CHANS-SD-YRB V1.0: A System Dynamics model of the coupled human-natural systems for the Yellow River Basin

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Abstract: Modeling the coupled human–natural systems (CHANS) is vital for understanding human–natural interactions and achieving regional sustainability, offering a powerful tool to alleviating human–water conflicts, ensuring food security, thereby supporting the region's pathway toward sustainable development. However, the scarcity of regional-scale CHANS models constrains progress in practical applications for regional sustainability. The Yellow River basin (YRB) is an ideal region

for modeling regional CHANS due to its closely coupled human and natural systems, which are stressed by water and ecosystem fragility. Here, we developed the CHANS-SD-YRB model using the System Dynamics approach, integrating 10 sectors essential for modeling human-water interactions of the basin, including five human sectors (Population, Economy, Energy, Food, and Water Demand) and five natural sectors (Water Supply, Sediment, Land, Carbon, and Climate). The model can simulate evolution and feedbacks of the YRB CHANS annually at provincial and sub-basin scales, while conserving hydrological connectivity between sub-basins. The model can accurately reproduce historical CHANS dynamics, achieving strong quantitative agreement with historical data ($R > 0.95$ for human sectors and $R > 0.7$ for natural sectors), which supports its applicability for scenario analyses and future projections. We applied the model to explore human–natural system dynamics under a future baseline scenario, assuming the continuation of existing policies and climate projection under middle of the road scenario (SSP–RCP 2-4.5). The future projections (2021-2100) indicate that achieving sustainable development in the YRB will remain challenging, though economic growth and food security are expected to improve. Emerging issues, such as ecological–human water trade-offs, labor shortages, reduced sediment load, and limited carbon absorption capacity, may hinder regional long-term sustainability.

Keywords: coupled human-natural systems, regional modeling, system dynamics, the Yellow River Basin

1 Introduction

Coupled human and natural systems (CHANS) emphasize the reciprocal feedback and co-evolution between human and natural systems, offering an integrated framework for diagnosing complex problems and guiding sustainable development (Fu and Li, 2016). By capturing dynamic interactions of interconnected components, CHANS theories and models enable more effective policies and interventions that align ecological integrity with socioeconomic progress (Motesharrei et al., 2016; Verburg et al., 2016). Numerous integrated modelling approaches have been developed to simulate

human–natural interactions at the global scale. These include system dynamics-based integrated assessment models (IAMs), such as ANEMI (Breach and Simonovic, 2021), FeliX (Rydzak et al., 2013; Ye et al., 2024), and FRIDA (Rajah et al., 2025), process-based and optimization-based IAMs (Vaidyanathan, 2021), and Earth system models with synchronously coupled human components, such as E3SM-GCAM (Di Vittorio et al., 2025) and integrated Earth system models (iESMs) (Jain et al., 2022). These models effectively characterize human–natural interactions at the global scale, and have been applied to assess the impacts of climate change on human society (e.g., agriculture (Monier et al., 2018), economic damage (Wang et al., 2020b), fatalities increase and welfare loss (Dottori et al., 2018)), and humans' feedback on the Earth system, such as those from climate mitigation on water and food security (Cheng et al., 2022; Fujimori et al., 2022).

Currently, most CHANS models are at the global scale (Calvin and Bond-Lamberty, 2018) with much fewer regional models. While global modeling research has deepened our knowledge of dynamic feedbacks among Earth's spheres and system evolution under climate change, sustainability challenges often manifest at regional scales, where social and ecological dynamics are more intricately intertwined (Liu et al., 2007). Compared to global CHANS, regional CHANS are open systems that continuously exchange energy and materials with other regions and the global system, e.g., water resources and electricity transfers (Dobbs et al., 2023; Zhang et al., 2022a) and trade (Ristaino et al., 2021), resulting in pericoupling and telecoupling of different systems (Liu, 2017). Besides, regional CHANS are shaped by more immediate and complex human influences and stressors, such as urban expansion (van Vliet, 2019), ecological protection (Xu et al., 2017; Yang et al., 2022), and water resource regulation policies (diversion and allocation) (Song et al., 2024), which alter regional CHANS dynamics. Due to their diverse ecological and socioeconomic resilience, regional CHANS exhibit heterogeneous responses to external weather events and climate change, as evidenced by differing responses in crop yield (Hasegawa et al., 2021) and

economic production to extreme heat and warming (Waidelich et al., 2024a). Furthermore, the coarse spatiotemporal resolution of global models limits their capacity to support effective decision-making for regional development (X. Li et al., 2018). As such, advancing regional CHANS modeling is essential for informing adaptive strategies in the face of growing regional environmental and societal pressures.

To address this limitation, many regional CHANS models have been developed at various regional scales (e.g., national, basin, and urban) using System Dynamics (SD) and agent-based modeling (ABMs) techniques. Notable examples include the ANIME-Yangtze model (Jiang et al., 2022), the T21-China (Qu et al., 2020), and the iSDG-Australia model (Allen et al., 2019), all based on SD, the Jordan Water Model (Yoon et al., 2021) with its core on ABM, as well as integrated models in the San Juan River Basin (Hyun et al., 2019) and the Heihe River Basin (Li et al., 2021). These models are designed to capture finer-scale dynamics and region-specific human–natural interactions, since they embed localized characteristics (e.g., fishing ban, reservoir operation strategies, demographic policies, transboundary flows) and account for heterogeneity overlooked by global models. As a result, regional CHANS models offer stronger policy relevance, providing actionable insights for national, basin, and urban decision-making, and advancing CHANS research across multiple scales.

The Yellow River Basin (YRB) in China is one of the regions where conflicts between human and natural systems are most acute and complex, particularly in terms of human–water relations, due to the severe imbalance between socioeconomic development and natural hydrological, ecosystem processes. The YRB faces severe water stress, with the water resource utilization rate exceeding 80% (Feng and Zhu, 2022; Zhang et al., 2022b). Intensive water extraction has triggered a series of ecological and environmental issues, including flow interruptions, water pollution, and declining groundwater levels, all of which in turn constrain socioeconomic development. The Yellow River traverses the Loess Plateau (Zhu et al., 2019), where severe soil erosion makes the Yellow River one of the most sediment-laden rivers

globally (Fu et al., 2011; Yin et al., 2021). The pronounced spatial and temporal variability in streamflow and sediment load leads to significant riverbed aggradation, frequent flooding, and disruption of agricultural production and other livelihood activities (Miao et al., 2016). Due to internal hydrological connectivity, sub-basins are highly interconnected and are susceptible to upstream influences. Upstream water overuse diminishes downstream availability (Wei et al., 2023), a factor that played a major role in flow interruptions during the 1990s (Changming and Shifeng, 2002; Wang et al., 2019). Ecological challenges differ across the subbasin, with the upstream facing ecosystem degradation and limited water retention (Ning et al., 2022), the midstream characterized by soil erosion and large-scale ecological restoration (Fu et al., 2011), and the downstream focusing on wetland conservation (Fu et al., 2023). Policy measures aimed at ecological restoration, such as afforestation and cropland conversion, have increased vegetation cover, reduced sediment loads but also decreased runoff, exacerbating water scarcity (Feng et al., 2016; Wang et al., 2016). These interlinked dynamics underscore the YRB as a complex coupled human–natural system, where addressing environmental challenges requires an integrated, systems-oriented approach.

The existing models for the YRB are typically designed for specific problems with a narrow application focus and only represent a limited set of human and natural components within the CHANS. These include limited nature-to-human impact pathways, e.g., low flows threatening farmers' livelihoods (Liu et al., 2008), the damage of floods and droughts on agriculture (Zhang et al., 2015), as well as human-to-nature impact pathways, e.g., effects of ecological restoration policy on hydropower and water–sediment–carbon dynamics (Wu et al., 2025; Yan et al., 2024), and the impacts of irrigation water-saving and salinity-control practices on crop yield and water productivity (Wu et al., 2023). These models focus on isolated components of CHANS, with limited consideration of fully coupled human–natural interactions, which limits their capacity to represent full human–natural interactions and support regional decision-making.

To address the gap in CHANS modeling for the YRB, following our previously proposed CHANS modeling framework for the basin (Sang et al., 2025b), we implemented the framework to develop the coupled human and natural systems model for the YRB (CHANS-SD-YRB) using the System Dynamics approach. Through dynamic interaction with policies, climate change, human activities, and environmental feedbacks, the CHANS-SD-YRB model provides a platform for predicting system dynamics, conducting scenario analyses, evaluating policies, and optimizing water-food-carbon synergies. This study offers both theoretical and practical insights for advancing regional CHANS modeling and promoting sustainable development in the YRB.

2 Description of the CHANS-SD-YRB

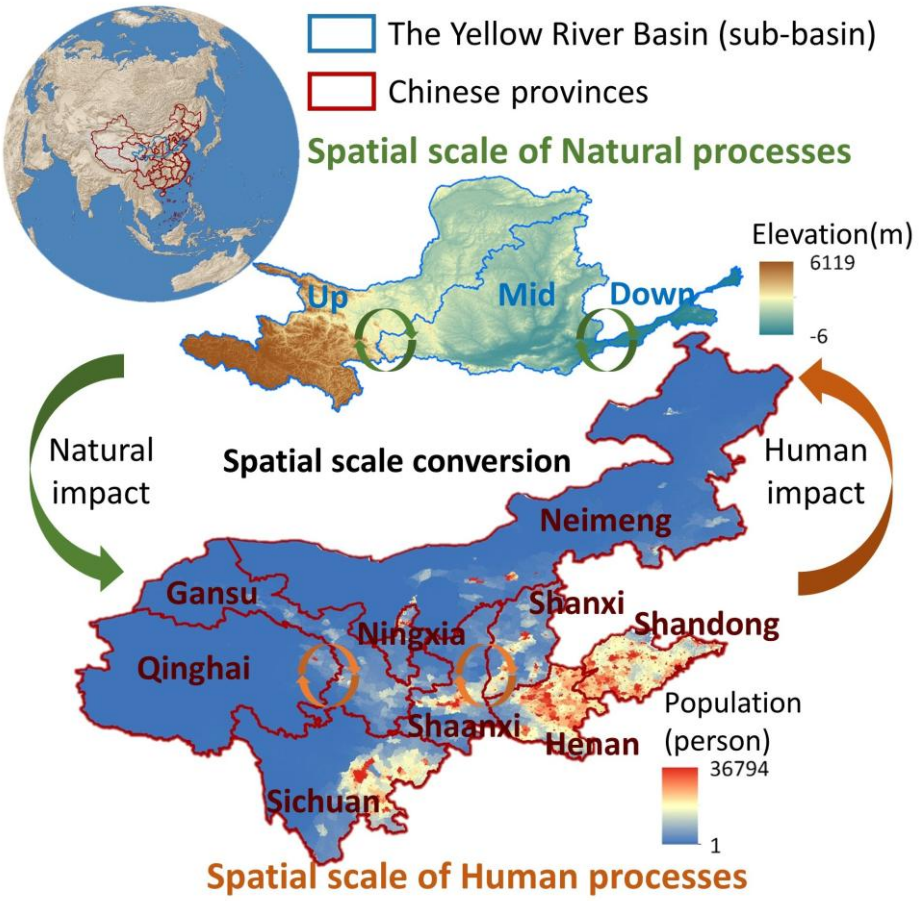

Fig.1 Geolocation of the Yellow River Basin and boundary of the natural and human processes in the CHANS-SD-YRB model. Natural processes are simulated at the sub-

basin scale (base map: elevation), and human processes at the provincial scale across nine provinces (base map: population density in 2020).

