

1 **Isotopic apportionment of sulfate aerosols between natural** 2 **and anthropogenic sources in the outflow of South Asia**

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15 **Abstract**

16 Sulfate aerosols cool the climate and thus temporarily mask climate warming, but at a cost to air quality. Their
17 short atmospheric lifetime leads to heterogeneous global coverage, with sulfate concentrations over South Asia
18 being especially elevated and continuing to increase. It remains challenging to constrain the relative importance
19 of different emission sources due to poor observational coverage and uncertainties in bottom-up technology-based
20 emission estimates. The stable sulfur isotope composition ($\delta^{34}\text{S-SO}_4^{2-}$) quantitatively distinguishes natural and
21 anthropogenic sources. This study aimed to constrain the sources of sulfate arriving at the Maldives Climate
22 Observatory Hanimaadhoo (MCOH), which is ideally situated for intercepting the outflow from airsheds over the
23 Indian subcontinent. The results show that anthropogenic sources of sulfate contributed $93 \pm 14 \%$, $87 \pm 10 \%$,
24 and $66 \pm 12 \%$ in winter (post-monsoon), spring (pre-monsoon), and summer (monsoon), respectively. There was
25 also a moderate to strong correlation ($r^2 = 0.75$, $p \ll 0.05$, $n = 7$) between continental anthropogenic (winter and
26 spring) sulfate ($\delta^{34}\text{S}$) and black carbon aerosols from fossil fuel combustion (pinpointed by $\Delta^{14}\text{C}$). This study
27 provides improved constraints on sulfate sources for South Asia – a key region for aerosol pollution and aerosol
28 masking of climate warming.

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33 **1 Introduction**

34 Anthropogenic aerosols cause substantial net negative climate forcing, which is currently attenuating the warming
35 caused by the emissions of greenhouse gases. Sulfate aerosols represent the largest component of this aerosol-
36 induced masking of climate warming and the health-affecting deterioration of air quality. Sulfate aerosols are
37 primarily secondary aerosols formed from the oxidation (via H₂O₂, O₃, OH, transition metals, NO₂) of sulfur
38 dioxide, hydrogen sulfide, and sulfur-bearing organic substances, emitted from numerous natural and
39 anthropogenic sources (e.g., Berresheim et al., 2002; Szopa et al., 2021). These aerosols then alter the climate
40 through the direct (scattering of light) and indirect effects (primarily alteration of cloud properties), leading to a
41 net cooling effect (e.g., Charlson et al., 1991; Szopa et al., 2021).

42 These climatic effects are associated with the ubiquitous presence of sulfate aerosols in the atmosphere, leading
43 to an effective radiative forcing (ERF) of -0.94 Wm⁻² [-1.63 to -0.25 Wm⁻²] (Szopa et al., 2021). However, the
44 aerosols' short atmospheric lifetime (1-2 weeks) is expected to lead to a rapid readjustment once emissions change.

45 There have been regional reductions in sulfate emissions in North America and Europe for several decades. While
46 these reductions are significant environmental legislative successes, emission patterns shifted eastwards towards
47 Asia. The rise in emissions in East Asia led to a brief increase in global emissions at the start of the 21st century
48 (McDuffie et al., 2020). However, subsequent successful mitigation efforts primarily in China have dramatically
49 decreased sulfate loadings in the recent decade (McDuffie et al., 2020). The emissions from India surpassed those
50 of China at the end of the last decade (Li et al., 2017); South Asia is currently the epicenter of sulfate emissions,
51 with emissions still believed to be on the rise (McDuffie et al., 2020).

52 South Asia experienced rapid industrialization and economic growth in the latter part of the 20th century that have
53 continued into the present day, with the unintended consequence of high aerosol emissions. These loadings are
54 shown by the increasing trend over the last few decades in the aerosol optical depth (AOD), with sulfate being a
55 large contributor to this increase (Aas et al., 2019; Gupta et al., 2023). These high aerosol loadings must be
56 addressed as they are a major health and environmental burden on South Asia, which is home to almost a quarter
57 of the world's population (Lelieveld et al., 2020). However, our knowledge about sulfate sources and emissions
58 remains uncertain with notable disparities between estimates from emission inventories and from remote sensing
59 (Elguindi et al., 2020; Sharma and Kumar, 2016). The understanding of sulfate emissions is further hampered by
60 uncertainties in natural emissions, such as oceanic dimethyl sulfide (DMS), which has seasonal and regional
61 emission fluctuations ranging by a factor of 10-100 (Norman et al., 1999, 2004; Shenoy and Kumar, 2007). These
62 uncertainties lead to large variations in estimates of natural versus anthropogenic contributions, especially in
63 locations surrounded by oceans such as South Asia (Norman et al., 1999, 2004; Shenoy and Kumar, 2007).
64 Quantitative top-down source-diagnostic isotopic composition, in combination with consideration of air-mass
65 origins, presents an opportunity to quantitatively apportion the relative contributions from anthropogenic and
66 natural sources of sulfate for the wider receptor atmosphere of South Asia.

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68 Isotopic composition allows for the quantitative separation of natural versus anthropogenic sources of sulfate
69 through distinct source end-member compositions (source fingerprints). The use of δ³⁴S for separating natural vs

70 anthropogenic sources is established, with a few pioneering studies now also in South Asia (Dasari and Widory,
71 2024; Rastogi et al., 2020; Sawlani et al., 2019).

