



1    **Numerical simulation of nitrous oxide over Asia using**  
2    **regional climate-chemistry-ecology coupling model**  
3    **RegCM-Chem-YIBs**

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11    **Abstract.** Nitrous oxide (N<sub>2</sub>O) is a significant greenhouse gas that not only contributes to global warming  
12    but also depletes the ozone layer. In our study, we enhanced a regional climate-chemistry-ecology model  
13    to better understand how N<sub>2</sub>O is emitted, transported, and dispersed in the atmosphere. We focused on  
14    East Asia, South Asia, and Southeast Asia, using two different datasets to analyze the patterns of N<sub>2</sub>O in  
15    2020. Our model showed good agreement with real-world observations, revealing that N<sub>2</sub>O levels vary  
16    seasonally and spatially. For example, the lowest concentrations were found in June, while the highest  
17    were in December. Certain areas, like the North China Plain and the Ganges River Basin, had higher  
18    N<sub>2</sub>O levels. We also found that N<sub>2</sub>O concentrations decrease with altitude. By validating our model, we  
19    gained insights into the complex interactions between N<sub>2</sub>O emissions and atmospheric processes. This  
20    research helps policymakers develop strategies to reduce N<sub>2</sub>O emissions. In the future, we aim to refine  
21    our model further to improve predictions of N<sub>2</sub>O emissions and distribution, which will support efforts  
22    to combat climate change and protect the ozone layer.

1 Introduction

23        Nitrous oxide (N<sub>2</sub>O) ranks as the third most prevalent greenhouse gas in the atmosphere, following  
24        carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>). It possesses a warming potential 273 times greater than that  
25        of carbon dioxide and contributes approximately 6% to the radiative forcing of long-lived greenhouse  
26        gases (WMO, 2023; NOAA, 2024), with an atmospheric lifetime of  $116 \pm 9$  years (Prather et al., 2015).  
27        By 2022, the global average concentration of N<sub>2</sub>O had reached  $335.8 \pm 0.1$  parts per billion (ppb),  
28        reflecting a 124% increase since the 1850s. This makes N<sub>2</sub>O the primary driver of the increase in the



29 effective radiative forcing of long-lived gases since 1990 (Wmo, 2023). In addition to its role as a potent  
30 greenhouse gas,  $\text{N}_2\text{O}$  is also a significant ozone-depleting substance. Its photolysis in the stratosphere  
31 generates  $\text{NO}_x$ , which contributes to ozone depletion (Mcelroy and McConnell, 1971; Unep, 2013). In  
32 2020, anthropogenic  $\text{N}_2\text{O}$  emissions, expressed in terms of CFC-11 equivalents, exceeded those of all  
33 CFCs by more than twofold, representing over 20% of the peak CFC emissions recorded in 1987 (Geneva,  
34 2022). Currently,  $\text{N}_2\text{O}$  emissions are the most substantial among ozone-depleting substances ( $\text{ODS}_s$ ) and  
35 are anticipated to remain the largest throughout the 21st century (Ravishankara et al., 2009). The  
36 escalating concentration of  $\text{N}_2\text{O}$  exerts a profound impact on global warming, air pollution, and ozone  
37 layer depletion, prompting extensive global concern. In response, the international community has  
38 implemented various conventions and agreements to regulate  $\text{N}_2\text{O}$  emissions, aiming to mitigate its  
39 detrimental effects on the climate and ozone layer. The United Nations Framework Convention on  
40 Climate Change (UNFCCC), along with its Kyoto Protocol and the Paris Agreement (Idowu et al., 2023),  
41 constitute three pivotal international legal instruments in the fight against climate change. These  
42 frameworks guide global efforts to reduce emissions. The UNFCCC is dedicated to controlling  
43 greenhouse gas emissions to combat global warming, the Kyoto Protocol establishes specific reduction  
44 targets for developed countries, including  $\text{N}_2\text{O}$ , and the Paris Agreement reinforces the global  
45 commitment to emission reduction by requiring nations to formulate and execute their Intended  
46 Nationally Determined Contributions (NDCs) to cap the increase in global average temperature.

47 As a greenhouse gas of significant global concern, accurately simulating the emission and  
48 concentration distribution of  $\text{N}_2\text{O}$  is essential for assessing the impacts of climate change and crafting  
49 effective emission reduction strategies (De Sisto et al., 2024; Zhou et al., 2020). Tian (2020) quantified  
50 global natural and anthropogenic sources and sinks of  $\text{N}_2\text{O}$  from 1980 to 2016, using both top-down and  
51 bottom-up approaches. Their findings indicate that global anthropogenic emissions, primarily driven by  
52 nitrogen inputs from agricultural activities, have increased by 30% during this period, leading to a  
53 growing atmospheric burden. The imbalance between  $\text{N}_2\text{O}$  sources and sinks has contributed to the rise  
54 in its atmospheric concentration. Previous research has predominantly concentrated on modeling  $\text{N}_2\text{O}$   
55 emissions to diminish uncertainty. This includes relatively simple empirical models such as the DAISY  
56 model (Hansen et al., 1991) and the CRISP model (Nielsen et al., 1999), as well as process-based models  
57 like DAYCENT (Parton et al., 1996), DNDC (Stange et al., 2000), ECOSYS (Grant et al., 2001), MiCNiT



58 (Blagodatsky et al., 2011), and the Dynamic Land Ecosystem Model (DLEM) (Tian et al., 2013). While  
59 the simulation of N<sub>2</sub>O emissions has seen consistent improvement, there remains a scarcity of models  
60 capable of accurately simulating N<sub>2</sub>O concentrations. Therefore, the development of high-resolution  
61 atmospheric chemical models is crucial for simulating the concentration distribution of N<sub>2</sub>O in the  
62 atmosphere. Precise simulation of N<sub>2</sub>O emissions and concentration dynamics is vital for forecasting its  
63 potential environmental impacts and providing a scientific foundation upon which to base the  
64 development of effective environmental policies.

