



1 **CHANS-SD-YRB V1.0: A System Dynamics model of**
2 **the coupled human-natural systems for the Yellow**
3 **River Basin**

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20
21 **Abstract:** Modeling the coupled human–natural systems (CHANS) is vital for
22 understanding human–natural interactions and achieving regional sustainability,
23 offering a powerful tool to alleviating human–water conflicts, ensuring food security,
24 thereby supporting the region’s pathway toward sustainable development. However,
25 the scarcity of regional-scale CHANS models constrains progress in practical
26 applications for regional sustainability. The Yellow River basin (YRB) is an ideal region



for modeling regional CHANS due to its closely coupled human and natural systems, which are stressed by water and ecosystem fragility. Here, we developed the CHANS-SD-YRB model using the System Dynamics approach, integrating 10 sectors essential for modeling human-water interactions of the basin, including five human sectors (Population, Economy, Energy, Food, and Water Demand) and five natural sectors (Water Supply, Sediment, Land, Carbon, and Climate). The model can simulate evolution and feedbacks of the YRB CHANS annually at provincial and sub-basin scales, while conserving hydrological connectivity between sub-basins. The model can accurately reproduce historical CHANS dynamics, achieving strong quantitative agreement with historical data ($R > 0.95$ for human sectors and $R > 0.7$ for natural sectors), which supports its applicability for scenario analyses and future projections. We applied the model to explore human–natural system dynamics under a future baseline scenario, assuming the continuation of existing policies and climate projection under middle of the road scenario (SSP–RCP 2-4.5). The future projections (2021-2100) indicate that achieving sustainable development in the YRB will remain challenging, though economic growth and food security are expected to improve. Emerging issues, such as ecological–human water trade-offs, labor shortages, reduced sediment load, and limited carbon absorption capacity, may hinder regional long-term sustainability.

Keywords: coupled human-natural systems, regional modeling, system dynamics, the Yellow River Basin

1 Introduction

Coupled human and natural systems (CHANS) emphasize the reciprocal feedback and co-evolution between human and natural systems, offering an integrated framework for diagnosing complex problems and guiding sustainable development (Fu and Li, 2016). While global-scale CHANS research has deepened our knowledge of dynamic feedbacks among Earth’s spheres and system evolution under climate change, sustainability challenges often manifest at regional scales, where social and ecological dynamics are more intricately intertwined (Liu et al., 2007). Compared to global



55 CHANS, regional CHANS are open systems that continuously exchange energy and
 56 materials with other regions and the global system, e.g., water resources and electricity
 57 transfers (Dobbs et al., 2023; Zhang et al., 2022a) and trade (Ristaino et al., 2021),
 58 resulting in pericoupling and telecoupling of different systems (Liu, 2017). Besides,
 59 regional CHANS are shaped by more immediate and complex human influences and
 60 stressors, such as urban expansion (van Vliet, 2019), ecological protection (Xu et al.,
 61 2017; Yang et al., 2022), and water resource regulation policies (diversion and
 62 allocation) (Song et al., 2024), which alter regional CHANS dynamics. Moreover, due
 63 to their diverse ecological and socioeconomic resilience, regional CHANS exhibit
 64 heterogeneous responses to external weather events and climate change, as evidenced
 65 by differing responses in crop yield (Hasegawa et al., 2021) and economic production
 66 to extreme heat and warming (Waidelich et al., 2024). By capturing dynamic
 67 interactions of interconnected components, CHANS theories and models enable more
 68 effective policies and interventions that align ecological integrity with socioeconomic
 69 progress (Motesharrei et al., 2016; Verburg et al., 2016). As such, advancing CHANS
 70 research is essential for informing adaptive strategies in the face of regional growing
 71 environmental and societal pressures.

72 Modeling of CHANS is a frontier area of geographical science and has become a
 73 key approach in resources, environment, and climate research. CHANS models serve
 74 as essential tools for analyzing the complex dynamics of human–natural interactions
 75 and guiding pathways toward sustainability. Numerous modeling approaches have been
 76 developed to simulate human–natural interactions at global scale, including system
 77 dynamics (SD) models (e.g., ANEMI (Breach and Simonovic, 2021), FeliX (Rydzak et
 78 al., 2013; Ye et al., 2024), and FRIDA (Rajah et al., 2025)), integrated assessment
 79 models (IAMs) (Vaidyanathan, 2021), integrated Earth system model (iESM) (Jain et
 80 al., 2022) and synchronously coupled human component in the earth system model
 81 (E3SM-GCAM) (Di Vittorio et al., 2025). These models effectively characterize
 82 human–natural interactions at the global scale, and have been applied to assess the



83 impacts of climate change on human society (e.g., agriculture (Monier et al., 2018),
84 economic damage (Wang et al., 2020b), fatalities increase and welfare loss (Dottori et
85 al., 2018)), and humans' feedback on the Earth system, such as those from climate
86 mitigation on water and food security (Cheng et al., 2022; Fujimori et al., 2022).
87 However, global models cannot represent the complex dynamics of human–natural
88 interactions at the regional level, due to the openness of regional CHANS (Verburg et
89 al., 2016), diversified regional human and natural processes (Clark et al., 2015), with
90 locally heterogeneous responses to disturbances (Baldos et al., 2023). Moreover, the
91 coarse spatiotemporal resolution of global models limits their capacity to support
92 effective decision-making for regional development (X. Li et al., 2018).

93 To address this limitation, many regional CHANS models have been developed at
94 various regional scales (e.g., national, basin, and urban) using System Dynamics (SD)
95 and agent-based modeling (ABMs) techniques. Notable examples include the ANIME-
96 Yangtze model (Jiang et al., 2022), the T21-China (Qu et al., 2020), and the iSDG-
97 Australia model (Allen et al., 2019), all based on SD, the Jordan Water Model (Yoon et
98 al., 2021) based on ABM, as well as integrated models in the San Juan River Basin
99 (Hyun et al., 2019) and the Heihe River Basin (Li et al., 2021). These models are
100 designed to capture finer-scale dynamics and region-specific human–natural
101 interactions, since they embed localized characteristics (e.g., fishing ban, reservoir
102 operation strategies, demographic policies, transboundary flows) and account for
103 heterogeneity overlooked by global models. As a result, regional CHANS models offer
104 stronger policy relevance, providing actionable insights for national, basin, and urban
105 decision-making, and advancing CHANS research across multiple scales.

106 The Yellow River Basin (YRB) in China is one of the regions where conflicts
107 between human and natural systems are most acute and complex, particularly in terms
108 of human–water relations, due to the severe imbalance between socioeconomic
109 development and natural hydrological, ecosystem processes. The YRB faces severe
110 water stress, with the water resource utilization rate exceeding 80% (Feng and Zhu,



2022; Zhang et al., 2022b). Intensive water extraction has triggered a series of ecological and environmental issues, including flow interruptions, water pollution, and declining groundwater levels, all of which in turn constrain socioeconomic development. The Yellow River traverses the Loess Plateau (Zhu et al., 2019), where severe soil erosion makes the Yellow River one of the most sediment-laden rivers globally (Fu et al., 2011; Yin et al., 2021). The pronounced spatial and temporal variability in streamflow and sediment load leads to significant riverbed aggradation, frequent flooding, and disruption of agricultural production and other livelihood activities (Miao et al., 2016). Due to internal hydrological connectivity, sub-basins are highly interconnected and are susceptible to upstream influences. Upstream water overuse diminishes downstream availability (Wei et al., 2023), a factor that played a major role in flow interruptions during the 1990s (Changming and Shifeng, 2002; Wang et al., 2019). Ecological challenges differ across the subbasin, with the upstream facing ecosystem degradation and limited water retention (Ning et al., 2022), the midstream characterized by soil erosion and large-scale ecological restoration (Fu et al., 2011), and the downstream focusing on wetland conservation (Fu et al., 2023). Policy measures aimed at ecological restoration, such as afforestation and cropland conversion, have increased vegetation cover, reduced sediment loads but also decreased runoff, exacerbating water scarcity (Feng et al., 2016; Wang et al., 2016). These interlinked dynamics underscore the YRB as a complex coupled human–natural system, where addressing environmental challenges requires an integrated, systems-oriented approach.

