



1 Characteristics and processing of aqueous secondary organic aerosols
2 during autumn in suburban Eastern China: role of aerosol liquid
3 water, aerosol acidity, and photochemistry

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17



18 **Abstract**

19 Aqueous-phase secondary organic aerosols (aqSOA) constitute a large fraction of SOA,
20 thereby exerting significant influence on air quality, climate, and human health.
21 However, its formation mechanisms remain unclear due to limited observational
22 evidence. We conducted field measurements of particulate matter (PM) composition by
23 deploying high-resolution aerosol mass spectrometry in a suburban environment during
24 autumn in Nanjing, China. The characteristics and formation pathways of aqSOA are
25 comprehensively investigated by using Positive Matrix Factorization (PMF) method.
26 Our results show that aqSOA accounted for 27.6% of oxidized organic aerosols,
27 exhibiting elevated O:C ratios (0.78) and strong correlations with nitrate and aerosol
28 liquid water (ALW). The important role of acid-catalyzed reactions is also revealed by
29 the enhanced production of aqSOA at lower aerosol pH conditions. Under elevated
30 nitrate and ALW levels, a pronounced morning aqSOA peak was frequently observed;
31 whereas a noon-time aqSOA peak was also observed on several days, likely governed
32 by photochemistry and aqueous-phase reactions. These findings highlight the critical
33 roles of nitrate, ALW, acidity, and photochemistry in driving aqSOA production in
34 polluted urban environments. This study advances the mechanistic understanding of
35 aqSOA formation and provides insights into the mitigation of SOA in Eastern China.



36 **1 Introduction**

37 Organic aerosols (OA) significantly affect air quality, human health, and climate by
38 influencing radiative forcing and cloud formation (Kanakidou et al., 2005; Zhou et al.,
39 2019). Secondary organic aerosols (SOA) can contribute 30–70% of total organic
40 aerosols (Huang et al., 2014; Xian et al., 2023). The formation of SOA has been
41 attributed mainly to gas-phase oxidation of volatile organic compounds (VOCs), where
42 the oxidized products subsequently partition into the aerosol phase (Hennigan et al.,
43 2009; Seinfeld and Pankow, 2003; Ziemann and Atkinson, 2012). Recently, growing
44 evidence highlights the important role of SOA generated through aqueous-phase
45 processes in cloud droplets, fog, and aerosol liquid water (ALW) (Ervens et al., 2011;
46 Kim et al., 2019; Sun et al., 2010). However, the formation mechanisms and sources of
47 aqueous-phase SOA (aqSOA) remain highly uncertain (Ervens et al., 2011; Huang et
48 al., 2025; McNeill, 2015) .

49 In recent years, concentrations of OA have gradually declined across many regions
50 in China due to the implementation of air pollution control measures, but the relative
51 contribution of SOA has increased markedly (Chen et al., 2024), with aqSOA
52 constituting a substantial fraction. Numerous studies have demonstrated that elevated
53 ALW can significantly promote aqSOA formation during winter haze episodes through
54 multiphase reactions (Chen et al., 2021; Feng et al., 2022; Liu et al., 2019; Peng et al.,
55 2021; Sun et al., 2016, 2019; Wang et al., 2023; Xiao et al., 2022; Xu et al., 2017; Zhao
56 et al., 2019). Still, the exact processes driving aqSOA formation is not fully
57 characterized. Field observations in urban Beijing have shown that ring-breaking



58 oxidation and functionalization of polycyclic aromatic hydrocarbons of fossil-fuel-
59 derived primary organic aerosols could lead to rapid aqSOA formation at high relative
60 humidity (RH) during winter haze episodes (Wang et al., 2021). Another field study
61 conducted in the North China Plain showed that the formation of aqSOA could be
62 largely enhanced under favorable photochemical conditions with precursors originated
63 from biomass burning activities (Kuang et al., 2020). Recently, field measurements in
64 Hebei reported that high nitrate may support the potential formation/transformation
65 from POA-related components to aqSOA (Gu et al., 2023). Laboratory studies reveal
66 that accretion reactions, which play a crucial role in SOA formation, are highly sensitive
67 to pH levels (Tilgner et al., 2021). Moreover, under the emerging dominance of nitrate
68 in aerosol composition (Huang et al., 2025), the interplay among nitrate, ALW, and pH
69 may complex the aqSOA formation and requires further investigation.

