

1 Global CO emissions and drivers of atmospheric CO trends constrained by 2 MOPITT satellite measurements

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18 19 20 **Abstract**

21 Carbon monoxide (CO), an important atmospheric pollutant produced by incomplete
22 combustion and hydrocarbon oxidation, significantly affects atmospheric oxidation capacity
23 and air quality. Accurate quantification of its global emissions and the underlying driver behind
24 its atmospheric trends is essential for understanding changes in global atmospheric
25 environment. Using 20 years (2003-2022) of data from the Measurement of Pollution in the
26 Troposphere (MOPITT) instrument, we analyze changes in global CO emissions and
27 atmospheric concentrations by applying a four-dimensional variational (4D-Var) assimilation
28 framework within the GEOS-Chem adjoint model. A posteriori simulations show good
29 agreement with independent surface and aircraft measurements compared to a priori
30 simulations. Sensitivity analyses further confirm that inferred emissions remain robust against
31 uncertainties associated with satellite vertical sensitivity and variations in hydroxyl radical (OH)
32 concentrations. Our results indicate a substantial decline in global anthropogenic CO emissions
33 of 14-17% (approximately 85-110 Tg yr⁻¹) over the two-decade period, largely driven by

34 emission reductions in the United States, Europe, and eastern China. Biomass burning
35 emissions exhibited strong interannual variability, with recent increases in Northern
36 Hemisphere high-latitude forests; in particular, the intense 2021 wildfires substantially offset
37 the anthropogenic emission-driven decline in atmospheric CO over the Northern Hemisphere.
38 This study provides a comprehensive assessment of global CO emissions and the mechanisms
39 governing atmospheric CO trends, offering a scientific basis for integrated policies addressing
40 both air pollution and climate change.

41

42 **1. Introduction**

43 Carbon monoxide (CO) is a key atmospheric pollutant produced from incomplete
44 combustion and the oxidation of hydrocarbons. As the main sink for the hydroxyl radical (OH),
45 CO critically influences the oxidative capacity of the atmosphere (Zhao et al., 2020; Tan et al.,
46 2022), and is an important precursor for tropospheric ozone (Whaley et al., 2015; Hu et al.,
47 2024). With a chemical lifetime of approximately one to two months, CO is frequently
48 employed as a valuable tracer for elucidating variations in anthropogenic activities and biomass
49 burning, providing critical insights into the long-range transport of atmospheric constituents
50 (Tang et al., 2019; Buchholz et al., 2022; Smoydzin and Hoor, 2022). By modulating the
51 abundance of OH, changes in CO concentrations indirectly affect the atmospheric lifetime of
52 methane (CH₄). Furthermore, CO shares common combustion sources with major greenhouse
53 gases like CH₄ and carbon dioxide (Worden et al., 2017; Zheng et al., 2023). Accurate
54 quantification of global CO emissions and a clear understanding of the drivers behind its
55 atmospheric trends are therefore essential for formulating effective policies to address the
56 challenges of air quality and climate change.

57 The advent of long-term satellite measurements has revolutionized our ability to monitor
58 global CO distributions (Warner et al., 2013; Worden et al., 2013; Hedelius et al., 2021),

59 enabling a shift from short-term, regional emission estimates (Arellano et al., 2004; Heald et
60 al., 2004; Kopacz et al., 2010) to analyses of decadal-scale changes. Numerous studies have
61 leveraged these records to report substantial declines in anthropogenic CO emissions (Fortems-
62 Cheiney et al., 2011; Jiang et al., 2017; Miyazaki et al., 2020), especially across the Northern
63 Hemisphere, contributing to improved air quality. However, a critical and emerging challenge
64 is to disentangle the competing influences on atmospheric CO concentrations. While
65 anthropogenic emissions are generally decreasing, biomass burning emissions exhibit strong
66 interannual variability. Thus, an important unanswered question is to what extent the recent
67 intensification of wildfires, particularly in high-latitude forests (Jain et al., 2024; Jones et al.,
68 2024), is offsetting the gains achieved from anthropogenic emission reductions. This has
69 profound implications, as a rise in wildfire CO signals a concurrent rise in wildfire greenhouse
70 gas emissions, which could offset part of the gains achieved from reductions in anthropogenic
71 greenhouse gas emissions.

72 Constraining global emissions and robustly attributing observed concentration trends
73 require the application of sophisticated inverse modeling approaches. These methods, which
74 include ensemble-based techniques (e.g., the ensemble Kalman filter) and variational methods
75 (e.g., four-dimensional variational, 4D-Var, data assimilation), provide powerful frameworks
76 for optimizing emission estimates by reconciling model simulations with satellite
77 measurements, while accounting for complex atmospheric transport and chemistry (Müller et
78 al., 2018; Miyazaki et al., 2020; Jiang et al., 2025). Among these, the 4D-Var data assimilation,
79 implemented within chemical transport models like GEOS-Chem and its adjoint (Henze et al.,
80 2007), has been widely and successfully applied to constrain CO emissions (Kopacz et al.,
81 2010; Jiang et al., 2015b; Tang et al., 2023), owing to its strengths in handling nonlinear
82 constraints and providing computationally efficient gradients. However, long-term multi-
83 decadal trend analyses based on this system has often been hindered by limitations such as

84 inconsistent meteorological inputs across years and the use of outdated a priori emission
85 inventories (Jiang et al., 2017; Qu et al., 2022).

86 To address these limitations, we employ a recent extension of the GEOS-Chem adjoint
87 model (Tang et al., 2023) that features support for consistent MERRA-2 meteorological data
88 and modern emission inventories via the Harmonized Emissions Component (HEMCO)
89 (Keller et al., 2014; Lin et al., 2021). By assimilating MOPITT (Measurements of Pollution in
90 the Troposphere) CO retrievals from 2003 to 2022, this study aims to provide an analysis with
91 the following specific objectives: (1) to quantify the long-term evolution of global CO
92 emissions; (2) to attribute the observed trends in atmospheric CO concentrations to changes in
93 emissions and meteorological variations, in particular, the effect of biomass burning emissions
94 on atmospheric CO decline driven by anthropogenic reductions; and (3) to evaluate the
95 sensitivity of inferred emissions to uncertainties in satellite vertical sensitivity and OH
96 concentrations. By doing so, this work aims to improve the understanding of key drivers behind
97 atmospheric CO changes and offer a refined emission inventory to support future air quality
98 and climate policies.

99 The paper is structured as follows: Section 2 describes the methodology, including the
100 assimilation framework, observational data, and the design of assimilation experiments.
101 Section 3 presents the results on the long-term emission trends, the robustness tests, and the
102 attribution of concentration changes. Conclusions are provided in Section 4.

103

104 **2. Methodology and Data**

105 **2.1 Assimilation framework**

106 We utilize the adjoint of the GEOS-Chem model (version 35n) with extended support for
107 MERRA-2 meteorological data and HEMCO emission inventories. The analysis is conducted
108 at a horizontal resolution of $2^\circ \times 2.5^\circ$ with 47 vertical levels (MERRA-2) up to 0.01 hPa and

109 employs a CO-only simulation (tagged-CO mode), in which the chemical sink of CO is
110 linearized with archived monthly mean OH fields. Two types of archived OH fields are used
111 in this study: fixed monthly OH fields for 2013 from the GEOS-Chem full chemistry simulation
112 (Fisher et al., 2017), and variable monthly OH fields for 2005-2020 from the Tropospheric
113 Chemistry Reanalysis version 2 (TCR-2, Miyazaki et al. (2020)). The TCR-2 OH fields have
114 been validated against various aircraft observations and show generally good agreement
115 (Miyazaki et al., 2020). Fig. S2 (see the SI) shows global mean tropospheric OH concentrations
116 from TCR-2, demonstrating a slight increasing trend ($1.0 \pm 0.6 \times 10^3 \text{ molec cm}^{-3} \text{ yr}^{-1}$) in 2005-
117 2020.

118 The global default anthropogenic emission inventory is the CEDS-CMIP6 (Community
119 Emissions Data System) (Hoesly et al., 2018). Regional emissions are replaced as follows:
120 MIX (Li et al., 2017) over Asia, NEI 2016 (National Emissions Inventory) over the United
121 States, DICE_AFRICA and EDGARv4.3 over Africa, and APEI over Canada. The contribution
122 of co-emitted anthropogenic VOC sources is considered by scaling up anthropogenic CO
123 emissions by 11%. Biogenic sources are simulated using the Model of Emissions of Gases and
124 Aerosols from Nature, version 2.0 (MEGANv2.0, Guenther et al. (2006)). CH₄ oxidation
125 source is considered by using a prescribed, spatially varying CH₄ field following the default
126 tagged-CO configuration (Fisher et al., 2017). The CO sources from both biogenic VOCs and
127 CH₄ oxidation are calculated online based on the assumption of instantaneous oxidation by OH
128 radicals. Biomass burning emissions are based on the Global Fire Emissions Database version
129 4 (GFED4, van der Werf et al. (2010)). For years beyond the end year of a specific inventory,
130 emissions from the last available year within that inventory's coverage were used to fill the
131 subsequent years. The distribution of the annual mean CO emissions from 2003 to 2022 is
132 shown in Figs. 1a-c.

133 The annual global sources are 536.3 Tg yr⁻¹ from anthropogenic emissions, 312.5 Tg yr⁻¹

134 from biomass burning, 623.2 Tg yr⁻¹ from the oxidation of biogenic VOCs, 922.5 Tg yr⁻¹ from
 135 the oxidation of CH₄, with a total sink (through the reaction with OH radicals) of approximately
 136 2395.0 Tg yr⁻¹ in a priori inventories in this work. For comparison, Zheng et al. (2019) reported
 137 inversion-based global CO budget estimates for 2005-2017 of approximately 700 Tg yr⁻¹ for
 138 anthropogenic emissions, 500 Tg yr⁻¹ for biomass burning, 300 Tg yr⁻¹ for biogenic VOC
 139 oxidation, and 900 Tg yr⁻¹ for CH₄ oxidation, with a total sink of approximately 2600 Tg yr⁻¹.
 140 Regarding anthropogenic emissions, the CEDS-CMIP6 inventory estimates an average of
 141 607.6 Tg yr⁻¹ for 2003-2014, while the updated CEDS-CMIP7 inventory yields lower values,
 142 averaging 480.2 Tg yr⁻¹ over 2003-2023. For biomass burning, the GFED5 inventory estimates
 143 an average of 518.3 Tg yr⁻¹ for 2003-2022.

144 The objective of the 4D-Var approach is to minimize the difference between simulations
 145 and observations by minimizing the cost function (Henze et al., 2007):

$$146 \quad J(\mathbf{x}) = \sum_{i=1}^N (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i)^T \mathbf{S}_\Sigma^{-1} (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i) + \gamma (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} (\mathbf{x} - \mathbf{x}_a) \quad (1)$$

147 where \mathbf{x} is the state vector of CO emissions, N is the number of observations distributed in
 148 time over the assimilation period, \mathbf{z}_i are the MOPITT CO retrievals, and $\mathbf{F}(\mathbf{x})$ is the forward
 149 model. Error estimates are assumed to be Gaussian: \mathbf{S}_Σ is the observational error covariance,
 150 which combines a 10% uniform error and the MOPITT CO retrieval error covariance; and \mathbf{S}_a
 151 is a priori error covariance. Here the combustion-related CO sources (fossil fuel, biofuel, and
 152 biomass burning) and the oxidation source from biogenic VOCs are combined, with a uniform
 153 a priori error of 50% assumed following previous studies (Jiang et al., 2013; Jiang et al., 2017).
 154 The CO source from CH₄ oxidation is optimized separately as an aggregated global source,
 155 with a priori uncertainty of 25%. The cost function is minimized by iteratively adjusting the
 156 CO emissions using the quasi-Newton gradient-based optimization L-BFGS-B algorithm (Zhu
 157 et al., 1997) and the adjoint gradients:

$$158 \quad \nabla_{\mathbf{x}} J(\mathbf{x}) = \sum_{k=1}^N \left[2 \left(\frac{\partial \mathbf{F}_i}{\partial \mathbf{x}} \right)^T \mathbf{S}_\Sigma^{-1} (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i) \right] + 2\gamma (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} \quad (2)$$

159 The LOGX2 method (Jiang et al., 2015a; Jiang et al., 2017) is employed to improve the
160 reduction of negative gradients.

