



1 **Magnetic separation reveals overestimation of soil organic matter due to undecomposed**
2 **particulate residues**

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11 **Abstract**

12 Soil organic matter (SOM) is a complex mixture of organic compounds derived from the
13 decomposition of plant and animal residues. Only after undergoing microbial transformation
14 and forming stable associations with minerals can it be considered “true” SOM, rather than a
15 simple mechanical accumulation of carbon-containing substances. According to current
16 understanding, particulate organic matter (POM) comprises both undecomposed and partially
17 decomposed organic residues. Of these, the undecomposed fraction does not qualify as SOM
18 in the strict sense. However, conventional analytical methods cannot fully distinguish fine
19 particulate residues from soil matrices, leading to an overestimation of POM-derived carbon
20 content. The extent and persistence of this “false increase” in SOM due to “disguised” POM
21 remains poorly understood. In this study, straw and biochar were magnetized via chemical
22 co-precipitation and applied to soils. The incompletely decomposed magnetized residues in
23 the soil were separated using magnetic separation at various time points, enabling more
24 accurate tracking of SOM dynamics. Five treatments were established: blank control (CK),
25 untreated straw (CS), untreated biochar with carbon input equivalent to untreated straw (Bc),
26 magnetized straw with carbon input equivalent to untreated straw (MCS), and magnetized
27 biochar with carbon input equal to untreated straw (MBc). The results showed that after the
28 application of organic materials into the soil, the recovery rate of magnetized straw residues
29 declined continuously, reaching 54.55% after 360 d, whereas biochar remained largely stable
30 at 92.48%. In CS and Bc treatments, the organic carbon content of POM fractions and their
31 proportion in the total SOM were consistently higher than in CK, particularly during early
32 incubation. However, this was attributable to overestimation from incompletely decomposed



33 residues. In contrast, MCS-D and MBc-D treatments (after magnetic residue removal)
34 showed minimal deviation from CK, confirming the contribution of incomplete
35 decomposition to SOM overestimation. On day 30, the apparent increase in the particulate
36 organic carbon (*POC*) content reached 63.48% for CS and 58.99% for Bc. Over time, the
37 overestimation in the CS treatment declined to 15.34% by day 360, whereas the
38 overestimation in the Bc treatment remained largely unchanged, with a 53.71% increase
39 persisting. These findings highlight the potential for SOM overestimation when POM
40 fractions are fully included without accounting for undecomposed inputs, particularly the
41 long-term persistence of recalcitrant organic materials, which may introduce systematic bias
42 in global SOM quantification.

43 **Keywords:** Particulate Organic Matter (POM); Soil Organic Matter (SOM); Magnetic
44 Materials; Straw; Biochar



45 **1. Introduction**

46 Soil organic matter (SOM) is a complex assemblage of organic compounds formed by
47 the decomposition and transformation of plant and animal residues, with the greatest binding
48 to minerals. It originates from partial microbial decomposition of plant detritus (Angst et al.,
49 2021; Cotrufo et al., 2013; Dou et al., 2020; Vendig et al., 2023). The core of SOM lies in
50 organic constituents with dynamic transformation characteristics that supply nutrients to soil,
51 sustain microbial activity, and influence soil structure (Feng et al., 2025), rather than in the
52 accumulation of carbon-containing substances. Distinct SOM components exhibit different
53 turnover rates and stabilization mechanisms (Sokol et al., 2022; Von Lützow et al., 2007).
54 Cambardella and Elliott (1992) proposed a physical fractionation method based on particle
55 size to extract particulate organic matter (POM) ranging from 2 mm to 53 μm . Conceptual
56 models for SOM formation and stabilization have been developed based on these fractions
57 (Christensen, 1992; Cotrufo and Lavalley, 2022; Guo et al., 2022; Lavalley et al., 2020; Rocci
58 et al., 2021; Witzgall et al., 2021).

59 In recent years, soil management and improvement measures have primarily aimed to
60 increase organic material inputs and promote microbial utilization to form SOM (Cotrufo et
61 al., 2013; Castellano et al., 2015). However, when application rates exceed microbial
62 decomposition capacity, substantial amounts of undecomposed organic material can
63 accumulate in POM over a certain period (Bhattacharyya et al., 2011; Brown et al., 2014;
64 Stewart et al., 2012), leading to sharp short-term increases in POM organic carbon content
65 (Hua et al., 2022; Liang et al., 2016; Mitchell et al., 2018). This increase is “spurious”, as
66 POM remains susceptible to decomposition and transformation even under the physical



67 protection of soil (Connell et al., 2025), and such short-term increases are unstable (Janzen,
68 2015; Powlson et al., 2014). Currently, an accurate assessment of the proportion and duration
69 of this “spurious increase” in POM mass and organic carbon content at various times
70 following organic material application is still lacking.

71 In routine experiments, methods such as heavy liquid separation, sieving, and
72 electrostatic attraction can isolate some undecomposed organic materials. However, they are
73 less effective for highly fragmented materials such as biochar, limiting evaluation of SOM
74 transformation processes. Therefore, new approaches for efficient separation of
75 undecomposed residues are required. Magnetized materials (e.g., iron-based materials such as
76 nano-zero-valent iron and iron sulfides) can be rapidly separated from soil under an external
77 magnetic field, enabling the efficient recovery of target substances (Li et al., 2024; Rana et al.,
78 2025; Zhang et al., 2025). Although biochar modified with magnetized materials has been
79 widely studied, most research has focused on heavy metal or pollutant adsorption, with no
80 application in SOM transformation. Among magnetized material preparation methods, the
81 chemical coprecipitation method has been widely used because of its operational simplicity,
82 high efficiency, and ease of impurity removal (Zhou et al., 2019). It offers excellent
83 biocompatibility, stability, and recyclability (Baragaño et al., 2020; Duan et al., 2022), which
84 facilitate the combined application of organic materials in soil.

85 In this study, a chemical coprecipitation method was used to composite straw (CS) and
86 straw biochar (Bc) into magnetized materials. At different incubation stages, undecomposed
87 magnetized organic residues were separated using an external magnetic field to eliminate
88 their interference in SOM determination. This approach allowed the accurate achievement of



89 three objectives: (i) to quantify and characterize the incompletely decomposed residues at
90 different times after organic material application; (ii) to determine the existence, proportion,
91 and duration of a “spurious increase” in POM organic carbon; and (iii) to assess the
92 proportion of organic residues ultimately transformed into stable SOM. The results provide
93 critical support for precise evaluation of POM organic carbon content and elucidation of the
94 mechanisms by which organic materials are transformed into stable SOM.

95

96 **2. Materials and Methods**

97 *2.1. Experimental materials*

98 The test soil was collected from the experimental station of Jilin Agricultural University,
99 located in the semi-humid region of Northeast China (43°48′43.57″N, 125°23′38.50″E). The
100 region has a temperate semi-humid climate, with an annual mean temperature of 4.6°C and
101 average annual precipitation ranging from 600 to 700 mm. The soil is classified as Black Soil
102 under the suborder of semi-moist temperature semi-eluvial soil in the Chinese soil
103 classification system, which is equivalent to Argiudolls in the USDA soil taxonomy. In
104 September 2023, 100 soil samples were randomly collected from the 0–20 cm layer using a
105 soil auger and combined to form a composite sample. After sampling, visible organic
106 residues were manually removed. The field-moist soil was air-dried and sieved through a 2
107 mm mesh for subsequent incubation. The basic properties of the soil were as follows: soil
108 organic matter, 22.76 g kg⁻¹; total nitrogen, 1.28 g kg⁻¹; available nitrogen, 132.21 mg kg⁻¹;
109 available phosphorus, 18.52 mg kg⁻¹; and available potassium, 99.32 mg kg⁻¹.

110 The corn stover (CS) used in the experiment was obtained from the Experimental



111 Station of the Jilin Agricultural University, Jilin Province, China (Changchun, China). The
112 entire CS was rinsed with deionized water to remove surface ash and soil, dried in an oven
113 for 24 h, ground using a grinder, and sieved through a 20-mesh sieve for later use.

114 The sieved straw powder (passing 20 mesh) was placed in a tubular furnace and
115 pyrolyzed at 500°C for 2 h under a nitrogen atmosphere at a heating rate of 5°C min⁻¹. After
116 cooling to room temperature, the resulting black solid was collected as straw biochar (Bc) for
117 further use.

