
1 **Global CO emissions and drivers of atmospheric CO trends constrained by**
2 **MOPITT satellite measurements**

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20 **Abstract**

21 Carbon monoxide (CO), an important atmospheric pollutant produced by incomplete
22 combustion and hydrocarbon oxidation, significantly affects atmospheric oxidation capacity
23 and air quality. Accurate quantification of its global emissions and the underlying driver behind
24 its atmospheric trends is essential for understanding changes in global atmospheric
25 environment. Using 20 years (2003-2022) of data from the Measurement of Pollution in the
26 Troposphere (MOPITT) instrument, we analyze changes in global CO emissions and
27 atmospheric concentrations by applying a four-dimensional variational (4D-Var) assimilation
28 framework within the GEOS-Chem adjoint model. A posteriori simulations show good
29 agreement with independent surface and aircraft measurements compared to a priori
30 simulations. Sensitivity analyses further confirm that inferred emissions remain robust against
31 uncertainties associated with satellite vertical sensitivity and variations in hydroxyl radical (OH)
32 concentrations. Our results indicate a substantial decline in global anthropogenic CO emissions
33 of 14-17% (approximately 85-110 Tg yr⁻¹) over the two-decade period, largely driven by

34 emission reductions in the United States, Europe, and eastern China. Biomass burning
35 emissions exhibited strong interannual variability, with recent increases in Northern
36 Hemisphere high-latitude forests; in particular, the intense 2021 wildfires substantially offset
37 the anthropogenic emission-driven decline in atmospheric CO over the Northern Hemisphere.
38 This study provides a comprehensive assessment of global CO emissions and the mechanisms
39 governing atmospheric CO trends, offering a scientific basis for integrated policies addressing
40 both air pollution and climate change.

41

42 **1. Introduction**

43 Carbon monoxide (CO) is a key atmospheric pollutant produced from incomplete
44 combustion and the oxidation of hydrocarbons. As the main sink for the hydroxyl radical (OH),
45 CO critically influences the oxidative capacity of the atmosphere (Zhao et al., 2020; Tan et al.,
46 2022), and is an important precursor for tropospheric ozone (Whaley et al., 2015; Hu et al.,
47 2024). With a chemical lifetime of approximately one to two months, CO is frequently
48 employed as a valuable tracer for elucidating variations in anthropogenic activities and biomass
49 burning, providing critical insights into the long-range transport of atmospheric constituents
50 (Tang et al., 2019; Buchholz et al., 2022; Smoydzin and Hoor, 2022). By modulating the
51 abundance of OH, changes in CO concentrations indirectly affect the atmospheric lifetime of
52 methane (CH₄). Furthermore, CO shares common combustion sources with major greenhouse
53 gases like CH₄ and carbon dioxide (Worden et al., 2017; Zheng et al., 2023). Accurate
54 quantification of global CO emissions and a clear understanding of the drivers behind its
55 atmospheric trends are therefore essential for formulating effective policies to address the
56 challenges of air quality and climate change.

57 The advent of long-term satellite measurements has revolutionized our ability to monitor
58 global CO distributions (Warner et al., 2013; Worden et al., 2013; Hedelius et al., 2021),

59 enabling a shift from short-term, regional emission estimates (Arellano et al., 2004; Heald et
60 al., 2004; Kopacz et al., 2010) to analyses of decadal-scale changes. Numerous studies have
61 leveraged these records to report substantial declines in anthropogenic CO emissions (Fortems-
62 Cheiney et al., 2011; Jiang et al., 2017; Miyazaki et al., 2020), especially across the Northern
63 Hemisphere, contributing to improved air quality. However, a critical and emerging challenge
64 is to disentangle the competing influences on atmospheric CO concentrations. While
65 anthropogenic emissions are generally decreasing, biomass burning emissions exhibit strong
66 interannual variability. Thus, an important unanswered question is to what extent the recent
67 intensification of wildfires, particularly in high-latitude forests (Jain et al., 2024; Jones et al.,
68 2024), is offsetting the gains achieved from anthropogenic emission reductions. This has
69 profound implications, as a rise in wildfire CO signals a concurrent rise in wildfire greenhouse
70 gas emissions, which could offset part of the gains achieved from reductions in anthropogenic
71 greenhouse gas emissions.

72 Constraining global emissions and robustly attributing observed concentration trends
73 require the application of sophisticated inverse modeling approaches. These methods, which
74 include ensemble-based techniques (e.g., the ensemble Kalman filter) and variational methods
75 (e.g., four-dimensional variational, 4D-Var, data assimilation), provide powerful frameworks
76 for optimizing emission estimates by reconciling model simulations with satellite
77 measurements, while accounting for complex atmospheric transport and chemistry (Müller et
78 al., 2018; Miyazaki et al., 2020; Jiang et al., 2025). Among these, the 4D-Var data assimilation,
79 implemented within chemical transport models like GEOS-Chem and its adjoint (Henze et al.,
80 2007), has been widely and successfully applied to constrain CO emissions (Kopacz et al.,
81 2010; Jiang et al., 2015b; Tang et al., 2023), owing to its strengths in handling nonlinear
82 constraints and providing computationally efficient gradients. However, long-term multi-
83 decadal trend analyses based on this system has often been hindered by limitations such as

84 inconsistent meteorological inputs across years and the use of outdated a priori emission
85 inventories (Jiang et al., 2017; Qu et al., 2022).

86 To address these limitations, we employ a recent extension of the GEOS-Chem adjoint
87 model (Tang et al., 2023) that features support for consistent MERRA-2 meteorological data
88 and modern emission inventories via the Harmonized Emissions Component (HEMCO)
89 (Keller et al., 2014; Lin et al., 2021). By assimilating MOPITT (Measurements of Pollution in
90 the Troposphere) CO retrievals from 2003 to 2022, this study aims to provide an analysis with
91 the following specific objectives: (1) to quantify the long-term evolution of global CO
92 emissions; (2) to attribute the observed trends in atmospheric CO concentrations to changes in
93 emissions and meteorological variations, in particular, the effect of biomass burning emissions
94 on atmospheric CO decline driven by anthropogenic reductions; and (3) to evaluate the
95 sensitivity of inferred emissions to uncertainties in satellite vertical sensitivity and OH
96 concentrations. By doing so, this work aims to improve the understanding of key drivers behind
97 atmospheric CO changes and offer a refined emission inventory to support future air quality
98 and climate policies.

99 The paper is structured as follows: Section 2 describes the methodology, including the
100 assimilation framework, observational data, and the design of assimilation experiments.
101 Section 3 presents the results on the long-term emission trends, the robustness tests, and the
102 attribution of concentration changes. Conclusions are provided in Section 4.