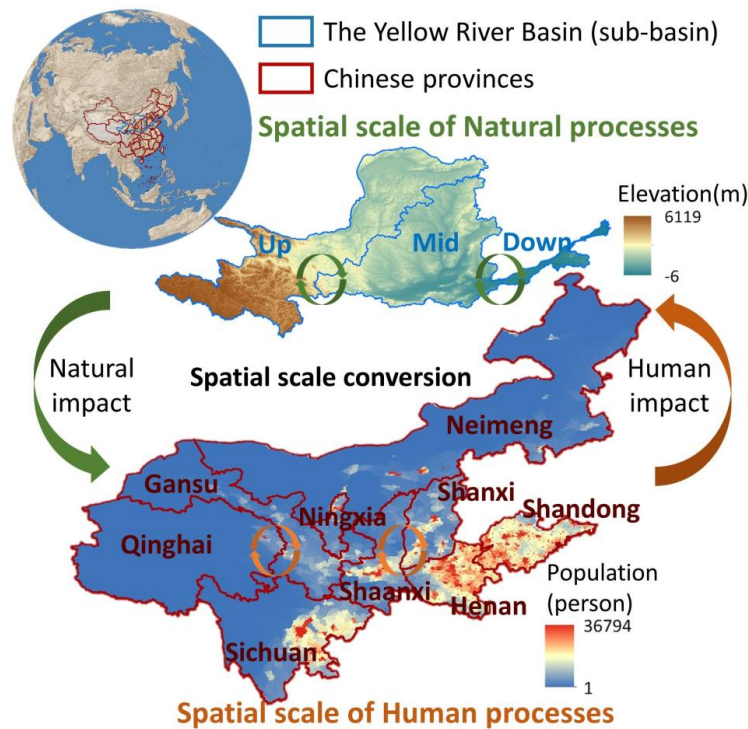
The existing models for the YRB are typically designed for specific problems with a narrow application focus and only represent a limited set of human and natural components within the CHANS. These include limited nature-to-human impact pathways, e.g., low flows threatening farmers’ livelihoods (Liu et al., 2008), the damage of floods and droughts on agriculture (Zhang et al., 2015), as well as human-to-nature impact pathways, e.g., effects of ecological restoration policy on hydropower and water–sediment–carbon dynamics (Wu et al., 2025; Yan et al., 2024), and the impacts



139 of irrigation water-saving and salinity-control practices on crop yield and water
140 productivity (Wu et al., 2023). These models focus on isolated components of CHANS,
141 with limited consideration of fully coupled human–natural interactions, which limits
142 their capacity to represent full human–natural interactions and support regional
143 decision-making.

144 To address the gap in CHANS modeling for the YRB, following our previously
145 proposed CHANS modeling framework for the basin (Sang et al., 2025), we
146 implemented the framework to develop the coupled human and natural systems model
147 for the YRB (CHANS-SD-YRB) using the System Dynamics approach. Through
148 dynamic interaction with policies, climate change, human activities, and environmental
149 feedbacks, the CHANS-SD-YRB model provides a platform for predicting system
150 dynamics, conducting scenario analyses, evaluating policies, and optimizing water-
151 food-carbon synergies. This study offers both theoretical and practical insights for
152 advancing regional CHANS modeling and promoting sustainable development in the
153 YRB.

154 **2 Description of the CHANS-SD-YRB**



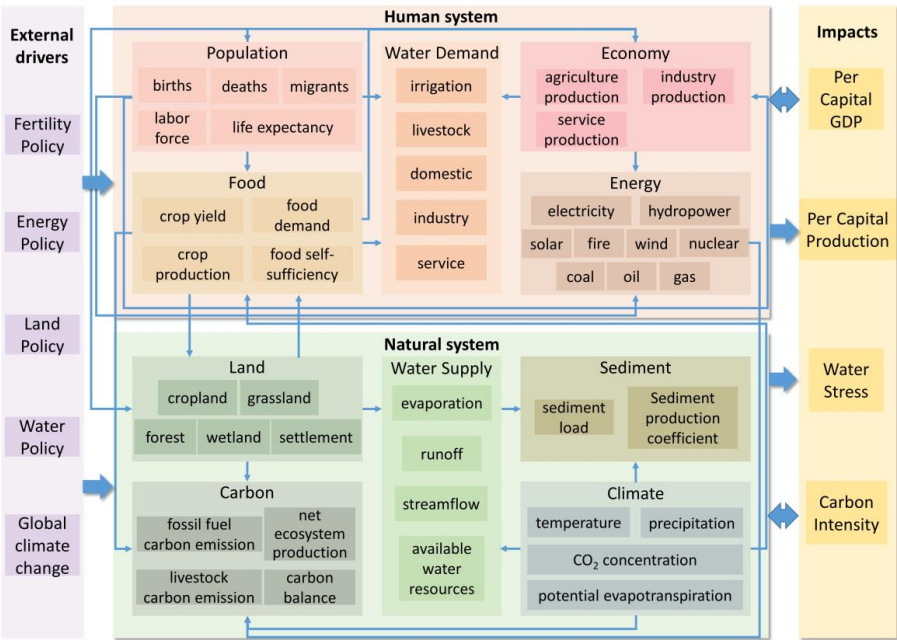
155
156 Fig.1 Geolocation of the Yellow River Basin and boundary of the natural and human
157 processes in the CHANS-SD-YRB model. Natural processes are simulated at the sub-
158 basin scale (base map: elevation), and human processes at the provincial scale across
159 nine provinces (base map: population density in 2020).

160 We developed the CHANS-SD-YRB based on system dynamics, a method well-
161 suited for capturing complex system behaviors characterized by nonlinearity, multi-
162 level structures, and feedback loops (Akhtar et al., 2013). The model was constructed
163 and implemented using the VENSIM DSS (Ventana Systems, 2023) software platform,
164 operating on an annual time step. The CHANS-SD-YRB simulates both human and
165 natural processes for historical simulations (1981–2020) and future projections (2021–
166 2100). Human processes are simulated at the provincial scale, covering the nine
167 provinces along the Yellow River (Qinghai, Sichuan, Gansu, Ningxia, Neimeng,



168 Shaanxi, Shanxi, Henan, and Shandong), while natural processes are simulated at the
169 sub-basin scale (up-, mid-, and downstream) (Fig.1). The model is designed to capture
170 the various interactions within and between different components of human system and
171 natural system across administrative and hydrological units. The spatial scale
172 conversion between provincial and sub-basin levels relies on weights (e.g., the
173 proportion of sub-basin-level values to provincial total values) derived from historical
174 high-resolution gridded datasets of human-related variables. These weights enable
175 disaggregation of provincial outputs to the sub-basin level (see Supporting Information
176 S4 for details). Given the availability of gridded data for human processes and the
177 strong correlations among relevant variables, gridded population and GDP data were
178 used as proxies to disaggregate provincial-level outputs to the sub-basin scale (Table
179 S3).

180 **2.1 Model structure**



181
182 Fig.2 Structure of the CHANS-SD-YRB, which shows sectors of human and natural
183 systems, their key processes and interactions.



184 Drawing on the modeling framework of CHANS in the YRB (Sang et al., 2025),
 185 we designed the CHANS-SD-YRB structure (Fig. 2), including five sectors related to
 186 human society (*Population, Economy, Energy, Food, and Water Demand*), and five
 187 sectors related to natural ecosystem (*Water Supply, Sediment, Land, Carbon, and*
 188 *Climate*).

189 These sectors are interconnected to represent various human-natural interactions
 190 as summarized below. Key interactions among human system modules are summarized
 191 below. The *Population* sector affects food demand (*Food*), residential water use (*Water*
 192 *Demand*), household electricity and gas consumption (*Energy*), and settlement land area
 193 (*Land*). The sector also interacts dynamically with the *Economy* sector, where economic
 194 output influences deaths and migrants, while the labor force, in turn, drives economic
 195 production. The *Economy* sector drives energy uses (electricity, coal, oil, and gas use)
 196 from *Energy*, as well as industrial and service water withdrawal from *Water Demand*.
 197 The gross agricultural production in the *Economy* is made up of crop and livestock
 198 production (*Food*). The *Energy* sector produces fossil fuel emissions in the *Carbon*
 199 sector. The *Food* sector is affected by the *Land* and *Climate* sector, and it also
 200 determines irrigation water withdrawal (*Water Demand*) and livestock-related
 201 emissions (*Carbon*). Additionally, the *Food* sector interacts closely with the *Land* sector,
 202 where crop production depends on cropland area, which, in turn, is influenced by food
 203 self-sufficiency. The *Water Demand* sector affects streamflow in the *Water Supply*
 204 sector through consumptive water use.

205 The key interactions among natural system modules are listed below. The *Land*
 206 sector influences evapotranspiration in the *Water Supply* sector through vegetation
 207 coverage, and it influences carbon absorption in *Carbon* sector through land use area.
 208 The carbon absorption is also affected by climatic variables including temperature,
 209 precipitation, and CO₂ concentration. Similarly, runoff in the *Water Supply* sector is
 210 affected by precipitation, precipitation intensity (mm/h, the rate of rainfall within one
 211 hour calculated from daily data), and potential evapotranspiration. Streamflow



212 influences sediment dynamics in the *Sediment* sector together with the *Climate* sector.

213 In addition, the CHANS dynamics in the YRB are modulated by external drivers,
214 including policies and global climate change that affect various modeled processes (e.g.,
215 fertility, energy, land use, and water). With comprehensive representation of CHANS
216 processes and their interactions, the CHANS-SD-YRB model is capable of generating
217 integrated indicators to assess the state of the coupled system, such as per capita GDP,
218 per capita food production, water stress, and carbon intensity. These indicators not only
219 serve as evaluation metrics but also feed back to influence the internal dynamics of the
220 human–natural system in the YRB.

221 **2.2 Sector description**

222 The CHANS-SD-YRB model focuses on the essential human–natural processes in
223 the YRB, with a particular emphasis on human–water interactions. To this end, it has a
224 comprehensive representation of the full range of natural and human processes that
225 influence water use and supply. Formulation of each sector aims to explicitly account
226 for cross-sectoral interactions as fully as possible while remaining sufficiently simple
227 to be implemented within the SD software. Considering data availability and the spatial
228 scales of process, human sectors (*Population*, *Economy*, *Energy*, *Food*, and *Water*
229 *Demand*) are simulated at the provincial level, while natural sectors (*Water Supply*,
230 *Sediment*, *Land*, *Carbon*, and *Climate*) are simulated at the sub-basin level within the
231 YRB. Next, we describe each sector and its key formulation and provide full details in
232 Supporting Information S2.



