_{crust} does not strictly identify atmospheric PM however, since values in the
395 range of 10² for alkaline elements can also reflect the high solubility of these elements and their
396 extraction from biological tissues of the phyllosphere during rainfall.

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397 Among tree species, spruce displayed lowest EFs for ⁷Be and ²¹⁰Pb (highest interception), but higher EFs
398 than either oak or pine for most elements [*p*<0.05]. By site, both spruce and oak at the SO site showed
399 higher EFs for ⁹Be, Ni, Sb, V and Zn, demonstrating a site effect for these elements which could be
400 related to higher pollutant loads near a higher population density, or alternatively to underlying mafic
401 lithology at this site [*p*<0.05]. By contrast, there were no species or site effects for the heavily cycled
402 elements Ca, Mg, K, P or Mn, or for metals including Cd, Cu, Hg, Pb, or U [*p*>0.05]. Tree species were
403 compared only during deciduous leaf-on months.

404 3.2. Interpreting canopy processes from FRN throughfall concentrations

405 Throughfall EFs evaluated on an annual basis described above show that the FRNs **absorb** strongly and
406 that the canopy is therefore a strong sink for secondary aerosol metals. For further insights into the
407 balance of processes acting to retain and release FRNs within the canopy at the event-scale, we also
408 evaluated the concentrations of FRNs as a function of precipitation depth for individual storms. Below 2
409 cm of precipitation both ⁷Be and ²¹⁰Pb concentrations in TF are generally greater than 100% of
410 concentrations in OF, and in some cases more than 300% higher than incident precipitation. This
411 indicates that net release of FRNs is the dominant process at low precipitation totals, which we attribute
412 to the removal of dry-deposited **PM** which **was** deposited during the antecedent dry period (Fig. 2a).

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413 Beyond 2 cm of precipitation the concentrations of both ⁷Be and ²¹⁰Pb are lower than openfall,
414 indicating that the canopy transitions to a net sink with increasing precipitation depth. This suggests that
415 at some threshold amount of precipitation, susceptible dry-deposited **PM** is fully removed, and wet
416 deposited **PM** continue to be absorbed by the canopy. The unambiguous net sorption of wet-deposited
417 **PM** is clear when concentrations are converted to fluxes (by multiplication with precipitation depth; Fig.
418 2b). At highest precipitation totals >6 cm there is some suggestion that sorption capacity of the canopy

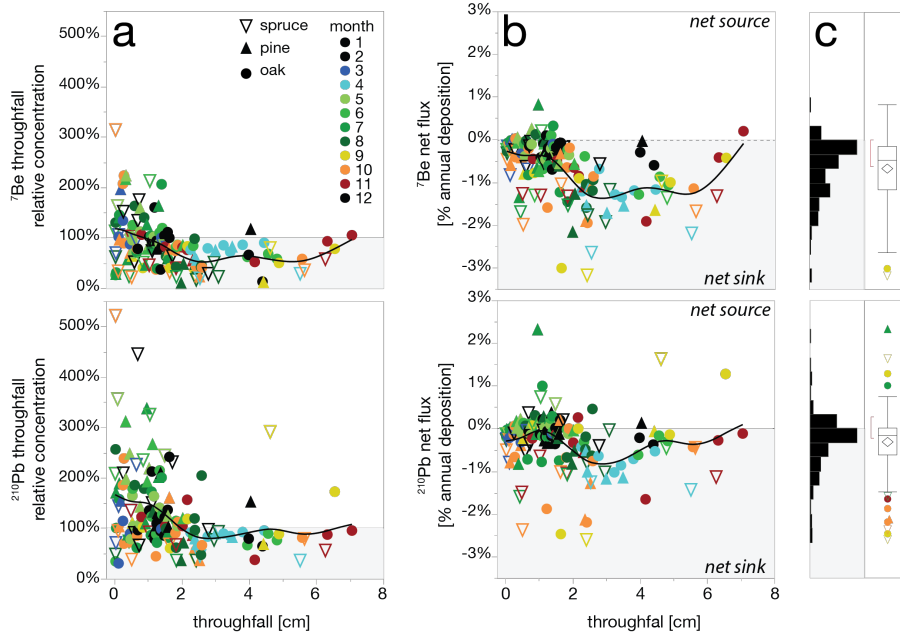
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426 may be exhausted and FRN yields from the canopy begin to increase, but data are too few for a
 427 confident interpretation.

