



1 **Ocean acidification alters phytoplankton diversity and community structure in the coastal**
 2 **water of the East China Sea**

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14 **Keywords:** CO₂, East China Sea, mesocosms; ocean acidification; phytoplankton; primary
 15 production

16 **Abstract**

17 Anthropogenic CO₂ emissions and their continuous dissolution into seawater lead to seawater
 18 pCO₂ rise and ocean acidification (OA). Phytoplankton groups are known to be differentially
 19 affected by carbonate chemistry changes associated with OA in different regions of contrasting
 20 physical and chemical features. To explore responses of phytoplankton to OA in the Chinese coastal
 21 waters, we conducted a mesocosm experiment in a eutrophic bay of the southern East China Sea
 22 under ambient (410 µatm, AC) and elevated (1000 µatm, HC) pCO₂ levels. The HC stimulated
 23 phytoplankton growth and primary production during the initial nutrient-replete stage, while the
 24 community diversity and evenness were reduced during this stage due to the rapid nutrient
 25 consumption and diatom blooms, and the subsequent shift from diatoms to hetero-dinoflagellates
 26 led to a decline in primary production during the mid and later phases under nutrient depletion. Such
 27 suppression of diatom-to-dinoflagellate succession occurred with enhanced remineralization of
 28 organic matter under the HC conditions, with smaller phytoplankton becoming dominant for the
 29 sustained primary production. Our findings indicate that, the impacts of OA on phytoplankton



30 diversity in the coastal water of the southern East China Sea depend on availability of nutrients,
 31 with primary productivity and biodiversity of phytoplankton reduced in the eutrophicated coastal
 32 water.

33 **1 Introduction**

34 It is commonly known that sequestration of CO₂ in coastal waters play important roles against
 35 global warming due to their high primary productivity (Rogelj et al., 2022), which resulted in faster
 36 CO₂ removal due to photosynthesis than dissolution of CO₂ from the air (Stukel et al., 2023). It has
 37 been assessed that, with the increasing anthropogenic CO₂ emissions, the oceanic CO₂ sink
 38 increased from 1.7 ± 0.4 pg C yr⁻¹ in the 1980s to 2.5 ± 0.6 pg C yr⁻¹ in the 2010s (Friedlingstein
 39 et al., 2025). Nevertheless, such apparent oceanic CO₂ uptake is altering carbonate chemistry in
 40 surface oceans, leading to a pH drop of by 0.017–0.027 units per decade, with a potential further
 41 drop by 0.3–0.4 units at the end of this century (Canadell et al., 2023; Gattuso et al., 2015). Such
 42 progressive ocean acidification (OA) has been shown to impact many marine organisms (Gattuso et
 43 al., 2015), including primary producers (Gao et al., 2020), subsequently feeding back on the CO₂
 44 sequestration efficiency in marine systems including coastal waters.

45 OA in eutrophic coastal waters are suggested to progress faster than in open oceans by roughly
 46 20 % due to CO₂ dissolution and enhanced remineralization of organic matters (Cai et al., 2011). The
 47 subsequent changes in carbonate chemistry may thus drive shifts in phytoplankton community
 48 structures/diversity and affect primary productions in differential ways due to regional
 49 environmental traits and species-specific physiology (Feng et al., 2024; Gao et al., 2012). While the
 50 effects of elevated *p*CO₂ on different phytoplankton assemblages have been demonstrated, positive,
 51 neutral and negative effects have been reported, reflecting differences in experimental approaches
 52 and/or phytoplankton compositions (Gao et al., 2020). Among the different approaches, field
 53 mesocosm experiments under elevated *p*CO₂ projected for future OA scenario have been employed
 54 to investigate the effects of OA on ecological processes, including primary production. Previous
 55 mesocosm experiments showed reduced growth of coccolithophores species under 710 µatm *p*CO₂
 56 during early summer in 2001 (Engel et al., 2005) and the loss of the ability to form blooms under
 57 1000–3000 µatm *p*CO₂ during early summer in 2011 (Riebesell et al., 2017). Under elevated *p*CO₂,
 58 the phytoplankton communities in Norwegian coastal mesocosms shifted from *Bathycoccus* to



59 *Micromonas* (Meakin and Wyman, 2011). By contrast, another mesocosm experiment carried out in
 60 northeast Atlantic showing that diatoms were insensitive to OA under oligotrophic conditions, but
 61 were positively affected under nutrient enrichment (Bach et al., 2019). Previously, we showed, by
 62 running a mesocosm experiment during spring of 2018 in the southern coastal water of the East
 63 China Sea, that elevated $p\text{CO}_2$ of 1000 μatm suppressed the succession from diatoms to
 64 dinoflagellates and increased the abundance of viruses and heterotrophic bacteria, thereby
 65 promoting refueling of nutrients for phytoplankton growth (Huang et al., 2021). These results
 66 indicate that the effects of OA on community structure can vary temporally and spatially. On the
 67 other hand, mesocosm experiments conducted in oligotrophic or mesotrophic regions showed that
 68 nutrients enrichment increased Chl *a* concentration under high $p\text{CO}_2$ condition (Riebesell et al.,
 69 2017; Tanaka et al., 2013; Schulz et al., 2013), however, mesocosm experiments conducted in highly
 70 eutrophic water showed that high $p\text{CO}_2$ did not affect Chl *a* concentration (Liu et al., 2017; Huang
 71 et al., 2021). Plankton communities supported by remineralized nutrients were more sensitive to OA
 72 than those having access to higher availability of inorganic nutrients (Bach et al., 2016; Bach et al.,
 73 2019). It is likely that availabilities or levels of eutrophication can modulate the effects of OA,
 74 alongside regional chemical and physical differences (Paul and Bach, 2020).

75 In coastal regions, changes in seawater carbonate chemistry influence primary production
 76 which in reverse affect the pH change due to faster photosynthetic CO_2 removal than its dissolution
 77 (Gao, 2021), resulting in increased pH during daytime and declined pH during nighttime. Such large
 78 diel pH fluctuation may require phytoplankton to invest more energy to maintain cellular
 79 homeostasis in response to the negative effects of increased hydrogen ion concentration (the acidic
 80 stress), thereby affecting the functioning of planktonic ecosystem (Raven and Beardall, 2020;
 81 Rokitta et al., 2012; Taylor et al., 2017). The documented positive effects of increased inorganic
 82 carbon substrates (e.g., CO_2 and HCO_3^-) and the negative effect of increased H^+ concentration may
 83 shape the responses of coastal ecosystems to OA, leading to controversial results (Wu et al., 2017;
 84 Vázquez et al., 2023; Chauhan et al., 2024). To better understand the consequences of OA in Chinese
 85 eutrophic coastal regions, we conducted a mesocosm experiment in the eutrophic coastal water of
 86 Wuyuan Bay, Xiamen, China, using *in situ* plankton communities during October-December, 2019
 87 and investigated how OA shapes the diversity of phytoplankton community and affects primary



88 production processes. Our results show that in the eutrophic coastal water of East China Sea, with
 89 the natural decrease of temperature, elevated $p\text{CO}_2$ increased primary production by promoting the
 90 phytoplankton biomass (indicated by Chl *a* concentration) under nutrient-replete condition, and
 91 promoted smaller phytoplankton's growth to sustain the primary production after the nutrient
 92 depletion and diatom bloom collapsed, though suppressed the emergence of dinoflagellates.

