

1 **Characteristics and processing of aqueous secondary organic aerosols**
2 **during autumn in suburban Eastern China: role of aerosol liquid**
3 **water, aerosol acidity, and photochemistry**

4 Qiu Wang¹, Tengyu Liu^{1,2,3*}, Weiqi Xu⁴, Jinbo Wang^{1,5}, Dafeng Ge¹, Caijun Zhu¹,
5 Chuanhua Ren¹, Jiaping Wang^{1,2}, Qiaozhi Zha^{1,2}, Ximeng Qi^{1,2}, Wei Nie^{1,2}, Xuguang
6 Chi^{1,2}, Sijia Lou^{1,2,3}, Xin Huang^{1,2,3}, Aijun Ding^{1,2,3}

7 ¹School of Atmospheric Sciences, Nanjing University, Nanjing, 210023, China

8 ²National Observation and Research Station for Atmospheric Processes and
9 Environmental Change in Yangtze River Delta, Nanjing, 210023, China

10 ³Frontiers Science Center for Critical Earth Material Cycling, Nanjing University,
11 Nanjing, 210023, China

12 ⁴State Key Laboratory of Atmospheric Environment and Extreme Meteorology,
13 Institute of Atmospheric Physics, Chinese Academy of Sciences, Beijing 100029, China

14 ⁵Now at National Satellite Meteorological Center, China Meteorological
15 Administration, Beijing 100081, China

16 *Corresponding author: Tengyu Liu (tengyu.liu@nju.edu.cn)

17

18 **Abstract**

19 Aqueous-phase secondary organic aerosols (aqSOA) constitute a large fraction of SOA,
20 thereby exerting significant influence on air quality, climate, and human health.
21 However, its formation pathways under real ambient conditions in Chinese urban
22 regions remain insufficiently constrained. We conducted field measurements of
23 particulate matters (PM) composition by deploying high-resolution aerosol mass
24 spectrometry in a suburban environment during autumn in Nanjing, China. The
25 characteristics and formation pathways of aqSOA are comprehensively investigated by
26 using Positive Matrix Factorization (PMF) method. Our results show that aqSOA
27 accounted for 27.6% of oxidized organic aerosols, exhibiting elevated O:C ratios (0.78)
28 and strong correlations with nitrate and aerosol liquid water (ALW). The important role
29 of acid-catalyzed reactions is also revealed by the enhanced production of aqSOA at
30 lower aerosol pH conditions. Under elevated nitrate and ALW levels, a pronounced
31 morning aqSOA peak was frequently observed; whereas a noon-time aqSOA peak was
32 also observed on several days, likely governed by photochemistry and aqueous-phase
33 reactions. These findings highlight the critical roles of nitrate, ALW, acidity, and
34 photochemistry in driving aqSOA production in polluted urban environments. This
35 study advances the mechanistic understanding of aqSOA formation and provides
36 insights into the mitigation of SOA in the Eastern China.

37 **1 Introduction**

38 Organic aerosols (OA) significantly affect air quality, human health, and climate by
39 influencing radiative forcing and cloud formation (Kanakidou et al., 2005; Zhou et al.,
40 2019). Secondary organic aerosols (SOA) can contribute 30–70% of total organic
41 aerosols (Huang et al., 2014; Xian et al., 2023). The formation of SOA has been
42 attributed mainly to gas-phase oxidation of volatile organic compounds (VOCs), where
43 the oxidized products subsequently partition into the aerosol phase (Hennigan et al.,
44 2009; Seinfeld and Pankow, 2003; Ziemann and Atkinson, 2012). Recently, growing
45 evidence highlights the important role of SOA generated through aqueous-phase
46 processes in cloud droplets, fog, and aerosol liquid water (ALW) (Ervens et al., 2011;
47 Kim et al., 2019; Sun et al., 2010). However, the formation mechanisms and sources of
48 aqueous-phase SOA (aqSOA) in Chinese urban regions remain highly uncertain
49 (Ervens et al., 2011; Huang et al., 2025; McNeill, 2015) .