We developed the CHANS-SD-YRB based on system dynamics, a method well-suited for capturing complex system behaviors characterized by nonlinearity, multi-level structures, and feedback loops (Forrester, 1968; Richardson, 2011). The model was constructed and implemented using the VENSIM DSS (Ventana Systems, 2023) software platform, operating on an annual time step. The CHANS-SD-YRB simulates both human and natural processes for historical simulations (1981–2020) and future projections (2021–2100). Human processes are simulated at the provincial scale, covering the nine provinces along the Yellow River (Qinghai, Sichuan, Gansu, Ningxia, Neimeng, Shaanxi, Shanxi, Henan, and Shandong), while natural processes are simulated at the sub-basin scale (up-, mid-, and downstream) (Fig.1). The model is designed to capture the various interactions within and between different components of human system and natural system across administrative and hydrological units. The spatial scale conversion between provincial and sub-basin levels relies on weights (e.g., the proportion of sub-basin-level values to provincial total values) derived from historical high-resolution gridded datasets of human-related variables. These weights enable disaggregation of provincial outputs to the sub-basin level (see Supporting Information S4 for details). Given the availability of gridded data for human processes and the strong correlations among relevant variables, gridded population and GDP data were used as proxies to disaggregate demographic variables and economic and human carbon emissions, respectively, from provincial-level to the sub-basin scale (Table S3).

2.1 Model structure

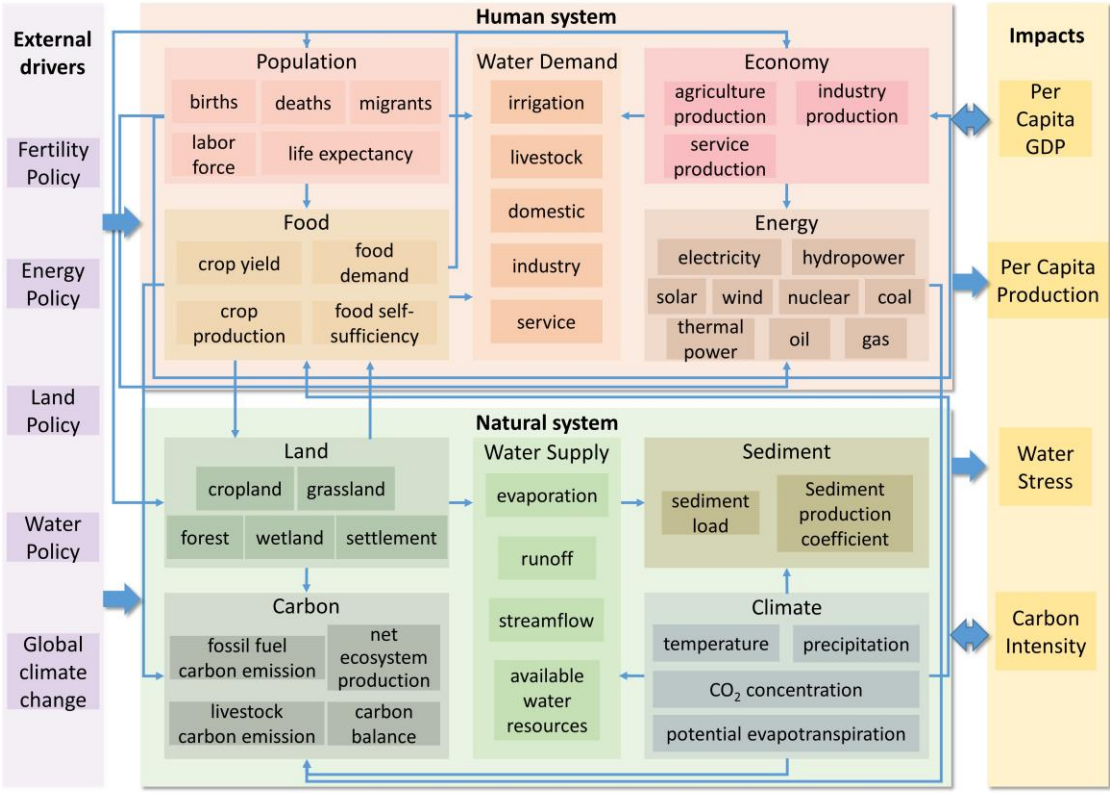

Fig.2 Structure of the CHANS-SD-YRB, which shows sectors of human and natural systems, their key processes and interactions.

Drawing on the modeling framework of CHANS in the YRB (Sang et al., 2025b), we designed the CHANS-SD-YRB structure (Fig. 2), including five sectors related to human society (*Population*, *Economy*, *Energy*, *Food*, and *Water Demand*), and five sectors related to natural ecosystem (*Water Supply*, *Sediment*, *Land*, *Carbon,* and *Climate*).

These sectors are interconnected to represent various human-natural interactions as summarized below. Key interactions among human system modules are summarized below. The *Population* sector affects food demand (*Food*), residential water use (*Water Demand*), household electricity and gas consumption (*Energy*), and settlement land area (*Land*). The sector also interacts dynamically with the *Economy* sector, where economic output influences deaths and migrants, while the labor force, in turn, drives economic production. The *Economy* sector drives energy uses (electricity, coal, oil, and gas use)

from *Energy*, as well as industrial and service water withdrawal from *Water Demand.* The gross agricultural production in the *Economy* is made up of crop and livestock production (*Food*). The *Energy* sector produces fossil fuel emissions in the *Carbon* sector. The *Food* sector is affected by the *Land* and *Climate* sector, and it also determines irrigation water withdrawal (*Water Demand*) and livestock-related emissions (*Carbon*). Additionally, the *Food* sector interacts closely with the *Land* sector, where crop production depends on cropland area, which, in turn, is influenced by food self-sufficiency. The *Water Demand* sector affects streamflow in the *Water Supply* sector through consumptive water use.

The key interactions among natural system modules are listed below. The *Land* sector influences evapotranspiration in the *Water Supply* sector through vegetation coverage, and it influences carbon absorption in *Carbon* sector through land use area. The carbon absorption is also affected by climatic variables including temperature, precipitation, and CO_2 concentration. Similarly, runoff in the *Water Supply* sector is affected by precipitation, precipitation intensity (mm/h, the rate of rainfall within one hour calculated from daily data), and potential evapotranspiration. Streamflow influences sediment dynamics in the *Sediment* sector together with the *Climate* sector.

In addition, the CHANS dynamics in the YRB are modulated by external drivers, including policies and global climate change that affect various modeled processes (e.g., fertility, energy, land use, and water). With comprehensive representation of CHANS processes and their interactions, the CHANS-SD-YRB model is capable of generating integrated indicators to assess the state of the coupled system, such as per capita GDP, per capita food production, water stress, and carbon intensity. These indicators not only serve as evaluation metrics but also feed back to influence the internal dynamics of the human–natural system in the YRB.

2.2 Sector description

The CHANS-SD-YRB model focuses on the essential human–natural processes in the YRB, with a particular emphasis on human–water interactions. To this end, it has a

comprehensive representation of the full range of natural and human processes that influence water use and supply. Formulation of each sector aims to explicitly account for cross-sectoral interactions as fully as possible while remaining sufficiently simple to be implemented within the SD software. Considering data availability and the spatial scales of process, human sectors (*Population*, *Economy*, *Energy*, *Food*, and *Water Demand*) are simulated at the provincial level, while natural sectors (*Water Supply*, *Sediment*, *Land*, *Carbon*, and *Climate*) are simulated at the sub-basin level within the YRB. Next, we describe each sector and its key formulation and provide full details in Supporting Information S2.

2.2.1 Population

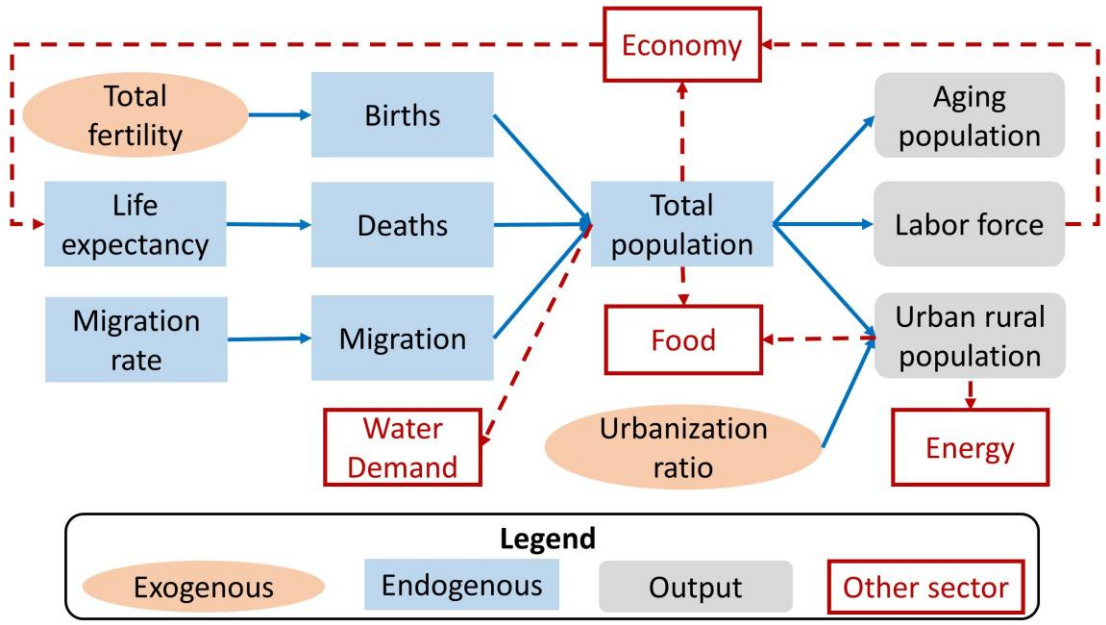

Fig.3 Structure of the *Population* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

Population dynamics are driven by births, deaths, and migration (Fig. 3). Births are calculated based on exogenous total fertility rates derived from historical data (Equation S2) to reflect the strong influence of China's Family Planning Policy. Deaths are determined by the life expectancy, which is modeled as a function of human well-being (Equations S3-S6). Migration is calculated based on historical statistical data and per capita GDP differences between YRB and the national average level (Equations S7-

S8). The total population is characterized by age and gender with exogenous urbanization ratios. The output of *Population* sector drives *Economy*, *Energy*, *Food*, and *Water Demand* sectors through labor force, urban and rural populations.