93 **2 Materials and methods**

94 **2.1 Mesocosms setup and sampling**

95 The *in situ* mesocosm experiment was conducted at the Facility for the Study of Ocean
 96 Acidification Impacts of Xiamen University (FOANIC-XMU) located in the subtropical coastal
 97 region Wuyuan Bay of southern East China Sea (24.52°N, 117.18°E) from 9th October (day 0
 98 relative to algal inoculation) to 14th November, 2019. Nine cylindrical and transparent
 99 thermoplastic polyurethane (TPU) mesocosm bags, each 3 m deep and 1.5 m in diameter, were filled
 100 with approximately 3000 L of *in situ* seawater that had been prefiltered (MU801-4T, Midea, China,
 101 pore size of 0.01 μm). The mesocosms were hooked to and secured within steel frames. Two $p\text{CO}_2$
 102 treatments were established to investigate the effects of ocean acidification on the *in situ*
 103 phytoplankton community: an ambient $p\text{CO}_2$ treatment (AC, 410 μatm ; 4 numbered bags) and a
 104 high $p\text{CO}_2$ treatment representative of end-of-century projections (HC, 1,000 μatm ; 5 numbered
 105 bags). To adjust CO_2 in seawater in the HC bags to the projected 1000 μatm in 2100s, approximately
 106 11 L of CO_2 -saturated seawater was added to each HC bag. The AC and HC $p\text{CO}_2$ levels were
 107 maintained by bubbling with ambient air and pre-mixed air- CO_2 (1000 $\mu\text{atm CO}_2$), respectively, at
 108 a rate of 5 L min^{-1} using a CO_2 Enricher (CE-100B, Wuhan Ruihua Instrument & Equipment Ltd,
 109 China). After the carbonate system had been stabilized (leveled pH), 720 L of *in situ* seawater were
 110 filtered by a 180 μm mash to exclude large zooplankton, and each mesocosm bag was inoculated
 111 with 80 L of it to initiate the coastal microbial community. Samples were taken from each bag at a
 112 depth of 0.5m at 9:00 a.m. by niskin bottles every 1–3 days for physical, chemical and biological
 113 analysis.

114 **2.2 Measurement of environmental factors**

115 Solar light intensity was monitored every minute throughout the experimental period using a
 116 real-time solar irradiance monitoring device (EKO, Japan). Salinity, temperature and pH in each



mesocosm were measured with a salinometer, a digital thermometer and a pH meter (Orino 2 STAR, Thermo Scientific, U. S. A, calibrated with standard NBS buffer), respectively. Dissolved inorganic carbon (DIC) was sampled and measured using an Environmental Water Analyzer (Ma et al., 2018), and total alkalinity (TA) measured using an Automated Spectrophotometric Analyzer (Li et al., 2013). Other seawater carbonate chemistry parameters were calculated with CO2SYS software with known parameters of $p\text{CO}_2$, salinity, pH, temperature, and nutrient concentration.

Nutrient samples of each bag were filtered through 0.45 mm cellulose acetate membrane, and the filtrate was divided into 2 subsamples; one was stored at $-20\text{ }^{\circ}\text{C}$ for the measurement of $\text{NO}_3^- + \text{NO}_2^-$, PO_4^{3-} , and NH_4^+ ; another stored at 4°C for the measurement of SiO_3^{2-} . Measurement of $\text{NO}_3^- + \text{NO}_2^-$, PO_4^{3-} , and SiO_3^{2-} concentration was conducted using an auto-analyzer (AA3, Seal, Germany), NH_4^+ was measured with indophenol blue spectrophotometry using a spectrophotometer (Tri-223, Spectrum, China) at $25\text{ }^{\circ}\text{C}$.

2.3 Measurement of chlorophyll *a* and particle organic matters

Water samples of each mesocosm (100 – 1000 mL, depending on the biomass in the mesocosms) were filtered onto GF/F filter (Whatman, United States) by suction filter with low vacuum pressure no more than 0.02 MPa and soaked in 5 mL pure methanol overnight. The extracts were centrifuged at $8000 \times g$ and $4\text{ }^{\circ}\text{C}$ for 10 min, then the absorption spectra of supernatants from 400 to 800 nm were measured using a UV-VIS spectrophotometer (DU 800, Beckman, U. S. A). The Chl *a* concentration was calculated according to the following equation (Ritchie, 2006):

$$\text{Chl } a (\mu\text{g L}^{-1}) = 16.29 \times (A_{665} - A_{750}) - 8.54 \times (A_{652} - A_{750})$$

where A_{750} , A_{665} , and A_{652} represents the absorbance of Chl *a* at wave length 750, 665, and 652 nm, respectively.

For the analysis of particulate organic carbon (POC) and nitrogen (PON) across two particle size fractions, water samples of known volume were first filtered through a $20\text{ }\mu\text{m}$ mesh to obtain subsamples containing particle organic matters smaller than $20\text{ }\mu\text{m}$. Particles larger than $20\text{ }\mu\text{m}$ (retained on $20\text{ }\mu\text{m}$ mesh) were backwashed using an equal volume of prefiltered ($0.22\text{ }\mu\text{m}$) *in situ* seawater, yielding in subsamples containing particulate organic matters larger than $20\text{ }\mu\text{m}$. All subsamples were then filtered on pre-combusted ($450\text{ }^{\circ}\text{C}$, 6 h) GF/F filter (Whatman) and stored at $-20\text{ }^{\circ}\text{C}$ until analysis. Before analyses, all filters were fumed over pure HCl for 12 h and dried at



146 60 °C to remove inorganic carbon, then packed in tin cups and measured with a CHNS elemental
 147 analyzer (Vario EL cube Elementar, Germany).