72 This study employs isotopic $\delta^{34}\text{S}$ source apportionment of SO_4^{2-} to quantify the anthropogenic sulfate
73 contributions to the expansive airshed outflow from South Asia (representative of the wider system of the regional
74 aerosol-climate effect). The present study has a much wider footprint and longer time coverage than earlier studies
75 in the region. Distinguishing the relative source contributions of the climate-affecting sulfate in the South Asia
76 region provides guidance for future mitigation strategies and also provides observation-based constraints useful
77 for climate models.

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79 **2. Methods**

80 **2.1 Site description and meteorological context**

81 South Asia encompasses several climate zones with its climate being primarily driven by the South Asian
82 monsoon. The monsoon season marks the onset of frequent precipitation events and the reversal of the prevailing
83 northerly winds. The southerly monsoon winds transport air masses from the Indian Ocean to the continent.
84 Seasonal oscillation also occurs in the zonal (east-west) wind component, with stronger easterlies during winter
85 and autumn (from the Bay of Bengal), and more westerly flow in spring (from the Arabian Sea).

86 Meteorologically, this leads to air masses from the high-emission region of the Indo-Gangetic Plain (IGP) being
87 transported into the Bay of Bengal in winter/autumn with dispersal of their anthropogenic load also over the
88 northern Indian Ocean (see Fig. 1). The IGP is a fertile, densely populated and highly industrialized region
89 spanning several countries. The IGP contributes the highest aerosol loading in South Asia and strongly impacts
90 aerosol loadings far out over the Indian Ocean (Aswini et al., 2020; Budhavant et al., 2024; Nair et al., 2023;
91 Ramanathan et al., 2001; Verma et al., 2012).

92 The Maldives Climate Observatory at Hanimaadhoo (MCOH; $6^\circ 46' 34''\text{N}$, $73^\circ 10' 59''\text{E}$; tower inlet at 15 m agl) is
93 ideally situated for intercepting air masses arriving from the polluted and polluting IGP, the subcontinent at large,
94 and additionally from the open Indian Ocean during the summer (see Fig. 1). As the IGP is the key source region
95 of anthropogenic aerosols, two additional sites were chosen in this emission region to constrain its anthropogenic
96 sulfate signature. These sites were the Delhi branch of the Indian Institute of Tropical Meteorology (IITM-Delhi;
97 $28^\circ 35'\text{N}$, $77^\circ 12'\text{E}$; 15 m agl) and the Bangladesh Climate Observatory Bhola (BCOB; $22^\circ 17' 00''\text{N}$, $90^\circ 42' 36''\text{E}$,
98 10 m agl) (see Fig. 1). Detailed information on the observatory sites can be found in previous studies (e.g., Bikkina
99 et al., 2019; Dasari et al., 2019).

100 **2.2 Collection of aerosols**

101 Samples for stable sulfur isotope measurements of sulfate in this study were hence collected from three sites in
102 South Asia. The urban $\text{PM}_{2.5}$ filters (47 mm diameter quartz filters, Millipore) were collected from the IITM-Delhi
103 in the winter of 2016 (January and February) using a high-volume sampler (APM 550 Envirotech, flow rate = 1
104 $\text{m}^3 \text{hr}^{-1}$, time = 12 hours). The IGP outflow $\text{PM}_{2.5}$ filters were collected at BCOB in the winter of 2016 (January),
105 using a high-volume sampler (model DH77, DIGITEL A.G., Switzerland, flow rate = 500L min^{-1} , time = 24
106 hours). The integrating filters of the South Asian outflow, intercepted over the Indian Ocean, were collected at
107 MCOH with sample dates spanning several years (winter = 2016, summer = 2013-2015, spring = 2013) (winter =

108 PM₁, spring and summer = PM_{2.5}) using a high-volume sampler (model DH77, DIGITEL A.G., Switzerland, flow
109 rate = 500 L min⁻¹, time = 24 hours). Detailed information on the collection of aerosol samples can be found in
110 previous studies (e.g., Kirillova et al., 2013, 2014).

111 **2.3 Ionic concentrations**

112 Ionic concentrations (Cl⁻, Br⁻, NO₂⁻, NO₃⁻, PO₄³⁻, SO₄²⁻, Na⁺, NH₄⁺, K⁺, Mg²⁺, Ca²⁺) were measured using a
113 Dionex Aquion ion chromatography (IC) system (Thermo Finnigan LLC, Dionex IonPac CS12A) anion (Dionex
114 IonPac AS22 fast). Cut-outs (1-4 cm²) of the filter samples were dissolved in 10 mL of Milli-Q water and analyzed
115 with a flow rate of 1 mL min⁻¹. Standards and field blanks were used to ensure quality control and minimize
116 external influences.

117 The sea spray fraction of SO₄²⁻ was removed following (Keene et al., 1986), as shown in Eq. (1):

$$118 \text{ nssSO}_4^{2-} = [\text{SO}_4^{2-}] - \left[\left(\frac{[\text{SO}_4^{2-}]}{[\text{Na}^+]} \right)_{\text{sea}} \times [\text{Na}^+] \right] \quad (1)$$

119 The nss-SO₄²⁻ is the non-sea salt sulfate, with the mass ratio of sulfate to sodium from sea water being 0.253
120 taken from (Keene et al., 1986). Seawater is enriched in sulfate due to long-term accumulation. This results in
121 natural emissions of sulfate aerosols from sea spray, which must be accounted for to enable accurate source
122 apportionment of nss-SO₄²⁻.