70 The Yangtze River Delta (YRD) region is one of the most densely populated and
71 economically developed areas in China, characterized by intensive industrial activity,
72 heavy traffic emissions, frequent regional pollution episodes, and high relative
73 humidity (Liu et al., 2025). Several previous studies have examined aqSOA processes
74 in this region (Wang et al., 2016; Wu et al., 2018; Xian et al., 2023), but uncertainties
75 regarding its sources, controlling factors, and formation mechanisms still remained.
76 Consequently, it is challenging for current models to precisely simulate aqSOA and its
77 contribution to the YRD region (Ervens et al., 2011; Rogers et al., 2025).

78 In this study, we conducted real-time measurements of OA at the National
79 Observation and Research Station for Atmospheric Processes and Environmental



80 Change in Yangtze River Delta (SORPES) located in suburban Nanjing in the western
81 YRD region during the autumn of 2020. The chemical characteristics and formation
82 mechanisms of aqSOA, as well as the roles of nitrate, ALW, and aerosol acidity are
83 investigated. By classifying diurnal variation patterns of aqSOA, we assessed the
84 relative roles of photochemical and aqueous-phase processes. The results provide new
85 insights into regional aqSOA formation in the YRD and have implications for the
86 development of effective air pollution control strategies.

87 **2 Materials and Methods**

88 **2.1 Sampling Site**

89 The field campaign was conducted from 13 October to 30 December in 2020 at
90 SORPES station (118°57'E, 32°07'N) located in the Xianlin campus of Nanjing
91 University in Nanjing, China. This is a representative station of western YRD,
92 surrounded by high vegetation cover, and also subject to more anthropogenic emissions
93 (Ding et al., 2016, 2019; Dou et al., 2025; Liu et al., 2025).

94 **2.2 Instrumentation**

95 Real-time non-refractory PM₁ composition was measured using a high-resolution time-
96 of-flight aerosol mass spectrometer (HR-ToF-AMS; hereafter, AMS; Aerodyne
97 Research Inc.). An aerodynamic PM₁ lens was used to focus the particle into a beam,
98 which was then impacted on the heated tungsten surface (~600°C) and flash-vaporized.
99 In our study, ambient aerosols were passed through a ~2 m long stainless-steel sampling
100 tube, dried by a Nafion drying tube, and then introduced into the AMS. In order to
101 obtain highly sensitive data, AMS was operated in V mode with a time resolution of 2



102 minutes (DeCarlo et al., 2006).

103 Other instruments were also employed at SORPES in support of these
104 measurements. Black carbon (BC) was measured by the photoacoustic extinctionmeter
105 (PAX, Droplet Measurement Technologies Inc., USA). The meteorological parameters
106 and gaseous pollutants were also measured simultaneously. Ozone (O_3), carbon
107 monoxide (CO), nitric oxide (NO), nitrogen oxides (NO_x) and sulfur dioxide (SO_2)
108 were measured using online analyzers (Thermo Fisher Scientific, USA). Ammonia
109 (NH_3) was measured by the Picarro G2103 gas analyzer (Picarro Inc., USA) (Liu et al.,
110 2024). Temperature, RH and other meteorological parameters were monitored by
111 meteorological sensors (GRWS100, Campbell, USA).

112 **2.3 Data Analysis**

113 The AMS data were processed by SQUIRREL (version 1.60P) and PIKA (version 1.20P)
114 from the ToF-AMS Software Downloads Web page ([http://cires.colorado.edu/jimenez-](http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html)
115 [group/ToFAMSResources/ToFSoftware/index.html](http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html)). The ionization efficiency (IE)
116 was calibrated using 300 nm pure ammonium nitrate before and after the campaign.
117 The relative ionization efficiency (RIE) of ammonium was determined from pure
118 ammonium nitrate, yielding a value of 3.52. RIE values for OA, nitrate, sulfate, and
119 chloride were set to their default values of 1.4, 1.1, 1.2, and 1.3, respectively
120 (Canagaratna et al., 2007). As well, the collection efficiency (CE) was assigned a typical
121 value of 0.5 for common environments. Element ratios, including H: C, O: C, N: C and
122 OM: OC, are calculated using the Improved-Ambient method (Canagaratna et al., 2015).

