



1 **Isotopic apportionment of sulfate aerosols between natural**
2 **and anthropogenic sources in the outflow of South Asia**

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14



15 **Abstract**

16 Sulfate aerosols cool the climate and thus temporarily mask climate warming, but at a cost to air quality. Their
17 short atmospheric lifetime leads to heterogeneous global coverage, with sulfate concentrations over South Asia
18 being especially elevated and continuing to increase. It remains challenging to constrain the relative importance
19 of different emission sources due to poor observational coverage and uncertainties in bottom-up technology-based
20 emission estimates. The stable sulfur isotope composition ($\delta^{34}\text{S-SO}_4^{2-}$) quantitatively distinguishes natural and
21 anthropogenic sources. This study aimed to constrain the sources of sulfate arriving at the Maldives Climate
22 Observatory Hanimaadhoo (MCOH), which is ideally situated for intercepting the outflow from airsheds over the
23 Indian subcontinent. The results show that anthropogenic sources of sulfate contributed $94 \pm 11\%$, $88 \pm 9\%$, and
24 $67 \pm 13\%$ in winter (post-monsoon), spring (pre-monsoon), and summer (monsoon), respectively. There was also
25 a moderate to strong correlation ($r^2 = 0.79$, $p < 0.05$, $n = 7$) between continental anthropogenic (winter and
26 spring) sulfate ($\delta^{34}\text{S}$) and black carbon aerosols from fossil fuel combustion (pinpointed by $\Delta^{14}\text{C}$). This study
27 provides improved constraints on sulfate sources for South Asia – a key region for aerosol pollution and aerosol
28 masking of climate warming.

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33 **1 Introduction**

34 Anthropogenic aerosols cause substantial net negative climate forcing, which is currently attenuating the warming
 35 caused by the emissions of greenhouse gases. Sulfate aerosols represent the largest component of this aerosol-
 36 induced masking of climate warming and the health-affecting deterioration of air quality. Sulfate aerosols are
 37 primarily secondary aerosols formed from the oxidation (via H_2O_2 , O_3 , OH, transition metals, NO_2) of sulfur
 38 dioxide, hydrogen sulfide, and sulfur-bearing organic substances, emitted from numerous natural and
 39 anthropogenic sources (e.g., Berresheim et al., 2002; Szopa et al., 2021). These aerosols then alter the climate
 40 through the direct (scattering of light) and indirect effects (primarily alteration of cloud properties), leading to a
 41 net cooling effect (e.g., Charlson et al., 1991; Szopa et al., 2021).

42 These climatic effects are associated with the ubiquitous presence of sulfate aerosols in the atmosphere, leading
 43 to an effective radiative forcing (ERF) of -0.94 Wm^{-2} [-1.63 to -0.25 Wm^{-2}] (Szopa et al., 2021). However, the
 44 aerosols' short atmospheric lifetime (1-2 weeks) is expected to lead to a rapid readjustment once emissions change.

45 There have been regional reductions in sulfate emissions in North America and Europe for several decades. While
 46 these reductions are significant environmental legislative successes, emission patterns shifted eastwards towards
 47 Asia. The rise in emissions in East Asia led to a brief increase in global emissions at the start of the 21st century
 48 (McDuffie et al., 2020). However, subsequent successful mitigation efforts primarily in China have dramatically
 49 decreased sulfate loadings in the recent decade (McDuffie et al., 2020). The emissions from India surpassed those
 50 of China at the end of the last decade (Li et al., 2017); South Asia is currently the epicenter of sulfate emissions,
 51 with emissions still believed to be on the rise (McDuffie et al., 2020).

52 South Asia experienced rapid industrialization and economic growth in the latter part of the 20th century that have
 53 continued into the present day, with the unintended consequence of high aerosol emissions. These loadings are
 54 shown by the increasing trend over the last few decades in the aerosol optical depth (AOD), with sulfate being a
 55 large contributor to this increase (Aas et al., 2019; Gupta et al., 2023). These high aerosol loadings must be
 56 addressed as they are a major health and environmental burden on South Asia, which is home to almost a quarter
 57 of the world's population (Lelieveld et al., 2020). However, our knowledge about sulfate sources and emissions
 58 remains uncertain with notable disparities between estimates from emission inventories and from remote sensing
 59 (Elguindi et al., 2020; Sharma and Kumar, 2016). The understanding of sulfate emissions is further hampered by
 60 uncertainties in natural emissions, such as oceanic dimethyl sulfide (DMS), which has seasonal and regional
 61 emission fluctuations ranging by a factor of 10-100 (Norman et al., 1999, 2004; Shenoy and Kumar, 2007). These
 62 uncertainties lead to large variations in estimates of natural versus anthropogenic contributions, especially in
 63 locations surrounded by oceans such as South Asia (Norman et al., 1999, 2004; Shenoy and Kumar, 2007).
 64 Quantitative top-down source-diagnostic isotopic composition, in combination with consideration of air-mass
 65 origins, presents an opportunity to quantitatively apportion the relative contributions from anthropogenic and
 66 natural sources of sulfate for the wider receptor atmosphere of South Asia.