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429 **Figure 2.** relative concentrations of ^7Be or ^{210}Pb in throughfall versus openfall ($\text{EF}=\text{TF}/\text{OF}$).
 430 (a) Oak is shown with circles, pine with triangles, spruce with open triangles. Black lines show a spline best-fit.
 431 (b) net throughfall fluxes (NTF) of ^7Be and ^{210}Pb as a percentage of total annual deposition. (c) histograms of
 432 ^7Be and ^{210}Pb net flux.

433 3.3. Throughfall $^7\text{Be}:^{210}\text{Pb}$ ratios as a metric of PM age

434 ^{210}Pb is systematically biased to higher yields than ^7Be despite the same patterns in canopy interactions,
 435 and as a result $^7\text{Be}:^{210}\text{Pb}$ ratios are significantly lower in throughfall versus openfall (Fig. 2a, Fig. S3).
 436 Mean throughfall $^7\text{Be}:^{210}\text{Pb}$ ratios across all species averaged 10.3 ± 0.3 versus mean openfall of 14.8 ± 0.5
 437 [mean $\pm$ SE]. Throughfall for each tree species was significantly lower than paired openfall
 438 measurements [$p < 0.0001$]. There was no correlation between throughfall pH and the $^7\text{Be}:^{210}\text{Pb}$ ratio that
 439 might indicate discrimination of the FRNs through aqueous chemistry [$R^2 = 0.001$, $p = 0.80$; Fig. S3b]. We
 440 next used multiple regression to identify independent explanators on throughfall $^7\text{Be}:^{210}\text{Pb}$ ratios and to
 441 quantify their independent effects on $^7\text{Be}:^{210}\text{Pb}$; we report independent effects as the % of total variance
 442 explained by each variable [in brackets] (Fig. S4). A predictive model for $^7\text{Be}:^{210}\text{Pb}$ showed these
 443 significant explanators: species [22%], season [12%], and $^7\text{Be}:^{210}\text{Pb}$ ratio of incident precipitation [9%]
 444 [$R^2 = 0.42$]. Adjusted for season and incident ratio, $^7\text{Be}:^{210}\text{Pb}$ ratios among species increased in order:
 445 spruce (7.9 ± 0.6)^A < pine < (9.5 ± 0.6)^B oak (11.9 ± 0.5)^C; values not sharing the same letter are
 446 significantly different [$p < 0.05$]. Adjusted for both species and incident ratio, seasonal ratios increased in
 447 order: autumn (7.6)^A < summer (9.9)^B < winter (10.5)^{BC} < spring (11.3)^C. Both species and seasonal

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453 patterns are consistent with the impact of tree phenology on long-term storage of ^{210}Pb , with the higher
454 LAI and older foliage of conifers storing larger reservoirs of ^{210}Pb , which is exported in small amounts
455 during subsequent storms.

456 The strongest discrimination between ^7Be and ^{210}Pb appears to be their characteristic half-lives, in
457 absence of any evidence for impacts by environmental factors. There was no effect on $^7\text{Be}:$ ^{210}Pb from
458 antecedent dry period [$R^2=0.01$, $p=0.27$], suggesting no discrimination in dry deposition between ^7Be
459 and ^{210}Pb . There was no positive correlation between rainfall depth and $^7\text{Be}:$ ^{210}Pb [$R^2=0.03$, $p=0.42$], in
460 contrast to OF where a strong positive correlation reflects the washout-rainout transition (Landis et al.,
461 2021). Importantly, there were no significant effects from factors that might indicate discrimination
462 between Be and Pb aqueous chemistries including: pH [$p=0.32$], $\log_{10}(K_D)$ for either ^7Be [$p=0.34$] or ^{210}Pb
463 [$p=0.18$], FPOM [$p=0.29$], DOC [$p=0.47$], or any MTE concentrations, fluxes, net fluxes, or enrichment
464 factors [$p>0.05$].