161 Following Jiang et al. (2017), we applied a two-step approach to mitigate the influence of
162 systematic biases in the model simulations. First, a sequential Kalman filter (Todling and Cohn,
163 1994; Tang et al., 2022) was used to assimilate MOPITT CO retrievals from October 1, 2002,
164 to December 31, 2022, providing optimized CO concentration fields with lower bias. As
165 illustrated in Fig. 2a, the GEOS-Chem model driven by the original monthly CO initial
166 conditions and a priori emission inventories (referred to as GC-original) substantially
167 underestimated column CO concentrations by approximately 30–40% (mean bias = $-39.4 \times$
168 10^{16} molec cm^{-2} ; Table 1). In contrast, simulations using the monthly CO initial conditions
169 derived from the sequential Kalman filter together with a priori emissions (GC-a priori) showed
170 markedly improved agreement with MOPITT CO retrievals (Fig. 2b), reducing the mean bias
171 to about 10% (mean bias = -9.7×10^{16} molec cm^{-2}). Similarly, the use of optimized monthly
172 CO initial conditions led to considerable improvement in model performance against
173 independent surface and aircraft measurements (Table 1). The mean bias decreased from -18.3
174 ppb (GC-original) to -1.4 ppb (GC-a priori) for World Data Centre for Greenhouse Gases
175 (WDCGG) surface observations; from -18.9 ppb to -3.8 ppb for HIAPER Pole-to-Pole
176 Observations (HIPPO) aircraft data; and from -16.2 ppb to -3.4 ppb for Atmospheric
177 Tomography Mission (ATom) aircraft measurements. These results suggest that the substantial
178 negative biases seen in Fig. 2a largely originate from the accumulation of biases over preceding
179 months.

180 Furthermore, ocean scenes (pink grids in Fig. S3) were defined as land boundary
181 conditions. The optimized CO fields from the Kalman filter were used to update CO
182 concentrations over the ocean at hourly intervals during the forward simulation within the 4D-
183 Var process. Meanwhile, the 4D-Var system constrained CO emissions over land without

184 modifying oceanic CO distributions. As demonstrated by Jiang et al. (2017), the use of
185 optimized CO land boundary conditions in 4D-Var assimilation effectively reduces systematic
186 biases associated with long-range transport. By adopting this two-step assimilation framework,
187 the inversion focuses on optimizing fresh continental CO emissions, while reducing the
188 influence of uncertainties arising from transport and chemical processes, which tend to exhibit
189 larger systematic biases. Consequently, a posteriori CO emissions estimated in this study are
190 expected to be lower than those derived without adjustments to the initial and boundary CO
191 conditions. This reflects both the specific inverse modeling setup and a possible
192 underestimation in our a posteriori emission estimates, attributable to the emphasis on
193 constraining fresh continental CO sources.

194 Based on this assimilation framework, three sets of CO emission inversion experiments
195 are designed:

196 (1) Column-FixOH: uses MOPITT CO column concentration data with default OH fields
197 fixed in 2013.

198 (2) Profile-FixOH: uses MOPITT CO profile data with default OH fields fixed in 2013.

199 (3) Column-VarOH: uses MOPITT CO column concentration data with variable OH fields
200 from the TCR-2 tropospheric chemistry reanalysis.

201 By comparing the results of Column-FixOH and Profile-FixOH, the influence of different
202 MOPITT CO data types on CO source estimates can be assessed. Similarly, comparing
203 Column-FixOH and Column-VarOH allows for evaluation of the impact of different OH fields
204 on CO source estimates.

205 **2.2 MOPITT CO retrievals**

206 The MOPITT instrument was launched on December 18, 1999, aboard the NASA Terra
207 spacecraft. The satellite follows a sun-synchronous polar orbit at 705 km altitude, crossing the
208 equator at 10:30 local time. The instrument made measurements over a 612 km cross-track

209 scan, with a footprint of 22 km × 22 km. The MOPITT data used in this study are from the
 210 joint retrieval (version 9J) of CO, which combines thermal infrared (TIR, 4.7μm) and near-
 211 infrared (NIR, 2.3μm) radiances using an optimal estimation approach (Worden et al., 2010;
 212 Deeter et al., 2022). The retrieved volume mixing ratios are reported as layer averages across
 213 10 pressure levels (surface, 900, 800, 700, 600, 500, 400, 300, 200, and 100 hPa). The
 214 relationship between the retrieved CO profile and the true atmospheric state is expressed as:

$$215 \quad \hat{\mathbf{z}} = \mathbf{z}_a + \mathbf{A}(\mathbf{z} - \mathbf{z}_a) + \mathbf{G}\boldsymbol{\epsilon} \quad (3)$$

216 where \mathbf{z}_a is the MOPITT a priori CO profile, \mathbf{z} is the true atmospheric state, $\mathbf{G}\boldsymbol{\epsilon}$ represents
 217 the retrieval error, and $\mathbf{A} = \partial\hat{\mathbf{z}}/\partial\mathbf{z}$ is the MOPITT averaging kernel matrix, indicating the
 218 sensitivity of the retrieval to the actual atmospheric CO. We only consider data with Cloud
 219 Description = 2 (cloud free) and exclude MOPITT data with CO column amounts less than
 220 5×10^{17} molec cm⁻². The threshold (5×10^{17} molec cm⁻²) was selected to prevent the influence of
 221 certain potentially inaccurate, extremely low-concentration observations, which may also have
 222 low observation errors in the cost function, on the 4D-Var assimilation (Jiang et al., 2013; Jiang
 223 et al., 2017). Since the NIR channel relies on reflected solar radiation, only daytime data are
 224 considered (Worden et al., 2010; Tang et al., 2024).

225 **2.3 Aircraft and surface CO measurements**

226 The HIPPO (Wofsy and HIPPO Science Team (2011)) were conducted using the
 227 Gulfstream V aircraft from 2009 to 2011. The flights primarily covered the Pacific Ocean,
 228 spanning latitudes from 67°S to 87°N, with continuous sampling from 0.2 to 12 km altitude.
 229 The ATom (Wofsy and Atom Science Team (2018)) used the DC-8 aircraft from 2016 to 2018.
 230 ATom covered similar altitude and latitudinal ranges as HIPPO but with broader spatial
 231 coverage, particularly over the Atlantic Ocean. For HIPPO, a total of 687 CO profiles from
 232 five missions were used directly. For ATom, CO measurements during continuous ascents and
 233 descents were used to construct 523 CO profiles from four missions. Surface CO measurements

234 from the WDCGG are also included in this analysis. The WDCGG, operated by the Japan
235 Meteorological Agency under the World Meteorological Organization's Global Atmosphere
236 Watch (GAW) program, collects, archives, and distributes atmospheric greenhouse gas data,
237 including CO, contributed by various institutions worldwide.

238

239 **3. Results and Discussion**

240 **3.1 Evaluation of assimilation system performance**

241 Before presenting the estimated emission trends, we first evaluate the performance of our
242 assimilation system. The evaluation involves comparing modeled CO concentrations from the
243 GC-original, GC-a priori, and a posteriori simulations (Column-FixOH, Profile-FixOH,
244 Column-VarOH) over the period 2003-2022 against MOPITT satellite measurements, as well
245 as independent surface observations from WDCGG and aircraft measurements from HIPPO
246 and ATom. As summarized in Table 1, a posteriori simulations exhibit mean biases relative to
247 MOPITT retrievals ranging from -5.1 to -7.3×10^{16} molec cm⁻². These values are notably
248 smaller than the biases in the GC-a priori simulation (-9.7×10^{16} molec cm⁻²) and the GC-
249 original simulation (-39.4×10^{16} molec cm⁻²). Similarly, for the HIPPO aircraft observations,
250 a posteriori simulations show mean biases between -2.5 and -2.1 ppb, improved compared to
251 the GC-a priori (-3.8 ppb) and GC-original (-18.9 ppb) simulations. For ATom aircraft data, a
252 posteriori mean biases range from -2.9 to -1.6 ppb, also lower than those from the GC-a priori
253 (-3.4 ppb) and GC-original (-16.2 ppb) simulations. In the case of surface CO concentrations,
254 a posteriori simulations yield mean biases between 0.3 and 1.9 ppb relative to WDCGG
255 observations (Table 1), which are reduced compared to GC-original (-18.3 ppb) simulations,
256 and comparable with the GC-a priori (-1.4 ppb) simulations. A posteriori simulations slightly
257 overestimate surface concentrations relative to WDCGG data, while underestimating CO in the
258 free troposphere according to MOPITT and aircraft measurements. This systematic

259 discrepancy may be attributable to uncertainties in convective transport parameterizations
260 within the model.

261 Overall, the good consistency between a posteriori simulations and multiple independent
262 observation platforms demonstrates the capability of our assimilation system to effectively
263 constrain CO emissions. Given this confidence in the system's performance, we now present
264 the central findings of this study: the long-term evolution of CO emissions. As mentioned in
265 Section 2.1, the combustion-related CO sources and the oxidation source from biogenic VOCs
266 are combined, and thus, the inverse system optimizes total CO emissions within each model
267 grid cell. The subsequent attribution of emissions to specific source types (e.g., anthropogenic,
268 biomass burning) in an individual grid cell is based on the relative contribution of each source
269 category from a priori emission inventories. Specifically, a posteriori emission for a given
270 source type in a grid cell is calculated by applying the grid-scale scaling factor (the ratio of a
271 posteriori to a priori total emissions) to the corresponding a priori emission of that source type.
272 Different sources can finally be calculated because each source category possesses distinct
273 spatial patterns and seasonal variations.

274

275 **3.2 Long-term evolution of global CO emissions**

276 **3.2.1 Anthropogenic CO emissions**

277 At the global scale, anthropogenic CO emissions based on three inversion configurations
278 are estimated to be 7-14% higher than a priori values (Table 2) in 2003-2022 and show a clear
279 declining trend (Fig. 3f). Under the Column-FixOH configuration, global anthropogenic
280 emissions from 2003 to 2022 ranged from 546.1 to 654.1 Tg yr⁻¹, with a multi-year average of
281 approximately 610 Tg yr⁻¹ and a total reduction of about 17%; similar emission ranges and
282 reduction rates (14-17%) were obtained under the Profile-FixOH and Column-VarOH
283 configurations. These results are broadly consistent with Zheng et al. (2019). The

284 CEDS-CMIP7 inventory (Hoesly et al., 2018) shows significantly lower global CO emissions
285 than those derived from inverse modeling, though its decreasing trend is comparable. As shown
286 in Fig. 4a, negative trends (blue) were concentrated in three major industrialized regions:
287 eastern North America, Europe, and eastern China, forming a "reduction belt". These regions
288 accounted for over 65% of global anthropogenic CO emissions, and their systematic reductions
289 constituted the principal driver of the global downward trend. In contrast, positive trends (red)
290 were primarily distributed in northern India (increases of 15.2-22.3%) and Central Africa,
291 corresponding to rapid urbanization and industrialization processes.

292 In the United States (US), emissions declined rapidly from 2003 to 2009, followed by a
293 period of slower reduction (Fig. 3a). Over the entire period (2003-2022), US CO emissions
294 decreased at rates of 2.0-2.2 Tg yr⁻¹, resulting in a cumulative reduction of 46-49% (Table S1).
295 This phased reduction pattern is consistent with the diminishing marginal effects of widespread
296 transportation control technologies, as supported by independent studies (Elguindi et al., 2020;
297 Miyazaki et al., 2020). Our estimated emission magnitude and decreasing trend are similar to
298 Zheng et al. (2019) and the CEDS-CMIP7 in the US. European CO emissions (Fig. 3b)
299 followed a similar pattern (cumulative reduction of 32–34% over 2003-2022). Estimated
300 emissions over Europe in Zheng et al. (2019) are substantially higher than ours and show
301 stronger interannual variability. In comparison, the CEDS-CMIP7 inventory shows good
302 agreement with our results during 2003-2017, but a faster decline after 2017; and Fortems-
303 Cheiney et al. (2024) suggests continuous decline in CO emissions in Europe in 2011-2021.
304 This discrepancy could be possibly attributable to differences in the processing of initial and
305 boundary CO conditions (e.g., the use of climatological CO concentrations in Fortems-Cheiney
306 et al. (2024)).

307 The evolution of eastern China's CO emissions can be divided into four stages (Fig. 3c):
308 (1) a slight growth until 2007, peaking around that time; (2) a sharp decline of approximately

309 7% during the 2008 global financial crisis; (3) a temporary rebound from 2008 to 2010 under
310 economic stimulus policies; and (4) a continuous decline phase after 2010. From 2003 to 2022,
311 anthropogenic CO emissions from eastern China decreased at an average rate of 3.0-4.0 Tg yr⁻¹
312 (Table 2), with a cumulative reduction of 23-32% (Table S1). Zhao et al. (2012) and Xia et
313 al. (2016) confirmed the trend reversal around 2007, attributing it to improved energy
314 efficiency and strengthened emission controls, while Lin and McElroy (2011) and Tong et al.
315 (2016) highlighted the suppressive impact of the 2008 economic recession. Both Zheng et al.
316 (2019) and the CEDS-CMIP7 emission dataset show a declining trend consistent with our
317 results, although their emission magnitudes are lower. During 2019-2022, the emission
318 reduction rate accelerated to 4.8-8.3 Tg yr⁻¹, reflecting not only the short-term impact of the
319 COVID-19 pandemic but also the long-term cumulative effects of clean air policies and energy
320 structure transformation.