148 **2.4 Net primary production and dark respiration**

149 Before sunrise, 120 mL of water samples from each mesocosm were collected and dispensed
 150 into six 25 mL borosilicate bottles (three bottles for 12 h incubations, and three for 24 h incubations).
 151 For each culture duration of each mesocosm, two bottles were illuminated under natural light and
 152 one bottle was wrapped tightly in aluminum foil as a dark control. After incubation, cells were
 153 filtered onto the GF/F filters (Whatman) under dim light and stored at −20 °C. Before measurement,
 154 filters were placed individually in 20 mL scintillation vials and exposed to HCl fumes overnight,
 155 dried in a constant temperature oven at 60 °C for over 6 h to remove any unincorporated $\text{H}^{14}\text{CO}_3^-$.
 156 The incorporated ^{14}C by algae was counted with a liquid scintillation counter (Beckman, LS6500,
 157 Germany) in the presence of 5 mL scintillation cocktail (Hisafe 3, Perkin-Elmer, United States).
 158 Nighttime respiratory carbon loss was calculated as the difference between carbon fixation over 12
 159 h (daytime primary productivity) and 24 h (daily net primary productivity).

160 **2.5 Determination of phytoplankton biomass and community structure**

161 Water samples (500–2000 mL) from each mesocosm were collected into polyethylene bottles
 162 and fixed with 1.5 % lugol's iodine. The samples were statically placed and concentrated into 50
 163 mL subsamples in the centrifuge tube using siphons within 3 days. The concentrated subsamples
 164 were examined with a microscopy (Nikon Eclipse Ns2) and a plankton counting chamber to assess
 165 phytoplankton abundance and diversity based on the morphological characteristics as previous
 166 described (Hasle and Syvertsen, 1997; Steidinger and Jangen, 1997; Yang and Liu, 2018). To
 167 distinguish whether dinoflagellates are autotrophic or heterotrophic, we observed living algal cells
 168 in the unstained water samples under the microscope for cell transparency and the presence of
 169 chloroplasts, as shown in Fig. S9.

170 For cell counts, an aliquot of 100 μL for each mesocosm was loaded onto a counting chamber
 171 for microscopic enumeration. In each aliquot, the count was deemed valid only when the total
 172 number of cells exceeded 200; otherwise, the subsample volume for microscopy was increased to
 173 achieve sufficient counts. For samples collected during the exponential growth phase that exhibited
 174 excessively high cell densities, appropriate dilution with 0.22 μm -filtered, sterilized seawater was



175 performed prior to counting (State Oceanic Administration 2005).

176 **2.6 Statistical analysis**

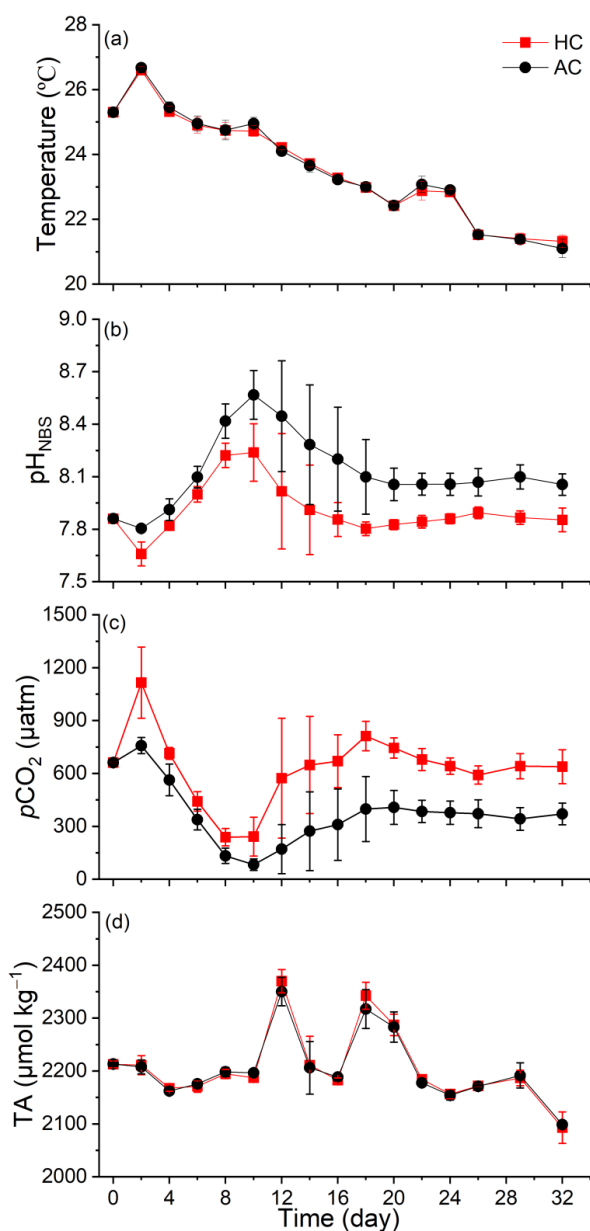
177 The data were all expressed by the mean and standard deviation (means $\pm$ SD) and plotted by
 178 Origin 2024. Independent-samples *t*-test was conducted to check the significant effects of increased
 179 $p\text{CO}_2$ at the level of $p < 0.05$ using SPSS 19. To evaluate α -diversity, Shannon diversity index was
 180 calculated based on the relative abundances of phytoplankton taxa using the estimate R and diversity
 181 functions from the vegan package (version 2.6-4) in R (Version 4.2.2). Shannon index incorporates
 182 both species diversity and evenness. Patterns of physiological parameters over time were
 183 emphasizing using generalized additive models (GAMs) and constructed using the ‘mgcv’ package
 184 in R to analyze changes in physiology through the experiment.

185 **3 Results**

186 **3.1 Environmental changes in the mesocosms**

187 Throughout the experiment, most days were sunny, with daytime mean PAR (12h-average
 188 photosynthetic active radiation) ranging from 200 to 850 $\mu\text{mol photons m}^{-2} \text{ s}^{-1}$ (Fig. S1). The
 189 environmental temperatures decreased gradually from 26.7 ± 0.05 °C at day 0 (9th October) to 21.1
 190 ± 0.28 °C at day 38 (14th November) (Fig. 1 a). Significant differences in pH_{NBS} and $p\text{CO}_2$ between
 191 HC and AC were maintained throughout most of the experimental period, while there’s no
 192 significant difference in total alkalinity (TA) between the two $p\text{CO}_2$ treatments ($p = 4.02 \times 10^{-8}$,
 193 2.87×10^{-12} , 0.549, respectively. Figs. 1 b-d, S5 a-c).

194 Following a sharp increase in phytoplankton biomass from Day 4 to Day 8 (Fig. 3 a), the pH_{NBS}
 195 in the HC and AC mesocosms increased and peaked at 8.24 ± 0.16 and 8.56 ± 0.14 , respectively
 196 (Fig. 1 b). Correspondingly, the $p\text{CO}_2$ value dropped to the lowest points of 238.48 ± 49.02 μatm
 197 (HC) and 82.82 ± 32.88 μatm (AC) (Fig. 1 c). Then, as the phytoplankton biomass decreased after
 198 day 8, pH_{NBS} gradually declined and $p\text{CO}_2$ increased, both stabilizing at relatively constant levels
 199 from day 18 until the end of experiment. There were no obvious temporal changes observed in total
 200 alkalinity (TA) throughout the experiment (Fig. 1 d).



