



1 **Differences in organic carbon fractions and stability explain limited accumulation in loam**
2 **and sandy loam under greenhouse conditions**

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18



19 **Abstract**

20 Manure is widely applied in greenhouses to enhance soil organic carbon (SOC) and improve
21 fertility. However, how SOC fractions and their chemical stability change under different soil
22 textures and long-term manure inputs remains unclear. We investigated greenhouse soils with
23 2–50 years of manure application in loam and sandy loam. SOC, easily oxidizable carbon (EOC),
24 microbial biomass carbon (MBC), dissolved organic carbon (DOC), particulate organic carbon
25 (POC), and mineral-associated organic carbon (MAOC) were quantified. The molecular structures
26 of SOC were analyzed via ¹³C NMR spectroscopy. Results showed that SOC in loam stabilized
27 after about 20 years of manure application, whereas sandy loam reached equilibrium within 2
28 years. In loam, aromatic C and carbonyl C in SOC increased, raising the aromaticity index (ARM);
29 in sandy loam, alkyl C increased, elevating A/OA and the hydrophobicity index (HI). Loam
30 contained higher SOC, EOC, POC, and MAOC contents than sandy loam, with SOC positively
31 correlated with EOC, POC, and MAOC, whereas in sandy loam SOC was positively correlated
32 only with EOC and MAOC, and negatively with fPOC. In loam, ARM and HI promoted SOC
33 accumulation by stimulating EOC and POC, which enhanced MAOC formation. In sandy loam,
34 HI mainly promoted SOC through increasing EOC, which enhanced MAOC formation. In loam,
35 MAOC formation was mainly associated with POC, whereas in sandy loam it was driven by EOC.
36 Overall, in greenhouses, long-term stability of SOC depends on the transformation of labile
37 carbon into stable fractions, with fine-textured soils exhibiting greater sequestration efficiency due
38 to higher structural stability and greater MAOC accumulation.

39 **Keywords:** organic carbon sequestration, organic carbon fractions, organic carbon stability,
40 greenhouse soils, long-term manure application



41 **1. Introduction**

42 Soil organic carbon (SOC) is a keystone determinant of soil fertility and sustainable
43 agricultural production; its quantity and composition directly regulate soil fertility and resilience
44 to environmental stress (Beillouin et al., 2023; Oldfield et al., 2019). SOC fractions differ
45 markedly in function. Labile carbon pools (e.g., soil labile organic carbon fractions [SLOCF] and
46 particulate organic carbon [POC]) are highly sensitive to environmental and management
47 perturbations, capturing short-term shifts in soil quality and carbon turnover; whereas stable
48 carbon pools (e.g., mineral-associated organic carbon [MAOC]) reflect the long-term stabilization
49 and preservation of SOC, functioning as the basis for maintaining structural stability and
50 long-term fertility (Das et al., 2023; Zheng et al., 2022; Lavalée et al., 2020; Witzgall et al.,
51 2021). Therefore, compared with SOC content, SOC composition can reveal soil productivity,
52 nutrient supply, and carbon sequestration potential in greater detail.

53 Manure application has been widely demonstrated to significantly enhance SOC content,
54 attributable not only to direct carbon inputs but also to its stimulation of microbial activity and the
55 consequent promotion of organic matter decomposition and resynthesis (Denoncourt et al., 2025;
56 Wu et al., 2024). A global meta-analysis by Gross and Glaser (2021) showed that manure
57 application, on average, enhanced SOC stocks by about 35.4%. Although total SOC increases, the
58 effects of manure application on different carbon fractions are not consistent. Manure application
59 preferentially enriches SLOC and POC, and these labile carbon pools are gradually incorporated
60 into stable pools through microbial reorganization and mineral mediation, serving as important
61 precursors for MAOC formation (Hao et al., 2025; Zhou et al., 2024). At the molecular level, ¹³C
62 Nuclear Magnetic Resonance (NMR) analyses revealed increases in alkyl C, declines in aromatic



63 C, higher A/OA, and lower ARM following manure application, suggesting a shift of SOC from
64 aromatic dominance to more hydrophobic and aliphatic structures (Zhao et al., 2023). The
65 efficiency of manure-derived carbon transformation into SOC depends on climatic conditions, soil
66 texture, and agricultural systems. Long-term manure application in non-tropical climates can
67 increase SOC stocks by about 12.8 Mg ha⁻¹, whereas in tropical regions the increase is only about
68 8.5 Mg ha⁻¹ (Gross and Glaser, 2021). However, Galluzzi et al. (2025) pointed out that although
69 rising temperatures usually accelerate SOC decomposition, SOC accumulation is simultaneously
70 constrained by mineralogical composition. Compared with coarse-textured soils, fine-textured
71 soils (such as loam) possess higher surface area and stronger sorptive capacity, which enable them
72 to more effectively bind low-molecular-weight fractions such as SLOCF and POC into MAOC,
73 thus enhancing SOC long-term storage (Kleber et al., 2015; Islam et al., 2022; Das et al., 2023).
74 Under similar temperature and soil texture conditions, SOC in the 0–20 cm plow layer of paddy
75 soils averages about 9 Mg ha⁻¹ higher than in upland soils., primarily due to the inhibition of SOC
76 mineralization under flooded conditions (Liu et al., 2021).

77 In contrast to open-field agriculture, greenhouses are maintained under consistently warm
78 and humid environments, with air temperatures typically ranging from 18–35°C and relative
79 humidity of approximately 65–80% (Tan et al., 2025; Yang et al., 2024). Such stable and
80 relatively high temperature-humidity regimes enhance microbial activity and substrate
81 decomposition, thereby accelerating SOC turnover. Consequently, to improve soil fertility,
82 manure inputs in greenhouses often reach 30–120 Mg ha⁻¹ yr⁻¹; however, SOC accumulation
83 remains considerably lower than anticipated (Zhang et al., 2022; Tan et al., 2025). Zhang et al.
84 (2022) found that continuous annual application of 37.5 Mg ha⁻¹ of chicken manure resulted in a



85 new SOC equilibrium after five years, with SOC content increasing by 1.27-fold, yet only 24% of
86 the exogenous carbon was retained. In addition, the majority of greenhouse studies span only
87 1–10 years, and although apparent equilibrium in total SOC has been observed, such short-term
88 stabilization does not represent an endpoint, as SOC may continue to accumulate or shift over
89 longer timescales. Evidence from open-field studies indicates that net SOC accumulation,
90 particularly MAOC, typically requires continuous organic inputs for more than 20–30 years
91 before becoming pronounced (Dietz et al., 2024; Just et al., 2023). To more comprehensively
92 evaluate the effects of manure application on SOC in greenhouses, it is necessary to analyze SOC
93 accumulation dynamics and carbon fraction distribution over much longer timescales.
94 Accordingly, this study examined greenhouse soils with loam and sandy loam spanning 2–50
95 years of manure application, employing analyses of SOC fractions (SLOCF, POC, MAOC)
96 combined with ^{13}C NMR spectroscopy to systematically compare the composition and molecular
97 structural features of SOC fractions across different manure application years. The aims of this
98 study were: (1) to compare variations in SOC content, structural features, and molecular
99 composition between loam and sandy loam soils in greenhouses under different years of manure
100 application; and (2) to determine the relationships between changes in SOC fractions and
101 molecular structural features with long-term SOC accumulation. The findings will provide a
102 scientific basis for formulating greenhouse manure application strategies and sustaining soil
103 fertility over the long term.

104 **2. Materials and methods**

105 **2.1 Site description**

106 The sampling sites for this study were located in Wujianfang Town ($41^{\circ}49'12''\text{N}$,



120°43'12"E) and Yangshibao Town (41°37'48"N, 122°45'36"E), which are two key regions for greenhouse agriculture in Liaoning Province, northeast China. Greenhouse agriculture has steadily developed in these two regions since 1970 and 1988, respectively, and has gradually expanded with the advancement of farming techniques and improved economic benefits. To improve production efficiency and support local agriculture, the government organized standardized planting schedules, crop rotation, and water-nutrient management training, thus providing ideal conditions for investigating the composition of SOC fractions and their stabilization characteristics under varying durations of manure application in greenhouses.

By 2020, the greenhouses in these two regions had planting histories ranging from 2–50 years (Table 1). The soil in Wujianfang Town is classified as Luvisols, while the soil in Yangshibao Town is classified as Cambisols (FAO, 2015). The soil texture was classified based on the proportion of sand, silt, and clay particles in the collected samples, using the United States Department of Agriculture soil texture triangle. Soils in Wujianfang and Yangshibao were classified as loam and sandy loam, respectively (Table 1). Each greenhouse sample plot was 60 m × 15 m, with an average daytime temperature of 22–36 °C and nighttime temperature of 15–18 °C. A three-crop rotation system of tomato-bean-tomato was implemented annually. Tomato is the main crop, with a planting density of 82,500 plants ha⁻¹.

The greenhouses were treated annually with approximately 41.25 Mg ha⁻¹ (dry weight) of mixed cow and pig manure composts, providing a carbon input of 11.58 Mg ha⁻¹ yr⁻¹. The compost contained 28.07% carbon and 2.24% nitrogen. Additionally, chemical fertilizers were applied at annual rates of 300–400 kg ha⁻¹ N, 300–400 kg ha⁻¹ P₂O₅, and 450 kg ha⁻¹ K₂O, in combination with a drip irrigation system. To compare the effects of the greenhouse, open-field



129 soils located within 100 m of the sampled greenhouse were selected as controls (CK). These
130 control fields followed the same tomato-soybean-tomato rotation, with similar chemical fertilizer
131 application rates but no manure application.

132 **2.2 Soil sampling and analysis**

133 Soil samples were collected in May 2020. Greenhouse plots with similar manure application
134 durations were grouped into loam (Wujianfang Town) and sandy loam (Yangshibao Town) soils.
135 Each group consisted of 6 to 8 samples, for a total of 90 samples. In each greenhouse plot, soil
136 samples were collected from five points arranged in an X-pattern at a depth of 0–20 cm and
137 combined to form a composite sample. To ensure consistency and reduce variability, samples with
138 marked differences in soil texture were excluded, and only soils with a neutral pH (6–8) were
139 retained. This approach was adopted to minimize potential bias from soil property heterogeneity
140 and to ensure the reliability of the experimental results. The same sampling and screening
141 methods for the CK were applied to the CK adjacent to the greenhouse plots. After screening, 42
142 samples remained. Among these, four samples were collected from greenhouse sandy loam aged
143 30–32 years, and loam aged 9–11 years and 28–31 years, while the remaining groups consisted of
144 three samples each (Table 1).