50 In recent years, concentrations of OA have gradually declined across many regions
51 in China due to the implementation of air pollution control measures, but the relative
52 contribution of SOA has increased markedly (Chen et al., 2024; Huang et al., 2025),
53 with aqSOA constituting a substantial fraction. Numerous studies have demonstrated
54 that elevated ALW can significantly promote aqSOA formation during winter haze
55 episodes through multiphase reactions (Chen et al., 2021; Feng et al., 2022; Liu et al.,
56 2019; Peng et al., 2021; Sun et al., 2016, 2019; Wang et al., 2023; Xiao et al., 2022; Xu
57 et al., 2017; Zhao et al., 2019). Still, the exact processes driving aqSOA formation is
58 not fully characterized. Field observations in urban Beijing have shown that ring-

59 breaking oxidation and functionalization of polycyclic aromatic hydrocarbons of fossil-
60 fuel-derived primary organic aerosols could lead to rapid aqSOA formation at high
61 relative humidity (RH) during winter haze episodes (Wang et al., 2021). Another field
62 study conducted in the North China Plain showed that the formation of aqSOA could
63 be largely enhanced under favorable photochemical conditions with precursors
64 originated from biomass burning activities (Kuang et al., 2020). Recently, field
65 measurements in Hebei reported that high nitrate may support the potential
66 formation/transformation from POA-related components to aqSOA (Gu et al., 2023).
67 Laboratory studies reveal that accretion reactions, which play a crucial role in SOA
68 formation, are highly sensitive to pH levels (Tilgner et al., 2021). Moreover, under the
69 emerging dominance of nitrate in aerosol composition (Huang et al., 2025), the
70 interplay among nitrate, ALW, and pH may complex the aqSOA formation and requires
71 further investigation.

72 The Yangtze River Delta (YRD) region is one of the most densely populated and
73 economically developed areas in China, characterized by intensive industrial activity,
74 heavy traffic emissions, frequent regional pollution episodes, and high relative
75 humidity (Liu et al., 2025). Several previous studies have examined aqSOA processes
76 in this region (Wang et al., 2016; Wu et al., 2018; Xian et al., 2023), but uncertainties
77 regarding its sources, controlling factors, and formation mechanisms still remained.
78 Consequently, it is challenging for current models to precisely simulate aqSOA and its
79 contribution to the YRD region (Ervens et al., 2011; Rogers et al., 2025).

80 In this study, we conducted real-time measurements of OA at the National

81 Observation and Research Station for Atmospheric Processes and Environmental
82 Change in Yangtze River Delta (SORPES) located in suburban Nanjing in the western
83 YRD region during the autumn of 2020. The chemical characteristics and formation
84 mechanisms of aqSOA, as well as the roles of nitrate, ALW, and aerosol acidity) are
85 investigated. By classifying diurnal variation patterns of aqSOA, we assessed the
86 relative roles of photochemical and aqueous-phase processes. The results provide new
87 insights into regional aqSOA formation in the YRD and have implications for the
88 development of effective air pollution control strategies.

89 **2 Materials and Methods**

90 **2.1 Sampling Site**

91 The field campaign was conducted from 13 October to 30 December in 2020 at
92 SORPES station (118°57'E, 32°07'N) located in the Xianlin campus of Nanjing
93 University in Nanjing, China. This is a representative station of western YRD,
94 surrounded by high vegetation cover, and also subject to abundant anthropogenic
95 emissions (Ding et al., 2016, 2019; Dou et al., 2025; Liu et al., 2025).

96 **2.2 Instrumentation**

97 Real-time non-refractory PM₁ composition was measured using a high-resolution time-
98 of-flight aerosol mass spectrometer (HR-ToF-AMS; hereafter, AMS; Aerodyne
99 Research Inc.). An aerodynamic PM₁ lens was used to focus the particle into a beam,
100 which was then impacted on the heated tungsten surface (~600°C) and flash-vaporized.
101 In our study, ambient aerosols were passed through a ~2 m long stainless-steel sampling
102 tube, dried by a Nafion drying tube, and then introduced into the AMS. In order to

103 obtain highly sensitive data, AMS was operated in V mode with a time resolution of 2
104 minutes (DeCarlo et al., 2006).

105 Other instruments were also employed at SORPES in support of these
106 measurements. Black carbon (BC) was measured by the photoacoustic extinctions
107 (PAX, Droplet Measurement Technologies Inc., USA). The meteorological parameters
108 and gaseous pollutants were also measured simultaneously. Ozone (O₃), carbon
109 monoxide (CO), nitric oxide (NO), nitrogen oxides (NO_x) and sulfur dioxide (SO₂)
110 were measured using online analyzers (Thermo Fisher Scientific, USA). Ammonia
111 (NH₃) was measured by the Picarro G2103 gas analyzer (Picarro Inc., USA) (Liu et al.,
112 2024). Temperature, RH and other meteorological parameters were monitored by
113 meteorological sensors (GRWS100, Campbell, USA).