123

124 **2.4 Isotopic analysis**

125 **2.4.1 Analysis of δ³⁴S**

126 The determination of the δ³⁴S composition of aerosols on 17 carefully selected samples was performed following
127 the method described in detail in Rodiouchkina, 2018. Filter cut-outs were placed in polypropylene tubes and
128 Milli-Q water was added to extract the sulfate (5 h in an ultrasonic bath). Sulfur was then isolated using anion
129 exchange chromatography (AG 1-X8, analytical grade, 200-400 mesh, chloride form, Bio-Rad Laboratories,
130 USA). The standards (S1, S3, and S4 from IAEA, Austria) and samples were then diluted to an S concentration
131 of 2 µg mL⁻¹ before being run in a solution matrix of 0.3 M HNO₃ (sub-boiled). Silicon (Si) was added to all
132 solutions as ammonium hexafluorosilicat (NH₄)₂SiF₆, at a concentration ratio of 1:1 (S:Si, µg mL⁻¹:µg mL⁻¹) for
133 internal standardization. Additionally, sodium (Na) was added as sodium carbonate (Na₂CO₃) at a molar ratio of
134 2 (Na/S) to all measurement solutions. The measurements were carried out using a multi-collector inductively
135 coupled plasma mass spectrometer (MC-ICP-MS, Nu Plasma II, Nu Instruments, UK) and run-in dry plasma mode
136 using a desolvator (Aridus II, CETAC, USA) at the Vegacenter facility of the Swedish Museum of Natural
137 History, Stockholm, Sweden. Detailed descriptions and typical operating parameters can be found in
138 Rodiouchkina, 2018. For instrument settings and parameters see Method S1.

139

140 **2.4.2 Analysis of radiocarbon (Δ¹⁴C) composition in Black Carbon (BC)**

141 A total of 14 samples for Δ¹⁴C were prepared for dual isotopic analysis of combustion-derived BC (MCOH spring
142 = 2, MCOH summer = 3, BCOB = 3, MCOH winter = 6). The winter samples comprise data taken from Dasari et
143 al., 2020. Inorganic carbonates were removed by acid fumigation (12 M HCl) for 24 h, then the samples were
144 dried at 60 °C for 1 h. The samples were subjected to thermal-oxidation separation of OC from BC using a TOT

145 analyzer (Thermal Optical Transmittance, Sunset Laboratory, Tigard, Oregon, USA) and the resulting carbon
 146 dioxide derived from the BC fraction was trapped using a custom-built system described and extensively tested
 147 earlier (e.g., Andersson et al., 2015; Chen et al., 2013; Winiger et al., 2015). The trapped CO₂ from the BC fraction
 148 was then sent to the collaborating Accelerator Mass Spectrometry facility (AMS) at Uppsala University (Sweden)
 149 for $\Delta^{14}\text{C}$ isotopic analysis. The source apportionment was calculated using Eq. (2) and Eq. (3), with end-member
 150 values taken from the literature (Andersson et al., 2015; Bikkina et al., 2019).

$$151 \quad \Delta^{14}\text{C}_{BC} = \left[F_m \times e^{\frac{-(x_{year}-1954)}{8267}} - 1 \right] \times 1000 \quad (2)$$

152

$$153 \quad \Delta^{14}\text{C}_{BC} = \Delta^{14}\text{C}_{biomass} \times F_{bio-BC} + \Delta^{14}\text{C}_{fossil} \times (1 - F_{bio-BC}) \quad (3)$$

154

155 **2.5 Model for apportioning between natural and anthropogenic sulfur input**

156 The approach to apportion natural versus anthropogenic sulfate was carried out in several steps. First, the data
 157 were corrected for the minor contribution from sea spray based on the well-known isotopic composition of sea
 158 water using Eq. (4). The sample MCOH PM₁-13-14/01/2016 was corrected using the fraction of non-sea salt
 159 contribution to tot-SO₄ from the other samples (98.9 ± 0.2 %) due to sample depletion.

$$160 \quad \delta^{34}\text{S}_{nss} = \left(\delta^{34}\text{S} - \left(\frac{100 - F_{nssSO_4}}{100} \right) \times \delta^{34}\text{S}_{sea\ water} \right) \div \frac{\%nssSO_4}{100} \quad (4)$$

161 The isotopic composition of $\delta^{34}\text{S}_{sea\ water}$ is +21 ± 0.2 ‰ from Böttcher et al., 2007 and Rees et al., 1978. The
 162 percentage of nss-SO₄²⁻ is determined from Eq. (1).

163 After this correction, the nss- $\delta^{34}\text{S}$ was used in a binary model to apportion the relative contributions from
 164 anthropogenic-fossil sources (e.g., ship emissions/IGP sources) vs the marine biogenic (DMS) source, using Eq.
 165 (5):

166

$$167 \quad \delta^{34}\text{S}_{nss} = \delta^{34}\text{S}_{DMS} \times F_{DMS-SO_4} + \delta^{34}\text{S}_{anthropogenic} \times (1 - F_{DMS-SO_4}) \quad (5)$$

168 where F represents the fraction (DMS and anthropogenic), $\delta^{34}\text{S}_{nss}$ represents the $\delta^{34}\text{S}$ of the sample, $\delta^{34}\text{S}_{DMS}$
 169 and $\delta^{34}\text{S}_{anthropogenic}$ refer to the mean isotopic composition of the end-members (DMS and anthropogenic). The
 170 DMS end-member was taken from Amrani et al., 2013, with a $\delta^{34}\text{S}$ composition of +18.8 ± 0.5 ‰. The
 171 anthropogenic end-member varies by season and is constrained by a combination of literature reports and new
 172 findings in this study. For samples where air masses originated from the continent (spring, winter), an IGP end-
 173 member was used. The IGP end-member had a $\delta^{34}\text{S}$ of 2.3 ± 1.7 ‰ calculated from this study and literature (Dasari
 174 and Widory, 2024; Sawlani et al., 2019, Table S1). The ship end-member had a $\delta^{34}\text{S}$ of 3 ± 3 ‰ taken from studies
 175 that measured $\delta^{34}\text{-SO}_4$ and $\delta^{34}\text{-SO}_2$ in the North Atlantic, which is thought to be representative of remote air
 176 masses with strong ship emissions from heavy fuel oil (Seguin et al., 2010, 2011; Wadleigh, 2004).