145 **2.3 SOC and sequestration analysis**

146 The SOC content was determined using the potassium dichromate ($K_2Cr_2O_7$) oxidation
147 method (Lu, 2000).

148 SOC stock (SOC_{stock} , $Mg\ C\ ha^{-1}$) was estimated by Eq. (1) (Liu et al., 2015):

$$149 \quad SOC_{stock} = SOC \times BD \times 0.20 \times 10^4 \times 10^{-3} \quad (1)$$



Where SOC is the SOC content (g kg^{-1} soil), BD is the soil bulk density (g cm^{-3}), and 0.20 is the depth of the plow horizon (m). 10^4 is the conversion factor for hectare to square meters (m^2), and 10^{-3} is the overall unit conversion factor for adjusting the units to Mg C ha^{-1} .

SOC sequestration rate (SOC_s , %) was estimated using Eq. (2) (Liu et al., 2015; Lu et al., 2009):

$$SOC_s = \frac{SOC_{stock-n} - SOC_{stock-CK}}{C-input_n} \times 100 \quad (2)$$

Where $SOC_{stock-n}$ and $SOC_{stock-CK}$ represent the carbon stocks (Mg C ha^{-1}) in greenhouse soil with n years of manure application and in the control group, respectively. $C-input_n$ represents the cumulative C input from manure application over n years (Mg C ha^{-1}).

2.4 Fraction of SOC

The SLOCF (EOC, DOC, and MBC), POC (cPOC and fPOC), and MAOC was separated from the soil samples and measured to evaluate the long-term impact of manure application on the turnover and accumulation of different forms of SOC in the greenhouse.

KMnO_4 oxidation method (Vieira et al., 2007) was used to measure EOC in the samples. The fumigation extraction method (Vance et al., 1987) was used to measure DOC and MBC in the samples. DOC content was determined using the extract from the non-fumigated samples. MBC content was estimated using Eq. (3):

$$MBC = \frac{OC_f - DOC}{0.45} \quad (3)$$

where OC_f and DOC represent the organic carbon (OC) contents of the fumigated and non-fumigated samples, respectively (g kg^{-1} soil), and 0.45 is a constant value representing the OC extraction efficiency.

The contents of POC and MAOC were measured using a particle size fractionation method.



172 An air-dried soil sample (10 g) was placed in a 50 mL centrifuge tube, and 35 mL of 1.85 g cm^{-3}
 173 NaI solution (85 g NaI dissolved in 100 mL of water) was added. After shaking the tube 10 times,
 174 the sample was centrifuged at 4000 rpm for 20 minutes. The NaI solution was then removed, and
 175 the residue on the tube wall was rinsed. The remaining soil was rinsed 2-3 times with deionized
 176 water and then treated with 30 mL of $5 \text{ g L}^{-1} \text{ Na}_6\text{P}_6\text{O}_{18}$ solution, followed by shaking at 200 rpm
 177 for 18 hours to disperse aggregates. The dispersion was successively sieved through 250 μm and
 178 53 μm mesh sieves. Different fractions were collected: cPOC $>250 \mu\text{m}$, fPOC 250–53 μm , and
 179 MAOC $<53 \mu\text{m}$. Each fraction was dried at 40°C to constant weight. The carbon content of each
 180 fraction was determined using the $\text{K}_2\text{Cr}_2\text{O}_7$ oxidation method.

181 The contributions of each carbon fractions (SLOCF, POC, and MAOC) to SOC (%) was
 182 estimated using Eq. (4):

$$\begin{aligned} & \text{Proportion of SLOCF, POC, or MAOC (\%)} \\ & = \frac{\text{SOC Fraction (SLOCF, POC, MAOC)}}{\text{SOC}} \times 100 \end{aligned} \quad (4)$$

185 2.5 ^{13}C NMR analysis and spectral indicators

186 The molecular structure of SOC was characterized using Cross-Polarization-Magic Angle
 187 Spinning (CPMAS) solid-state ^{13}C NMR spectroscopy (Bruker AVANCE III 400, Switzerland).
 188 To remove paramagnetic materials and concentrate OC, soil samples were pretreated with 10%
 189 hydrofluoric acid (HF) solution (Schmidt et al., 2011). The processed samples were dried to
 190 constant weight at 40°C , sieved through a 60-mesh sieve, and stored for NMR analysis.

191 The processed samples were analyzed on a Bruker AVANCE III 400 NMR spectrometer
 192 (Switzerland), operating at a ^{13}C resonance frequency of 100 MHz, equipped with a 4 mm
 193 CPMAS probe. The measurement parameters included a 5 kHz MAS rotation rate, 1 ms contact



time, 0.8 s recovery delay, and 140,000 scans. The spectra were divided into 4 chemical shift regions, with each chemical shift region and its corresponding label as follows: alkyl C (0–45 ppm), O-alkyl C (45–110 ppm), aromatic C (110–160 ppm) and carbonyl C (160–220 ppm). The relative intensities of each SOC structure were calculated by integrating the areas of these regions. Three SOC stability indicators were calculated: the aromaticity index (ARM), hydrophobicity index (HI) (Mustafa et al., 2022), and the ratio of alkyl C to O-alkyl C (A/OA) (Kubar et al., 2018).

2.6 Statistical analysis

Statistical analyses were performed using IBM SPSS Statistics 22 software (IBM, Armonk, New York, USA). Significant differences among treatments were analyzed using one-way analysis of variance (ANOVA) with Tukey’s HSD post hoc test for multiple comparisons. Data visualization and Pearson correlation analysis were performed using OriginPro 2023 (OriginLab Corp., Northampton, MA, USA). Pearson correlation coefficients (r) and their significance levels (P) were calculated and visualized as a heatmap. The circle size indicates the strength of the correlation, and the color gradient (red for positive and blue for negative correlations) represents the direction. Statistically significant correlations are marked with asterisks (* for $P < 0.05$, ** for $P < 0.01$, *** for $P < 0.001$). Solid-state ^{13}C NMR spectra were processed and integrated using MestreNova software (Mestrelab Research, Santiago de Compostela, Spain).

The partial least squares structural equation model (PLS-SEM) was used to explore the pathways by which SOC fractions (SLOCF, POC, and MAOC) and structural stability indices of organic carbon (ARM, HI, and A/OA) influence SOC content in greenhouse. Based on our hypotheses, an a priori model was first constructed (Fig. S1). The model was subsequently



216 modified and simplified according to external model quality (heterotrait–monotrait ratio,
 217 composite reliability, Cronbach’s alpha, and convergent validity), internal model quality
 218 (explained variance, R^2), and model fit (standardized root mean square residual, SRMR, and
 219 normed fit index, NFI). The model was constructed using the R package “plspm” (Hair et al.,
 220 2022).

221 **3. Results**

222 **3.1 SOC content and sequestration in the greenhouse**

223 In the greenhouse, SOC content did not continuously increase with the prolonged application
 224 of manure, and the SOC stock remained much lower than the carbon input from manure
 225 application. In the loam, after 20 years of manure application, SOC content stabilized between
 226 33.1–37.0 g kg⁻¹, with a SOC stock of approximately 77.5 Mg C ha⁻¹, and SOC_s at 16.1% (Fig. 1a,
 227 b). Although both SOC content and stock stabilized after 20 years, SOC_s continued to decrease
 228 annually with the prolonged application of manure. After 48–50 years of manure application,
 229 SOC_s decreased to 7.2% (Fig. 1c). In contrast, the effect of manure application on SOC content
 230 was relatively weak in the sandy loam. After 2 years of manure application, the SOC content and
 231 stock in the sandy loam were 13.3 g kg⁻¹ and 33.0 Mg C ha⁻¹, respectively, representing only
 232 42.2% and 42.6% of those in the loam (Fig. 1a, b). Although SOC_s was higher in the sandy loam
 233 (54.9%) after 2 years of manure application, it markedly decreased as the application years
 234 increased. After approximately 30 years, SOC_s in the sandy loam decreased to 4.0%, only 49.2%
 235 of that in the loam (Fig. 1c).

236 **3.2 Dynamic changes in SLOCF, POC, and MAOC in the greenhouse**

237 The application of manure enhanced SOC content of various fractions in the greenhouse.



238 SLOCF took the shortest time to reach a new equilibrium, followed by POC, with MAOC
 239 requiring the longest. Specifically, DOC and MBC contents in both soil textures stabilized after 2
 240 years of manure application in greenhouses. EOC stabilized after 9–11 years in loam (5.77–6.92 g
 241 kg⁻¹), whereas sandy loam required 21–24 years to reach equilibrium (3.67–4.60 g kg⁻¹),
 242 representing only 32% of the loam's (Fig. 2c). In loam, cPOC and fPOC contents reached their
 243 peak (5.01 and 6.80 g kg⁻¹) after 28–31 and 20–23 years of manure application, respectively, and
 244 then slightly decreased (Fig. 2d, f). In sandy loam, cPOC content remained at 1.19–1.62 g kg⁻¹
 245 within 2–32 years; the fPOC content reached its peak (3.08 g kg⁻¹) at 2 years and significantly
 246 decreased to 1.81 g kg⁻¹ at 30–32 years (Fig. 2d, f). In loam, MAOC content gradually stabilized
 247 after 28 years of manure application, and reached highest (21.83 g kg⁻¹) after 48–50 years (Fig.
 248 2h). In sandy loam, MAOC content tended to stabilize after 4 years of manure application,
 249 maintaining at 6.21–7.54 g kg⁻¹ (Fig. 2h). Compared to sandy loam, loam exhibited significantly
 250 higher contents of all OC fractions except DOC and MBC. Sandy loam showed 62.38%, 60.55%,
 251 and 65.41% reductions in cPOC, fPOC, and MAOC contents, respectively, relative to loam (Fig.
 252 2a, b).

253 In the greenhouse, the proportion of OC fractions in SOC follows the order: MAOC >
 254 EOC > fPOC > cPOC > DOC > MBC. In SLOCF, the proportion of DOC and MBC in SOC
 255 peaked at 2 years of manure application in both soil textures and then declined to levels similar to
 256 those of CK. The proportion of EOC in SOC significantly increased in the greenhouse, with the
 257 increase in the sandy loam being significantly higher than that in the loam (Fig. 3a, b). In the loam,
 258 the proportion of cPOC and fPOC in SOC reached 14.1% and 20.0%, respectively, after 20–23
 259 years of manure application, and then decreased to 12.2% and 16.3% after 48–50 years. In



contrast, the proportion of MAOC in SOC gradually increased. Notably, when manure application was less than 48 years, the proportion of MAOC remained lower than that in CK (Fig. 3c). In the sandy loam, the proportion of cPOC and fPOC peaked at 12.7% and 24.5%, respectively, after 2 years of manure application, and then declined. Over 4–32 years, the proportions of cPOC and fPOC remained consistently at 9%–10% and 14%–15%, respectively. MAOC reached its maximum value of 52.3% after 4–6 years of manure application and then stabilized between 47% and 51% thereafter (Fig. 3e).