114 **2.3 Data Analysis**

115 The AMS data were processed by SQUIRREL (version 1.60P) and PIKA (version 1.20P)
116 from the ToF-AMS Software Downloads Web page ([http://cires.colorado.edu/jimenez-](http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html)
117 [group/ToFAMSResources/ToFSoftware/index.html](http://cires.colorado.edu/jimenez-group/ToFAMSResources/ToFSoftware/index.html)). The ionization efficiency (IE)
118 was calibrated using 300 nm pure ammonium nitrate before and after the campaign.
119 The relative ionization efficiency (RIE) of ammonium was determined from pure
120 ammonium nitrate, yielding a value of 3.52. RIE values for OA, nitrate, sulfate, and
121 chloride were set to their default values of 1.4, 1.1, 1.2, and 1.3, respectively
122 (Canagaratna et al., 2007). As well, the collection efficiency (CE) was assigned a typical
123 value of 0.5 for common environments, although it is recognized that this constant value
124 may introduce additional uncertainty during periods with high nitrate content. Element

125 ratios, including H: C, O: C, N: C and OM: OC, are calculated using the Improved-
126 Ambient method (Canagaratna et al., 2015).

127 In addition, ALW content and aerosol acidity that are associated with inorganic
128 species were estimated by the Extended Aerosol Inorganics Model (E-AIM), which is
129 a well-known inorganic thermodynamic model without simplifying assumptions (Clegg
130 et al. 1998; Wexler and Clegg 2002; Pye et al. 2020). The key inputs to the E-AIM
131 model include temperature, relative humidity, and the concentrations of inorganic ions
132 (sulfate, nitrate, ammonium, chloride) as well as gas-phase ammonia. The main sources
133 of uncertainty stem from thermodynamic parameters and the uncertainties of measured
134 inorganic ions. In this study, the ALW associated with the hygroscopicity of organics
135 was estimated following the method of Nguyen et al. (2016) using the κ -Kohler theory
136 and the Zdanovskii–Stokes–Robinson mixing rule. The estimated organic-associated
137 ALW was only 10.7% of that associated with the inorganic aerosols. Therefore, the
138 organic-associated ALW was neglected in the subsequent analysis.

139 **2.4 Source Apportionment of OA**

140 Positive Matrix Factorization (PMF) analysis was applied to the high-resolution mass
141 spectra of organic matrix for m/z 12 – 120 to resolve distinct OA factors from specific
142 sources (Paatero and Tapper, 1994; Ulbrich et al., 2009). The data and error matrices
143 were treated according to the procedures detailed in DeCarlo et al. (2010). By
144 comparing the mass spectral profiles with previous studies and correlations with time
145 series of tracers, five OA factors with $f_{\text{peak}} = 0$ were selected, including one primary
146 organic aerosol (POA) factor, one nitrogenous OA (NOA) factor, and three SOA factors,

147 namely, less-oxidized oxygenated OA (LO-OOA), more-oxidized OOA (MO-OOA),
148 and aqSOA. The detailed diagnostic plots are shown in Figures S1 and S2 in the
149 supporting information.

150 AqSOA exhibited strong correlations with unique fragment ions that are widely
151 recognized as markers of aqueous-phase secondary products (Sun et al., 2016; Xu et al.,
152 2019). For instance, significant correlations were observed with $C_2O_2^+$ (m/z 56, $r =$
153 0.77), a typical fragment of oxalate-related species, and with CH_3SO^+ (m/z 63, $r = 0.90$),
154 which is indicative of organosulfur compounds (Figure S3). These results provide
155 chemical evidence supporting the aqueous-phase origin of aqSOA in Nanjing. Among
156 these five OA factors, aqSOA exhibits typical characteristics of highly oxidized organic
157 aerosols: fraction of m/z 44 (CO_2^+ , primarily from carboxylic acids and highly oxidized
158 compounds) (Heald et al., 2010) in total signals exceeded that of m/z 43 (typically
159 representing less oxidized compounds) (Figure S2). Elemental analysis further
160 indicated a strongly oxygenated character, with an average H:C ratio of 1.80 and an
161 O:C ratio of 0.78. These values are comparable to those reported in other regions, such
162 as northern Italy and Beijing (Gilardoni et al., 2016; Xu et al., 2019). The N:C ratio
163 (0.05) was also elevated and similar to wintertime values observed in Beijing (N:C =
164 0.045; Xu et al. 2017), suggesting nitrogen-containing compound formation via
165 aqueous-phase processing.