2.2.1 Population

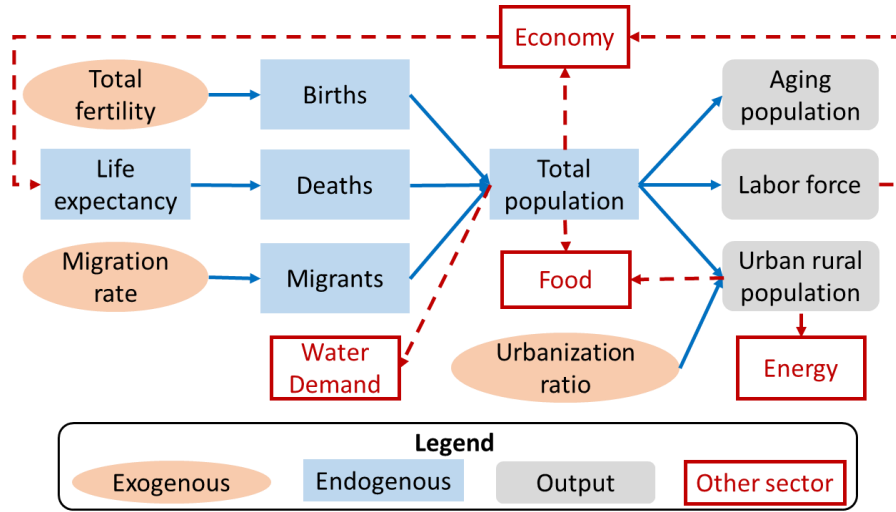


Fig.3 Structure of the *Population* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

Population dynamics are represented by birth, death, and migrants, which are determined by a series of social and economic factors. The total population is characterized by age and gender with exogenous urbanization ratios. The output of *Population* sector drives *Economy*, *Energy*, *Food*, and *Water Demand* sectors through labor force, urban and rural populations.

The key variable, total population, is modeled using the age-structured mathematical method (Kemei et al., 2024), which categorizes individuals by one-year age group and gender (Equation 1),

$$Pop_{g,a} = IniPop_{g,a} + \begin{cases} \int (B_{g,a} + NM_{g,a} - D_{g,a})dt, & a = 0 \\ \int (Pop_{g,a} + NM_{g,a} - D_{g,a})dt, & 1 \leq a \leq 99, a = 100 \text{ and over} \end{cases} \quad (1)$$

where *Pop* is population, subscript *g* and *a* are gender (*g* = M is male, *g* = F is female) and age (*a* = 0-100 and over); *IniPop_{g,a}* is the population in the initial year, *B_{g,a}* is the births (Equation S2), *D_{g,a}* is the deaths (Equations S3-S6), and *NM_{g,a}* is the net migrants (i.e., immigrants – emigrations, Equations S7-S8).



2.2.2 Economy

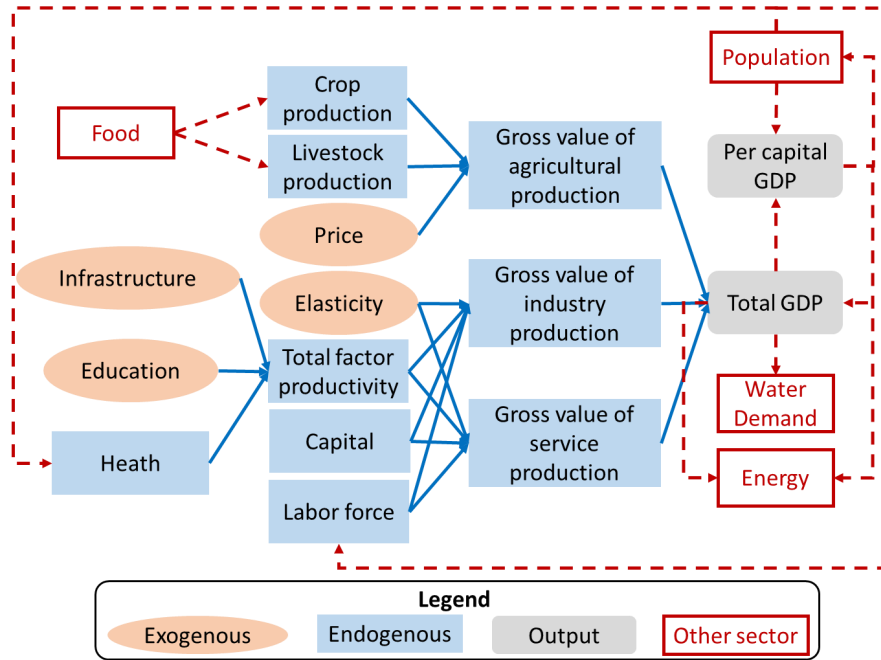


Fig.4 Structure of the *Economy* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Economy* sector simulates production activities in agriculture, industry, and services, which in turn drive changes in the *Population*, *Energy*, and *Water Demand* sectors.

The gross product of industry and services is calculated using the Cobb-Douglas production function (Cobb and Douglas, 1928) to account for multiple factors in the economy (Equation 2). Exogenous variables (infrastructure, education, and elasticity), and endogenous variables (health and labor force) from the *Population* sector (Equations S10-S17) all affect the gross domestic product in industry and service (GDP_s),

$$GDP_s = IniGDP_s \times TFP_s \times \left(\frac{L_s}{IniL_s}\right)^{1-\alpha_s} \times \left(\frac{K_s}{IniK_s}\right)^{\alpha_s} \quad (2)$$

where subscript s represents industry and service, $IniGDP_s$ is the initial GDP; TFP_s is the total factor productivity; L_s represents the labor force and $IniL_s$ is its initial value



(from *Population* sector); α_s is capital and $1-\alpha_s$ is labor elasticities, from T21-China (Qu et al., 2020); K_s and $IniK_s$ refer to the capital stock and its initial level.

Gross agricultural production includes crop and livestock production from *Food* sector, calculated by their respective prices (Equations S18-S19).

2.2.3 Energy

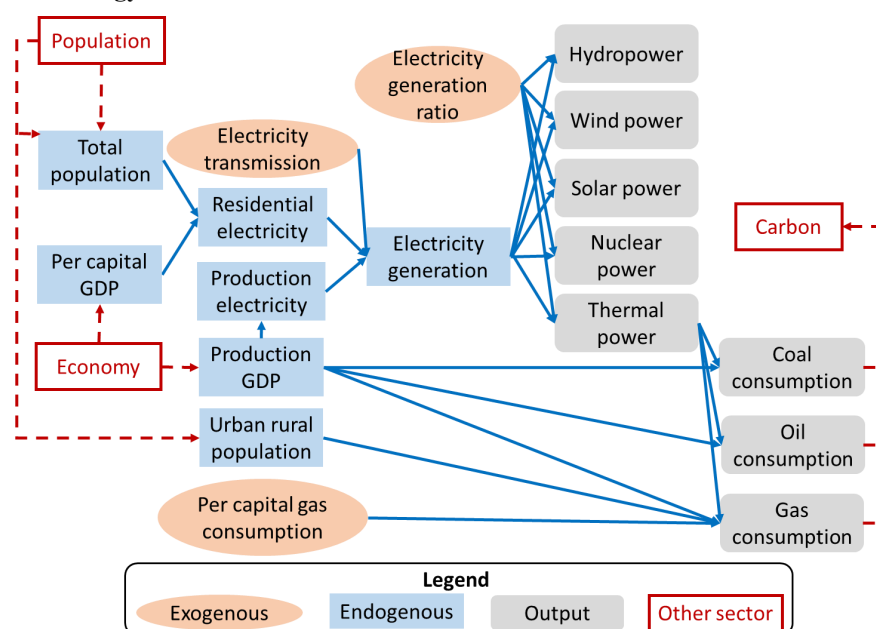


Fig.5 Structure of the *Energy* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Energy* sector simulates production, consumption, and the structure of energy, encompassing coal, oil, gas, and electricity. Coal, oil, and gas consumption are derived from linear relationships between historical sectoral GDP and corresponding consumption (Equations S22-S24). Electricity generation is divided into residential and industrial uses, the former is estimated from the linear fit of historical per capita GDP and residential demand, and the latter is calculated using electricity intensity data from the China Energy Statistical Yearbook (NBSC, 2020a) (Equations S20-S21). Cross-provincial electricity transmission is also incorporated according to the same yearbook. The shares of electricity generated from thermal, hydro, wind, solar, and nuclear



sources are determined by exogenously specified ratios, as reported in the China Energy Statistical Yearbook (NBSC, 2020a). Fossil fuel consumption drives carbon emissions in the *Carbon* sector.