3.3 Correlations among SOC, SLOCF, POC, and MAOC in the greenhouse

In greenhouses with loam, SOC was highly positively correlated with EOC, cPOC, fPOC, and MAOC ($P < 0.001$). Among them, MAOC showed the strongest correlation with SOC. Additionally, SOC was significantly positively correlated with cPOC/SOC and fPOC/SOC, and negatively correlated with DOC/SOC and MBC/SOC ($P < 0.01$, Fig. 4a). In contrast, SOC was only highly positively correlated with MAOC in the sandy loam ($P < 0.001$). No significant correlation was found between cPOC and either SOC or MAOC, while fPOC was negatively correlated with both SOC and MAOC ($P < 0.05$). The MAOC/SOC was significantly positively correlated with SOC ($P < 0.05$), while cPOC/SOC and fPOC/SOC were significantly negatively correlated with SOC ($P < 0.01$, Fig. 4b).

3.4 Chemical composition and stability of SOC based on ^{13}C NMR analysis

The solid-state ^{13}C NMR spectra and the relative contributions of various carbon functional groups in SOC are shown in Fig. 5a and 5b, and Table 2. The 0–45 ppm region corresponds to the alkyl C group, predominantly consisting of CH_2 bonds, which are characteristic of branched aliphatic C functional groups. The 45–110 ppm region corresponds to the O-alkyl C group,



282 typically associated with methoxy C/N-alkyl compounds, mainly derived from lignin derivatives
 283 such as guaiacol and vanillin, as well as C-N bonds in amino acids. The aromatic C group is
 284 prominent in the 110–160 ppm range and is strongly associated with aromatic C and phenolic C
 285 structures. The 160–220 ppm region is predominantly dominated by the carbonyl C group,
 286 including amides, carboxylic acids, and esters.

287 Long-term manure application significantly altered the distribution of OC functional groups
 288 in greenhouse soils, with notable variations observed between different soil textures (Fig. 5a, b).
 289 Compared to CK, the proportion of alkyl C in the loam decreased by 13.9% at 9–11 years of
 290 manure application, and subsequently increased by 2.6% at 28–31 years. The proportion of
 291 O-Alkyl C exhibited a gradual decline, decreasing by 16.7% at 28–31 years. In contrast, the
 292 proportions of aromatic C and carbonyl C steadily increased, rising by 17.5% and 10.5%,
 293 respectively, at 28–31 years (Table 2). In the sandy loam, manure application increased the
 294 proportion of alkyl C while decreasing the proportions of O-alkyl C and aromatic C. Compared to
 295 CK, the proportion of alkyl C increased by 12.9% at 30–32 years, while the proportions of
 296 O-alkyl C and aromatic C decreased by 6.2% and 5.3%, respectively. The proportion of carbonyl
 297 C remained relatively stable, maintaining at 15.56%–19.11% (Table 2).

298 In the loam, long-term manure application primarily increased ARM, followed by HI and
 299 A/OA. ARM increased with the extension of manure application years. HI increased only between
 300 28–31 years, remaining relatively stable at 2 years and 9–11 years. A/OA decreased at 2 and 9–11
 301 years. Compared to CK, after 28–31 years of manure application in the greenhouse, ARM, HI,
 302 and A/OA increased by 249.44%, 29.39%, and 49.90%, respectively, with an overall increase of
 303 54.61% in SOC stability (Fig. 5c). In contrast, the SOC stability indicators in the sandy loam were



304 significantly lower than those in the loam. In the sandy loam, ARM decreased with the extended
 305 manure application years, while HI and A/OA gradually increased. Compared to CK, although the
 306 ARM of greenhouse soils after 30–32 years of manure application decreased by 0.12, HI and
 307 A/OA increased by 0.29 and 0.99, respectively (Fig. 5d).

308 There were significant differences in the correlation patterns between the two soil textures. In
 309 the loam, SOC and all its fractions were significantly negatively correlated with alkyl C and
 310 O-alkyl C, but showed highly significant positive correlations with aromatic C and carbonyl C (P
 311 < 0.001 , Fig. 6a). All SOC fractions were significantly positively correlated with ARM, but
 312 negatively with A/OA. HI was significantly positively correlated only with SOC, POC, and
 313 MAOC, but not with SLOCF (EOC, DOC, and MBC, Fig. 6a). In contrast, in the sandy loam,
 314 POC showed generally weak correlations with functional groups and stability indices; only cPOC
 315 was significantly positively correlated with alkyl C and HI, and negatively with O-alkyl C (Fig.
 316 6b). SOC, SLOCF, and MAOC were positively correlated with alkyl C, but generally negatively
 317 with O-alkyl C, aromatic C, and carbonyl C. Notably, DOC showed no significant correlation with
 318 either aromatic C or carbonyl C. The strongest correlation was observed between carbonyl C and
 319 MAOC ($P < 0.001$), but correlations with other SOC fractions were weak. SOC, SLOCF, and
 320 MAOC were all significantly positively correlated with HI, but negatively with ARM. A/OA
 321 exhibited significant positive correlations with SOC, EOC, MBC, and MAOC, but was not
 322 significantly associated with DOC (Fig. 6b).

323 3.5 Pathways of long-term SOC storage in greenhouses

324 In loam, SOC content was primarily driven by MAOC and EOC. Although no direct path
 325 was found from POC to SOC, SOC stability indices (ARM and HI) exerted significant positive



effects on POC accumulation, and POC indirectly contributed to SOC content through its promotion of MAOC formation. By contrast, the path from MAOC to EOC was not significant (Fig. 7a). In sandy loam, SOC content was jointly influenced by MAOC and POC. SOC stability (HI) significantly enhanced EOC accumulation, which subsequently promoted MAOC formation and ultimately increased SOC content. However, the path between SOC stability (HI) and cPOC showed relatively low significance, and no direct association was detected between cPOC and MAOC (Fig. 7b).

4. Discussion

4.1 Effects of manure application on SOC composition and chemical structure in greenhouse soils

Manure application in greenhouses significantly increased SOC content, playing a key role in enhancing soil fertility. However, with continuous application, SOC in loam stabilized after 20 years, whereas in sandy loam equilibrium was reached within only 2 years (Fig. 1). This stabilization likely reflects a new balance in SOC, where additional carbon inputs are offset by the higher mineralization rates characteristic of greenhouse soils (Stewart et al., 2007). The difference in equilibrium time can be attributed to soil texture: fine-textured soils (e.g., loam) provide more binding sites that enhance SOC stability and slow decomposition, while sandy loam, with limited binding capacity, undergoes faster SOC decomposition and therefore reaches equilibrium more rapidly (Wiesmeier et al., 2019; Mao et al., 2024).

Manure application significantly increased the contents of SOC fractions in greenhouse soils. Among them, MAOC represented the highest content and proportion (Fig. 2, 3), indicating that MAOC plays a central role in long-term carbon sequestration in greenhouse soils. This finding is



348 consistent with previous studies highlighting the dominant role of MAOC in SOC stabilization.
349 Manure application promotes the complexation of OC with mineral oxides, thereby significantly
350 enhancing the stability of MAOC and effectively facilitating SOC sequestration (Zhang et al.,
351 2011). With prolonged manure application, EOC, cPOC, and fPOC and their relative proportions
352 in SOC tended to stabilize (Fig. 2, 3). This phenomenon can be attributed to the rapid turnover
353 characteristics of EOC and POC combined with greenhouse conditions. Manure-derived EOC and
354 POC are more prone to microbial decomposition under elevated temperatures, and with prolonged
355 manure application, their inputs and losses eventually reach a dynamic balance, thereby
356 suppressing further increases in their contents and relative contributions to SOC (Zheng et al.,
357 2022; Zhou et al., 2024). In contrast to the other carbon fractions, DOC and MBC contents
358 showed no significant differences across manure application durations or soil textures (Fig. 2e, g),
359 accounting for less than 2.5% of SOC, indicating that they are mainly regulated by the amount of
360 manure applied. Although their contributions are limited, they could serve as indicators of
361 microbial metabolism and carbon transformation processes (Francioli et al., 2016; Li et al., 2021;
362 Yan et al., 2023). For example, Wu et al. (2024) reported that organic matter addition altered the
363 relative abundances of Bacteroidetes and Acidobacteria, thereby affecting the expression of key
364 carbon-cycle genes such as cbbL and cbbM, which in turn regulated the distribution of carbon
365 fractions.

366 Although the trends of SOC fraction dynamics were generally consistent between the two
367 soil textures, their distribution patterns differed significantly. Loam exhibited significantly higher
368 levels of EOC, POC, and MAOC, and the sustained increase in the proportion of MAOC with
369 prolonged manure application reflects an enhanced capacity for soil organic matter accumulation



370 and stabilization within persistent carbon pools. In contrast, sandy loam contained lower levels of
371 these fractions, indicating a markedly limited potential for long-term SOC sequestration. This is
372 consistent with the findings of Zhou et al. (2024), who reported that light-textured soils are less
373 effective in sequestering SOC due to insufficient clay content and adsorption sites. A global
374 cropland study by Ling et al. (2025) further demonstrated that fine-textured soils with higher
375 proportions of clay and silt provide more binding and protective sites for organic matter, thereby
376 promoting the formation and stabilization of MAOC. In contrast, sandy soils with limited binding
377 sites tend to store SOC predominantly in free particulate forms, which further explains the
378 observed differences in SOC fractions between loam and sandy loam in this study. Manure
379 application not only altered SOC and its fractions but also significantly modified the composition
380 of SOC functional groups. In the loam, the application of manure decreased the proportion of
381 alkyl C and O-alkyl C in SOC, while increasing the proportions of aromatic C and carbonyl C
382 (Fig. 5a, Table 2). O-alkyl C mainly originates from cellulose, peptides, and other unstable
383 organic components. Alkyl C derives from primary plant bio-polymers and metabolic products of
384 soil microorganisms. In comparison with the more resistant aromatic C and carbonyl C, they can
385 be preferentially biodegraded (Panettieri et al., 2014; Lan et al., 2022). Alkyl C and O-alkyl C
386 gradually transform into more stable forms of OC through oxidation by enzymes secreted by
387 microorganisms and reactive oxygen species. For instance, Yao et al. (2022) found that in the
388 process of OC decomposition by microorganisms, oxidation reactions and free radical reactions in
389 the soil can also convert compounds containing alkyl C into carbonyl-containing compounds
390 (such as ketones and aldehydes), thereby promoting the formation of aromatic C structures. At the
391 early stage of manure application (approximately 2 years), the ARM of SOC in loam increased,



392 while the HI and A/OA were significantly lower than those in the CK (Fig. 5c). This may be due
 393 to the rapid decomposition of large amounts of labile components (e.g., O-alkyl C and alkyl C)
 394 under the high-temperature conditions of greenhouses, leading to a relative increase in the
 395 proportion of aromatic C (Xu et al., 2020). With the extension of manure application years, the
 396 alkyl C and carbonyl C structures in SOC gradually rose, ultimately improving the stability of
 397 SOC.