118 The magnetized straw (MCS) and magnetized biochar (MBc) were prepared using the
119 chemical coprecipitation method (Zhou et al., 2019) as follows: 2.5 g of FeCl₃·6H₂O and 1.5
120 g of FeSO₄·7H₂O (Fe³⁺:Fe²⁺ molar ratio of 2:1) were weighed into a beaker. Subsequently,
121 2.0 g of dried CS or Bc was introduced to 100 mL of ultrapure water. The mixture was
122 thoroughly stirred at room temperature for 30 s using a magnetic stirrer. An excess of
123 ammonia solution (NH₃·H₂O) was subsequently added to adjust the pH to 10. After the
124 reaction, the magnetic materials in the suspension were separated from the liquid phase using
125 an external magnet. The magnetized samples were collected, dried in a vacuum oven at 60°C,
126 weighed, and designated as magnetized straw (MCS) and magnetized biochar (MBc).

127

128 2.2. Experimental Design

129 Prior to the incubation experiment, the collected soil was pretreated by thorough mixing
130 and sieving through a 2 mm mesh. Fine roots and other visible plant residues were carefully
131 removed, and all iron-containing particles were extracted using a magnetic rod to minimize
132 potential experimental interference. The study included five treatments: (1) control (CK): no



133 organic amendment; (2) straw treatment (CS): non-magnetized straw; (3) biochar treatment
134 (Bc): non-magnetized biochar with carbon content equivalent to CS; (4) magnetized straw
135 treatment (MCS): magnetized straw at the same carbon input as CS; and (5) magnetized
136 biochar treatment (MBc): magnetized biochar with carbon equivalent to CS.

137 For each treatment, the respective materials (CS, Bc, MCS, and MBc) were thoroughly
138 mixed with soil. Specifically, 400 g of soil was placed in PVC containers, and amendments
139 were applied based on a full straw return rate of 11 t ha⁻¹. Accordingly, 1.95 g of straw was
140 added to the CS treatment, while 1.16 g of biochar was applied to the Bc treatment to match
141 the carbon input of straw. The amount of magnetized straw and magnetized biochar were
142 adjusted according to their preparation yields, resulting in 2.67 and 1.63 g for the MCS and
143 MBc treatments, respectively.

144 To ensure homogeneous mixing, a small portion of amendment and air-dried soil was
145 first combined in a glass beaker using a plastic spoon. After thorough mixing, the remaining
146 soil was gradually added and continuously mixed until a uniform soil amendment mixture
147 was obtained (Shi et al., 2024). The CK, CS, and Bc treatments each included three replicates.
148 Each of the MCS and MBc treatments had six replicates divided into two subgroups: three
149 without magnetic residue separation (MCS-O and MBc-O) and three with magnetic residue
150 separation prior to soil and parameter analyses (MCS-D and MBc-D). During incubation, soil
151 moisture was maintained at 25% by frequent weighing and watering, and all samples were
152 incubated at 30°C in a constant-temperature incubator.

153 The incubation began in April 2024, with destructive sampling at 30, 60, 180, and 360 d
154 after the start. The samples from each treatment were retrieved, air-dried, and sieved through



155 a 2 mm mesh for subsequent analyses.

156

157 2.3. Fractionation of POM and MAOM

158 SOM was fractionated into particulate organic matter (POM) and mineral-associated
159 organic matter (MAOM) following the wet sieving and particle-size fractionation methods
160 described by Cambardella and Elliott (1992). Specifically, 20 g of air-dried soil was weighed
161 and mixed with 60 mL of 5 g L⁻¹ sodium hexametaphosphate solution. The mixture was
162 shaken for 18 h at 25°C and 180 rpm. The dispersed suspension was then passed through a 53
163 µm sieve and washed repeatedly with small volumes of deionized water until the filtrate
164 became clear and colorless. The material retained on the sieve (>53 µm) was considered as
165 sand particles and POM, while the fraction passing through the sieve (<53 µm) consisted of
166 silt- and clay-sized particles along with MAOM.

167 The POM and MAOM fractions were collected separately in glass beakers. Within the
168 soil-water suspension, a strong external magnetic iron rod was used to separate
169 undecomposed magnetized straw and magnetized biochar residues from the POM and
170 MAOM fractions in liquid form. These separated materials were designated as magnetized
171 residue components within the POM and MAOM fractions, respectively. Both the soil
172 fractions and magnetized residue fractions of POM and MAOM were dried at 60°C, weighed,
173 and ground through a 60-mesh sieve (Liu et al., 2024). After the complete removal of
174 undecomposed magnetized organic residues, the soil organic matter (SOC) content of the
175 original soil and each fraction was determined using the potassium dichromate oxidation
176 method with external heating (Nelson and Sommers, 1982). The organic carbon content of



organic residue samples and soil fractions at different incubation times was measured using an elemental analyzer (Vario EL III, Hanau, Germany). Data were corrected on an ash-free and moisture-free basis (Ndzelu et al., 2021). The concentration of particulate organic carbon (POC) expressed in g kg^{-1} was calculated as follows:

$$POC = M_p \times OC_p \quad (1)$$

$$POM (\%) = POC / SOC \times 100 \quad (2)$$

where M_p denotes the relative mass proportion of the POM fraction (%), OC_p represents the organic carbon content of the POM fraction (g kg^{-1}), and SOC denotes the organic carbon content of the undisturbed soil in soil samples (g kg^{-1}).

2.4. Calculation of organic residue retention rate

The cumulative retention rate of dry matter from straw (CS) and biochar (Bc) residues within the MCS and MBc fractions was calculated as follows:

$$L(\%) = \frac{M_d - M_{Fe}}{M_1} \times 100\% \quad (3)$$

where L is the mass retention rate of the organic residue (%), M_d is the dry mass of the recovered magnetized material at different decomposition times (g), M_{Fe} is the dry mass of the Fe-related products in the applied magnetized material (g), and M_1 is the dry mass of the applied straw or biochar (g).

2.5. Data analysis

All data were preprocessed using Microsoft Excel 2022. Statistical analyses were performed using IBM SPSS Statistics 25 (IBM Corporation, Armonk, NY, USA).



199 Differences between treatments were assessed using Duncan's multiple range test at a
200 significance level of $p < 0.05$. Graphs were generated using Origin 2022.

201

202 **3. Results and analysis**

203 *3.1. Differences between magnetized and original organic materials*

204 As shown in Table 1, no significant differences were observed in the molar ratios of
205 carbon to nitrogen (C/N), hydrogen to carbon (H/C), or oxygen to carbon (O/C) before and
206 after magnetic modification of the organic materials. This result indicates that the
207 magnetization process does not substantially alter the elemental composition of organic
208 materials. The C/N, H/C, and O/C ratios serve as key indicators of the chemical properties
209 and structural characteristics of organic materials. Specifically, the C/N ratio could be closely
210 associated with the decomposition rate of organic substrates, the H/C ratio reflects
211 aromaticity, and the O/C ratio represents the oxidation level (Ndzelu et al., 2021). The
212 near-constant values of these ratios before and after magnetization suggest that the
213 fundamental chemical attributes and structural features of the organic materials remained
214 unchanged by magnetic treatment.

215 Therefore, the magnetized organic materials exhibited high chemical and structural
216 consistency with their non-magnetized counterparts, making them reliable representatives of
217 original organic substrates. This conclusion could present a sound theoretical basis for using
218 magnetized organic materials in subsequent experiments to investigate the behavior of
219 undecomposed organic residues in soil. It also ensured the reliability and accuracy of results
220 related to the assessment of "false increases" in soil organic matter through the separation of



221 magnetized organic residues.

222

223 **Table 1** Comparison of elemental composition of organic materials before and after
 224 magnetization.

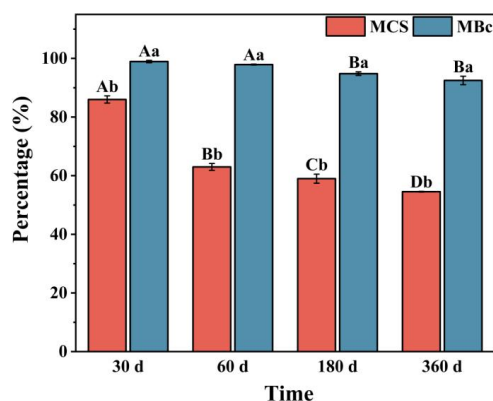
Treatment	N	C	H	O	C/N	H/C	O/C
	%				Ratio	Ratio	Ratio
CS	0.99	46.20	6.84	45.96	54.32	1.78	0.75
Bc	0.78	79.35	3.65	16.22	118.49	0.55	0.15
MCS	1.00	46.68	6.55	45.77	54.43	1.68	0.74
MBc	0.80	79.37	3.54	16.30	115.65	0.53	0.15

225

226 3.2. Temporal changes in magnetized organic residues in soil

227 As shown in Fig. 1, the retention rate of straw residues in MCS gradually decreased over
 228 the incubation period, with values of 85.98%, 63.00%, 58.99%, and 54.55% at 30, 60, 180,
 229 and 360 d, respectively. In contrast, the biochar fraction in MBc exhibited relatively minor
 230 changes, with retention rates of 98.92%, 97.88%, 94.80%, and 92.48% at the corresponding
 231 time points. These results indicate that the straw component in MCS underwent
 232 decomposition in soil and, with increasing incubation time, was eventually transformed into
 233 stabilized organic matter. Conversely, the biochar fraction in MBc demonstrated high
 234 stability and low decomposition in soil, suggesting that it could not represent truly
 235 decomposed organic matter.