65 Research (Tian et al., 2020; Weber et al., 2024) indicates that East Asia (EA) and South Asia (SA)  
66 are the primary contributors to global N<sub>2</sub>O emissions. Agricultural activities in these regions, particularly  
67 the application of fertilizers and the use of livestock manure, are highly concentrated, which are the  
68 predominant sources of N<sub>2</sub>O (De Sisto et al., 2024; Zhang et al., 2022). In China and India, emissions  
69 from fertilizer use are dominant, with agriculture accounting for approximately 70% of anthropogenic  
70 N<sub>2</sub>O emissions (Tian et al., 2020). Consequently, agricultural practices in these areas have a substantial  
71 impact on the global N<sub>2</sub>O emission profile. EA is responsible for 71-79% of global aquaculture N<sub>2</sub>O  
72 emissions, while SA and Southeast Asia (SEA) together contribute 10-20% (Beusen et al., 2016; Macleod,  
73 2019). The densely populated and rapidly developing economies of EA, SA and SEA have led to an  
74 increased demand for food and animal feed, further intensifying agricultural activities and the associated  
75 N<sub>2</sub>O emissions (Tian et al., 2020). As both a significant greenhouse gas and a stratospheric ozone-  
76 depleting substance, N<sub>2</sub>O presents additional challenges to environmental sustainability and regional  
77 climate in EA (Bi et al., 2013). Therefore, monitoring and modeling the concentration distribution of  
78 N<sub>2</sub>O is essential for comprehending and predicting climate change trends (Ma et al., 2021). Satellite data  
79 inversion reveals that N<sub>2</sub>O concentrations exhibit significant temporal and spatial variability, influenced  
80 not only by local emissions but also by regional transport from external areas. This underscores the  
81 importance of regional and global cooperation in mitigating N<sub>2</sub>O emissions and safeguarding the ozone  
82 layer. Consequently, simulating and studying N<sub>2</sub>O concentration patterns in EA, SA and SEA will not  
83 only assist in pinpointing key emissions and transport pathways but also provide a scientific foundation  
84 for developing effective regional climate policies and measures to protect the ozone layer (Weber et al.,  
85 2024).

86 In this study, we employ an advanced coupled regional climate-chemical-ecosystem model,



87 RegCM-Chem-YIBs, to simulate and analyze  $\text{N}_2\text{O}$  concentrations in EA, SA and SEA during 2020. The  
88 primary objective is to evaluate the atmospheric response to the rapid increase in terrestrial  $\text{N}_2\text{O}$   
89 emissions. Our research is designed to accurately simulate the fluctuations in  $\text{N}_2\text{O}$  concentrations,  
90 thereby enabling the prediction of their potential environmental impacts and laying a scientific  
91 groundwork for the development of impactful environmental policies. Furthermore, the findings of this  
92 study are poised to offer invaluable insights for the formulation of greenhouse gas (GHG) reduction  
93 strategies. The methodologies and datasets utilized in this research are meticulously outlined in Section  
94 2, while the ensuing results are thoroughly examined in Section 3. Concluding remarks, summarizing  
95 our key findings and their implications, are presented in Section 4.

## 96 **2 Model and Methods**

### 97 **2.1 Model description**

98 This study employs the regional climate-chemistry-ecology coupling model (RegCM-Chem-  
99 YIBs) (Xie et al., 2019) which analyze the temporal and spatial distribution of short-lived air pollutants  
100 and their interactions with the long-lived greenhouse gas  $\text{CO}_2$ . Additionally, it investigates the  
101 processes and mechanisms underlying the interactions between air pollutants, greenhouse gases, and  
102 regional climate. To enhance air pollution control and respond to climate change more effectively, the  
103 regional climate chemical model RegCM-Chem (Shalaby et al., 2012) is integrated with the ecological  
104 model YIBs (Yue and Unger, 2015), resulting in the development of a regional climate-chemical-  
105 ecological coupling framework. Employing this advanced model, the authors have unveiled the spatial  
106 and temporal distribution characteristics and influencing factors of terrestrial  $\text{CO}_2$  concentration and  
107 terrestrial carbon flux in China over recent years (Xie et al., 2019; Xie et al., 2020). They have also  
108 quantitatively analyzed the interaction effects of  $\text{O}_3$ ,  $\text{PM}_{2.5}$ , and  $\text{CO}_2$  on the East Asian summer  
109 monsoon climate (Ma et al., 2023b; Ma et al., 2023a). Drawing on the representative concentration  
110 pathway climate scenario RCP4.5, the study simulates and forecasts the impacts of future regional  
111 emission reduction policies and global climate change on air pollution and climate change in China,  
112 aligning with the "dual-carbon" goal (Xu et al., 2022; Xu et al., 2023). Furthermore, the model was  
113 utilized to explore the interaction dynamics of  $\text{CO}_2$ , ozone, and particulate matter through terrestrial



114 vegetation (Xie et al., 2024). The incorporation of heterogeneous methane ( $\text{CH}_4$ ) allows for a  
115 comprehensive assessment of the radiative and chemical impacts of  $\text{CH}_4$  on terrestrial carbon fluxes  
116 across East Asia (EA), South Asia (SA), and Southeast Asia (SEA) in 2010 (Zhang et al., 2024).

117  $\text{N}_2\text{O}$  in the atmosphere is governed by a complex interplay of factors, such as sources, sinks,  
118 transport processes, and chemical reactions, leading to a non-uniform spatiotemporal distribution. In  
119 this study, we introduce a new species of  $\text{N}_2\text{O}$  into the coupled RegCM-Chem-YIBs model, taking into  
120 account the emissions, atmospheric transport, and diffusion processes of  $\text{N}_2\text{O}$ . The enhanced model  
121 was employed to simulate  $\text{N}_2\text{O}$  over a one-year period, and a comprehensive analysis of its spatial and  
122 temporal distribution in the atmosphere was conducted.

## 123 **2.2 Experimental design**

124 The model utilized in this study features a horizontal resolution of 36 kilometers, with the upper  
125 boundary of the model domain set at 50 hPa. It is composed of a total of 18 vertical layers, centered on  
126 the grid point ( $31^\circ \text{ N}$ ,  $101^\circ \text{ E}$ ). Employing the Lambert projection, the model simulates the entire region  
127 of EA, as well as the majority of SEA and SA. The simulation period spans one year, from January 2020  
128 to December 2020, capturing the spatial and temporal distribution of  $\text{N}_2\text{O}$  within the study area for that  
129 year. For this experiment, the model integrates several advanced schemes and modules: the CBM-Z vapor  
130 phase chemistry scheme (Zaveri and Peters, 1999) for chemical processes, the Holtslag PBL boundary  
131 layer scheme (Holtslag et al., 1990) for boundary layer dynamics, the Grell cumulus convection scheme  
132 (Grell, 1993) for convective processes, the CLM4.5 land surface process module (Oleson et al., 2008;  
133 Stöckli et al., 2008) for terrestrial processes, and the CCM3 radiative transfer scheme (Zhang et al., 1998;  
134 Giorgi et al., 2012; Giorgi and Mearns, 1999) for simulating the radiative effects within the atmosphere.