The key variable, total population, is modeled using the age-structured mathematical method (Kemei et al., 2024), which categorizes individuals by one-year age group and gender (Equation 1),

$$Pop_{g,a} = IniPop_{g,a} + \begin{cases} \int \left(B_{g,a} + NM_{g,a} - D_{g,a}\right)dt, & a = 0 \\ \int \left(Pop_{g,a} + NM_{g,a} - D_{g,a}\right)dt, & 1 \le a \le 99, a = 100 \ and \ over \end{cases} \quad (1)$$

where *Pop* is population, subscript *g* and *a* are gender (g = M is male, g = F is female) and age (a = 0-100 and over); $IniPop_{g,a}$ is the population in the initial year, $B_{g,a}$ is the births, $D_{g,a}$ is the deaths, and $NM_{g,a}$ is the net migrants (i.e., immigrants – emigrations).

2.2.2 Economy

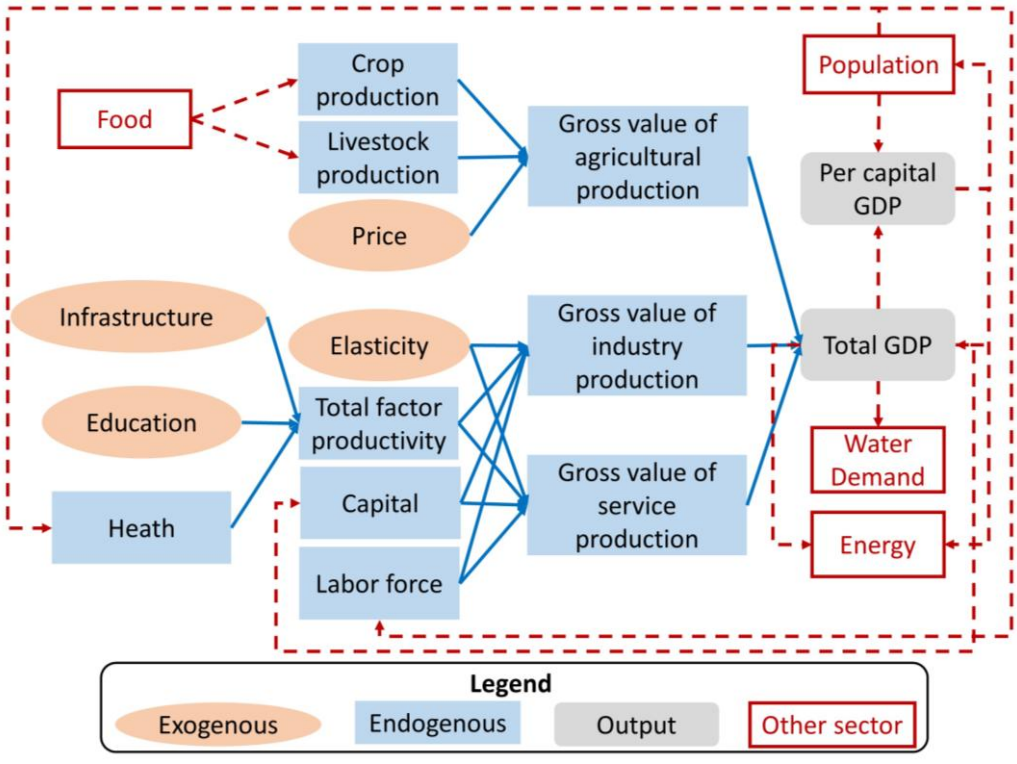

Fig.4 Structure of the *Economy* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Economy* sector simulates production activities in agriculture, industry, and

services, which in turn drive changes in the *Population*, *Energy*, and *Water Demand* sectors (Fig. 4).

The gross product of industry and services is calculated using the Cobb-Douglas production function (Cobb and Douglas, 1928) to account for multiple factors in the economy (Equation 2). Exogenous variables (infrastructure, education, and elasticity), and endogenous variables (health and labor force) from the *Population* sector (Equations S10-S21) all affect the gross domestic product in industry and service (GDP_s),

$$GDP_s = IniGDP_s \times TFP_s \times \left(\frac{L_s}{IniL_s}\right)^{1-\alpha_s} \times \left(\frac{K_s}{IniK_s}\right)^{\alpha_s} \quad (2)$$

where subscript *s* represents industry and service, $IniGDP_s$ is the initial GDP; TFP_s is the total factor productivity calculated from exogenous (infrastructure, education, and elasticity) and endogenous variables (health and labor force) (Equations S10-S13); L_s represents the labor force and $IniL_s$ is its initial value (from *Population* sector); α_s and $1\text{-}\alpha_s$ are capital and labor elasticities from T21-China (Qu et al., 2020) and calibrated using historical data; K_s and $IniK_s$ refer to the capital stock and its initial level.

Gross agricultural production includes crop and livestock production from *Food* sector, calculated by their respective prices (Equations S22-S23).

2.2.3 Energy

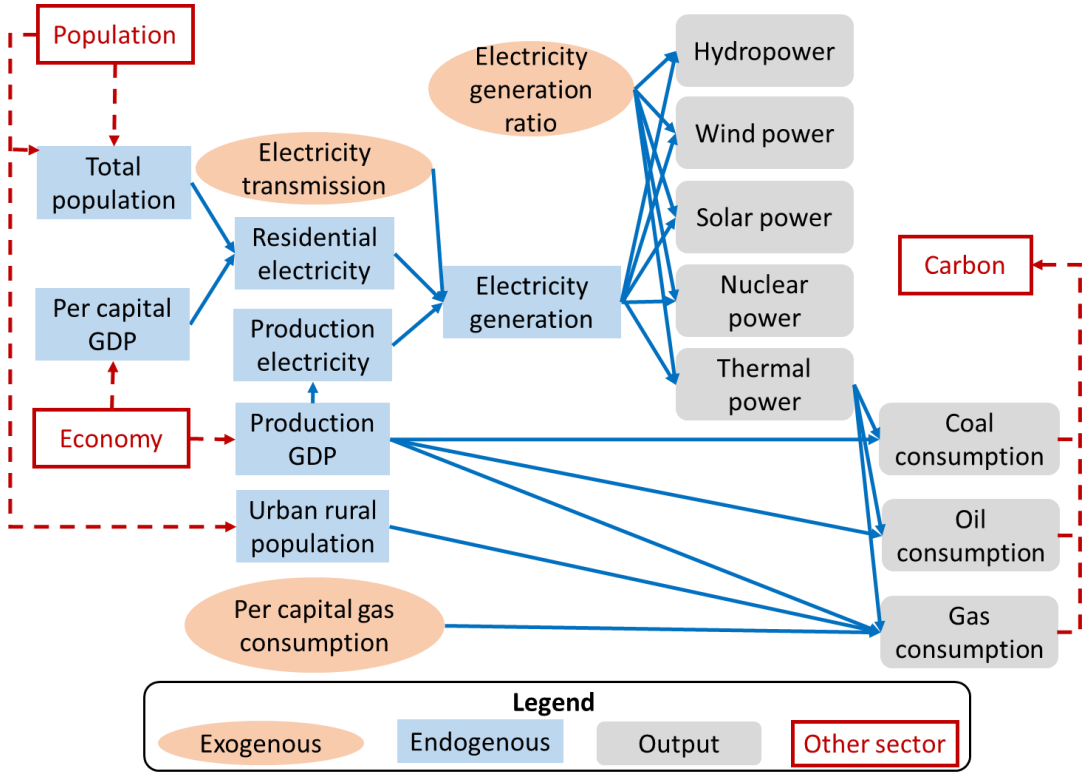

Fig.5 Structure of the *Energy* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Energy* sector simulates production, consumption, and the structure of the energy system, encompassing fossil fuels (coal, oil, and gas) and electricity (Fig. 5). Energy production and consumption are always balanced in this sector. Electricity generation is divided into residential and production uses, the former is estimated from the linear fit of historical per capita GDP and residential demand, and the latter is calculated using GDP and electricity intensity data from the China Energy Statistical Yearbook (NBSC, 2020a) (Equations S24-S25). Cross-provincial electricity transmission is also incorporated according to the same yearbook. The shares of electricity generated from thermal, hydro, wind, solar, and nuclear sources are determined by exogenously specified ratios, as reported in the China Energy Statistical Yearbook (NBSC, 2020a). The fossil fuel consumption by economic production of industry and service sectors is modeled as a linear function of sectoral GDP based on historical data (Equations S26-S28). It combines with fossil fuel consumption for

electricity generation and residential gas consumption to drive carbon emissions in the *Carbon* sector.

2.2.4 Food

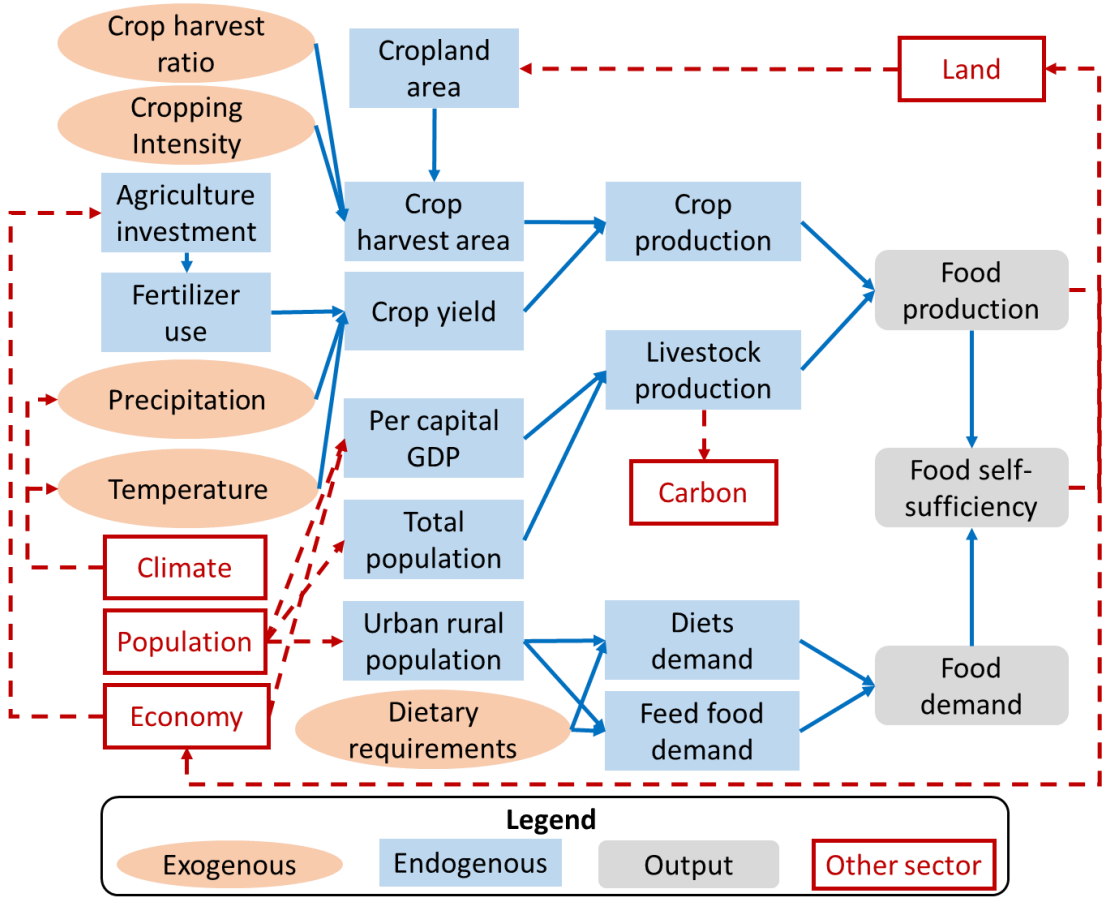

Fig.6 Structure of the *Food* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Food* sector simulates the production of livestock and crops, and food demand, and these productions directly affect gross agricultural production in the *Economy* sector (Fig. 6). Livestock production ($Livestock_{pro}$) is calculated by an empirical function, driven by economic development and population growth (Equation 3),

$$Livestock_{pro} = Pop \times \left(Para_{GL1} + (Para_{GL2} - Para_{GL1}) \times \frac{PC_{GDP}}{PC_{GDP} + Const_{GL}}\right) \quad (3)$$

where $Para_{GL1}$、$Para_{GL2}$ and $Const_{GL}$ are parameters obtained by fitting the historical per capita meat production and per capita GDP (PC_{GDP}) data.