236



237

238 **Fig. 1.** Residual rates of undecomposed magnetized straw (MCS) and magnetized biochar
 239 (MBc) separated from soil at different incubation times.

240 Note: MCS refers to magnetized straw; MBc refers to magnetized biochar. Different
 241 uppercase letters indicate significant differences among sampling times within the same
 242 organic residue ($p < 0.05$), while different lowercase letters indicate significant differences
 243 between organic residues at the same sampling time ($p < 0.05$).

244

245 The results in Fig. 2 further demonstrate that the two types of magnetized organic
 246 materials differed not only in their retention rates but also in the extent of mass changes after
 247 decomposition. Compared with the soil MAOM fraction, undecomposed MCS residues in the
 248 early incubation stage exhibited higher H/C and C/N ratios, closer to those of the soil POM
 249 fraction. As shown in Fig. 2a, the H/C ratio of MCS residues decreased gradually over time,
 250 approaching that of MAOM by day 360, whereas the O/C ratio exhibited a slow increase. In
 251 contrast, these trends were not evident in MBc residue samples. A decrease in the H/C ratio
 252 indicates a reduced aliphatic character of the organic residues (Banach-Szott et al., 2014; Dou
 253 and Li, 2010), while an increase in the O/C ratio could reflect increased oxidation



(Mohammed et al., 2023). These findings suggest that with prolonged incubation, the molecular conjugation of the organic residues increased, accompanied by changes in the electronic environment of oxygen-containing functional groups, thereby enhancing aromatic structural features. Fig. 2b illustrates that the C/N ratio and carbon concentration of MCS residues declined continuously, gradually approaching those of the soil MAOM fraction. This trend reflects the progressive humification of organic residues (Abakumov and Eskov, 2023), indicating an increasing degree of humification. Notably, the C/N ratio of MCS residues at 360 d approximated that of MAOM. Conversely, the changes in these parameters for MBc residues were relatively small, indicating that the organic components of MBc residues are more stable and less prone to decomposition and transformation.

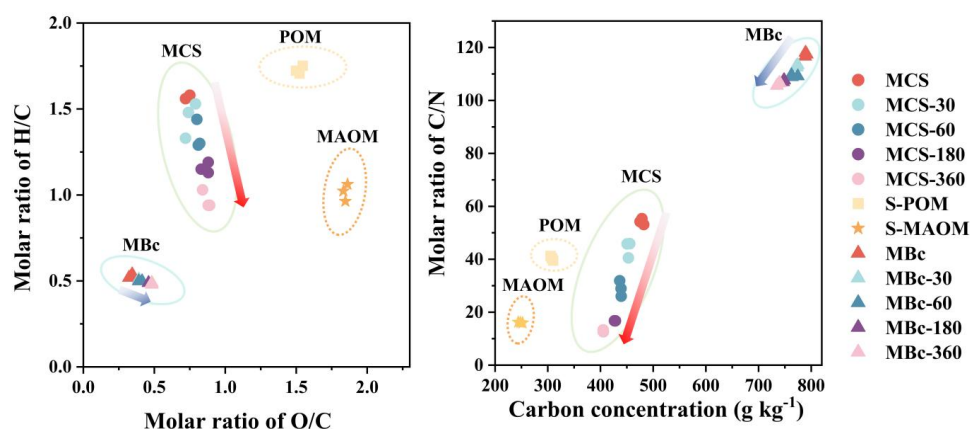


Fig. 2. Van Krevelen diagram of atomic H/C and O/C ratios (a), and comparison of C/N ratio and carbon concentration (b) of magnetized organic residues, soil POM, and MAOM fractions at different incubation times.

Note: MCS refers to magnetized straw; MBc refers to magnetized biochar. MCS-30, MCS-60, MCS-180, and MCS-360 represent undecomposed magnetized straw residues separated from



271 soil at 30, 60, 180, and 360 days, respectively; MBc-30, MBc-60, MBc-180, and MBc-360
 272 represent undecomposed magnetized biochar residues separated from soil at the same
 273 respective time points. S-POM and S-MAOM denote soil samples of particulate organic
 274 matter and mineral-associated organic matter fractions, respectively. All data are corrected on
 275 an ash-free and moisture-free basis.

276

277 *3.3. Organic residues cause false increases in the mass proportion and organic carbon* 278 *content of soil POM fraction*

279 Fig. 3 presents the relative mass proportions of the soil POM fraction at different
 280 incubation times across treatments. Clear differences among treatments were evident.
 281 Notably, no significant differences were observed between the MCS-O and MBc-O
 282 treatments and the CS and Bc treatments at any sampling time, supporting the applicability of
 283 magnetized materials in soil applications.

284 At 30, 60, 180, and 360 d of incubation, the mass proportion of the POM fraction (M_P)
 285 in the CS treatment increased by 18.94%, 11.97%, 8.78%, and 7.05%, respectively, compared
 286 with CK. For the Bc treatment, the corresponding increases were 17.22%, 17.16%, 16.95%,
 287 and 16.83%, respectively. However, after the removal of magnetized organic residues from
 288 the soil, no significant changes in POM mass proportion were observed in the MCS-D and
 289 MBc-D treatments compared with CK. These results indicate that the increases in POM mass
 290 proportion in the CS and Bc treatments were “artificial elevations” caused by undecomposed
 291 organic residues. Further analysis revealed that in the CS treatment, this artificial elevation
 292 decreased gradually over time, stabilizing at approximately day 180, whereas in the Bc



293 treatment, it remained nearly constant throughout the incubation period.

294 Similarly, at different time points, the organic carbon content of the POM fraction (OC_P)

295 in the CS treatment increased by 37.87%, 26.99%, 15.94%, and 7.92%, respectively,

296 compared with CK. For the Bc treatment, the increases were 35.86%, 33.83%, 31.93%, and

297 31.10%, respectively. At the same time points, the OC_P in the CS treatment exceeded that in

298 the MCS-D treatment by 37.68%, 26.53%, 15.25%, and 7.48%, respectively. Moreover, the

299 OC_P in the Bc treatment was higher than in the MBc-D treatment by 35.80%, 33.96%,

300 31.93%, and 31.10%. These results demonstrated that both the carbon content and mass

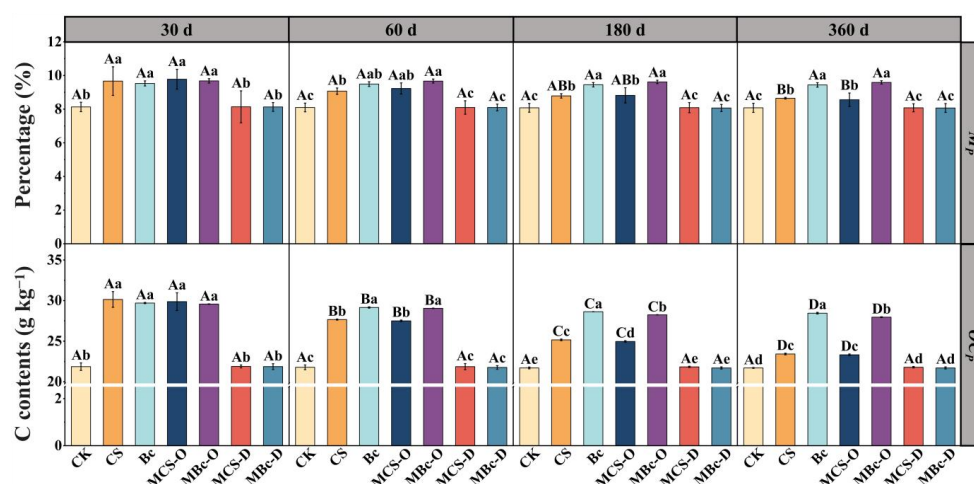
301 proportion of the POM fraction exhibited a “false increase”. As no significant difference was

302 identified between the MCS-D and CK treatments, the reduction in the false increase for the

303 CS treatment was attributable not to an increase in organic carbon content in the MCS-D

304 treatment but to the decomposition of organic residues within the fraction.

305



306

307 **Fig. 3.** Relative mass proportion (M_p) and organic carbon content (OC_P) of the soil POM

308 fraction at different incubation times across treatments.