201

202 Figure 1. Temporal variation of seawater temperature (a), pH_{NBS} (b), pCO₂ (c) and TA (d) in HC
 203 (1000 μatm) and AC (410 μatm) mesocosms. The pCO₂ was estimated from the measured pH_{NBS}
 204 and DIC concentration using the CO2SYS program. Data are means ± SD of 5 replicates for HC
 205 and 4 replicates for AC mesocosms.

206

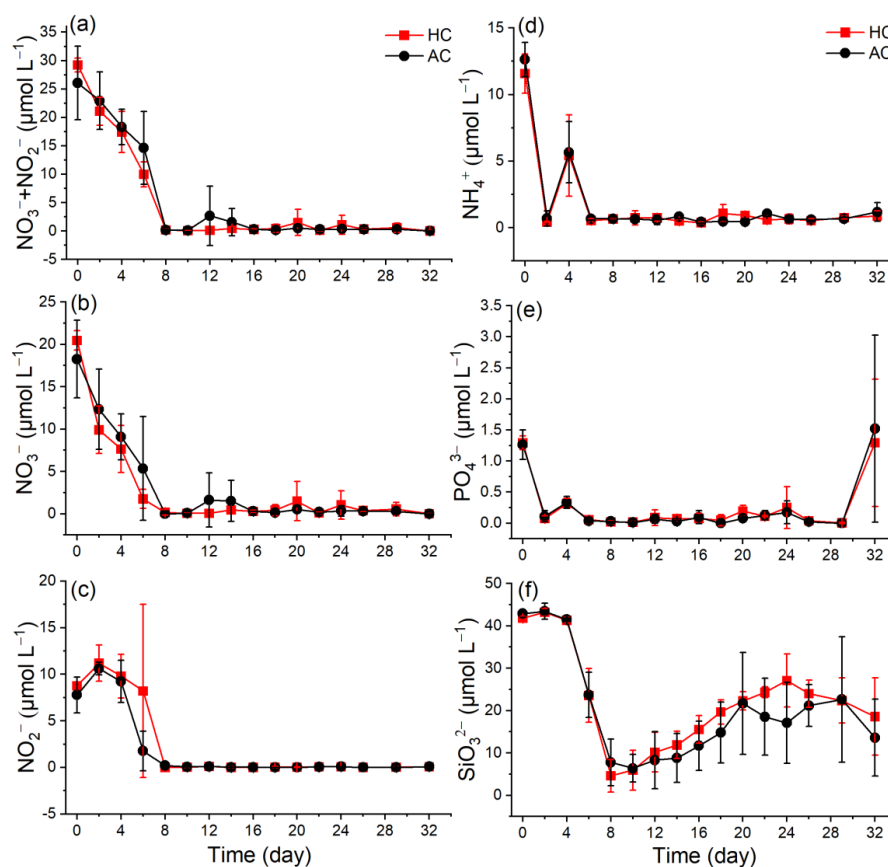
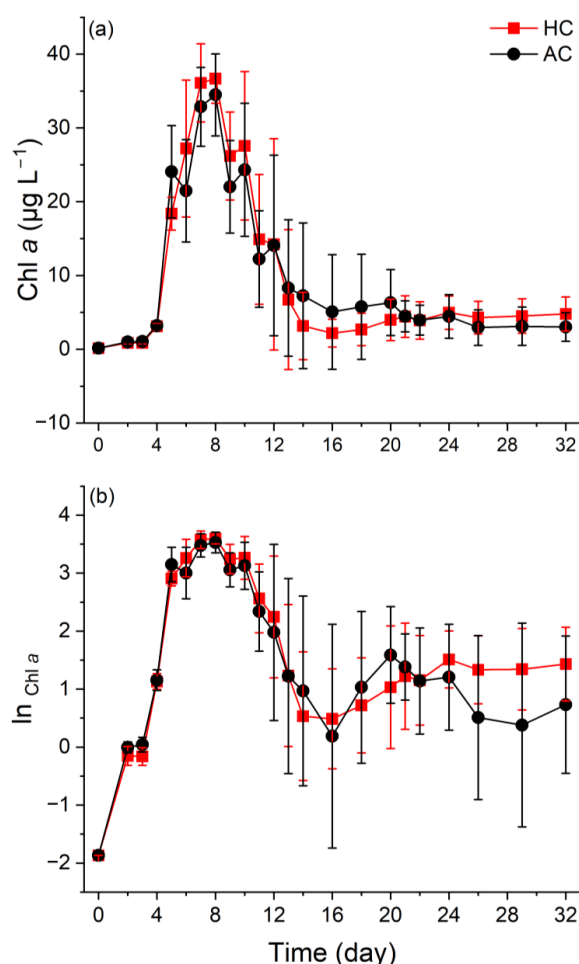


Figure 2. Temporal variation of nutrients (NO_3^- , NO_2^- , $\text{NO}_3^- + \text{NO}_2^-$, NH_4^+ , PO_4^{3-} , and SiO_3^{2-}) in HC (1000 μatm) and AC (410 μatm) mesocosms. Data are means $\pm$ SD of 5 replicates for HC and 4 replicates for AC mesocosms.

The initial nutrient concentrations reflected the eutrophic condition in the coastal seawater ($\text{NO}_3^- + \text{NO}_2^-$: 27 μM , PO_4^{3-} : 1.4 μM). In the mesocosm bags, nutrient concentrations declined dramatically in the early phase (up to day 8, Fig. 2). The $\text{NO}_3^- + \text{NO}_2^-$ and NO_3^- concentrations decreased sharply to nearly 0 by day 8 (Fig. 2 a, b). In contrast, the NO_2^- concentration experienced a slight increase on day 2, then declined to nearly 0 by day 8 and remained at a low level until the end of experiment in both HC and AC mesocosms (Fig. 2 c). Under HC condition, the NO_3^- concentration decreased more rapidly than that under the AC until day 20, although the difference



219 was not significant ($p = 0.423$, Fig. S5 d). Both NH_4^+ and PO_4^{3-} concentrations dropped to nearly 0
 220 after 4 days and remain relatively stable thereafter. There were no significant difference observed
 221 between HC and AC mesocosms ($p = 0.579$ and 0.631 , respectively, Figs. 2 d, e, S5 e, f). The SiO_3^{2-}
 222 concentration decreased from $41.81 \pm 0.48 \mu\text{M}$ in HC and $42.88 \pm 0.91 \mu\text{M}$ in AC on day 0 to a
 223 minimum of $4.62 \pm 3.82 \mu\text{M}$ and $7.79 \pm 5.52 \mu\text{M}$ by day 8, respectively. Thereafter, SiO_3^{2-}
 224 concentrations in both HC and AC mesocosms gradually increased until the end of experiment ($p =$
 225 0.343 , Figs. 2 f, S5 g).