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68 Isotopic composition allows for the quantitative separation of natural versus anthropogenic sources of sulfate
 69 through distinct source end-member compositions (source fingerprints). The use of $\delta^{34}\text{S}$ for separating natural vs



anthropogenic sources is established, with a few pioneering studies now also in South Asia (Dasari and Widory, 2024; Rastogi et al., 2020; Sawlani et al., 2019).

This study employs isotopic $\delta^{34}\text{S}$ source apportionment of SO_4^{2-} to quantify the anthropogenic sulfate contributions to the expansive airshed outflow from South Asia (representative of the wider system of the regional aerosol-climate effect). The present study has a much wider footprint and longer time coverage than earlier studies in the region. Distinguishing the relative source contributions of the climate-affecting sulfate in the South Asia region provides guidance for future mitigation strategies and also provides observation-based constraints useful for climate models.

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2. Methods

2.1 Site description and meteorological context

South Asia encompasses several climate zones with its climate being primarily driven by the South Asian monsoon. The monsoon season marks the onset of frequent precipitation events and the reversal of the prevailing northerly winds. The southerly monsoon winds transport air masses from the Indian Ocean to the continent. Seasonal oscillation also occurs in the zonal (east-west) wind component, with stronger easterlies during winter and autumn (from the Bay of Bengal), and more westerly flow in spring (from the Arabian Sea).

Meteorologically, this leads to air masses from the high-emission region of the Indo-Gangetic Plain (IGP) being transported into the Bay of Bengal in winter/autumn with dispersal of their anthropogenic load also over the northern Indian Ocean (see Fig. 1). The IGP is a fertile, densely populated and highly industrialized region spanning several countries. The IGP contributes the highest aerosol loading in South Asia and strongly impacts aerosol loadings far out over the Indian Ocean (Aswini et al., 2020; Budhavant et al., 2024; Nair et al., 2023; Ramanathan et al., 2001; Verma et al., 2012).

The Maldives Climate Observatory at Hanimaadhoo (MCOH; $6^\circ 46' 34''\text{N}$, $73^\circ 10' 59''\text{E}$; tower inlet at 15 m agl) is ideally situated for intercepting air masses arriving from the polluted and polluting IGP, the subcontinent at large, and additionally from the open Indian Ocean during the summer (see Fig. 1). As the IGP is the key source region of anthropogenic aerosols, two additional sites were chosen in this emission region to constrain its anthropogenic sulfate signature. These sites were the Delhi branch of the Indian Institute of Tropical Meteorology (IITM-Delhi; $28^\circ 35'\text{N}$, $77^\circ 12'\text{E}$; 15 m agl) and the Bangladesh Climate Observatory Bhola (BCOB; $22^\circ 17' 00''\text{N}$, $90^\circ 42' 36''\text{E}$, 10 m agl) (see Fig. 1). Detailed information on the observatory sites can be found in previous studies (e.g., Bikkina et al., 2019; Dasari et al., 2019).

2.2 Collection of aerosols

Samples for stable sulfur isotope measurements of sulfate in this study were hence collected from three sites in South Asia. The urban $\text{PM}_{2.5}$ filters (47 mm diameter quartz filters, Millipore) were collected from the IITM-Delhi in the winter of 2016 (January and February) using a high-volume sampler (APM 550 Envirotech, flow rate = $1 \text{ m}^3 \text{ hr}^{-1}$, time = 12 hours). The IGP outflow $\text{PM}_{2.5}$ filters were collected at BCOB in the winter of 2016 (January), using a high-volume sampler (model DH77, DIGITEL A.G., Switzerland, flow rate = 500 L min^{-1} , time = 24 hours). The integrating filters of the South Asian outflow, intercepted over the Indian Ocean, were collected at MCOH with sample dates spanning several years (winter = 2016, summer = 2013-2015, spring = 2013) (winter =



PM₁, spring and summer = PM_{2.5}) using a high-volume sampler (model DH77, DIGITEL A.G., Switzerland, flow rate = 500 L min⁻¹, time = 24 hours). Detailed information on the collection of aerosol samples can be found in previous studies (e.g., Kirillova et al., 2013, 2014).

2.3 Ionic concentrations

Ionic concentrations (Cl⁻, Br⁻, NO₂⁻, NO₃⁻, PO₄³⁻, SO₄²⁻, Na⁺, NH₄⁺, K⁺, Mg²⁺, Ca²⁺) were measured using a Dionex Aquion ion chromatography (IC) system (Thermo Finnigan LLC). Cut-outs (1-4 cm²) of the filter samples were dissolved in 10 mL of Milli-Q water and analyzed with a flow rate of 1 mL min⁻¹. Standards and field blanks were used to ensure quality control and minimize external influences.