465 3.4. Multiple-regression mass balance to quantify dry deposition

466 We used multiple-regression mass balance to quantify the independent influences of wet and dry
467 deposition and canopy interactions that determine net throughfall (NTF) of the FRNs (Eq. 5). Overall
468 model fits for NTF are good for both ^7Be [$R^2=0.49$, $p<0.0001$] and for ^{210}Pb [$R^2=0.30$, $p<0.0001$] (Fig. 3).
469 Both ^7Be and ^{210}Pb show positive effects with antecedent dry period (θ_1) [$p<0.05$], and inverse effects
470 from precipitation depth (θ_2) apparent from Fig. 2 [$p<0.05$]. For ^7Be the dry deposition rate (θ_1) is
471 equivalent to $213 \pm 53 \text{ Bq m}^{-2}$ per year or $13 \pm 4\%$ of total annual deposition, (assuming that all dry
472 deposition is fully removed by subsequent storms). The multiple regression dry deposition estimate
473 does not incorporate ^7Be decay that occurs during the antecedent period. However, with a decay rate of
474 1.3% per day, and median antecedent period of 7 days (mode of 2 days), underestimation of ^7Be dry
475 deposition by ignoring radioactive decay is likely $<10\%$ of estimated dry deposition or $<20 \text{ Bq m}^{-2} \text{ y}^{-1}$. For
476 ^{210}Pb , θ_1 is equivalent to $19 \pm 6 \text{ Bq m}^{-2}$ per year or $15 \pm 5\%$ of total annual deposition.

477 It is important to observe from Fig. 2 that dry deposition is not fully removed from the canopy by
478 subsequent rainfall since throughfall FRN concentrations may be higher than incident precipitation for
479 up to 2 cm of precipitation. During and following rain events it is thus likely that a complex exchange is
480 occurring at the leaf surface between wet and dry deposited PM, and both components are likely to be
481 absorbed by the leaf especially at low rainfall totals. In this case the multiple regression estimate of dry
482 deposition is underestimated by a factor equal to the efficiency with which dry deposition is absorbed
483 by the canopy. Implications for dry deposition absorption are discussed below in Sect. 4.1.

484 Dry deposition to the canopy is likely derived from ambient PM and not from resuspension of dust. The
485 $^7\text{Be}:$ ^{210}Pb ratio of dry deposition to the canopy based on multiple regression coefficients $=11 \pm 4$, which is
486 comparable to our prior measurements of ambient PM10 ($=10 \pm 2$), and somewhat lower than openfall
487 wet deposition [flux-weighted average = 15.1 ± 0.6 ; $p=0.22$], but significantly higher than bulk dry
488 deposition [$=4.2 \pm 0.6$, $p<0.0001$] (Landis et al., 2021).

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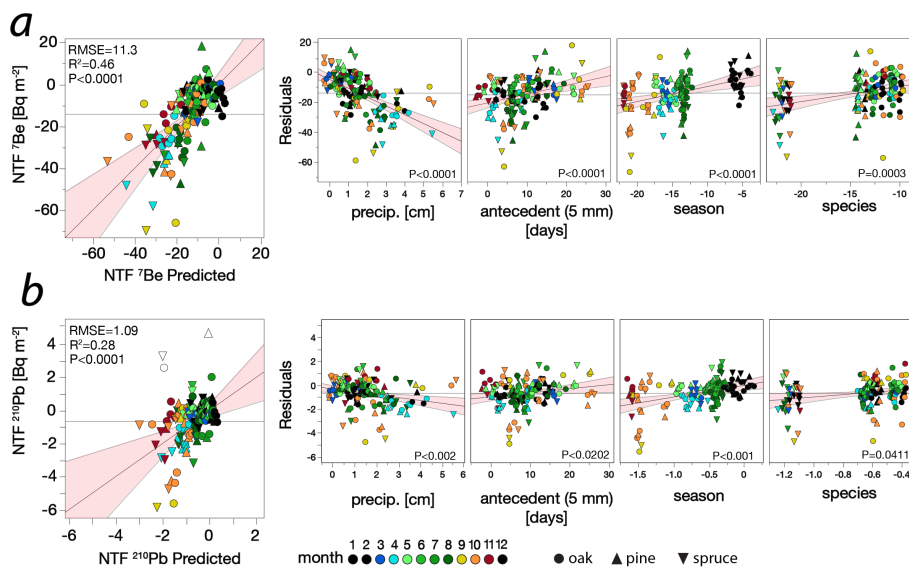


Figure 3. multiple regressions for prediction of net throughfall (NTF) of FRNs. (a) ^7Be and (b) ^{210}Pb . NTF is modeled with explainers including precipitation depth, antecedent dry period (since 5 mm precipitation), season, and tree species. Models are shown without transformation to retain interpretable units on the data but this does not impact significance of explainers. Three outliers omitted in ^{210}Pb regression are indicated with open symbols.