226
 227 Figure 3. Temporal variations of Chl *a* concentration (a) and the LN scale of Chl *a* concentration (b)
 228 in HC (1000 μatm) and AC (410 μatm) mesocosms. Data are means $\pm$ SD of 5 replicates for HC and
 229 4 replicates for AC mesocosms.



3.2 Chlorophyll *a* concentration

Phytoplankton biomass, indirectly indicated by Chl *a* concentration, increased to peak at $35.88 \pm 3.25 \mu\text{g L}^{-1}$ in the HC and 36.54 ± 4.88 in the AC mesocosms on day 8, and then gradually decreased to $1.12 \pm 0.43 \mu\text{g L}^{-1}$ in HC by day 14 and $0.58 \pm 0.31 \mu\text{g L}^{-1}$ in AC by day 16, followed by a slight increase by the end of experiment (Fig. 3 a). Based on the natural logarithm (ln) scale of Chl *a* concentration, the phytoplankton growth kinetics under the two $p\text{CO}_2$ treatments showed the following phases (Fig. 3 b): the exponential phase (from day 0 to day 5), the stationary phase (from day 6 to day 10), the decline phase (from day 11 to day 16), and a second exponential phase from day 17 to day 24 in the HC and to day 20 in the AC mesocosms. Then, phytoplankton assemblages in the HC mesocosms entered a second stationary phase until the end of experiment, while in the AC ones, they entered a decline phase until day 29, followed by a slight increase on day 32.

Throughout the experiment, the elevated $p\text{CO}_2$ resulted in higher average value of Chl *a* concentration at most sampling times, although the differences were not statistically significant ($p = 0.142$, Fig. S5 h).

3.3 Primary production and dark respiration

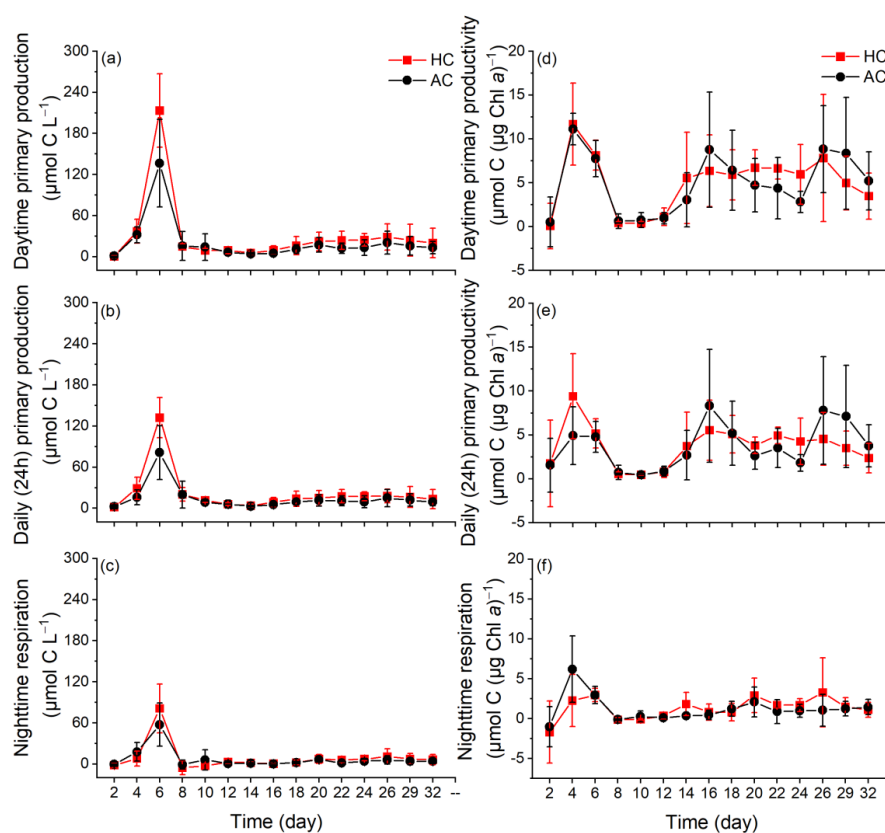
The primary production and night-respiratory per water volume showed patterns similar to those of phytoplankton biomass (indicated by Chl *a* concentration) (Fig. 4, a, b and c). They reached their maximal values on day 6, which corresponded to the end of exponential phase. As the phytoplankton communities entered the stationary phase, daytime (12 h) primary production, daily (24 h) net primary production and nighttime respiration per water volume progressively decreased, and then slightly increased again when the phytoplankton communities underwent the second exponential phase. The elevated $p\text{CO}_2$ increased both daytime and daily net primary production during the middle phase of the experiment, although the positive effect on 24 h primary production tended to decline by the end of experiment ($p = 0.038$ and 0.012 , Fig. S6 a, b). The nighttime respiration of phytoplankton was suppressed before day 8 and enhanced thereafter under the elevated $p\text{CO}_2$, though no significant difference was observed ($p = 0.444$, Fig. S6 c).

Primary productivity per Chl *a* increased sharply on day 4, and decreased to the lowest values on day 8. On day 12, both daytime and 24 h primary productivity in the HC increased drastically and then remained relatively stable until the end of experiment (Fig. 4 d, e). In contrast, two



259 additional peaks were observed in the AC mesocosms on days 16 and 26. The elevated $p\text{CO}_2$
 260 appeared to have enhanced primary productivity from day 2 to day 20, though these effects were
 261 not statistically significant ($p = 0.946$ for daytime and $p = 0.985$ for 24 h, Figs. 4 d, e, S6 d, e).

262 Nighttime respiration per μg Chl *a* initially increased on day 4, then decreased to nearly zero
 263 in both the HC and AC mesocosms on day 8 and remained relatively stable till the end of experiment.
 264 The elevated $p\text{CO}_2$ had a negative effect on phytoplankton respiration before day 12, but increased
 265 it thereafter, though no significant difference was observed between the HC and AC treatments ($p =$
 266 0.834 , Figs. 4 f, S6 f).



267
 268 Figure 4. The changes of daytime primary production (a) and primary productivity (b), daily (24h)
 269 primary production (c) and primary productivity (d), nighttime respiration per water volume (e) and
 270 per Chl *a* (f) in HC (1000 μatm) and AC (410 μatm) mesocosms. Data are means $\pm$ SD of 5 replicates
 271 for HC and 4 replicates for AC mesocosms.