398 The changes in the proportions of SOC functional groups in the sandy loam exhibited
 399 characteristics distinct from those in the loam. The proportion of alkyl C gradually increased with
 400 prolonged manure application, whereas the proportions of O-alkyl C, aromatic C, and carbonyl C
 401 show an overall decreasing trend (Fig. 5b, Table 2). This is likely because of the lower content of
 402 clay and silt in the sandy loam, which reduces the physical protection of soil particles on SOC
 403 (An et al., 2021; Yao et al., 2022). In sandy loam, manure application significantly altered HI,
 404 A/OA, and ARM, with HI and A/OA increasing but ARM decreasing, in contrast to loam (Fig.
 405 5d). This indicates that, although manure application in the sandy loam does not favor the
 406 formation of the most stable aromatic C, it has a positive effect on SOC stability by enhancing its
 407 hydrophobicity and increasing the proportion of alkyl C. Previous studies have also demonstrated
 408 that alkyl C, as a major contributor to thermal stability and hydrophobicity, can enhance the
 409 resistance of SOC to decomposition in soil (Almendros and González-Pérez, 2025). These results
 410 suggest that loam, due to its higher clay and silt contents, not only facilitates the gradual
 411 transformation of EOC and POC into MAOC but also achieves long-term SOC stabilization
 412 through the accumulation of recalcitrant structures such as aromatic C and carbonyl C. In contrast,
 413 sandy loam, constrained by limited binding sites, exhibits lower levels of SOC fractions and



414 structural stability, relying more on hydrophobic stabilization driven by alkyl C, which results in
 415 markedly lower SOC sequestration potential compared with loam.

416 **4.2 Correlations between SOC fractions and chemical structures and sequestration** 417 **pathways in greenhouse soils**

418 In the loam, the higher clay and silt contents significantly enhanced the capacity for SOC
 419 accumulation, mainly attributable to mineral protection driven by soil texture (Datta et al., 2017;
 420 Yan et al., 2023). Although SLOCF and POC are generally difficult to preserve in soils over the
 421 long term, these fractions continuously associate with mineral particles during decomposition and
 422 form more stable MAOC through processes of adsorption, aggregation, and re-precipitation (Zhou
 423 et al., 2024; Six et al., 2002). Consequently, MAOC showed significant positive correlations with
 424 EOC, cPOC, and fPOC (Fig. 4a). Previous studies have shown that cPOC and fPOC initially
 425 attach to the surfaces of clay and silt particles, gradually forming stable MAOC through
 426 physicochemical interactions such as cation bridging, hydrogen bonding, and van der Waals
 427 forces (Six et al., 2002; Rocci et al., 2021; Zhou et al., 2024). Additionally, SLOCF, particularly
 428 POC derived from DOC, can also enter the POC pool and further enhance SOC stability by
 429 facilitating MAOC formation (Si et al., 2024). These precursor-product linkages directly
 430 influence the long-term stability of SOC in loam (Ji et al., 2024; Si et al., 2024).

431 In contrast, the sandy loam, with its low clay and silt contents, exhibited a weaker SOC
 432 accumulation capacity, where SOC was positively correlated only with EOC and MAOC, but
 433 negatively correlated with fPOC (Fig. 4b). This finding differs from the results of Rocci et al.
 434 (2021), who reported that increases in both MAOC and POC directly contributed to SOC pool
 435 enhancement. However, in the sandy loam soil under greenhouses in this study, the negative



436 correlation between POC and MAOC suggests that the accumulation of POC may not have
 437 contributed to effective stabilization, but rather may have accelerated the decomposition of
 438 existing MAOC. This interpretation is supported by Angst et al. (2023), who indicated that
 439 continuous addition of exogenous organic matter to soils approaching carbon saturation could
 440 trigger a microbial priming effect, thereby accelerating MAOC mineralization. Furthermore, ^{13}C
 441 tracer experiments have shown that low-quality litter (e.g., plant residues with high lignin content
 442 and high C/N ratios) can induce strong priming effects, accelerating MAOC loss and offsetting
 443 the newly formed MAOC (Elias et al., 2024). In addition, Liang et al. (2023) found that in soils
 444 dominated by low-reactivity clays, rhizodeposition-derived bioavailable carbon inputs can deplete
 445 the MAOC pool, indicating that limited mineral protection capacity increases the susceptibility of
 446 MAOC to carbon-induced mobilization.

447 This study also found that the correlations between SOC fractions, functional groups, and
 448 stability indices differed significantly between loam and sandy loam (Fig. 6). Previous studies
 449 have shown that fine-textured soils, particularly within the MAOC fraction, tend to accumulate
 450 microbial transformation products and chemically restructured stable compounds such as carboxyl,
 451 aromatic, and long-chain alkyl groups, whereas coarse-textured soils preferentially retain partially
 452 decomposed plant residues, leading to higher proportions of O-alkyl C (Jindaluang et al., 2013;
 453 Niu et al., 2024; Witzgall et al., 2021). In this study, EOC, POC, and MAOC in loam showed
 454 significant positive correlations with SOC stability indices such as ARM and HI (Fig. 6a),
 455 indicating that even metabolically active fractions may contribute to long-term SOC stabilization
 456 through mineral associations when mineral adsorption sites are abundant. In contrast, the organic
 457 carbon fractions in sandy loam were mainly positively correlated with alkyl C, HI, and A/OA,



458 while fPOC showed no significant relationships with SOC chemical structures or stability indices
459 (Fig. 6b), indicating that these particulate organic carbons were not effectively adsorbed or
460 protected. This finding aligns with Witzgall et al. (2021), who observed in controlled litter
461 decomposition experiments that fine-textured soils preferentially retained microbially restructured
462 SOC, whereas coarse-textured soils released more CO₂ and retained only limited plant-derived
463 fragments.

464 In loam and sandy loam, MAOC represents the primary driver of SOC accumulation, but the
465 contribution pathways of EOC and POC differ markedly (Fig. 7). In loam, although EOC can be
466 partially transformed into residues and incorporated into MAOC through microbial metabolism,
467 its contribution is not significant. This does not imply the absence of such a process; rather, the
468 abundant clay and mineral binding sites in loam make the microbial reprocessing and mineral
469 association of POC the predominant sources of MAOC, thereby masking the effect of EOC
470 incorporation (Liu et al., 2025; Zhou et al., 2024). Therefore, the role of EOC in loam is mainly
471 reflected in its short-term enhancement of SOC content, while the long-term accumulation of
472 MAOC relies more on the transformation of POC and mineral protection. In contrast, in sandy
473 loam, the limited clay and sorption sites markedly reduce the efficiency of POC conversion into
474 MAOC, resulting in POC remaining largely in the form of particulate residues (Christy et al.,
475 2023; Fohrafellner et al., 2024). This explains why there was no significant path detected between
476 POC and MAOC in sandy loam, while POC still directly promoted SOC content through its own
477 accumulation. Under such conditions, the molecular characteristics of EOC, such as small
478 molecules rich in carboxyl and phenolic hydroxyl groups, enable rapid diffusion and preferential
479 occupation of scarce mineral binding sites, making it a key precursor of MAOC (Christy et al.,



2023). This is consistent with several studies reporting that in soils with low fine particle content, inputs of labile or soluble organic carbon contribute relatively more to the stable carbon pool (Peng et al., 2025; Simon et al., 2025). In summary, the results of this study indicate that the long-term stabilization of SOC in greenhouse soils is not the accumulation of a single carbon pool, but rather a synergistic process in which labile carbon fractions serve as substrates that are gradually incorporated into stable pools through microbial metabolism and transformation, thereby achieving long-term preservation.

The results of this study provide guidance for the management of manure application in greenhouse soils with different textures. It should be noted that this study did not include structural characterization of submicron-scale mineral–OC complexes (e.g., X-ray absorption fine structure spectroscopy), and therefore could not distinguish the differences among clay mineral types (e.g., montmorillonite vs. kaolinite) in the formation and stabilization of MAOC. In addition, this study primarily analyzed the relationship between SOC fractions and stabilization, lacking direct evidence on the microbial communities driving carbon transformation. Future studies could integrate metagenomic sequencing techniques to elucidate the community structure and metabolic traits of functional microorganisms that play key roles in carbon transformation across soils with different textures, thereby advancing the understanding of SOC stabilization processes in greenhouses through a multi-scale approach.

5. Conclusions

In greenhouse soils, SOC content gradually stabilized with increasing years of manure application, with loam reaching equilibrium after approximately 20 years, whereas sandy loam attained a new balance within only 2 years. Under equivalent manure inputs, loam contained



502 significantly higher amounts of EOC, POC, and MAOC than sandy loam, with SOC stocks being
 503 approximately 2–3 times higher. In loam, SOC showed significant positive correlations with EOC,
 504 POC, and MAOC, indicating their synergistic role in long-term SOC stabilization. By contrast, in
 505 sandy loam, SOC was positively associated only with EOC and MAOC, but negatively with fPOC,
 506 reflecting the limited capacity of sandy loam to incorporate POC into the stable pool under
 507 constrained clay and sorption sites. Further analysis revealed that manure application in loam
 508 enriched aromatic and carbonyl C, a structural feature conferring greater stability and enhancing
 509 the efficiency of EOC and POC transformation into MAOC, thereby facilitating long-term SOC
 510 accumulation. Conversely, SOC in sandy loam was dominated by hydrophobic alkyl C, where
 511 increases in HI primarily stimulated EOC accumulation, and SOC stabilization largely depended
 512 on the continuous replenishment of EOC and its derivatives. Taken together, these findings
 513 demonstrate that long-term SOC sequestration is not determined solely by the quantity of external
 514 inputs, but is strongly governed by the efficiency with which labile fractions are transformed into
 515 MAOC. Overall, the response of greenhouse SOC to long-term manure application is essentially
 516 regulated by both the distribution of C fractions and the structural stability of their chemical
 517 composition.