## 135 **2.3 Input data**

136 The meteorological initial conditions for our model are sourced from the ERA-Interim reanalysis  
137 dataset (Dee, 2011), encompassing parameters such as temperature, humidity, potential height, and wind.  
138 These data are ingested every six hours to inform the model's meteorological boundary conditions,  
139 thereby constraining the simulated meteorological fields (Zhang et al., 2024). The model also receives  
140 average sea surface temperature data from the National Oceanic and Atmospheric Administration



141 (NOAA) (Reynolds et al., 2002) on a weekly basis, while the chemical boundary conditions reflecting  
142 the climate state are derived from the time-varying outputs of the Global Chemical Transport Model  
143 (MOZART) (Hauglustaine et al., 1998; Horowitz et al., 2003). This model operates on a 6-hourly time  
144 resolution. The anthropogenic emission inventory is based on the MIX Asia emission inventory (Li et al.,  
145 2017), which offers a spatial resolution of  $0.25^{\circ} \times 0.25^{\circ}$ . Within the scope of this paper, we utilize two  
146 distinct datasets for  $\text{N}_2\text{O}$  flux data: the Copernicus Atmosphere Monitoring Service (CAMS) (CAMS,  
147 2023) and the Emission Database for Global Atmospheric Research (EDGAR)  $\text{N}_2\text{O}$  flux datasets (Crippa,  
148 2024).

149 CAMS, which is operated by the European Centre for Medium-Range Weather Forecasts (ECMWF)  
150 on behalf of the European Commission, is one of the six services comprising the Copernicus program.  
151 The  $\text{N}_2\text{O}$  flux data for CAMS is derived using the atmospheric inversion framework PyVAR- $\text{N}_2\text{O}$ . This  
152 framework relies on observations of  $\text{N}_2\text{O}$  mixing ratios to estimate the flux that best accounts for these  
153 observations, while being informed by a priori flux estimates. The CAMS Global Inversion Optimized  
154 Greenhouse Gas (GHG) fluxes and concentrations dataset encompasses net surface fluxes, model-level  
155 mixing ratios, and column average mixing ratios for  $\text{CO}_2$ ,  $\text{CH}_4$  and  $\text{N}_2\text{O}$ . These net fluxes encompass  
156 contributions from both the natural biosphere, such as vegetation and wetlands, as well as anthropogenic  
157 sources, including fossil fuel emissions and rice fields. The optimized  $\text{N}_2\text{O}$  flux data from CAMS are  
158 archived as NetCDF files, with each file containing a month's worth of time-varying flux data at a spatial  
159 resolution of  $2.5^{\circ} \times 1.26^{\circ}$ . Additionally, a three-dimensional  $\text{N}_2\text{O}$  concentration field, generated using  
160 the optimized flux data, is also saved in NetCDF format. Each monthly file includes 79 vertical layers of  
161  $\text{N}_2\text{O}$  concentration data with a horizontal resolution of  $2.5^{\circ} \times 1.26^{\circ}$ , recorded every 3 hours. The  
162 CAMS  $\text{N}_2\text{O}$  concentration data serve as the initial and boundary conditions for model initialization, while  
163 the  $\text{N}_2\text{O}$  flux data are used as the emission input within the model to simulate  $\text{N}_2\text{O}$  emissions.

164 EDGAR is a comprehensive, independent global database that documents anthropogenic  
165 greenhouse gas emissions and air pollution worldwide. Utilizing international statistics and consistent  
166 IPCC methodologies, EDGAR offers independent emissions estimates, providing a valuable comparison  
167 to those reported by European member states or parties under the United Nations Framework Convention  
168 on Climate Change (UNFCCC). The database provides a grid of total national emissions at a global scale  
169 with a resolution of  $0.1^{\circ} \times 0.1^{\circ}$ , including annual, monthly, and hourly data. For this study, we have

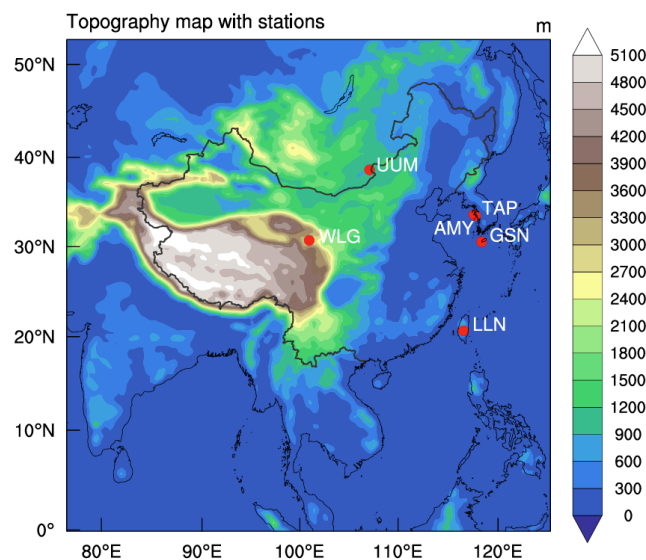


170 utilized EDGAR's monthly N<sub>2</sub>O emission grid for 2020, categorized by sector, in units of kg/m<sup>2</sup>/s, with  
171 a spatial resolution of 0.1° x 0.1° . The sectors covered include: the power industry for electric and  
172 thermal power plants (both public and private), industrial combustion for manufacturing processes,  
173 small-scale non-industrial fixed combustion in buildings, transportation for mobile combustion sources  
174 (road, rail, ship and aviation), agricultural soils, crop residue burning, enteric fermentation, manure  
175 management, and agricultural indirect N<sub>2</sub>O emissions. Additionally, it includes emissions from fuel  
176 production, fuel extraction (such as conversion and refining), industrial processes (e.g. cement, steel,  
177 aluminum, chemicals, solvents), and the waste industry, encompassing solid waste treatment and  
178 wastewater management. The 12-month emissions for each sector in 2020, as provided by EDGAR, were  
179 consolidated into a total flux file. This file was then input into the model as an alternative set of N<sub>2</sub>O  
180 emissions for simulation purposes.



181     **2.4 Validation data**

182         The World Greenhouse Gas Data Center (WDCGG) (Lan, 2023) is a World Data Center (WDC)  
183         operated by the Japan Meteorological Agency (JMA) under the Global Atmospheric Watch (GAW)  
184         program of the World Meteorological Organization (WMO). WDCGG collects, archives, and distributes  
185         data on greenhouse gases (e.g. CO<sub>2</sub>, CH<sub>4</sub>, CFCs, N<sub>2</sub>O) and related gases (e.g. CO) from the atmosphere  
186         and other sources, provided by various contributors. In this study, data from six N<sub>2</sub>O observation sites in  
187         WDCGG were utilized to validate the model simulation results. These sites include Anmyeon-do (AMY),  
188         Gosan (GSN), and Tae-ahn Peninsula (TAP) in the Republic of Korea, Ulaan Uul (UUM) in Mongolia,  
189         and Mt. Waliguan (WLG) and Lulin (LLN) in China. The N<sub>2</sub>O concentration data from WDCGG  
190         encompass monthly data for the six sites, daily data for AMY and GSN, and hourly data for AMY and  
191         GSN. Both WLG and LLN are located on plateaus, while UUM is situated in a vast and sparsely  
192         populated area, less influenced by human activities, thus providing a better representation of natural N<sub>2</sub>O  
193         concentrations. The AMY and GSN stations are near the ocean and are significantly affected by marine  
194         sources. The spatial distribution of these N<sub>2</sub>O monitoring stations is depicted in Fig.1, and their latitude  
195         and longitude coordinates are listed in Table 1.