321 India exhibited a continuous growth in anthropogenic CO emissions from 2003 to 2009,
322 followed by a period of slower increase, with an average annual increase of 0.5-0.8 Tg yr⁻¹ in
323 2003-2022. Our estimated emission magnitude and trend are similar to Zheng et al. (2019) in
324 India. In comparison, the CEDS-CMIP7 inventory shows a similar trend, but its emission levels
325 are lower than those derived from inverse modeling. In Southeast Asia, anthropogenic CO
326 emissions exhibited a relatively stable and slow upward trend over the study period, though a
327 noticeable decline occurred from 2019 to 2022, which is likely associated with the impact of
328 the COVID-19 pandemic. The emission trend derived from our inversion is generally
329 consistent with that reported by Zheng et al. (2019) for this region, although their estimates
330 show stronger interannual variability. Compared with the CEDS-CMIP7 inventory, the trend
331 in CO emissions is similar to our results, but the emission magnitude in CEDS-CMIP7 is lower
332 than that derived from inverse modeling.

333 **3.2.2 Biomass burning CO emissions**

334 Globally, biomass burning CO emissions were 4-11% higher than a priori estimate in
335 2003-2022 (Table 3) and reached a historical peak of approximately 500 Tg yr⁻¹ in 2021 (Fig.
336 5g). In contrast to the clear decline of anthropogenic emissions, the trend in global biomass
337 burning CO emissions remains insignificant (Table 3). A comparison with the GFED5 emission
338 inventory (van der Werf et al., 2025) reveals noticeable differences: GFED5 estimates are
339 generally higher than our results (Fig. 5g), and do not show the 2021 peak. Spatial analysis
340 revealed a pronounced latitudinal differentiation in the changes of biomass burning CO
341 emissions (Figs. 4d-f): positive trends (red) were concentrated in Northern Hemisphere high-
342 latitude coniferous forests, while negative trends (blue) dominated tropical and subtropical
343 regions. This pattern is consistent with the "global fire emission geographic reconstruction"
344 observed by Zheng et al. (2023), reflecting the differential impacts of climate change across
345 latitudinal zones.

346 Emissions from high-latitude coniferous forests have shown different long-term trends
347 between boreal North America and boreal Asia over the past two decades (Figs. 5a-b). Peak
348 fire activity in boreal North America occurs during June-August (Fig. 6a); boreal Asia
349 experiences its primary fire season in June-August, with a secondary peak often observed in
350 March-May (Fig. 6b). When excluding the exceptional wildfire year of 2021, summertime
351 biomass burning CO emissions in boreal North America exhibited an overall declining trend
352 from 2003 to 2022 (Fig. 7a). In contrast, boreal Asia experienced a general increase in
353 summertime biomass burning CO emissions during the same period, even when 2021 is
354 omitted, though the trend is less pronounced than when including that extreme year (Fig. 7b).
355 The peak in wildfire emissions from high-latitude coniferous forests in 2021 was triggered by
356 severe, concurrent droughts across the Northern Hemisphere (Zheng et al., 2023). The
357 pronounced latitudinal amplification of emissions is consistent with higher carbon emission
358 density of boreal forests, which is 4-10 times greater than that of grasslands (Zheng et al., 2021).

359 GFED5 data suggests that boreal Asia's wildfire emissions peaked in 2012, different from
360 emission inversion results in this work and Zheng et al. (2023).

361 A notable decline in fire activity in South America occurred after 2010 (Fig. 5c),
362 particularly in August-September (Fig. 7c) coinciding with the peak wildfire season in South
363 America (Fig. 6c). The trend shift in CO emissions are consistent with the sharp reductions in
364 annual deforestation rates in the Brazilian Amazon from 25396 km² yr⁻¹ in 2003 to 7000 km²
365 yr⁻¹ in 2010 (Deeter et al., 2018). Africa experiences its primary fire season in June-September,
366 with a secondary peak often observed in December-February (Fig. 6d). Biomass burning CO
367 emissions in Africa exhibited a modest increasing trend overall (Fig. 5d), particularly in
368 February (Fig. 7c). Pronounced regional differentiation occurred, with increases in central
369 Africa and decreases in surrounding areas (Figs. 4d-f), reflecting the "strong contrast" pattern
370 described by Andela et al. (2017). Compared to the GFED5 inventory, our inversion results
371 generally show lower CO emission intensities.

372 Peak fire activity in Australia occurs during August-December (Fig. 6f); Southeast Asia
373 experiences its primary fire season in August-October, with a secondary peak often observed
374 in February-March (Fig. 6e). Emission patterns in Southeast Asia (Fig. 5e) and Australia (Fig.
375 5f) highlighted their sensitivity to large-scale climate oscillations. Major fire events in
376 Indonesia in 2006, 2009, 2015, and 2019 were closely linked to El Niño-induced droughts
377 (Page, 2009; Field et al., 2016). Australia's extreme fires in 2019 resulted from compound
378 extreme climate conditions influenced by the El Niño-Southern Oscillation, the Southern
379 Annular Mode, and the Indian Ocean Dipole (Deb et al., 2020). Building upon the observed
380 sensitivity to large-scale climate oscillations, the long-term interannual trend of wildfire
381 emissions across Southeast Asia and Australia remains insignificant, despite decline in August-
382 October in Southeast Asia (Fig. 7e) and September-November in Australia (Fig. 7f). Our

383 emission inversion shows lower CO emissions than GFED5 inventory in Southeast Asia and
384 Australia.

385 **3.2.3 Difference between combustion and biogenic NMVOC sources**

386 CO from combustion sources in the Northern Hemisphere showed strong regional
387 differentiation (Fig. 4), reflecting a dynamic redistribution between declining anthropogenic
388 sources and increasing biomass burning sources. Positive trends were densely distributed in
389 high-latitude regions, mainly due to increases in wildfires; Negative trends dominated mid-to-
390 low latitude industrialized areas. Tropical regions showed a mixed pattern, while the Southern
391 Hemisphere exhibited generally weaker trends. This spatial heterogeneity confirms a net global
392 decrease in combustion-related CO, revealing a clear contrast between increases at high
393 northern latitudes and decreases at mid-latitudes, reflecting the compound influences of
394 industrialization, policy interventions and climate change.

395 In contrast, CO produced from the oxidation of biogenic VOCs remained relatively stable
396 from 2003 to 2022 (Figs. 4g-i). This stability aligns with findings by Messina et al. (2016),
397 suggesting that global-scale biogenic VOC sources are less sensitive to short-term climate and
398 land cover changes. The global stability of biogenic VOC-derived CO is important for
399 atmospheric chemistry, as these compounds are key reactants for OH radicals and play a
400 regulatory role in atmospheric oxidation capacity. This stable background provides a crucial
401 baseline for understanding changes in atmospheric oxidation processes. The weaker trends
402 compared to those reported by Jiang et al. (2017) may be associated with our use of continuous
403 MERRA-2 meteorological data, which enhances consistency in long-term analysis.

404

405 **3.3 Long-term evolution and drivers of global CO concentrations**

406 Building on the emission estimates evaluated above, this section investigates their ultimate
407 influence in the atmosphere by analyzing the spatiotemporal patterns and trends of CO

408 concentrations. We first present the mean state and long-term changes in CO concentrations,
409 and then quantitatively attribute these changes to their underlying drivers: emissions and
410 meteorology. Figs. 8a-c show the mean surface CO concentrations in 2003-2022 from a
411 posteriori simulations and WDCGG surface observations. Higher CO concentrations are
412 evident in regions with strong anthropogenic emissions, such as East Asia, India, and Southeast
413 Asia, as well as in areas with significant biomass burning, i.e., Central Africa and South
414 America. The long-term trends in surface CO (Figs. 8d-f) reveal declining concentrations over
415 North America, Europe, East Asia, and South America, which contrast with rising trends over
416 India, Boreal Asia, Central Africa, and Australia. The 20-year mean CO columns (Figs. 9a-c)
417 show a consistent spatial pattern, with the highest column concentrations over East Asia and
418 Central Africa, followed by South America, India, and Southeast Asia. In contrast, the long-
419 term trend of CO columns (Figs. 9d-f) exhibits a more uniform decrease across the Northern
420 Hemisphere, lacking the distinct regional hotspots observed in the surface trends. This suggests
421 that changes in CO are more thoroughly mixed within the column.

422 To quantitatively attribute the concentration trends to specific drivers, we conducted a
423 series of sensitivity experiments. The experimental design isolates the influence of individual
424 emission sectors by building a baseline scenario in which all emissions are fixed at 2003 levels
425 to reflect the impact of meteorological condition changes in 2003-2022. Three more sensitivity
426 experiments were then conducted in 2003-2022 in which only one emission category, i.e.,
427 anthropogenic, biomass burning, or biogenic VOC sources, was allowed to vary over time,
428 respectively. The time-varying sources in these sensitivity experiments were prescribed from
429 the Column-FixOH a posteriori inversion.

430 The results indicate that meteorological influences induced positive trends in surface CO
431 concentrations in regions such as central Africa, Southeast Asia, and the Tibetan Plateau (0.6-
432 1.8% yr⁻¹), along with slight negative trends in areas such as South America (Fig. 10a). The

433 meteorological impact on CO column concentrations was comparatively weaker (Fig. 10b),
434 showing positive trends of $0.45\% \text{ yr}^{-1}$ over central Africa and the Tibetan Plateau. This vertical
435 differentiation implies that meteorological influences may primarily alter the vertical
436 distribution of CO through changes in convective transport, with a more limited effect on larger
437 horizontal scales. The derived meteorological impact is noticeably weaker than that reported
438 by Jiang et al. (2017), a discrepancy likely attributable to our use of consistent MERRA-2
439 meteorological fields, which enhances the reliability of the long-term trend analysis. Similarly,
440 the impact of biogenic VOC changes on CO concentrations (Figs. 10g, 10h) was markedly
441 weaker than in Jiang et al. (2017).

442 Anthropogenic emission changes were identified as the principal driver behind declining
443 CO levels, inducing strong negative trends in industrial regions of the Northern Hemisphere,
444 such as eastern North America, Europe, and eastern China. This signal is consistent across both
445 surface and column concentrations (Figs. 10c-d). Globally, anthropogenic emission changes
446 led to an average annual decrease of $0.27\% \text{ yr}^{-1}$ in CO column concentrations, with a more
447 pronounced decline rate of $0.51\% \text{ yr}^{-1}$ in the Northern Hemisphere (Table 4). Regionally, the
448 US, Europe, and eastern China exhibited the most substantial decreases, at $-0.57\% \text{ yr}^{-1}$, -0.69%
449 yr^{-1} and $-0.69\% \text{ yr}^{-1}$, respectively. In contrast, India experienced a slight concentration increase
450 ($0.03\% \text{ yr}^{-1}$) due to rising emissions, while Southeast Asia showed a more moderate decline ($-$
451 $0.19\% \text{ yr}^{-1}$) compared to other major industrial regions.

452 Conversely, changes in biomass burning emissions generally contributed to positive trends
453 in CO, particularly at high latitudes (Figs. 10e-f). At global and Northern Hemispheric scales,
454 this positive trend was largely attributable to extreme wildfire activity in 2021. When 2021 is
455 excluded, the long-term trend in CO columns due to biomass burning becomes statistically
456 insignificant at these broad scales (Figs. 7g-h), with only regionally and seasonally confined
457 increases remaining apparent, notably over Boreal Asia (July-August, Fig. 7b) and Africa

458 (January-April, Fig. 7d). In the full record (including 2021), biomass burning emissions led to
459 an average annual increase of $0.10\% \text{ yr}^{-1}$ in global CO columns, and $0.24\% \text{ yr}^{-1}$ in the Northern
460 Hemisphere (Table 4). It is noteworthy that the CO concentration response lagged behind
461 emission pulses by about one month and persisted longer. In the Northern Hemisphere, for
462 instance, enhanced emissions occurred mainly from July to September, whereas the significant
463 concentration response extended from August to December (Fig. 7g). This lag and prolonged
464 influence were primarily attributable to the delayed response over Boreal North America (Fig.
465 7a). At the regional scale, increases occurred in Boreal North America ($0.43\% \text{ yr}^{-1}$) and Boreal
466 Asia ($0.48\% \text{ yr}^{-1}$). In contrast, South America, Australia, and Southeast Asia experienced
467 declining trends ranging from $-0.13\% \text{ yr}^{-1}$ to $-0.22\% \text{ yr}^{-1}$, while Africa showed a slight increase
468 of $0.09\% \text{ yr}^{-1}$.