123 In addition, ALW content and aerosol acidity that are associated with inorganic



species were estimated by the Extended Aerosol Inorganics Model (E-AIM), which is a well-known inorganic thermodynamic model without simplifying assumptions (Clegg et al. 1998; Wexler and Clegg 2002; Pye et al. 2020).

2.4 Source Apportionment of OA

Positive Matrix Factorization (PMF) analysis was applied to the high-resolution mass spectra of organic matrix for m/z 12 – 120 to resolve distinct OA factors from specific sources (Paatero and Tapper, 1994; Ulbrich et al., 2009). The data and error matrices were treated according to the procedures detailed in DeCarlo et al. (2010). By comparing the mass spectral profiles with previous studies and correlations with time series of tracers, five OA factors with $f_{\text{peak}} = 0$ were selected, including one primary organic aerosol (POA) factor, one nitrogenous OA (NOA) factor, and three SOA factors, namely, less-oxidized oxygenated OA (LO-OOA), more-oxidized OOA (MO-OOA), and aqSOA. The detailed diagnostic plots are shown in Figures S1 and S2 in the supporting information.

Among these five OA factors, aqSOA exhibits typical characteristics of highly oxidized organic aerosols: fraction of m/z 44 (CO_2^+ , primarily from carboxylic acids and highly oxidized compounds) (Heald et al., 2010) in total signals exceeded that of m/z 43 (typically representing less oxidized compounds) (Figure S2). Elemental analysis further indicated a strongly oxygenated character, with an average H:C ratio of 1.80 and an O:C ratio of 0.78. These values are comparable to those reported in other regions, such as northern Italy and Beijing (Gilardoni et al., 2016; Xu et al., 2019). The N:C ratio (0.05) was also elevated and similar to wintertime values observed in Beijing



(N:C = 0.045; Xu et al. 2017), suggesting nitrogen-containing compound formation via aqueous-phase processing. AqSOA exhibited strong correlations with unique fragment ions that are widely recognized as markers of aqueous-phase secondary products (Sun et al., 2016; Xu et al., 2019). For instance, significant correlations were observed with C_2O_2^+ (m/z 56, $r = 0.77$), a typical fragment of oxalate-related species, and with CH_3SO^+ (m/z 63, $r = 0.90$), which is indicative of organosulfur compounds (Figure S3). These results provide chemical evidence supporting the aqueous-phase origin of aqSOA in Nanjing.

3.Results and Discussions

3.1 General Characteristics of aqSOA

The meteorological conditions in Nanjing were overall stable during the field campaign (October-December), with an average temperature and RH of 14.6 ± 4.6 °C and $65.1 \pm 17.9\%$, respectively (Figure S4). The diurnal variation of RH ranged from 50% to 85%, resulting in relatively humid air conditions that favored aqueous-phase chemical reactions. The average wind speed of the prevailing northerly wind was relatively low (0.21 ± 0.16 m/s), resulting in the accumulation of local pollutants. In particular, the average NR-PM₁ concentration was 37.3 ± 20.6 µg/m³, indicating frequently occurred particulate matter pollution during this period.

Analysis of aerosol chemical composition revealed that organic aerosol was the dominant NR-PM₁ component in Nanjing, accounting for 40.9% of the total PM₁ concentration, evidently higher than nitrate (30.7%) and sulfate (13.9%; Figure S4). The time series of aqSOA showed significant variation, with peak concentrations up to



15.7 $\mu\text{g}/\text{m}^3$ (Figure 1a). The average concentration of aqSOA during the campaign was 3.1 $\mu\text{g}/\text{m}^3$, accounting for 20.2% (62.7% in maximum) of the total OA and 27.6% (78.2% in maximum) of the total SOA (Figure 1b). The average fraction of aqSOA in OA was much higher than Beijing (13-17%) (Zhao et al., 2019), indicating an important role of aqueous-phase processes in SOA formation in Nanjing. Moreover, both aqSOA concentrations and their relative contribution to total organic aerosol increased significantly with increasing RH (Figure S5a), which was likely due to the increased availability of the aqueous reaction medium enhanced by water uptake by aerosols, which is crucial for the aqSOA formation. This is further confirmed with the sharply increased contribution of aqSOA to OA (up to 62.7%) at RH levels above 80%, suggesting that high-humidity environments in suburban Nanjing substantially promoted aqSOA formation through enhanced aqueous-phase chemistry.