177

178

179 **2.6 Satellite imagery and computational analysis**

180 The Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) model was used to run 7-day backward
181 trajectories from the NOAA Air Research Laboratory (available at <http://ready.arl.noaa.gov/HYSPLIT.php>). Back
182 trajectories for seasonal clusters were calculated for summer (June–September 2015), winter (December–March
183 2015–2016) and spring (April–May 2015).

184 MERRA-2 (Modern-Era Retrospective Analysis for Research and Applications, version 2) and TROPES
185 (TROpospheric Ozone and its Precursors from Earth System Sounding) were used to retrieve model-estimated
186 surface concentrations of black carbon and sulfate. MERRA-2 is a reanalysis product that uses remote sensing
187 aerosol optical depth (satellite- and ground-based) in combination with the Goddard Earth Observing System
188 Model, Version 5 to calculate aerosol concentrations (Buchard et al., 2017; Gelaro et al., 2017; Randles et al.,
189 2017). TROPES includes additional satellite measurements (TES, AIRS, TROPOMI, and OMPS) combined
190 with a retrieval algorithm to obtain sulfate surface concentrations (Miyazaki, 2024).

191

192 **2.7 Emission inventory**

193 The Community Emissions Data System (CEDSV2021_04_21) is a bottom-up emission inventory that provides
194 gridded and national fluxes of short-lived climate pollutants (Hoesly et al., 2018). The national inventory dataset
195 is available at <https://zenodo.org/records/4741285>.

196 **3 Results and discussion**

197 **3.1 Constraints on isotopic signatures of different sources**

198 Reports on the isotopic composition of sulfate from different sources and regions display a large range, although
199 marine natural sources are clearly enriched in ^{34}S compared to anthropogenic sources. Natural sources considered
200 in this study were sea spray and dimethyl sulfide (DMS), the dominant marine biogenic source. Sea spray and
201 DMS have a well-defined isotopic composition of $+21 \pm 0.2 \text{ ‰}$ and $+19.7 \pm 0.5 \text{ ‰}$, respectively (Amrani et al.,
202 2013; Böttcher et al., 2007; Rees et al., 1978). In contrast, fossil fuel end-members can vary (-35 to $+33 \text{ ‰}$)
203 depending on fuel type (liquid vs solid), as well as having strong geographical differences (Jongebloed et al.,
204 2023; Lee et al., 2023). Additional details on sulfate sources and considerations in sulfur-isotopic attribution
205 studies are provided in the Supplementary Information (Text S1).

206 To account for the potential variability in fossil fuel isotope end-members, an anthropogenic end-member was
207 determined through sampling in the regional source-integrating air masses of the IGP, with end-member
208 uncertainty accounted for through error propagation (see Method S2). The samples were first split based on air
209 mass origin into a continental anthropogenic end-member and an oceanic anthropogenic end-member. The oceanic
210 anthropogenic input was expected to be predominantly from ship emissions, with a $\delta^{34}\text{S}$ signature of $3 \pm 3 \text{ ‰}$
211 taken from the literature on the marine anthropogenic end-member (Seguin et al., 2010, 2011; Wadleigh, 2004,
212 Text S1). The oceanic anthropogenic and continental anthropogenic end-members were treated as distinct to
213 reflect their different origins. Ship emissions may contribute to the continental end-member, but available top-
214 down (MERRA-2) and bottom-up (CEDS) constraints indicate that these contributions are much smaller than
215 land-based continental emissions (Buchard et al., 2017; Randles et al., 2017; Hoesly et al., 2018). In any case, the
216 choice of end-member changes the inferred anthropogenic contribution by at most $\sim 3\%$.

217 The integrated continental anthropogenic signature was calculated directly from $\delta^{34}\text{S}$ (SO_4^{2-}) measured on aerosol
218 filters in the IGP (Delhi and BCOB), with samples measured in this study and from the literature, resulting in an
219 end-member value of $2.3 \pm 1.7 \text{ ‰}$ (see Fig. 2, Table S1; Dasari & Widory, 2024; Sawlani et al., 2019, $n = 50$)
220 (see Fig. 2, Table S1; Dasari & Widory, 2024; Sawlani et al., 2019, $n = 50$). Analysing source aerosols captures
221 isotope fractionation during SO_2 oxidation to sulfate, which can be substantial ($\approx 3\text{ ‰}$ enrichment in urban
222 environments, See SI Text S1; Lee et al., 2023). The IGP end-member in this study was assumed to be
223 predominantly anthropogenic. Potential crustal and biogenic contributions are expected to be minor (Dasari &
224 Widory, 2024). Our sensitivity analysis shows that including a 5–10% non-anthropogenic sulfate fraction does
225 not change the source apportionment beyond the propagated uncertainty (see SI Text S3). We also investigated
226 the use of a Keeling plot to determine the anthropogenic end-member, but did not apply this approach because
227 mixed and variable sources could bias the inferred end-member toward depleted values (see SI Text 3 and SI Fig.
228 3). The average from this study was slightly lower due to one sample showing a near-zero $\delta^{34}\text{S}$ value (-0.07 ‰).
229 This $\delta^{34}\text{S}$ -depleted sample likely reflects sulfate formation through oxidation catalyzed by transition metal ions
230 (TMI) and is associated with poor air quality, reported during hazy conditions (Harris et al., 2013; Sawlani et al.,
231 2019). Alternatively, this type of isotope depletion has been proposed to be related to increased coal input or
232 terrigenous sulfate (Dasari and Widory, 2024). These depleted oxidative pathways/sources are likely to occur as
233 a continuum, motivating the inclusion of the depleted sample.