469 This attribution analysis highlights the substantial impact of extreme wildfire years on the
470 CO budget. Although anthropogenic emission reductions lowered Northern Hemisphere CO
471 columns by approximately $0.51\% \text{ yr}^{-1}$, the intense biomass burning emissions in 2021
472 introduced a positive perturbation of about $0.24\% \text{ yr}^{-1}$ in the full-record trend, thereby offsetting
473 a considerable fraction of the anthropogenic-driven decline. As a result, the net concentration
474 decline was reduced to approximately $0.27\% \text{ yr}^{-1}$ in the analysis including 2021. This implies
475 that nearly half (47%) of the potential air quality improvement from anthropogenic emission
476 controls can be offset by wildfire emissions. This finding provides a clear mechanistic
477 explanation for the decline in atmospheric CO concentrations in recent years, and underscores
478 the growing role of extreme wildfire events in modulating regional to hemispheric air
479 composition.

480

481 **3.4 Impacts of systematic errors on inferred CO emissions**

482 The MOPITT instrument provides retrievals for both CO total column and vertical profile.

483 The degrees of freedom for signal (DFS) for MOPITT multi-spectral profile retrievals
484 (TIR+NIR) is approximately 1.5-2.0 over land, reducing to about 1.0 when converted to a total
485 column (Worden et al., 2010). The discrepancy between a posteriori emission estimates
486 constrained by CO column (Column-FixOH) and profile (Profile-FixOH) data helps evaluate
487 the influence of systematic errors associated with the vertical sensitivity of the satellite
488 retrievals (Tang et al., 2024). Globally, a posteriori anthropogenic and biomass burning CO
489 emissions from Profile-FixOH were both slightly lower than those from Column-FixOH, with
490 average differences of -6.6% and -5.5%, respectively, over the period 2003-2022 (Table 2).
491 The two configurations also showed broadly consistent long-term trends in inferred
492 anthropogenic emissions, both indicating a global decline of approximately $-0.9\% \text{ yr}^{-1}$. Larger
493 regional discrepancies were observed over eastern China ($-2.1\% \text{ yr}^{-1}$ for Profile-FixOH vs. -
494 $1.6\% \text{ yr}^{-1}$ for Column-FixOH) and India ($1.1\% \text{ yr}^{-1}$ vs. $0.7\% \text{ yr}^{-1}$). Similarly, the trends in global
495 biomass burning CO emissions were consistent ($0.3\% \text{ yr}^{-1}$ for Column-FixOH and $0.5\% \text{ yr}^{-1}$
496 for Profile-FixOH), though regional differences were more pronounced for boreal North
497 America ($3.1\% \text{ yr}^{-1}$ vs. $4.9\% \text{ yr}^{-1}$) and Australia ($-1.5\% \text{ yr}^{-1}$ vs. $-0.7\% \text{ yr}^{-1}$). The limited
498 differences in inferred emissions between the two configurations resulted in a consistent
499 declining trend in simulated CO columns ($-0.5\% \text{ yr}^{-1}$ for both).

500 OH concentrations in model simulations significantly influence the inverse analysis of CO
501 emissions (Jiang et al., 2011; Müller et al., 2018). By assimilating MOPITT CO column data,
502 we compared the inverted CO emission estimates driven by fixed (Column-FixOH) and
503 variable (Column-VarOH) OH fields to investigate the potential influence. As shown in Fig.
504 11c, OH concentrations from the TCR-2 reanalysis are broadly 10-40% lower than the fixed
505 climatological OH concentrations over land (differences over the ocean are not considered here
506 due to the use of CO land boundary conditions in the 4D-Var assimilation). Lower OH
507 concentrations over land lead to reduced chemical loss, which is compensated by lower global

508 anthropogenic CO emissions in Column-VarOH inversion (590.1 Tg yr⁻¹) in 2003-2022,
509 approximately 3.7% lower than in Column-FixOH (612.8 Tg yr⁻¹).

510 Variations in OH concentrations influence the oxidation of biogenic VOCs to CO and their
511 subsequent chemical loss. These two counteracting processes establish a complex balance,
512 ultimately reflected in the inverted estimates of biogenic CO sources. Specifically, the Column-
513 VarOH inversion yields an average global biogenic CO sources of 391.4 Tg yr⁻¹ in 2003-2022,
514 approximately 3.9% lower than the 407.6 Tg yr⁻¹ in Column-FixOH inversion. The sensitivity
515 experiments described above address the third objective of this study, which is to evaluate the
516 robustness of our central findings against potential systematic errors associated with satellite
517 retrieval vertical sensitivity and OH concentrations. This robustness can be attributed, in part,
518 to our two-step inversion framework, which mitigates systematic biases through optimized
519 initial and boundary CO conditions.

520

521 **4. Conclusions**

522 This study provides a comprehensive, quantitative analysis of global CO emissions and
523 drivers governing atmospheric CO trends over the past two decades (2003-2022). By
524 employing a 4D-Var assimilation framework within GEOS-Chem adjoint model, constrained
525 by long-term MOPITT satellite measurements, we have generated an observationally
526 constrained CO emission inventory. A central methodological strength lies in the use of
527 continuous MERRA-2 meteorological fields and modern a priori emission inventories, which
528 significantly enhanced the long-term consistency and reliability of our trend analysis. The
529 implementation of a two-step bias mitigation strategy, optimizing both initial conditions and
530 land boundary conditions for CO, effectively reduced the accumulated impacts of transport and
531 chemistry uncertainties. The optimized emissions yield simulated CO concentrations that show
532 good agreement with independent surface measurements from the WDCGG network and

533 aircraft-based profiles from the HIPPO and ATom campaigns. The mean bias in simulated CO
534 concentrations (model minus observation) was reduced from -3.8 ppb in a priori simulation to
535 between -2.5 and -2.1 ppb in a posteriori simulation for HIPPO, and from -3.4 ppb in a priori
536 simulation to between -2.9 and -1.6 ppb in a posteriori simulation for ATOM.

537 Our results demonstrate a significant 14-17% decline (approximately 85-110 Tg yr⁻¹) in
538 global anthropogenic CO emissions over the 20-year period. This reduction was predominantly
539 driven by pollution control policies in major industrialized regions, with cumulative reductions
540 of 46-49% in the US, 32-34% in Europe, and 23-32% in eastern China. The decline in
541 anthropogenic CO emissions is consistent with the trends reported in the CEDS-CMIP7
542 inventory and the inversion results of Zheng et al. (2019), and is identified as the dominant and
543 statistically significant driver behind the observed decrease in atmospheric CO concentrations.
544 For biomass burning, our emission estimates suggested the historical peak of approximately
545 500 Tg yr⁻¹ in 2021, while the overall CO emissions in GFED5 inventory are higher than our
546 estimates in 2003-2022. Biomass burning emissions exhibited strong interannual variability
547 without a statistically significant long-term trend at the global and Northern Hemispheric scales,
548 although regionally and seasonally trends, such as in the boreal Northern Hemisphere during
549 summer, were evident in certain periods.

550 A central finding of this work is the substantial impact of extreme wildfire events,
551 particularly the record-breaking 2021 burning season in Northern Hemisphere high latitudes.
552 Our attribution analysis reveals that these wildfires introduced a strong positive perturbation to
553 atmospheric CO, offsetting nearly half (47%) of the concentration decline driven by
554 anthropogenic reductions in the Northern Hemisphere over our study period (2003-2022). This
555 finding highlights that while not a persistent trend, extreme wildfire events can counteract a
556 large fraction of the gains achieved from decades of emission control efforts. Our analysis thus
557 clarifies the past evolution of global CO emissions and concentrations, highlighting an

558 increasingly critical challenge: climate change is amplifying the intensity and impact of
559 extreme wildfire events, which can periodically undermine emission control efforts. This
560 underscores the need for integrated policies that address both anthropogenic sources and the
561 climate-driven amplification of natural emissions.

562

563 **Code and data availability:** The MOPITT CO data can be downloaded from
564 <https://asdc.larc.nasa.gov/data/MOPITT/>. The WDCGG data can be downloaded from
565 <https://gaw.kishou.go.jp>. The HIPPO data can be downloaded from
566 <https://data.eol.ucar.edu/dataset/112.123>. The ATom data can be downloaded from
567 <https://www.earthdata.nasa.gov/data/catalog/ornl-cloud-atom-merge-v2-1925-2.0>. The adjoint
568 of GEOS-Chem model can be downloaded from [http://wiki.seas.harvard.edu/geos-](http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_Adjoint)
569 [chem/index.php/GEOS-Chem_Adjoint](http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_Adjoint). The GFED5 data can be downloaded from
570 <https://www.globalfiredata.org/data.html>. The CEDS data can be downloaded from
571 <https://geos-chem.s3.amazonaws.com/index.html#HEMCO/CEDS/>. The emission data from
572 Zheng et al. (2019) can be downloaded from
573 [https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454](https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454453/1)
574 [453/1](https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454453/1). A posteriori CO emission estimates (Column-FixOH, Profile-FixOH and Column-
575 VarOH) derived in this work can be downloaded from
576 <https://doi.org/10.5281/zenodo.17221834>.

577

578 **Author Contributions:** Z.J. designed the research. Z.T. developed the model code and
579 performed the research. Z.J. and Z.T. wrote the manuscript. All authors contributed to
580 discussions and editing the manuscript.

581

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583

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588

589 **Tables and Figures**

590 **Table 1.** Mean biases of modeled CO concentrations relative to satellite (MOPITT) and in-situ
591 (WDCGG, HIPPO, ATom) observations for five model simulations over the period 2003-2022.
592 Biases are in units of 10^{16} molec cm^{-2} for MOPITT and ppb for other datasets.

593

594 **Table 2.** Mean anthropogenic CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-2022:
595 A comparison of a priori inventories with those constrained by MOPITT retrievals under
596 different configurations. The region definition is shown in Figure S1e.

597

598 **Table 3.** Mean biomass burning CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-
599 2022: A comparison of a priori inventories (GFED4) with those constrained by MOPITT
600 retrievals under different configurations. The region definition is shown in Figure S1f.

601

602 **Table 4.** Attribution of trends in column CO concentrations ($\% \text{ yr}^{-1}$) from 2003 to 2022 to
603 changes in anthropogenic and biomass burning emissions, based on sensitivity simulations
604 using the Column-FixOH inversion results.

605

606 **Fig. 1.** Spatial patterns of a priori CO sources and a posteriori scaling factors for the period
607 2003-2022. (a-c) Mean a priori CO sources (10^{12} molec $\text{cm}^{-2} \text{ s}^{-1}$). (d-l) Scaling factors (ratio of
608 a posteriori to a priori sources) derived from the three inversion experiments: (d-f) Column-
609 FixOH, (g-i) Profile-FixOH, and (j-l) Column-VarOH.

610

611 **Fig. 2.** Relative bias in column CO for 2003-2022, calculated as $(\text{Model} - \text{MOPITT}) / \text{MOPITT}$
612 for GC-original (a), GC-a priori (b), and a posteriori simulations (c-e).

613

614 **Fig. 3.** Time series of anthropogenic CO emissions from 2003 to 2022 across major regions,
615 comparing a priori inventories, a posteriori inversions from this study, and independent
616 estimates from CEDS-CMIP6/7 and Zheng et al. (2019).

617

618 **Fig. 4.** Long-term trends in CO sources (10^{10} molec cm^{-2} s^{-1} yr^{-1}) as constrained by the three
619 inversion experiments (2003-2022). For anthropogenic and biogenic VOC trends, months
620 dominated by biomass burning (>50% contribution) were excluded in this figure. This
621 approach ensures that the derived trends more accurately reflect the actual changes in
622 anthropogenic and biogenic VOC sources, without being biased by short-term, seasonal
623 biomass burning signals.

624

625 **Fig. 5.** Time series of biomass burning CO emissions from 2003 to 2022, comparing the a priori
626 inventory (GFED4), a posteriori inversions, and the GFED5 inventory.

627

628 **Fig. 6.** Climatological monthly cycle (2003-2022 average) of biomass burning CO emissions
629 across different regions, comparing a priori (GFED4) and a posteriori estimates with GFED5.

630

631 **Fig. 7.** Monthly trends in biomass burning CO emissions (based on Column-FixOH) and their
632 impact on column CO concentrations (2003-2022). Solid lines show trends including 2021,
633 while dashed lines exclude 2021 to illustrate the impact of extreme fire year. Please check Fig.
634 S4 and S5 for the standard deviation of the trends.