103

104 **2. Methodology and Data**

105 **2.1 Assimilation framework**

106 We utilize the adjoint of the GEOS-Chem model (version 35n) with extended support for
107 MERRA-2 meteorological data and HEMCO emission inventories. The analysis is conducted
108 at a horizontal resolution of $2^\circ \times 2.5^\circ$ with 47 vertical levels (MERRA-2) up to 0.01 hPa and

109 employs a CO-only simulation (tagged-CO mode), in which the chemical sink of CO is
110 linearized with archived monthly mean OH fields. Two types of archived OH fields are used
111 in this study: fixed monthly OH fields for 2013 from the GEOS-Chem full chemistry simulation
112 (Fisher et al., 2017), and variable monthly OH fields for 2005-2020 from the Tropospheric
113 Chemistry Reanalysis version 2 (TCR-2, Miyazaki et al. (2020)). The TCR-2 OH fields have
114 been validated against various aircraft observations and show generally good agreement
115 (Miyazaki et al., 2020). Fig. S2 (see the SI) shows global mean tropospheric OH concentrations
116 from TCR-2, demonstrating a slight increasing trend ($1.0 \pm 0.6 \times 10^3 \text{ molec cm}^{-3} \text{ yr}^{-1}$) in 2005-
117 2020.

118 The global default anthropogenic emission inventory is the CEDS-CMIP6 (Community
119 Emissions Data System) (Hoesly et al., 2018). Regional emissions are replaced as follows:
120 MIX (Li et al., 2017) over Asia, NEI 2016 (National Emissions Inventory) over the United
121 States, DICE_AFRICA and EDGARv4.3 over Africa, and APEI over Canada. The contribution
122 of co-emitted anthropogenic VOC sources is considered by scaling up anthropogenic CO
123 emissions by 11%. Biogenic ~~sources~~ are simulated using the Model of Emissions of Gases and
124 Aerosols from Nature, version 2.0 (MEGANv2.0, Guenther et al. (2006)). CH₄ oxidation
125 source is considered by using a prescribed, spatially varying CH₄ field following the default
126 tagged-CO configuration (Fisher et al., 2017). The CO sources from both biogenic VOCs and
127 CH₄ oxidation are calculated online based on the assumption of instantaneous oxidation by OH
128 radicals. Biomass burning emissions are based on the Global Fire Emissions Database version
129 4 (GFED4, van der Werf et al. (2010)). For years beyond the end year of a specific inventory,
130 emissions from the last available year within that inventory's coverage were used to fill the
131 subsequent years. The distribution of the annual mean CO emissions from 2003 to 2022 is
132 shown in Figs. 1a-c.

133 The annual global sources are 536.3 Tg yr⁻¹ from anthropogenic emissions, 312.5 Tg yr⁻¹

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135 from biomass burning, 623.2 Tg yr⁻¹ from the oxidation of biogenic VOCs, 922.5 Tg yr⁻¹ from
 136 the oxidation of CH₄, with a total sink (through the reaction with OH radicals) of approximately
 137 2395.0 Tg yr⁻¹ in a priori inventories in this work. For comparison, Zheng et al. (2019), reported
 138 inversion-based global CO budget estimates for 2005-2017 of approximately 700 Tg yr⁻¹ for
 139 anthropogenic emissions, 500 Tg yr⁻¹ for biomass burning, 300 Tg yr⁻¹ for biogenic VOC
 140 oxidation, and 900 Tg yr⁻¹ for CH₄ oxidation, with a total sink of approximately 2600 Tg yr⁻¹.
 141 Regarding anthropogenic emissions, the CEDS-CMIP6 inventory estimates an average of
 142 607.6 Tg yr⁻¹ for 2003-2014, while the updated CEDS-CMIP7 inventory yields lower values,
 143 averaging 480.2 Tg yr⁻¹ over 2003-2023. For biomass burning, the GFED5 inventory estimates
 144 an average of 518.3 Tg yr⁻¹ for 2003-2022.

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145 The objective of the 4D-Var approach is to minimize the difference between simulations
 146 and observations by minimizing the cost function (Henze et al., 2007):

$$147 \quad J(\mathbf{x}) = \sum_{i=1}^N (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i)^T \mathbf{S}_\Sigma^{-1} (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i) + \gamma (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} (\mathbf{x} - \mathbf{x}_a) \quad (1)$$

148 where \mathbf{x} is the state vector of CO emissions, N is the number of observations distributed in
 149 time over the assimilation period, \mathbf{z}_i are the MOPITT CO retrievals, and $\mathbf{F}(\mathbf{x})$ is the forward
 150 model. Error estimates are assumed to be Gaussian: \mathbf{S}_Σ is the observational error covariance,
 151 which combines a 10% uniform error and the MOPITT CO retrieval error covariance; and \mathbf{S}_a
 152 is a priori error covariance. Here the combustion-related CO sources (fossil fuel, biofuel, and
 153 biomass burning) and the oxidation source from biogenic VOCs are combined, with a uniform
 154 a priori error of 50% assumed following previous studies (Jiang et al., 2013; Jiang et al., 2017).
 155 The CO source from CH₄ oxidation is optimized separately as an aggregated global source,
 156 with a priori uncertainty of 25%. The cost function is minimized by iteratively adjusting the
 157 CO emissions using the quasi-Newton gradient-based optimization L-BFGS-B algorithm (Zhu
 158 et al., 1997) and the adjoint gradients:

$$159 \quad \nabla_{\mathbf{x}} J(\mathbf{x}) = \sum_{k=1}^N \left[2 \left(\frac{\partial \mathbf{F}_i}{\partial \mathbf{x}} \right)^T \mathbf{S}_\Sigma^{-1} (\mathbf{F}_i(\mathbf{x}) - \mathbf{z}_i) \right] + 2\gamma (\mathbf{x} - \mathbf{x}_a)^T \mathbf{S}_a^{-1} \quad (2)$$

162 The LOGX2 method (Jiang et al., 2015a; Jiang et al., 2017) is employed to improve the
163 reduction of negative gradients.