The sea spray fraction of SO₄²⁻ was removed following (Keene et al., 1986), as shown in Eq. (1):

$$nssSO_4^{2-} = [SO_4^{2-}] - \left(\frac{[SO_4^{2-}]}{[Na^+]} \right)_{sea} \times [Na^+] \quad (1)$$

The nss-SO₄²⁻ is the non-sea salt sulfate, with the ratio of sulfate to sodium from sea water being 0.253 taken from (Keene et al., 1986). Seawater is enriched in sulfate due to long-term accumulation. This results in natural emissions of sulfate aerosols from sea spray, which must be accounted for to enable accurate source apportionment of nss-SO₄²⁻.

2.4 Isotopic analysis

2.4.1 Analysis of δ³⁴S

The determination of the δ³⁴S composition of aerosols on 17 carefully selected samples was performed following the method described in detail in Rodiouchkina, 2018. Filter cut-outs were placed in polypropylene tubes and Milli-Q water was added to extract the sulfate (5 h in an ultrasonic bath). Sulfur was then isolated using anion exchange chromatography (AG 1-X8, analytical grade, 200-400 mesh, chloride form, Bio-Rad Laboratories, USA). The standards (S1, S3, and S4 from IAEA, Austria) and samples were then diluted to an S concentration of 2 µg mL⁻¹ before being run in a solution matrix of 0.3 M HNO₃ (sub-boiled). Silicon (Si) was added to all solutions at a concentration ratio of 1:1 (S:Si, µg mL⁻¹:µg mL⁻¹) for internal standardization. Additionally, sodium (Na) was added at a molar ratio of 2 (Na/S) to all measurement solutions. The measurements were carried out using a multi-collector inductively coupled plasma mass spectrometer (MC-ICP-MS, Nu Plasma II, Nu Instruments, UK) and run-in dry plasma mode using a desolvator (Aridus II, CETAC, USA) at the Vegacenter facility of the Swedish Museum of Natural History, Stockholm, Sweden. Detailed descriptions and typical operating parameters can be found in Rodiouchkina, 2018. For instrument settings and parameters see Method S1.

2.4.2 Analysis of radiocarbon (Δ¹⁴C) composition in Black Carbon (BC)

A total of 14 samples for Δ¹⁴C were prepared for dual isotopic analysis of combustion-derived BC (MCOH spring = 2, MCOH summer = 3, BCOB = 3, MCOH winter = 6). The winter samples comprise data taken from Dasari et al., 2020. Inorganic carbonates were removed by acid fumigation (12 M HCl) for 24 h, then the samples were dried at 60 °C for 1 h. The samples were subjected to thermal-oxidation separation of OC from BC using a TOT analyzer (Thermal Optical Transmittance, Sunset Laboratory, Tigard, Oregon, USA) and the resulting carbon



dioxide derived from the BC fraction was trapped using a custom-built system described and extensively tested earlier (e.g., Andersson et al., 2015; Chen et al., 2013; Winiger et al., 2015). The trapped CO₂ from the BC fraction was then sent to the collaborating Accelerator Mass Spectrometry facility (AMS) at Uppsala University (Sweden) for $\Delta^{14}\text{C}$ isotopic analysis. The source apportionment was calculated using Eq. (2) and Eq. (3), with end-member values taken from the literature (Andersson et al., 2015; Bikkina et al., 2019).

$$\Delta^{14}\text{C}_{BC} = \left[F_m \times e^{\frac{-(x_{\text{year}} - 1954)}{8267}} - 1 \right] \times 1000 \quad (2)$$

$$\Delta^{14}\text{C}_{BC} = \Delta^{14}\text{C}_{\text{biomass}} \times F_{\text{bio-BC}} + \Delta^{14}\text{C}_{\text{fossil}} \times (1 - F_{\text{bio-BC}}) \quad (3)$$

2.5 Model for apportioning between natural and anthropogenic sulfur input

The approach to apportion natural versus anthropogenic sulfate was carried out in several steps. First, the data were corrected for the minor contribution from sea spray based on the well-known isotopic composition of sea water using Eq. (4). The sample MCOH PM₁-13-14/01/2016 was corrected using the fraction of non-sea salt contribution to tot-SO₄ from the other samples (98.9 ± 0.2 %) due to sample depletion.

$$\delta^{34}\text{S}_{\text{nss}} = (\delta^{34}\text{S} - \left(\frac{100 - \% \text{nssSO}_4}{100} \right) \times \delta^{34}\text{S}_{\text{sea water}}) \div \frac{\% \text{nssSO}_4}{100} \quad (4)$$

The isotopic composition of $\delta^{34}\text{S}_{\text{sea water}}$ is +21 ± 0.2 ‰ from Böttcher et al., 2007 and Rees et al., 1978. The percentage of nss-SO₄²⁻ is determined from Eq. (1).