635

636 **Fig. 8.** Modeled surface CO concentrations (ppb) and their trends ($\% \text{yr}^{-1}$) from 2003 to 2022.
637 (a-c) Mean concentrations from WDCGG observations and model simulations. (d-f) Spatial
638 pattern of the long-term trend. Only stations with 14 year observations (the time range between
639 the first and last observations) during 2003-2022 are included.

640

641 **Fig. 9.** Modeled column CO concentrations (10^{18} molec cm^{-2}) and their trends ($\% \text{yr}^{-1}$) from
642 2003 to 2022. (a-c) Spatial distribution of the 20-year mean. (d-f) Spatial pattern of the long-
643 term trend.

644

645 **Fig. 10.** Attribution of trends ($\% \text{yr}^{-1}$) in surface and column CO concentrations to individual
646 drivers, derived from sensitivity simulations based on Column-FixOH inversion (2003-2022).
647 Trends are shown for scenarios with: (a, b) all sources fixed at 2003 levels; (c, d) only

648 anthropogenic emissions varying over time; (e, f) only biomass burning emissions varying; (g,
649 h) only biogenic VOC sources varying.

650

651 **Fig. 11.** Sensitivity of inverted CO emissions to OH fields. (a-c) Tropospheric OH columns
652 (10^{12} molec cm^{-2}) from fixed and variable (TCR-2) fields and their difference. (d-f)
653 Corresponding scaling factors from the Column-FixOH and Column-VarOH inversions and
654 their difference. Please note that due to the use of land boundary conditions, differences in OH
655 concentrations over the ocean in the left column figures have a negligible effect on the
656 differences in scaling factors shown in the right column figures.

657

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Observations	GC-original	GC-a priori	Column-FixOH	Profile-FixOH	Column-VarOH
MOPITT	-39.4	-9.7	-7.3	-6.8	-5.1
WDGCG	-18.3	-1.4	1.8	1.9	0.3
HIPPO	-18.9	-3.8	-2.5	-2.1	-2.2
ATOM	-16.2	-3.4	-2.1	-2.9	-1.6

Table 1. Mean biases of modeled CO concentrations relative to satellite (MOPITT) and in-situ (WDCGG, HIPPO, ATom) observations for five model simulations over the period 2003-2022. Biases are in units of 10^{16} molec cm^{-2} for MOPITT and ppb for other datasets.

Anthropogenic emissions		US	Europe	E. China	India	SE. Asia	Global
GC-a priori	Emissions	53.5	33.7	179.6	72.7	26.1	536.3
	Trends	-1.8 ± 0.1	-0.7 ± 0.1	-1.4 ± 0.3	0.3 ± 0.1	0.1 ± 0.0	-3.6 ± 0.4
Column-FixOH	Emissions	61.6	39.2	195.8	73.3	26.9	612.8
	Trends	-2.2 ± 0.3	-0.9 ± 0.2	-4.0 ± 0.8	0.8 ± 0.3	0.0 ± 0.1	-5.6 ± 0.9
Profile-FixOH	Emissions	56.7	36.3	188.3	69.9	26.8	572.2
	Trends	-2.0 ± 0.3	-0.8 ± 0.2	-3.0 ± 0.4	0.5 ± 0.2	0.1 ± 0.1	-5.0 ± 0.4
Column-VarOH	Emissions	58.2	38.5	186.7	67.7	26.3	590.1
	Trends	-2.0 ± 0.3	-0.9 ± 0.2	-3.7 ± 0.8	0.7 ± 0.2	-0.0 ± 0.1	-5.3 ± 0.8

Table 2. Mean anthropogenic CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-2022: A comparison of a priori inventories with those constrained by MOPITT retrievals under different configurations. The region definition is shown in Figure S1e.

Biomass burning		Boreal N. America	Boreal Asia	S. America	Africa	SE. Asia	Australia	Global
GC-a priori	Emissions	21.5	39.5	44.0	136.9	25.0	12.0	312.5
	Trends	1.1 ± 1.1	1.3 ± 1.8	-0.6 ± 1.7	-1.4 ± 0.5	-0.8 ± 1.6	-0.0 ± 0.6	-0.6 ± 3.2
Column-FixOH	Emissions	20.4	45.4	37.4	167.5	18.0	15.1	345.6
	Trends	0.6 ± 1.3	1.6 ± 1.9	-0.9 ± 1.4	0.5 ± 0.8	-0.5 ± 0.8	-0.2 ± 0.6	0.9 ± 3.4
Profile-FixOH	Emissions	21.2	46.7	38.4	148.8	19.7	13.7	326.6
	Trends	1.1 ± 1.5	2.1 ± 2.7	-1.1 ± 1.4	0.3 ± 0.8	-0.6 ± 1.1	-0.1 ± 0.6	1.7 ± 4.4
Column-VarOH	Emissions	18.0	41.0	33.3	159.5	19.0	15.8	325.7
	Trends	0.3 ± 0.8	1.1 ± 1.6	-1.0 ± 1.2	0.1 ± 0.8	-0.1 ± 0.8	0.1 ± 0.5	-0.1 ± 2.4

Table 3. Mean biomass burning CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-2022: A comparison of a priori inventories (GFED4) with those constrained by MOPITT retrievals under different configurations. The region definition is shown in Figure S1f.

Region	Anthropogenic	Region	Biomass burning
United States	-0.57 ± 0.11	Boreal N. America	0.43 ± 0.49
Europe	-0.69 ± 0.09	Boreal Asia	0.48 ± 0.53
Eastern China	-0.69 ± 0.22	South America	-0.13 ± 0.27
India	0.03 ± 0.16	Africa	0.09 ± 0.19
Southeast Asia	-0.19 ± 0.18	Southeast Asia	-0.22 ± 0.41
\	\	Australia	-0.12 ± 0.22
Northern Hemisphere	-0.51 ± 0.12	Northern Hemisphere	0.24 ± 0.33
Global	-0.27 ± 0.11	Global	0.10 ± 0.25

Table 4. Attribution of trends in column CO concentrations ($\% \text{ yr}^{-1}$) from 2003 to 2022 to changes in anthropogenic and biomass burning emissions, based on sensitivity simulations using the Column-FixOH inversion results.

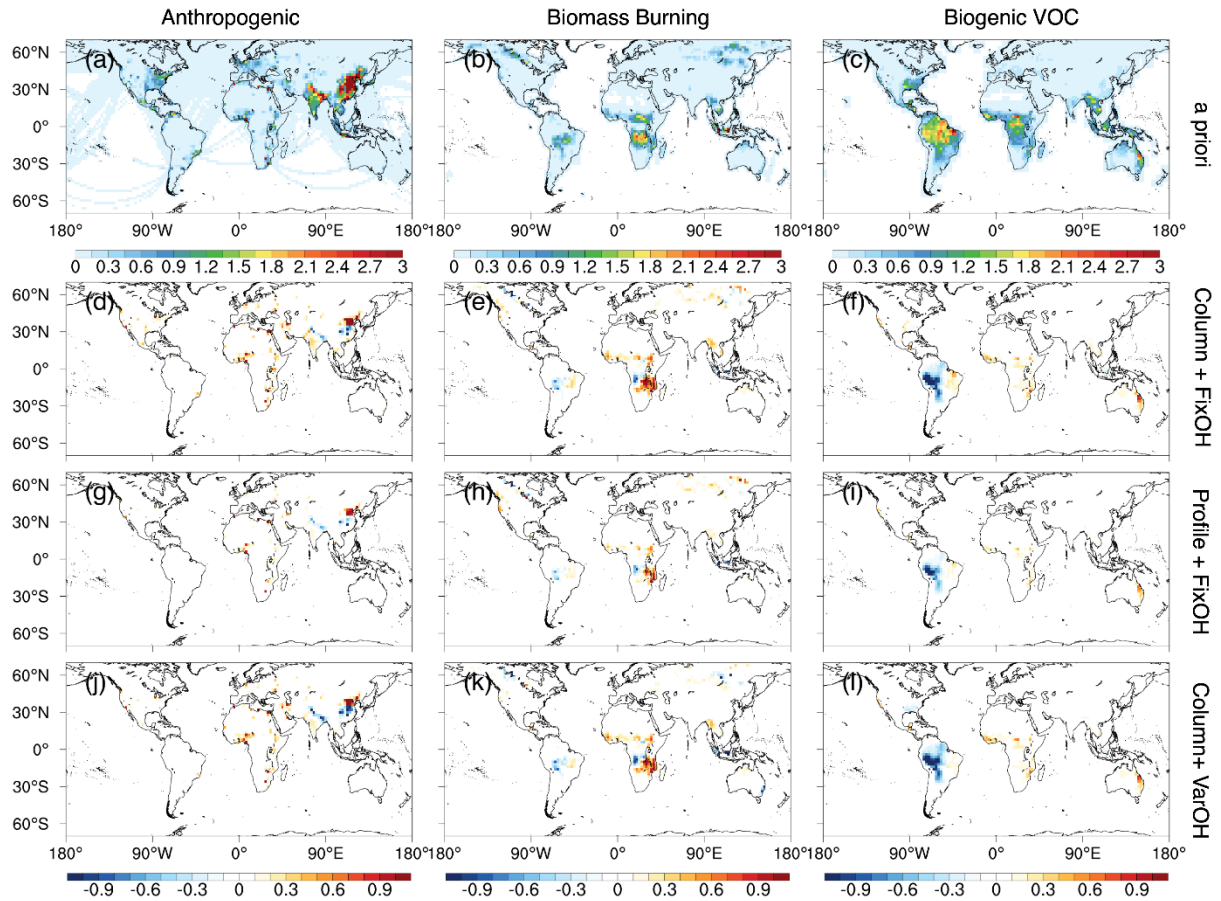


Fig. 1. Spatial patterns of a priori CO sources and a posteriori scaling factors for the period 2003-2022. (a-c) Mean a priori CO sources ($10^{12} \text{ molec cm}^{-2} \text{ s}^{-1}$). (d-l) Scaling factors (ratio of a posteriori to a priori sources) derived from the three inversion experiments: (d-f) Column-FixOH, (g-i) Profile-FixOH, and (j-l) Column-VarOH.

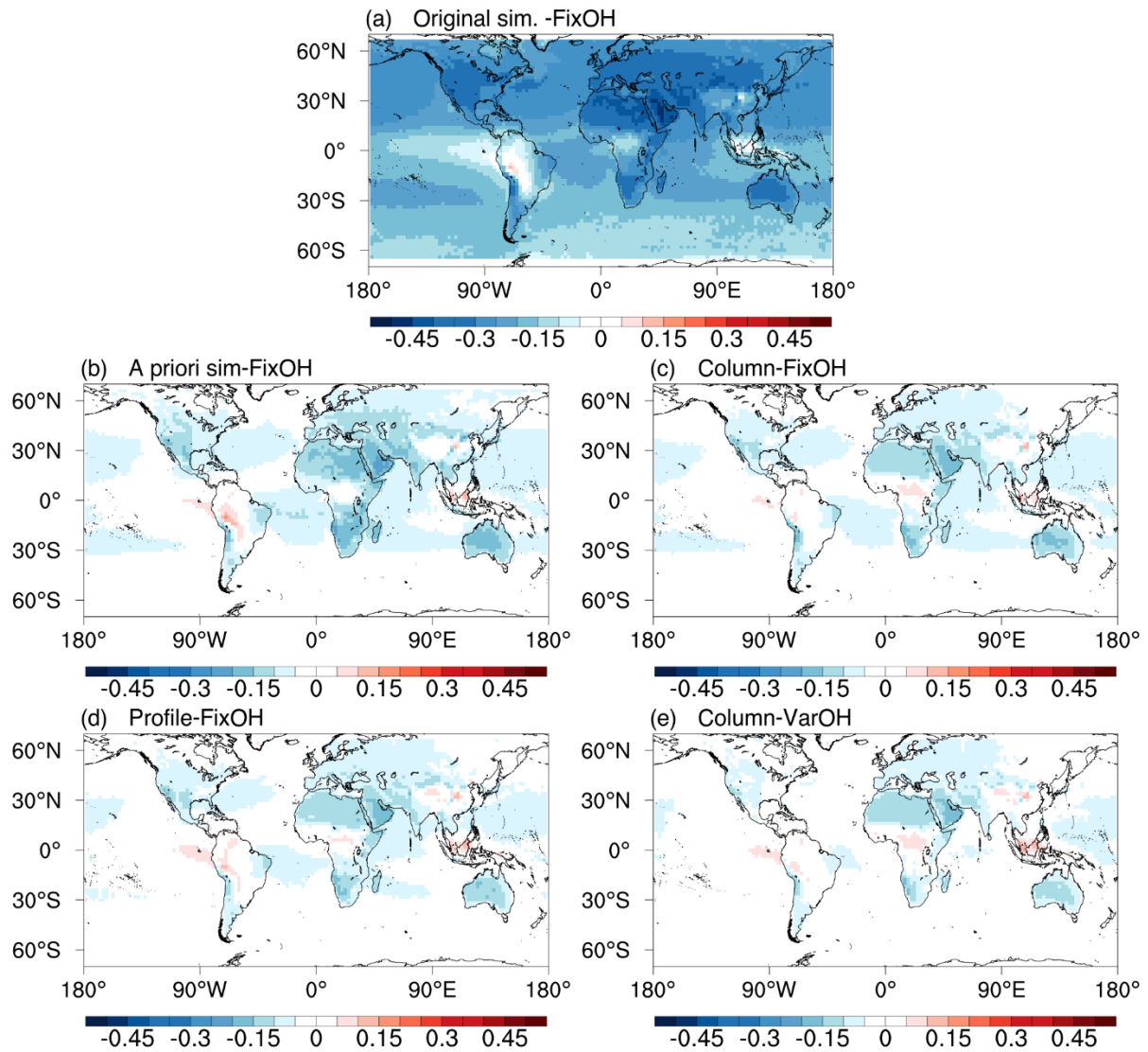


Fig. 2. Relative bias in column CO for 2003-2022, calculated as $(\text{Model} - \text{MOPITT}) / \text{MOPITT}$ for GC-original (a), GC-a priori (b), and a posteriori simulations (c-e).