164 Following Jiang et al. (2017), we applied a two-step approach to mitigate the influence of
165 systematic biases in the model simulations. First, a sequential Kalman filter (Todling and Cohn,
166 1994; Tang et al., 2022) was used to assimilate MOPITT CO retrievals from October 1, 2002,
167 to December 31, 2022, providing optimized CO concentration fields with lower bias. As
168 illustrated in Fig. 2a, the GEOS-Chem model driven by the original monthly CO initial
169 conditions and a priori emission inventories (referred to as GC-original) substantially
170 underestimated column CO concentrations by approximately 30–40% (mean bias = $-39.4 \times$
171 10^{16} molec cm^{-2} ; Table 1). In contrast, simulations using the monthly CO initial conditions
172 derived from the sequential Kalman filter together with a priori emissions (GC-a priori) showed
173 markedly improved agreement with MOPITT CO retrievals (Fig. 2b), reducing the mean bias
174 to about 10% (mean bias = -9.7×10^{16} molec cm^{-2}). Similarly, the use of optimized monthly
175 CO initial conditions led to considerable improvement in model performance against
176 independent surface and aircraft measurements (Table 1). The mean bias decreased from -18.3
177 ppb (GC-original) to -1.4 ppb (GC-a priori) for World Data Centre for Greenhouse Gases
178 (WDCGG) surface observations; from -18.9 ppb to -3.8 ppb for HIAPER Pole-to-Pole
179 Observations (HIPPO) aircraft data; and from -16.2 ppb to -3.4 ppb for Atmospheric
180 Tomography Mission (ATom) aircraft measurements. These results suggest that the substantial
181 negative biases seen in Fig. 2a largely originate from the accumulation of biases over preceding
182 months.

183 Furthermore, ocean scenes (pink grids in Fig. S3) were defined as land boundary
184 conditions. The optimized CO fields from the Kalman filter were used to update CO
185 concentrations over the ocean at hourly intervals during the forward simulation within the 4D-
186 Var process. Meanwhile, the 4D-Var system constrained CO emissions over land without

187 modifying oceanic CO distributions. As demonstrated by Jiang et al. (2017), the use of
188 optimized CO land boundary conditions in 4D-Var assimilation effectively reduces systematic
189 biases associated with long-range transport. By adopting this two-step assimilation framework,
190 the inversion focuses on optimizing fresh continental CO emissions, while reducing the
191 influence of uncertainties arising from transport and chemical processes, which tend to exhibit
192 larger systematic biases. Consequently, a posteriori CO emissions estimated in this study are
193 expected to be lower than those derived without adjustments to the initial and boundary CO
194 conditions. This reflects both the specific inverse modeling setup and a possible
195 underestimation in our a posteriori emission estimates, attributable to the emphasis on
196 constraining fresh continental CO sources.

197 Based on this assimilation framework, three sets of CO emission inversion experiments
198 are designed:

199 (1) Column-FixOH: uses MOPITT CO column concentration data with default OH fields
200 fixed in 2013.

201 (2) Profile-FixOH: uses MOPITT CO profile data with default OH fields fixed in 2013.

202 (3) Column-VarOH: uses MOPITT CO column concentration data with variable OH fields
203 from the TCR-2 tropospheric chemistry reanalysis.

204 By comparing the results of Column-FixOH and Profile-FixOH, the influence of different
205 MOPITT CO data types on CO source estimates can be assessed. Similarly, comparing
206 Column-FixOH and Column-VarOH allows for evaluation of the impact of different OH fields
207 on CO source estimates.

208 **2.2 MOPITT CO retrievals**

209 The MOPITT instrument was launched on December 18, 1999, aboard the NASA Terra
210 spacecraft. The satellite follows a sun-synchronous polar orbit at 705 km altitude, crossing the
211 equator at 10:30 local time. The instrument made measurements over a 612 km cross-track

212 scan, with a footprint of $22 \text{ km} \times 22 \text{ km}$. The MOPITT data used in this study are from the
213 joint retrieval (version 9J) of CO, which combines thermal infrared (TIR, $4.7\mu\text{m}$) and near-
214 infrared (NIR, $2.3\mu\text{m}$) radiances using an optimal estimation approach (Worden et al., 2010;
215 Deeter et al., 2022). The retrieved volume mixing ratios are reported as layer averages across
216 10 pressure levels (surface, 900, 800, 700, 600, 500, 400, 300, 200, and 100 hPa). The
217 relationship between the retrieved CO profile and the true atmospheric state is expressed as:

$$218 \quad \hat{\mathbf{z}} = \mathbf{z}_a + \mathbf{A}(\mathbf{z} - \mathbf{z}_a) + \mathbf{G}\boldsymbol{\epsilon} \quad (3)$$

219 where \mathbf{z}_a is the MOPITT a priori CO profile, \mathbf{z} is the true atmospheric state, $\mathbf{G}\boldsymbol{\epsilon}$ represents
220 the retrieval error, and $\mathbf{A} = \partial\hat{\mathbf{z}}/\partial\mathbf{z}$ is the MOPITT averaging kernel matrix, indicating the
221 sensitivity of the retrieval to the actual atmospheric CO. We only consider data with Cloud
222 Description = 2 (cloud free) and exclude MOPITT data with CO column amounts less than
223 $5 \times 10^{17} \text{ molec cm}^{-2}$. The threshold ($5 \times 10^{17} \text{ molec cm}^{-2}$) was selected to prevent the influence of
224 certain potentially inaccurate, extremely low-concentration observations, which may also have
225 low observation errors in the cost function, on the 4D-Var assimilation (Jiang et al., 2013; Jiang
226 et al., 2017). Since the NIR channel relies on reflected solar radiation, only daytime data are
227 considered (Worden et al., 2010; Tang et al., 2024).

228 **2.3 Aircraft and surface CO measurements**

229 The HIPPO (Wofsy and HIPPO Science Team (2011)) were conducted using the
230 Gulfstream V aircraft from 2009 to 2011. The flights primarily covered the Pacific Ocean,
231 spanning latitudes from 67°S to 87°N , with continuous sampling from 0.2 to 12 km altitude.
232 The ATom (Wofsy and Atom Science Team (2018)) used the DC-8 aircraft from 2016 to 2018.
233 ATom covered similar altitude and latitudinal ranges as HIPPO but with broader spatial
234 coverage, particularly over the Atlantic Ocean. For HIPPO, a total of 687 CO profiles from
235 five missions were used directly. For ATom, CO measurements during continuous ascents and
236 descents were used to construct 523 CO profiles from four missions. Surface CO measurements

237 from the WDCGG are also included in this analysis. The WDCGG, operated by the Japan
238 Meteorological Agency under the World Meteorological Organization's Global Atmosphere
239 Watch (GAW) program, collects, archives, and distributes atmospheric greenhouse gas data,
240 including CO, contributed by various institutions worldwide.