2.2.4 Food

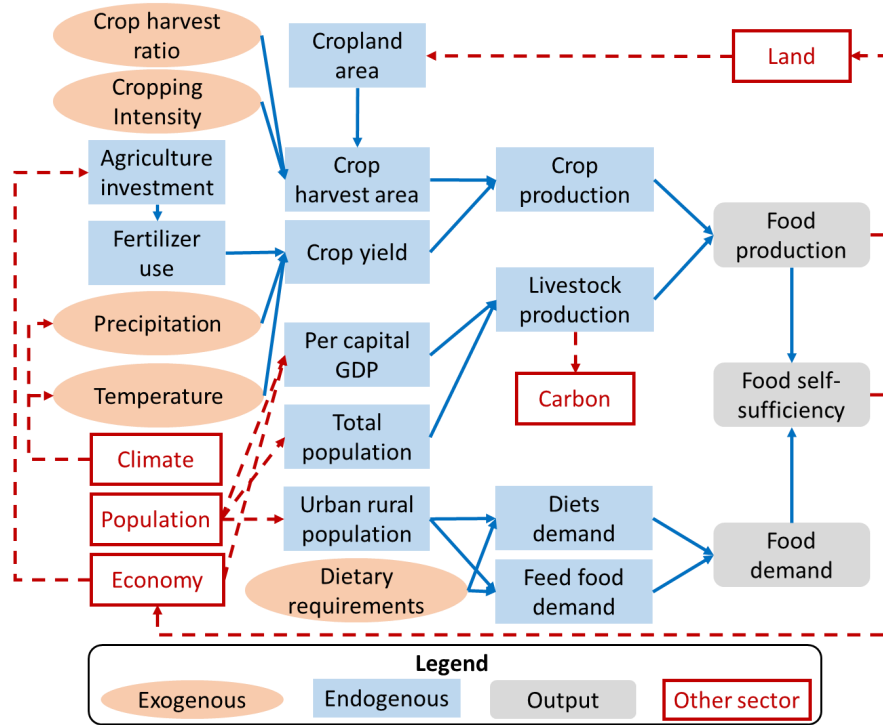


Fig.6 Structure of the *Food* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Food* sector simulates the production of livestock and crops, and food demand, and these productions directly affect gross agricultural production in the *Economy* sector. Livestock production ($Livestock_{pro}$) is calculated by an empirical function, driven by economic development and population growth (Equation 3),

$$Livestock_{pro} = Pop \times \left(Para_{GL1} + (Para_{GL2} - Para_{GL1}) \times \frac{PC_{GDP}}{PC_{GDP} + Const_{GL}} \right) \quad (3)$$

where $Para_{GL1}$, $Para_{GL2}$ and $Const_{GL}$ are parameters obtained by fitting the historical per capita meat production and per capita GDP (PC_{GDP}) data.

Crop production ($Crop_{pro}$) is determined by the yield ($Yield_c$) and harvest area (PA_c)



of seven major crop types (subscript c): rice, wheat, corn, soybeans, cotton, potatoes, and oil crops (Equation 4). The harvest area is influenced by cropland area from the *Land* sector, exogenous cropping intensity and crop harvest ratio. Crop yields are positively affected by agricultural investment from the *Economy* sector, along with effects of precipitation, temperature, and CO_2 concentration from the *Climate* sector (Equations S26-S31).

$$\text{Crop}_{pro} = \sum_{c=1}^7 \text{Yield}_c \times \text{PA}_c \quad (4)$$

Food demand encompasses both staple and feed grain demand, which are determined by population size from *Population* sector and dietary patterns (Equations S33-S36).

2.2.5 Water Demand

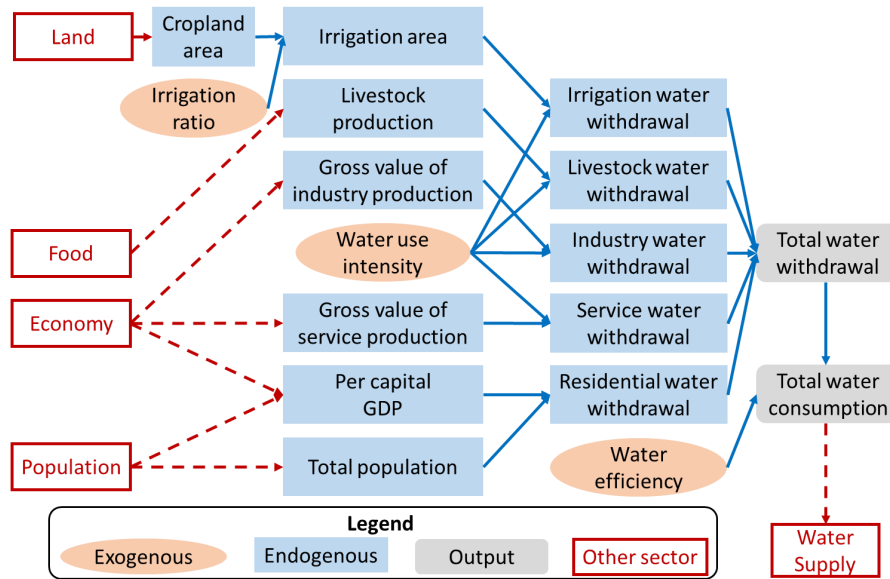


Fig.7 Structure of the *Water Demand* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Water Demand* sector simulates water withdrawal (WW) and consumption (WC) across multiple uses—irrigation, livestock, industry, services, and residential—in the nine provinces of the YRB. Water consumption is derived as the product of water



316 withdrawal and water use efficiency reported in the Water Resources Bulletin
 317 (YRCCMWR, 2020) (Equations S37-S38), which affects the *Water Supply* sector.

318 Irrigation water withdrawal (WW_{irr}) is estimated based on exogenous irrigation
 319 water use intensity (WWI_{irr}), and irrigation cropland ratios (IR) from the China
 320 Agricultural Yearbook (MAARA, 2020), and cropland area ($Area_{Cropland}$) provided by
 321 the *Land* sector (Equation 5).

$$322 \quad WW_{irr} = WWI_{irr} \times Area_{Cropland} \times IR \quad (5)$$

323 The water withdrawal for livestock, industry, and services is also driven by
 324 exogenous sectoral water use intensities collected from the China Statistical Yearbook
 325 and National Long-term Water Use Dataset of China (Zhou et al., 2020), in combination
 326 with livestock production from the *Food* sector, economic output from the *Economy*
 327 sector (Equations S40-S43).

328 Residential water withdrawal (WW_{res}) is obtained from the empirical function,
 329 driven by economic development and population growth (Equation 6),

$$330 \quad WW_{res} = Pop \times \left(Para_{GD1} + (Para_{GD2} - Para_{GD1}) \times \frac{PC_{GDP}}{PC_{GDP} + Const_{GD}} \right) \quad (6)$$

331 where $Para_{GD1}$, $Para_{GD2}$, and $Const_{GD}$ are parameters obtained by fitting the historical
 332 per capita domestic water use with per capita GDP (PC_{GDP}).

333 2.2.6 Water Supply

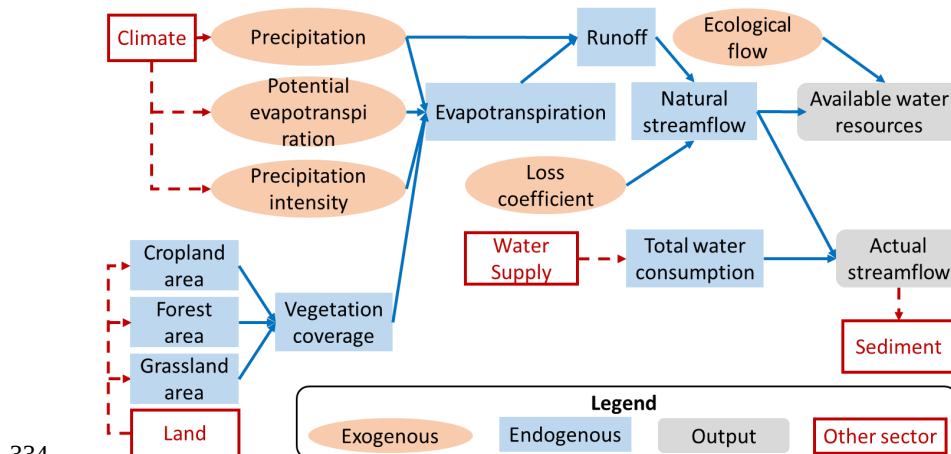




Fig.8 Structure of the *Water Supply* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Water Supply* sector simulates runoff, discharge, and their changes in each sub-basin. Runoff (R) is determined by precipitation (Pre) and evapotranspiration (ET) based on the water balance principle derived from the Budyko equation, which is suitable for non-humid regions of China (Yang et al., 2009) (Equation 7).

$$R = Pre - ET \quad (7)$$

The ET is calculated by various exogenous climate variables from the *Climate* sector, and vegetation coverage from the *Land* sector (Equations 8-9, S46-S48),

$$ET = \frac{PET \times Pre}{(Pre^n + EP^n)^{\frac{1}{n}}} \quad (8)$$

$$n = Para_{E1} \times \left(\frac{Ks}{PI}\right)^{Para_{E2}} \times FVC^{Para_{E3}} \times e^{Para_{E4} \tan \beta} \quad (9)$$

where PET is potential evapotranspiration at sub-basin, from the *Climate* sector; n is a parameter reflecting the basin landscape characteristics, related to saturated hydraulic conductivity (Ks), precipitation intensity (PI), average slope (β), and fraction of vegetation coverage (FVC); $Para_{E1}$, $Para_{E2}$, $Para_{E3}$, and $Para_{E4}$ are parameters fitted from historical data.