3.5. Importance of DOC and FPOM in MTE multiple regressions

For MTEs we similarly estimated dry deposition and canopy interactions that contribute to their enrichment in throughfall using multiple regression. All MTEs were found to have positive correlations with both antecedent dry period and rainfall depth (Table S1). Importantly, both FPOM and DOC contribute significantly to all FRN and MTE models. This suggests that carbon plays a significant role in release of both FRNs and MTEs from the canopy. DOC is the stronger explainer and typically supersedes FPOM in the model (Table S1). Because they have strong predictive power in FRN and MTE models, we also examined FPOM and DOC using multiple regression (Fig. S5). The FPOM model [$R^2=0.35$] has significant explainers including season [19%], antecedent duration [11%], and rainfall depth [5%]. Surprisingly, there was no effect from species [$p>0.05$]. The strongest effect from antecedent duration uses a 15 mm rather than 5 mm precipitation threshold. Including DOC as an explainer strengthens the FPOM model [$R^2=0.58$] as DOC becomes the strongest predictor [21%]. A model for DOC [$R^2=0.31$] shows the same explainers as for FPOM, with season [14%], antecedent [7%], and rainfall depth [6%] each significant. DOC production is higher relative to FPOM in spring and fall [$p<0.05$], which is consistent with tree phenology and peak DOC production with both new foliar growth and senescence (Van Stan et al., 2012).

3.6. Dissolved Organic Carbon regulates FRN and MTE Net Throughfall

The coefficient θ_{DOC} obtained from multiple regression mass balance quantifies the increase in NTF per increase in DOC ($\mu\text{mol-MTE } \mu\text{mol-C}^{-1}$). This provides an empirical metric of FRN or MTE affinity for

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throughfall DOC (Table S1). It can be expressed as an association constant (K_{DOC}) by normalizing each multiple regression coefficient by the average dissolved FRN or MTE concentration ($\mu\text{mol L}^{-1}$). K_{DOC} is significantly correlated with EF across all elements, which supports the notion that DOC aids in regulating export of MTEs from the canopy (Fig. 4; $R^2=0.32$, $p=0.003$; see also Hou et al., 2005 and Gandois et al., 2010). Moreover, there are additional noteworthy associations within the FRNs and MTEs. They are bound at lower EFs by group (1) consisting of ^7Be , ^{210}Pb , Be^+ , Pb^+ , Hg^+ , Al , and Fe [slope $=0.381 \pm 0.056$; $R^2=0.90$, $p=0.001$]; these are elements with strongest oxide-over-humate preferences, and thus likely to be most weakly solubilized by DOC (Takahashi et al., 1999).

Separation of FRNs from their stable isotope counterparts within this relationship, with both lower K_{DOC} and lower EF, suggests that some age-dependence of PM regulates the metal-DOC association (Discussion S2, Fig. S8). Within the ^7Be - Be^+ and ^{210}Pb - Pb^+ pairs elemental chemistry is conserved, and differences in K_{DOC} and EF must be attributed to different PM ages/sources. The distinction made here is between secondary aerosol that characterizes ^7Be and ^{210}Pb and resuspended dust or aged PM that more likely influences Be^+ , Pb^+ , (as well as Hg^+ , Al , Fe , and U). Elsewhere we have described this as a particle-age effect to describe how PM metals become increasingly and irreversibly particulate-bound with increasing PM age (Landis et al., 2021b).

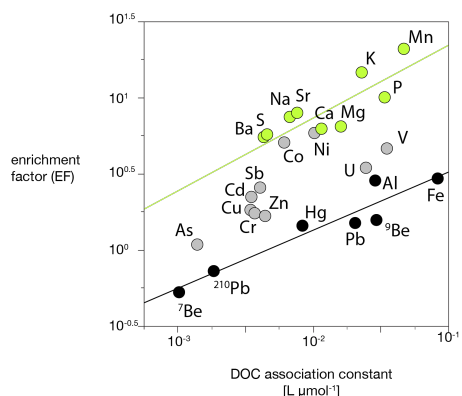


Figure 4. Dependence of throughfall EF on DOC binding constant, K_{DOC} . K_{DOC} was estimated as the increase in metal NTF per unit increase in DOC, normalized to metal concentration. Elements in black have large oxide (K_{OH}) over oxalate binding constants (K_{Ox}), elements in green have oxalate preference (Takahashi et al., 1999).

Alkaline and strongly cycled nutrient elements form a second boundary of group (2) elements at high EFs [slope $=0.478 \pm 0.133$; $R^2=0.72$, $p=0.0159$]. The alkaline elements Ca , Sr and Ba have highest preference for humate over oxide ligands (Takahashi et al., 1999). K_{DOC} alone does not clearly separate groups 1 and 2 due, in part, to the particle age effect. This suggests that EF distinguishes metal source, with atmospheric deposition controlling group 1 and leaching controlling group 2. Trace metals in intermediate group (3) are increasingly likely to have metabolic contributions from canopy leaching with larger EFs.