196  
197     **Figure 1.** The model domain of the RegCM-Chem-YIBs model and location of WDCGG N<sub>2</sub>O  
198     measurement sites.





**Table 1.** Geographic information of the observation sites

Site	Abbreviation	Longitude(°E)	Latitude(°N)	Altitude(m)
Anmyeon-do	AMY	126.33	36.54	42
Gosan	GSN	126.16	33.29	71
Lulin	LLN	120.87	23.47	2862
Tae-ahn Peninsula	TAP	126.13	36.73	20
Ulaan Uul	UUM	111.08	44.44	992
Mt. Waliguan	WLG	100.90	36.29	3810

### 3 Results and discussion

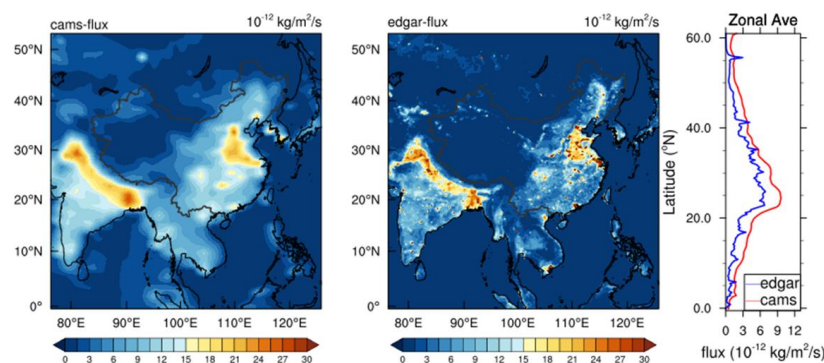
#### 3.1 Distribution characteristics of two different emissions

In this trial, we employed two distinct emission datasets: the CAMS net surface N<sub>2</sub>O flux derived from ECMWF inversion and EDGAR's sectoral N<sub>2</sub>O emission inventory. To facilitate a more direct comparison of emission differences between these two sources, we averaged the emissions at the grid points within the simulated region and compared the average monthly emissions for each month of 2020 between the two datasets (as shown in Fig.7). The analysis reveals that for the flux of EDGAR remains relatively stable throughout the year, with no significant seasonal variation. The average annual nitrification flux is  $2.475 \times 10^{-12}$  kg/m<sup>2</sup>/s, with a fluctuation amplitude of 0.27 ppb. In contrast, the net flux from CAMS exhibits greater variability than the EDGAR, with a notable increase in flux during July, August, and September. The N<sub>2</sub>O emission is comparatively lower in winter, with a fluctuation amplitude of 0.96 ppb. The average annual emission flux of nitrous oxide for CAMS is  $3.965 \times 10^{-12}$  kg/m<sup>2</sup>/s, which is 1.6 times that of the EDGAR. The accounting of anthropogenic and natural sources of N<sub>2</sub>O in China indicates that anthropogenic emissions in China for the year 2020 amount to  $6.3 \times 10^{-12}$  kg/m<sup>2</sup>/s, while the combined natural and anthropogenic emissions total  $7.5 \times 10^{-12}$  kg/m<sup>2</sup>/s (Liang et al., 2024b). These figures are more aligned with the emission utilized in this study.

We have depicted the annually spatial distributions and zonally averaged N<sub>2</sub>O fluxes for both datasets across the year in Fig.2. We selected January, April, July and October to represent the four seasons of spring, summer, autumn and winter to analyze the seasonal characteristics of N<sub>2</sub>O, as shown in Fig.S1. As observed in Fig.2, within the study area, North China and the Ganges River basin in India exhibit elevated N<sub>2</sub>O fluxes, surpassing  $1.5 \times 10^{-11}$  kg/m<sup>2</sup>/s. Besides these two extensive regions with



222 significant  $\text{N}_2\text{O}$  emissions, China and India also feature numerous localized areas with high  $\text{N}_2\text{O}$   
 223 emissions. This could be attributed to the intensive use of nitrogen fertilizers in these regions, as they  
 224 are a primary source of  $\text{N}_2\text{O}$  from agricultural lands (Shengji Yan, 2022). Furthermore, urbanization  
 225 can lead to reduced flow rates and dissolved oxygen levels, as well as increased nitrogen availability  
 226 and denitrification processes, potentially contributing to higher concentrations and fluxes of  $\text{N}_2\text{O}$  (Chen  
 227 et al., 2025). In contrast, Northwest China, the Tibetan Plateau, and Inner Mongolia are characterized  
 228 by relatively lower  $\text{N}_2\text{O}$  emissions. This may be due to less agricultural intensification, reduced  
 229 nitrogen fertilizer application, distinct climatic conditions and soil properties, and a more pristine and  
 230 less polluted natural environment in these areas (Wang et al., 1998; Du et al., 2016). The pattern of high  
 231 and low emission regions aligns with the findings of previous studies (Liang et al., 2024b; Zhou et al.,  
 232 2014; Garg et al., 2011; Wang et al., 1998). It is noteworthy that in the CAMS flux distribution, the  
 233 western Pacific region along China's coast displays a relatively pronounced negative  $\text{N}_2\text{O}$  flux during  
 234 winter. This indicates that the ocean along China's coast acts as a weak sink for  $\text{N}_2\text{O}$  during late autumn  
 235 and throughout winter and early spring. In summary, it is evident that the EDGAR datasets, with their  
 236 higher resolution of  $0.1^\circ \times 0.1^\circ$ , provide a more detailed distribution of  $\text{N}_2\text{O}$  in the target area compared  
 237 to the CAMS flux data at  $2.5^\circ \times 2.5^\circ$ . Despite the CAMS datasets being more accurate and closer to  
 238 reality, they exhibit significant seasonal variations and fluctuations.