The Van Krevelen diagram (Heald et al., 2010) provided further insights into the chemical aging of OA (Figure S6). The H:C vs. O:C slope was near -0.5, indicating OA oxidation primarily involving carboxylic acid and peroxide or alcohol functional groups addition without fragmentation and/or the addition of carboxylic acid functional groups with fragmentation (Ng et al., 2011). In particular, OA observed under high RH conditions (>80%) clustered in the upper plot region, suggesting distinct chemical evolution of OA when aqueous-phase reactions were involved. Specifically, the nearly zero slope of the relationship among POA, LO-OOA and aqSOA suggests that the observed increase in the O:C ratio of aqSOA may be related to oligomerization and hydroxyl formation through dark chemistry processes (Lim et al., 2010).

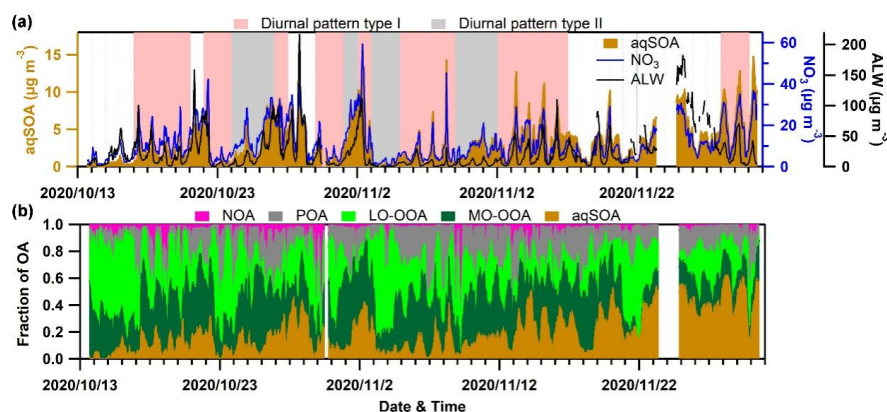


Figure 1. (a) Time series of aqSOA and nitrate concentrations. The pink area represents diurnal pattern type I, while the grey area represents diurnal pattern type II. (b) Time series of the fraction of five OA factors.

3.2 Enhanced aqSOA Formation Driven by Nitrate, ALW, and Acid Catalysis

As a strongly hygroscopic component, nitrate aerosol can enhance aerosol water uptake, thereby modifying the aqueous microenvironment for SOA production (Hodas et al., 2014; Sullivan et al., 2016). We observed strong correlation between aqSOA and nitrate concentrations ($R^2 = 0.83$; Figure 2a) during the observation period, and the variation of aqSOA was highly consistent with that of nitrate aerosol (Figure 1a). This indicates that, nitrate aerosols play an important role in the aqSOA formation in Nanjing, likely through influencing ALW and/or aqueous reactions.

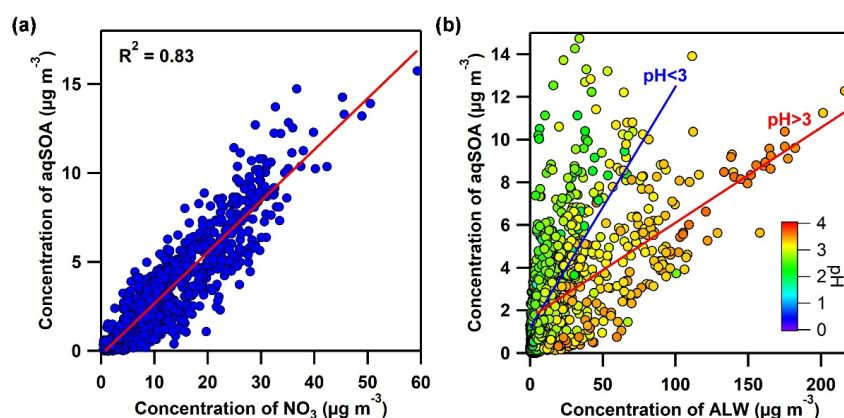


Figure 2. (a) Scatter plot of aqSOA and nitrate concentrations. (b) Scatter plot of aqSOA and ALW concentrations, colored by aerosol pH. The blue line represents the fitted line for data with aerosol pH < 3, while the red line represents the fitted line for data with aerosol pH > 3.

