



1       **Unexpected enhancement of new particle formation by lactic acid**  
2       **sulfate resulting from SO<sub>3</sub> loss in forested and agricultural regions**

3               **Rui Wang, Shuqin Wei<sup>‡</sup>, Zeyao Li<sup>‡</sup>, Kaiyu Xue, Rui Bai, Tianlei Zhang\***  
4       *Shaanxi Key Laboratory of Catalysis, School of Chemical & Environment Science, Shaanxi*  
5       *University of Technology, Hanzhong, Shaanxi 723001, P. R. China*

6       **Abstract**

7       Organosulfates (OSs) are key components of atmospheric aerosols and serve as tracers for  
8       secondary organic aerosol (SOA) formation. Among these, lactic acid sulfate (LAS) has been  
9       increasingly detected in the atmosphere. However, its molecular formation pathways and its role in  
10      new particle formation (NPF) remain poorly understood. In this work, we investigate the gas-phase  
11      formation mechanism of LAS via the reaction between lactic acid (LA) and SO<sub>3</sub>, and assess its  
12      impact on sulfuric acid-ammonia (SA-A) driven NPF using quantum chemical calculations and  
13      Atmospheric Cluster Dynamics Code (ACDC) kinetic modeling. Our results show that SA and H<sub>2</sub>O  
14      significantly catalyze the LA-SO<sub>3</sub> reaction, enhancing the effective rate coefficient by 7-10 orders  
15      of magnitude within the temperature range of 280-320 K. Further molecular-level analysis using the  
16      ACDC reveals that LAS not only significantly enhances the clustering stability of SA and A up to  
17      10<sup>8</sup>-fold, but also plays a significant and direct role in SA-A nucleation under conditions typical of  
18      forested and agricultural regions. Notably, LAS-SA-A clusters contribute to 97% of the overall  
19      cluster formation pathways in regions with high LAS concentrations like Centreville, Alabama.  
20      Additionally, our findings show that the nucleation potential of LAS-SA-A clusters is stronger than  
21      that of LA-SA-A clusters, aligning with field observations, even though LAS concentrations are  
22      typically three orders of magnitude lower than LA. These findings imply that OSs formed through  
23      SO<sub>3</sub> consumption may significantly contribute to the enhanced NPF rates observed in continental  
24      regions.

\* Corresponding authors Tel: +86-0916-2641083, Fax: +86-0916-2641083.

E-mail addresses: ztianlei88@163.com (T. L Zhang)

<sup>‡</sup> Shuqin Wei and Zeyao Li contributed equally to this work.



## 25    **1 Introduction**

26        Atmospheric aerosol particles pose significant risks to public health, adversely affecting both  
27        the respiratory and cardiovascular systems (Anderson et al., 2012; Xing et al., 2016; Zhang et al.,  
28        2023b). Beyond health implications, these particles contribute to global warming by reducing  
29        visibility and disrupting the Earth's radiative balance (Lund et al., 2019; Zheng et al., 2018). As a  
30        major source of atmospheric aerosols, new particle formation (NPF), accounts for over 50% of the  
31        total particle number concentration and is strongly associated with severe haze events in megacities  
32        across China (Kulmala et al., 2004; Brean et al., 2020). Despite its significance, accurately  
33        characterizing the NPF process remains a considerable challenge, primarily due to limitations in  
34        current measurement techniques and an incomplete comprehension of the underlying mechanisms.  
35        While field observations and CLOUD chamber experiments (Kulmala et al., 2004; Dai et al., 2023;  
36        Lee et al., 2019; Hirsikko et al., 2011; Zhang et al., 2015) have provided valuable insights, they are  
37        insufficient to fully elucidate these processes. To address these gaps, a molecular-level approach is  
38        essential, as it allows for a more precise understanding of nucleation mechanisms (Yang et al., 2021;  
39        Li et al., 2017). This approach enables the detailed determination of molecular cluster geometries,  
40        the strengths of intermolecular interactions, and the pathways of cluster formation (Long et al., 2013;  
41        Zu et al., 2024b; Rong et al., 2020b). Such molecular insights are critical to evaluating the impacts  
42        of aerosols on the atmosphere and for devising effective strategies to mitigate haze formation.