Crop production ($Crop_{pro}$) is determined by the yield ($Yield_c$) and harvest area (PA_c) of seven major crop types (subscript c): rice, wheat, corn, soybeans, cotton, potatoes,

and oil crops (Equation 4). The harvest area is influenced by cropland area from the *Land* sector, exogenous cropping intensity and crop harvest ratio. Crop yields are positively affected by agricultural investment from the *Economy* sector, along with effects of precipitation, temperature, and CO_2 concentration from the *Climate* sector (Equations S30-S36).

$$Crop_{pro} = \sum_{c=1}^{7} Yield_c \times PA_c \quad (4)$$

Food demand encompasses both staple and feed grain demand, which are determined by population size from *Population* sector and dietary patterns (Equations S37-S41).

2.2.5 Water Demand

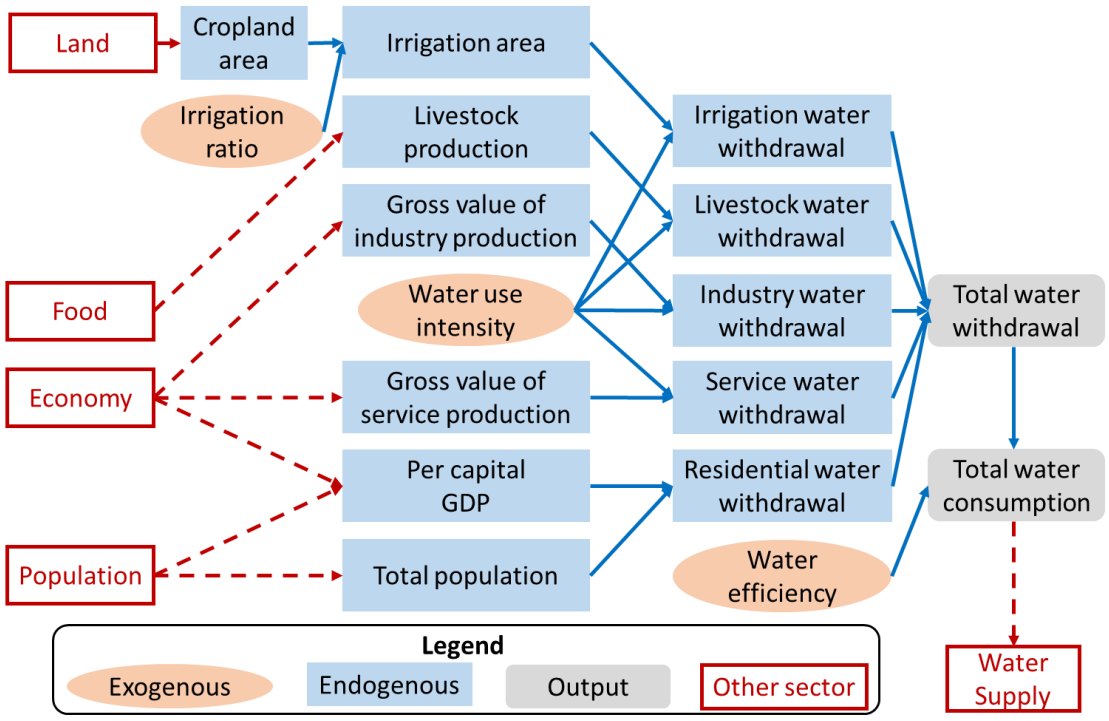

Fig.7 Structure of the *Water Demand* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Water Demand* sector simulates water withdrawal (*WW*) and consumption (*WC*) across multiple uses—irrigation, livestock, industry, services, and residential—in the nine provinces of the YRB (Fig. 7). Water consumption is derived as the product of water withdrawal and water use efficiency reported in the Water Resources Bulletin

(YRCCMWR, 2020) (Equations S42-S43), which affects the *Water Supply* sector.

Irrigation water withdrawal (WW_{irr}) is estimated through a physically-based function of exogenous irrigation water use intensity (WWI_{irr}), cropland irrigation ratios (IR) from the China Agricultural Yearbook (MAARA, 2020), and cropland area ($Area_{Cropland}$) provided by the *Land* sector (Equation 5).

$$WW_{irr} = WWI_{irr} \times Area_{Cropland} \times IR \quad (5)$$

The water withdrawal for livestock, industry, and services is also driven by exogenous sectoral water use intensities collected from the China Statistical Yearbook and National Long-term Water Use Dataset of China (Zhou et al., 2020), in combination with livestock production from the *Food* sector, economic output from the *Economy* sector (Equations S45-S48).

Residential water withdrawal (WW_{res}) is derived from a non-linear empirical function of economic development and population growth with an upper limit of per capita domestic water use (Flörke et al., 2013) because water demand per person cannot increase indefinitely (Equation 6),

$$WW_{res} = Pop \times \left(Para_{GD1} + (Para_{GD2} - Para_{GD1}) \times \frac{PC_{GDP}}{PC_{GDP} + Const_{GD}}\right) \quad (6)$$

where $Para_{GD1}$, $Para_{GD2}$, and $Const_{GD}$ are parameters obtained by fitting the historical per capita domestic water use with per capita GDP (PC_{GDP}).

2.2.6 Water Supply

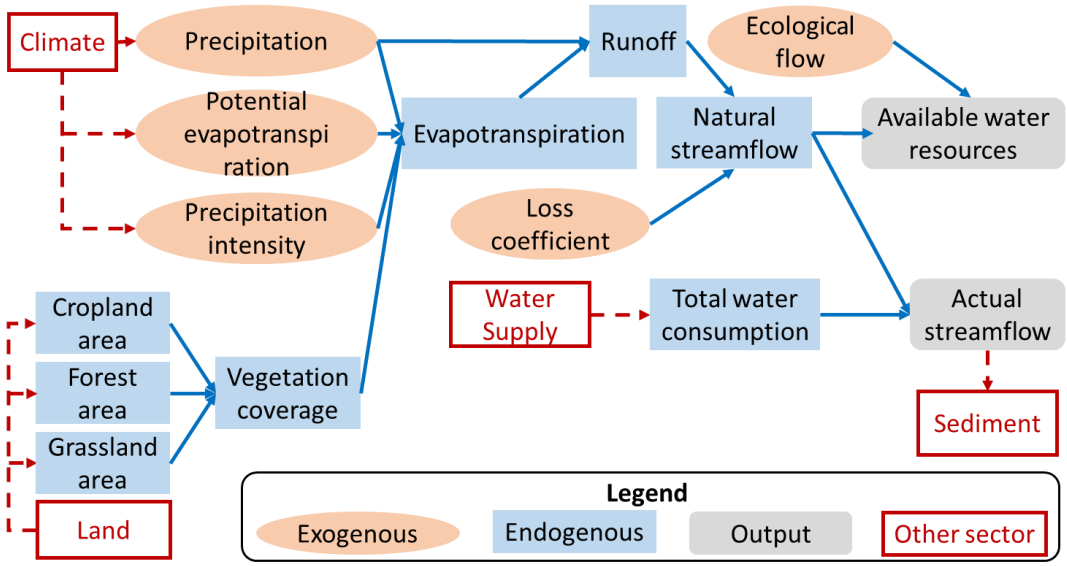

Fig.8 Structure of the *Water Supply* sector. Blue lines indicate connections inside the sector, red dotted lines indicate connections with other sectors.

The *Water Supply* sector simulates runoff, discharge, and their changes in each sub-basin (Fig. 8). Runoff (R) is determined by precipitation (Pre) and evapotranspiration (ET) based on the water balance principle derived from the Budyko equation, which is suitable for non-humid regions of China (Yang et al., 2009) (Equation 7).

$$R = Pre - ET \tag{7}$$

The ET is calculated by various exogenous climate variables from the *Climate* sector, and vegetation coverage from the *Land* sector (Equations 8-9, S51-S53),

$$ET = \frac{PET \times Pre}{(Pre^n + EP^n)^{\frac{1}{n}}} \tag{8}$$

$$n = Para_{E1} \times \left(\frac{Ks}{PI}\right)^{Para_{E2}} \times FVC^{Para_{E3}} \times e^{Para_{E4} tan\beta} \tag{9}$$

where PET is potential evapotranspiration at sub-basin, from the *Climate* sector; n is a parameter reflecting the basin landscape characteristics, related to saturated hydraulic conductivity (Ks), precipitation intensity (PI), average slope (β), and fraction of vegetation coverage (FVC); $Para_{E1}$, $Para_{E2}$, $Para_{E3}$, and $Para_{E4}$ are parameters fitted from historical data.

Runoff and the loss coefficient (defined by the ratio of natural streamflow to runoff) determine the natural streamflow due to the water loss during the confluence process (Equation S54). Natural streamflow and ecological flow constraints define the upper limit of available water resources. By integrating human water consumption from the *Water Demand* sector across sub-basins, the model calculates the actual streamflow (Equations S55-S57). Actual streamflow is transferred through hydrological connectivity from upstream to midstream and then to downstream, ultimately reaching the sea, which governs sediment transport processes within the *Sediment* sector.

2.2.7 Sediment

The *Sediment* sector estimates sediment load (*Sed*) for each sub-basin using an empirical model in the literature (Yin et al., 2023b), which links actual streamflow from the *Water Supply* sector to sediment transport,

$$Sed = Para_{SS} \times AS + Const_{SS} \quad (10)$$

where $Para_{SS}$ and $Const_{SS}$ are derived from linear fitting of historical hydrological station data on actual streamflow and sediment load.