309 Note: CK denotes the control treatment without organic amendments; CS denotes the
 310 treatment with normal straw application; Bc denotes the treatment with biochar applied at an
 311 equivalent carbon amount to straw; MCS denotes the treatment with magnetized straw
 312 applied at an equivalent carbon amount; and MBc denotes the treatment with magnetized
 313 biochar applied at an equivalent carbon amount. MCS-O and MBc-O refer to treatments in
 314 which magnetized organic materials were not removed at the end of incubation; MCS-D and
 315 MBc-D refer to treatments in which magnetized organic residues were separated from the soil
 316 before testing the remaining soil samples. Different uppercase letters indicate significant
 317 differences among sampling times within the same treatment, whereas different lowercase
 318 letters indicate significant differences among treatments at the same sampling time ($p < 0.05$).

319

320 3.4. Organic residues cause false increases in POC and SOC contents

321 As shown in Fig. 4, the *POC* contents in the MCS-O and MBc-O treatments were
 322 slightly lower than those in the CS and Bc treatments, although the differences were not
 323 statistically significant. Specifically, at 30, 60, 180, and 360 d of incubation, the *POC*
 324 contents in both the CS and Bc treatments were significantly higher than those in CK. The
 325 increase in *POC* content compared with CK in the CS treatment showed a clear decreasing
 326 trend, with increases of 63.48%, 42.61%, 26.29%, and 16.00%, respectively. In contrast,
 327 although the *SOC* content in the Bc treatment also declined during incubation, the decrease
 328 was less pronounced, with increases of 58.99%, 57.38%, 54.86%, and 53.71% at the
 329 respective time points. No significant differences in *POC* content were observed between the
 330 MCS-D and MBc-D treatments, indicating that the elevated *POC* contents in the CS and Bc



331 treatments originated from undecomposed organic residues. The *POC* contents in the CS
332 treatment exceeded that in the MCS-D treatment by 63.48%, 41.80%, 24.86%, and 15.34% at
333 the respective time points, whereas the *POC* contents in the Bc treatment were higher than
334 that in the MBc-D treatment by 58.99%, 42.61%, 54.86%, and 53.71%, respectively.

335 Across all incubation periods, the *SOC* contents in the MCS-O and MBc-O treatments
336 were comparable to those in the CS and Bc treatments, indicating strong consistency between
337 the magnetized organic materials and the original organic materials during incubation.
338 Specifically, at 30, 60, 180, and 360 d, the *SOC* contents in the CS and Bc treatments were
339 significantly higher than in CK. The *SOC* content in the CS treatment showed a decreasing
340 trend, with increases of 11.95%, 8.40%, 5.71%, and 4.50%, respectively. In contrast,
341 although the *SOC* content in the Bc treatment also declined over time, the decrease was less
342 pronounced, with increases of 12.41%, 12.35%, 10.89%, and 11.06% at the respective time
343 points. After 360 d of incubation, the *SOC* content in the MCS-D treatment showed an
344 increasing trend, whereas the *SOC* content in the MBc-D treatment remained similar to that
345 in CK without significant changes, maintaining a relatively stable level throughout the
346 incubation period.

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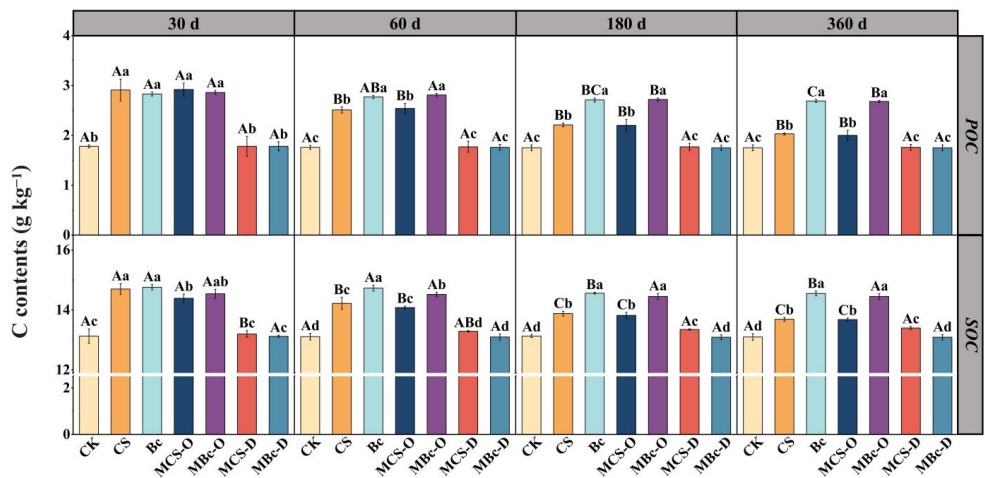


Fig. 4. *POC* and *SOC* contents of different treatments at various incubation times.

Note: CK denotes the control treatment without organic amendments; CS denotes the treatment with normal straw application; Bc denotes the treatment with biochar applied at an equivalent carbon amount to straw; MCS denotes the treatment with magnetized straw applied at an equivalent carbon amount; and MBc denotes the treatment with magnetized biochar applied at an equivalent carbon amount. MCS-O and MBc-O refer to treatments in which magnetized organic materials were not removed at the end of incubation; MCS-D and MBc-D refer to treatments in which magnetized organic residues were separated from the soil before testing the remaining soil samples. Different uppercase letters indicate significant differences among sampling times within the same treatment, whereas different lowercase letters indicate significant differences among treatments at the same sampling time ($p < 0.05$).

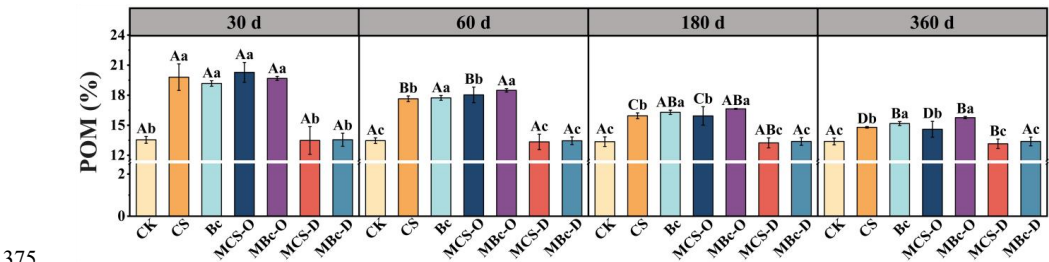
3.5. Organic residues cause false increases in the proportion of POM in total SOM

As shown in Fig. 5, the proportion of POM to total SOM (POC/SOC) in the CS and Bc treatments was significantly higher than that in CK. At different incubation times, the ratios



364 in the CS treatment increased by 46.23%, 31.05%, 19.31%, and 10.54%, whereas those in the
365 Bc treatment increased by 41.58%, 39.60%, 39.15%, and 38.04%, respectively. However, the
366 false elevation of the POM proportion in the SOM persisted. At different time points, the
367 ratios in the CS treatment were 46.88%, 32.23%, 20.39%, and 12.47% higher than those in
368 the MCS-D treatment, whereas in the Bc treatment, they were 41.47%, 39.70%, 38.94%, and
369 38.04% higher than those in the MBc-D treatment. Notably, the POM proportion in SOM in
370 the MCS-D treatment was lower than that in CK, which was attributed to the greater
371 conversion of straw residues into MAOM during decomposition, thereby increasing the
372 MAOM proportion in SOM. In contrast, no significant changes were observed in biochar in
373 the MBc-D treatment, and the proportion of POM in SOM was similar to that in CK.

374



375
376 **Fig. 5.** Proportion of POM content in total SOM for different treatments at various incubation
377 times.

378 Note: CK denotes the control treatment without organic amendments; CS denotes the
379 treatment with normal straw application; Bc denotes the treatment with biochar applied at an
380 equivalent carbon amount to straw; MCS denotes the treatment with magnetized straw
381 applied at an equivalent carbon amount; and MBc denotes the treatment with magnetized
382 biochar applied at an equivalent carbon amount. MCS-O and MBc-O refer to treatments in



383 which magnetized organic materials were not removed at the end of incubation; MCS-D and
384 MBc-D refer to treatments in which magnetized organic residues were separated from the soil
385 before testing the remaining soil samples. Different uppercase letters indicate significant
386 differences among sampling times within the same treatment, whereas different lowercase
387 letters indicate significant differences among treatments at the same sampling time ($p < 0.05$).