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3.7. PM storage, and change in storage (ΔS), in the forest canopy

Our assumptions for FRN canopy mass balance underlying the derivation of Eq. 10 appear to be observed, with no detected differences in ^7Be and ^{210}Pb geochemical behaviors or rates of dry deposition. Lower $^7\text{Be}:$ ^{210}Pb ratios and excess of ^{210}Pb over ^7Be in TF should thus be interpreted as a change in storage (ΔS) derived from within the non-metabolic materials of the [tree](#) canopy. $^{210}\text{Pb}_{\Delta S}$ averaged 0.70 Bq m^{-2} per storm or 29% of annual net throughfall flux. In contrast to net throughfall ($^{210}\text{Pb}_{\text{NTF}}$) which decreases with total precipitation, $^{210}\text{Pb}_{\Delta S}$ increases significantly with total precipitation at a rate of $0.22 \pm 0.04 \text{ Bq m}^{-2} \text{ cm}^{-1}$, which is equivalent to $22 \pm 4 \text{ Bq m}^{-2}$ per year [$R^2=0.19$, $p<0.0001$]. $^{210}\text{Pb}_{\Delta S}$ was not influenced by antecedent dry period [$r^2=0.009$, $p=0.25$], which lends confidence that $^{210}\text{Pb}_{\Delta S}$ does not simply reflect different depositional dynamics for ^7Be and ^{210}Pb . Multiple regression for $^{210}\text{Pb}_{\Delta S}$ yielded a strong model with significant, independent effects from rainfall depth [13%], species [9%], season [7%], and DOC [15%] [$R^2=0.46$, $n=137(8)$] (Fig. 5). $^{210}\text{Pb}_{\Delta S}$ was highest in autumn and summer, and lowest in winter, irrespective of tree species and including the conifer species [$p<0.05$]. Spruce contributed larger change in storage than either pine or oak [$p<0.05$]. Linear, positive dependence on p_D demonstrates that $^{210}\text{Pb}_{\Delta S}$ supply is not readily exhausted as is the case for dry deposition (i.e., as shown in Fig. 2a).

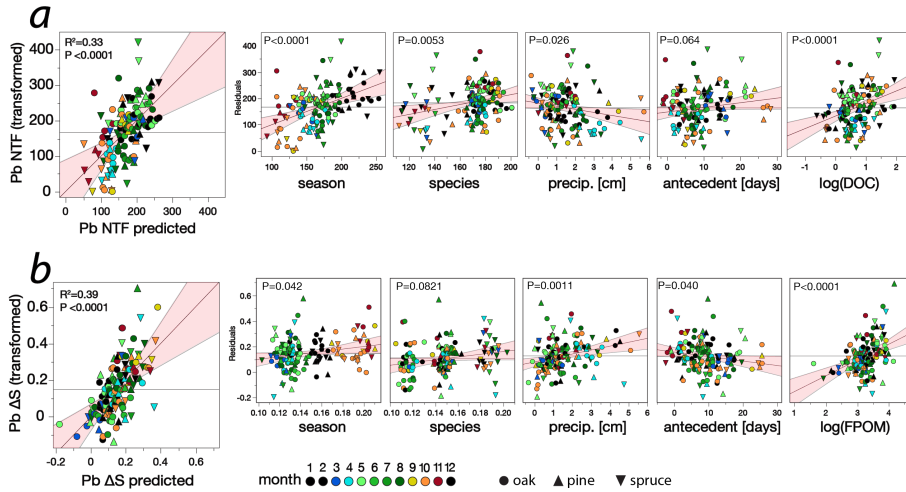


Figure 5. Multiple regressions for (a) ^{210}Pb net throughfall, (b) ^{210}Pb change in storage. Both variables are transformed to provide normal distributions.