After this correction, the nss- $\delta^{34}\text{S}$ was used in a binary model to apportion the relative contributions from anthropogenic-fossil sources (e.g., ship emissions/IGP sources) vs the marine biogenic (DMS) source, using Eq. (5):

$$\delta^{34}\text{S}_{\text{nss}} = \delta^{34}\text{S}_{\text{DMS}} \times F_{\text{DMS-SO}_4} + \delta^{34}\text{S}_{\text{anthropogenic}} \times (1 - F_{\text{DMS-SO}_4}) \quad (5)$$

where F represents the fraction (DMS and anthropogenic), $\delta^{34}\text{S}_{\text{nss}}$ represents the $\delta^{34}\text{S}$ of the sample, $\delta^{34}\text{S}_{\text{DMS}}$ and $\delta^{34}\text{S}_{\text{anthropogenic}}$ refer to the mean isotopic composition of the end-members (DMS and anthropogenic). The DMS end-member was taken from Amrani et al., 2013, with a $\delta^{34}\text{S}$ composition of +19.7 ± 0.5 ‰. The anthropogenic end-member varies by season and is constrained by a combination of literature reports and new findings in this study. For samples where air masses originated from the continent (spring, winter), an IGP end-member was used. The IGP end-member had a $\delta^{34}\text{S}$ of 2.3 ± 1.7 ‰ calculated from this study and literature (Dasari and Widory, 2024; Sawlani et al., 2019, Table S1). The ship end-member had a $\delta^{34}\text{S}$ of 3 ± 3 ‰ taken from studies that measured $\delta^{34}\text{S}$ -SO₄ and $\delta^{34}\text{S}$ -SO₂ in the North Atlantic, which is thought to be representative of remote air masses with strong ship emissions from heavy fuel oil.



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179 **2.6 Satellite imagery and computational analysis**

180 The Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) model was used to run 7-day backward
181 trajectories from the NOAA Air Research Laboratory (available at <http://ready.arl.noaa.gov/HYSPLIT.php>). Back
182 trajectories for seasonal clusters were calculated for summer (June–September 2015), winter (December–March
183 2015–2016) and spring (April–May 2015).

184 MERRA-2 (Modern-Era Retrospective Analysis for Research and Applications, version 2) and TROPES
185 (TROpospheric Ozone and its Precursors from Earth System Sounding) were used to retrieve model-estimated
186 surface concentrations of black carbon and sulfate. MERRA-2 is a reanalysis product that uses remote sensing
187 aerosol optical depth (satellite- and ground-based) in combination with the Goddard Earth Observing System
188 Model, Version 5 to calculate aerosol concentrations (Buchard et al., 2017; Gelaro et al., 2017; Randles et al.,
189 2017). TROPES includes additional satellite measurements (TES, AIRS, TROPOMI, and OMPS) combined
190 with a retrieval algorithm to obtain sulfate surface concentrations (Miyazaki, 2024).

191

192 **2.7 Emission inventory**

193 The Community Emissions Data System (CEDSV2021_04_21) is a bottom-up emission inventory that provides
194 gridded and national fluxes of short-lived climate pollutants (Hoesly et al., 2018). The national inventory dataset
195 is available at <https://zenodo.org/records/4741285>.



196 **3 Results and discussion**

197 **3.1 Constraints on isotopic signatures of different sources**

198 Reports on the isotopic composition of sulfate from different sources and regions display a large range, although
 199 marine natural sources are clearly enriched in ^{34}S compared to anthropogenic sources. Natural sources
 200 considered in this study were sea spray and dimethyl sulfide (DMS), the dominant marine biogenic source. Sea
 201 spray and DMS have a well-defined isotopic composition of $+21 \pm 0.2 \text{ ‰}$ and $+19.7 \pm 0.5 \text{ ‰}$, respectively
 202 (Amrani et al., 2013; Böttcher et al., 2007; Rees et al., 1978). In contrast, fossil fuel end-members can vary (-35
 203 to $+33 \text{ ‰}$) depending on fuel type (liquid vs solid), as well as having strong geographical differences
 204 (Jongebloed et al., 2023; Lee et al., 2023). Additional details on sulfate sources and considerations in sulfur-
 205 isotopic attribution studies are provided in the Supplementary Information (Text S1).

206 To account for the potential variability in fossil fuel isotope end-members, an anthropogenic end-member was
 207 determined through sampling in the regional source-integrating air masses of the IGP, with end-member
 208 uncertainty accounted for through error propagation (see Method S2). The samples were first split based on air
 209 mass origin into a continental anthropogenic end-member and an oceanic anthropogenic end-member. The
 210 oceanic anthropogenic input was expected to be predominantly from ship emissions, with a $\delta^{34}\text{S}$ signature of $3 \pm$
 211 3 ‰ taken from the literature on the marine anthropogenic end-member (Seguin et al., 2010, 2011; Wadleigh,
 212 2004, Text S1).