Runoff and the loss coefficient (defined by the ratio of natural streamflow to runoff) determine the natural streamflow due to the water loss during the confluence process (Equation S49). Natural streamflow and ecological flow constraints define the upper limit of available water resources. By integrating human water consumption from the *Water Demand* sector across sub-basins, the model calculates the actual streamflow (Equations S50-S52). Actual streamflow is transferred through hydrological connectivity from upstream to midstream and then to downstream, ultimately reaching the sea, which governs sediment transport processes within the *Sediment* sector.

2.2.7 Sediment

The *Sediment* sector estimates sediment load (Sed) for each sub-basin using an empirical model in the literature (Yin et al., 2023b), which links actual streamflow from



the *Water Supply* sector to sediment transport,

$$Sed = Para_{SS} \times AS + Const_{SS} \tag{10}$$

where $Para_{SS}$ and $Const_{SS}$ are derived from linear fitting of historical hydrological station data on actual streamflow and sediment load.

2.2.8 Land

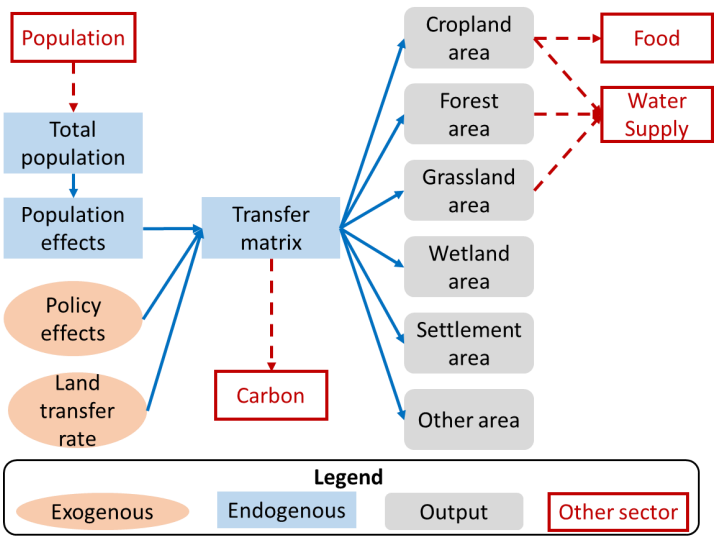


Fig.9 Structure of the *Land* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Land* sector simulates the area changes of six land use types (cropland, forest, grassland, wetland, settlement, and others) based on the land transfer matrix obtained from historical remote sensing data (Xu et al., 2018). The land transfer matrix calculates the inflow and outflow of each land category and can be configured to represent the influence of future land use drivers. This sector outputs vegetation area (including forest, grassland, and cropland), which influences the *Water Supply* sector. Cropland area changes impact the *Food* sector, while land use conversion also plays a role in the *Carbon* sector.

Based on the initial land use area ($IniArea_i$), the transfer matrix determines the area ($Area_i$) allocated to each land use type (Equations 11, S55-S61),



$$Area_i = IniArea_i + \int (\sum_{i=1}^6 FTM_{i,j} - \sum_{j=1}^6 FTM_{i,j}) dt \quad (11)$$

where $FTM_{i,j}$ is the final land use transfer matrix, indicating the area of land use i transferred to land use j , i and j represent six land use type.

2.2.9 Carbon

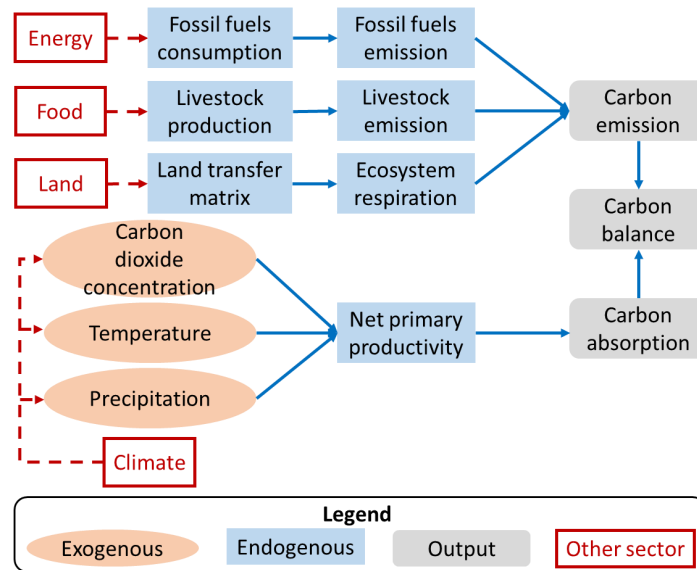


Fig.10 Structure of the *Carbon* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Carbon* sector simulates the basin's carbon balance processes, including carbon emission and absorption, adapted from the carbon cycle of ANEMI (Davies and Simonovic, 2011). Carbon emissions (CE) encompass fossil fuel emissions, and livestock emission, and ecosystem respiration, which are influenced by outputs from the *Energy*, *Food*, and *Land* sectors.

Fossil fuel and livestock emissions are calculated based on fossil fuel consumption, livestock production, and their respective emission coefficients. Ecosystem emissions include carbon released from burning as well as from the decomposition of biomass, litter, humus, and charcoal, which are determined by carbon pools' lifespans, decomposition factor, and respiration coefficients (Equations S64-S88).



397 Carbon absorption (CA) is determined by net primary productivity (NPP) and land
398 use area. NPP is influenced by climate factors, including CO_2 concentration,
399 temperature, and precipitation (Equations 12, S63),

400
$$CA = \sum_{i=1}^6 NPP_i \times Area_i \quad (12)$$

401 where i represent land use type, land use area ($Area_i$) from the *Land* sector.

402 **2.2.10 Climate**

403 The *Climate* sector supplies both historical and projected climate data essential for
404 the *Water Supply*, *Food*, and *Carbon* sectors. The key climate variables comprise
405 temperature and precipitation (at the sub-basin and provincial levels), as well as
406 potential evapotranspiration, precipitation intensity, and CO_2 concentration (at the sub-
407 basin level). These variables are treated as exogenous inputs, without accounting for
408 potential feedbacks from human activities or natural system responses on regional
409 climate patterns.

410 **2.3 Data sources**

411 In the model, there are more than 100 exogenous variables, some of which are
412 initial variables that drive the simulation. These exogenous variables are derived either
413 from historical statistical data or from fitted results based on historical data (Table S2).
414 All data sources required for the model simulation are listed in Table 1.

415 Table 1. Summary of data sources for the YRB model. Updated from Sang et al.
416 (2025).

Sector	Variables*	Data sources	Spatial and temporal scale	Time range
Population	age-specific population, births, deaths, migrants	- China Statistical Yearbook (NBSC, 2020c), - China Population Census Yearbook (NBSC, 2020b)	provincial, annual	1981–2023
Economy	GDP, investment, employment	- China Agricultural Yearbook (MAARA, 2020), - China Statistical	provincial, annual	1981–2023



Yearbook (NBSC, 2020c)				
Energy	electricity, coal, oil, gas consumption	China Energy Statistical Yearbook (NBSC, 2020a)	provincial, annual	1981–2023
Food	crop production, fertilizer, irrigated area	China Agricultural Yearbook (MAARA, 2020)	provincial, annual	1981–2023
Water demand	Industrial, domestic, irrigation water withdrawal	- National Long-term Water Use Dataset of China (NLWUD) (Zhou et al., 2020) - China Statistical Yearbook (NBSC, 2020c)	prefectural, annual	1965–2013 1981–2023
Water supply	runoff, streamflow, water stress	- China Natural Runoff Dataset (CNRD) (Gou et al., 2021), - Gauge-based Natural Streamflow Dataset (Miao et al., 2022) - Yellow River Water Resources Bulletin (YRCCMWR, 2020)	0.25° grid, Monthly 0.25° grid, annual - provincial, annual	1961–2018 1961–2018 1998–2022
Sediment	sediment load	Yellow River Water Resources Bulletin (YRCCMWR, 2020)	provincial, annual	1998–2022
Land	cropland, forest, grassland, wetland, settlement	- China land use remote sensing monitoring dataset (CNLUCC) (Xu et al., 2018) - Long-term global land surface satellite fractional vegetation cover product (Jia et al., 2015, 2019)	1 km grid, annual	1980, 1990, 1995, 2000, 2005, 2010, 2015, 2020
Carbon	fossil fuel carbon emissions, net primary productivity	- Carbon Emission Accounts and Datasets (Shan et al., 2018), - Global Vegetation	- Provincial, annual 5 km grid, 8 day	1997–2021 1981–2018



	Productivity Products (GVPP) (Cui et al., 2016; Wang et al., 2020a; Yu et al., 2018)		
Climate	temperature, precipitation, CO ₂	• Gridded daily observation dataset over China region (Wu and Gao, 2013), • European Environment Agency (EEA, 2019), • China Meteorological Data Service Center (https://data.cma.cn/)	• 0.25° grid, Annual • Global, Annual • Meteorological stations, Monthly
	concentration, potential evapotranspiration and precipitation intensity		• 1961–2020 • 1800–2020 • 1956–2018