234

235 **3.2 Abundance of sulfate aerosols over the northern Indian Ocean**

236 There are only limited reports of measured sulfate concentrations over the Indian Ocean, an area integral to South
237 Asia's monsoon/hydrological cycles and aerosol-affected regional climate forcing. This reduces the confidence in
238 our understanding of the regional loadings of sulfate and the effective regional climate forcing. The loadings for
239 the non-sea salt sulfate (nss-SO₄²⁻) over the northern Indian Ocean for winter, spring, and summer in this study
240 were $10 \pm 4 \mu\text{g m}^{-3}$ ($96 \pm 6 \%$ nss-SO₄²⁻, full period $n = 43$ from Dasari et al., 2019), $5.5 \pm 1 \mu\text{g m}^{-3}$ ($98.5 \pm 0.1 \%$
241 nss-SO₄²⁻), $1 \pm 0.2 \mu\text{g m}^{-3}$ ($85.5 \pm 5.5 \%$ nss-SO₄²⁻), respectively, with these values broadly agreeing with previous
242 reports from the MCOH receptor of the South Asian outflow (Budhavant et al., 2023, 2024b).

243 Seasonal differences are due to both fluctuating emissions and varying seasonal air transport from the source
244 regions (Fig. 1). These meteorological factors (wind speed, boundary layer height, humidity, etc.) cause winter
245 conditions to be favorable for long-range aerosol transport from the polluted IGP out over the northern Indian
246 Ocean (Aswini et al., 2020; Dasari et al., 2019; Kesti et al., 2020; Ram et al., 2012). With a shortage of in situ
247 measurements, estimates from remote sensing provide a picture of sulfate loadings for remote regions; expansion
248 of in situ atmospheric observatory measurements is critical to the validation of the accuracy of derived estimates
249 from remote sensing.

250 Remote sensing estimations of sulfate loadings are commonly derived from optical measurements of columnar
251 data, which can differ greatly from ground conditions. The indirect estimates of sulfate surface loadings from
252 remote sensing (MERRA-2, TROPES at $p = 1000$ hPa) were 2-5 times lower than the in situ sulfate measurements
253 in this study (see Fig. S1). These discrepancies need to be confirmed by forthcoming longer and more complete
254 in situ measurement records of sulfate in this critical region for sulfate-climate effects. Long-term in situ records
255 remain pivotal in understanding long-term changes, which become even more valuable when combined with
256 additional analysis such as isotopic source apportionment.

257

258 **3.3 Source-diagnostic isotopic composition of sulfate aerosols over the northern Indian Ocean**

259 Refined constraints on sulfate sources improve our understanding of the atmospheric cycle of climate-forcing
260 anthropogenic sulfate over South Asia. The study results show that anthropogenic sulfate dominated in the South
261 Asian outflow intercepted over the northern Indian Ocean, yet there is also substantial input of natural sulfate,
262 especially in the monsoon period. The input of anthropogenic sources to the nss-SO₄²⁻ loading was $93 \pm 12 \%$, 87
263 $\pm 10 \%$, $66 \pm 14 \%$ for winter, spring (pre-monsoon), and summer (monsoon), respectively (Fig. 2; Table S1).

264 These fractions translate into anthropogenic sulfate concentrations of $0.6 \pm 0.2 \mu\text{g m}^{-3}$ (summer), $4.6 \pm 0.7 \mu\text{g m}^{-3}$
265 (spring) and $9.1 \pm 5 \mu\text{g m}^{-3}$ (samples with $\delta^{34}\text{S}$; winter). Although the summer conditions are cleanest, the region
266 is still highly affected by anthropogenic input, with freight ships during that time likely being an important source.
267 Three major shipping lanes cross the northern Indian Ocean. During the study period, the global shipping
268 emissions were nearly equal to the cumulative emissions of South Asia (McDuffie et al., 2020). These loadings
269 are elevated given the remote location, even with respect to health guidelines for exposure to fine particulate
270 matter.

271 The World Health Organization (WHO) annual guideline for PM_{2.5} is $5 \mu\text{g m}^{-3}$ (World Health Organization, 2021),
272 a limit which is frequently exceeded in both winter and spring even this far out over the Indian Ocean, despite

273 local emissions being documented to be low (Budhavant et al., 2015). It is worth noting that a portion of the sulfate
274 loading in the winter air masses from the Bay of Bengal may possibly be influenced by input from mangroves,
275 which emit hydrogen sulfide (H₂S). This gas is rapidly oxidized and typically has a δ³⁴S below 0 ‰ (Jamieson &
276 Wadleigh, 1999). The H₂S flux from mangroves is potentially significant with emissions estimates ranging from
277 approximately 10 – 25 % of South Asia’s anthropogenic emissions (Ganguly et al., 2018; Hoesly et al., 2018).
278 However, analysis of the back trajectories suggests that sulfate loadings were influenced more by seasonal
279 variation than by specific trajectory pathways (see Fig. S2). Sulfate loadings reaching MCOH are elevated and
280 are attributed primarily to anthropogenic sources.