166 **3.Results and Discussions**

167 **3.1 General Characteristics of aqSOA**

168 The meteorological conditions in Nanjing were overall stable during the field campaign

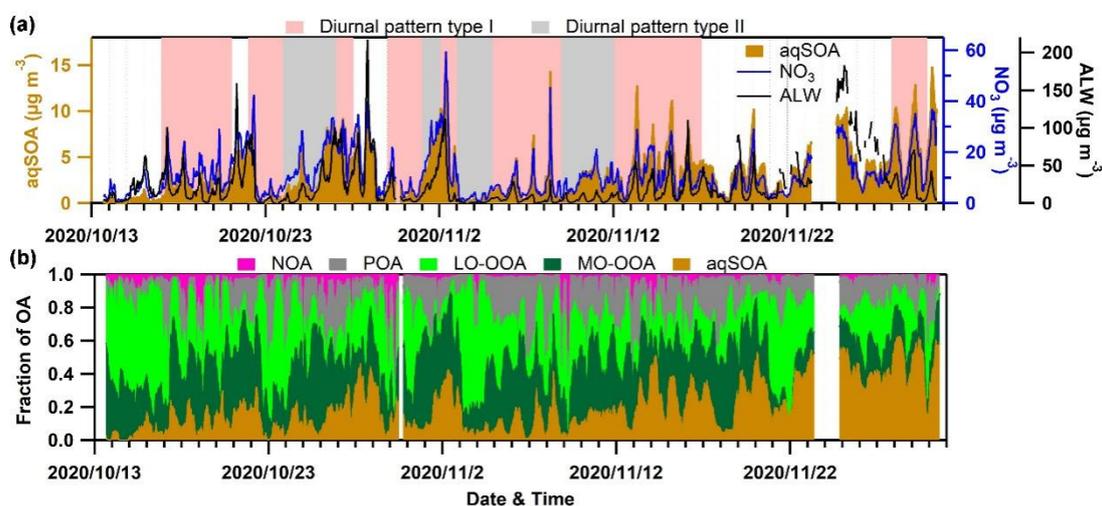
169 (October-December), with an average temperature and RH of 14.6 ± 4.6 °C and $65.1 \pm$
170 17.9% , respectively (Figure S6). The diurnal variation of RH ranged from 50% to 85%,
171 resulting in relatively humid air conditions that favored aqueous-phase chemical
172 reactions. The average wind speed of the prevailing northerly wind was relatively low
173 (0.21 ± 0.16 m/s), resulting in the accumulation of local pollutants. In particular, the
174 average NR-PM₁ concentration was 37.3 ± 20.6 µg/m³, indicating frequently occurred
175 particulate matter pollution during this period.

176 Analysis of aerosol chemical composition revealed that organic aerosol was the
177 dominant NR-PM₁ component in Nanjing, accounting for 40.9% of the total PM₁
178 concentration, evidently higher than nitrate (30.7%) and sulfate (13.9%; Figure S6).
179 The time series of aqSOA showed significant variation, with peak concentrations up to
180 15.7 µg/m³ (Figure 1a). The average concentration of aqSOA during the campaign was
181 3.1 µg/m³, accounting for 20.2% (62.7% in maximum) of the total OA and 27.6% (78.2%
182 in maximum) of the total SOA (Figure 1b). The average fraction of aqSOA in OA was
183 much higher than Beijing (13-17%) (Zhao et al., 2019), indicating an important role of
184 aqueous-phase processes in SOA formation in Nanjing. Moreover, both aqSOA
185 concentrations and their relative contribution to total organic aerosol increased
186 significantly with increasing RH (Figure S7a), which was likely due to the increased
187 availability of the aqueous reaction medium enhanced by water uptake by aerosols,
188 which is crucial for the aqSOA formation. This is further confirmed with the sharply
189 increased contribution of aqSOA to OA (up to 62.7%) at RH levels above 80%,
190 suggesting that high-humidity environments in suburban Nanjing substantially

191 promoted aqSOA formation through enhanced aqueous-phase chemistry.

192 The Van Krevelen diagram (Heald et al., 2010) provided further insights into the
193 chemical aging of OA (Figure S8). The H:C vs. O:C slope was near -0.5, indicating OA
194 oxidation primarily involving carboxylic acid and peroxide or alcohol functional groups
195 addition without fragmentation and/or the addition of carboxylic acid functional groups
196 with fragmentation (Ng et al., 2011). In particular, OA observed under high RH
197 conditions (>80%) clustered in the upper plot region, suggesting distinct chemical
198 evolution of OA when aqueous-phase reactions were involved. Specifically, the nearly
199 zero slope of the relationship among POA, LO-OOA and aqSOA suggests that the
200 observed increase in the O:C ratio of aqSOA may be related to oligomerization and
201 hydroxyl formation through dark chemistry processes (Lim et al., 2010).