* Listed are some typical variables of each sector

3 Model validation and application

3.1 Historical model validation

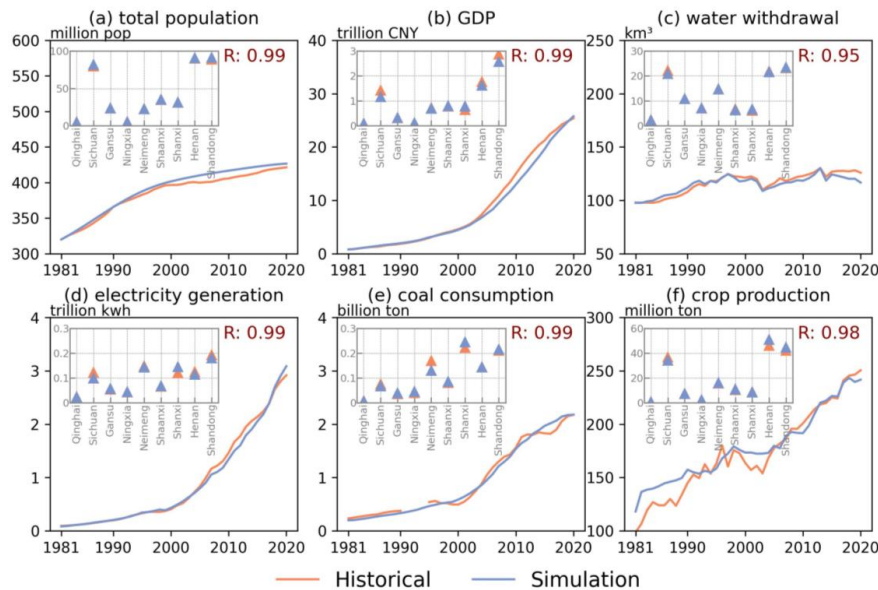


Fig.11 Validation of human system processes during historical period of 1981-2020.

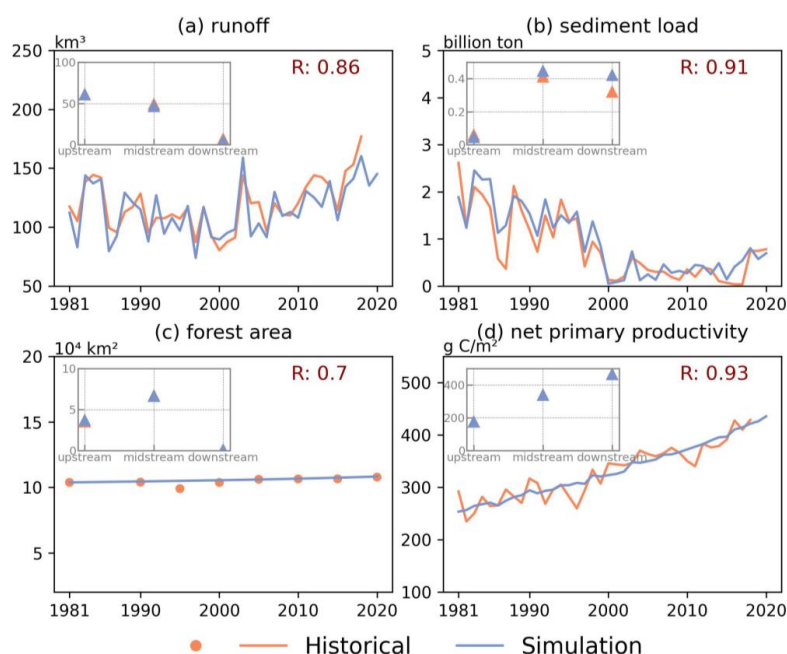
(a) total population in *Population* sector; (b) GDP in *Economy* sector; (c) water withdrawal in *Water Demand* sector; (d) electricity generation and (e) coal consumption in *Energy* sector; (f) crop production in *Food* sector. Pink triangles in the upper left sub-image represents the average of historical and simulation value in



1981–2020 of each province in the YRB.

We compared model performance during the historical period (1981–2020) against historical data. These data are sourced from multiple channels, including statistical yearbooks, hydrological stations, remote sensing observations, and outputs from other models (Table S3).

For human sectors, the simulation accuracy of selected key variables from each sector is consistently high ($R > 0.95$), including total population, gross domestic product (GDP), water withdrawal, electricity generation, coal consumption, and crop production across the nine provinces. However, the accuracy varies among provinces, particularly for GDP, electricity generation, coal consumption, and crop production. In general, Shanxi exhibits lower simulation accuracy compared to the others. Among all sectors, the greatest uncertainty arises in simulating coal consumption, because the absence of a clear historical trend results in poor fitting performance. Nevertheless, the model performs reasonably well in the human sectors, effectively capturing the historical dynamics of human systems in the YRB.





442 Fig.12 Validation of the natural system processes during historical period of 1981-
 443 2020. (a) runoff in *Water Supply* sector; (b) sediment load in *Sediment* sector; (c)
 444 forest area in *Land* sector, and the historical dataset is discontinuous; (d) net primary
 445 productivity in *Carbon* sector. Pink triangles in the upper left sub-image represents
 446 the average of historical and simulation value in 1981-2020 of each sub-basin in the
 447 YRB.

448 For natural sectors, the simulation accuracy for runoff, sediment load, forest area,
 449 and NPP in the YRB is relatively high ($R > 0.7$). At the sub-basin level, the overall
 450 performance is satisfactory; however, the model shows lower accuracy in simulating
 451 sediment transport in the mid- and downstream. Compared to the human sectors, the
 452 natural sectors generally exhibit lower correlations with historical data, largely due to
 453 the simplifications required when modeling physically complex natural processes.
 454 Nevertheless, the model is still capable of capturing the historical dynamics of natural
 455 processes in the YRB.

456 **3.2 Model application for future projection**

457 **3.2.1 Future baseline scenario**

458 The future baseline scenario represents a trajectory in which existing plans and
 459 policies continue to operate without substantial changes in external environments. The
 460 development of the baseline scenario primarily relies on variables from the *Population*,
 461 *Economy*, *Water Demand*, and *Land* sector based on available government planning
 462 documents, historical trends, and other projections (Table 2). Future climate data (2015-
 463 2100) are from the ensemble mean of 11 CMIP6 models (ACCESS-CM2, CESM2,
 464 CMCC-ESM2, GFDL-ESM4, INM-CM4-8, INM-CM5-0, IPSL-CM6A-LR, MIROC6,
 465 MRI-ESM2-0, UKESM1-0-LL, and HadGEM-GC31-LL) (available at
 466 <https://cds.climate.copernicus.eu/datasets/projections-cmip6?tab=download>) for the
 467 SSP 2-4.5 because this scenario aligns more closely with the climate trends in the YRB.
 468 To ensure temporal consistency, CMIP6 historical climate data (1981–2014) were used
 469 instead of observed historical records. We run the model under the future baseline



470 scenario to project the evolution of CHANS in the basin from 2021 to 2100 (see
 471 Supporting Information S5 for details of scenario design).

472 Table 2 Settings of key variables in the future baseline scenario

Sector	Variables	Future baseline
Population	• Total fertility	• 2024 United Nations Population Projections
	• Gender ratio	• 2024 United Nations Population Projections
	• Average years of schooling	• The national maximum in 2023 (Germany) as the upper limit
	• Urban and rural population ratio	• Analysis and Forecast of Urbanization Trends in China's Modernization by 2020
	• Labor force	• Delay retirement age
Energy	• Electricity generation sources ratio	• Projected provincial-level energy structure meeting China's carbon peaking and carbon neutrality goals (Li et al., 2024)
	• Irrigation intensity	• Irrigation water use intensity declines linearly (1%/yr) to its minimum value.
Water Demand	• Industry, service, livestock water use intensity	• Industry, service and livestock water use intensity declines linearly (1%/yr) to its minimum value.
Land	• Land use area	• Minimum cropland area above the red line for cropland protection.
		• Other lands are allowed to be fully converted into other categories.
Climate	• Temperature, Precipitation, Potential evapotranspiration, precipitation intensity,	• SSP 2-4.5 — CMIP 6