2.2.8 Land

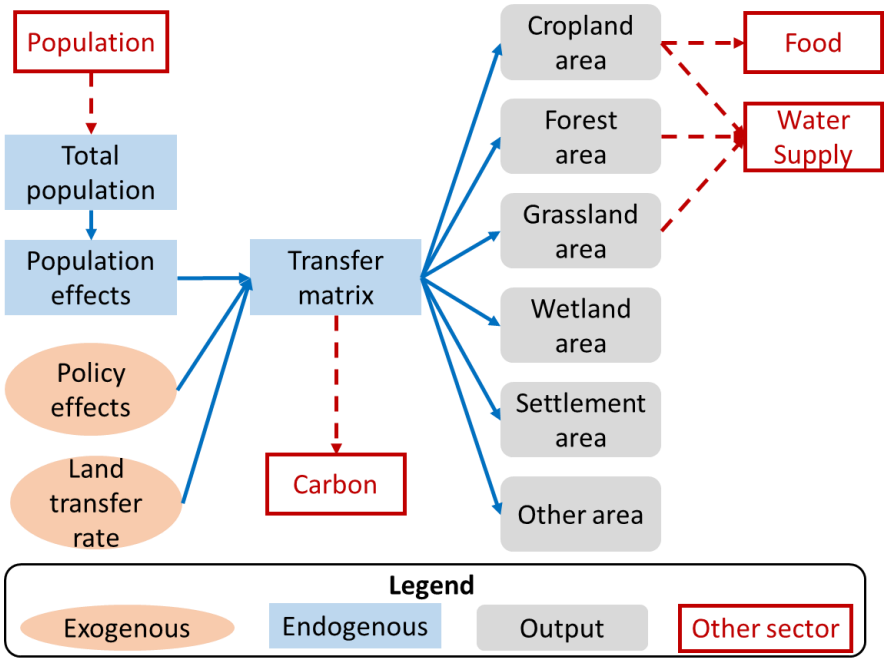

Fig.9 Structure of the *Land* sector. Blue lines indicate connections inside the sector,

red dotted lines indicate connections with other sectors.

The *Land* sector simulates the area changes of six land use types (cropland, forest, grassland, wetland, settlement, and others) based on the land transfer matrix obtained from historical remote sensing data (Xu et al., 2018). The land transfer matrix calculates the inflow and outflow of each land category and can be configured to represent the influence of future land use drivers (Fig. 9). This sector outputs vegetation area (including forest, grassland, and cropland), which influences the *Water Supply* sector. Cropland area changes impact the *Food* sector, while land use conversion also plays a role in the *Carbon* sector.

Based on the initial land use area ($IniArea_i$), the transfer matrix determines the area ($Area_i$) allocated to each land use type (Equations 11, S60-S66),

$$Area_i = IniArea_i + \int \left(\sum_{i=1}^{6} FTM_{i,j} - \sum_{j=1}^{6} FTM_{i,j}\right)dt \quad (11)$$

where $FTM_{i,j}$ is the finial land use transfer matrix, indicating the area of land use i transferred to land use j, i and j represent six land use type.

2.2.9 Carbon

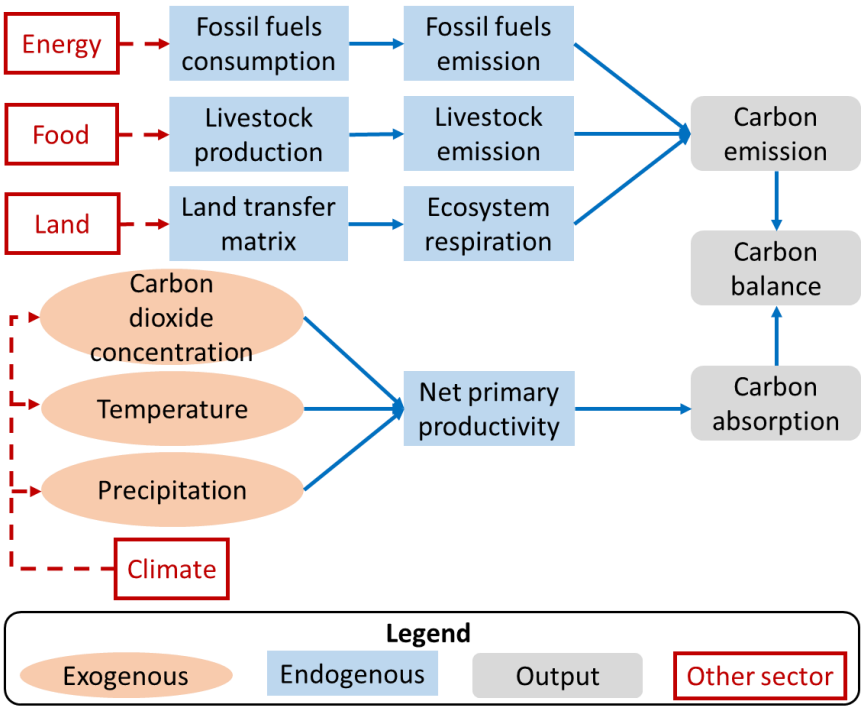

Fig.10 Structure of the *Carbon* sector. Blue lines indicate connections inside the

sector, red dotted lines indicate connections with other sectors.

The *Carbon* sector simulates the basin's carbon balance processes, including carbon emission and absorption, adapted from the carbon cycle of ANEMI (Davies and Simonovic, 2011). Carbon emissions (*CE*) encompass fossil fuel emissions, and livestock emission, and ecosystem respiration, which are influenced by outputs from the *Energy*, *Food*, and *Land* sectors (Fig. 10).

Fossil fuel and livestock emissions are calculated based on fossil fuel consumption, livestock production, and their respective emission coefficients. Ecosystem emissions include carbon released from burning as well as from the decomposition of biomass, litter, humus, and charcoal, which are determined by carbon pools' lifespans, decomposition factor, and respiration coefficients (Equations S67-S93).

Carbon absorption (*CA*) is determined by net primary productivity (*NPP*) and land use area. NPP is influenced by climate factors, including CO_2 concentration, temperature, and precipitation (Equations 12, S63),

$$CA = \sum_{i=1}^{6} NPP_i \times Area_i \quad (12)$$

where i represent land use type, land use area ($Area_i$) from the *Land* sector.

2.2.10 Climate

The *Climate* sector supplies both historical and projected climate data essential for the *Water Supply*, *Food*, and *Carbon* sectors. The key climate variables comprise temperature and precipitation (at the sub-basin and provincial levels), as well as potential evapotranspiration, precipitation intensity, and CO_2 concentration (at the sub-basin level). These variables are treated as exogenous inputs, without accounting for potential feedbacks from human activities or natural system responses on regional climate patterns.

2.3 Data sources

In the model, there are more than 100 exogenous variables, some of which are initial variables that drive the simulation. These exogenous variables are derived either from historical statistical data or from fitted results based on historical data (Table S2).

All data sources required for the model simulation are listed in Table 1.

Table 1. Summary of data sources for the YRB model. Updated from Sang et al. (2025).

Sector	Variables*	Data sources	Spatial and temporal scale	Time range
Population	age-specific population, births, deaths, migrants	• China Statistical Yearbook (NBSC, 2020c), • China Population Census Yearbook (NBSC, 2020b)	• provincial, annual	• 1981–2023
Economy	GDP, investment, employment	• China Agricultural Yearbook (MAARA, 2020), • China Statistical Yearbook (NBSC, 2020c)	• provincial, annual	• 1981–2023
Energy	electricity, coal, oil, gas consumption	• China Energy Statistical Yearbook (NBSC, 2020a)	• provincial, annual	• 1981–2023
Food	crop production, fertilizer, irrigated area	• China Agricultural Yearbook (MAARA, 2020)	• provincial, annual	• 1981–2023
Water demand	Industrial, domestic, irrigation water withdrawal	• National Long-term Water Use Dataset of China (NLWUD) (Zhou et al., 2020) • China Statistical Yearbook (NBSC, 2020c)	• prefectural, annual	• 1965–2013 • 1981–2023
Water supply	runoff, streamflow, water stress	• China Natural Runoff Dataset (CNRD) (Gou et al., 2021), • Gauge-based Natural Streamflow Dataset (Miao et al., 2022) • Yellow River Water Resources Bulletin (YRCCMWR, 2020)	• 0.25° grid, Monthly • 0.25° grid, annual • provincial, annual	• 1961–2018 • 1961–2018 • 1998–2022
Sediment	sediment load	• Yellow River Water	• provincial,	• 1998–2022

		Resources Bulletin (YRCCMWR, 2020)	annual	
Land	cropland, forest, grassland, wetland, settlement	• China land use remote sensing monitoring dataset (CNLUCC) (Xu et al., 2018) • Long-term global land surface satellite fractional vegetation cover product (Jia et al., 2015, 2019)	• 1 km grid, annual	• 1980, 1990, 1995, 2000, 2005, 2010, 2015, 2020
Carbon	fossil fuel carbon emissions, net primary productivity	• Carbon Emission Accounts and Datasets (Shan et al., 2018), • Global Vegetation Productivity Products (GVPP) (Cui et al., 2016; Wang et al., 2020a; Yu et al., 2018)	• Provincial, annual • 5 km grid, 8 day	• 1997–2021 • 1981–2018
Climate	temperature, precipitation, CO_2 concentration, potential evapotranspiration and precipitation intensity	• Gridded daily observation dataset over China region (Wu and Gao, 2013), • European Environment Agency (EEA, 2019), • China Meteorological Data Service Center (https://data.cma.cn/)	• 0.25° grid, Annual • Global, Annual • Meteorological stations, Monthly	• 1961–2020 • 1800-2020 • 1956-2018