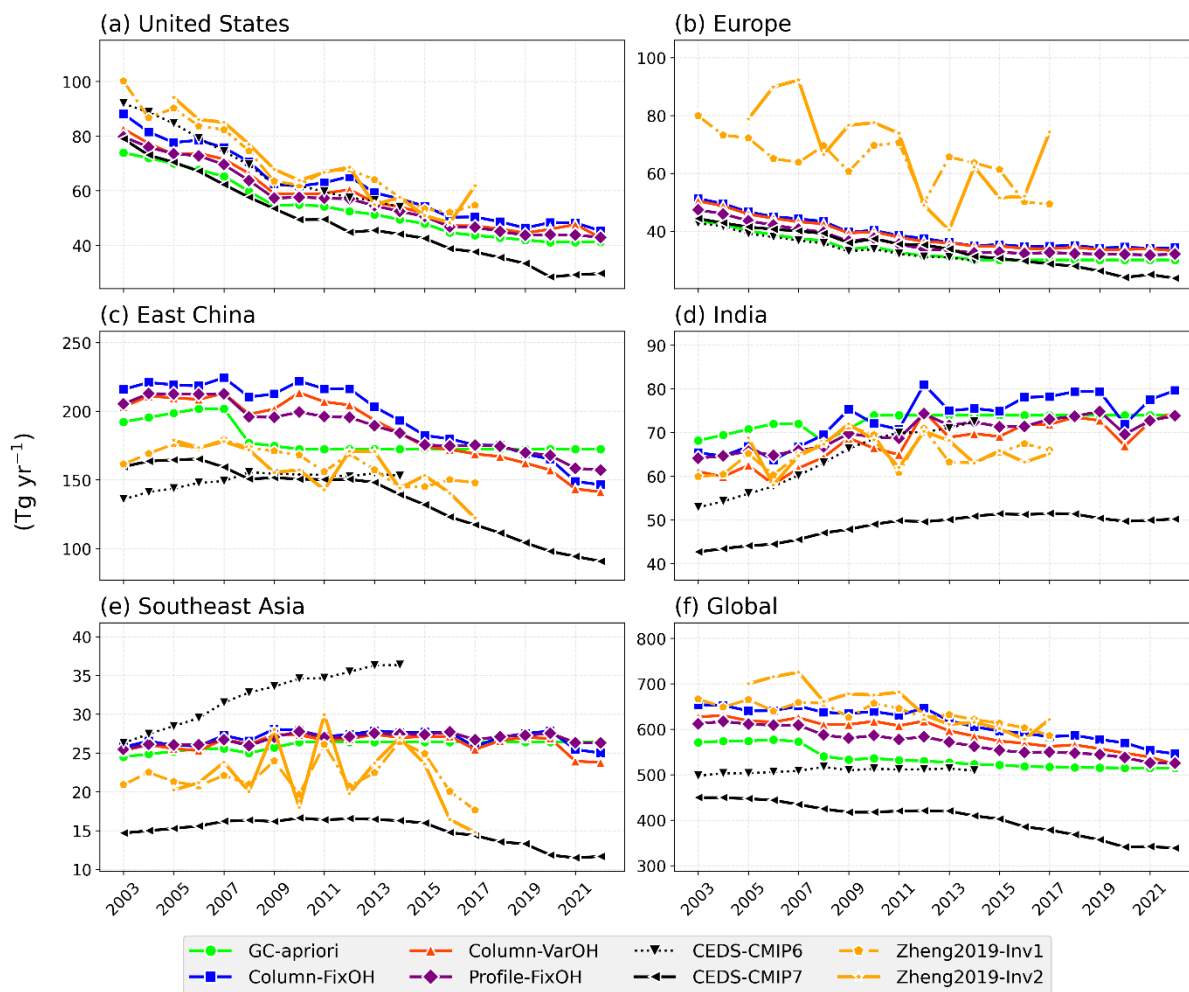


Fig. 3. Time series of anthropogenic CO emissions from 2003 to 2022 across major regions, comparing a priori inventories, a posteriori inversions from this study, and independent estimates from CEDS-CMIP6/7 and Zheng et al. (2019).

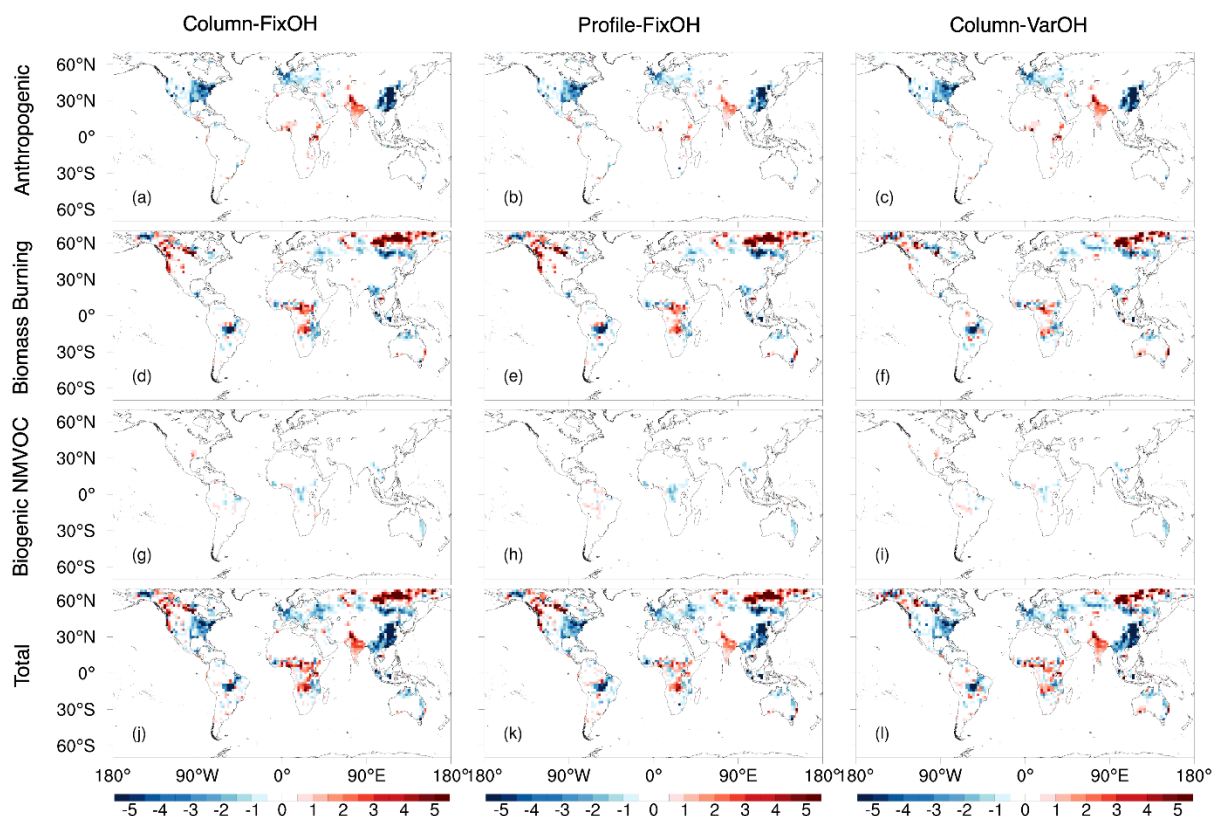


Fig. 4. Long-term trends in CO sources (10^{10} molec cm^{-2} s^{-1} yr^{-1}) as constrained by the three inversion experiments (2003-2022). For anthropogenic and biogenic VOC trends, months dominated by biomass burning (>50% contribution) were excluded in this figure. This approach ensures that the derived trends more accurately reflect the actual changes in anthropogenic and biogenic VOC sources, without being biased by short-term, seasonal biomass burning signals.

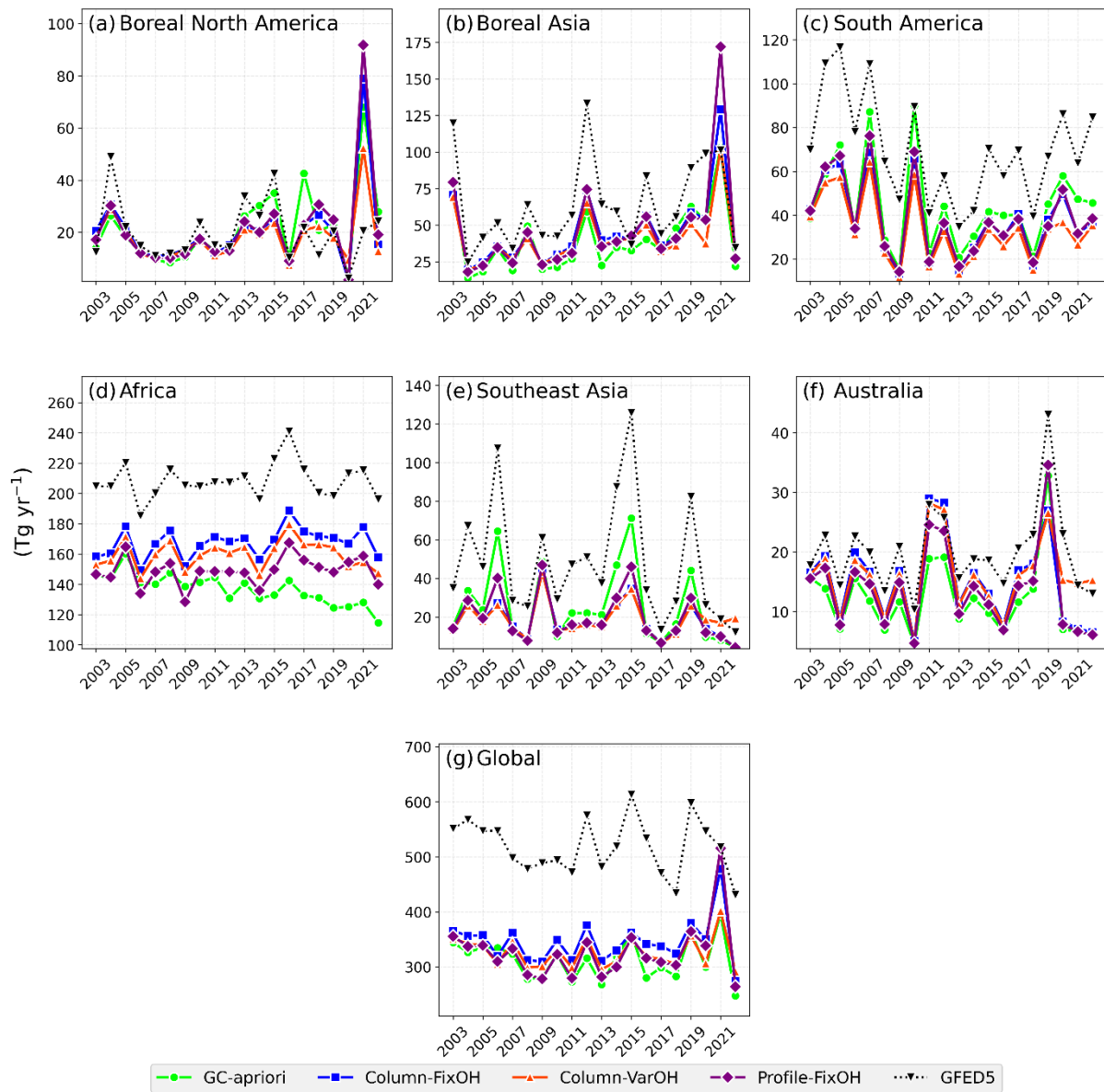


Fig. 5. Time series of biomass burning CO emissions from 2003 to 2022, comparing the a priori inventory (GFED4), a posteriori inversions, and the GFED5 inventory.