43        Gaseous sulfuric acid (SA), derived from the oxidation of SO<sub>2</sub>, has long been recognized as a  
44        key NPF precursor (Kirkby et al., 2011; Zhao et al., 2024). Molecular-level studies have shown that  
45        various nucleation precursors, including water (H<sub>2</sub>O) (Zhang et al., 2012b), ammonia (NH<sub>3</sub>) (Kirkby  
46        et al., 2011; Zhang et al., 2015), methylamine (MA) (Shen et al., 2020), dimethylamine (DMA) (Cai  
47        et al., 2021; Kurtén et al., 2008), monoethanolamine (MEA) (Shen et al., 2019), piperazine (PZ)  
48        (Ma et al., 2019) and iodic acid (Sipilä et al., 2016), are involved in SA-driven binary nucleation,  
49        which serves as a primary initiator of NPF. However, binary nucleation mechanisms alone cannot  
50        fully account for the discrepancies observed between measured and modeled global NPF rates  
51        (Hodshire et al., 2019; Kirkby et al., 2016), suggesting the involvement of additional gaseous  
52        species. In response to this, several studies have explored the role of ternary nucleation driven by  
53        SA-A, which involves a broader array of atmospheric species, including ammonia (NH<sub>3</sub>) (Li et al.,



2020b; Yin et al., 2021a), organic amines (Li et al., 2017; Li et al., 2018a), organic and inorganic acids (Wang et al., 2011; Liu et al., 2018), and highly oxidized multifunctional compounds (HOMs) (Liu et al., 2021a; Liu et al., 2019a; Ning and Zhang, 2022; Yin et al., 2021b; Zhang et al., 2018). Despite recognizing the enhancement provided by SA-A-driven ternary nucleation, the nucleation rates predicted by these mechanisms still fall short when compared to field observations (Kirkby et al., 2016; Hodshire et al., 2019; Yin et al., 2021a). The persistent underestimation underscores the need for further investigation into the role of additional gaseous species to better understand the complex mechanisms driving NPF.

Organosulfates (OSs), formed through the chemical transformation of organic acids, constitute a major portion of organosulfur species in atmospheric aerosols, contributing 5-30% to the organic mass in PM<sub>10</sub> (Sun et al., 2025; Brüggemann et al., 2017). These compounds are prevalent in atmospheric particles and are commonly employed as markers to track the formation of secondary organic aerosols (SOAs) in environmental research (Tan et al., 2022; Zhang et al., 2012a; Froyd et al., 2010a; Brüggemann et al., 2017; Mutzel et al., 2015; Glasius et al., 2017). Recent research has led to the identification and characterization of various OSs in fine particulate matter samples from regions including the United States, China, Mexico City and Pakistan (Hettiyadura et al., 2017; Wang et al., 2018; Olson et al., 2011). Meanwhile, studies suggest that the cycloaddition of SO<sub>3</sub> to organic acids could be a key mechanism for OSs formation resulting in compounds with lower vapor pressures than their parent carboxylic acids and increased inter-molecular interaction sites (Smith et al., 2020; Tan et al., 2020; Yao et al., 2020; Zhang et al., 2023a). Notably, lactic acid sulfate (LAS) has been identified as the dominant OSs species across all these field observations (Darer et al., 2011; Riva et al., 2015; Kundu et al., 2013). However, the specific formation mechanism of LAS from the reaction of lactic acid (LA) with SO<sub>3</sub> remains largely unexplored. Additionally, SA and water (H<sub>2</sub>O) (Tan et al., 2022; Zhang et al., 2025; Li et al., 2018b), both prevalent in the atmosphere, act as strong hydrogen atom donors/acceptors, facilitating proton transfer reactions and potentially catalyzing the LA-SO<sub>3</sub> reaction.

The reaction products of SO<sub>3</sub> with major atmospheric trace species have been shown proven to significantly influence the formation of NPF. For instance, compounds such as sulfamic acid (Li et al., 2018a), oxalic sulfuric anhydride (Yang et al., 2021), methyl hydrogen sulfate (Liu et al., 2019a), glyoxylic sulfuric anhydride (Rong et al., 2020a) and formic acid sulfate (Wang et al., 2025),