423 * Listed are some typical variables of each sector

3 Model validation and application

3.1 Historical model validation

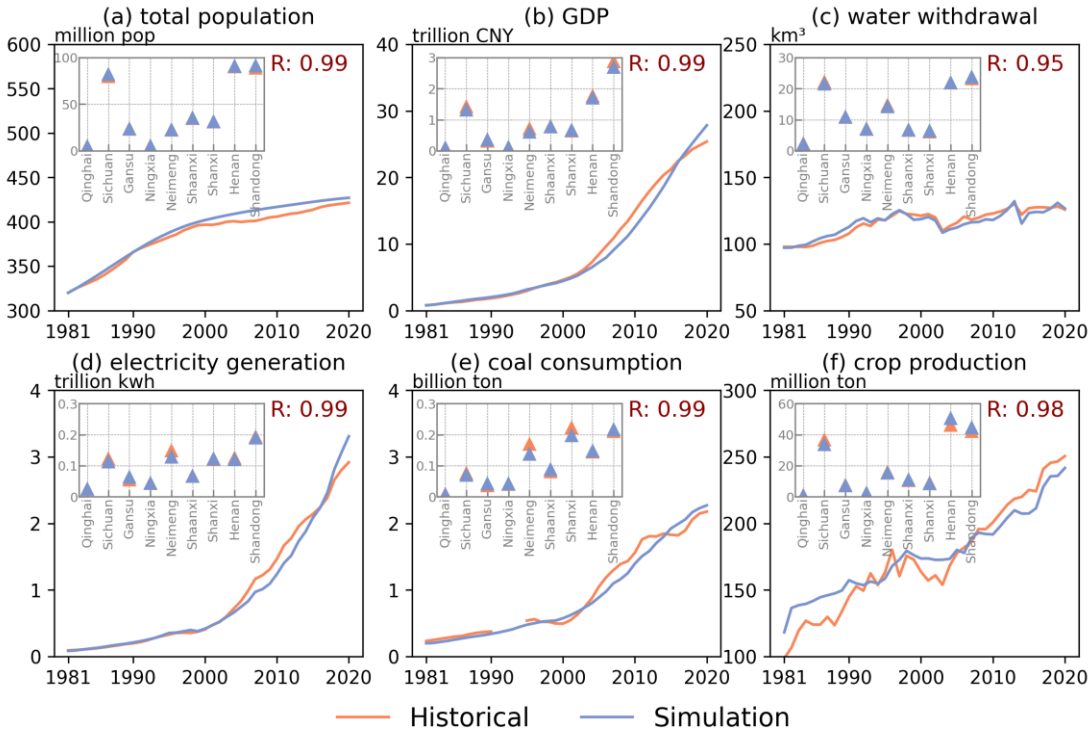

Fig.11 Validation of human system processes during historical period of 1981-2020. (a) total population in *Population* sector; (b) GDP in *Economy* sector; (c) water withdrawal in *Water Demand* sector; (d) electricity generation and (e) coal consumption in *Energy* sector; (f) crop production in *Food* sector. Pink triangles in the upper left sub-image represents the average of historical and simulation value in 1981-2020 of each province in the YRB.

We compared model performance during the historical period (1981–2020) against historical data. These data are sourced from multiple channels, including statistical yearbooks, hydrological stations, remote sensing observations, and outputs from other models (Table S3).

For human sectors, the simulation accuracy of selected key variables from each sector is consistently high ($R > 0.95$), including total population, gross domestic product (GDP), water withdrawal, electricity generation, coal consumption, and crop production across the nine provinces (Fig.11). However, the accuracy varies among

provinces, particularly for GDP, electricity generation, coal consumption, and crop production. Some provinces, like Neimeng, exhibit lower simulation accuracy for certain indicators than others, probably due to the simplification of relevant modelled processes, imperfect parameterizations, and external policy interventions. Among all sectors, the greatest uncertainty arises in simulating coal consumption, because the absence of a clear historical trend results in poor fitting performance. Nevertheless, the model performs reasonably well in the human sectors, effectively capturing the historical dynamics of human systems in the YRB.

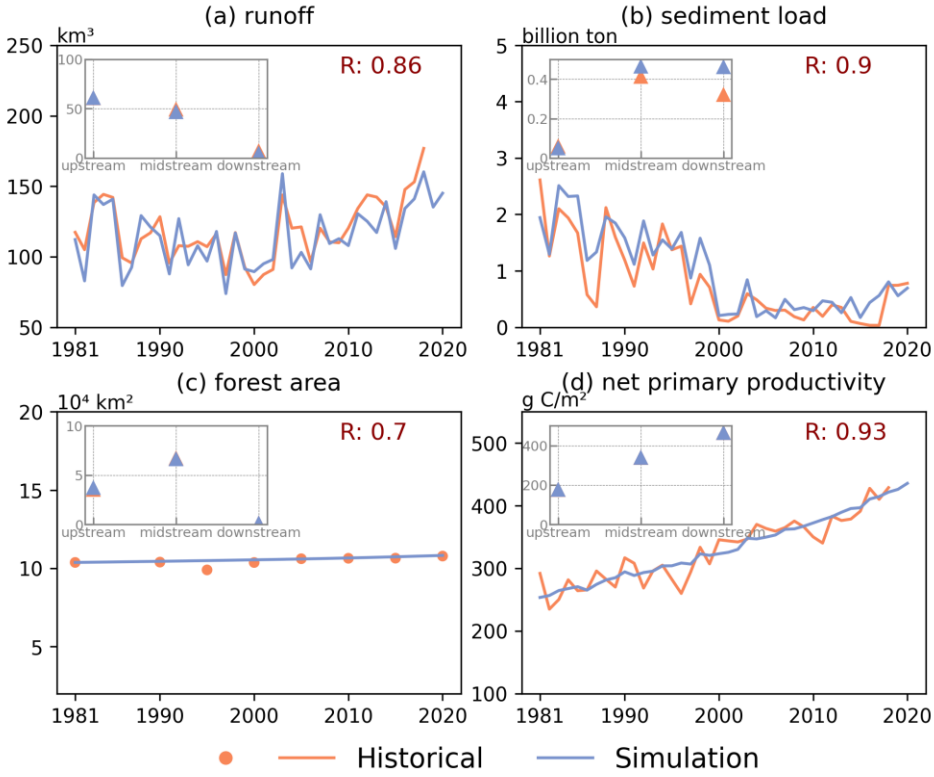

Fig.12 Validation of the natural system processes during historical period of 1981-2020. (a) runoff in *Water Supply* sector; (b) sediment load in *Sediment* sector; (c) forest area in *Land* sector, and the historical dataset is discontinuous; (d) net primary productivity in *Carbon* sector. Pink triangles in the upper left sub-image represents the average of historical and simulation value in 1981-2020 of each sub-basin in the YRB.

For natural sectors, the simulation accuracy for runoff, sediment load, forest area,

and NPP in the YRB is relatively high (R > 0.7) (Fig.12). At the sub-basin level, the overall performance is satisfactory; however, the model shows lower accuracy in simulating sediment transport in the mid- and downstream. Compared to the human sectors, the natural sectors generally exhibit lower correlations with historical data, largely due to the simplifications required when modeling physically complex natural processes. Nevertheless, the model is still capable of capturing the historical dynamics of natural processes in the YRB.

Sensitivity analyses with randomly generated parameter values reveal growing uncertainty in the long-term trajectories of socio-economic and natural variables; the model maintains behavioral robustness across all runs without catastrophic collapse or unrealistic oscillations (see Supporting Information S3 for details).

3.2 Model application for future projection

3.2.1 Future baseline scenario

The future baseline scenario represents a trajectory in which existing plans and policies continue to operate without substantial changes in external environments. The development of the baseline scenario primarily relies on variables from the *Population*, *Economy*, *Water Demand*, and *Land* sector based on available government planning documents, historical trends, and other projections (Table 2). For future projections, variables or parameters not specified in the table are held constant at historical levels from their most recent year. Future climate data (2015-2100) are from the ensemble mean of 11 CMIP6 models (ACCESS-CM2, CESM2, CMCC-ESM2, GFDL-ESM4, INM-CM4-8, INM-CM5-0, IPSL-CM6A-LR, MIROC6, MRI-ESM2-0, UKESM1-0-LL, and HadGEM-GC31-LL) (available at https://cds.climate.copernicus.eu/datasets/projections-cmip6?tab=download) for the SSP 2-4.5 because this scenario aligns more closely with the climate trends in the YRB. To ensure temporal consistency, CMIP6 historical climate data (1981–2014) were used instead of observed historical records. To reduce systemic biases in the raw CMIP6 data, we applied bias correction using CN05 and ground weather stations observations from

1981 to 2014. By statistically aligning the mean of the CMIP6 data with observation records, systemic bias is removed while retaining the temporal consistency required for long-term simulation.

We run the model under the future baseline scenario to project the evolution of CHANS in the basin from 2021 to 2100 (see Supporting Information S6 for details of future scenario design and supplementary spreadsheet for exogenous variables).

Table 2 Settings of key variables in the future baseline scenario

Sector	Variables	Future baseline
Population	• Total fertility	• 2024 United Nations Population Projections
	• Gender ratio	• 2024 United Nations Population Projections
	• Average years of schooling	• The national maximum in 2023 (Germany) as the upper limit
	• Urban and rural population ratio	• Analysis and Forecast of Urbanization Trends in China's Modernization by 2020
	• Labor force	• Delay retirement age
Energy	• Electricity generation sources ratio	• Projected provincial-level energy structure meeting China’s carbon peaking and carbon neutrality goals (Li et al., 2024)
	• Irrigation intensity	• Irrigation water use intensity declines linearly (1%/yr) to its minimum value.
Water Demand	• Industry, service, livestock water use intensity	• Industry, service and livestock water use intensity declines linearly (1%/yr) to its minimum value.
Land	• Land use area	• Cropland, forest, and grassland are allowed to be converted into each other, but not to unused land; unused land can be converted to the rest land; wetlands are fixed. • Minimum cropland area above the red line

		for cropland protection.
Climate	• Temperature • Precipitation • Potential evapotranspiration • Precipitation intensity • CO_2 concentration,	• SSP 2-4.5 —— CMIP 6

3.2.2 Projection of CHANS dynamics in future baseline scenario

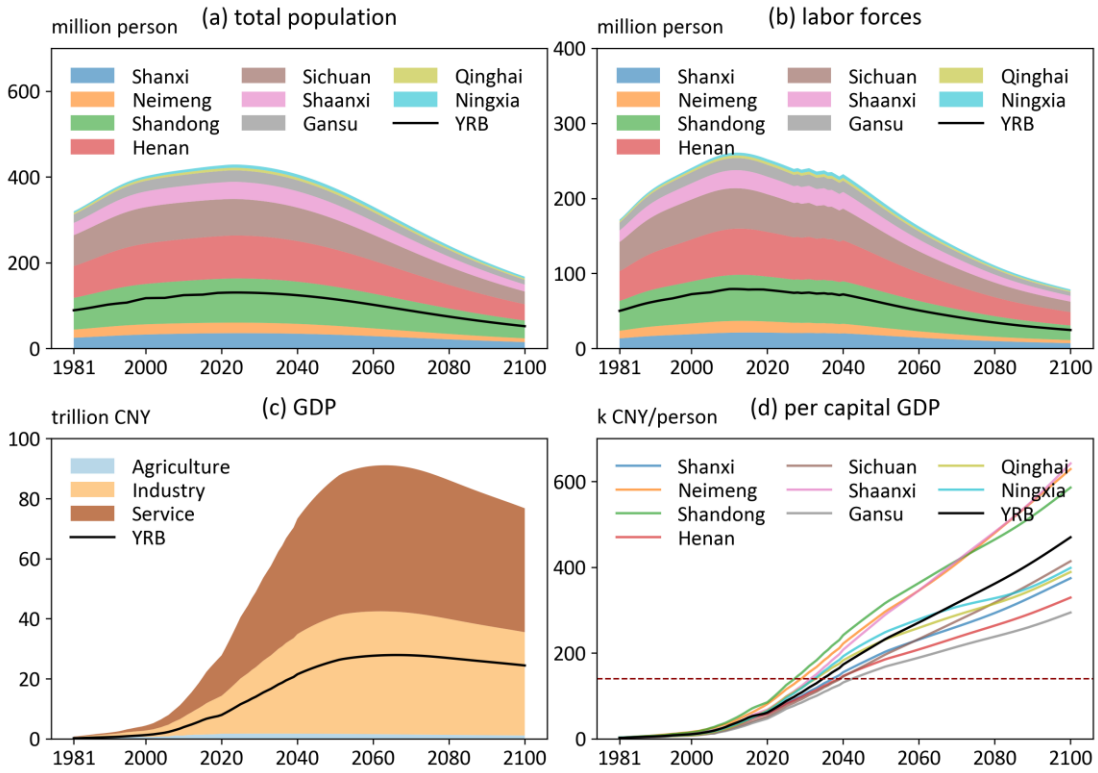

Fig.13 Changes in key variables of the *Population* and *Economy* sector in the future baseline scenario: (a) the total population, (b) the total labor forces, (c) the gross output of agriculture, industry, and service, and the total gross domestic output, (d) the per capita GDP in nine provinces and YRB.