202



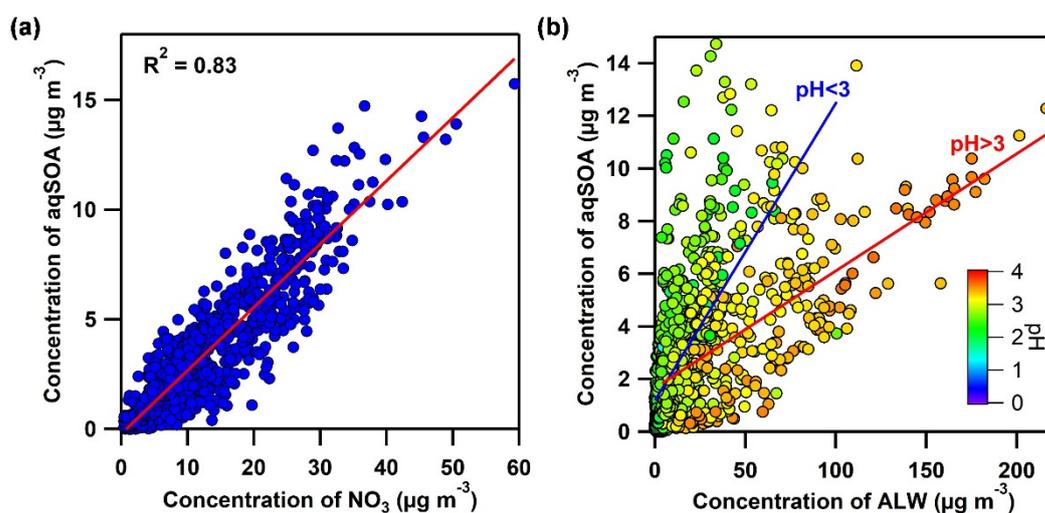
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204 **Figure 1.** (a) Time series of aqSOA and nitrate concentrations. The pink area represents
205 diurnal pattern type I, while the grey area represents diurnal pattern type II. (b) Time
206 series of the fraction of five OA factors.

207 3.2 Enhanced aqSOA Formation Driven by Nitrate, ALW, and Acid

208 Catalysis

209 As a strongly hygroscopic component, nitrate aerosol can enhance aerosol water uptake,
210 thereby modifying the aqueous microenvironment for SOA production (Hodas et al.,
211 2014; Sullivan et al., 2016). We observed strong correlation between aqSOA and nitrate
212 concentrations ($R^2 = 0.83$; Figure 2a) during the observation period, and the variation
213 of aqSOA was highly consistent with that of nitrate aerosol (Figure 1a). This suggests
214 that nitrate aerosols may contribute importantly to aqSOA formation in Nanjing,
215 potentially mediated by their influence on ALW and/or aqueous reactions. We estimated
216 that the average contribution of organic nitrates to the total nitrate was approximately
217 9.5% by assuming a NO^+ to NO_2^+ ratio of 10 for organic nitrates (Farmer et al., 2010).
218 Thus, organic nitrates can be considered negligible in this study.
219



220
221 **Figure 2.** (a) Scatter plot of aqSOA and nitrate concentrations. (b) Scatter plot of
222 aqSOA and ALW concentrations, colored by aerosol pH. The blue line represents the
223 fitted line for data with aerosol $\text{pH} < 3$, while the red line represents the fitted line for

224 data with aerosol pH > 3.

225

226 The correlation between aqSOA and aerosol liquid water (ALW; $r = 0.63$; Figure
227 2b), which is mainly derived from the hygroscopic growth of inorganic salts such as
228 nitrate, was moderate compared to nitrate. Still, their positive relationship indicates that
229 higher ALW levels may enhance aqueous-phase chemical processes by providing the
230 medium in which multiphase reactions can occur (Hodas et al., 2014). The presence of
231 ALW promotes the partitioning of water-soluble organic precursors and allows for
232 subsequent aqueous-phase reactions contributing to aqSOA formation. The results are
233 consistent with previous studies in humid urban environments (Chen et al., 2021; Duan
234 et al., 2022; Kuang et al., 2020).