213 The integrated continental anthropogenic signature was calculated from the IGP samples in this study and from
 214 the literature, resulting in an end-member value of $2.3 \pm 1.7 \text{ ‰}$ (see Fig. 2, Table S1; Dasari & Widory, 2024;
 215 Sawlani et al., 2019, $n = 50$). Anthropogenic sulfate also includes minor contributions from crustal and biomass
 216 burning sources, which have not been separated in this study (Dasari and Widory, 2024). The average from this
 217 study was slightly lower due to one sample showing a near-zero $\delta^{34}\text{S}$ value (-0.07 ‰). This $\delta^{34}\text{S}$ -depleted
 218 sample likely reflects sulfate formation through oxidation catalyzed by transition metal ions (TMI) and is
 219 associated with poor air quality, reported during hazy conditions (Harris et al., 2013; Sawlani et al., 2019).
 220 Alternatively, this type of isotope depletion has been proposed to be related to increased coal input or
 221 terrigenous sulfate (Dasari and Widory, 2024). These depleted oxidative pathways/sources are likely to occur as
 222 a continuum, motivating the inclusion of the depleted sample.

223

224 **3.2 Abundance of sulfate aerosols over the northern Indian Ocean**

225 There are only limited reports of measured sulfate concentrations over the Indian Ocean, an area integral to South
 226 Asia's monsoon/hydrological cycles and aerosol-affected regional climate forcing. This reduces the confidence in
 227 our understanding of the regional loadings of sulfate and the effective regional climate forcing. The loadings for
 228 the non-sea salt sulfate (nss-SO_4^{2-}) over the northern Indian Ocean for winter, spring, and summer in this study
 229 were $10 \pm 4 \text{ } \mu\text{g m}^{-3}$ ($96 \pm 6 \text{ ‰ nss-SO}_4^{2-}$, full period $n = 43$ from Dasari et al., 2019), $5.5 \pm 1 \text{ } \mu\text{g m}^{-3}$ ($98.5 \pm 0.1 \text{ ‰}$
 230 nss-SO_4^{2-}), $1 \pm 0.2 \text{ } \mu\text{g m}^{-3}$ ($85.5 \pm 5.5 \text{ ‰ nss-SO}_4^{2-}$), respectively, with these values broadly agreeing with previous
 231 reports from the MCOH receptor of the South Asian outflow (Budhavant et al., 2023, 2024b).

232 Seasonal differences are due to both fluctuating emissions and varying seasonal air transport from the source
 233 regions (Fig. 1). These meteorological factors (wind speed, boundary layer height, humidity, etc.) cause winter
 234 conditions to be favorable for long-range aerosol transport from the polluted IGP out over the northern Indian



235 Ocean (Aswini et al., 2020; Dasari et al., 2019; Kesti et al., 2020; Ram et al., 2012). With a shortage of in situ
 236 measurements, estimates from remote sensing provide a picture of sulfate loadings for remote regions; expansion
 237 of in situ atmospheric observatory measurements is critical to the validation of the accuracy of derived estimates
 238 from remote sensing.

239 Remote sensing estimations of sulfate loadings are commonly derived from optical measurements of columnar
 240 data, which can differ greatly from ground conditions. The indirect estimates of sulfate surface loadings from
 241 remote sensing (MERRA-2, TROPES at $p = 1000$ hPa) were 2-5 times lower than the in situ sulfate measurements
 242 in this study (see Fig. S1). These discrepancies need to be confirmed by forthcoming longer and more complete
 243 in situ measurement records of sulfate in this critical region for sulfate-climate effects. Long-term in situ records
 244 remain pivotal in understanding long-term changes, which become even more valuable when combined with
 245 additional analysis such as isotopic source apportionment.

246

247 **3.3 Source-diagnostic isotopic composition of sulfate aerosols over the northern Indian Ocean**

248 Refined constraints on sulfate sources improve our understanding of the atmospheric cycle of climate-forcing
 249 anthropogenic sulfate over South Asia. The study results show that anthropogenic sulfate dominated in the South
 250 Asian outflow intercepted over the northern Indian Ocean, yet there is also substantial input of natural sulfate,
 251 especially in the monsoon period. The input of anthropogenic sources to the nss-SO_4^{2-} loading was 94 ± 11 %, 88
 252 ± 9 %, 67 ± 13 % for winter, spring (pre-monsoon), and summer (monsoon), respectively (Fig. 2; Table S1).