518 **Data availability**

519 Data will be made available on request.

520 **Author contribution:**

521 Conceptualization: Yanli Yi, Song Li, Wei Han; Data curation: Boyuan Tan; Formal analysis:
 522 Boyuan Tan, Lu Yang, Xun Xiao, Lingxuan Meng; Funding acquisition: Song Li, Wei Han;
 523 Investigation: Boyuan Tan, Lu Yang, Lingxuan Meng, Jing Tan; Methodology: Boyuan Tan;



524 Project administration: Song Li, Wei Han; Resources: Song Li, Wei Han; Software: Boyuan Tan;
525 Supervision: Yanli Yi, Wei Han; Validation: Boyuan Tan; Visualization: Boyuan Tan, Lu Yang;
526 Writing – original draft: Boyuan Tan; Writing – review & editing: Xue Liu, Fengkui Qiani, Na Li,
527 Chunji Li, Ning An, Song Li, Wei Han.

528 **Competing interests**

529 The authors declare that they have no known competing financial interests or personal
530 relationships that could have appeared to influence the work reported in this paper.

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541 **References**

542 Almendros, G., González-Pérez, J. A.: Soil organic carbon sequestration mechanisms and the
543 chemical nature of soil organic matter—A review, *Sustainability*, 17, 6689,
544 <https://doi.org/10.3390/su17156689>, 2025.



- 545 An, Z., Bernard, G. M., Ma, Z., Plante, A. F., Michaelis, V. K., Bork, E. W., Carlyle, C. N.,
 546 Baah-Acheamfour, M., Chang, S. X.: Forest land-use increases soil organic carbon
 547 quality but not its structural or thermal stability in a hedgerow system, *Agric. Ecosyst.*
 548 *Environ.*, 321, 107617, <https://doi.org/10.1016/j.agee.2021.107617>, 2021.
- 549 Angst, G., Mueller, K. E., Castellano, M. J., Vogel, C., Wiesmeier, M., Mueller, C. W.:
 550 Unlocking complex soil systems as carbon sinks: multi-pool management as the key, *Nat.*
 551 *Commun.*, 14, 2967, <https://doi.org/10.1038/s41467-023-38700-5>, 2023.
- 552 Beillouin, D., Corbeels, M., Demenois, J., Berre, D., Boyer, A., Fallot, A., Feder, F., Cardinael, R.:
 553 A global meta-analysis of soil organic carbon in the Anthropocene, *Nat. Commun.*, 14,
 554 3700, <https://doi.org/10.1038/s41467-023-39338-z>, 2023.
- 555 Christy, I., Moore, A., Myrold, D., Kleber, M.: A mechanistic inquiry into the applicability of
 556 permanganate oxidizable carbon as a soil health indicator, *Soil Sci. Soc. Am. J.*, 87,
 557 1083–1095, <https://doi.org/10.1002/saj2.20569>, 2023.
- 558 Das, A., Purakayastha, T. J., Ahmed, N., Das, R., Biswas, S., Shivay, Y. S., Sehgal, V. K., Rani,
 559 K., Trivedi, A., Tigga, P., et al.: Influence of clay mineralogy on soil organic carbon
 560 stabilization under tropical climate, *Indian J. Soil Sci. Plant Nutr.*, 23, 1003–1018,
 561 <https://doi.org/10.1007/s42729-022-01099-x>, 2023.
- 562 Datta, R., Kelkar, A., Baraniya, D., Molaei, A., Moulick, A., Meena, R. S., Formanek, P.:
 563 Enzymatic degradation of lignin in soil: A Review, *Sustainability*, 9, 1163,
 564 <https://doi.org/10.3390/su9071163>, 2017.
- 565 Denoncourt, C., Chantigny, M. H., Angers, D. A., Maillard, É., Halde, C.: Animal manure
 566 application promotes nitrogen and organic carbon accumulation in soil organic matter



567 fractions: A global meta-analysis, *Sci. Total Environ.*, 996, 180097,
 568 <https://doi.org/10.1016/j.scitotenv.2025.180097>, 2025.

569 Dietz, C. L., Jackson, R. D., Ruark, M. D., Sanford, G. R.: Soil carbon maintained by perennial
 570 grasslands over 30 years but lost in field crop systems in a temperate Mollisol, *Commun.*
 571 *Earth Environ.*, 5, 360, <https://doi.org/10.1016/j.scitotenv.2025.180097>, 2024.

572 Elias, D. M. O., Mason, K. E., Goodall, T., Taylor, A., Zhao, P., Otero-Fariña, A., Chen, H.,
 573 Peacock, C. L., Ostle, N. J., Griffiths, R., et al.: Microbial and mineral interactions
 574 decouple litter quality from soil organic matter formation, *Nat. Commun.*, 15, 10063,
 575 <https://doi.org/10.1038/s41467-024-54446-0>, 2024.

576 Fohrafellner, J., Keiblinger, K. M., Zechmeister-Boltenstern, S., Murugan, R., Spiegel, H.,
 577 Valkama, E.: Cover crops affect pool specific soil organic carbon in cropland – A
 578 meta-analysis, *Eur. J. Soil Sci.*, 75, e13472, <https://doi.org/10.1111/ejss.13472>, 2024.

579 FAO (Food and Agriculture Organization of the United Nations): World reference base for soil
 580 resources 2014: International soil classification system for naming soils and creating
 581 legends for soil maps, World Soil Resources Reports No. 106, FAO, Rome, 2014.

582 Francioli, D., Schulz, E., Lentendu, G., Wubet, T., Buscot, F., Reitz, T.: Mineral vs. organic
 583 amendments: Microbial community structure, activity and abundance of agriculturally
 584 relevant microbes are driven by long-term fertilization strategies, *Front. Microbiol.*, 7,
 585 1446, <https://doi.org/10.3389/fmicb.2016.01446>, 2016.

586 Galluzzi, G., Plaza, C., Giannetta, B., Priori, S., Zaccone, C.: Time and climate roles in driving
 587 soil carbon distribution and stability in particulate and mineral-associated organic matter



- 588 pools, *Sci. Total Environ.*, 963, 178511, <https://doi.org/10.1016/j.scitotenv.2025.178511>,
 589 2025.
- 590 Gross, A., Glaser, B.: Meta-analysis on how manure application changes soil organic carbon
 591 storage, *Sci. Rep.*, 11, 5516, <https://doi.org/10.1038/s41598-021-82739-7>, 2021.
- 592 Hair, J. F., Hult, G. T. M., Ringle, C. M., Sarstedt, M.: A primer on partial least squares structural
 593 equation modeling (PLS-SEM), 3rd ed., Sage Publications, Thousand Oaks, CA, USA,
 594 2022.
- 595 Hao, X., Ma, X., Sun, L., Liu, S., Ji, J., Zhou, B., Zhao, Y., Zheng, Y., Kuang, E., Liu, Y., et al.:
 596 High ratio of manure substitution enhanced soil organic carbon storage via increasing
 597 particulate organic carbon and nutrient availability, *Plants*, 14, 2045,
 598 <https://doi.org/10.3390/plants14132045>, 2025.
- 599 Islam, M. R., Singh, B., Dijkstra, F. A.: Stabilisation of soil organic matter: interactions between
 600 clay and microbes, *Biogeochemistry*, 160, 145–158,
 601 <https://doi.org/10.1007/s10533-022-00956-2>, 2022.
- 602 Ji, Y., Zhao, Y., Han, X., Chen, X., Yan, J., Lu, X., Zhu, Y., Zou, W.: Effect of long-term
 603 fertilization practices on the stability of soil organic matter in the northeast black soil
 604 region in China, *Agronomy*, 14, 2272, <https://doi.org/10.3390/agronomy14102272>, 2024.
- 605 Jindaluang, W., Kheoruenromne, I., Suddhiprakarn, A., Singh, B. P., Singh, B.: Influence of soil
 606 texture and mineralogy on organic matter content and composition in physically
 607 separated fractions soils of Thailand, *Geoderma*, 195–196, 207–219,
 608 <https://doi.org/10.1016/j.geoderma.2012.12.003>, 2013.



- 609 Just, C., Armbruster, M., Barkusky, D., Baumecker, M., Diepolder, M., Döring, T. F., Heigl, L.,
 610 Honermeier, B., Jate, M., Merbach, I., et al.: Soil organic carbon sequestration in
 611 agricultural long-term field experiments as derived from particulate and
 612 mineral-associated organic matter, *Geoderma*, 434, 116472,
 613 <https://doi.org/10.1016/j.geoderma.2023.116472>, 2023.
- 614 Kleber, M., Eusterhues, K., Keiluweit, M., Mikutta, C., Mikutta, R., Nico, P. S.: Chapter one -
 615 Mineral–organic associations: formation, properties, and relevance in soil environments,
 616 in: *Advances in Agronomy*, Vol. 130, 1–140,
 617 <https://doi.org/10.1016/bs.agron.2014.10.005>, 2015.
- 618 Kubar, K. A., Huang, L., Lu, J., Li, X., Xue, B., Yin, Z.: Long-term tillage and straw returning
 619 effects on organic C fractions and chemical composition of SOC in rice-rape cropping
 620 system, *Arch. Agron. Soil Sci.*, 65, 125–137,
 621 <https://doi.org/10.1080/03650340.2018.1490726>, 2018.
- 622 Lan, X., Shan, J., Huang, Y., Liu, X., Lv, Z., Ji, J., Hou, H., Xia, W., Liu, Y.: Effects of long-term
 623 manure substitution regimes on soil organic carbon composition in a red paddy soil of
 624 southern China, *Soil Tillage Res.*, 221, 105395,
 625 <https://doi.org/10.1016/j.still.2022.105395>, 2022.
- 626 Lavallee, J. M., Soong, J. L., Cotrufo, M. F.: Conceptualizing soil organic matter into particulate
 627 and mineral-associated forms to address global change in the 21st century, *Glob. Change*
 628 *Biol.*, 26, 261–273, <https://doi.org/10.1111/gcb.14859>, 2020.