CO₂ concentration,

3.2.2 Projection of CHANS dynamics in future baseline scenario

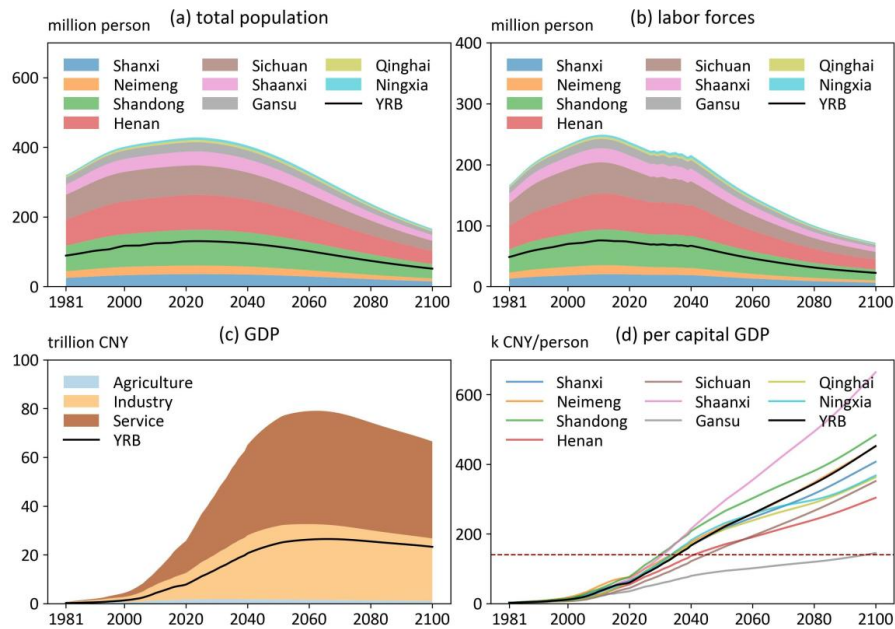
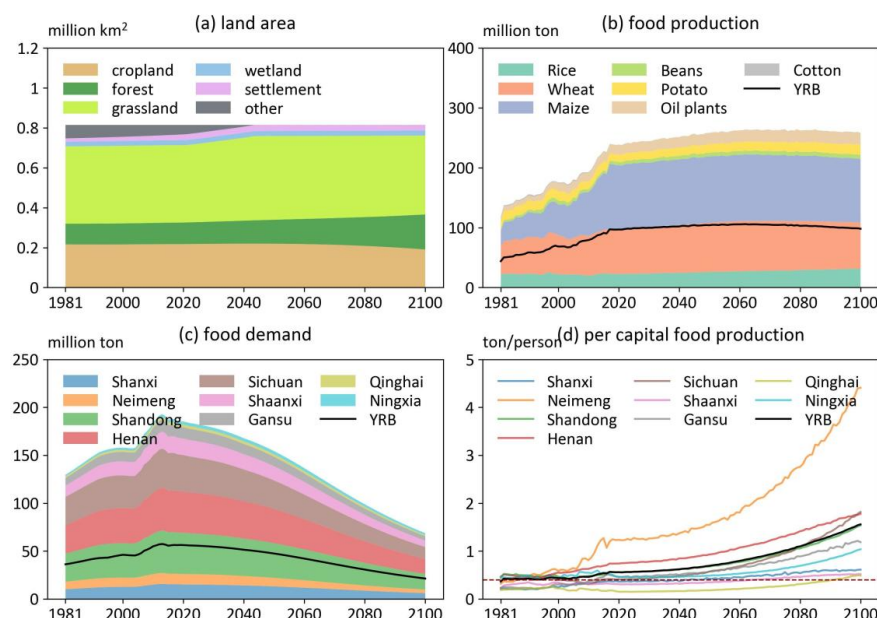


Fig.13 Changes in key variables of the *Population* and *Economy* sector in the future baseline scenario: (a) the total population, (b) the total labor forces, (c) the gross output of agriculture, industry, and service, and the total gross domestic output, (d) the per capital GDP in nine provinces and YRB.

The model simulates human and natural system dynamics under the future baseline scenario and produces outputs for nine provinces and their corresponding areas within the YRB. The simulation results are reported either at the provincial level (nine provinces) for human system sectors or at the basin level (basin boundary) for natural system sectors. The total population in the nine provinces and the YRB is projected to peak in 2023 and 2024, at 428 million and 131 million, respectively, driven by declining fertility rates. The labor force peaks earlier, reaching 249 million in the nine provinces in 2011 and 76 million in the YRB in 2010. After 2025, the labor force is projected to increase again due to the delayed retirement policy. GDP in the nine provinces is



488 expected to increase until a peak in 2062, at 79 trillion CNY (three times the 2020 level,
 489 in 2020 constant prices), under the influence of a labor force decline, with the YRB
 490 reaching its peak four years later. In contrast to total GDP, per capita GDP in all regions
 491 is projected to continue rising throughout the future period. Although all provinces
 492 demonstrate improvements in economic status and living standards, considerable
 493 regional disparities persist. Among the nine provinces, Shaanxi demonstrates the largest
 494 growth in average per capita GDP from 2021 to 2100 relative to the historical period
 495 (1981-2020), with a more than sixteenfold increase, while the YRB as a whole shows
 496 an over twelvefold increase.



497
 498 Fig.14 Land and food in the future baseline scenario. (a) land area in the YRB; (b-d)
 499 food production of seven crop types, food demand and per capital food production in
 500 nine provinces and the YRB, red dotted line represents the 0.4 tons international food
 501 security threshold.

502 Under the future baseline scenario, land use patterns remain relatively stable based
 503 on historical trends, as no new land policies are introduced and cropland area needs to
 504 stay above the red line. The forest area is projected to increase gradually, reaching 62%



above the 2021 level by 2100, while the cropland area is expected continue declining, falling 12% below the 2021 level. Total crop production in nine provinces and YRB is projected to peak in 2079 and 2062, respectively, driven by the combined effects of declining cropland area and increasing crop yields. Food demand in both the nine provinces and the YRB as a whole peaked in 2013 (192 million tons and 57 million tons, respectively), largely driven by population dynamics. In the future, per capita food production in the YRB is projected to consistently exceed the international food security threshold of 0.4 tons per person. However, in certain years, provinces such as Qinghai and Shaanxi could fall below this standard.

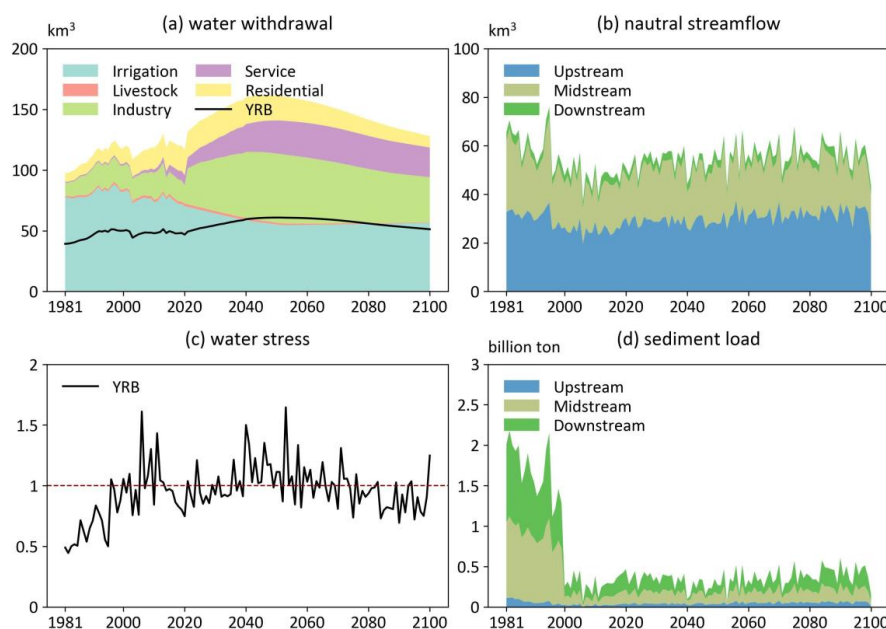


Fig.15 Water demand and supply in the future baseline scenario. (a) water withdrawal across five water-use sectors in nine provinces and the YRB; (b) natural streamflow in up-, mid- and downstream; (c) water stress in the YRB, red dotted line represents the water security threshold; (d) sediment load in the up-, mid- and downstream.

Total water withdrawal in the nine provinces and the YRB is projected to peak in 2047 and 2051, reaching 162 km³ and 61 km³, respectively. Irrigation water withdrawal is expected to decline, driven by reductions in cropland area and irrigation intensity, the



latter resulting from improvements in irrigation efficiency assumed in the scenario. Peaks in residential, industrial, service, and livestock water withdrawals are primarily associated with projected peaks in population and GDP. Natural streamflow is projected to follow a fluctuating upward trend, with the future annual mean nearly identical to the historical average (a 0.7% increase), and the largest increase occurring in the upstream region (8%). Considering the ecological flow requirement of 18.7 km³ (YRCCMWR, 2015), water stress (calculated as water consumption divided by (natural discharge - ecological flow)) is expected to decline in the latter half of the century, largely due to the peak and subsequent reduction in water withdrawal. However, overall water stress is projected to exceed historical levels in 2041 and will fall below 1 in 54% of the years, reflecting mitigation of persistent human–water tensions and future trade-offs between ecological and human water use. As a result of reduced actual streamflow, sediment transport in the Yellow River is projected to decline sharply, with a 63% decrease in sediment load relative to the historical period.