84 generated through reactions of  $\text{SO}_3$  with ammonia, oxalic acid, methanol, glyoxylic acid and formic  
85 acid, all exhibit catalytic effects on NPF in aerosols. Structurally, LAS, the product of the  $\text{SO}_3 + \text{LA}$   
86 reaction, contains both  $-\text{COOH}$  and  $-\text{SO}_3\text{H}$  functional groups, which facilitate additional hydrogen  
87 bonding with atmospheric particle precursors (Yao et al., 2020). However, the role of LAS in  
88 enhancing SA-A nucleation remains underexplored, limiting our ability to comprehensively  
89 evaluate its impact on NPF processes. Furthermore, LA, a highly oxidized  $\alpha$ -hydroxy acid with both  
90  $-\text{OH}$  and  $-\text{COOH}$  groups (Mochizuki et al., 2019), can enhance the stability of SA-A clusters and  
91 facilitate NPF (Li et al., 2017). Given its relatively larger atmospheric concentrations, particularly  
92 in regions with elevated organic acid pollution, LA may also significantly influence NPF. So,  
93 understanding the distinct contributions of LAS and LA to SA-A nucleation is crucial, as this will  
94 advance our understanding of NPF events, particularly in agricultural and forested regions.

95 In this work, we utilized quantum chemical calculations together with master equation analysis  
96 to investigate the gas-phase reaction of  $\text{SO}_3$  with LA that forms LAS, with  $\text{H}_2\text{O}$  and SA serving as  
97 catalysts. The role of LAS in enhancing SA-A nucleation was then explored by examining the  
98 formation mechanisms of the  $(\text{LAS})_x(\text{SA})_y(\text{A})_z$  ( $0 \leq z \leq x + y \leq 3$ ) system using the Atmospheric  
99 Clusters Dynamic Code (ACDC) kinetic model. Additionally, the potential influence of LAS on  
100 atmospheric nucleation and particle formation (NPF) was assessed across diverse global regions.  
101 Finally, a comparative study of LA and LAS was also conducted to elucidate the respective roles of  
102 organic acids and OSs in enhancing SA-A nucleation, focusing on the formation mechanisms of  
103 both LA-SA-A and LAS-SA-A systems.

## 104 **2 Methodology**

### 105 **2.1 Quantum chemical calculations**

106 The gas-phase reaction of  $\text{SO}_3$  with LA to form LAS, both in the absence and presence of  
107  $\text{H}_2\text{O}$  and SA as catalysts, was systematically optimized and calculated using the Gaussian 09  
108 program (Faloona et al., 2009) at the M06-2X/6-311++G(2df,2pd) level (Stewart, 2007; Walker  
109 et al., 2013). Intrinsic reaction coordinate analyses (Hratchian and Schlegel, 2005) were carried  
110 out at the same computational level to verify the connection between transition states and their  
111 respective pre-reactive complexes and products. Furthermore, single-point energy calculations  
112 were refined at the CCSD(T)-F12/cc-pVDZ-F12 level with the ORCA program (Neese, 2012),



113 employing the optimized geometries as input.

114 To identify the global minimum energy configurations of  $(\text{SA})_x(\text{A})_y(\text{LAS})_z$  clusters ( where 0  
115  $\leq y \leq x + z \leq 3$ ), we utilized the ABCluster program (Zhang and Dolg, 2016) to systematically  
116 generate initial structures for various clusters combinations. Specifically, using the ABCluster  
117 procedure and the CHARMM force field, a diverse set of initial structures  $n \times 1000$  ( $1 < n \leq 4$ ) were  
118 randomly produced. Initially, the primary structures were optimized and their energies were ranked  
119 using the PM6 method in MOPAC 2016 (Partanen et al., 2016; Stewart, 2007). After the initial  
120 sampling, considering the excellent performance of the M06-2X method in accurately  
121 characterizing the geometries of atmospheric clusters (Walker et al., 2013; Lu et al., 2020), up to  
122 1000 favorable configurations were selected for rigorous re-optimization at the M06-2X/3-21G\*  
123 level of theory. Subsequently, the 100 lowest-energy configurations were further optimized using  
124 the M06-2X/6-31G(*d, p*) level of theory, from which the 10 configurations with the lowest energies  
125 were identified. Finally, to accurately determine the global minimum, the M06-2X/6-311++G(2*df*,  
126 2*pd*) method was applied to refine these 10 lowest-energy configurations.