The model simulates human and natural system dynamics under the future baseline scenario and produces outputs for nine provinces and their corresponding areas within the YRB. The simulation results are reported either at the provincial level (nine

provinces) for human system sectors or at the basin level (basin boundary) for natural system sectors. The total population in the nine provinces and the YRB is projected to peak in 2023 and 2024, at 429 million and 131 million, respectively, driven by declining fertility rates (Fig.13 (a)). The labor force peaks earlier, reaching 261 million in the nine provinces in 2012 and 79 million in the YRB in 2011 (Fig.13 (b)). After 2025, the labor force is projected to increase again due to the delayed retirement policy. GDP in the nine provinces is expected to increase until a peak in 2063, at 91 trillion CNY (three times the 2020 level, in 2020 constant prices), under the influence of a labor force decline, with the YRB reaching its peak four years later (Fig.13 (c)). In contrast to total GDP, per capita GDP in all regions is projected to continue rising throughout the future period (Fig.13 (d)). Although all provinces demonstrate improvements in economic status and living standards, considerable regional disparities persist. Among the nine provinces, Shaanxi demonstrates the largest growth in average per capita GDP from 2021 to 2100 relative to the historical period (1981-2020), with a more than seventeenfold increase, while the YRB as a whole shows an over thirteenfold increase.

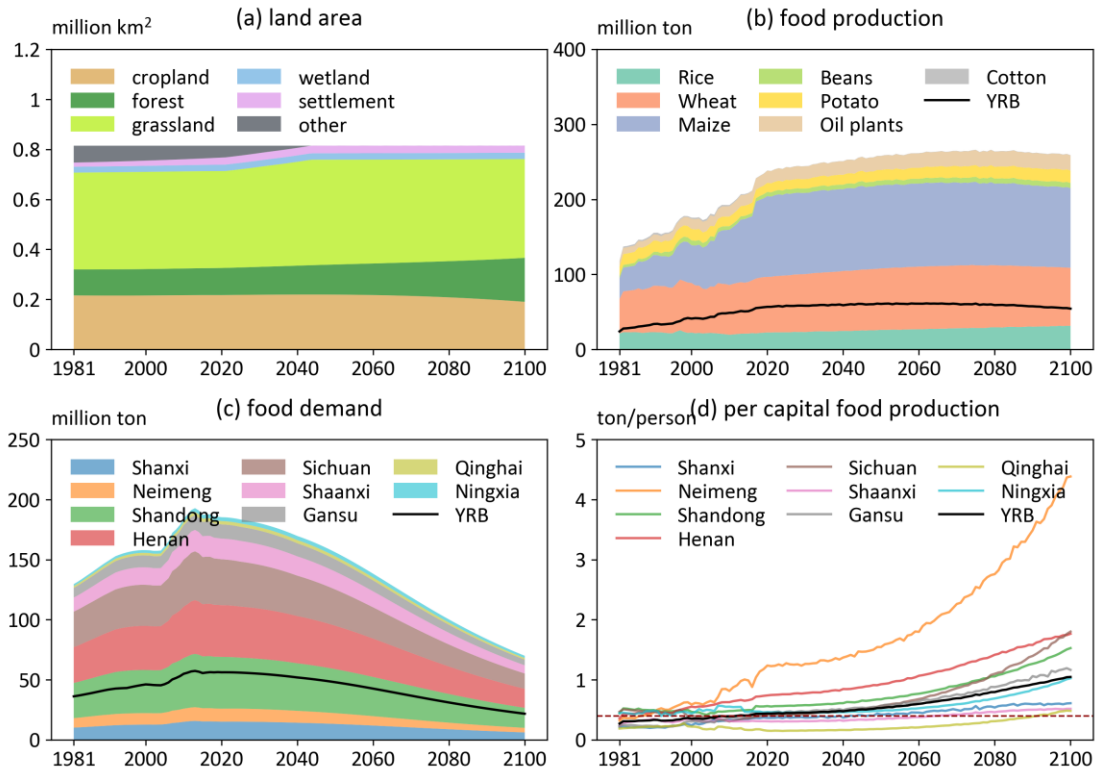

Fig.14 Land and food in the future baseline scenario. (a) land area in the YRB; (b-d) food production of seven crop types, food demand and per capita food production in nine provinces and the YRB, red dotted line represents the 0.4 tons international food security threshold.

Under the future baseline scenario, land use patterns remain relatively stable based on historical trends, as no new land policies are introduced and cropland area has to stay above the red line (a mandatory minimum cropland area for each province for food security) (Fig.14 (a)). The forest area is projected to increase gradually, reaching 62% above the 2021 level by 2100, while the cropland area is expected continue declining, falling 12% below the 2021 level. Total crop production in nine provinces and YRB is projected to peak in 2079 and 2062 (Fig.14 (b)), respectively, driven by the combined effects of declining cropland area and increasing crop yields. Food demand in both the nine provinces and the YRB as a whole peaked in 2013 (193 million tons and 57 million tons, respectively) (Fig.14 (c)), largely driven by population dynamics. In the future, per capita food production in the YRB is projected to consistently exceed the international food security threshold of 0.4 tons per person (Fig.14 (d)). However, in certain years, provinces such as Qinghai and Shaanxi could fall below this standard.

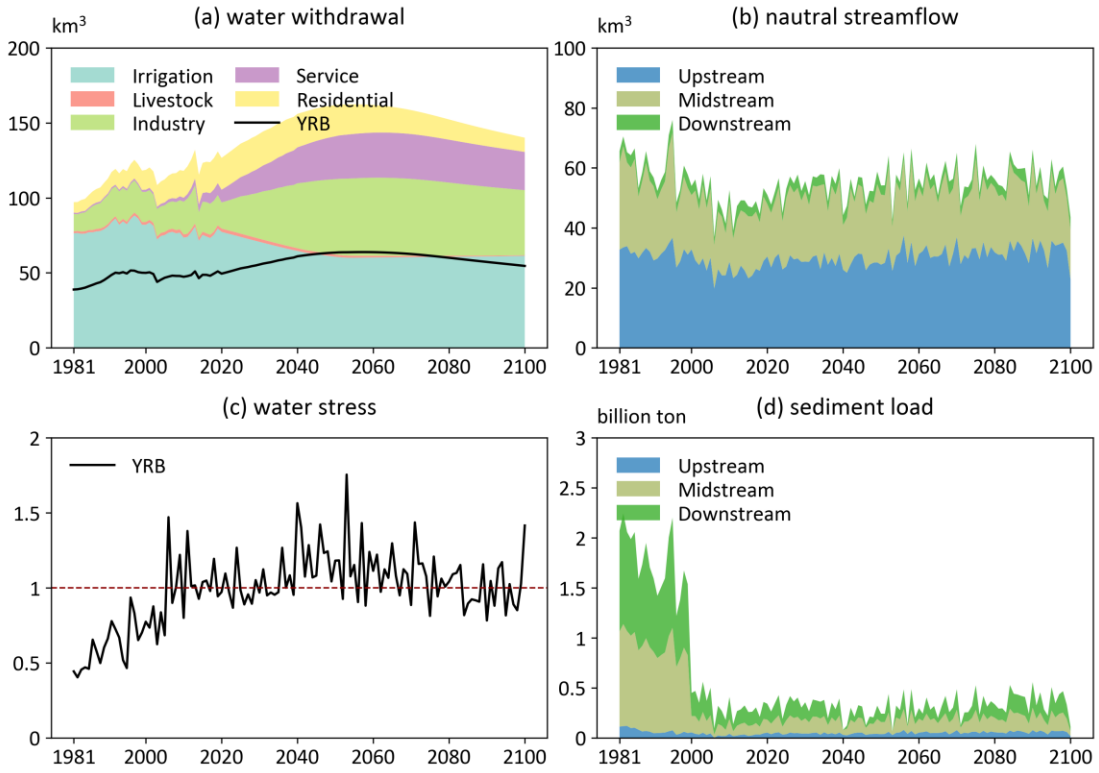

Fig.15 Water demand and supply in the future baseline scenario. (a) water withdrawal across five water-use sectors in nine provinces and the YRB; (b) natural streamflow in up-, mid- and downstream; (c) water stress in the YRB, red dotted line represents the water security threshold; (d) sediment load in the up-, mid- and downstream.

Total water withdrawal in the nine provinces and the YRB is projected to peak in 2056 and 2058, reaching 162 km³ and 64 km³, respectively (Fig.15 (a)). Irrigation water withdrawal is expected to decline, driven by reductions in cropland area and irrigation intensity, the latter resulting from improvements in irrigation efficiency assumed in the scenario. Peaks in residential, industrial, service, and livestock water withdrawals are primarily associated with projected peaks in population and GDP. Natural streamflow is projected to exhibit a fluctuating upward trend of 0.073 km³/year basin-wide (Fig. 15(b)), with the most significant growth in the upstream region (0.068 km³/year). Considering the ecological flow requirement of 18.7 km³ (YRCCMWR, 2015), water stress (calculated as water consumption divided by (natural discharge - ecological flow)) is expected to decline in the latter half of the century, largely due to the peak and

subsequent reduction in water withdrawal. However, overall water stress is projected to exceed historical levels in 2041 and will fall below 1 in 38% of the years (Fig.15 (c)), reflecting persistent human–water tensions and future trade-offs between ecological and human water use. As a result of reduced actual streamflow, sediment transport in the Yellow River is projected to decline sharply (Fig.15 (d)), with a 71% decrease in sediment load relative to the historical period.