241

242 **3. Results and Discussion**

243 **3.1 Evaluation of assimilation system performance**

244 Before presenting the estimated emission trends, we first evaluate the performance of our
245 assimilation system. The evaluation involves comparing modeled CO concentrations from the
246 GC-original, GC-a priori, and a posteriori simulations (Column-FixOH, Profile-FixOH,
247 Column-VarOH) over the period 2003-2022 against MOPITT satellite measurements, as well
248 as independent surface observations from WDCGG and aircraft measurements from HIPPO
249 and ATom. As summarized in Table 1, a posteriori simulations exhibit mean biases relative to
250 MOPITT retrievals ranging from -5.1 to -7.3×10^{16} molec cm⁻². These values are notably
251 smaller than the biases in the GC-a priori simulation (-9.7×10^{16} molec cm⁻²) and the GC-
252 original simulation (-39.4×10^{16} molec cm⁻²). Similarly, for the HIPPO aircraft observations,
253 a posteriori simulations show mean biases between -2.5 and -2.1 ppb, improved compared to
254 the GC-a priori (-3.8 ppb) and GC-original (-18.9 ppb) simulations. For ATom aircraft data, a
255 posteriori mean biases range from -2.9 to -1.6 ppb, also lower than those from the GC-a priori
256 (-3.4 ppb) and GC-original (-16.2 ppb) simulations. In the case of surface CO concentrations,
257 a posteriori simulations yield mean biases between 0.3 and 1.9 ppb relative to WDCGG
258 observations (Table 1), which are reduced compared to GC-original (-18.3 ppb) simulations,
259 and comparable with the GC-a priori (-1.4 ppb) simulations. A posteriori simulations slightly
260 overestimate surface concentrations relative to WDCGG data, while underestimating CO in the
261 free troposphere according to MOPITT and aircraft measurements. This systematic

262 discrepancy may be attributable to uncertainties in convective transport parameterizations
263 within the model.

264 Overall, the good consistency between a posteriori simulations and multiple independent
265 observation platforms demonstrates the capability of our assimilation system to effectively
266 constrain CO emissions. Given this confidence in the system's performance, we now present
267 the central findings of this study: the long-term evolution of CO emissions. As mentioned in
268 Section 2.1, the combustion-related CO sources and the oxidation source from biogenic VOCs
269 are combined, and thus, the inverse system optimizes total CO emissions within each model
270 grid cell. The subsequent attribution of emissions to specific source types (e.g., anthropogenic,
271 biomass burning) in an individual grid cell is based on the relative contribution of each source
272 category from a priori emission inventories. Specifically, a posteriori emission for a given
273 source type in a grid cell is calculated by applying the grid-scale scaling factor (the ratio of a
274 posteriori to a priori total emissions) to the corresponding a priori emission of that source type.
275 Different sources can finally be calculated because each source category possesses distinct
276 spatial patterns and seasonal variations.

277

278 **3.2 Long-term evolution of global CO emissions**

279 **3.2.1 Anthropogenic CO emissions**

280 At the global scale, anthropogenic CO emissions based on three inversion configurations
281 are estimated to be 7-14% higher than a priori values (Table 2) in 2003-2022 and show a clear
282 declining trend (Fig. 3f). Under the Column-FixOH configuration, global anthropogenic
283 emissions from 2003 to 2022 ranged from 546.1 to 654.1 Tg yr⁻¹, with a multi-year average of
284 approximately 610 Tg yr⁻¹ and a total reduction of about 17%; similar emission ranges and
285 reduction rates (14-17%) were obtained under the Profile-FixOH and Column-VarOH
286 configurations. These results are broadly consistent with Zheng et al. (2019). The

287 CEDS-CMIP7 inventory (Hoesly et al., 2018) shows significantly lower global CO emissions
288 than those derived from inverse modeling, though its decreasing trend is comparable. As shown
289 in Fig. 4a, negative trends (blue) were concentrated in three major industrialized regions:
290 eastern North America, Europe, and eastern China, forming a "reduction belt". These regions
291 accounted for over 65% of global anthropogenic CO emissions, and their systematic reductions
292 constituted the principal driver of the global downward trend. In contrast, positive trends (red)
293 were primarily distributed in northern India (increases of 15.2-22.3%) and Central Africa,
294 corresponding to rapid urbanization and industrialization processes.

295 In the United States (US), emissions declined rapidly from 2003 to 2009, followed by a
296 period of slower reduction (Fig. 3a). Over the entire period (2003-2022), US CO emissions
297 decreased at rates of 2.0-2.2 Tg yr⁻¹, resulting in a cumulative reduction of 46-49% (Table S1).
298 This phased reduction pattern is consistent with the diminishing marginal effects of widespread
299 transportation control technologies, as supported by independent studies (Elguindi et al., 2020;
300 Miyazaki et al., 2020). Our estimated emission magnitude and decreasing trend are similar to
301 Zheng et al. (2019) and the CEDS-CMIP7 in the US. European CO emissions (Fig. 3b)
302 followed a similar pattern (cumulative reduction of 32–34% over 2003-2022). Estimated
303 emissions over Europe in Zheng et al. (2019) are substantially higher than ours and show
304 stronger interannual variability. In comparison, the CEDS-CMIP7 inventory shows good
305 agreement with our results during 2003-2017, but a faster decline after 2017; and Fortems-
306 Cheiney et al. (2024) suggests continuous decline in CO emissions in Europe in 2011-2021.
307 This discrepancy could be possibly attributable to differences in the processing of initial and
308 boundary CO conditions (e.g., the use of climatological CO concentrations in Fortems-Cheiney
309 et al. (2024)).

310 The evolution of eastern China's CO emissions can be divided into four stages (Fig. 3c):
311 (1) a slight growth until 2007, peaking around that time; (2) a sharp decline of approximately

312 7% during the 2008 global financial crisis; (3) a temporary rebound from 2008 to 2010 under
313 economic stimulus policies; and (4) a continuous decline phase after 2010. From 2003 to 2022,
314 anthropogenic CO emissions from eastern China decreased at an average rate of 3.0-4.0 Tg yr⁻¹
315 (Table 2), with a cumulative reduction of 23-32% (Table S1). Zhao et al. (2012) and Xia et
316 al. (2016) confirmed the trend reversal around 2007, attributing it to improved energy
317 efficiency and strengthened emission controls, while Lin and McElroy (2011) and Tong et al.
318 (2016) highlighted the suppressive impact of the 2008 economic recession. Both Zheng et al.
319 (2019) and the CEDS-CMIP7 emission dataset show a declining trend consistent with our
320 results, although their emission magnitudes are lower. During 2019-2022, the emission
321 reduction rate accelerated to 4.8-8.3 Tg yr⁻¹, reflecting not only the short-term impact of the
322 COVID-19 pandemic but also the long-term cumulative effects of clean air policies and energy
323 structure transformation.