272

273 3.4 Changes in Phytoplankton community and diversity

274 A total of 47 genera identified microscopically include 33 genera of diatoms, 7 of
 275 dinoflagellates, 2 of cyanobacteria, 2 of chlorophyta, 2 of cryptophyta and 1 of euglenophyta. In all
 276 mesocosms, the dominant species included *Cerataulina pelagica*, *Eucampia cornuta*, *Guinardia*
 277 *delicatula*, *Leptocylindrus danicus*, *Skeletonema costatum*, *Protoperidinium* sp., *Gyrodinium*
 278 *spirale*, *Cryptophyta* sp. and *Pyramimonas* sp. (Fig. S4).

279 Phytoplankton communities underwent dynamic succession in the mesocosms (Fig. 5).
 280 Diatoms (mainly *Cerataulina pelagica*) dominated the phytoplankton communities during the early
 281 and middle stages of the experiment, as indicated by the similar temporal trends in total
 282 phytoplankton and diatom cell counts compared with Chl *a* concentration (Figs. 5 a, b, S4 a). Diatom
 283 density was lower in the HC than in the AC mesocosms, though the difference was not statistically
 284 significant ($p = 0.259$, Fig. S7 a). Autotrophic dinoflagellates began to emerge on day 8 and rapidly
 285 declined on day 12 in both HC and AC enclosures (Fig. 5 c). Except for days 6 to 18, the elevated
 286 $p\text{CO}_2$ increased the biomass of autotrophic dinoflagellates, though the difference was insignificant
 287 ($p = 0.505$, Fig. S7 b). Hetero-dinoflagellates began to emerge on day 6, with their abundance
 288 peaked on day 12 in the AC and on day 14 in the HC mesocosms, then decreased by day 22. The
 289 elevated $p\text{CO}_2$ did not result in any significant change in terms of their cell numbers ($p = 0.785$,
 290 Figs. 5 d, S7 c). On day 26, the biomass of hetero-dinoflagellates increased again in the HC
 291 treatment, while it remained constant in the AC treatment ($p = 0.729$, Independent-samples *t*-test).

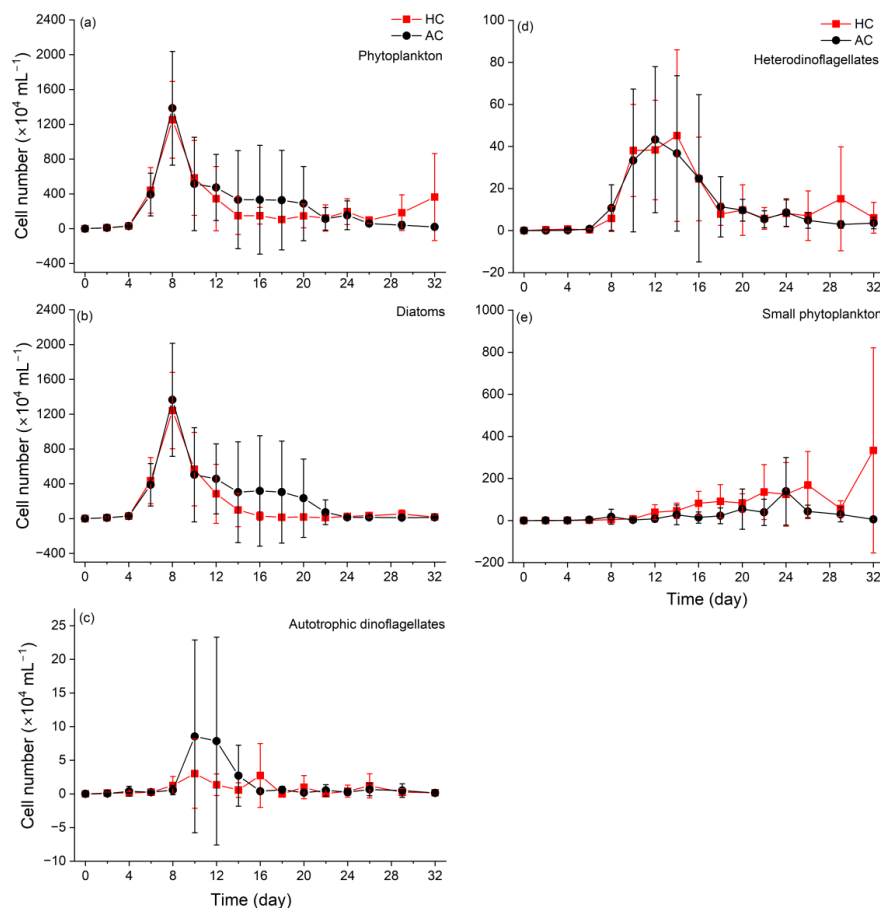
292 The biomass of small taxa (Cyanobacteria, Chlorophyta, Cryptophyta and Euglenophyta)
 293 started to increase on day 8, the HC treatment significantly increased the total biomass of these
 294 small phytoplankton species thereafter ($p = 0.019$, Figs. 5 e, S4 h, i, S7 d). From day 22, when
 295 diatoms biomass decreased to the lowest level, the temporal variation in small taxa biomass became
 296 the main factor controlling overall phytoplankton dynamics (Figs. 5 e, S4 h, i). Accordingly, the
 297 positive effect of HC on the small phytoplankton species led to an earlier transition of phytoplankton
 298 from the large diatoms and dinoflagellate (mainly fall within the micro size fraction) to the smaller
 299 ones (Fig. 6 a, b). This accelerated transition in the HC treatment was also evidenced by higher
 300 concentration of POC and PON in the $<20\text{ }\mu\text{m}$ fraction and lower concentration in the $>20\text{ }\mu\text{m}$



fraction (Figs. S2, S3).