The correlation between aqSOA and aerosol liquid water (ALW; $r = 0.63$; Figure 2b), which is mainly derived from the hygroscopic growth of inorganic salts such as nitrate, was moderate compared to nitrate. Still, their positive relationship indicates that higher ALW levels may enhance aqueous-phase chemical processes by providing the medium in which multiphase reactions can occur (Hodas et al., 2014). The presence of ALW promotes the partitioning of water-soluble organic precursors and supports subsequent aqueous-phase reactions leading to aqSOA formation. The results are consistent with previous studies in humid urban environments (Chen et al., 2021; Duan et al., 2022; Kuang et al., 2020).

In addition to ALW, aerosol acidity also exerted a strong influence on aqSOA yields during the campaign (Lim et al., 2010; Tilgner et al., 2021). To illustrate, the



dataset was separated by aerosol pH ($\text{pH} < 3$ vs. $\text{pH} > 3$) (Figure 2b), the slopes of ALW-aqSOA correlations decreased with increasing pH, indicating that aqSOA production was higher under more acidic conditions at the same ALW level. This pattern demonstrates the importance of acid-catalyzed reactions in driving aqSOA formation. Previous studies have demonstrated that under acidic conditions (low pH, high H^+ concentration), non-oxidative aqueous organic chemical processes, such as accretion reactions (aldol condensation, hemiacetal and acetal formation, and the esterification of carboxylic acids) are important formation pathways of aqSOA (Freedman et al., 2019; Tilgner et al., 2021). For example, the hydration of methylglyoxal and its subsequent acetal formation are highly pH-dependent, requiring a pH of less than 3.5 to occur (Yasmeen et al., 2010). As aerosol pH increases (lower H^+ availability), the catalytic efficiency of these acid-driven reactions diminishes, leading to a reduction in aqSOA production efficiency. Collectively, these findings suggest that aerosol acidity plays a pivotal regulatory role in aqSOA formation, linking aqueous chemistry to the broader context of aerosol physicochemical properties in Nanjing.

Thus, our measurements revealed a synergistic interplay among nitrate, ALW, and aerosol acidity in regulating aqSOA formation in Nanjing. Nitrate enhances the ALW content, which in turn promotes aqueous-phase reactions, while aerosol acidity governs the efficiency of these chemical processes. These processes together constituted the formation mechanism of aqSOA during the observation period.



244 **3.3 Synergistic Role of Aqueous and Photochemical Processes**

245 During the campaign, the average diurnal variation of aqSOA in Nanjing exhibited a
246 distinct morning peak at approximately 09:00 local time (Figure S2d). This pattern is
247 different from most of the previous urban observations, where aqSOA concentrations
248 typically peak during nighttime periods (Sun et al., 2016; Xu et al., 2019; Gu et al.,
249 2023). To address this, the time series of aqSOA was analyzed on a daily basis, with
250 rainy days excluded. Based on this analysis, two distinct diurnal variation patterns were
251 identified: type I (22 days) and type II (9 days) (Figure 1a). For diurnal pattern type I,
252 the daily variation of aqSOA exhibits a pronounced morning peak, whereas the daytime
253 concentration of aqSOA shows a noon peak for diurnal pattern type II (Figure 3). These
254 two distinct diurnal patterns of aqSOA could be attributed to multiple formation
255 mechanisms of aqSOA or different meteorological conditions.