253 These fractions translate into anthropogenic sulfate concentrations of $0.6 \pm 0.2 \mu\text{g m}^{-3}$ (summer), $4.6 \pm 0.7 \mu\text{g m}^{-3}$
 254 (spring) and $9.1 \pm 5 \mu\text{g m}^{-3}$ (samples with $\delta^{34}\text{S}$; winter). Although the summer conditions are cleanest, the region
 255 is still highly affected by anthropogenic input, with freight ships during that time likely being an important source.
 256 Three major shipping lanes cross the northern Indian Ocean. During the study period, the global shipping
 257 emissions were nearly equal to the cumulative emissions of South Asia (McDuffie et al., 2020). These loadings
 258 are elevated given the remote location, even with respect to health guidelines for exposure to fine particulate
 259 matter.

260 The World Health Organization (WHO) annual guideline for $\text{PM}_{2.5}$ is $5 \mu\text{g m}^{-3}$ (World Health Organization, 2021),
 261 a limit which is frequently exceeded in both winter and spring even this far out over the Indian Ocean, despite
 262 local emissions being documented to be low (Budhavant et al., 2015). It is worth noting that a portion of the sulfate
 263 loading in the winter air masses from the Bay of Bengal may possibly be influenced by input from mangroves,
 264 which emit hydrogen sulfide (H_2S). This gas is rapidly oxidized and typically has a $\delta^{34}\text{S}$ below 0 ‰ (Jamieson &
 265 Wadleigh, 1999). The H_2S flux from mangroves is potentially significant with emissions estimates ranging from
 266 approximately 10 – 25 % of South Asia's anthropogenic emissions (Ganguly et al., 2018; Hoesly et al., 2018).
 267 However, analysis of the back trajectories suggests that sulfate loadings were influenced more by seasonal
 268 variation than by specific trajectory pathways (see Fig. S2). Sulfate loadings reaching MCOH are elevated and
 269 are attributed primarily to anthropogenic sources.

270 This isotope-based source apportionment of sulfate in the receptor-integrated South Asian outflow appears
 271 broadly consistent with other regional studies. A study over the northern Bay of Bengal reported $\delta^{34}\text{S} = 4.5 \pm$
 272 1.3 ‰ in PM_{10} during February–April 2013 (Rastogi et al., 2020). Applying our $\delta^{34}\text{S}$ two-end-member mixing



model to those Port Blair PM_{10} values indicates that at least ~90% of sulfate was anthropogenic. This should be viewed as a conservative lower bound because the sea-salt contribution for the isotopic data was not reported (so no sea-salt correction could be applied), and PM_{10} generally contains a larger sea-salt fraction. Similarly, the ICARB-2018 ship campaign estimated anthropogenic sulfate to make up 96 % of the $nss-SO_4^{2-}$ in the Arabian Sea and Indian Ocean (Aswini et al., 2020), using the ratio of methanesulfonic acid (MSA)/ $nss-SO_4^{2-}$. These results suggest a quite limited influence of DMS on the total sulfate mass balance over the northern Indian Ocean, even though globally DMS emissions are expected to be a quarter the size of anthropogenic emissions (Lana et al., 2011). These results provide quantitative constraints on sources, showing a strong dominance of anthropogenic sources of sulfate in the northern Indian Ocean, yet with seasonal variations.

3.4 Relationships between sulfate and black carbon

Sulfate precursors and black carbon (BC) are both emitted in large quantities from fossil fuel combustion, yet they also have other separate sources. We explored the extent of co-emission of these two short-lived climate pollutants (SLCPs), having opposing signs in their climate forcing. By comparing isotope-based source indicators ($\delta^{34}S-SO_4$ and $\Delta^{14}C-BC$), we observe a positive correlation ($r^2 = 0.79$, $p \ll 0.05$, $n = 7$) between anthropogenic sulfate and the fraction of BC attributed to fossil fuel combustion for the outflow receptor observatory at MCOH (winter and spring) (see Fig. 3 inset). Samples from BCOB, which were used to represent the anthropogenic end-member of sulfate, were assumed to be predominantly anthropogenic in origin (~100 %), although minor contributions from dust and biomass are expected (Dasari and Widory, 2024). Summer data were too limited for robust conclusions, although a similar pattern was observed, potentially reflecting differences in emission sources (see Fig. S3). The positive correlation between the isotope fingerprints of these SLCPs lends further credence to the $\delta^{34}S$ -based findings that fossil fuel combustion is indeed a major source of SO_4^{2-} over South Asia, with additional insight provided by examining the relationship between SO_4^{2-} and climate-warming BC.