388

389 **4. Discussion**

390 *4.1. Differences in the transformation of organic materials with different qualities in soil*

391 As shown in Fig. 1, the proportion of undecomposed straw residues significantly
392 decreased over different incubation periods, with a residue rate of only 54.55% after 360 d. In
393 sharp contrast, biochar exhibited almost no decomposition, maintaining a high residue rate of
394 92.48%. Fig. 2 indicates that the molecular conjugation degree of magnetic straw residues
395 was enhanced, with changes occurring in the electronic environment of oxygen-containing
396 functional groups, which can strengthen the aromatic structure and stabilize the framework.
397 The composition of organic residues may gradually shift towards humification (Yu et al.,
398 2022), whereas MBc residues exhibit negligible changes throughout the incubation period.
399 These outcomes can be attributed to the lower C/N ratio of straw compared with the
400 relatively higher C/N ratio of biochar, which can cause greater reactivity and a faster
401 mineralization rate, enabling straw to decompose more rapidly in soil (Chen et al., 2023).
402 Chen et al. (2010) reported that after straw incorporation into soil, labile components rapidly
403 decomposed within 0–60 d, followed by slower decomposition during 60–180 d, which can
404 be consistent with the present findings. This could be because during the early decomposition



405 stage, highly biodegradable components such as carbohydrates, organic acids, and amino
406 acids are extensively consumed, while more recalcitrant components, such as aromatic and
407 polymeric fractions, are retained (Ren et al., 2021). On the other hand, biochar is produced
408 via pyrolysis of biomass at 300–700°C under high-pressure anaerobic conditions (Dungait et
409 al., 2012). Through pyrolysis, the labile cellulose-C in straw can be converted into aromatic
410 biochar-C, greatly enhancing its structural stability and resistance to rapid decomposition in
411 soil (Yin et al., 2022), thus remaining largely as a residue (Bornø et al., 2019). As a
412 carbon-rich soil amendment with slow decomposition and a surface conducive to organic
413 molecule aggregation, biochar offers unique advantages for soil carbon sequestration and
414 structural improvement (Cao et al., 2022; Fan et al., 2021; Wang et al., 2025; Zhang et al.,
415 2024). When applied to soil, it can reduce bulk density (Zhang et al., 2021), increase porosity
416 (He et al., 2022), alleviate acidification (Shi et al., 2023), retain moisture (Khaledi et al.,
417 2023), and enhance nutrient absorption efficiency and nutrient cycling coordination (Burgeon
418 et al., 2022). Owing to its porosity, alkalinity, strong adsorption capacity, and large specific
419 surface area, biochar has been widely applied for environmental remediation and sustainable
420 agricultural production (Papageorgiou et al., 2021; Pathy et al., 2023; Xu et al., 2021).
421 Therefore, despite its slow decomposition in soil, its unique role in improving soil structure
422 and function remains significant.

423 *4.2. Residual undecomposed organic matter leads to a false increase in the POM fraction*

424 The analysis of the weight proportion and organic carbon content of soil fractions
425 presented in Figs. 3 and 4 clearly demonstrated that under the CS and Bc treatments, both the
426 relative mass and organic carbon content of the POM fraction were consistently higher than



427 those in the CK treatment. This finding aligns with those reported by Xie et al. (2014). The
428 POM fraction mainly consists of partially decomposed, chemically recalcitrant polymeric
429 structures, such as acid-insoluble fibers formed through fragmentation, which could primarily
430 originate from organic materials. Owing to its rapid responsiveness to environmental changes,
431 POM can be highly sensitive to agricultural management practices (Christensen, 1992;
432 Cotrufo et al., 2022; Guo et al., 2022; Rocci et al., 2021; Witzgall et al., 2021). Xie et al.
433 demonstrated that increasing the input of organic materials directly influenced both SOM
434 content and its proportion within the POM fraction. They attributed this phenomenon to the
435 continuous accumulation of organic residues in soil induced by organic amendments (Xie et
436 al., 2014). However, residual undecomposed organic residues should not be equated with true
437 organic matter, as treating these residues as part of SOM can lead to biased results and
438 misinterpretation.

439 In this experiment, the POM mass proportion and organic carbon content for the MCS-D
440 and MBc-D treatments were obtained by first applying the magnetic materials to the soil for a
441 period of incubation, then extracting the magnetic residues from the soil, and subsequently
442 testing the soil samples after removal of the undecomposed materials. The results showed that
443 after the magnetic materials were extracted, the POM mass proportion and organic carbon
444 content in the MCS and MBc treatments did not exhibit significant increases compared with
445 the CK treatment. This proved that the increases in POM mass proportion and organic carbon
446 content observed under the CS and Bc treatments were largely attributable to the direct input
447 of straw and biochar materials, with most undecomposed organic residues remaining within
448 the POM fraction. Moreover, the “false elevation” in both POM mass proportion and organic



449 carbon content under the CS treatment decreased over the incubation period, whereas the
450 corresponding values under the Bc treatment remained nearly constant. These findings
451 suggest that the quantity, quality, and incubation duration of organic residues are key factors
452 driving the increase in the POM mass proportion and organic carbon content. Additionally,
453 the extent of increase in the POM fraction was closely related to the amount and source of
454 organic material applied.

455 The results shown in Fig. 4 revealed a pronounced decreasing trend in the POM fraction
456 organic carbon content (*POC*) under CS treatment. This confirmed that the effect of organic
457 material addition in the short term was predominantly reflected in the POM fraction, whereas
458 a gradual increase in the MAOM fraction was observed. This aligned with the conclusions of
459 Bhattacharyya et al. (2011), Brown et al. (2014), and Stewart et al. (2012) who reported that
460 organic amendments were primarily retained in the POM fraction, which could be more
461 prone to mineralization, while gains in the MAOM fraction remained limited. The MAOM
462 fraction in soil has been mainly formed over decades to centuries through long-term
463 weathering processes involving interactions between organic matter and secondary minerals.
464 Due to this extremely slow formation process, MAOM accumulation can be difficult to
465 achieve in the short term (Kleber et al., 2007; Slessarev et al., 2022). Moreover, because
466 microorganisms struggle to utilize chemically recalcitrant components within plant residues,
467 decomposition of these highly processed structural organic residues and POM components
468 has been reported to cause MAOM formation (Cotrufo et al., 2015). This explains why the
469 organic carbon data for the Bc treatment in this study (Figs. 3 and 4) indicated that most
470 undecomposed organic residues remained preserved within the POM fraction, thereby



471 reducing the MAOM contribution to the soil. This also accounted for the consistently higher
472 MAOM contribution observed in the CS and Bc treatments than in the MCS-D and MBc-D
473 treatments.

474 Currently, some studies have suggested that abundant POM can be crucial for
475 agroecosystem functioning and crop productivity, thereby advocating for greater research
476 focusing on POM increments (Wood et al., 2016). However, the results of this study
477 indicated that within the POM fraction, the dominant influencing factors were the quantity
478 and quality of undecomposed organic residues, with temporal factors exerting a significant
479 impact. Although the POM fraction plays an important role in nutrient supply, microbial
480 activity promotion, and soil structure regulation, the indiscriminate addition of organic
481 materials to soil primarily increases the amount of undecomposed organic residues, most of
482 which reside in the POM fraction over short time periods. While this practice directly raises
483 the measured SOM content, it represents a superficial increase, as the majority of SOM gain
484 is derived from unstable undecomposed organic residues rather than stabilized organic matter.
485 Considering the POM fraction as part of SOM without distinction carries inherent risks, as
486 this “false elevation” can lead to the overestimation of global SOM content.

487

488 **5. Conclusion**

489 Based on the comprehensive results of this study, magnetic treatment exerted minimal
490 influence on the elemental composition of organic materials, indicating that magnetized
491 organic materials can serve as valid representatives of normal organic materials and that the
492 related experimental outcomes are reliable. The straw component in MCS decomposed



493 readily in soil, with its residue rate markedly decreasing during the incubation period. The
494 H/C ratio of the residues decreased, the O/C ratio increased, and both the C/N ratio and
495 carbon concentration decreased continuously, indicating reduced aliphaticity, enhanced
496 oxidation, and a molecular structural shift toward increased aromaticity and greater
497 humification. Furthermore, the proportion of residues in the POM fraction sharply declined
498 with incubation time, approaching the characteristics of MAOM after 360 d. In contrast, the
499 biochar component in MBc exhibited high stability in soil, showing minor changes in the
500 residue rate, elemental ratios, and a relatively gradual decline in the proportion of the POM
501 fraction, reflecting greater resistance to decomposition. After organic material incorporation,
502 the unseparated residues in the CS and Bc treatments resulted in higher POM organic carbon
503 content and a larger proportion of total SOM than in CK. However, this increase represented
504 a “false elevation”, most of which pronounced during the early incubation period. After
505 residue separation, the MCS-D and MBc-D treatments displayed little difference from CK,
506 indicating that the “false elevation” was primarily attributable to incompletely decomposed
507 residues. On day 30, the false increases in *POC* content reached 63.48% and 58.99% for the
508 CS and Bc treatments, respectively. Over time, the false elevation in CS gradually diminished,
509 decreasing to 15.34% after 360 d, whereas the false elevation in Bc remained largely
510 unchanged and stable, still reaching 53.71% after 360 d. Owing to the greater conversion of
511 straw residues into MAOM during decomposition, the POM fraction contribution in MCS-D
512 was lower than that in CK. Concurrently, biochar in MBc-D exhibited no significant change,
513 with a POM contribution comparable to that of CK. These results confirmed the risk of
514 overestimation of SOM when the POM fraction was fully counted, particularly because the



515 long-term persistence of recalcitrant organic materials can cause systematic bias in global
516 SOM content assessments. These findings could provide important references for the precise
517 evaluation of soil organic matter transformation processes and contents.