- 629 Li, Y., Li, Z., Cui, S., Liang, G., Zhang, Q.: Microbial-derived carbon components are critical for
 630 enhancing soil organic carbon in no-tillage croplands: A global perspective, *Soil Tillage*
 631 *Res.*, 205, 104758, <https://doi.org/10.1016/j.still.2020.104758>, 2021.
- 632 Liang, G., Stark, J., Waring, B. G.: Mineral reactivity determines root effects on soil organic
 633 carbon, *Nat. Commun.*, 14, 4962, <https://doi.org/10.1038/s41467-023-40768-y>, 2023.
- 634 Ling, J., Dungait, J. A. J., Delgado-Baquerizo, M., Cui, Z., Zhou, R., Zhang, W., Gao, Q., Chen,
 635 Y., Yue, S., Kuzyakov, Y., et al.: Soil organic carbon thresholds control fertilizer effects
 636 on carbon accrual in croplands worldwide, *Nat. Commun.*, 16, 3009,
 637 <https://doi.org/10.1038/s41467-025-57981-6>, 2025.
- 638 Liu, L., Yang, J., Wang, J., Yu, Q., Wei, C., Jiang, L., Huang, J., Zhang, Y., Jiang, Y., Zhang, H., et
 639 al.: Increase in mineral-associated organic carbon does not offset the decrease in
 640 particulate organic carbon under long-term nitrogen enrichment in a steppe ecosystem,
 641 *Soil Biol. Biochem.*, 202, 109695, <https://doi.org/10.1016/j.soilbio.2024.109695>, 2025.
- 642 Liu, Y., Ge, T., van Groenigen, K. J., Yang, Y., Wang, P., Cheng, K., Zhu, Z., Wang, J., Li, Y.,
 643 Guggenberger, G., et al.: Rice paddy soils are a quantitatively important carbon store
 644 according to a global synthesis, *Commun. Earth Environ.*, 2, 154,
 645 <https://doi.org/10.1038/s43247-021-00229-0>, 2021.
- 646 Liu, Y., Zhou, Z., Zhang, X., Xu, X., Chen, H., Xiong, Z.: Net global warming potential and
 647 greenhouse gas intensity from the double rice system with integrated soil–crop system
 648 management: A three-year field study, *Atmos. Environ.*, 116, 92–101,
 649 <https://doi.org/10.1016/j.atmosenv.2015.06.018>, 2015.



- 650 Lu, F., Wang, X., Han, B., Ouyang, Z., Duan, X., Hua, Z., Miao, H.: Soil carbon sequestrations by
 651 nitrogen fertilizer application, straw return and no-tillage in China's cropland, *Glob.*
 652 *Change Biol.*, 15, 281–305, <https://doi.org/10.1111/j.1365-2486.2008.01743.x>, 2009.
- 653 Lu, R. K.: Analysis methods of soil agricultural chemistry, China Agricultural Science and
 654 Technology Press, Beijing, (in Chinese), 2000.
- 655 Mao, H. R., Cotrufo, M. F., Hart, S. C., Sullivan, B. W., Zhu, X., Zhang, J., Liang, C., Zhu, M.:
 656 Dual role of silt and clay in the formation and accrual of stabilized soil organic carbon,
 657 *Soil Biol. Biochem.*, 192, 109390, <https://doi.org/10.1016/j.soilbio.2024.109390>, 2024.
- 658 Mustafa, A., Frouz, J., Naveed, M., Ping, Z., Nan, S., Minggang, X., Núñez-Delgado, A.: Stability
 659 of soil organic carbon under long-term fertilization: Results from ^{13}C NMR analysis and
 660 laboratory incubation, *J. Integr. Agric.*, 205, 112476,
 661 <https://doi.org/10.1016/j.envres.2021.112476>, 2022.
- 662 Niu, Y., Li, Y., Lou, M., Cheng, Z., Ma, R., Guo, H., Zhou, J., Jia, H., Fan, L., Wang, T.: Microbial
 663 transformation mechanisms of particulate organic carbon to mineral-associated organic
 664 carbon at the chemical molecular level: Highlighting the effects of ambient temperature
 665 and soil moisture, *Soil Biol. Biochem.*, 195, 109454,
 666 <https://doi.org/10.1016/j.soilbio.2024.109454>, 2024.
- 667 Oldfield, E. E., Bradford, M. A., Wood, S. A.: Global meta-analysis of the relationship between
 668 soil organic matter and crop yields, *Soil*, 5, 15–32, <https://doi.org/10.5194/soil-5-15-2019>,
 669 2019.
- 670 Panettieri, M., Knicker, H., Murillo, J. M., Madejón, E., Hatcher, P. G.: Soil organic matter
 671 degradation in an agricultural chronosequence under different tillage regimes evaluated



672 by organic matter pools, enzymatic activities and CPMAS ^{13}C NMR, Soil Biol. Biochem.,
 673 78, 170–181, <https://doi.org/10.1016/j.soilbio.2014.07.021>, 2014.

674 Peng, Y., Biswas, A., Adamowski, J. F., Zhang, X., Li, Y., Li, L., Peng, Y., Cao, J.: Sources and
 675 stability of particulate organic matter (POM) and mineral-associated organic matter
 676 (MAOM) on the Loess Plateau: Implications for soil carbon management, Ecol. Indic.,
 677 179, 114162, <https://doi.org/10.1016/j.ecolind.2025.114162>, 2025.

678 Rocci, K. S., Lavalley, J. M., Stewart, C. E., Cotrufo, M. F.: Soil organic carbon response to
 679 global environmental change depends on its distribution between mineral-associated and
 680 particulate organic matter: A meta-analysis, Sci. Total Environ., 793, 148569,
 681 <https://doi.org/10.1016/j.scitotenv.2021.148569>, 2021.

682 Schmidt, M. W. I., Torn, M. S., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I. A., Kleber,
 683 M., Kögel-Knabner, I., Lehmann, J., Manning, D. A. C., et al.: Persistence of soil organic
 684 matter as an ecosystem property, Nat., 478, 49–56, <https://doi.org/10.1038/nature10386>,
 685 2011.

686 Si, Q., Chen, K., Wei, B., Zhang, Y., Sun, X., Liang, J.: Dissolved carbon flow to particulate
 687 organic carbon enhances soil carbon sequestration, Soil, 10, 441–450,
 688 <https://doi.org/10.5194/soil-10-441-2024>, 2024.

689 Simon, C., Miltner, A., Mulder, I., Kaiser, K., Lorenz, M., Thiele-Bruhn, S., Lechtenfeld, O.:
 690 Long-term effects of manure addition on soil organic matter molecular composition:
 691 Carbon transformation as a major driver of energetic potential, Soil Biol. Biochem., 205,
 692 109755, <https://doi.org/10.1016/j.soilbio.2025.109755>, 2025.



- 693 Six, J., Callewaert, P., Lenders, S., De Gryze, S., Morris, S. J., Gregorich, E. G., Paul, E. A.,
 694 Paustian, K.: Measuring and understanding carbon storage in afforested soils by physical
 695 fractionation, *Soil Sci. Soc. Am. J.*, 66, 1981–1987,
 696 <https://doi.org/10.2136/sssaj2002.1981>, 2002.
- 697 Stewart, C. E., Paustian, K., Conant, R. T., Plante, A. F., Six, J.: Soil carbon saturation: concept,
 698 evidence and evaluation, *Biogeochemistry*, 86, 19–31,
 699 <https://doi.org/10.1007/s10533-007-9140-0>, 2007.
- 700 Tan, B., Xiao, X., Yang, L., Meng, L., Li, S., An, N., Lu, Y., Wang, W., Han, W., Yi, Y.:
 701 Mechanisms of soil structure stability in greenhouse: the balancing role of binding and
 702 dispersing agents, *J. Soils Sediments*, 25, 2371–2386,
 703 <https://doi.org/10.1007/s11368-025-04079-7>, 2025.
- 704 Vance, E. D., Brookes, P. C., Jenkinson, D. S.: An extraction method for measuring soil microbial
 705 biomass C, *Soil Biol. Biochem.*, 19, 703–707,
 706 [https://doi.org/10.1016/0038-0717\(87\)90052-6](https://doi.org/10.1016/0038-0717(87)90052-6), 1987.
- 707 Vieira, F. C. B., Bayer, C., Zanatta, J. A., Dieckow, J., Mielniczuk, J., He, Z. L.: Carbon
 708 management index based on physical fractionation of soil organic matter in an Acrisol
 709 under long-term no-till cropping systems, *Soil Till. Res.*, 96, 195–204,
 710 <https://doi.org/10.1016/j.still.2007.06.007>, 2007.
- 711 Wiesmeier, M., Urbanski, L., Hobley, E., Lang, B., von Lützow, M., Marin-Spiotta, E., van
 712 Wesemael, B., Rabot, E., Ließ, M., Garcia-Franco, N., et al.: Soil organic carbon storage
 713 as a key function of soils – a review of drivers and indicators at various scales,
 714 *Geoderma*, 333, 149–162, <https://doi.org/10.1016/j.geoderma.2018.07.026>, 2019.



- 715 Witzgall, K., Vidal, A., Schubert, D. I., Höschen, C., Schweizer, S. A., Buegger, F., Pouteau, V.,
 716 Chenu, C., Mueller, C. W.: Particulate organic matter as a functional soil component for
 717 persistent soil organic carbon, *Nat. Commun.*, 12, 4115,
 718 <https://doi.org/10.1038/s41467-021-24192-8>, 2021.
- 719 Wu, B., Zhang, M., Zhai, Z., Dai, H., Yang, M., Zhang, Y., Liang, T.: Soil organic carbon, carbon
 720 fractions, and microbial community under various organic amendments, *Sci. Rep.*, 14,
 721 25431, <https://doi.org/10.1038/s41598-024-75771-w>, 2024.
- 722 Xu, L., Wang, M., Tian, Y., Shi, X., Shi, Y., Yu, Q., Xu, S., Pan, J., Li, X., Xie, X.: Relationship
 723 between macropores and soil organic carbon fractions under long-term organic manure
 724 application, *Land Degrad. Dev.*, 31, 1344–1354, <https://doi.org/10.1002/ldr.3525>, 2020.
- 725 Yan, H., Fan, W., Wu, J.: Effects of continuous manure application on the microbial community
 726 and labile organic carbon fractions, *Agriculture*, 13, 2096,
 727 <https://doi.org/10.3390/agriculture13112096>, 2023.
- 728 Yang, L., Han, W., Tan, B., Wu, Y., Li, S., Yi, Y.: Effects of nutrient accumulation and microbial
 729 community changes on tomato fusarium wilt disease in greenhouse soil, *Sustainability*,
 730 16, 7756, <https://doi.org/10.3390/su16177756>, 2024.
- 731 Yao, B., Zeng, X., Pang, L., Kong, X., Tian, K., Ji, Y., Sun, S., Tian, X.: The photodegradation of
 732 lignin methoxyl C promotes fungal decomposition of lignin aromatic C measured with
 733 ¹³C-CPMAS NMR, *J. Fungi*, 8, 900, <https://doi.org/10.3390/jof8090900>, 2022.
- 734 Zhang, G., Cao, Z. P., Hu, C. J.: Soil organic carbon fractionation methods and their applications
 735 in farmland ecosystem research: a review, *Chin. J. Appl. Ecol.*, 22, 1921–1930,
 736 <https://www.cjae.net/EN/Y2011/V22/I07/1921>, 2011.