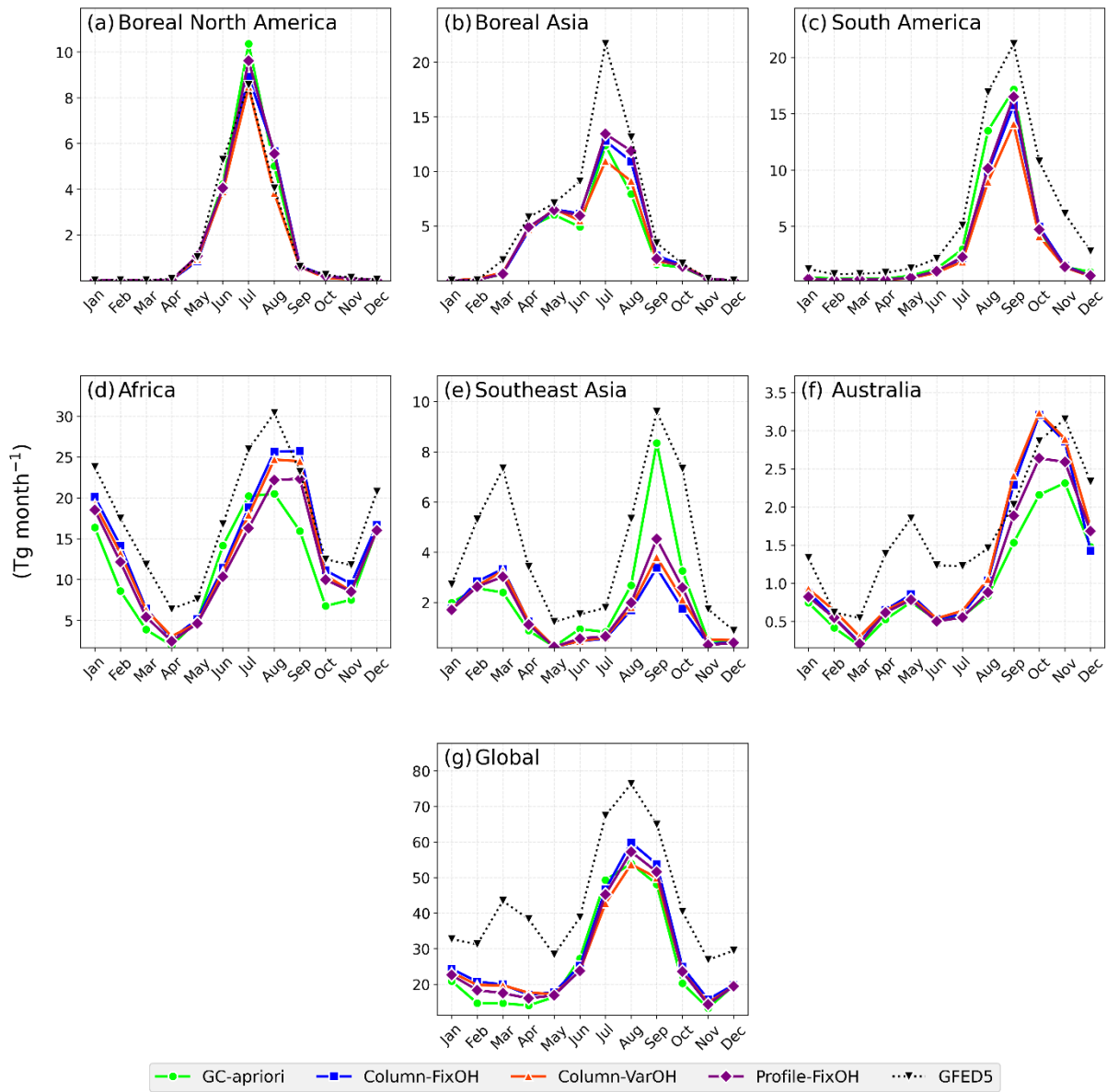


Fig. 6. Climatological monthly cycle (2003-2022 average) of biomass burning CO emissions across different regions, comparing a priori (GFED4) and a posteriori estimates with GFED5.

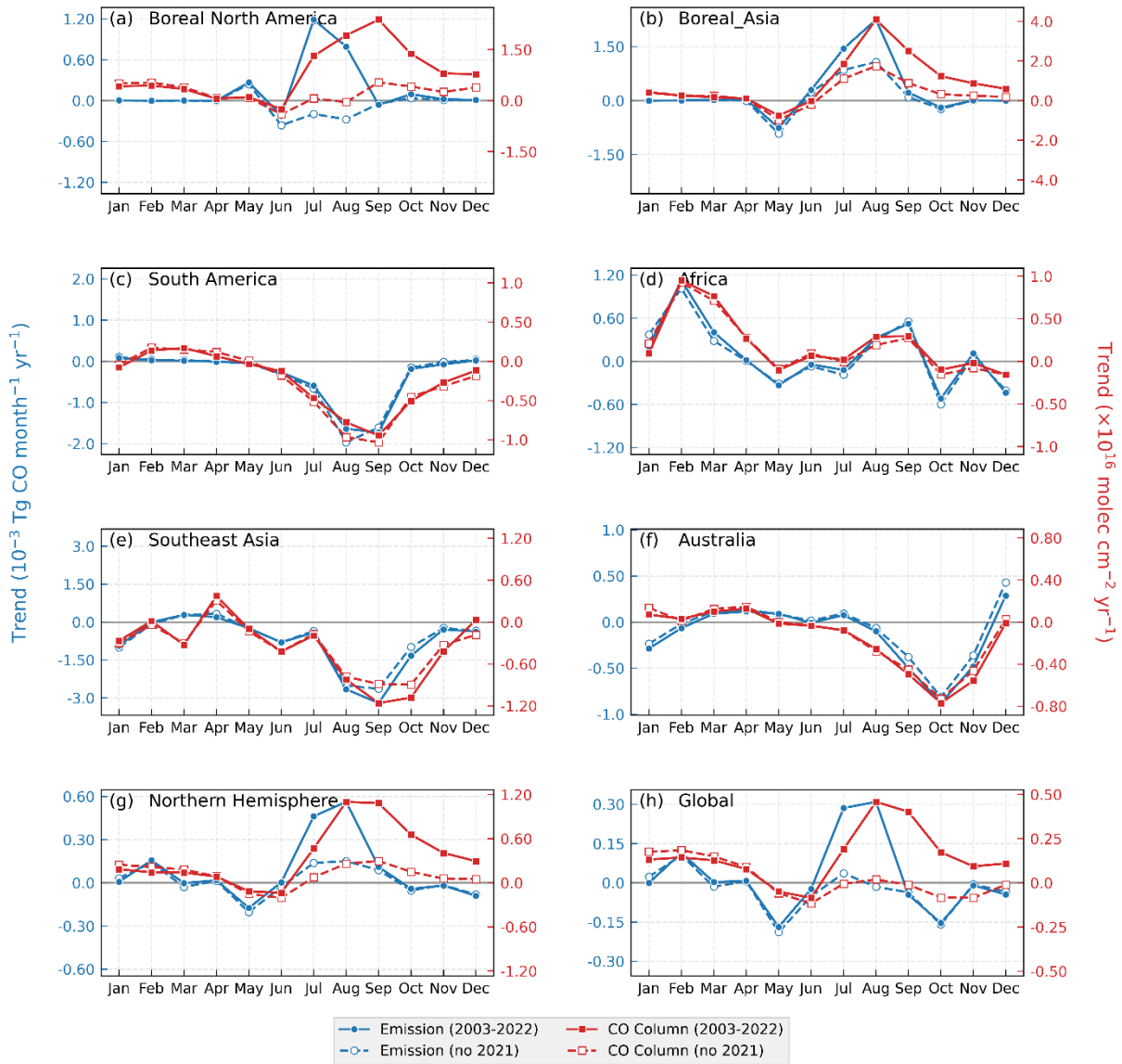


Fig. 7. Monthly trends in biomass burning CO emissions (based on Column-FixOH) and their impact on column CO concentrations (2003-2022). Solid lines show trends including 2021, while dashed lines exclude 2021 to illustrate the impact of extreme fire year. Please check Fig. S4 and S5 for the standard deviation of the trends.

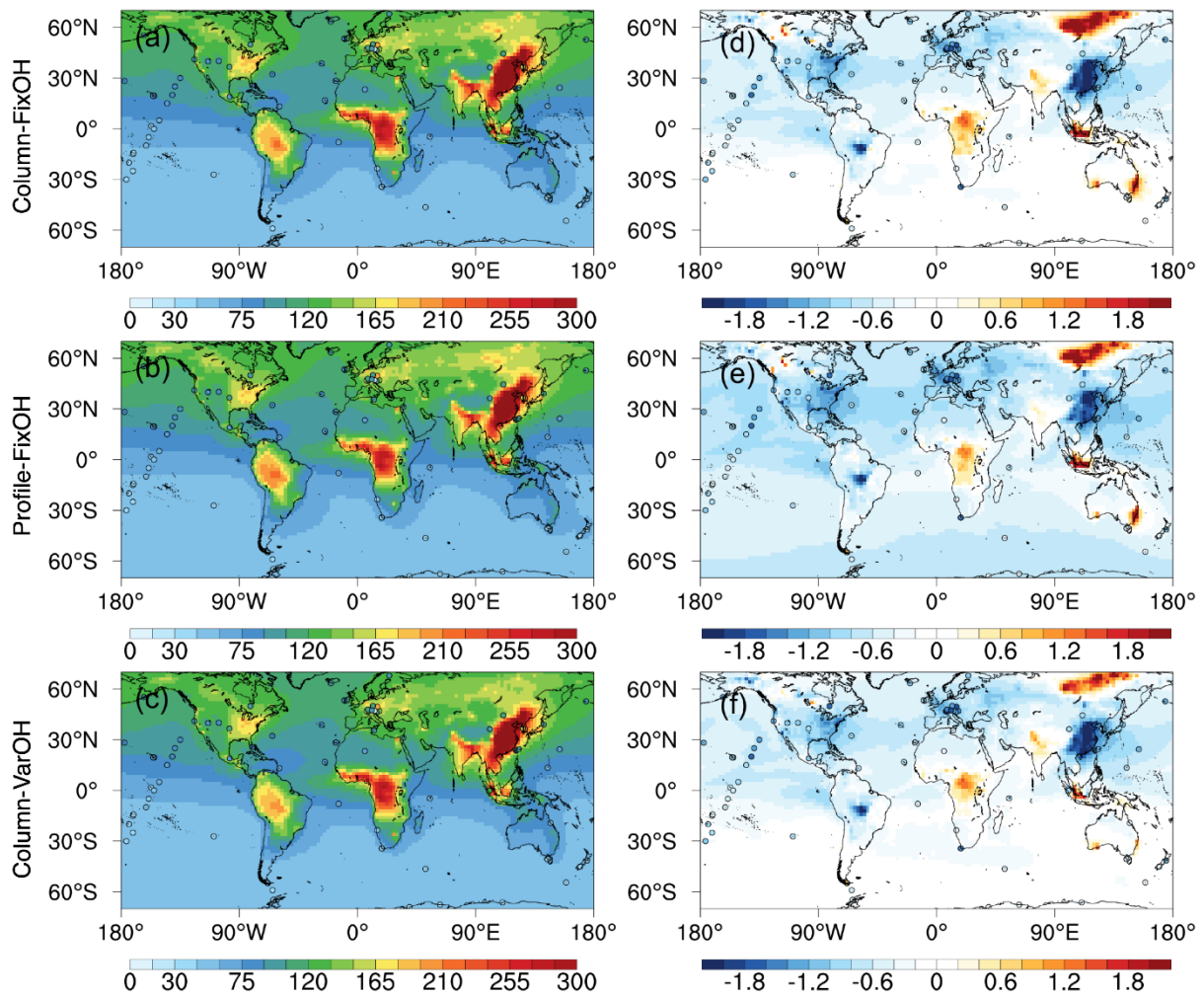


Fig. 8. Modeled surface CO concentrations (ppb) and their trends (% yr⁻¹) from 2003 to 2022. (a-c) Mean concentrations from WDCGG observations and model simulations. (d-f) Spatial pattern of the long-term trend. Only stations with 14 year observations (the time range between the first and last observations) during 2003-2022 are included.