When we compare regressions for $^{210}\text{Pb}_{\text{NTF}}$ and $^{210}\text{Pb}_{\Delta S}$, the importance of DOC is replaced by stronger correlation with FPOM in the ΔS model. In addition, where modeling of $^{210}\text{Pb}_{\text{NTF}}$ found no significant relationships with other metals, for the $^{210}\text{Pb}_{\Delta S}$ multiple regression multiple trace metals also provide explanatory power at the expense of season, DOC, and rainfall depth [$R^2=0.65$, $n=102(3)$]. MTEs with strong explanatory power were $\text{Cu}_{\Delta S}$ [26%], $\text{Fe}_{\Delta S}$ [26%], $\text{Al}_{\Delta S}$ [24%], $\text{P}_{\Delta S}$ [22%] and $\text{Hg}_{\Delta S}$ [14%]. Correlation with Hg also removes autumn seasonality in $^{210}\text{Pb}_{\Delta S}$. These strong correlations among ΔS for multiple metals suggest that they are entrained in a common weathering process from the canopy. It is

important to note that ${}^7\text{Be}_{\text{NTF}}$ provided no explanatory power in the ${}^{210}\text{Pb}_{\Delta\text{S}}$ model [$p=0.35$], confirming the decoupling of ${}^7\text{Be}$ and ${}^{210}\text{Pb}_{\Delta\text{S}}$ through the mass balance model.

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Among all MTEs, ΔS increases asymptotically with EF (Fig. S6). Sulfur is a notable exception because high EF is attributable to dry deposition of gaseous SO_2 rather than ΔS . Change in storage was estimated for MTEs as the relative contribution to annual NTF: $\text{SO}_4=7\%$; ${}^{210}\text{Pb}=17\%$; $\text{Cd}=40\%$; ${}^9\text{Be}=45\%$; $\text{As}=58\%$; $\text{Hg}=60\%$; $\text{Pb}=63\%$; $\text{Fe}=79\%$; $\text{Al}=79\%$; $\text{P}=91\%$ (Table S1).

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4. Discussion

4.1 Verifying throughfall multiple-regression mass balances

Throughfall mass balances and the derivation of ΔS can be verified against ecosystem mass balance for the FRNs since, due to their short half-lives, the FRNs measured in vegetation and soil have unambiguous atmospheric sources. For example, long-term ${}^{210}\text{Pb}$ fluxes calculated from regional steady-state soil inventories average $173 \pm 24 \text{ Bq m}^{-2} \text{ y}^{-1}$ ($n=8$; (Landis et al., 2016, 2024b)). Long-term bulk deposition measurements average 177 ± 27 ($n=12$) but have decreased 4% per year from 2011 to present and the most recent 4 years have averaged 158 ± 26 (Landis et al., 2021). The sum of event-based W plus D fluxes estimated from multiple regression reported here over two years equates to $151 \text{ Bq m}^{-2} \text{ y}^{-1}$ which is somewhat lower than expected from soil inventories. However, the multiple regression estimate of D assumes that dry deposition is fully removed by each subsequent precipitation event (Lovett and Lindberg, 1984). Our analysis of throughfall FRN concentrations in storms has already shown that this is not correct (Fig. 2). Therefore, the multiple regression approach underestimates D and thereby total annual deposition as well. From a whole tree mass balance approach (Landis, 2024b), an absorption rate of D was estimated to be 49%, meaning the multiple regression underestimates D by a factor of about 2. Adoption of this absorption factor hereafter requires a proportionate increase in annual dry deposition, almost doubling the estimate to $65 \text{ Bq m}^{-2} \text{ y}^{-1}$ and leading to an annual deposition total of $174 \text{ Bq m}^{-2} \text{ y}^{-1}$. This more robust quantitation of dry deposition stands in better agreement with soil records. The overall conclusion we reach is that both wet and dry deposition are well retained by the forest canopy.

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FRNs also provide an opportunity to evaluate the suitability of conventional filtering throughfall mass balance for MTEs. In contrast to the multiple-regression approach, the rate of ${}^{210}\text{Pb}$ dry deposition predicted by the filtering approach (Eq. 4, with Al as reference) is higher by a factor of twelve, and total annual ${}^{210}\text{Pb}$ deposition is higher by a factor of two over what is observed in local soils ($360 \text{ Bq m}^{-2} \text{ y}^{-1}$). The total ${}^7\text{Be}$ flux predicted by filtering is higher than observed by a factor of 2.6 (4700 Bq m^{-2}). The filtering approach thus grossly overestimates secondary aerosol metal deposition, irrespective of whether Na, Al or U are used as reference elements. These elements must therefore have substantial contributions from canopy leaching or resuspended dust that do not impact FRNs and other secondary aerosol metals. Total rates of deposition for all MTEs estimated by the filtering approach are higher than the multiple-regression approach on average by a factor of 5 (Discussion S3).