256 During diurnal pattern type I, aqSOA concentrations exhibited a pronounced
257 single peak between 09:00 and 10:00 local time, following a period of continuous
258 nighttime accumulation from approximately 20:00 to 07:00 (Figure 3a). Notably, the
259 aqSOA peak lagged behind the nitrate peak by about one hour, and nocturnal increases
260 in nitrate, ALW, and aqSOA were highly synchronized. Compared with type II,
261 nighttime meteorological conditions during type I were characterized by higher RH
262 (73.1% vs. 61.2%) and ALW concentrations ($34.2 \mu\text{g}/\text{m}^3$ vs. $12.6 \mu\text{g}/\text{m}^3$). Such humid
263 conditions favor the formation of both nitrate and ALW, thereby enhancing aqSOA
264 production through aqueous-phase reactions. Moreover, the higher RH conditions
265 observed during type I events could further promote aqueous reaction pathways, as the



nitrate formation from N_2O_5 hydrolysis on aqueous aerosol surfaces is strongly facilitated under such conditions (Sun et al., 2018). Overall, these observations indicate that aqSOA production under type I conditions is predominantly controlled by aerosol aqueous-phase processes, highlighting the important role of nocturnal multiphase chemistry in driving morning aqSOA peaks.

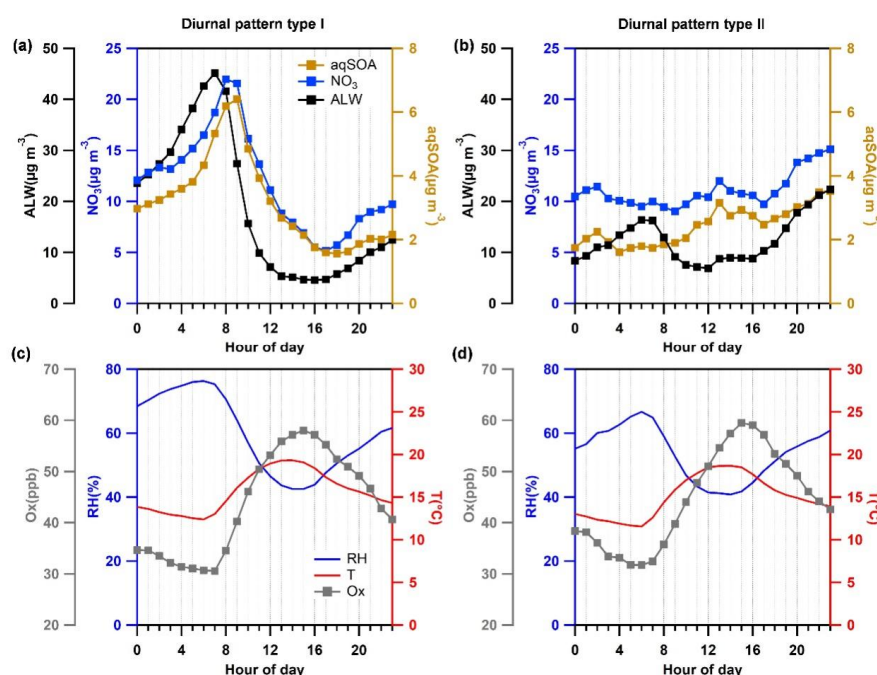
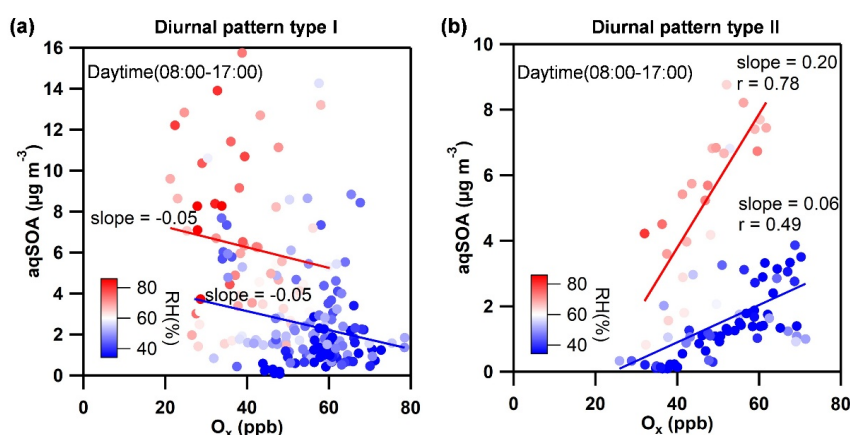


Figure 3. Diurnal variations for SOA, ALW, and nitrate in diurnal pattern types I (a) and II (b). Diurnal variations for temperature, humidity, and O_x concentration in diurnal pattern types I (c) and II (d).