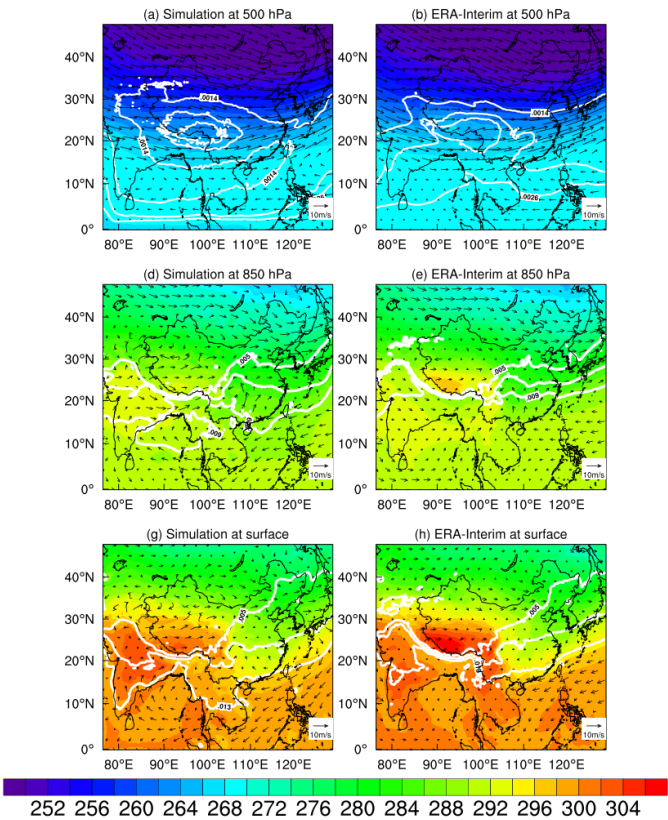


239  
 240 **Figure 2.** Comparison of annually averaged spatial distribution and zonal mean nitrous oxide flux  
 241 between two datasets.



### 242 3.2 Model evaluation

243 By comparing the model simulation results with the observations, we evaluated the performance of  
244 the model based on a number of meteorological factors, including air temperature (T), relative humidity  
245 (RH), and wind speed. As shown in Fig.3 and Fig.S2, as well as Table 2, the RegCM-Chem-YIBs model  
246 has successfully captured the spatial distribution and intensity of both temperature and specific humidity  
247 at the 500hPa, 850hPa, and 1000hPa pressure levels. Notably, the model's simulated correlation  
248 coefficient for wind at 500hPa height reached 0.95, indicating the precision of the model's simulation  
249 capabilities. The simulation correlation between the model and observational data for the middle and  
250 lower atmospheric layers is also significant, with values ranging from 0.36 to 0.44, as detailed in Table  
251 2. Within these layers, the model's simulation of wind speed over the Qinghai-Tibet Plateau and sea areas  
252 is relatively accurate, and the wind direction is largely correctly predicted. Prior research (Xie et al., 2019;  
253 Xie et al., 2020; Xu et al., 2023; Xu et al., 2022; Ma et al., 2023a; Ma et al., 2023b; Zhang et al., 2024;  
254 Xie et al., 2024; Yin et al., 2015; Gao et al., 2023) has similarly validated the model's proficiency in  
255 accurately representing air quality, weather fields, and major greenhouse gases in Asia. When comparing  
256 the two distinct sources, it was observed that the meteorological field simulated using the CAMS flux  
257 model aligns more closely with observational outcomes. Given that CAMS flux data are derived from  
258 the inversion of model results, they are deemed more precise than the EDGAR emission inventories,  
259 especially under the dual constraints of observational and model-based accuracy. Furthermore, different  
260 models may exhibit varying sensitivities to emissions, which is contingent upon how chemical and  
261 physical processes are handled within the model framework. Models that exhibit heightened sensitivity  
262 to emissions may yield more precise simulation results under specific conditions.



**Figure 3.** Annual average temperature (color-filled; units: K), specific humidity (contours; units: kg kg<sup>-1</sup>) and wind field (streamlines; units: m/s) at 500 hPa (a, b), 850 hPa (c, d) and surface (e, f) for the ERA-Interim reanalysis data (a, c, e) and model simulation (b, d, f). This is the result of the CAMS dataset.

**Table 2.** Statistical indicators for comparing model simulation results with reanalysis data Statistical characteristics of simulated (Sim) and observed (Obs) monthly N<sub>2</sub>O mixing ratios at the seven stations

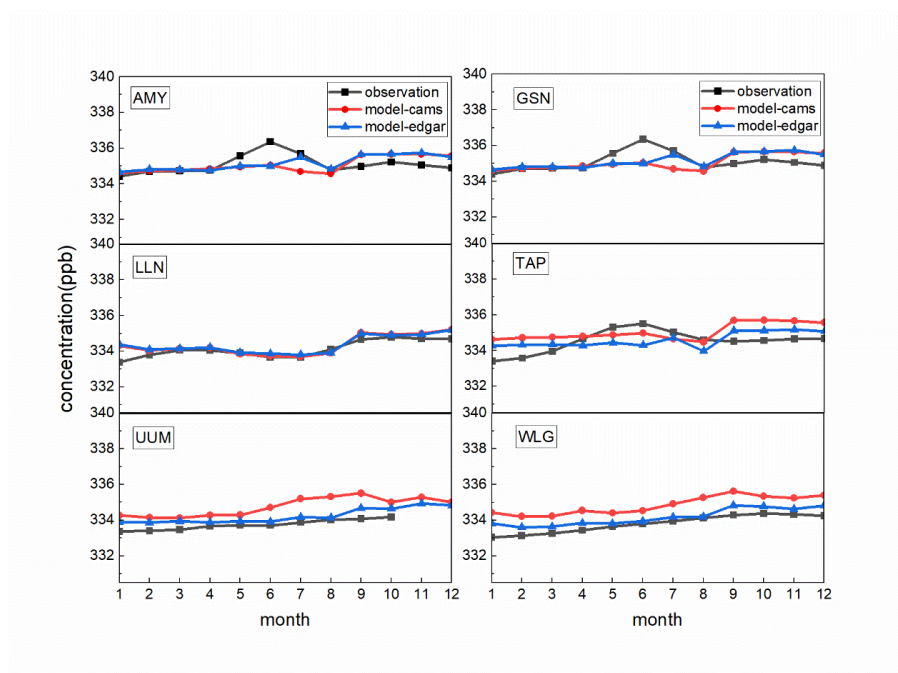
Factors	Altitude	<sup>a</sup> R	<sup>b</sup> R
Temperature (K)	500hPa	0.998	0.997
	850hPa	0.989	0.988
	1000hPa	0.988	0.988
Specific humidity (Kg Kg <sup>-1</sup> )	500hPa	0.854	0.831
	850hPa	0.985	0.983
	1000hPa	0.988	0.986
Wind speed (m/s)	500hPa	0.963	0.952
	850hPa	0.442	0.424
	1000hPa	0.363	0.368

(<sup>a</sup>R indicates the correlation coefficient of CAMS, <sup>b</sup>R indicates the correlation coefficient of EDGAR.)