4.2. PM age and residence time in the phyllosphere

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The strongest control on ${}^7\text{Be}:$ ${}^{210}\text{Pb}$ ratios in throughfall was tree species. We attribute this to an age effect and the storage of PM in phyllosphere by conifers with high foliar masses and LAI. Decreasing throughfall ${}^7\text{Be}:$ ${}^{210}\text{Pb}$ ratios in order oak > pine > spruce is consistent with characteristic leaf retention times of each species, about 6 months for oak, 1-2 years for white pine and up to 5-8 years for spruce. This trend is also consistent with the size of total ${}^{210}\text{Pb}$ inventories likely to be stored in these trees based on whole-tree mass balance, increasing from 9-times the annual flux in oak, 11-times in pine, and 17-times the ${}^{210}\text{Pb}$ annual flux in spruce (Landis, 2024b). Given these large estimates of ${}^{210}\text{Pb}$ storage in

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the canopy, and assuming that the canopy is at steady-state with respect to inputs and export, the mean residence time of ^{210}Pb storage in the canopy (S) can be described as follows, with the corresponding analytical solution:

$$\frac{\partial S}{\partial t} = A - kS - \lambda S \quad \text{Eq. 15}$$

$$S(t) = \frac{A}{k+\lambda} [1 - e^{-(k+\lambda)t}] \quad \text{Eq. 16}$$

where S is canopy inventory (Bq m^{-2}), A is the rate of canopy absorption of ^{210}Pb (which elsewhere is estimated as the product of wet and dry fluxes with their characteristic sorption coefficients of approximately 50% each, minus litter losses, summing to $57 \text{ Bq m}^{-2} \text{ y}^{-1}$ (Landis, 2024b)). Here λ is the ^{210}Pb decay constant, and k is a characteristic physicochemical weathering rate. k can be approximated if we observe a steady-state $S = 825 \text{ Bq m}^{-2}$ based on whole-tree measurements, which is an average for the two throughfall trees at the BM site. With these assumptions the mean residence time of **PM metals** in the phyllosphere with respect to weathering is 32 years (Fig. S7). Independent age estimates of individual components of an adjacent red oak tree have also been made with ^7Be : ^{210}Pb chronometry for leaves (0.5 years), twigs (1.4 years), branches (7 years), moss (28 years), lichen (21 years), live surface (20-52 years), and bark (18 to >70 years). These ages are consistent with the superposition or ordering of, e.g. twig versus branch, and canopy crown versus tree base, and overall confirm that canopy storage of metals persists for decades (Landis, 2024b).

4.3. Implications for rates of MTE deposition to forest vegetation

The FRNs stand apart from MTEs in canopy mass balance, having EFs <1 and NTF that scales inversely with rainfall depth. What distinguishes the FRNs from other PM metals is radioactive decay, and the obvious cause for their diverging behaviors is a difference in characteristic ecosystem residence times. Whereas the FRNs are quintessential atmospheric metals and tracers of $\text{PM}_{2.5}$, other trace metal cycles are therefore dominated by recycled pathways rather than novel PM deposition. This is illustrated clearly for the pairs ^{210}Pb versus Pb^T , and ^7Be versus ^9Be . Within each pair the FRN shows net sorption to the canopy while the stable isotope shows net release. Within each pair elemental chemistry must be conserved, and differences must therefore derive from half-life restrictions which in turn relate to different PM sources. In the absence of these insights from FRN mass balance, for example, large EFs for both ^9Be and Pb^T in throughfall would be attributed to novel dry deposition rather than to a change in storage from recycled materials and legacy deposition. This has obvious consequences for estimating PM budgets and dry deposition by canopy mass balance, which may otherwise be overestimated by factors ranging from 2-10 (Discussion S3).

A critical question thus emerges: from where does a change in storage originate? Possible sources cannot be discriminated by our current methods but include resuspension of aged **PM**, e.g., dust, from the surrounding forest floor, as well as regional sources such as vehicular traffic and road dust; excretion of metabolic products from flora of the phyllosphere; or weathering of PM from legacy deposition to living and dead surfaces of the phyllosphere. For all MTEs, the strong dependence of both NTF and ΔS on DOC indicates that the immediate metal source is within the phyllosphere. Production of DOC in the phyllosphere in this case appears to act as a weathering mechanism, meaning that **PM, metals**, and dust are efficiently captured in the canopy, their constituents solubilized over some characteristic time frame that may span years to decades, likely altering their chemical speciation in association with organic carbon, and exporting them from the canopy in an operationally dissolved phase by DOC (Gandois et al., 2010; Roulier et al., 2021). Poorly soluble metals such as Fe and Al require solubilization by DOC to generate high throughfall EFs through ΔS (Hou et al., 2005). In conclusion, the strong correlations of ΔS among many TMs suggest that ΔS may be an emergent ecosystem property through which metal,

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679 carbon, and hydrologic cycles converge (Van Stan and Stubbins, 2018). The phyllosphere thus plays a
680 critical role in regulating timing, bioavailability, and reactivity of metals and carbon subsequently
681 transferred to underlying soils. Future work should aim to extend the FRN mass balance to additional
682 biomes and climate gradients so that the importance of phyllosphere storage for PM dynamics can be
683 evaluated on a global scale in a context of climate change.