281 This isotope-based source apportionment of sulfate in the receptor-integrated South Asian outflow appears
282 broadly consistent with other regional studies. A study over the northern Bay of Bengal reported δ³⁴S = 4.5 ±
283 1.3‰ in PM₁₀ during February–April 2013 (Rastogi et al., 2020). Applying our δ³⁴S two-end-member mixing
284 model to those Port Blair PM₁₀ values indicates that at least ~90% of sulfate was anthropogenic. This should be
285 viewed as a conservative lower bound because the sea-salt contribution for the isotopic data was not reported (so
286 no sea-salt correction could be applied), and PM₁₀ generally contains a larger sea-salt fraction. Similarly, the
287 ICARB-2018 ship campaign estimated anthropogenic sulfate to make up 96 % of the nss-SO₄²⁻ in the Arabian Sea
288 and Indian Ocean (Aswini et al., 2020), using the ratio of methanesulfonic acid (MSA)/nss-SO₄²⁻. These results
289 suggest a quite limited influence of DMS on the total sulfate mass balance over the northern Indian Ocean, even
290 though globally DMS emissions are expected to be a quarter the size of anthropogenic emissions (Lana et al.,
291 2011). These results provide quantitative constraints on sources, showing a strong dominance of anthropogenic
292 sources of sulfate in the northern Indian Ocean, yet with seasonal variations.

293

294 **3.4 Relationships between sulfate and black carbon**

295 Sulfate precursors and black carbon (BC) are both emitted in large quantities from fossil fuel combustion, yet they
296 also have other separate sources. We explored the extent of co-emission of these two short-lived climate pollutants
297 (SLCPs), having opposing signs in their climate forcing. By comparing isotope-based source indicators (δ³⁴S–SO₄²⁻
298 and Δ¹⁴C–BC), we observe a positive correlation ($r^2 = 0.75$, $p \ll 0.05$, $n = 7$) between anthropogenic sulfate and
299 the fraction of BC attributed to fossil fuel combustion for the outflow receptor observatory at MCOH (winter and
300 spring) (see Fig. 3 inset). Samples from BCOB, which were used to represent the anthropogenic end-member of
301 sulfate, were assumed to be predominantly anthropogenic in origin (~100 %), although minor contributions from
302 dust and biomass are expected (Dasari and Widory, 2024). Summer data were too limited for robust conclusions,
303 although a similar pattern was observed, potentially reflecting differences in emission sources (see Fig. S3). The
304 positive correlation between the isotope fingerprints of these SLCPs lends further credence to the δ³⁴S-based
305 findings that fossil fuel combustion is indeed a major source of SO₄²⁻ over South Asia, with additional insight
306 provided by examining the relationship between SO₄²⁻ and climate-warming BC.

307 Black carbon and sulfate, in part co-emitted, have opposite climatic effects, with BC enhancing climate warming
308 while sulfate masks climate warming. Their opposing climatic effects lead to large uncertainties in radiative
309 forcing due to the ratio of scattering vs absorbing vectors, especially in regions with high loadings of both BC and
310 SO₄²⁻, such as South Asia (Li et al., 2022). The BC/SO₄²⁻ ratio for the study period was 0.075 ± 0.03 (spring and
311 winter; summer was excluded due to high variability). The BC/SO₄²⁻ ratio in the surface layer estimated from

312 remote sensing by MERRA-2 was much higher for winter during which it was predicted to be 0.16 (2016) and for
313 spring (April 2013) during which it was predicted to be 0.15; both were twice overestimated compared to our
314 direct in situ measurements. The observed BC/SO₄²⁻ ratio was compared with an inventory-based BC/SO₂ ratio
315 from the Community Emissions Data System (CEDS v2021-04-21). The BC/SO₂ ratio was obtained by
316 aggregating BC and SO₂ fluxes over India, Pakistan, and Bangladesh. As inventories report SO₂ (not sulfate),
317 BC/SO₂ is used as a proxy for BC/SO₄²⁻. The ratio from CEDS was 0.097 (2015-2019 average; Hoesly et al.,
318 2018), which is close to the BC/SO₄²⁻ ratio of 0.075 ± 0.03 constrained by in situ measurements in the current
319 study. When we instead compare to the in situ anthropogenic sulfate fraction (BC/anthro-SO₄²⁻), the agreement
320 improves (0.082 ± 0.03; spring and winter). The uncertainty reflects measurement variability only and does not
321 include additional uncertainty propagated from the δ³⁴S end-member constraints. Deviation between emission
322 inventory and in-situ may reflect the unaccounted biogenic sulfur inputs (lowering BC/SO₄²⁻), while preferential
323 removal of hygroscopic sulfate during transport would act in the opposite direction by increasing BC/SO₄²⁻ (e.g.,
324 Budhavant et al., 2020). There is a need for increased comparison of BC/SO₄²⁻ (and their sources) simulated by
325 models and predicted by emission inventories with greater datasets of in situ observations in South Asia, given
326 the expected changes in their respective emissions and that reduced SO₄²⁻ emissions will unmask current climate
327 warming for the region.

328

329 **4 Conclusion and outlook**

330 There are very large sulfate loadings over South Asia compared with the global average. However, there is poor
331 spatial coverage of actual measurements of both SO_4^{2-} and its source-diagnostic isotopic composition. This study
332 provides the first multi-seasonal isotope-based source apportionment of sulfate for the integrated outflow airshed
333 of South Asia. The quantitative results show that anthropogenic nss-sulfate is the major contributor to sulfate
334 loadings in this region, yet with clear seasonal and spatial variations and with some contributions also from
335 natural-biogenic sources (e.g., from marine plankton and possibly from mangroves). Quantitative isotope-based
336 observational constraints on the relative contributions of different sources to atmospheric sulfate are vital for both
337 understanding total emissions of this important climate forcer and for guiding strategies to mitigate the
338 anthropogenic components.