270 Figure 4 juxtaposes the monthly average mixing ratios of  $\text{N}_2\text{O}$  at the six observation sites, as both  
271 observed and simulated by the model, with detailed statistical insights provided in Table 3. The figure  
272 reveals that the model's simulated  $\text{N}_2\text{O}$  concentrations closely match the observed values, particularly at  
273 the LLN station. Notably, the  $\text{N}_2\text{O}$  mixing ratios at TAP, AMY, and GSN exhibit pronounced seasonality,  
274 peaking in spring or summer and plummeting in winter. This pattern is attributed to higher air-sea fluxes  
275 during spring and summer, with the East China Sea and South China Sea regions serving as net sources  
276 of  $\text{N}_2\text{O}$  (Chen et al., 2021). At inland locations such as UUM and WLG, the simulated and observed  $\text{N}_2\text{O}$   
277 concentrations follow a similar upward trend. However, the  $\text{N}_2\text{O}$  concentrations simulated with both  
278 sources exceed the observed concentrations, with those using EDGAR emissions being closer to the  
279 actual measurements. For the LLN site, the simulation outcomes from the two emissions are highly  
280 congruent, mirroring the observational data. The LLN site exhibits subdued seasonality, with  $\text{N}_2\text{O}$   
281 concentrations being lower in summer and winter and slightly elevated in spring and autumn.

282 Based on the fitting results from the model simulations and observations at the six sites, as detailed  
283 in Table S1, the correlation coefficients for LLN, WLG, and UUM are exceptionally high. Generally, the  
284 simulation results using the CAMS emission source outperform those using EDGAR, with correlation  
285 coefficients ranging from 0.82 to 0.92. In contrast, the correlation for the other three coastal sites is  
286 comparatively low. This is likely due to the significant influence of marine environmental factors such  
287 as land and sea breezes, sea fog, and salt fog on meteorological conditions in coastal areas (Qiu and Fan,  
288 2013). The complexity of these factors may contribute to the less accurate simulation outcomes in these  
289 regions. For these six sites, the discrepancy between the CAMS-calculated concentration values and the  
290 observed values is minimal, varying from -0.09% to 0.31%. Similarly, the deviation between the  
291 concentration values calculated by EDGAR and the observed values is also very small, ranging from -  
292 0.18% to 0.14%. Regarding the annual variation in simulated  $\text{N}_2\text{O}$  concentrations, UUM, WLG and LLN  
293 sites exhibit smaller fluctuations, while TAP, AMY and GSN sites display larger annual variations. In  
294 coastal regions, the sea-air exchange flux is substantial, and the transfer of  $\text{N}_2\text{O}$  from the ocean to the  
295 atmosphere is more dynamic, potentially leading to greater annual variability of  $\text{N}_2\text{O}$  at coastal sites  
296 (Wang et al., 2023).



297

298 **Figure 4.** The monthly mean mixing ratio of nitrous oxide at the six sites observed and simulated by the  
 299 model.

300 **Table 3.** Statistical characteristics of simulated (Sim) and observed (Obs) monthly N<sub>2</sub>O mixing ratios at  
 301 the seven stations Correlation coefficients and RMSE between the simulated results and the observed  
 302 values for the two emission sources

Site	Mean					Annual amplitude			Standard deviation		
	Sim1	Sim2	Obs	Bias1	Bias2	Sim1	Sim2	Obs	Sim1	Sim2	Obs
AMY	335.06	335.16	335.09	-0.01	0.02	1.09	1.08	1.95	0.44	0.41	0.52
GSN	334.43	334.12	334.73	-0.09	-0.18	1.37	1.34	1.80	0.49	0.51	0.57
LLN	334.34	334.36	334.12	0.07	0.07	1.52	1.38	1.41	0.56	0.50	0.46
TAP	335.04	334.60	334.53	0.15	0.02	1.22	1.21	2.10	0.47	0.42	0.60
UUM	334.76	334.23	333.75	0.30	0.14	1.39	1.07	0.82	0.52	0.41	0.27
WLG	334.84	334.17	333.80	0.31	0.11	1.41	1.23	1.33	0.51	0.47	0.47

303  $(Bias_1 = \frac{Sim_1 - Obs}{Obs} \times 100\% \cdot Bias_2 = \frac{Sim_2 - Obs}{Obs} \times 100\% \cdot$  Subscript 1 represents CAMS flux and  
 304 subscript 2 represents EDGAR flux. Sim1 represents the nitrous oxide mixing ratio modeled using  
 305 CAMS flux, Sim2 represents the nitrous oxide mixing ratio modeled using EDGAR flux, the units of  
 306 Sim1, 2 and Obs are ppb, the unit of Bias is %)



### 3.3 Spatiotemporal distribution of N<sub>2</sub>O

Figure 5 illustrates the spatial distribution of N<sub>2</sub>O surface concentrations simulated using two different fluxes. In a similar vein, we utilize the months of January, April, July, and October to depict the seasonal fluctuation of nitrous oxide. In combination with Fig.S3, the surface N<sub>2</sub>O concentration simulated by two different emissions exhibit a relatively consistent trend: from spring to winter, the concentrations initially decrease and then increase. The monthly average surface N<sub>2</sub>O concentrations simulated by CAMS and EDGAR reach their lowest values in June, at 334.01 ppb and 333.79 ppb, respectively, and their highest values in December, at 335.42 ppb and 335.15 ppb, respectively. Regardless of the month, N<sub>2</sub>O concentrations are high in central and eastern China, particularly in the North China Plain, Sichuan Basin, and the Ganges Valley in India. In contrast, the Tibetan Plateau, the Indian Ocean, and the South China Sea are characterized by low concentrations, aligning with the intensity of imported N<sub>2</sub>O emissions. The North China Plain and Sichuan Basin are major agricultural regions in China, where extensive use of nitrogen fertilizers leads to increased N<sub>2</sub>O emissions (Liang et al., 2024b; Shengji Yan, 2022). In China, in particular, the intensive scale development of agriculture to support 22% of the world's population has resulted in heavy nitrogen fertilizer usage, further boosting N<sub>2</sub>O emissions. The central and eastern regions of China are densely populated and industrially advanced, with industrial emissions and municipal waste from these areas also contributing significantly to N<sub>2</sub>O emissions (Bu et al., 2024; Luo et al., 2019).