235 In addition to ALW, aerosol acidity also exerted a strong influence on aqSOA
236 yields during the campaign (Lim et al., 2010; Tilgner et al., 2021). To illustrate, the
237 dataset was separated by aerosol pH (pH < 3 vs. pH > 3) (Figure 2b), the slopes of
238 ALW-aqSOA correlations decreased with increasing pH, indicating that aqSOA
239 production was higher under more acidic conditions at the same ALW level. This
240 pattern demonstrates the importance of acid-catalyzed reactions in driving aqSOA
241 formation. Previous studies have demonstrated that under acidic conditions (low pH,
242 high H^+ concentration), non-oxidative aqueous organic chemical processes, such as
243 accretion reactions (aldol condensation, hemiacetal and acetal formation, and the
244 esterification of carboxylic acids) are important formation pathways of aqSOA
245 (Freedman et al., 2019; Tilgner et al., 2021). For example, the hydration of

246 methylglyoxal and its subsequent acetal formation are highly pH-dependent, requiring
247 a pH of less than 3.5 to occur (Yasmeen et al., 2010). As aerosol pH increases (lower
248 H⁺ availability), the catalytic efficiency of these acid-driven reactions diminishes,
249 leading to a reduction in aqSOA production efficiency. Collectively, these findings
250 suggest that aerosol acidity plays a pivotal regulatory role in aqSOA formation, linking
251 aqueous chemistry to the broader context of aerosol physicochemical properties in
252 Nanjing.

253 Thus, our measurements revealed a synergistic interplay among nitrate, ALW, and
254 aerosol acidity in regulating aqSOA formation in Nanjing. Nitrate enhances the ALW
255 content, which in turn promotes aqueous-phase reactions, while aerosol acidity governs
256 the efficiency of these chemical processes. These processes together constituted the
257 formation mechanism of aqSOA during the observation period.

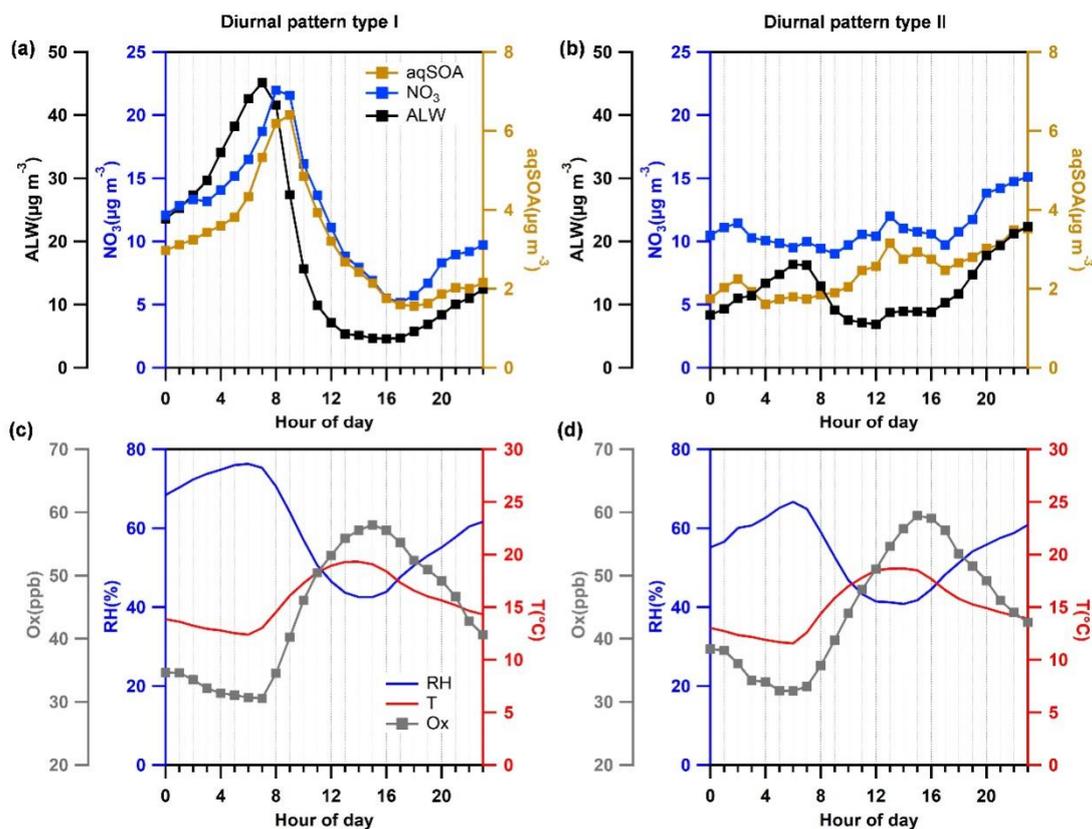
258 **3.3 Synergistic Role of Aqueous and Photochemical Processes**