518

519 **Conflicts of Interest:** The authors declare no conflict of interest.

520

521 **CRedit authorship contribution statement :** Yuhan Xia: Writing – original draft,
522 Visualization, Validation, Methodology, Investigation, Formal analysis, Data curation,
523 Conceptualization. Sen Dou: Writing – review & editing, Supervision, Resources, Project
524 administration, Methodology, Funding acquisition, Conceptualization. Guan Song: Writing –
525 review & editing, Supervision, Resources, Project administration, Methodology, Funding
526 acquisition, Conceptualization. Dilimulati Yalikhong: Writing – review & editing,
527 Methodology, Investigation.

528

529 **Funding:** This work was supported by the National Key Research and Development Program
530 of China (2024YFD1500502-04).

531

532 **References**

533 Abakumov, E., Eskov, A., 2023. Organic matter structural composition of vascular epiphytic
534 suspended soils of South Vietnam. *Applied Sciences* 13, 4473.
535 <https://doi.org/10.3390/app13074473>.
536 Angst, G., Mueller, K.E., Nierop, K.G.J., Simpson, M.J., 2021. Plant- or microbial-derived?
537 A review on the molecular composition of stabilized soil organic matter. *Soil Biology*



- 538 and Biochemistry 156, 108189. <https://doi.org/10.1016/j.soilbio.2021.108189>.
- 539 Banach-Szott, M., Debska, B., Rosa, E., 2014. Effect of soil pollution with polycyclic
 540 aromatic hydrocarbons on the properties of humic acids. Journal of Soils and Sediments
 541 14, 1169–1178. <https://doi.org/10.1007/s11368-014-0873-9>.
- 542 Baragaño, D., Alonso, J., Gallego, J.R., Lobo, M.C., Gil-Díaz, M., 2020. Magnetite
 543 nanoparticles for the remediation of soils co-contaminated with As and PAHs. Chemical
 544 Engineering Journal 399, 125809. <https://doi.org/10.1016/j.cej.2020.125809>.
- 545 Bhattacharyya, R., Kundu, S., Srivastva, A.K., Gupta, H.S., Prakash, V., Bhatt, J.C., 2011.
 546 Long term fertilization effects on soil organic carbon pools in a sandy loam soil of the
 547 Indian sub-Himalayas. Plant and Soil 341, 109–124.
 548 <https://doi.org/10.1007/s11104-010-0627-4>.
- 549 Bornø, M.L., Müller-Stöver, D.S., Liu, F.L., 2019. Biochar properties and soil type drive the
 550 uptake of macro- and micronutrients in maize (*Zea mays* L.). Journal of Plant Nutrition
 551 and Soil Science 182, 149–158. <https://doi.org/10.1002/jpln.201800228>.
- 552 Brown, K.H., Bach, E.M., Drijber, R.A., Hofmockel, K.S., Jeske, E.S., Sawyer, J.E.,
 553 Castellano, M.J., 2014. A long-term nitrogen fertilizer gradient has little effect on soil
 554 organic matter in a high-intensity maize production system. Global Change Biology 20,
 555 1339–1350. <https://doi.org/10.1111/gcb.12519>.
- 556 Burgeon, V., Fouché, J., Garré, S., Dehkordi, R.H., Colinet, G., Cornelis, J.T., 2022. Young
 557 and century-old biochars strongly affect nutrient cycling in a temperate agroecosystem.
 558 Agriculture, Ecosystems & Environment 328, 107847.
 559 <https://doi.org/10.1016/j.agee.2021.107847>.



- 560 Cambardella, C.A., Elliott, E.T., 1992. Particulate soil organic-matter changes across a
 561 grassland cultivation sequence. *Soil Science Society of America Journal* 56, 777–783.
 562 <https://doi.org/10.2136/sssaj1992.03615995005600030017x>.
- 563 Cao, L.Y., Zhang, X.Y., Xu, Y., Xiang, W., Wang, R., Ding, F.J., Hong, P.Z., Gao, B., 2022.
 564 Straw and wood based biochar for CO₂ capture: Adsorption performance and governing
 565 mechanisms. *Separation and Purification Technology* 287, 120592.
 566 <https://doi.org/10.1016/j.seppur.2022.120592>.
- 567 Castellano, M.J., Mueller, K.E., Olk, D.C., Sawyer, J.E., Six, J., 2015. Integrating plant litter
 568 quality, soil organic matter stabilization, and the carbon saturation concept. *Global*
 569 *Change Biology* 21, 3200–3209. <https://doi.org/10.1111/gcb.12982>.
- 570 Chen, H.L., Zhou, J.M., Xiao, B.H., 2010. Characterization of dissolved organic matter
 571 derived from rice straw at different stages of decay. *Journal of Soils and Sediments* 10,
 572 915–922. <https://doi.org/10.1007/s11368-010-0210-x>.
- 573 Chen, L.M., Sun, S.L., Zhou, Y.Y., Zhang, B.X., Peng, Y.T., Zhuo, Y.C., Ai, W.K., Gao, C.F.,
 574 Wu, B., Liu, D.W., Sun, C.R., 2023. Straw and straw biochar differently affect fractions
 575 of soil organic carbon and microorganisms in farmland soil under different water
 576 regimes. *Environmental Technology & Innovation* 32, 103412.
 577 <https://doi.org/10.1016/j.eti.2023.103412>.
- 578 Christensen, B.T., 1992. Physical fractionation of soil and organic matter in primary particle
 579 size and density separates. In: Stewart, B.A. (Ed.), *Advances in Soil Science*. Springer
 580 New York, New York, NY, pp. 1–90. https://doi.org/10.1007/978-1-4612-2930-8_1.
- 581 Connell, R.K., James, T.Y., Blesh, J., 2025. A legume-grass cover crop builds



582 mineral-associated organic matter across variable agricultural soils. *Soil Biology and*
 583 *Biochemistry* 203, 109726. <https://doi.org/10.1016/j.soilbio.2025.109726>.

584 Cotrufo, M.F., Haddix, M.L., Kroeger, M.E., Stewart, C.E., 2022. The role of plant input
 585 physical-chemical properties, and microbial and soil chemical diversity on the formation
 586 of particulate and mineral-associated organic matter. *Soil Biology and Biochemistry* 168,
 587 108648. <https://doi.org/10.1016/j.soilbio.2022.108648>.

588 Cotrufo, M.F., Soong, J.L., Horton, A.J., Campbell, E.E., Haddix, M.L., Wall, D.H., Parton,
 589 W.J., 2015. Formation of soil organic matter via biochemical and physical pathways of
 590 litter mass loss. *Nature Geoscience* 8, 776–779. <https://doi.org/10.1038/ngeo2520>.

591 Cotrufo, M.F., Wallenstein, M.D., Boot, C.M., Deneff, K., Paul, E., 2013. The Microbial
 592 Efficiency- Matrix Stabilization (MEMS) framework integrates plant litter
 593 decomposition with soil organic matter stabilization: Do labile plant inputs form stable
 594 soil organic matter? *Global Change Biology* 19, 988–995.
 595 <https://doi.org/10.1111/gcb.12113>.

596 Dou, S., Li, K., 2010. Effect of organic matter application on CP-MAS-13C-NMR spectra of
 597 humic acids from a brown soil. In: Xu, J.M., Huang, P.M. (Eds.), *Molecular*
 598 *Environmental Soil Science at the Interfaces in the Earth's Critical Zone*. Springer
 599 Berlin Heidelberg, Berlin, Heidelberg, pp. 29–31.
 600 https://doi.org/10.1007/978-3-642-05297-2_9.

601 Dou, S., Shan, J., Song, X.Y., Cao, R., Wu, M., Li, C.L., Guan, S., 2020. Are humic
 602 substances soil microbial residues or unique synthesized compounds? A perspective on
 603 their distinctiveness. *Pedosphere* 30, 159–167.