In contrast, aqSOA concentrations gradually increased during the daytime and peaked around 13:00 local time (Figure 3b) during diurnal pattern type II, indicating an



279 influence of photochemical processes. The diurnal variation of aqSOA is consistent
 280 with that of nitrate aerosols. This suggest that photochemistry leads to the formation of
 281 nitrate aerosols and likely further mediate the aqSOA formation through water uptake.
 282 O_x is commonly regarded as a tracer for photochemical processes. The relationship
 283 between daytime (08:00–17:00) aqSOA concentrations and O_x levels was examined for
 284 both diurnal patterns (Figure 4). For diurnal type I, no significant correlation was
 285 observed, suggesting that daytime photochemistry played a minimal role in aqSOA
 286 production under these conditions. In contrast, diurnal type II displayed a positive
 287 correlation, implying that aqSOA formation was partially related to the photochemical
 288 processes.
 289



290
 291 **Figure 4.** Scatter plots of aqSOA and O_x for Diurnal pattern types I (a) and II (b). Data
 292 points are colored by RH, with red and blue lines representing fits for RH > 60% and
 293 RH < 60% respectively.

294



295 Further analysis revealed that the slope of the aqSOA-O_x relationship under high
296 RH conditions in diurnal type II was approximately three times higher than that under
297 low RH conditions. This is because higher RH likely enhanced the partitioning of semi-
298 volatile photochemical oxidation products into the aqueous phase and promoted
299 aqueous-phase photochemical reactions, thereby facilitating aqSOA formation. These
300 results indicate that daytime aqSOA production during type II is governed by the
301 combined effects of photochemical oxidation and humidity-dependent aqueous
302 processing.

303

304 **4. Conclusions**

305 We analyzed the characteristics and formation mechanisms of aqSOA in the
306 autumn season in Nanjing, where aqSOA was found to constitute a significant fraction
307 of organic aerosols (20.2%). The results demonstrate that nitrate, ALW, and aerosol
308 acidity act in concert to regulate aqSOA formation, with nitrate serving as a critical
309 driver through its strong hygroscopicity. We observed the occurrence of a distinct
310 morning peak of aqSOA in the YRD region, a feature that differs from most of the
311 previously reported studies in urban environments. Based on diurnal variation analysis,
312 two distinct diurnal patterns were identified: type I, dominated by nighttime aqueous-
313 phase chemistry linked to nitrate and ALW accumulation; and Type II, shaped primarily
314 by daytime photochemical oxidation. These findings highlight that aqSOA formation
315 in this megacity is governed by the synergistic effects of both nocturnal aqueous
316 reactions and daytime photochemical processes.



317 China is strictly controlling SO₂ emissions, which has led to a decrease in sulfate
318 content in atmospheric particulate matter. Nitrate is becoming a more important
319 inorganic component of PM_{2.5}, and its role in promoting aqSOA formation will become
320 even more significant. Although this study did not directly resolve the specific
321 precursor compounds of aqSOA, our results still highlight the importance of paying
322 greater attention to nitrate-driven aqueous processes in future air quality assessments.
323 In particular, coordinated management of nitrogen oxides (NO_x) alongside traditional
324 particulate matter controls may provide an effective pathway for mitigating aqSOA
325 burdens in a post-sulfate-dominated atmosphere.

326

327 **Data availability**

328 All data are available from the corresponding author upon request.

329 **Author contributions**

330 T.L. and A.D. designed the research project; J.W., D.G., C.Z., C.R., W.X., J.W., W.N.,
331 X.C., S.L. and X.H. performed the research; Q.W., Q.Z., and X.Q. analyzed data; Q.W.,
332 Q.Z., and T.L. wrote the paper. All authors participated in the relevant scientific
333 discussion and commented on the manuscript.

334 **Competing interests**

335 The authors declare no competing interests.

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