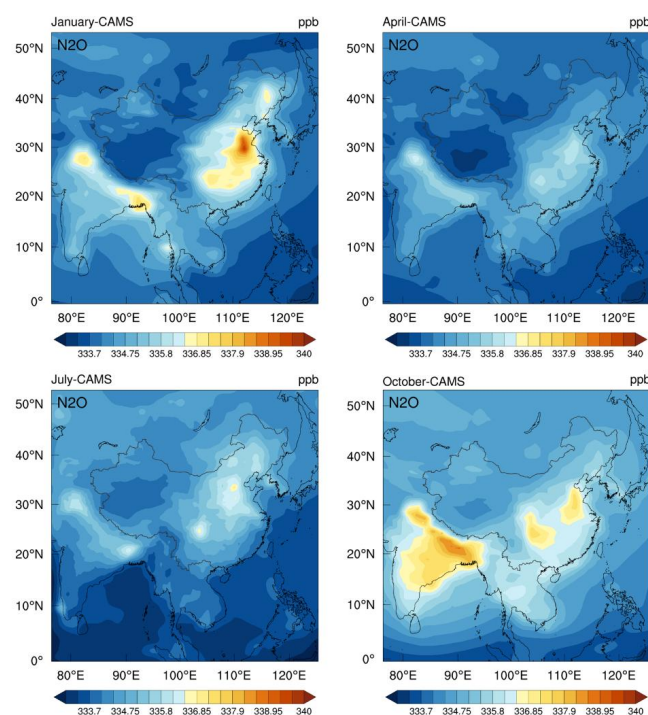
The findings indicate that ground-level N<sub>2</sub>O emissions significantly influence its atmospheric concentration, exhibiting a strong correlation. In the Qinghai-Tibet Plateau, the high altitude and low temperatures lead to reduced microbial activity, resulting in lower N<sub>2</sub>O production. Additionally, the thin air at high altitudes contributes to the relatively low concentration of N<sub>2</sub>O. In the Indian Ocean and the South China Sea, ocean dynamics primarily dictate N<sub>2</sub>O levels. While the gas mainly originates from denitrification processes in seafloor sediments, the unique marine environment keeps its concentration





331 relatively low (Craig and Gordon, 1963).

332



333

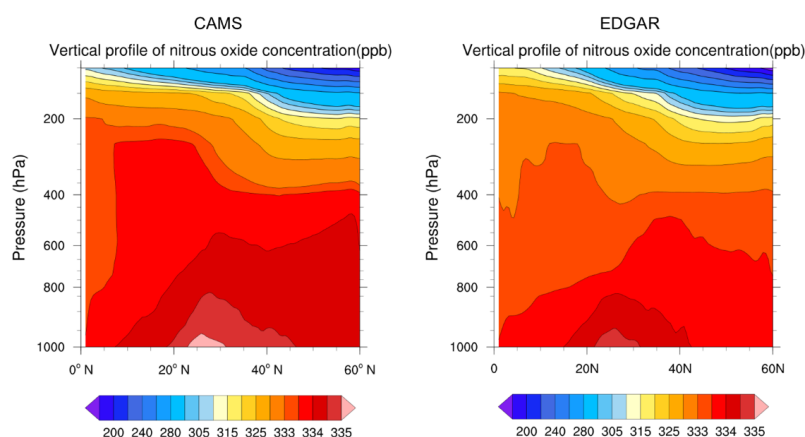
334 **Figure 5.** The distribution of the seasonal average ground-level  $N_2O$  concentration in 2020, as simulated  
 335 by the CAMS datasets

336 We selected January, April, July and October to represent the four seasons of spring, summer,  
 337 autumn and winter, respectively, to investigate the seasonal variation characteristics of the vertical  
 338 distribution of  $N_2O$  (Fig.S4). Additionally, we obtained the annual vertical profile of nitrous oxide for  
 339 the study period, depicted in Fig.6. According to Fig.6 and Fig.S4, regardless of whether using the CAMS  
 340 flux or the EDGAR flux,  $N_2O$  is uniformly distributed at high concentrations in the middle and lower  
 341 troposphere. From the upper troposphere to the lower stratosphere, the concentration of  $N_2O$  gradually  
 342 decreases (Krysztofiak et al., 2023; Tsai et al., 2012). Overall, the concentration of  $N_2O$  at the same  
 343 pressure height gradually decreases from the equator to the high latitudes of the northern hemisphere,





such as in spring and autumn. However, in summer, the high-value area shifts northward to around 30° N latitude, with a relative low-value area in the troposphere over the equatorial region. In winter, the high-value area of the troposphere is near the latitude of 15°N. The RegCM-Chem-YIBs model used in this study extends up to 50hPa, covering the middle and lower stratosphere, but it does not include any N<sub>2</sub>O loss processes. Therefore, future research should incorporate the atmospheric chemical sink of N<sub>2</sub>O in the stratosphere so as to better study the vertical change trends of N<sub>2</sub>O.



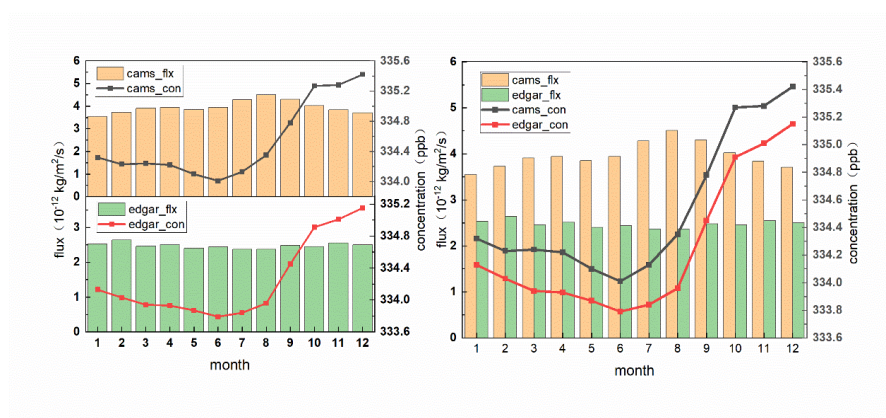
**Figure 6.** Vertical profiles of N<sub>2</sub>O in 2020 from two emissions simulations.