- 737 Zhang, Y., Hagedorn, F., Guidi, C., Yang, L.: Limited soil carbon storage under long-term organic
 738 fertilization in solar greenhouses, *Soil Use Manage.*, 38, 1614–1627,
 739 <https://doi.org/10.1111/sum.12832>, 2022.
- 740 Zhao, Z., Mao, Y., Gao, S., Lu, C., Pan, C., Li, X.: Organic carbon accumulation and aggregate
 741 formation in soils under organic and inorganic fertilizer management practices in a
 742 rice-wheat cropping system, *Sci. Rep.*, 13, 3665,
 743 <https://doi.org/10.1038/s41598-023-30541-y>, 2023.
- 744 Zheng, W., Rao, C., Wu, Q., Wang, E., Jiang, X., Xu, Y., Hu, L., Chen, Y., Liang, X., Yan, W.:
 745 Changes in the soil labile organic carbon fractions following bedrock exposure rate in a
 746 karst context, *Forests*, 13, 516, <https://doi.org/10.3390/f13040516>, 2022.
- 747 Zhou, Z., Ren, C., Wang, C., Delgado-Baquerizo, M., Luo, Y., Luo, Z., Du, Z., Zhu, B., Yang, Y.,
 748 Jiao, S., et al.: Global turnover of soil mineral-associated and particulate organic carbon,
 749 *Nat. Commun.*, 15, 5329, <https://doi.org/10.1038/s41467-024-49743-7>, 2024.
- 750



751 **Table 1.** Manure application years and soil mechanical composition at greenhouse sampling sites.

Sampling Sites	Manure application years	Points	Soil bulk density g cm ⁻³	pH	Sand %	Silt %	Clay %	Soil texture
				1:2.5				
Wujianfang Town	CK	3	1.12±0.03	7.06±0.23	43.6±1.5	21.0±1.8	35.4±1.7	Loam
	2	3	1.19±0.06	6.98±0.22	43.7±3.0	20.7±3.3	34.9±0.9	
	9-11	4	1.13±0.03	7.05±0.19	42.3±1.6	20.9±2.8	36.8±4.1	
	20-23	3	1.15±0.04	7.00±0.15	42.6±1.5	23.3±1.9	34.1±2.0	
	28-31	4	1.10±0.04	6.93±0.20	43.5±1.2	20.0±2.3	36.5±1.2	
	38-40	3	1.07±0.03	6.36±0.21	43.2±1.7	20.7±2.5	36.1±2.2	
	48-50	3	1.14±0.06	6.60±0.13	44.4±1.0	19.6±0.9	36.0±1.3	
Yangshipu Town	CK	3	1.24±0.03	6.42±0.24	70.9±2.1	8.2±1.9	20.9±2.8	Sandy Loam
	2	3	1.25±0.03	6.84±0.23	70.1±3.0	8.6±0.9	21.3±3.1	
	4-6	3	1.27±0.01	7.02±0.33	70.6±2.6	8.3±2.7	21.1±0.9	
	11-13	3	1.23±0.02	7.15±0.28	70.43±2.6	9.4±0.9	20.2±2.1	
	21-24	3	1.21±0.01	7.09±0.06	69.4±2.7	7.1±1.5	23.5±4.2	
	30-32	4	1.25±0.04	7.14±0.24	68.7±2.6	7.5±1.4	23.9±3.5	

752

753 **Table 2.** Percentage (%) contributions of C functional groups based on integrated CPMAS ¹³C NMR spectroscopy of bulk soil in the greenhouse.

Soil texture	Treatment	Alkyl C (0-45 ppm)	O-Alkyl C (45-110 ppm)			Aromatic C (110-160 ppm)			Carbonyl C (160-220 ppm)			
			N-alkyl/methoxyl	O-alkyl	di-O-alkyl	Total	Aryl C/O-aryl	Phenolic-C	Total	Amide-C	Ketone-C	Total
	years		C	C	C							
Loam	CK	42.40	8.70	20.82	2.31	31.83	7.88	5.90	13.78	6.75	5.24	11.99
	2 yr	34.20	7.78	19.73	4.29	31.80	10.56	7.75	18.31	8.89	6.79	15.68
	9-11 yr	28.55	4.71	14.31	4.22	23.24	16.62	9.54	26.16	12.81	9.25	22.05
	28-31 yr	31.16	3.81	7.74	3.60	15.15	19.89	11.34	31.23	12.82	9.65	22.46
Sandy	CK	25.67	6.67	22.34	6.63	35.65	13.92	5.65	19.57	13.00	6.11	19.11
	2 yr	30.58	8.16	17.72	6.73	32.62	13.91	4.51	18.42	14.23	4.16	18.39
Loam	11-13 yr	38.54	8.90	13.05	7.46	29.42	12.01	2.30	14.31	13.66	4.07	17.73
	30-32 yr	37.54	8.88	13.14	7.50	29.52	12.13	2.45	14.58	12.88	4.86	17.74

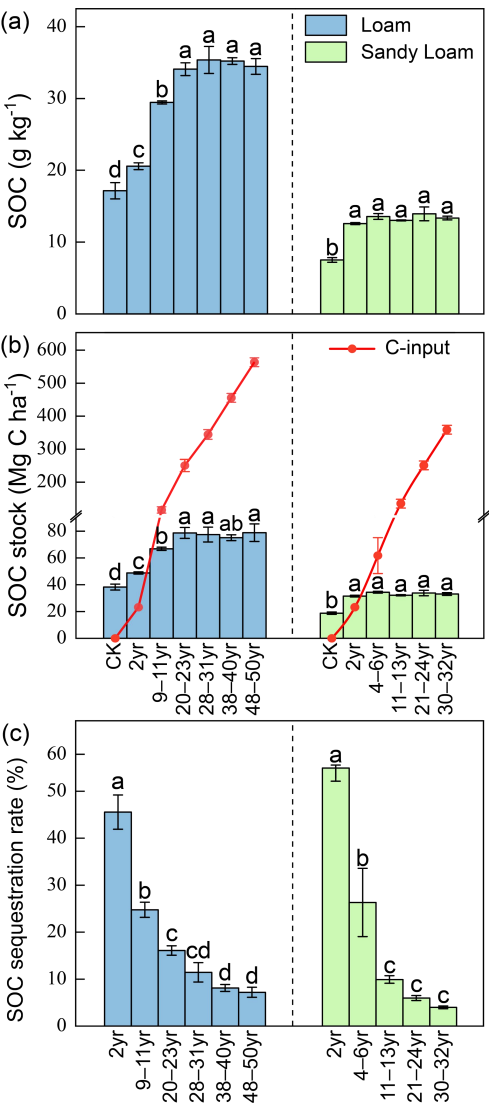


Fig. 1. Changes in (a) soil organic carbon (SOC), (b) manure-derived carbon input and SOC stock, and (c) SOC sequestration rate following manure application over various years in greenhouses. Bars represent the standard errors. Within each soil texture, bars with different lowercase letters indicate significant differences among years of manure application ($P < 0.05$).

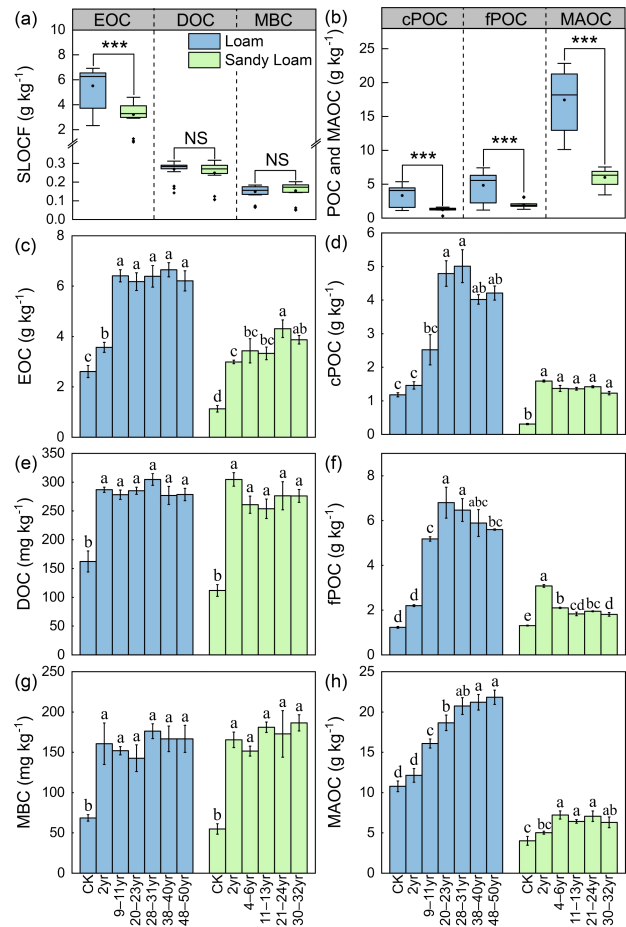


Fig. 2. Content and variations of (a) soil labile organic carbon fractions (SLOCF), (b) particulate organic carbon (POC), and mineral-associated organic carbon (MAOC) in greenhouse soils (loam and sandy loam); and dynamic changes in (c) easily oxidizable organic carbon (EOC), (e) dissolved organic carbon (DOC), (g) microbial biomass carbon (MBC), (d) coarse particulate organic carbon (cPOC), (f) fine particulate organic carbon (fPOC), and (h) mineral-associated organic carbon (MAOC) across different years of manure application in greenhouse soils. Bars represent the standard errors. Within each soil texture, bars with different lowercase letters indicate significant differences among years of manure application ($P < 0.05$).