324 India exhibited a continuous growth in anthropogenic CO emissions from 2003 to 2009,
325 followed by a period of slower increase, with an average annual increase of 0.5-0.8 Tg yr⁻¹ in
326 2003-2022. Our estimated emission magnitude and trend are similar to Zheng et al. (2019) in
327 India. In comparison, the CEDS-CMIP7 inventory shows a similar trend, but its emission levels
328 are lower than those derived from inverse modeling. In Southeast Asia, anthropogenic CO
329 emissions exhibited a relatively stable and slow upward trend over the study period, though a
330 noticeable decline occurred from 2019 to 2022, which is likely associated with the impact of
331 the COVID-19 pandemic. The emission trend derived from our inversion is generally
332 consistent with that reported by Zheng et al. (2019) for this region, although their estimates
333 show stronger interannual variability. Compared with the CEDS-CMIP7 inventory, the trend
334 in CO emissions is similar to our results, but the emission magnitude in CEDS-CMIP7 is lower
335 than that derived from inverse modeling.

336 **3.2.2 Biomass burning CO emissions**

337 Globally, biomass burning CO emissions were 4-11% higher than a priori estimate in
338 2003-2022 (Table 3) and reached a historical peak of approximately 500 Tg yr⁻¹ in 2021 (Fig.
339 5g). In contrast to the clear decline of anthropogenic emissions, the trend in global biomass
340 burning CO emissions remains insignificant (Table 3). A comparison with the GFED5 emission
341 inventory (van der Werf et al., 2025) reveals noticeable differences: GFED5 estimates are
342 generally higher than our results (Fig. 5g), and do not show the 2021 peak. Spatial analysis
343 revealed a pronounced latitudinal differentiation in the changes of biomass burning CO
344 emissions (Figs. 4d-f): positive trends (red) were concentrated in Northern Hemisphere high-
345 latitude coniferous forests, while negative trends (blue) dominated tropical and subtropical
346 regions. This pattern is consistent with the "global fire emission geographic reconstruction"
347 observed by Zheng et al. (2023), reflecting the differential impacts of climate change across
348 latitudinal zones.

349 Emissions from high-latitude coniferous forests have shown different long-term trends
350 between boreal North America and boreal Asia over the past two decades (Figs. 5a-b). Peak
351 fire activity in boreal North America occurs during June-August (Fig. 6a); boreal Asia
352 experiences its primary fire season in June-August, with a secondary peak often observed in
353 March-May (Fig. 6b). When excluding the exceptional wildfire year of 2021, summertime
354 biomass burning CO emissions in boreal North America exhibited an overall declining trend
355 from 2003 to 2022 (Fig. 7a). In contrast, boreal Asia experienced a general increase in
356 summertime biomass burning CO emissions during the same period, even when 2021 is
357 omitted, though the trend is less pronounced than when including that extreme year (Fig. 7b).
358 The peak in wildfire emissions from high-latitude coniferous forests in 2021 was triggered by
359 severe, concurrent droughts across the Northern Hemisphere (Zheng et al., 2023). The
360 pronounced latitudinal amplification of emissions is consistent with higher carbon emission
361 density of boreal forests, which is 4-10 times greater than that of grasslands (Zheng et al., 2021).

362 GFED5 data suggests that boreal Asia's wildfire emissions peaked in 2012, different from
363 emission inversion results in this work and Zheng et al. (2023).

364 A notable decline in fire activity in South America occurred after 2010 (Fig. 5c),
365 particularly in August-September (Fig. 7c) coinciding with the peak wildfire season in South
366 America (Fig. 6c). The trend shift in CO emissions are consistent with the sharp reductions in
367 annual deforestation rates in the Brazilian Amazon from 25396 km² yr⁻¹ in 2003 to 7000 km²
368 yr⁻¹ in 2010 (Deeter et al., 2018). Africa experiences its primary fire season in June-September,
369 with a secondary peak often observed in December-February (Fig. 6d). Biomass burning CO
370 emissions in Africa exhibited a modest increasing trend overall (Fig. 5d), particularly in
371 February (Fig. 7c). Pronounced regional differentiation occurred, with increases in central
372 Africa and decreases in surrounding areas (Figs. 4d-f), reflecting the "strong contrast" pattern
373 described by Andela et al. (2017). Compared to the GFED5 inventory, our inversion results
374 generally show lower CO emission intensities.

375 Peak fire activity in Australia occurs during August-December (Fig. 6f); Southeast Asia
376 experiences its primary fire season in August-October, with a secondary peak often observed
377 in February-March (Fig. 6e). Emission patterns in Southeast Asia (Fig. 5e) and Australia (Fig.
378 5f) highlighted their sensitivity to large-scale climate oscillations. Major fire events in
379 Indonesia in 2006, 2009, 2015, and 2019 were closely linked to El Niño-induced droughts
380 (Page, 2009; Field et al., 2016). Australia's extreme fires in 2019 resulted from compound
381 extreme climate conditions influenced by the El Niño-Southern Oscillation, the Southern
382 Annular Mode, and the Indian Ocean Dipole (Deb et al., 2020). Building upon the observed
383 sensitivity to large-scale climate oscillations, the long-term interannual trend of wildfire
384 emissions across Southeast Asia and Australia remains insignificant, despite decline in August-
385 October in Southeast Asia (Fig. 7e) and September-November in Australia (Fig. 7f). Our

386 emission inversion shows lower CO emissions than GFED5 inventory in Southeast Asia and
387 Australia.

388 3.2.3 Difference between combustion and biogenic NMVOC sources

389 CO from combustion sources in the Northern Hemisphere showed strong regional
390 differentiation (Fig. 4), reflecting a dynamic redistribution between declining anthropogenic
391 sources and increasing biomass burning sources. Positive trends were densely distributed in
392 high-latitude regions, mainly due to increases in wildfires; Negative trends dominated mid-to-
393 low latitude industrialized areas. Tropical regions showed a mixed pattern, while the Southern
394 Hemisphere exhibited generally weaker trends. This spatial heterogeneity confirms a net global
395 decrease in combustion-related CO, revealing a clear contrast between increases at high
396 northern latitudes and decreases at mid-latitudes, reflecting the compound influences of
397 industrialization, policy interventions and climate change.

398 In contrast, CO produced from the oxidation of biogenic VOCs remained relatively stable
399 from 2003 to 2022 (Figs. 4g-i). This stability aligns with findings by Messina et al. (2016),
400 suggesting that global-scale biogenic VOC ~~sources~~ are less sensitive to short-term climate and
401 land cover changes. The global stability of biogenic VOC-derived CO is important for
402 atmospheric chemistry, as these compounds are key reactants for OH radicals and play a
403 regulatory role in atmospheric oxidation capacity. This stable background provides a crucial
404 baseline for understanding changes in atmospheric oxidation processes. The weaker trends
405 compared to those reported by Jiang et al. (2017) may be associated with our use of continuous
406 MERRA-2 meteorological data, which enhances consistency in long-term analysis.