- 604 [https://doi.org/10.1016/S1002-0160\(20\)60001-7](https://doi.org/10.1016/S1002-0160(20)60001-7).
- 605 Duan, L.C., Wang, Q.H., Li, J.N., Wang, F.H., Yang, H., Guo, B.L., Hashimoto, Y., 2022.
- 606 Zero valent iron or Fe₃O₄-loaded biochar for remediation of Pb contaminated sandy soil:
- 607 Sequential extraction, magnetic separation, XAFS and ryegrass growth. *Environmental*
- 608 Pollution 308, 119702. <https://doi.org/10.1016/j.envpol.2022.119702>.
- 609 Dungait, J.A.J., Hopkins, D.W., Gregory, A.S., Whitmore, A.P., 2012. Soil organic matter
- 610 turnover is governed by accessibility not recalcitrance. *Global Change Biology* 18,
- 611 1781–1796. <https://doi.org/10.1111/j.1365-2486.2012.02665.x>.
- 612 Fan, Y.V., Klemes, J.J., Lee, C.T., 2021. Environmental performance and techno-economic
- 613 feasibility of different biochar applications: An overview. *Chemical Engineering*
- 614 Transactions 83, 469–474. <https://doi.org/10.3303/CET2183079>.
- 615 Feng, H.L., Han, X.Z., Biswas, A., Zhang, M., Zhu, Y.C., Ji, Y.X., Lu, X.C., Chen, X., Yan, J.,
- 616 Zou, W.X., 2025. Long-term organic material application enhances black soil
- 617 productivity by improving aggregate stability and dissolved organic matter dynamics.
- 618 *Field Crops Research* 328, 109946. <https://doi.org/10.1016/j.fcr.2025.109946>.
- 619 Guo, X.W., Viscarra Rossel, R.A., Wang, G.C., Xiao, L.J., Wang, M.M., Zhang, S., Luo, Z.K.,
- 620 2022. Particulate and mineral-associated organic carbon turnover revealed by modelling
- 621 their long-term dynamics. *Soil Biology and Biochemistry* 173, 108780.
- 622 <https://doi.org/10.1016/j.soilbio.2022.108780>.
- 623 He, W., Wang, H., Ye, W.H., Tian, Y.L., Hu, G.Q., Lou, Y.H., Pan, H., Yang, Q.G., Zhuge, Y.P.,
- 624 2022. Distinct stabilization characteristics of organic carbon in coastal salt-affected soils
- 625 with different salinity under straw return management. *Land Degradation &*



- 626 Development 33, 2246–2257. <https://doi.org/10.1002/ldr.4276>.
- 627 Hua, F.Y., Bruijnzeel, L.A., Meli, P., Martin, P.A., Zhang, J., Nakagawa, S., Miao, X., Wang,
 628 W., McEvoy, C., Peña-Arancibia, J.L., Brancalion, P.H.S., Smith, P., Edwards, D.P.,
 629 Balmford, A., 2022. The biodiversity and ecosystem service contributions and trade-offs
 630 of forest restoration approaches. *Science* 376, 839–844.
 631 <https://doi.org/10.1126/science.abl4649>.
- 632 Janzen, H.H., 2015. Beyond carbon sequestration: soil as conduit of solar energy. *European J*
 633 *Soil Science* 66, 19–32. <https://doi.org/10.1111/ejss.12194>.
- 634 Khaledi, S., Delbari, M., Galavi, H., Bagheri, H., Chari, M.M., 2023. Effects of biochar
 635 particle size, biochar application rate, and moisture content on thermal properties of an
 636 unsaturated sandy loam soil. *Soil and Tillage Research* 226, 105579.
 637 <https://doi.org/10.1016/j.still.2022.105579>.
- 638 Kleber, M., Sollins, P., Sutton, R., 2007. A conceptual model of organo-mineral interactions
 639 in soils: Self-assembly of organic molecular fragments into zonal structures on mineral
 640 surfaces. *Biogeochemistry* 85, 9–24. <https://doi.org/10.1007/s10533-007-9103-5>.
- 641 Lavallee, J.M., Soong, J.L., Cotrufo, M.F., 2020. Conceptualizing soil organic matter into
 642 particulate and mineral-associated forms to address global change in the 21st century.
 643 *Global Change Biology* 26, 261–273. <https://doi.org/10.1111/gcb.14859>.
- 644 Li, X.N., Li, R.P., Zhan, M.Q., Hou, Q., Zhang, H.Y., Wu, G.Q., Ding, L.Q., Lv, X.F., Xu, Y.,
 645 2024. Combined magnetic biochar and ryegrass enhanced the remediation effect of soils
 646 contaminated with multiple heavy metals. *Environment International* 185, 108498.
 647 <https://doi.org/10.1016/j.envint.2024.108498>.



648 Liang, J.J., Crowther, T.W., Picard, N., Wiser, S., Zhou, M., Alberti, G., Schulze, E.D.,
 649 McGuire, A.D., Bozzato, F., Pretzsch, H., de-Miguel, S., Paquette, A., Hérault, B.,
 650 Scherer-Lorenzen, M., Barrett, C.B., Glick, H.B., Hengeveld, G.M., Nabuurs, G.J.,
 651 Pfautsch, S., Viana, H., Vibrans, A.C., Ammer, C., Schall, P., Verbyla, D., Tchebakova,
 652 N., Fischer, M., Watson, J.V., Chen, H.Y.H., Lei, X., Schelhaas, M.J., Lu, H., Gianelle,
 653 D., Parfenova, E.I., Salas, C., Lee, E., Lee, B., Kim, H.S., Bruelheide, H., Coomes, D.A.,
 654 Piotto, D., Sunderland, T., Schmid, B., Gourlet-Fleury, S., Sonké, B., Tavani, R., Zhu, J.,
 655 Brandl, S., Vayreda, J., Kitahara, F., Searle, E.B., Neldner, V.J., Ngugi, M.R., Baraloto,
 656 C., Frizzera, L., Bałazy, R., Oleksyn, J., Zawila-Niedźwiecki, T., Bouriaud, O., Bussotti,
 657 F., Finér, L., Jaroszewicz, B., Jucker, T., Valladares, F., Jagodzinski, A.M., Peri, P.L.,
 658 Gonmadje, C., Marthy, W., O'Brien, T., Martin, E.H., Marshall, A.R., Rovero, F.,
 659 Bitariho, R., Niklaus, P.A., Alvarez-Loayza, P., Chamuya, N., Valencia, R., Mortier, F.,
 660 Wortel, V., Engone-Obiang, N.L., Ferreira, L.V., Odeke, D.E., Vasquez, R.M., Lewis,
 661 S.L., Reich, P.B., 2016. Positive biodiversity-productivity relationship predominant in
 662 global forests. *Science* 354, aaf8957. <https://doi.org/10.1126/science.aaf8957>.
 663 Liu, J.X., Sun, P., Chen, Y.Y., Guo, J.M., Liu, L.C., Zhao, X.Y., Xin, J., Liu, X.L., 2024. The
 664 regulation pathways of biochar and microorganism in soil-plant system by multiple
 665 statistical methods: The forms of carbon participation in coastal wetlands. *Chemosphere*
 666 362, 142918. <https://doi.org/10.1016/j.chemosphere.2024.142918>.
 667 Mitchell, E., Scheer, C., Rowlings, D., Conant, R.T., Cotrufo, M.F., Grace, P., 2018. Amount
 668 and incorporation of plant residue inputs modify residue stabilisation dynamics in soil
 669 organic matter fractions. *Agriculture, Ecosystems & Environment* 256, 82–91.



- 670 <https://doi.org/10.1016/j.agee.2017.12.006>.
- 671 Mohammed, I., Kodaolu, B., Zhang, T.Q., Wang, Y.T., Audette, Y., Longstaffe, J., 2023.
- 672 Analysis of molecular structure changes in humic acids from manure-amended soils
- 673 over 17 years using elemental analysis and solid-state ¹³C nuclear magnetic resonance
- 674 spectroscopy. *Soil Systems* 7, 76. <https://doi.org/10.3390/soilsystems7030076>.
- 675 Ndzelu, B.S., Dou, S., Zhang, X.W., Zhang, Y.F., Ma, R., Liu, X., 2021. Tillage effects on
- 676 humus composition and humic acid structural characteristics in soil aggregate-size
- 677 fractions. *Soil and Tillage Research* 213, 105090.
- 678 <https://doi.org/10.1016/j.still.2021.105090>.
- 679 Nelson, D.W., Sommers, L.E., 1982. Total carbon, organic carbon, and organic matter. In:
- 680 Page, A.L. (Ed.), *Methods of Soil Analysis: Part 2 Chemical and Microbiological*
- 681 Properties. Wiley, Hoboken, pp. 539–579.
- 682 <https://doi.org/10.2134/agronmonogr9.2.2ed.c29>.
- 683 Papageorgiou, A., Azzi, E.S., Enell, A., Sundberg, C., 2021. Biochar produced from wood
- 684 waste for soil remediation in Sweden: Carbon sequestration and other environmental
- 685 impacts. *Science of The Total Environment* 776, 145953.
- 686 <https://doi.org/10.1016/j.scitotenv.2021.145953>.
- 687 Pathy, A., Pokharel, P., Chen, X.L., Balasubramanian, P., Chang, S.X., 2023. Activation
- 688 methods increase biochar’s potential for heavy-metal adsorption and environmental
- 689 remediation: A global meta-analysis. *Science of The Total Environment* 865, 161252.
- 690 <https://doi.org/10.1016/j.scitotenv.2022.161252>.
- 691 Powlson, D.S., Stirling, C.M., Jat, M.L., Gerard, B.G., Palm, C.A., Sanchez, P.A., Cassman,