339 Future investigations should aim to constrain the long-term trend (decades) of wintertime and summer sources of
340 sulfate to the South Asian airshed. This will improve the understanding of emission patterns from the heavily
341 polluted IGP, as well as provide potential insights into atmospheric processing due to South Asia's changing
342 meteorology. Aerosol loadings have already been shown to strongly affect the monsoon period (Krishnan et al.,
343 2016; Ramarao et al., 2023). Moreover, both current and past emissions will be needed to understand future
344 changes, with emissions expected to decrease due to socio-economic pressures. Thus, there is an urgent need for
345 improved understanding of all facets of sulfate aerosols to help improve resilience against future climatic impacts,
346 including the extent to which we may need to anticipate enhanced warming (with South Asia being particularly
347 sensitive to heat extremes) due to the demasking of climate warming caused by decreasing sulfate loadings.

348 **Additional information**

349 Additional information can be found in the attached Supplementary Information.

350

351 **Data availability**

352 The datasets generated and analyzed in this study can be accessed from the following link:
353 <https://doi.org/10.17043/clarke-2026-sulfate-aerosols-1>. The authors are willing to provide additional
354 information for data that may be of interest to readers upon request.

355

356 **Author contributions**

357 This study was conceptualized by Ö.G. and H.H. Atmospheric sampling strategies and campaign execution were
358 realized by K.B., H.H. and Ö.G. Laboratory procedures, data quality assessments and calculations were
359 performed by M.R., S.C., H.H., K.R. and E.K. Interpretations and the first draft of the manuscript were
360 primarily contributed by H.H., Ö.G., S.C. Finally, S.C. produced all display items and wrote the manuscript
361 with all authors contributing to detailed data interpretation and writing.

362

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379

380 **Competing interests**

381 The authors declare that they have no conflict of interest.

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383
384 **Figures**
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386

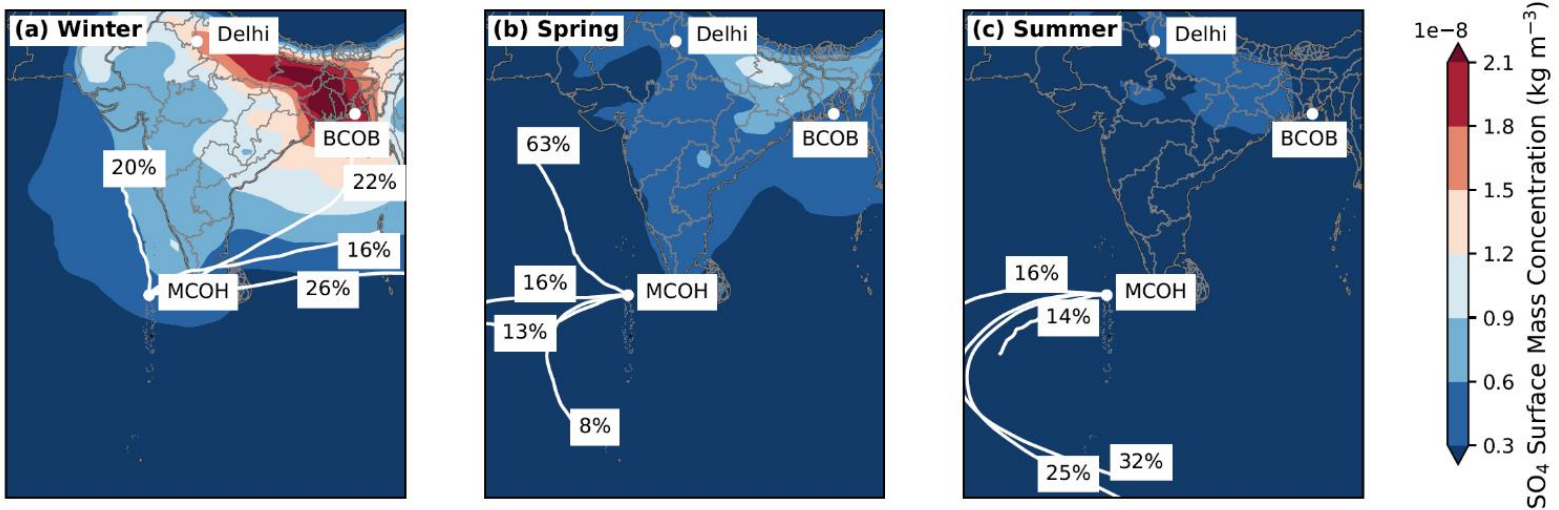
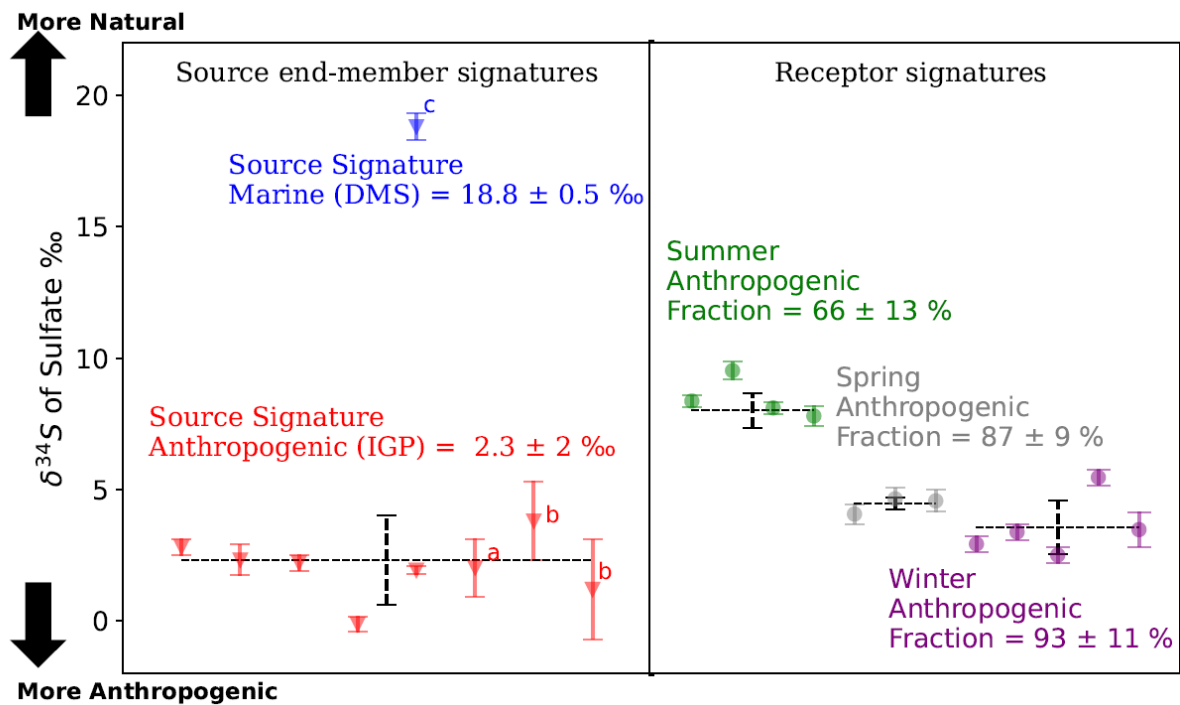