407

408 3.3 Long-term evolution and drivers of global CO concentrations

409 Building on the emission estimates evaluated above, this section investigates their ultimate
410 influence in the atmosphere by analyzing the spatiotemporal patterns and trends of CO

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412 concentrations. We first present the mean state and long-term changes in CO concentrations,
413 and then quantitatively attribute these changes to their underlying drivers: emissions and
414 meteorology. Figs. 8a-c show the mean surface CO concentrations in 2003-2022 from a
415 posteriori simulations and WDCGG surface observations. Higher CO concentrations are
416 evident in regions with strong anthropogenic emissions, such as East Asia, India, and Southeast
417 Asia, as well as in areas with significant biomass burning, i.e., Central Africa and South
418 America. The long-term trends in surface CO (Figs. 8d-f) reveal declining concentrations over
419 North America, Europe, East Asia, and South America, which contrast with rising trends over
420 India, Boreal Asia, Central Africa, and Australia. The 20-year mean CO columns (Figs. 9a-c)
421 show a consistent spatial pattern, with the highest column concentrations over East Asia and
422 Central Africa, followed by South America, India, and Southeast Asia. In contrast, the long-
423 term trend of CO columns (Figs. 9d-f) exhibits a more uniform decrease across the Northern
424 Hemisphere, lacking the distinct regional hotspots observed in the surface trends. This suggests
425 that changes in CO are more thoroughly mixed within the column.

426 To quantitatively attribute the concentration trends to specific drivers, we conducted a
427 series of sensitivity experiments. The experimental design isolates the influence of individual
428 emission sectors by building a baseline scenario in which all emissions are fixed at 2003 levels
429 to reflect the impact of meteorological condition changes in 2003-2022. Three more sensitivity
430 experiments were then conducted in 2003-2022 in which only one emission category, i.e.,
431 anthropogenic, biomass burning, or biogenic VOC sources, was allowed to vary over time,
432 respectively. The time-varying sources in these sensitivity experiments were prescribed from
433 the Column-FixOH a posteriori inversion.

434 The results indicate that meteorological influences induced positive trends in surface CO
435 concentrations in regions such as central Africa, Southeast Asia, and the Tibetan Plateau (0.6-
436 1.8% yr⁻¹), along with slight negative trends in areas such as South America (Fig. 10a). The

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439 meteorological impact on CO column concentrations was comparatively weaker (Fig. 10b),
440 showing positive trends of $0.45\% \text{ yr}^{-1}$ over central Africa and the Tibetan Plateau. This vertical
441 differentiation implies that meteorological influences may primarily alter the vertical
442 distribution of CO through changes in convective transport, with a more limited effect on larger
443 horizontal scales. The derived meteorological impact is noticeably weaker than that reported
444 by Jiang et al. (2017), a discrepancy likely attributable to our use of consistent MERRA-2
445 meteorological fields, which enhances the reliability of the long-term trend analysis. Similarly,
446 the impact of biogenic VOC changes on CO concentrations (Figs. 10g, 10h) was markedly
447 weaker than in Jiang et al. (2017).

448 Anthropogenic emission changes were identified as the principal driver behind declining
449 CO levels, inducing strong negative trends in industrial regions of the Northern Hemisphere,
450 such as eastern North America, Europe, and eastern China. This signal is consistent across both
451 surface and column concentrations (Figs. 10c-d). Globally, anthropogenic emission changes
452 led to an average annual decrease of $0.27\% \text{ yr}^{-1}$ in CO column concentrations, with a more
453 pronounced decline rate of $0.51\% \text{ yr}^{-1}$ in the Northern Hemisphere (Table 4). Regionally, the
454 US, Europe, and eastern China exhibited the most substantial decreases, at $-0.57\% \text{ yr}^{-1}$, -0.69%
455 yr^{-1} and $-0.69\% \text{ yr}^{-1}$, respectively. In contrast, India experienced a slight concentration increase
456 ($0.03\% \text{ yr}^{-1}$) due to rising emissions, while Southeast Asia showed a more moderate decline ($-$
457 $0.19\% \text{ yr}^{-1}$) compared to other major industrial regions.

458 Conversely, changes in biomass burning emissions generally contributed to positive trends
459 in CO, particularly at high latitudes (Figs. 10e-f). At global and Northern Hemispheric scales,
460 this positive trend was largely attributable to extreme wildfire activity in 2021. When 2021 is
461 excluded, the long-term trend in CO columns due to biomass burning becomes statistically
462 insignificant at these broad scales (Figs. 7g-h), with only regionally and seasonally confined
463 increases remaining apparent, notably over Boreal Asia (July-August, Fig. 7b) and Africa

464 (January-April, Fig. 7d). In the full record (including 2021), biomass burning emissions led to
465 an average annual increase of $0.10\% \text{ yr}^{-1}$ in global CO columns, and $0.24\% \text{ yr}^{-1}$ in the Northern
466 Hemisphere (Table 4). It is noteworthy that the CO concentration response lagged behind
467 emission pulses by about one month and persisted longer. In the Northern Hemisphere, for
468 instance, enhanced emissions occurred mainly from July to September, whereas the significant
469 concentration response extended from August to December (Fig. 7g). This lag and prolonged
470 influence were primarily attributable to the delayed response over Boreal North America (Fig.
471 7a). At the regional scale, increases occurred in Boreal North America ($0.43\% \text{ yr}^{-1}$) and Boreal
472 Asia ($0.48\% \text{ yr}^{-1}$). In contrast, South America, Australia, and Southeast Asia experienced
473 declining trends ranging from $-0.13\% \text{ yr}^{-1}$ to $-0.22\% \text{ yr}^{-1}$, while Africa showed a slight increase
474 of $0.09\% \text{ yr}^{-1}$.

475 This attribution analysis highlights the substantial impact of extreme wildfire years on the
476 CO budget. Although anthropogenic emission reductions lowered Northern Hemisphere CO
477 columns by approximately $0.51\% \text{ yr}^{-1}$, the intense biomass burning emissions in 2021
478 introduced a positive perturbation of about $0.24\% \text{ yr}^{-1}$ in the full-record trend, thereby offsetting
479 a considerable fraction of the anthropogenic-driven decline. As a result, the net concentration
480 decline was reduced to approximately $0.27\% \text{ yr}^{-1}$ in the analysis including 2021. This implies
481 that nearly half (47%) of the potential air quality improvement from anthropogenic emission
482 controls can be offset by wildfire emissions. This finding provides a clear mechanistic
483 explanation for the decline in atmospheric CO concentrations in recent years, and underscores
484 the growing role of extreme wildfire events in modulating regional to hemispheric air
485 composition.