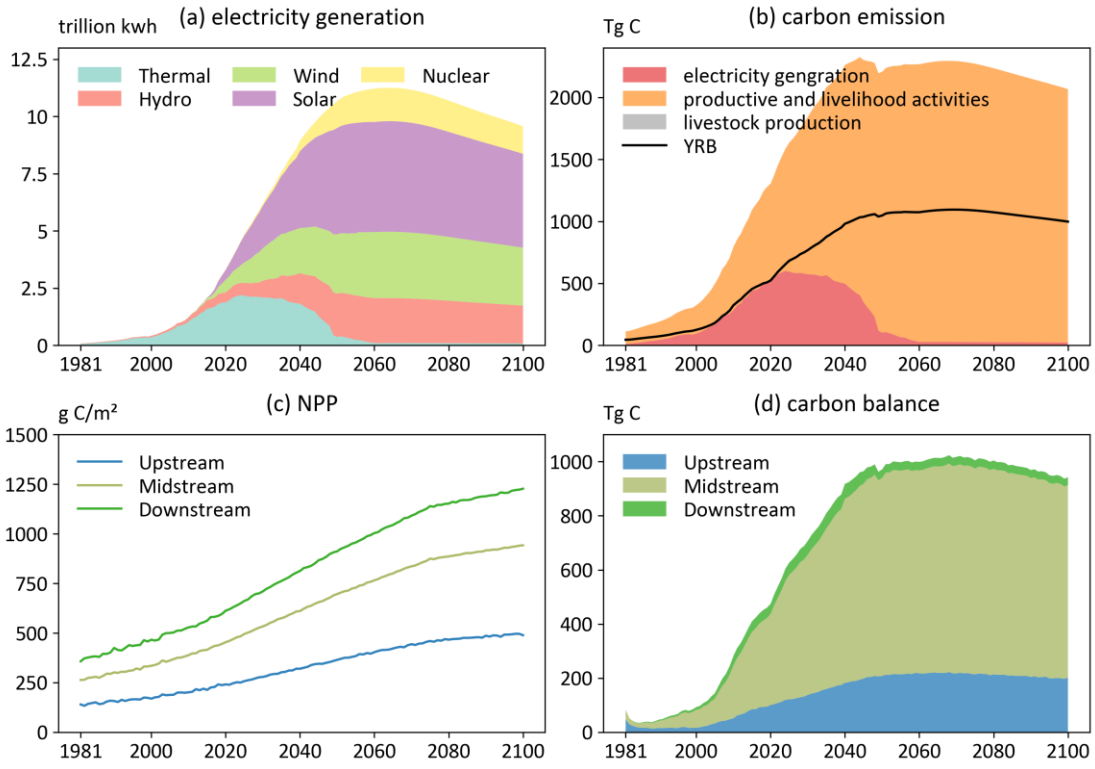

Fig.16 Electricity and carbon in the future baseline scenario. (a) the electricity generation from five sources in nine provinces; (b) carbon emission from three sources in nine provinces and the YRB; (c-d) the net primary productivity and carbon balance in the up-, mid-, and downstream.

Since the energy system is transitioning to carbon neutrality goals under the future baseline scenario, thermal power generation is projected to decline sharply, while clean energy (hydropower, wind, solar, and nuclear) will become the dominant source for electricity generation (Fig.16 (a)). The large-scale adoption of clean energy will lead to a peak in carbon emissions from electricity generation in 2024 (Fig.16 (b)). In contrast, emissions from productive and livelihood activities are projected to rise substantially

over time, with total human emissions in the YRB reaching a peak in 2044. NPP is anticipated to increase markedly in response to CO_2 concentration increase (Fig.16 (c)). Although the carbon balance of the YRB has remained positive (net carbon source), it is expected to gradually decline after 2068 (Fig.16 (d)), indicating that the basin's ecosystems alone cannot fully offset human-induced carbon emissions.

The above projections of future baseline scenarios highlight persistent challenges to achieving sustainable development in the YRB, underscoring the need for integrated policy responses to address these challenges. Priority should be given to enhancing water-use efficiency and establishing adaptive water resource allocation mechanisms that balance ecological and human water needs. Sediment management strategies, which combine ecological restoration with hydraulic regulation, are necessary to maintain river stability and delta health. To counter the decline in labor force, policies should promote industrial upgrading and technological innovation. Strengthening the basin's carbon mitigation capacity requires accelerating the clean energy transition and expanding ecological restoration to enhance carbon sequestration. Finally, a multi-sectoral, CHANS-based governance framework should be established to coordinate water, land, energy, and carbon management, enabling evidence-based scenario analysis and adaptive policy design for long-term sustainability.

4 Discussion and conclusions

The development of the CHANS-SD-YRB model involves a series of key considerations in translating a conceptual framework into a functioning model. The framework focuses on human–water interactions in the YRB, integrating bidirectional feedback between human and natural processes. SD was chosen to implement the framework, as it effectively captures feedback, non-linear relationships, and cross-sectoral linkages within complex human–natural systems, while remaining practical and straightforward to use. For quantitative representation of human–natural processes, we prefer theoretical methods that incorporate key interactions within the system suitable for implementation using SD. For example, the Budyko framework connects

595 the *Land*, *Climate*, *Water Demand*, and *Water Supply* sectors. Relative to distributed
596 hydrological models, it offers lower computational complexity and reduced data
597 requirements. The Cobb–Douglas production function links the *Population* and
598 *Economy* sectors. Compared with Computable General Equilibrium models (Fujimori
599 et al., 2014a), it is simpler to construct and requires fewer parameters and data inputs.
600 When no suitable theoretical framework/method is available to describe a process, we
601 rely on empirical relationships derived from historical data. For example, crop yields
602 are estimated using fitted relationships between climate variables, fertilizer application,
603 and historical yields. For processes that cannot be represented through empirical
604 functions, we apply literature-based estimates to obtain approximate quantitative
605 relationships, such as fossil fuel emission factors. For processes influenced by external
606 drivers that cannot be endogenously expressed within the model, we quantify them
607 using exogenous parameters derived from historical statistics, such as water intensity.
608 Given the heterogeneous human–natural interactions in the YRB, we represent all
609 processes at the provincial and sub-basin scales using scale-specific parameters, except
610 where parameters or data at these levels are unavailable.

611 The CHANS-SD-YRB model explicitly couples multiple human and natural
612 sectors, enabling a more integrated representation of feedbacks across population,
613 economy, energy, food, water, sediment, land, carbon, and climate. Unlike models that
614 focus on specific sectors or isolated subsystems—such as eco-hydrological models and
615 sediment transport models, which may well capture individual processes but cannot
616 represent the complex human–nature interactions driving system dynamics. Our
617 model's integration of human and natural sectors provides a robust framework for
618 addressing regional CHANS challenges and offers practical guidance for sustainable
619 development. The CHANS-SD-YRB model serves as a comprehensive platform for
620 conducting system dynamics prediction, scenario analysis, policy evaluation, and
621 optimization, to alleviate human–water conflicts, ensure food security, and achieve
622 long-term sustainability. For example, analysis of the impacts of the 1987 Yellow River

623 water allocation policy (Song et al., 2024) offer valuable insights for adjusting
624 interprovincial water distribution to promote sustainable water governance; assessment
625 of ecological restoration policies (Li et al., 2015) can guide future ecological
626 engineering; spatiotemporal dynamics of future water gaps can serve as a valuable
627 reference for planning inter-basin water transfers. The model's flexibility also allows
628 for the incorporation of additional feedback, for example, linking water scarcity to
629 industrial output or climate warming to agricultural productivity. These could include
630 the effects of global warming on human health (Yin et al., 2023a) and economic
631 activities (Nordhaus, 2017), water constraints on production, dietary shifts influencing
632 carbon emissions and land use (Ren et al., 2023), and the trade-offs between carbon
633 mitigation and food security (Xu et al., 2022).

634 Nevertheless, the model remains subject to further refinement. The current
635 simplifications of natural processes could be replaced with more sophisticated models
636 to enhance simulation accuracy. For instance, the YRB hydrological processes involve
637 highly complex human interventions, including reservoir, conservation, and
638 revegetation projects (Wang et al., 2025). The refined runoff simulations by distributed
639 hydrological models improve water supply assessments through their better
640 characterization of spatial heterogeneity in soil, precipitation, and snowmelt (Cong et
641 al., 2009). For human processes, the *Energy* and *Economy* sectors could be refined to
642 model more detailed industry subsectors and emerging trends in energy demand driven
643 by electrification. Moreover, there are still important feedbacks absent from the current
644 coupling framework. Notable examples include the effects of land use change on
645 climate, the effects of climate change on economic growth, and the effects of pricing
646 on energy use and carbon emissions. These missing feedbacks could be incorporated
647 based on recent studies, including the land use feedback on precipitation through
648 moisture recycling (Sang et al., 2025a), socioeconomic losses from impacts of climate
649 change and extremes (Waidelich et al., 2024b), and integration with Computable
650 General Equilibrium (CGE) models (Fujimori et al., 2014b). Additionally, the model's

spatial and temporal resolutions are relatively coarse due to the inherent mismatch in spatiotemporal scales between human and natural processes. The scale at which we make the model represents a practical compromise, constrained by data availability and aligned with the study's objectives. Simulations at finer scales (e.g., monthly, daily, or gridded) would enhance the representation of natural processes (e.g., hydrological processes, land use changes, and food production) and the associated spatiotemporal heterogeneity (e.g., daily simulations can assess the impacts of extreme weather). However, it may also exacerbate the scale mismatch with socioeconomic processes, making it challenging to analyze cross-sectoral dynamics. Technically, the inherent limitations of the SD software restrict the model's ability to represent temporal fluctuations and spatial variations. To overcome this, transitioning from the VENSIM platform to a code-based implementation will be necessary, which would also facilitate coupling with other models. Future research should prioritize these improvements to strengthen both the performance and applicability of the model.

Drawing on the conceptual framework of Sang et al. (2025), this study implements it into a fully functional, validated System Dynamics model tool, CHANS-SD-YRB 1.0. The model fills the gap in CHANS modeling for the YRB. It integrates the dynamics of ten interconnected sectors: *Population*, *Economy*, *Energy*, *Food*, *Water Demand*, *Water Supply*, *Sand*, *Land*, *Carbon*, and *Climate*, achieving reciprocal feedback between human and natural systems. This model can serve as a robust tool to inform policy decisions that influence the evolution of coupled human–natural systems and to explore pathways for optimizing these systems toward sustainability. Furthermore, the modeling process provides valuable experience for regional CHANS modeling and contributes to advancing the broader development of CHANS models at the regional scale.

Code and data availability

The CHANS-SD-YRB V1.0 model (Vensim DSS format), along with the input data and simulation outputs used in this study, are openly accessible at

https://doi.org/10.5281/zenodo.17568963 (Sang, 2025), and https://github.com/sangshan-ss/CHANS_SD_YRB.

Author contributions

Conceptualization: YL, SS[1], BF. Data curation: SS[1]. Formal analysis: SS[1]. Funding acquisition: BF, YL. Investigation: YL, SS[1]. Methodology: SS[1], YL, BF. Project administration: YL. Resources: YL. Software: SS[1], SZ, LY. Supervision: YL, BF. Validation: SS[1], SZ, LY. Visualization: SS[1]. Writing – original draft: SS[1]. Writing – review and editing: SS[1], YL, SW, YXL, XTW, SS[2], WZ, XHW. (Note: SS[1] refers to Shan Sang, and SS[2] refers to Shuang Song, to distinguish between authors with identical initials.)

Competing interests

The authors declare that they have no conflict of interest.

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