- 692 K.G., 2014. Limited potential of no-till agriculture for climate change mitigation. *Nature*
 693 *Climate Change* 4, 678–683. <https://doi.org/10.1038/nclimate2292>.
- 694 Rana, P., Soni, V., Sharma, S., Poonia, K., Patial, S., Singh, P., Selvasembian, R., Chaudhary,
 695 V., Hussain, C.M., Raizada, P., 2025. Harnessing nitrogen doped magnetic biochar for
 696 efficient antibiotic adsorption and degradation. *Journal of Industrial and Engineering*
 697 *Chemistry* 148, 174–195. <https://doi.org/10.1016/j.jiec.2025.01.025>.
- 698 Ren, Z.G., Zhang, H.Y., Wang, Y.W., Lu, L., Ren, D., Wang, J.J., 2021. Multiple roles of
 699 dissolved organic matter released from decomposing rice straw at different times in
 700 organic pollutant photodegradation. *Journal of Hazardous Materials* 401, 123434.
 701 <https://doi.org/10.1016/j.jhazmat.2020.123434>.
- 702 Rocci, K.S., Lavalley, J.M., Stewart, C.E., Cotrufo, M.F., 2021. Soil organic carbon response
 703 to global environmental change depends on its distribution between mineral-associated
 704 and particulate organic matter: A meta-analysis. *Science of The Total Environment* 793,
 705 148569. <https://doi.org/10.1016/j.scitotenv.2021.148569>.
- 706 Shi, H.Q., Liu, G., An, X.B., Zhao, Y.J., Zheng, F.L., Li, H.R., Zhang, X.C. (John), Pan, X.C.,
 707 Wu, B.L., Wang, X.S., 2024. Tracing soil erosion with Fe₃O₄ magnetic powder:
 708 Principle and application. *International Soil and Water Conservation Research* 12,
 709 419–431. <https://doi.org/10.1016/j.iswcr.2023.08.002>.
- 710 Shi, R.Y., Ni, N., Wang, R.H., Nkoh, J.N., Pan, X.Y., Dong, G., Xu, R.K., Cui, X.M., Li, J.Y.,
 711 2023. Dissolved biochar fractions and solid biochar particles inhibit soil acidification
 712 induced by nitrification through different mechanisms. *Science of The Total*
 713 *Environment* 874, 162464. <https://doi.org/10.1016/j.scitotenv.2023.162464>.



- 714 Slessarev, E.W., Chadwick, O.A., Sokol, N.W., Nuccio, E.E., Pett-Ridge, J., 2022. Rock
 715 weathering controls the potential for soil carbon storage at a continental scale.
 716 Biogeochemistry 157, 1–13. <https://doi.org/10.1007/s10533-021-00859-8>.
- 717 Sokol, N.W., Whalen, E.D., Jilling, A., Kallenbach, C., Pett-Ridge, J., Georgiou, K., 2022.
 718 Global distribution, formation and fate of mineral-associated soil organic matter under a
 719 changing climate: A trait-based perspective. Functional Ecology 36, 1411–1429.
 720 <https://doi.org/10.1111/1365-2435.14040>.
- 721 Stewart, C.E., Follett, R.F., Wallace, J., Pruessner, E.G., 2012. Impact of biosolids and tillage
 722 on soil organic matter fractions: Implications of carbon saturation for conservation
 723 management in the virginia coastal plain. Soil Science Society of America Journal 76,
 724 1257–1267. <https://doi.org/10.2136/sssaj2011.0165>.
- 725 Vendig, I., Guzman, A., De La Cerda, G., Esquivel, K., Mayer, A.C., Ponisio, L., Bowles,
 726 T.M., 2023. Quantifying direct yield benefits of soil carbon increases from cover
 727 cropping. Nature Sustainability 6, 1125–1134.
 728 <https://doi.org/10.1038/s41893-023-01131-7>.
- 729 Von Lützow, M., Kögel-Knabner, I., Ekschmitt, K., Flessa, H., Guggenberger, G., Matzner, E.,
 730 Marschner, B., 2007. SOM fractionation methods: Relevance to functional pools and to
 731 stabilization mechanisms. Soil Biology and Biochemistry 39, 2183–2207.
 732 <https://doi.org/10.1016/j.soilbio.2007.03.007>.
- 733 Wang, J.S., Li, S.C., Yin, H.M., Riaz, M., Liu, X.W., Zhang, M.Y., 2025. Biochar suppresses
 734 Clubroot disease in Chinese cabbage by improving soil nutrient conditions and
 735 recruiting beneficial microorganisms. Applied Soil Ecology 210, 106107.



- 736 <https://doi.org/10.1016/j.apsoil.2025.106107>.
- 737 Witzgall, K., Vidal, A., Schubert, D.I., Höschen, C., Schweizer, S.A., Buegger, F., Pouteau, V.,
 738 Chenu, C., Mueller, C.W., 2021. Particulate organic matter as a functional soil
 739 component for persistent soil organic carbon. *Nature Communications* 12, 4115.
 740 <https://doi.org/10.1038/s41467-021-24192-8>.
- 741 Wood, S.A., Sokol, N., Bell, C.W., Bradford, M.A., Naeem, S., Wallenstein, M.D., Palm,
 742 C.A., 2016. Opposing effects of different soil organic matter fractions on crop yields.
 743 *Ecological Applications* 26, 2072–2085. <https://doi.org/10.1890/16-0024.1>.
- 744 Xie, H.T., Li, J.W., Zhu, P., Peng, C., Wang, J.K., He, H.B., Zhang, X.D., 2014. Long-term
 745 manure amendments enhance neutral sugar accumulation in bulk soil and particulate
 746 organic matter in a Mollisol. *Soil Biology and Biochemistry* 78, 45–53.
 747 <https://doi.org/10.1016/j.soilbio.2014.07.009>.
- 748 Xu, C.B., Tan, X., Zhao, J.W., Cao, J.M., Ren, M., Xiao, Y., Lin, A.J., 2021. Optimization of
 749 biochar production based on environmental risk and remediation performance: Take
 750 kitchen waste for example. *Journal of Hazardous Materials* 416, 125785.
 751 <https://doi.org/10.1016/j.jhazmat.2021.125785>.
- 752 Yin, J.X., Zhao, L., Xu, X.Y., Li, D.P., Qiu, H., Cao, X.D., 2022. Evaluation of long-term
 753 carbon sequestration of biochar in soil with biogeochemical field model. *Science of The*
 754 *Total Environment* 822, 153576. <https://doi.org/10.1016/j.scitotenv.2022.153576>.
- 755 Yu, W.J., Huang, W.J., Weintraub-Leff, S.R., Hall, S.J., 2022. Where and why do particulate
 756 organic matter (POM) and mineral-associated organic matter (MAOM) differ among
 757 diverse soils? *Soil Biology and Biochemistry* 172, 108756.



- 758 <https://doi.org/10.1016/j.soilbio.2022.108756>.
- 759 Zhang, B.L., Jin, Y.P., Qi, J.X., Chen, H., Chen, G., Tang, S.S., 2021. Porous carbon materials
 760 based on *Physalis alkekengi* L. husk and its application for removal of malachite green.
 761 Environmental Technology & Innovation 21, 101343.
 762 <https://doi.org/10.1016/j.eti.2020.101343>.
- 763 Zhang, B.L., Li, R.Q., Zheng, Y.Y., Chen, S.J., Su, Y.J., Zhou, W., Sui, Q., Liang, D.D., 2024.
 764 Biochar composite with enhanced performance prepared through microbial modification
 765 for water pollutant removal. International Journal of Molecular Sciences 25, 11732.
 766 <https://doi.org/10.3390/ijms252111732>.
- 767 Zhang, G.X., Ren, R., Yan, X.R., Zhu, Y., Zhang, H.Y., Yan, G.Y., 2025. The key role of
 768 magnetic iron-to-biochar mass ratios in the dissipation of oxytetracycline and its
 769 resistance genes in soils with and without biodegradable microplastics. Journal of
 770 Environmental Management 377, 124658.
 771 <https://doi.org/10.1016/j.jenvman.2025.124658>.
- 772 Zhou, J.Y., Liu, Y.Y., Han, Y.T., Jing, F.Q., Chen, J.W., 2019. Bone-derived biochar and
 773 magnetic biochar for effective removal of fluoride in groundwater: Effects of synthesis
 774 method and coexisting chromium. Water Environment Research 91, 588–597.
 775 <https://doi.org/10.1002/wer.1068>.