### 3.4 Seasonality of ground N<sub>2</sub>O concentrations

Our enhanced RegCM-Chem-YIBs model simulates the emission process of N<sub>2</sub>O, as well as its advection, diffusion, and other transport processes in the atmosphere. The resulting concentration distribution of N<sub>2</sub>O exhibits pronounced seasonality (Nevison et al., 2011; Yang et al., 2020; Jiang et al., 2007; Sun et al., 2024; Liao et al., 2004), primarily due to the combined effects of ground-level emissions, atmospheric transport processes, and stratosphere-troposphere gas exchange. We utilized two distinct sets of emissions to investigate their impacts on N<sub>2</sub>O concentrations, as illustrated in Fig.7. The EDGAR emission inventory indicates that N<sub>2</sub>O emissions show minimal variation throughout the 12 months of 2020, with only slight fluctuations. The monthly average changes in N<sub>2</sub>O flux data derived from CAMS reanalysis are slightly more pronounced than those from EDGAR, yet there is no significant seasonal fluctuation. However, the monthly average concentration of surface N<sub>2</sub>O obtained from these two sources



display the same variation trend: the concentration reaches its lowest value in June and its highest value in December, exhibiting clear seasonality, which aligns with the average seasonal variation observed at Mt. Waliguan (Liang et al., 2024a). This suggests that in the studied Northern Hemisphere region, higher emissions correspond to higher surface  $N_2O$  concentrations. However, the seasonal relationship between surface  $N_2O$  concentration and surface emissions is relatively weak, while the seasonal relationship with atmospheric transport is more significant. Nevertheless, the seasonal characteristics of  $N_2O$  over oceans and land differ, with distinct dominant influencing factors (Jiang et al., 2007; Sun et al., 2024; Liao et al., 2004), necessitating further analysis specific to these regions.



**Figure 7.** Plots of the monthly average emissions of the two emissions and their corresponding monthly mean  $N_2O$  surface concentrations

#### 4 Summary and conclusions

In this study, the RegCM-Chem-YIBs model was improved to incorporate  $N_2O$  as a species, allowing for the simulation of its emission, transport, and diffusion processes in the atmosphere, thereby obtaining the distribution of  $N_2O$ . We employed this improved model to simulate  $N_2O$  in East Asia and most parts of South and Southeast Asia in 2020 utilizing two different emissions: the CAMS model inversion flux and the EDGAR emission inventory. Subsequently, we verify the model's ability to simulate the spatial and temporal distribution of  $N_2O$  in the target region, and evaluate the model's performance using the WDCGG ground observation data. Thereafter, the spatiotemporal distribution of  $N_2O$  in the model output were analyzed. Finally, the causes of  $N_2O$  seasonality were investigated using



384 the simulations from the two emissions datasets.

385 The research indicates that between spring and winter, the average surface N<sub>2</sub>O in the study area  
386 decreases initially, then rises again. The monthly average surface N<sub>2</sub>O concentration simulated by CAMS  
387 and EDGAR reached their lowest values in June, at 334.01 ppb and 333.79 ppb, respectively, and their  
388 highest values in December, at 335.42 ppb and 335.15 ppb, respectively. Regardless of the month, the  
389 concentrations of N<sub>2</sub>O are high in central and eastern China, particularly in the North China Plain,  
390 Sichuan Basin and Ganges valley of India. In contrast, the Tibetan Plateau, the Indian Ocean, the South  
391 China Sea and other sea areas are in a state of low concentration, which is consistent with the intensity  
392 of the input N<sub>2</sub>O emissions. In the vertical profile, N<sub>2</sub>O is uniformly distributed at high concentrations in  
393 the middle and lower troposphere. From the upper troposphere to the lower stratosphere. The  
394 concentration of N<sub>2</sub>O is markedly higher near the ground and shows a gradual reduction with height. On  
395 the whole, the mixing ratio of N<sub>2</sub>O at the same pressure level gradually decreases from the equator to the  
396 higher latitudes in the northern Hemisphere. By comparing the simulation results of two different  
397 emissions, it is evident that the seasonal relationship between surface N<sub>2</sub>O concentration and surface  
398 emission is relatively weak. The model used in this study does not include the stratospheric chemical  
399 sink of N<sub>2</sub>O, so the observed seasonality is primarily related to the atmospheric transport of N<sub>2</sub>O. For  
400 this model, the temporal and spatial distribution characteristics of N<sub>2</sub>O presented by the simulation results  
401 of two emissions are largely consistent, with the results being more accurate when using the CAMS  
402 emission.

403 In summary, this study improves a reliable high-resolution regional climate-chemistry-ecology  
404 model. This enhanced model can be used to develop a high-resolution inverse modeling system for N<sub>2</sub>O  
405 in future studies, as well as to investigate its climate and environmental impacts. Additionally, there is  
406 still potential for further refinement in subsequent research. First, model may overestimate the actual  
407 concentration of N<sub>2</sub>O in the atmosphere because they do not account for loss processes, such as photolysis  
408 and chemical reactions, which reduce N<sub>2</sub>O. The photolysis of N<sub>2</sub>O in the stratosphere is a significant  
409 chemical sink, and the material exchange (STE) between the stratosphere and the troposphere greatly  
410 influences the distribution of N<sub>2</sub>O in both the troposphere and at the surface (Ray et al., 2019; Jiang et  
411 al., 2007). This exchange contributes significantly to the seasonal cycle of N<sub>2</sub>O in the troposphere of the  
412 Northern Hemisphere. Secondly, it is essential to include heterogeneous N<sub>2</sub>O in the calculation of



413 longwave radiation This would enable the estimation of the N<sub>2</sub>O radiative forcing, providing a more  
414 accurate evaluation of its radiative and climate effects. Finally, we aim to further optimize N<sub>2</sub>O flux,  
415 accurately describing its sources and sinks. These improvements will help refine the RegCM-Chem-YIBs  
416 model, allowing it to more precisely predict N<sub>2</sub>O emissions and distribution, thereby offering stronger  
417 support for global climate change mitigation and ozone layer protection.

#### 418 **Code and data availability**

419 The input data and source code for RegCM-Chem-YIBs have been archived on Zenodo at  
420 <https://doi.org/10.5281/zenodo.15043206> (Zeng et al., 2025). Data used in this study are listed as  
421 follows: the European Centre for Medium Range Weather Forecasts (ECMWF) ERA-Interim reanalysis  
422 are provided at <https://doi.org/10.24381/cds.f2f5241d> (Dee, 2011), the Copernicus Atmosphere  
423 Monitoring Service(CAMS) N<sub>2</sub>O fluxes are available at  
424 <https://ads.atmosphere.copernicus.eu/datasets/cams-global-greenhouse-gas-inversion?tab=overview>  
425 (CAMS, 2023), the EDGAR's N<sub>2</sub>O emission inventory is provided at  
426 [https://edgar.jrc.ec.europa.eu/dataset\\_ghg2024#p4](https://edgar.jrc.ec.europa.eu/dataset_ghg2024#p4) (Crippa, 2024) and the World Data Centre for  
427 Greenhouse Gases (WDCGG) N<sub>2</sub>O observation data are accessed at <https://doi.org/10.15138/53g1-x417>  
428 (Lan, 2023).

#### 429 **Author contributions**

430 TW and XZ designed this study. XZ modified the model, performed the experiments, and drafted the  
431 manuscript. TW revised the manuscript, and all authors contributed to the review and editing of the  
432 manuscript.

#### 433 **Competing interests**

434 The corresponding author has stated that all the authors have no conflicts of interest.

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