486

487 **3.4 Impacts of systematic errors on inferred CO emissions**

488 The MOPITT instrument provides retrievals for both CO total column and vertical profile.

489 The degrees of freedom for signal (DFS) for MOPITT multi-spectral profile retrievals
490 (TIR+NIR) is approximately 1.5-2.0 over land, reducing to about 1.0 when converted to a total
491 column (Worden et al., 2010). The discrepancy between a posteriori emission estimates
492 constrained by CO column (Column-FixOH) and profile (Profile-FixOH) data helps evaluate
493 the influence of systematic errors associated with the vertical sensitivity of the satellite
494 retrievals (Tang et al., 2024). Globally, a posteriori anthropogenic and biomass burning CO
495 emissions from Profile-FixOH were both slightly lower than those from Column-FixOH, with
496 average differences of -6.6% and -5.5%, respectively, over the period 2003-2022 (Table 2).
497 The two configurations also showed broadly consistent long-term trends in inferred
498 anthropogenic emissions, both indicating a global decline of approximately $-0.9\% \text{ yr}^{-1}$. Larger
499 regional discrepancies were observed over eastern China ($-2.1\% \text{ yr}^{-1}$ for Profile-FixOH vs. -
500 $1.6\% \text{ yr}^{-1}$ for Column-FixOH) and India ($1.1\% \text{ yr}^{-1}$ vs. $0.7\% \text{ yr}^{-1}$). Similarly, the trends in global
501 biomass burning CO emissions were consistent ($0.3\% \text{ yr}^{-1}$ for Column-FixOH and $0.5\% \text{ yr}^{-1}$
502 for Profile-FixOH), though regional differences were more pronounced for boreal North
503 America ($3.1\% \text{ yr}^{-1}$ vs. $4.9\% \text{ yr}^{-1}$) and Australia ($-1.5\% \text{ yr}^{-1}$ vs. $-0.7\% \text{ yr}^{-1}$). The limited
504 differences in inferred emissions between the two configurations resulted in a consistent
505 declining trend in simulated CO columns ($-0.5\% \text{ yr}^{-1}$ for both).

506 OH concentrations in model simulations significantly influence the inverse analysis of CO
507 emissions (Jiang et al., 2011; Müller et al., 2018). By assimilating MOPITT CO column data,
508 we compared the inverted CO emission estimates driven by fixed (Column-FixOH) and
509 variable (Column-VarOH) OH fields to investigate the potential influence. As shown in Fig.
510 11c, OH concentrations from the TCR-2 reanalysis are broadly 10-40% lower than the fixed
511 climatological OH concentrations over land (differences over the ocean are not considered here
512 due to the use of CO land boundary conditions in the 4D-Var assimilation). Lower OH
513 concentrations over land lead to reduced chemical loss, which is compensated by lower global

514 anthropogenic CO emissions in Column-VarOH inversion (590.1 Tg yr^{-1}) in 2003-2022,
515 approximately 3.7% lower than in Column-FixOH (612.8 Tg yr^{-1}).

516 Variations in OH concentrations influence the oxidation of biogenic VOCs to CO and their
517 subsequent chemical loss. These two counteracting processes establish a complex balance,
518 ultimately reflected in the inverted estimates of biogenic CO sources. Specifically, the Column-
519 VarOH inversion yields an average global biogenic CO sources of 391.4 Tg yr^{-1} in 2003-2022,
520 approximately 3.9% lower than the 407.6 Tg yr^{-1} in Column-FixOH inversion. The sensitivity
521 experiments described above address the third objective of this study, which is to evaluate the
522 robustness of our central findings against potential systematic errors associated with satellite
523 retrieval vertical sensitivity and OH concentrations. This robustness can be attributed, in part,
524 to our two-step inversion framework, which mitigates systematic biases through optimized
525 initial and boundary CO conditions.

526

527 **4. Conclusions**

528 This study provides a comprehensive, quantitative analysis of global CO emissions and
529 drivers governing atmospheric CO trends over the past two decades (2003-2022). By
530 employing a 4D-Var assimilation framework within GEOS-Chem adjoint model, constrained
531 by long-term MOPITT satellite measurements, we have generated an observationally
532 constrained CO emission inventory. A central methodological strength lies in the use of
533 continuous MERRA-2 meteorological fields and modern a priori emission inventories, which
534 significantly enhanced the long-term consistency and reliability of our trend analysis. The
535 implementation of a two-step bias mitigation strategy, optimizing both initial conditions and
536 land boundary conditions for CO, effectively reduced the accumulated impacts of transport and
537 chemistry uncertainties. The optimized emissions yield simulated CO concentrations that show
538 good agreement with independent surface measurements from the WDCGG network and

539 aircraft-based profiles from the HIPPO and ATom campaigns. The mean bias in simulated CO
540 concentrations (model minus observation) was reduced from -3.8 ppb in a priori simulation to
541 between -2.5 and -2.1 ppb in a posteriori simulation for HIPPO, and from -3.4 ppb in a priori
542 simulation to between -2.9 and -1.6 ppb in a posteriori simulation for ATOM.

543 Our results demonstrate a significant 14-17% decline (approximately 85-110 Tg yr⁻¹) in
544 global anthropogenic CO emissions over the 20-year period. This reduction was predominantly
545 driven by pollution control policies in major industrialized regions, with cumulative reductions
546 of 46-49% in the US, 32-34% in Europe, and 23-32% in eastern China. The decline in
547 anthropogenic CO emissions is consistent with the trends reported in the CEDS-CMIP7
548 inventory and the inversion results of Zheng et al. (2019), and is identified as the dominant and
549 statistically significant driver behind the observed decrease in atmospheric CO concentrations.
550 For biomass burning, our emission estimates suggested the historical peak of approximately
551 500 Tg yr⁻¹ in 2021, while the overall CO emissions in GFED5 inventory are higher than our
552 estimates in 2003-2022. Biomass burning emissions exhibited strong interannual variability
553 without a statistically significant long-term trend at the global and Northern Hemispheric scales,
554 although regionally and seasonally trends, such as in the boreal Northern Hemisphere during
555 summer, were evident in certain periods.

556 A central finding of this work is the substantial impact of extreme wildfire events,
557 particularly the record-breaking 2021 burning season in Northern Hemisphere high latitudes.
558 Our attribution analysis reveals that these wildfires introduced a strong positive perturbation to
559 atmospheric CO, offsetting nearly half (47%) of the concentration decline driven by
560 anthropogenic reductions in the Northern Hemisphere over our study period (2003-2022). This
561 finding highlights that while not a persistent trend, extreme wildfire events can counteract a
562 large fraction of the gains achieved from decades of emission control efforts. Our analysis thus
563 clarifies the past evolution of global CO emissions and concentrations, highlighting an

564 increasingly critical challenge: climate change is amplifying the intensity and impact of
565 extreme wildfire events, which can periodically undermine emission control efforts. This
566 underscores the need for integrated policies that address both anthropogenic sources and the
567 climate-driven amplification of natural emissions.