Black carbon and sulfate, in part co-emitted, have opposite climatic effects, with BC enhancing climate warming while sulfate masks climate warming. Their opposing climatic effects lead to large uncertainties in radiative forcing due to the ratio of scattering vs absorbing vectors, especially in regions with high loadings of both BC and SO_4^{2-} , such as South Asia (Li et al., 2022). The BC/SO_4^{2-} ratio for the study period was 0.075 ± 0.03 (spring and winter; summer was excluded due to high variability). The BC/SO_4^{2-} ratio in the surface layer estimated from remote sensing by MERRA-2 was much higher for winter during which it was predicted to be 0.16 (2016) and for spring (April 2013) during which it was predicted to be 0.15; both were twice overestimated compared to our direct in situ measurements. The observed BC/SO_4^{2-} ratio was compared with an inventory-based BC/SO_2 ratio from the Community Emissions Data System (CEDS v2021-04-21). The BC/SO_2 ratio was obtained by aggregating BC and SO_2 fluxes over India, Pakistan, and Bangladesh. As inventories report SO_2 (not sulfate), BC/SO_2 is used as a proxy for BC/SO_4^{2-} . The ratio from CEDS was 0.097 (2015-2019 average; Hoesly et al., 2018), which is close to the BC/SO_4^{2-} ratio of 0.075 ± 0.03 constrained by in situ measurements in the current study, with MERRA-2 being higher. There is a need for increased comparison of BC/SO_4^{2-} (and their sources) simulated by models and predicted by emission inventories with greater datasets of in situ observations in South Asia, given the expected changes in their respective emissions and that reduced SO_4^{2-} emissions will unmask current climate warming for the region.



313 **4 Conclusion and outlook**

314 There are very large sulfate loadings over South Asia compared with the global average. However, there is poor
 315 spatial coverage of actual measurements of both SO_4^{2-} and its source-diagnostic isotopic composition. This study
 316 provides the first multi-seasonal isotope-based source apportionment of sulfate for the integrated outflow airshed
 317 of South Asia. The quantitative results show that anthropogenic nss-sulfate is the major contributor to sulfate
 318 loadings in this region, yet with clear seasonal and spatial variations and with some contributions also from
 319 natural-biogenic sources (e.g., from marine plankton and possibly from mangroves). Quantitative isotope-based
 320 observational constraints on the relative contributions of different sources to atmospheric sulfate are vital for both
 321 understanding total emissions of this important climate forcer and for guiding strategies to mitigate the
 322 anthropogenic components.

323 Future investigations should aim to constrain the long-term trend (decades) of wintertime and summer sources of
 324 sulfate to the South Asian airshed. This will improve the understanding of emission patterns from the heavily
 325 polluted IGP, as well as provide potential insights into atmospheric processing due to South Asia's changing
 326 meteorology. Aerosol loadings have already been shown to strongly affect the monsoon period (Krishnan et al.,
 327 2016; Ramarao et al., 2023). Moreover, both current and past emissions will be needed to understand future
 328 changes, with emissions expected to decrease due to socio-economic pressures. Thus, there is an urgent need for
 329 improved understanding of all facets of sulfate aerosols to help improve resilience against future climatic impacts,
 330 including the extent to which we may need to anticipate enhanced warming (with South Asia being particularly
 331 sensitive to heat extremes) due to the demasking of climate warming caused by decreasing sulfate loadings.

332 **Additional information**

333 Additional information can be found in the attached Supplementary Information. Data are available in the Bolin
 334 Centre Database (DOI: 10.17043/clarke-2025-v3m0nk-1).

336 **Author contributions**

337 This study was conceptualized by Ö.G. and H.H. Atmospheric sampling strategies and campaign execution were
 338 realized by K.B., H.H. and Ö.G. Laboratory procedures, data quality assessments and calculations were
 339 performed by M.R., S.C., H.H., K.R. and E.K. Interpretations and the first draft of the manuscript were
 340 primarily contributed by H.H., Ö.G., S.C. Finally, S.C. produced all display items and wrote the manuscript
 341 with all authors contributing to detailed data interpretation and writing.