259 During the campaign, the average diurnal variation of aqSOA in Nanjing exhibited a
260 distinct morning peak at approximately 09:00 local time (Figure S2d). This pattern is
261 different from most of the previous urban observations, where aqSOA concentrations
262 typically peak during nighttime periods (Sun et al., 2016; Xu et al., 2019; Gu et al.,
263 2023). To address this, the time series of aqSOA was analyzed on a daily basis, with
264 rainy days excluded. Based on this analysis, two distinct diurnal variation patterns were
265 identified: type I (22 days) and type II (9 days) (Figure 1a). For diurnal pattern type I,
266 the daily variation of aqSOA exhibits a pronounced morning peak, whereas the daytime
267 concentration of aqSOA shows a noon peak for diurnal pattern type II (Figure 3). These

268 two distinct diurnal patterns of aqSOA could be attributed to multiple formation
269 mechanisms of aqSOA or different meteorological conditions. It should be noted that
270 the observed morning peak around 09:00–10:00 may be modulated not only by
271 chemical production but also by concurrent physical processes, such as morning
272 boundary-layer evolution and mixing/dilution effects. Nevertheless, the diurnal patterns
273 and classification discussed herein are interpreted primarily under chemically
274 constrained conditions, and the main conclusions about aqSOA formation pathways
275 rely chiefly on chemical relationships and correlations.

276 During diurnal pattern type I, aqSOA concentrations exhibited a pronounced
277 single peak between 09:00 and 10:00 local time, following a period of continuous
278 nighttime accumulation from approximately 20:00 to 07:00 (Figure 3a). Notably, the
279 aqSOA peak lagged behind the nitrate peak by about one hour (based on the average
280 peak-time difference), and nocturnal increases in nitrate, ALW, and aqSOA were highly
281 synchronized. This ~1 h lag may be attributed to the kinetic limitations of the nitrate-
282 mediated aqSOA formation processes. Compared with type II, nighttime
283 meteorological conditions during type I were characterized by higher RH (73.1% vs.
284 61.2%) and ALW concentrations (34.2 $\mu\text{g}/\text{m}^3$ vs. 12.6 $\mu\text{g}/\text{m}^3$). Such humid conditions
285 favor the formation of both nitrate and ALW, thereby enhancing aqSOA production
286 through aqueous-phase reactions. Moreover, the higher RH conditions observed during
287 type I events could further promote aqueous reaction pathways, as the nitrate formation
288 from N_2O_5 hydrolysis on aqueous aerosol surfaces is strongly facilitated under such
289 conditions (Sun et al., 2018). Overall, these observations indicate that aqSOA

290 production under type I conditions is predominantly controlled by aerosol aqueous-
 291 phase processes, highlighting the important role of nocturnal multiphase chemistry in
 292 driving morning aqSOA peaks.
 293



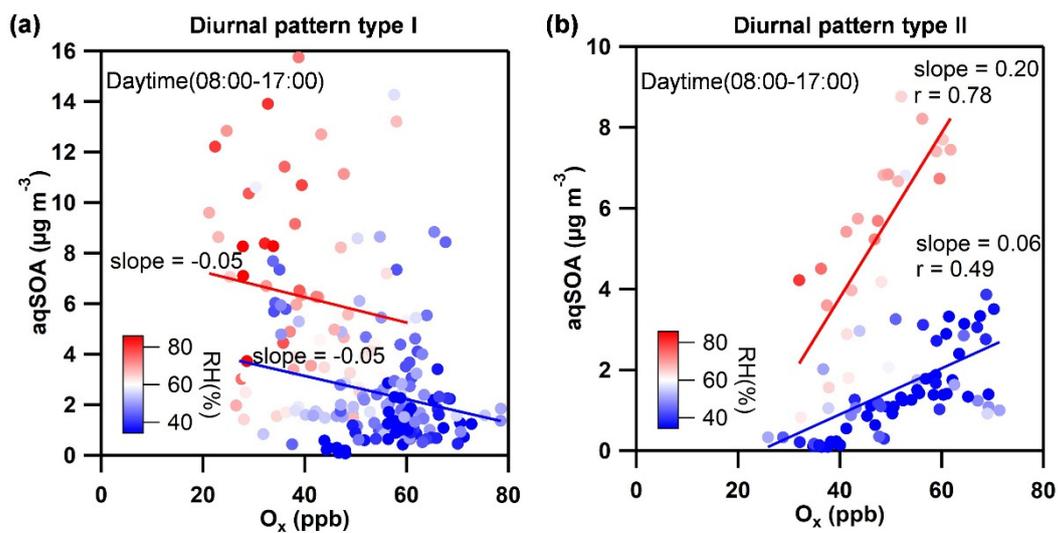
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 295 **Figure 3.** Diurnal variations for SOA, ALW, and nitrate in diurnal pattern types I (a)
 296 and II (b). Diurnal variations for temperature, humidity, and O_x concentration in diurnal
 297 pattern types I (c) and II (d).