568

569 **Code and data availability:** The MOPITT CO data can be downloaded from
570 <https://asdc.larc.nasa.gov/data/MOPITT/>. [The WDCGG data can be downloaded from](#)
571 <https://gaw.kishou.go.jp>. [The HIPPO data can be downloaded from](#)
572 <https://data.eol.ucar.edu/dataset/112.123>. [The ATom data can be downloaded from](#)
573 <https://www.earthdata.nasa.gov/data/catalog/ornl-cloud-atom-merge-v2-1925-2.0>. The adjoint
574 of GEOS-Chem model can be downloaded from [http://wiki.seas.harvard.edu/geos-](http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_Adjoint)
575 [chem/index.php/GEOS-Chem_Adjoint](http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_Adjoint). [The GFED5 data can be downloaded from](#)
576 <https://www.globalfiredata.org/data.html>. [The CEDS data can be downloaded from](#)
577 <https://geos-chem.s3.amazonaws.com/index.html#HEMCO/CEDS/>. [The emission data from](#)
578 [Zheng et al. \(2019\) can be downloaded from](#)
579 [https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454](https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454453/1)
580 [453/1](https://figshare.com/collections/Global_atmospheric_carbon_monoxide_budget_2000_2017/4454453/1). A posteriori CO emission estimates (Column-FixOH, Profile-FixOH and Column-
581 VarOH) [derived in this work](#) can be downloaded from
582 <https://doi.org/10.5281/zenodo.17221834>.

583

584 **Author Contributions:** Z.J. designed the research. Z.T. developed the model code and
585 performed the research. Z.J. and Z.T. wrote the manuscript. All authors contributed to
586 discussions and editing the manuscript.

587

588 **Competing interests:** The authors declare that they have no conflicts of interest.

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595

596 **Tables and Figures**

597 **Table 1.** Mean biases of modeled CO concentrations relative to satellite (MOPITT) and in-situ
598 (WDCGG, HIPPO, ATom) observations for five model simulations over the period 2003-2022.
599 Biases are in units of 10^{16} molec cm^{-2} for MOPITT and ppb for other datasets.

600

601 **Table 2.** Mean anthropogenic CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-2022:
602 A comparison of a priori inventories with those constrained by MOPITT retrievals under
603 different configurations. The region definition is shown in Figure S1e.

604

605 **Table 3.** Mean biomass burning CO emissions (Tg yr^{-1}) and their trends (Tg yr^{-2}) in 2003-
606 2022: A comparison of a priori inventories (GFED4) with those constrained by MOPITT
607 retrievals under different configurations. The region definition is shown in Figure S1f.

608

609 **Table 4.** Attribution of trends in column CO concentrations ($\% \text{ yr}^{-1}$) from 2003 to 2022 to
610 changes in anthropogenic and biomass burning emissions, based on sensitivity simulations
611 using the Column-FixOH inversion results.

612

613 **Fig. 1.** Spatial patterns of a priori CO sources and a posteriori scaling factors for the period
614 2003-2022. (a-c) Mean a priori CO sources (10^{12} molec $\text{cm}^{-2} \text{ s}^{-1}$). (d-l) Scaling factors (ratio of
615 a posteriori to a priori sources) derived from the three inversion experiments: (d-f) Column-
616 FixOH, (g-i) Profile-FixOH, and (j-l) Column-VarOH.

617

618 **Fig. 2.** Relative bias in column CO for 2003-2022, calculated as $(\text{Model} - \text{MOPITT}) / \text{MOPITT}$
619 for GC-original (a), GC-a priori (b), and a posteriori simulations (c-e).

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626 **Fig. 3.** Time series of anthropogenic CO emissions from 2003 to 2022 across major regions,
627 comparing a priori inventories, a posteriori inversions from this study, and independent
628 estimates from CEDS-CMIP6/7 and Zheng et al. (2019).

629
630 **Fig. 4.** Long-term trends in CO sources (10^{10} molec cm^{-2} s^{-1} yr^{-1}) as constrained by the three
631 inversion experiments (2003-2022). For anthropogenic and biogenic VOC trends, months
632 dominated by biomass burning (>50% contribution) were excluded in this figure. This
633 approach ensures that the derived trends more accurately reflect the actual changes in
634 anthropogenic and biogenic VOC sources, without being biased by short-term, seasonal
635 biomass burning signals.

636
637 **Fig. 5.** Time series of biomass burning CO emissions from 2003 to 2022, comparing the a priori
638 inventory (GFED4), a posteriori inversions, and the GFED5 inventory.

639
640 **Fig. 6.** Climatological monthly cycle (2003-2022 average) of biomass burning CO emissions
641 across different regions, comparing a priori (GFED4) and a posteriori estimates with GFED5.

642
643 **Fig. 7.** Monthly trends in biomass burning CO emissions (based on Column-FixOH) and their
644 impact on column CO concentrations (2003-2022). Solid lines show trends including 2021,
645 while dashed lines exclude 2021 to illustrate the impact of extreme fire year. Please check Fig.
646 S4 and S5 for the standard deviation of the trends.

647
648 **Fig. 8.** Modeled surface CO concentrations (ppb) and their trends ($\% \text{yr}^{-1}$) from 2003 to 2022.
649 (a-c) Mean concentrations from WDCGG observations and model simulations. (d-f) Spatial
650 pattern of the long-term trend. Only stations with 14 year observations (the time range between
651 the first and last observations) during 2003-2022 are included.

652
653 **Fig. 9.** Modeled column CO concentrations (10^{18} molec cm^{-2}) and their trends ($\% \text{yr}^{-1}$) from
654 2003 to 2022. (a-c) Spatial distribution of the 20-year mean. (d-f) Spatial pattern of the long-
655 term trend.

656
657 **Fig. 10.** Attribution of trends ($\% \text{yr}^{-1}$) in surface and column CO concentrations to individual
658 drivers, derived from sensitivity simulations based on Column-FixOH inversion (2003-2022).
659 Trends are shown for scenarios with: (a, b) all sources fixed at 2003 levels; (c, d) only

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665 anthropogenic emissions varying over time; (e, f) only biomass burning emissions varying; (g,
666 h) only biogenic VOC [sources](#) varying.

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667
668 **Fig. 11.** Sensitivity of inverted CO emissions to OH fields. (a-c) Tropospheric OH columns
669 (10^{12} molec cm^{-2}) from fixed and variable (TCR-2) fields and their difference. (d-f)
670 Corresponding scaling factors from the Column-FixOH and Column-VarOH inversions and
671 their difference. Please note that due to the use of land boundary conditions, differences in OH
672 concentrations over the ocean in the left column figures have a negligible effect on the
673 differences in scaling factors shown in the right column figures.

674

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