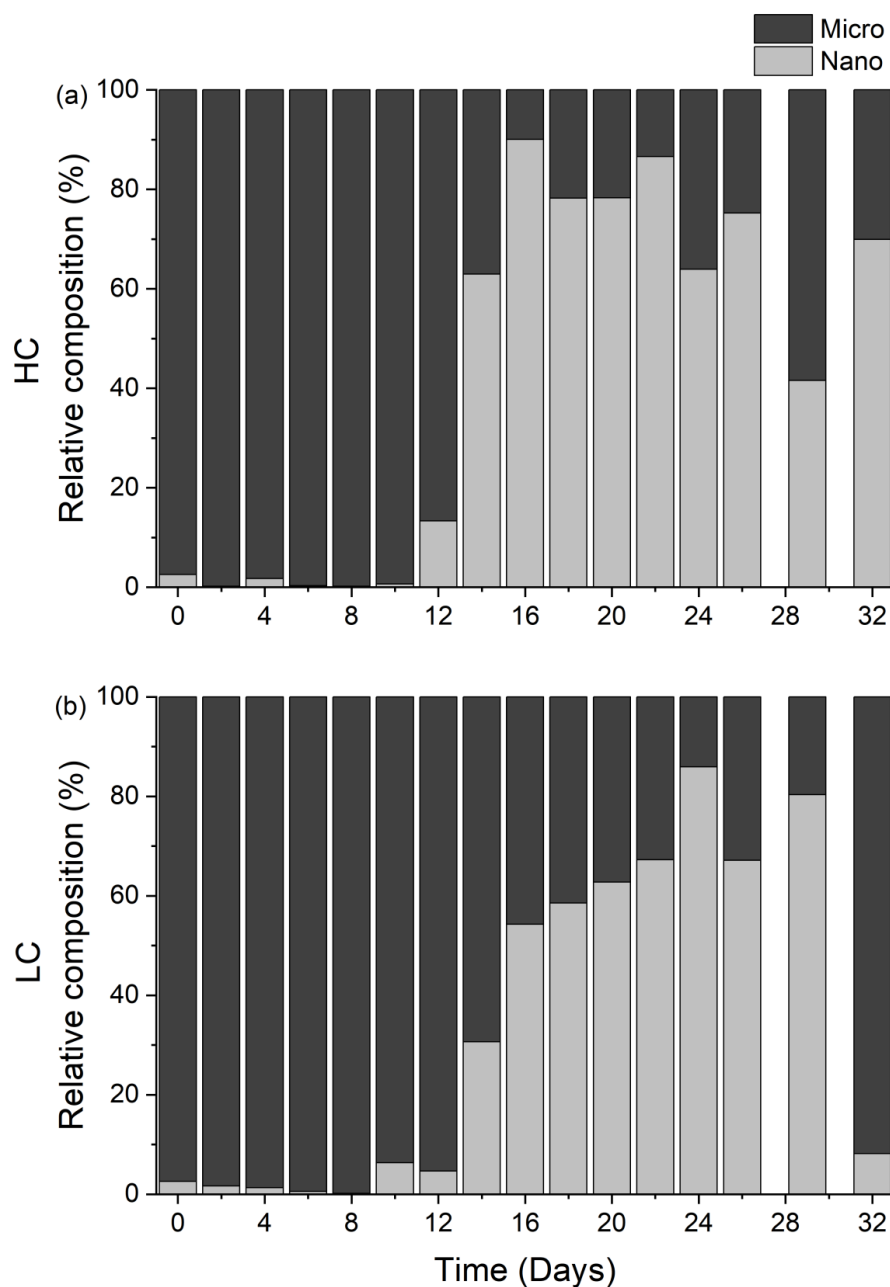
In both HC and AC mesocosms, Shannon diversity index decreased sharply from day 2, reaching the lowest values on day 8 in AC mesocosms and on day 10 in HC mesocosms (Fig. S8 a). Before day 22, Shannon diversity index increased under elevated $p\text{CO}_2$, whereas it is lowered under elevated $p\text{CO}_2$ level since day 24, although the differences were not statistically significant ($p = 0.161$, Fig. S8 b).

307



308

Figure 5. Temporal variations of phytoplankton (a), diatoms (b), photodinoflagellates (c), heterodinoflagellates (d) and (e) small phytoplankton (Cyanobacteria, Chlorophyta, Cryptophyta and Euglenophyta) cell numbers in HC (1000 μatm) and AC (410 μatm) mesocosms. Data are means $\pm$ SD of 5 replicates for HC and 4 replicates for AC mesocosms.



313
 314 Figure 6. Temporal variations of the relative composition of diatoms + dinoflagellates (Diat + Dino,
 315 black), Cyanobacteria + Chlorophyta + Cryptophytes + Euglenophyta (Cyano + Chlo + Cryp + Eugl,
 316 grey) in HC (1000 µatm, a) and AC (410 µatm, b) mesocosms. Data are means of 5 replicates for



317 HC and 4 replicates for AC mesocosms.

318

319 **4 Discussion**

320 Ocean global changes have been suggested to alter community structure and reduce the
 321 diversity of phytoplankton due to physicochemical environmental changes (Henson et al., 2021;
 322 Yuan et al., 2020). Specifically, there appears a growing trend of increasing dinoflagellates
 323 abundance relative to diatoms (Carreto et al., 2018). Our mesocosm experiment, conducted in the
 324 highly eutrophic Wuyuan Bay in the southern East China Sea during late autumn, also indicated that
 325 elevated $p\text{CO}_2$, along with the natural decrease of surface water temperature and declined nutrient
 326 availability, altered the structure and diversity of phytoplankton community. The diatom dominance
 327 corresponded to the decreased diversity and evenness of phytoplankton community, while the
 328 diversity and evenness were recovered when the diatom dominance was replaced by dinoflagellates,
 329 however, this shift was relatively suppressed under elevated $p\text{CO}_2$ conditions. In our mesocosms,
 330 the dinoflagellates that emerged during the mid-phase (e.g., *Protoperidinium* sp.,
 331 *Pentapharsodinium dalei* and *Heterocapsa* sp. Fig. S9 a) were predominantly small ($<20\ \mu\text{m}$, Fig.
 332 S4 g) (Gu et al., 2013; Hanifah et al., 2022). Subsequently, these dinoflagellates were soon replaced
 333 by an even smaller size fraction, including Cyanobacteria, Chlorophyta, Cryptophytes, and
 334 Euglenophyta (Figs. 5, 6, S4). Ultimately, these smaller taxa maintained the primary production of
 335 phytoplankton communities after nutrient depletion (Fig. 4).

336 When diatoms dominated the phytoplankton community (before day 8), elevated $p\text{CO}_2$ had an
 337 insignificant negative effect on their biomass (Figs. 5 b, S7 a), but increased the primary production
 338 per water volume and per $\mu\text{g Chl } a$ (Fig. 4 a, b, d and e). These may be attributed to the competitive
 339 advantages conferred by CO_2 -concentrating mechanisms (CCMs) in diatoms: their higher CO_2
 340 affinity and CCMs plasticity may help diatoms gain a competitive advantage in DIC uptake under
 341 ocean acidification scenarios (Huang et al., 2021; Raven and Beardall, 2020). Furthermore, the
 342 down-regulated of CCMs in diatoms can save energy for other physiology processes and thereby
 343 fuel their primary production (see the review by Gao and Campbell, 2014 and the references therein).
 344 These benefits resulting from elevated $p\text{CO}_2$ level also led to higher diversity and evenness (Fig. S8
 345 a, b), suggesting that more diatom species were benefited from the elevated $p\text{CO}_2$, though the total



346 diatom biomass was lower in HC mesocosms (Figs. S4, S7 a). While previous works have
 347 demonstrated a positive effect of elevated $p\text{CO}_2$ on the photosynthetic carbon fixation by diatoms
 348 grown under low light and phytoplankton assemblages in waters of higher nutrient availability (Gao
 349 et al., 2022), our results (Figs. 5 b, S1) indicate that nutrient limitation can override or even reverse
 350 to the positive effects of elevated $p\text{CO}_2$ on diatoms (Boyd et al., 2016; Li et al., 2018).