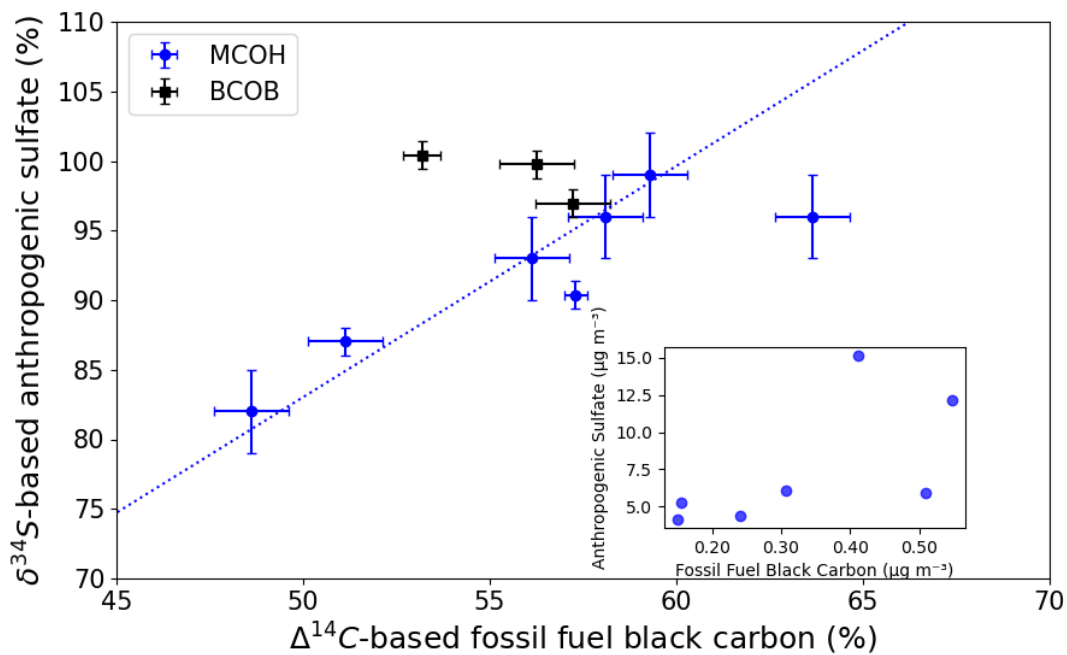
Fig. 1. Remote sensing estimates of sulfate surface concentration (MERRA-2) and average back trajectories (7-day clusters; HYSPLIT, white lines depict the mean paths of the HYSPLIT back-trajectory clusters) for South Asia, for three seasons: (a) winter (December–March 2015–2016); (b) spring (April–May 2013); (c) summer (June–September 2015).



387

388 *Fig. 2. Sulfate $\delta^{34}\text{S}$ composition for MCOH and end-members: IGP source signature (this study, $n = 5$; a = Dasari &*
 389 *Widory, 2024, $n = 15$; b = Sawlani et al., 2019, $n = 30$), DMS source signature (blue; c = Amrani et al., 2013, $n = 16$).*
 390 *MCOH summer (green), spring (grey), and winter (purple). Error bars for this study's samples show $\pm 4\sigma$ (instrumental*
 391 *analytical precision, per measurement). Literature points and seasonal aggregates show $\pm 2\sigma$ (between-sample standard*
 392 *deviation).*

393



394

395 **Fig. 3.** Relationships between fossil fuel black carbon and anthropogenic sulfate for MCOH (spring,
 396 summer; blue circles) and BCOB (black squares), calculated using isotopic composition ($\Delta^{14}\text{C}$, $\delta^{34}\text{S}$).
 397 The inset graph shows the actual concentrations of fossil fuel black carbon and anthropogenic sulfate
 398 for MCOH.

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