298

299 In contrast, aqSOA concentrations gradually increased during the daytime and
 300 peaked around 13:00 local time (Figure 3b) during diurnal pattern type II, indicating an
 301 influence of photochemical processes. The diurnal variation of aqSOA is consistent
 302 with that of nitrate aerosols. This suggests that enhanced photochemical activity is

303 associated with nitrate aerosol formation, which may subsequently mediate aqSOA
 304 production via water uptake. O_x is commonly regarded as a tracer for photochemical
 305 processes. The relationship between daytime (08:00–17:00) aqSOA concentrations and
 306 O_x levels was examined for both diurnal patterns (Figure 4). For diurnal type I, no
 307 significant correlation was observed, suggesting that daytime photochemistry played a
 308 minimal role in aqSOA production under these conditions. In contrast, diurnal type II
 309 displayed a positive correlation, implying that aqSOA formation was partially related
 310 to the photochemical processes.

311



312

313 **Figure 4.** Scatter plots of aqSOA and O_x for Diurnal pattern types I (a) and II (b). Data
 314 points are colored by RH, with red and blue lines representing fits for $RH > 60\%$ and
 315 $RH < 60\%$ respectively. For diurnal pattern type I, the correlation coefficient was -0.22
 316 ($N = 151$, $p < 0.01$) during daytime with $RH < 60\%$, and $r = -0.14$ ($N = 59$, $p = 0.3$)
 317 when $RH > 60\%$. For diurnal pattern type II, the correlation coefficient was 0.49 ($N =$
 318 67 , $p < 0.01$) under $RH < 60\%$, and $r = 0.78$ ($N = 23$, $p < 0.01$) under $RH > 60\%$.

319

320 Further analysis revealed that the slope of the aqSOA-O_x relationship under high
321 RH conditions in diurnal type II was approximately three times higher than that under
322 low RH conditions. This is because higher RH likely enhanced the partitioning of semi-
323 volatile photochemical oxidation products into the aqueous phase and promoted
324 aqueous-phase photochemical reactions, thereby facilitating aqSOA formation. These
325 results indicate that daytime aqSOA production during type II is governed by the
326 combined effects of photochemical oxidation and humidity-dependent aqueous
327 processing.

328

329 **4. Conclusions**

330 We analyzed the characteristics and formation mechanisms of aqSOA in the
331 autumn season in Nanjing, where aqSOA was found to constitute a significant fraction
332 of organic aerosols (20.2%). The results demonstrate that nitrate, ALW, and aerosol
333 acidity act in concert to regulate aqSOA formation, with nitrate serving as a critical
334 driver through its strong hygroscopicity. We observed the occurrence of a distinct
335 morning peak of aqSOA in the YRD region, a feature that differs from most of the
336 previously reported studies in urban environments. Based on diurnal variation analysis,
337 two distinct diurnal patterns were identified: type I, dominated by nighttime aqueous-
338 phase chemistry linked to nitrate and ALW accumulation; and Type II, shaped primarily
339 by daytime photochemical oxidation. These findings highlight that aqSOA formation
340 in this megacity is governed by the synergistic effects of both nocturnal aqueous

341 reactions and daytime photochemical processes.

342 China is strictly controlling SO₂ emissions, which has led to a decrease in sulfate
343 content in atmospheric particulate matter. Nitrate is becoming a more important
344 inorganic component of PM_{2.5}, and its role in promoting aqSOA formation will become
345 even more significant. Although this study did not directly resolve the specific
346 precursor compounds of aqSOA, our results still highlight the importance of paying
347 greater attention to nitrate-driven aqueous processes in future air quality assessments.
348 In particular, coordinated management of nitrogen oxides (NO_x) alongside traditional
349 particulate matter controls may provide an effective pathway for mitigating aqSOA
350 burdens in a post-sulfate-dominated atmosphere.

351

352

353 **Author contributions**

354 T.L. and A.D. designed the research project; J.W., D.G., C.Z., C.R., W.X., J.W., W.N.,
355 X.C., S.L. and X.H. performed the research; Q.W., Q.Z., and X.Q. analyzed data; Q.W.,
356 Q.Z., and T.L. wrote the paper. All authors participated in the relevant scientific
357 discussion and commented on the manuscript.

358 **Code/Data availability**

359 All data are available upon request to the corresponding author.

360 **Competing interests**

361 The authors declare no competing interests.

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366

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