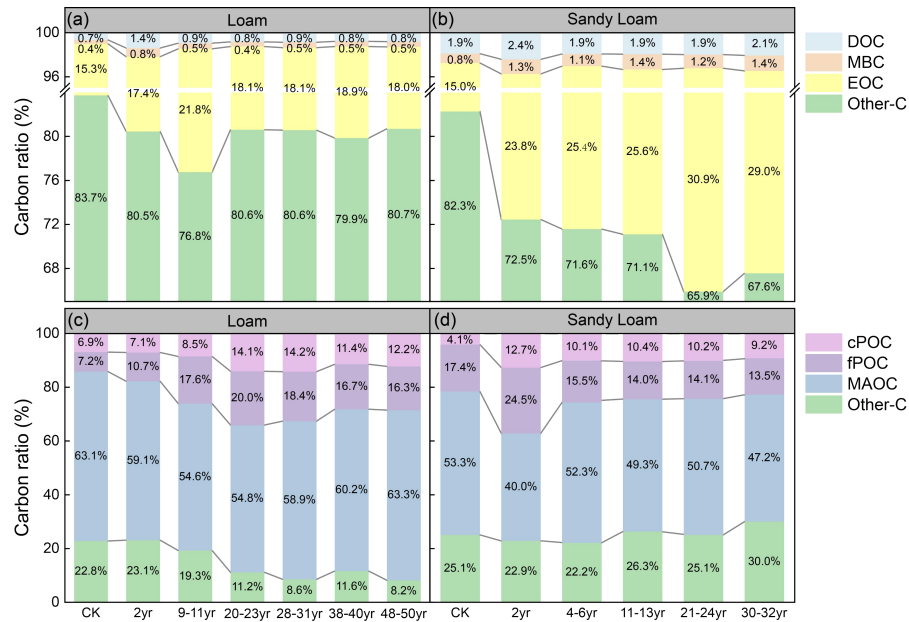


Fig. 3. Proportions of soil labile organic carbon fractions (SLOCF: easily oxidizable organic carbon [EOC], microbial biomass carbon [MBC], and dissolved organic carbon [DOC]), particulate organic carbon (POC: coarse POC [cPOC] and fine POC [fPOC]), and mineral-associated organic carbon (MAOC) to soil organic carbon (SOC) in greenhouse soils under different manure application years in (a, c) loam and (b, d) sandy loam.

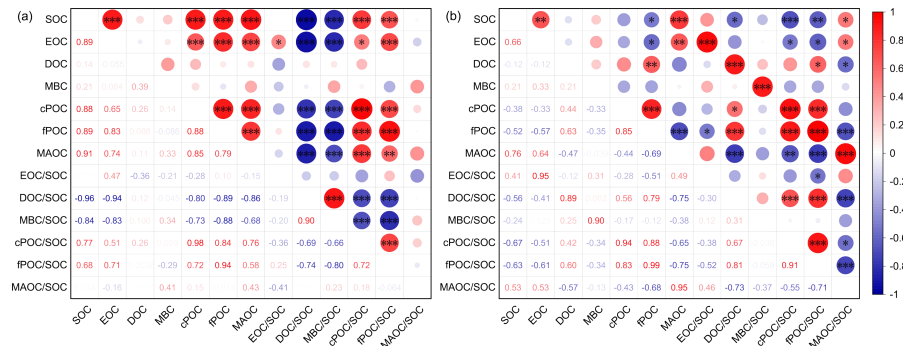


Fig. 4. Pearson correlations among soil organic carbon (SOC) fractions and their relative proportions in greenhouses for (a) loam and (b) sandy loam. Note: SOC fractions include total SOC, easily oxidizable organic carbon (EOC), dissolved organic carbon (DOC), microbial biomass carbon (MBC), coarse particulate organic carbon (cPOC), fine particulate organic carbon (fPOC), and mineral-associated organic carbon (MAOC). Their relative proportions to SOC include EOC/SOC, DOC/SOC, MBC/SOC, cPOC/SOC, fPOC/SOC, and MAOC/SOC. * Correlation was significant at $P \leq 0.05$; ** Correlation was significant at $P \leq 0.01$; *** Correlation was significant at $P \leq 0.001$.

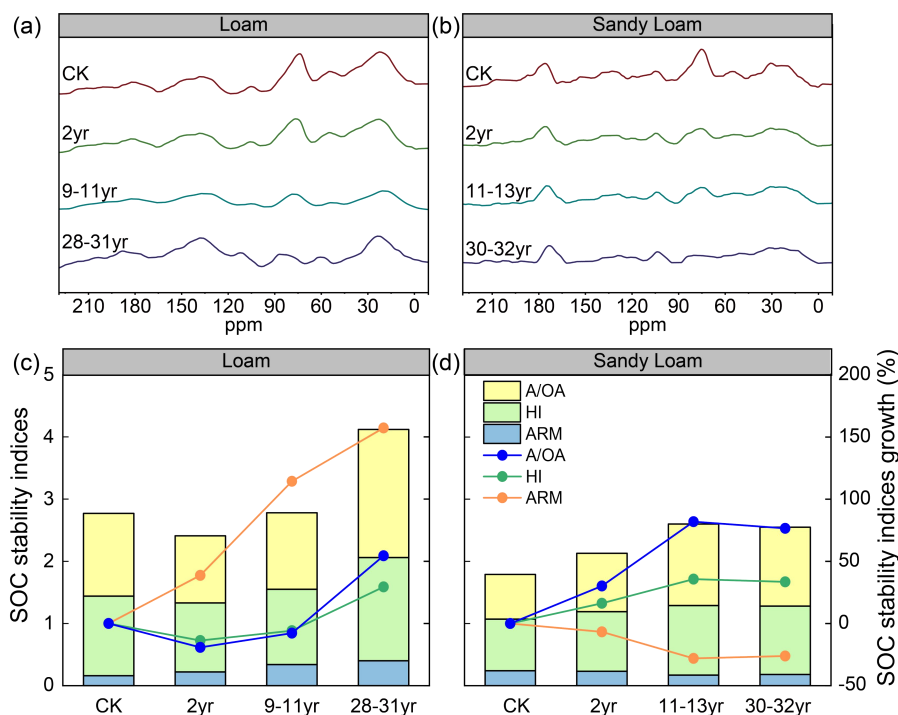


Fig. 5. Solid-state CPMAS ^{13}C NMR spectroscopy of greenhouse soils under varying years of manure application in (a) loam and (b) sandy loam. Soil organic carbon (SOC) stability indicators during different manure application years in (c) loam and (d) sandy loam under greenhouse conditions. A/OA: the ratio of alkyl C to O-alkyl; HI: hydrophobicity index; ARM: aromaticity index.

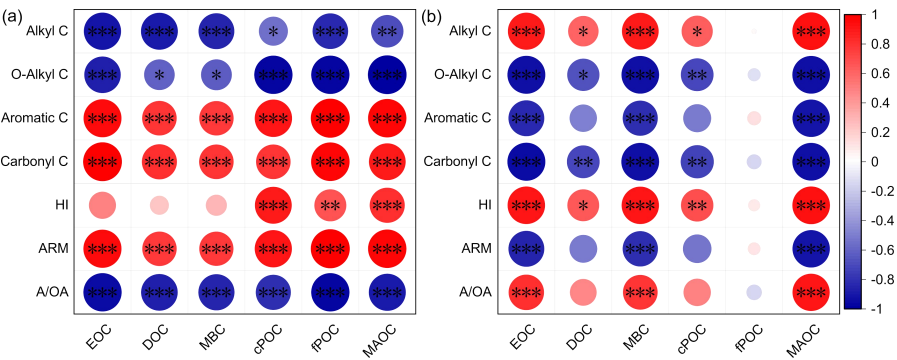


Fig. 6. Pearson correlations between soil organic carbon (SOC) fractions and SOC chemical structure and stability indices in greenhouses (a) loam and (b) sandy loam. Note: SOC fractions include total SOC, easily oxidizable organic carbon (EOC), dissolved organic carbon (DOC), microbial biomass carbon (MBC), coarse particulate organic carbon (cPOC), fine particulate organic carbon (fPOC), and mineral-associated organic carbon (MAOC). SOC stability indices include aromaticity index (ARM), humification index (HI), and the ratio of alkyl C to O-alkyl (A/OA). * Correlation was significant at $P \leq 0.05$; ** Correlation was significant at $P \leq 0.01$; *** Correlation was significant at $P \leq 0.001$.

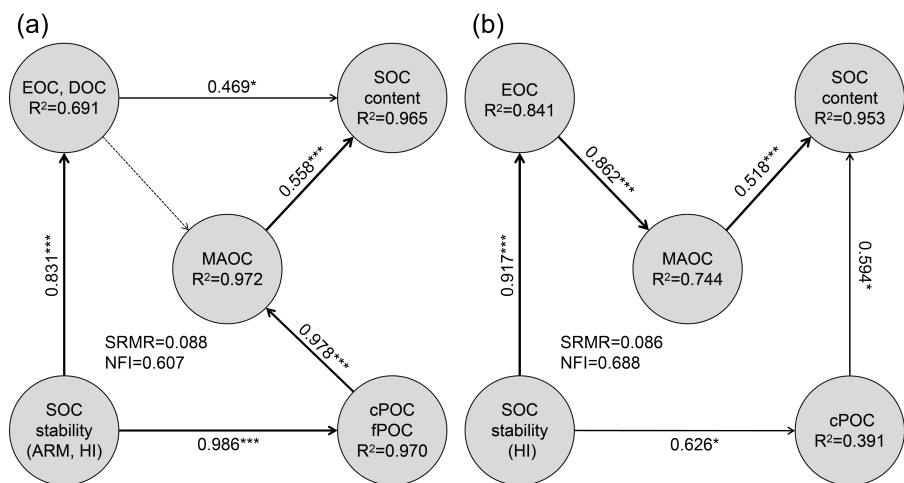


Fig. 7. Partial least squares structural equation models (PLS-SEM) illustrating the pathways through which soil organic carbon (SOC) chemical stability indices influence SOC fractions and SOC content in greenhouse soils: (a) loam; (b) sandy loam. Numbers on arrows denote path coefficients, while numbers adjacent to arrows represent weight contributions. Solid and dashed lines indicate significant and non-significant relationships, respectively. Note: SOC fractions include easily oxidizable organic carbon (EOC), dissolved organic carbon (DOC), coarse particulate organic carbon (cPOC), fine particulate organic carbon (fPOC), and mineral-associated organic carbon (MAOC). SOC chemical stability indices include aromaticity index (ARM) and humification index (HI).