684 The role of organic particulate matter or FPOM is important in all MTE canopy cycles and warrants
685 greater attention in future throughfall mass balance (Van Stan et al., 2021). Including the FPOM fraction
686 in mass budgets increases the net throughfall of each MTE on average by 26%, but this number is higher
687 for particle reactive elements including Hg (73%) or Pb (45%). Export of FPOM from the canopy mirrors
688 processes that we observe for all MTEs. Its dependence on both precipitation and antecedent dry
689 period, in turn, obscures the impact of these factors on MTEs which would otherwise be attributed to
690 either leaching or novel dry deposition, respectively. Canopy FPOM export is highly seasonal, peaking
691 strongly in summer to reflect productivity of the biological engine in the canopies of all tree species
692 measured here (notably, both deciduous and coniferous). The production of FPOM via herbivory likely
693 plays an important role (Frost and Hunter, 2004; Michalzik and Stadler, 2005). FPOM moreover figures
694 directly in predictive models for DOC export and may thus represent a source of soluble carbon. As such,
695 FPOM and DOC together influence trace metal cycling in the canopy and unite many trace metals in
696 carbon cycling within the phyllosphere.

697 4.4. Implications of aerosol ΔS for trace metal throughfall budgets

698 PM storage in the forest canopy introduces complexity to metal exchange processes that control
699 throughfall mass balance. Our compilation of literature TM throughfall highlights the widely varying
700 behaviors of individual metals to transit through the forest canopy, ranging from strong net sorption to
701 strong net release (Fig. 1). This variation is unlikely to result from elemental solubility behavior,
702 metabolic cycling, or the vagaries of event-scale interaction between precipitation and canopy. Instead,
703 we suggest that throughfall export from the canopy reflects the complexity of aerosol sources, how
704 these have changed through time in both source strength and PM characteristics (Cho et al., 2011), and
705 transformations that occur during PM residence both in the canopy and underlying forest soil. Thus,
706 time and history are critical determinants in throughfall metal measurements.

707 With global anthropogenic emissions of many metals reduced in recent decades through emissions
708 controls, the canopy is unlikely to be in steady-state with contemporary atmospheric deposition.
709 Throughfall is thus likely to be a biased metric of deposition if a change in storage from prior deposition
710 cannot be assessed. Future insights into the storage and solubilization of aerosol metals in the
711 phyllosphere will require a combination of approaches to unravel varying sources and timescales of
712 aerosol deposition to the canopy. We suggest that FRN mass balance and micrometeorology methods
713 for, e.g., Hg deposition, are complementary approaches that can combined to verify both PM
714 depositional processes and fluxes (Landis et al., 2024b). Here we have shown that multiple-regression
715 mass balance, coupled with new FRN mass balance, and in combination whole-tree measurements
716 (Landis, 2024b), can accurately quantify both new PM and TM deposition and their release from long-
717 term storage. We anticipate that coupling throughfall, whole-tree mass balances, and FRN chronometry
718 with stable isotope systems of, e.g., Cd, Cu, Pb, Hg, will yield new insights into metal dynamics in forest
719 canopies and should be a focus of future efforts to understand fate and redistribution of
720 atmospherically-derived metals.

Deleted: , as well,

Deleted: particulate matter

Deleted: in generation of FPOM

Deleted: trace metal

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731 **Data Availability**

732 The data prepared for this manuscript are freely available to the public [Landis, Joshua (2025),
733 “Throughfall mass balance with fallout radionuclides and major/trace elements”, Mendeley Data, V1,
734 doi: 10.17632/r9kpgp76xh.1].

735

736 **Supplementary Materials**

737 Supplementary Materials supporting this manuscript are available at the publisher website.

738 **Author Contributions**

739 This work was completed wholly by JDL in the course of Ph.D dissertation research.

740

741 **Competing Interests**

742 The author declares that no competing interests influenced the conduct or conclusions of this research.

743

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751

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