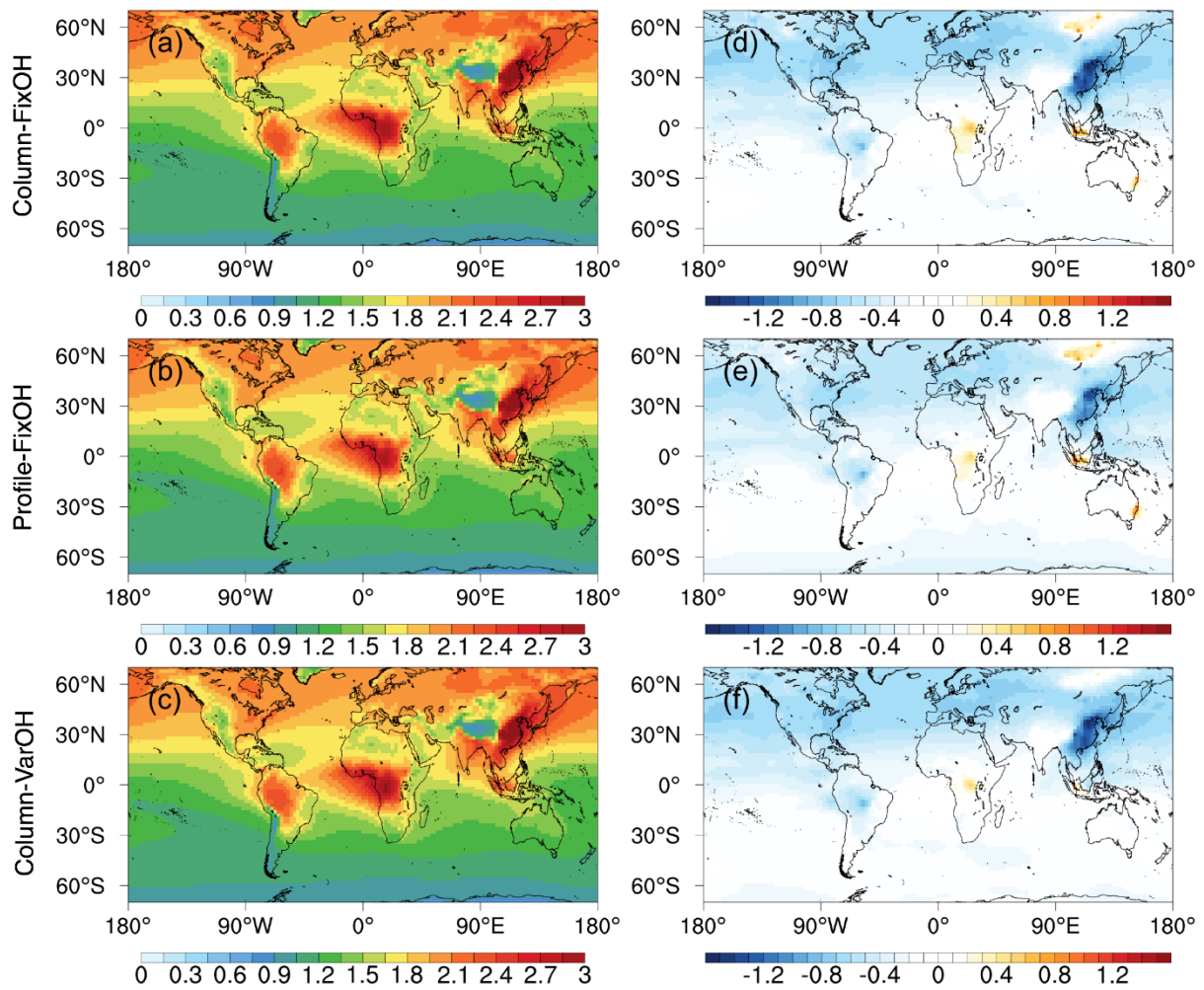


Fig. 9. Modeled column CO concentrations (10^{18} molec cm^{-2}) and their trends ($\% \text{ yr}^{-1}$) from 2003 to 2022. (a-c) Spatial distribution of the 20-year mean. (d-f) Spatial pattern of the long-term trend.

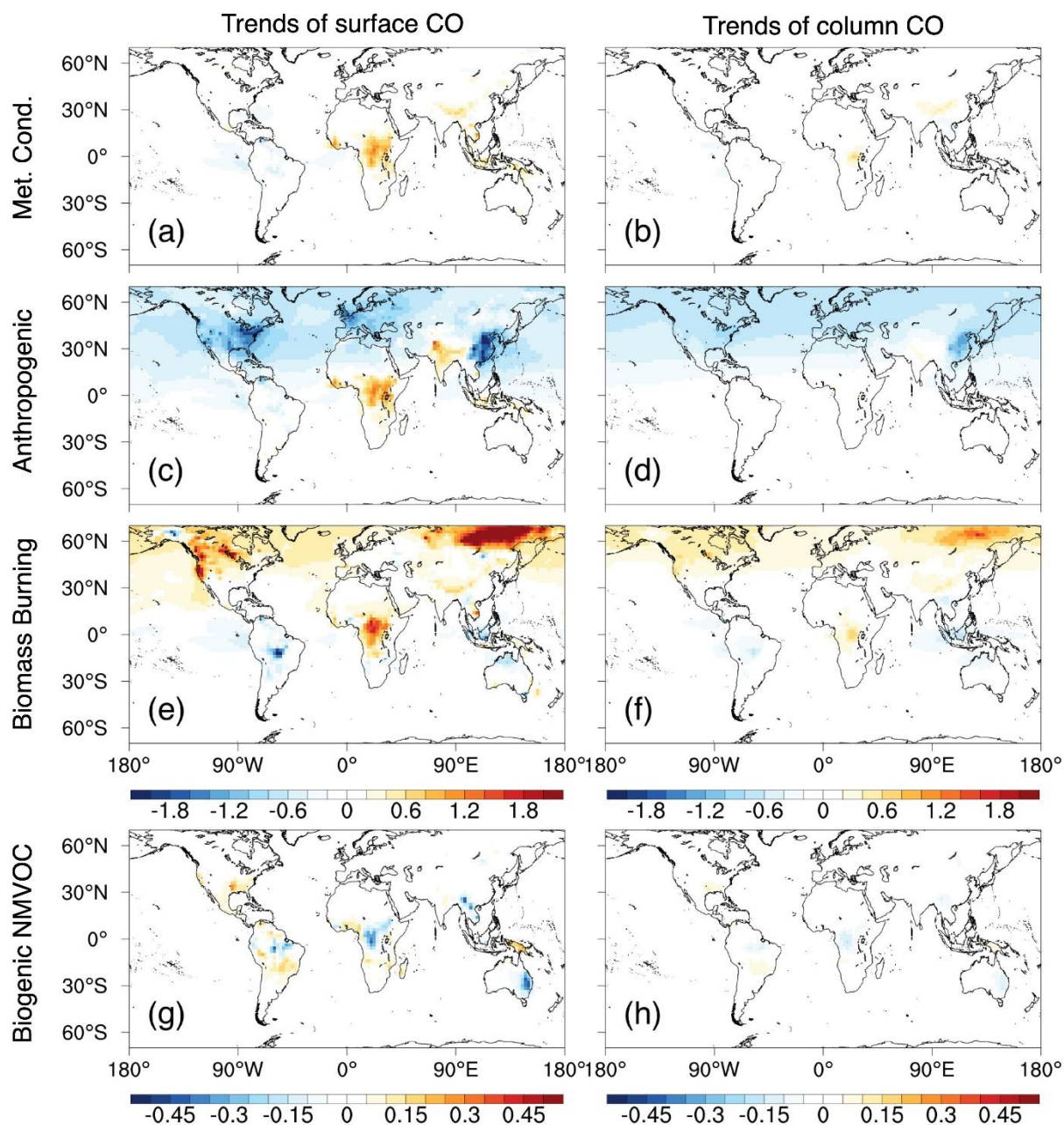


Fig. 10. Attribution of trends ($\% \text{ yr}^{-1}$) in surface and column CO concentrations to individual drivers, derived from sensitivity simulations based on Column-FixOH inversion (2003-2022). Trends are shown for scenarios with: (a, b) all sources fixed at 2003 levels; (c, d) only anthropogenic emissions varying over time; (e, f) only biomass burning emissions varying; (g, h) only biogenic VOC sources varying.

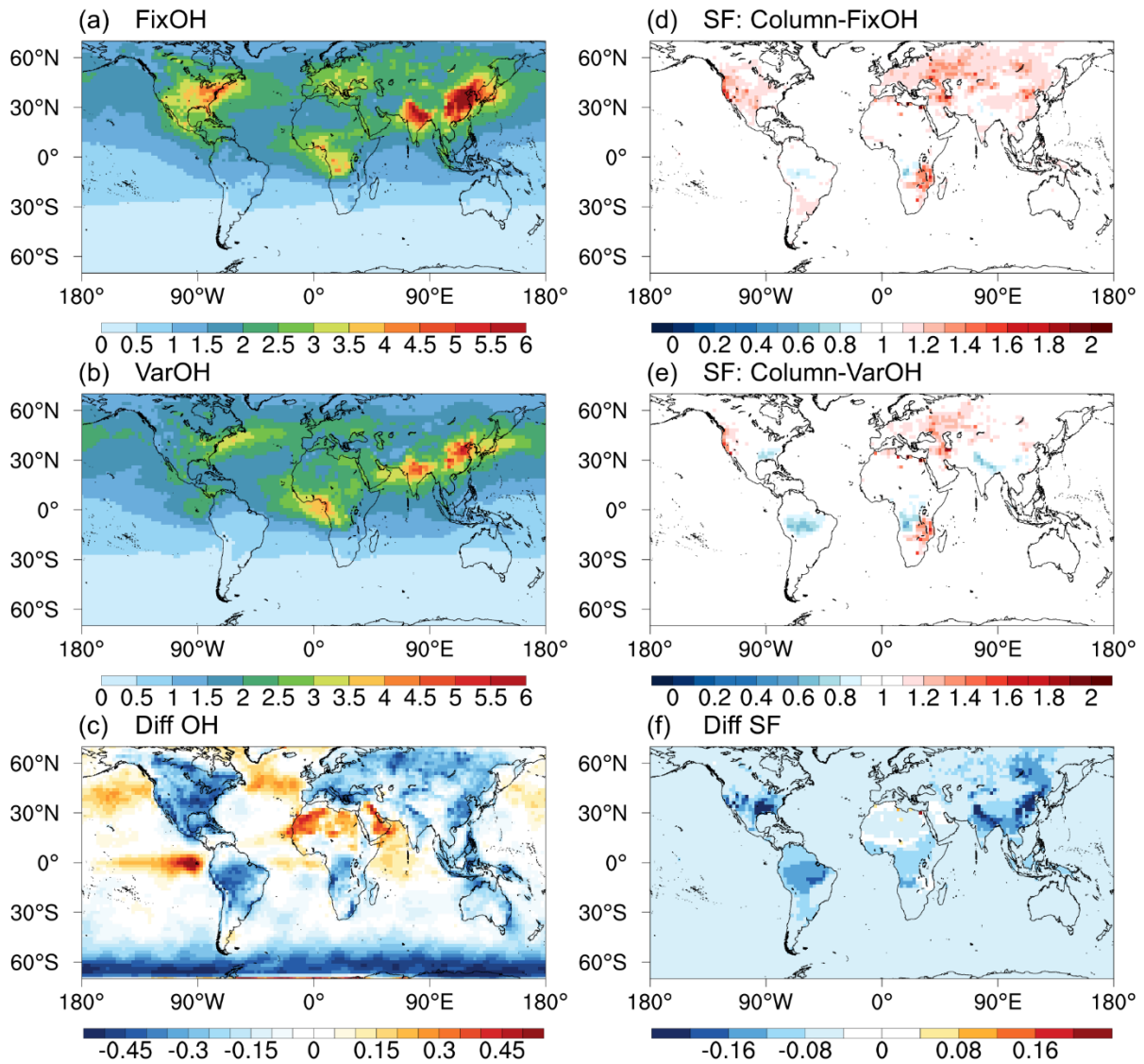


Fig. 11. Sensitivity of inverted CO emissions to OH fields. (a-c) Tropospheric OH columns (10^{12} molec cm^{-2}) from fixed and variable (TCR-2) fields and their difference. (d-f) Corresponding scaling factors from the Column-FixOH and Column-VarOH inversions and their difference. Please note that due to the use of land boundary conditions, differences in OH concentrations over the ocean in the left column figures have a negligible effect on the differences in scaling factors shown in the right column figures.