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350

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359

360 **Competing interests**

361 The authors declare that they have no conflict of interest.

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 364 **Figures**
 365

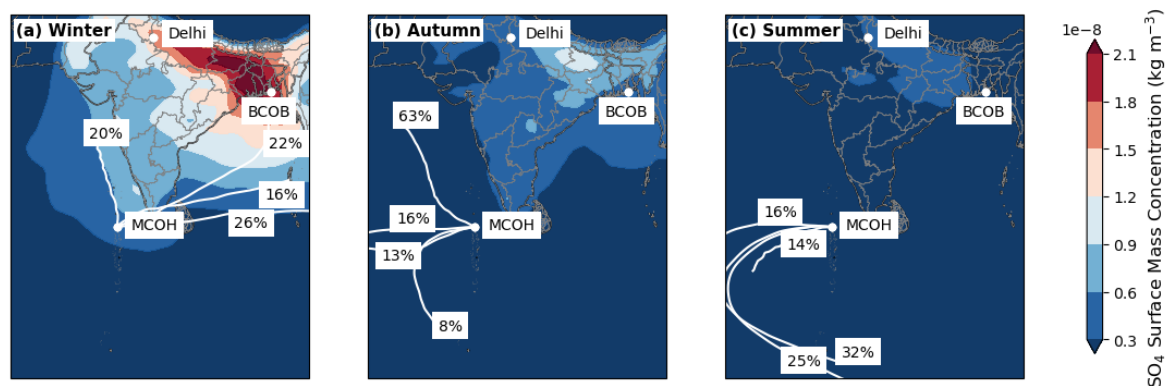
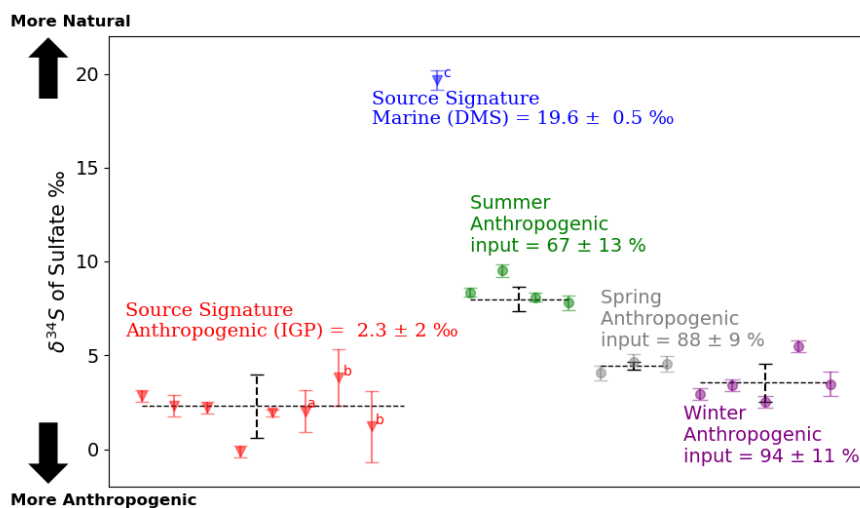
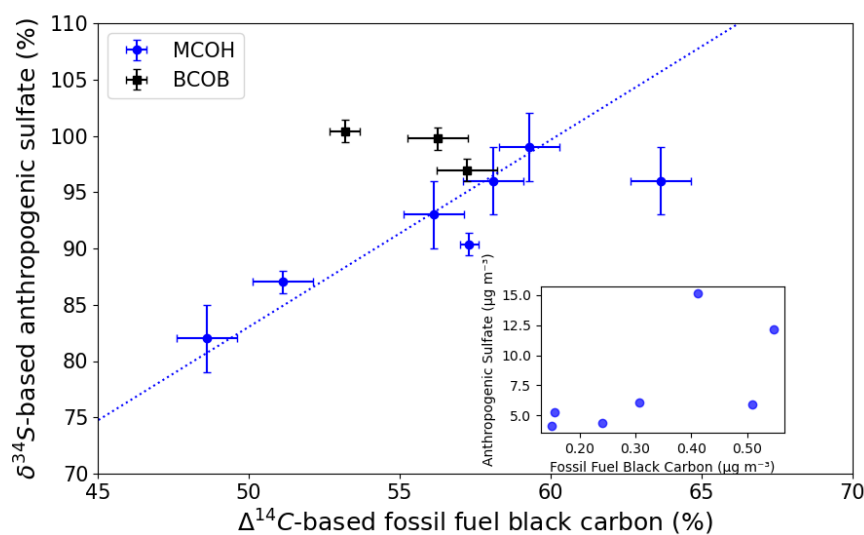


Fig. 1. Remote sensing estimates of sulfate surface concentration (MERRA-2) and average back trajectories (7-day clusters; HYSPLIT, white lines depict the mean paths of the HYSPLIT back-trajectory clusters) for South Asia, for three seasons: (a) winter (December–March 2015–2016); (b) spring (April–May 2013); (c) summer (June–September 2015).



367
 368 *Fig. 2. Sulfate $\delta^{34}\text{S}$ composition for MCOH and end-members: IGP source signature (this study, $n = 5$; a = Dasari &*
 369 *Widory, 2024, $n = 15$; b = Sawlani et al., 2019, $n = 30$), DMS source signature (blue; c = Amrani et al., 2013, $n = 16$).*
 370 *MCOH summer (green), spring (grey), and winter (purple). Error bars for this study's samples show $\pm 4\sigma$ (instrumental*
 371 *analytical precision, per measurement). Literature points and seasonal aggregates show $\pm 2\sigma$ (between-sample standard*
 372 *deviation).*

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374

375 **Fig. 3.** Relationships between fossil fuel black carbon and anthropogenic sulfate for MCOH (spring,
 376 summer; blue circles) and BCOB (black squares), calculated using isotopic composition ($\Delta^{14}\text{C}$, $\delta^{34}\text{S}$).
 377 The inset graph shows the actual concentrations of fossil fuel black carbon and anthropogenic sulfate
 378 for MCOH.

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