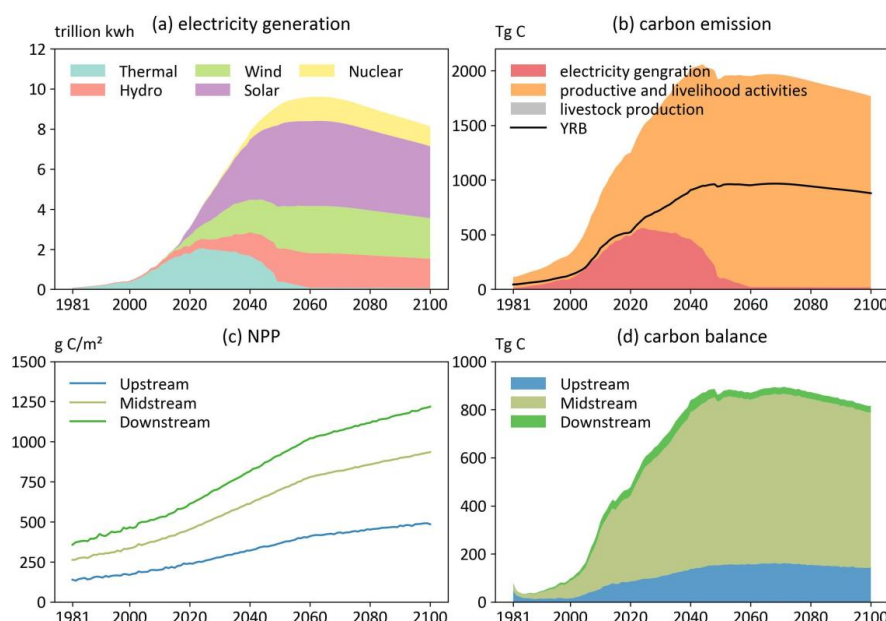


Fig.16 Electricity and carbon in the future baseline scenario. (a) the electricity generation from five sources in nine provinces; (b) carbon emission from three



539 sources in nine provinces and the YRB; (c-d) the net primary productivity and carbon
540 balance in the up-, mid-, and downstream.

541 Since the energy system is transitioning to carbon neutrality goals under the future
542 baseline scenario, thermal power generation is projected to decline sharply, while clean
543 energy (hydropower, wind, solar, and nuclear) will become the dominant source for
544 electricity generation. The large-scale adoption of clean energy will lead to a peak in
545 carbon emissions from electricity generation in 2024. In contrast, emissions from
546 productive and livelihood activities are projected to rise substantially over time, with
547 total human emissions in the YRB reaching a peak in 2044. NPP is anticipated to
548 increase markedly in response to CO₂ concentration increase. Although the carbon
549 balance of the YRB has remained positive (net carbon source), it is expected to
550 gradually decline after 2071, indicating that the basin's ecosystems alone cannot fully
551 offset human-induced carbon emissions.

552 The above projections of future baseline scenarios highlight persistent challenges
553 to achieving sustainable development in the YRB, underscoring the need for integrated
554 policy responses to address these challenges. Priority should be given to enhancing
555 water-use efficiency and establishing adaptive water resource allocation mechanisms
556 that balance ecological and human water needs. Sediment management strategies,
557 which combine ecological restoration with hydraulic regulation, are necessary to
558 maintain river stability and delta health. To counter the decline in labor force, policies
559 should promote industrial upgrading and technological innovation. Strengthening the
560 basin's carbon mitigation capacity requires accelerating the clean energy transition and
561 expanding ecological restoration to enhance carbon sequestration. Finally, a multi-
562 sectoral, CHANS-based governance framework should be established to coordinate
563 water, land, energy, and carbon management, enabling evidence-based scenario
564 analysis and adaptive policy design for long-term sustainability.

565 **4 Discussion and conclusions**

566 The development of the CHANS-SD-YRB model involves a series of key



567 considerations in translating a conceptual framework into a functioning model. The
 568 framework focuses on human–water interactions in the YRB, integrating bidirectional
 569 feedback between human and natural processes. SD was chosen to implement the
 570 framework, as it effectively captures feedback, non-linear relationships, and cross-
 571 sectoral linkages within complex human–natural systems, while remaining practical
 572 and straightforward to use. For quantitative representation of human–natural processes,
 573 we prefer theoretical methods that incorporate key interactions within the system
 574 suitable for implementation using SD. For example, the Budyko framework connects
 575 the *Land*, *Climate*, *Water Demand*, and *Water Supply* sectors. Relative to distributed
 576 hydrological models, it offers lower computational complexity and reduced data
 577 requirements. The Cobb–Douglas production function links the *Population* and
 578 *Economy* sectors. Compared with Computable General Equilibrium models (Fujimori
 579 et al., 2014), it is simpler to construct and requires fewer parameters and data inputs.
 580 When no suitable theoretical framework/method is available to describe a process, we
 581 rely on empirical relationships derived from historical data. For example, crop yields
 582 are estimated using fitted relationships between climate variables, fertilizer application,
 583 and historical yields. For processes that cannot be represented through empirical
 584 functions, we apply literature-based estimates to obtain approximate quantitative
 585 relationships, such as fossil fuel emission factors. For processes influenced by external
 586 drivers that cannot be endogenously expressed within the model, we quantify them
 587 using exogenous parameters derived from historical statistics, such as water intensity.
 588 Given the heterogeneous human–natural interactions in the YRB, we represent all
 589 processes at the provincial and sub-basin scales using scale-specific parameters, except
 590 where parameters or data at these levels are unavailable.

591 The CHANS-SD-YRB model explicitly couples multiple human and natural
 592 sectors, enabling a more integrated representation of feedbacks across population,
 593 economy, energy, food, water, sediment, land, carbon, and climate. Unlike models that
 594 focus on specific sectors or isolated subsystems—such as eco-hydrological models and



595 sediment transport models, which may well capture individual processes but cannot
596 represent the complex human–nature interactions driving system dynamics. Our
597 model’s comprehensive coupling broadens its scope of application to address complex
598 regional CHANS challenges and provides practical guidance for sustainable
599 development. The CHANS-SD-YRB model serves as a comprehensive platform for
600 conducting system dynamics prediction, scenario analysis, policy evaluation, and
601 optimization, to alleviate human–water conflicts, ensure food security, and achieve
602 long-term sustainability. For example, analysis of the impacts of the 1987 Yellow River
603 water allocation policy (Song et al., 2024) offer valuable insights for adjusting
604 interprovincial water distribution to promote sustainable water governance; assessment
605 of ecological restoration policies (Li et al., 2015) can guide future ecological
606 engineering; spatiotemporal dynamics of future water gaps can serve as a valuable
607 reference for planning inter-basin water transfers. The model’s flexibility also allows
608 for the incorporation of additional feedback, for example, linking water scarcity to
609 industrial output or climate warming to agricultural productivity. These could include
610 the effects of global warming on human health (Yin et al., 2023a) and economic
611 activities (Nordhaus, 2017), water constraints on production, dietary shifts influencing
612 carbon emissions and land use (Ren et al., 2023), and the trade-offs between carbon
613 mitigation and food security (Xu et al., 2022).

614 Nevertheless, the model remains subject to further refinement. The current
615 simplifications of natural processes could be replaced with more sophisticated models
616 to enhance simulation accuracy. For instance, the YRB hydrological processes involve
617 highly complex human interventions, including reservoir, conservation, and
618 revegetation projects (Wang et al., 2025). The refined runoff simulations by distributed
619 hydrological models improve water supply assessments through their better
620 characterization of spatial heterogeneity in soil, precipitation, and snowmelt (Cong et
621 al., 2009). The model’s spatial and temporal resolutions are relatively coarse due to the
622 inherent mismatch in spatiotemporal scales between human and natural processes. The



scale at which we make the model represents a practical compromise, constrained by data availability and aligned with the study's objectives. Simulations at finer scales (e.g., monthly, daily, or gridded) would enhance the representation of natural processes (e.g., hydrological processes, land use changes, and food production) and the associated spatiotemporal heterogeneity (e.g., daily simulations can assess the impacts of extreme weather). However, it may also exacerbate the scale mismatch with socioeconomic processes, making it challenging to analyze cross-sectoral dynamics. Technically, the inherent limitations of the SD software restrict the model's ability to represent temporal fluctuations and spatial variations. To overcome this, transitioning from the VENSIM platform to a code-based implementation will be necessary, which would also facilitate coupling with other models. Future research should prioritize these improvements to strengthen both the performance and applicability of the model.

Overall, the CHANS-SD-YRB model fills the gap in CHANS modeling for the YRB. It integrates the dynamics of ten interconnected sectors: *Population, Economy, Energy, Food, Water Demand, Water Supply, Sand, Land, Carbon, and Climate*, achieving reciprocal feedback between human and natural systems. This model can serve as a robust tool to inform policy decisions that influence the evolution of coupled human–natural systems and to explore pathways for optimizing these systems toward sustainability. Furthermore, the modeling process provides valuable experience for regional CHANS modeling and contributes to advancing the broader development of CHANS models at the regional scale.

Code and data availability

The CHANS-SD-YRB V1.0 model (Vensim DSS format), along with the input data and simulation outputs used in this study, are openly accessible at <https://doi.org/10.5281/zenodo.17568963> (Sang, 2025).

Author contributions

Conceptualization: YL, SS¹, BF. Data curation: SS¹. Formal analysis: SS¹. Funding acquisition: BF, YL. Investigation: YL, SS¹. Methodology: SS¹, YL, BF.



651 Project administration: YL. Resources: YL. Software: SS¹, SZ, LY. Supervision: YL,
 652 BF. Validation: SS¹, SZ, LY. Visualization: SS¹. Writing – original draft: SS¹. Writing –
 653 review and editing: SS¹, YL, SW, YXL, XTW, SS², WZ, XHW. (Note: SS¹ refers to
 654 Shan Sang, and SS² refers to Shuang Song, to distinguish between authors with identical
 655 initials.)

656 **Competing interests**

657 The authors declare that they have no conflict of interest.

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