351 It appeared that dinoflagellates were less sensitive to the depletion of nutrients compared to
 352 diatoms, with autotrophic dinoflagellates were more sensitive to elevated $p\text{CO}_2$ (Figs. 5 c, S7 b).
 353 These suppressed transition from diatoms to autotrophic dinoflagellates under the elevated $p\text{CO}_2$
 354 was consistent with our previous works (Huang et al., 2021), which is likely due to the different
 355 CCMs efficiency and/or acidic resilience between dinoflagellates and diatoms. Since the affinity of
 356 ribulose 1, 5-diphosphate carboxylase/oxidase (Rubisco) for CO_2 is much lower in autotrophic
 357 dinoflagellates than in diatoms (Reinfelder, 2011), elevated $p\text{CO}_2$ must have benefitted the former
 358 more compared the latter, though invisible growth advantage was observed on the autotrophic
 359 dinoflagellates between days 8 and 16. The heterodinoflagellates can utilize organic matters (Glibert
 360 and Legrand, 2006) and prey on microbes including bacteria and smaller microalgae (Jeong et al.,
 361 2010). This versatile nutrition strategy supported their rapid bloom starting from day 8, leading to
 362 the replacement of autotrophic ones from day 12 onward (Fig. 5 c, d). Although they were shown
 363 to be insensitive to ocean acidification (Meunier et al., 2017), their respiration was depressed due
 364 to the acidic stress, raising their resilience in terms of energetic cost (Wang and Gao, 2024). These
 365 mechanisms explain the observed insignificant effects of HC on hetero-dinoflagellates.

366 The increases in $\text{NO}_3^- + \text{NO}_2^-$, PO_4^{3-} and SiO_3^{2-} concentrations from day 10 onward (Fig. 2 a,
 367 e, f) should be attributed to remineralization by heterotrophic bacteria (Aristegui et al., 2009; Bunse
 368 and Pinhassi, 2017). The gradual increase in SiO_3^{2-} concentration, a nutrient exclusively required
 369 by diatoms, coincided with their decline, confirming the low abundance of diatoms in the mid and
 370 late phase of experiment. These regenerated $\text{NO}_3^- + \text{NO}_2^-$ and PO_4^{3-} subsequently refueled the
 371 growth of small phytoplankton taxa, recover the diversity and evenness in the phytoplankton
 372 communities (Figs. 5 e, S8 a, b) (Thingstad and Rassoulzadegan, 1995). Alternatively, it is plausible
 373 that grazing activity by zooplankton, which was not quantified in this study, also contributed to the
 374 apparent rise in diversity and evenness, as grazers tend to consume dominant phytoplankton taxa



375 (Thingstad and Rassoulzadegan, 1999; Calbet and Landry, 2004).

376 The dominant small taxa, such as *Cryptophyta* sp. and green microalga *Pyramimonas* sp. (Fig.
 377 S4 h, i) during days 16–24, achieved primary productivity (per μg Chl *a*) comparable to the diatom-
 378 dominated community observed on days 4–6 (Fig. 4 b, d). The success of these small taxa can be
 379 attributed to their small size and larger surface-to-volume ratio (Finkel et al., 2009; Giordano et al.,
 380 2005), which might enable them with higher efficiency in nutrients uptake and CO_2 diffusion.
 381 Furthermore, the higher abundance of viruses and heterotrophic bacteria in the HC mesocosms
 382 (Huang et al., 2021; Lin et al., 2018) intensified nutrient remineralization, subsidizing these small,
 383 fast-growing phototrophs and leading to their earlier emergence on day 16 compared to day 24 in
 384 the AC mesocosms (Fig. 6). While it's possible that picophytoplankton originally present in this
 385 region (Zhong et al., 2020) were missed by microscope-based identification in our mesocosm
 386 experiment, it is reasonable to infer that they also contributed to the late-phase phytoplankton
 387 communities. Previous studies indicated that, after diatom/dinoflagella blooms and nutrient
 388 depletion, remineralized nutrients in the seawater may also favor the growth of picophytoplankton
 389 (Nishibe et al., 2015; Fu et al., 2009) and elevated $p\text{CO}_2$ would further benefit their growth. Thus,
 390 it is likely that, picophytoplankton also dominated the phytoplankton communities in the HC
 391 mesocosms.

392 Previous studies have suggested that the shift from diatom to dinoflagellate dominance was
 393 generally associated with declines in primary productivity (Huang et al., 2021; Cloern, 1996). The
 394 results from the present autumn experiment reveals the same pattern with the previous spring
 395 mesocosm experiment (Huang et al., 2021), indicating that it is not the seasonal temperature
 396 trajectories but the availability of nutrients that controlled the shift. Such consistency underscores
 397 that nutrient availability and stoichiometry are the primary determinants of phytoplankton
 398 community composition, usually exerting stronger and more immediate effects on taxonomic and
 399 functional group dominance (Karl et al., 1996; Paerl and Paul, 2012; Ptacnik et al., 2008; Meyer et
 400 al., 2016), though thermal and acidic stresses can impact photosynthesis and respiration to greater
 401 extent under nutrient limitation (Li et al., 2018; Gao et al., 2022).

402 Reduced nutrient availability usually decreases phytoplankton community richness (Gazeau et
 403 al., 2017), although ocean acidification appeared to partly offset such effects (Fig. S8). However,



404 these compensatory effects diminished once both the initial and regenerated nitrogen sources were
405 exhausted (after day 24, Fig. S8 b). At that point, only a few small phytoplankton taxa tolerant to
406 low pH remained dominant, indicating a loss of diversity in the community and less stable
407 ecosystems (Mccann, 2000) under combination of acidic stress and nutrient limitation. Beyond
408 compensating previous works, our study further demonstrated that progressive ocean acidification
409 is likely to reduce primary productivity and phytoplankton diversity in the eutrophicated coastal
410 water of the southern East China Sea.

411 **Data availability Statement**

412 All relevant data are presented in the paper and its Supporting Information file, and will be available
413 upon request to the corresponding author Kunshan Gao.

414 **Conflict of Interest**

415 The authors declare no competing interests.

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422 experiment.

423 **Author's Contributions**

424 Kunshan Gao, Guang Gao and Xin Lin designed the mesocosm experiment; Yuming Rao, Na Wang,
425 Jiazhen Sun, Xiaowen Jiang, Di Zhang, Liming Qu, He Li, Qianqian Fu, Xuyang Wang, Cong Zhou,
426 Zichao Deng, Yang Tian, Xiangqi Yi, Ruiping Huang performed the mesocosm experiment; Yuming
427 Rao analyzed the data and wrote the manuscript; Na Wang performed microscopy observation;
428 Kunshan Gao edited the manuscript; All authors reviewed and contributed to revision of the
429 manuscript.

430

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