## 127 2.2 Rate coefficients calculations

128 Rate constants for the  $\text{SO}_3 + \text{LA}$  reaction, both without and with  $\text{H}_2\text{O}$  and  $\text{H}_2\text{SO}_4$  as catalysts,  
129 were determined via Rice-Ramsperger-Kassel-Marcus (RRKM) theory (Glowacki et al., 2012;  
130 Wardlaw and Marcus, 1984) within the Master Equation (ME/RRKM) framework in MESMER  
131 (Master Equation Solver for Multi-Energy Well Reactions) code (Glowacki et al., 2012;  
132 Klippenstein and Marcus, 1988). Specifically, in the MESMER calculations, the rate constants for  
133 the barrierless formation of pre-reactive complexes from reactants were determined using the  
134 Inverse Laplace Transform (ILT) method (Horváth et al., 2020), whereas the subsequent conversion  
135 of these complexes to products via transition states was evaluated using RRKM theory (Mai et al.,  
136 2018). The ILT method and RRKM theory can be represented in Eqs. (1) and (2), respectively:

$$137 \quad k^{\text{rr}}(\beta) = \frac{1}{Q(\beta)} \int_0^\infty k(E) \rho(E) \exp(-\beta E) dE \quad (1)$$

$$138 \quad k(E) = \frac{W(E-E_0)}{h\rho(E)} \quad (2)$$

139 Here,  $h$  represents Planck's constant,  $\rho(E)$  indicates the density of accessible states for the reactant  
140 at energy  $E$ ,  $E_0$  is the reaction threshold energy and  $W(E-E_0)$  refers to the rovibrational states of the



141 transition state, excluding motion along the reaction coordinate. Geometries, vibrational frequencies,  
142 and rotational constants were obtained at the M06-2X/6-311++G(2df,2pd) level, with single-point  
143 energies refined at the method of CCSD(T)-F12/cc-pVDZ-F12.

### 144 2.3 ACDC kinetics simulation

145 The ACDC was utilized to investigate the molecular-level collision coefficient ( $\beta$ ,  $\text{cm}^3 \text{s}^{-1}$ ),  
146 evaporation coefficient ( $\gamma$ ,  $\text{s}^{-1}$ ) and cluster formation rates ( $J$ ,  $\text{cm}^{-3} \text{s}^{-1}$ ). Thermodynamic parameters  
147 and structural information for cluster formation, obtained from quantum chemical calculations  
148 performed by M06-2X/6-311++G(2df,2pd), served as input parameters for the ACDC model. The  
149 MATLAB-R2014a platform, leveraging its odel5s solver (Shampine and Reichelt, 1997), performed  
150 numerical integration of the birthdeath equation for the ACDC model, thereby elucidating the  
151 kinetics of cluster growth over time. The general form of the birth-death equation for the  
152 concentration  $c_i$  of cluster  $i$  given by,

$$153 \quad \frac{dc_i}{dt} = \frac{1}{2} \sum_{j < i} \beta_{j,(i-j)} c_j c_{(i-j)} + \sum_j \gamma_{(i+j) \rightarrow i} c_{i+j} - \sum_j \beta_{i,j} c_i c_j - \frac{1}{2} \sum_{j < i} \gamma_{i \rightarrow j} c_i + Q_i - S_i \quad (3)$$

154 In this formulation,  $\beta_{i,j}$  corresponds to the collision frequency factor between clusters of sizes  $i$  and  
155  $j$ ,  $\gamma_{(i+j) \rightarrow i}$  quantifies the fragmentation rate of composite clusters into their constituent monomers  $i$   
156 and  $j$ . The system's open nature is accounted for through  $Q_i$ , representing the external flux of cluster  
157  $i$ , and  $S_i$ , characterizing its removal rate. Here, the condensation sink coefficient was assigned  $2.6 \times$   
158  $10^{-3}$  (Liu et al., 2021b).

## 159 3 Results and discussions

### 160 3.1 Formation of LAS via the reaction of $\text{SO}_3$ with LA

161 In the direct cycloaddition pathway (Channel LAS) illustrated in Fig. 1, the hydroxyl (-OH)  
162 group of LA reacts with the sulfur atom of  $\text{SO}_3$ , leading to the formation of LAS via proton transfer  
163 from LA to  $\text{SO}_3$ . However, the resulting  $\text{SO}_3 \cdots \text{LA}$  complex (denoted as IM) is thermodynamically  
164 unstable, primarily due to the significant ring strain in the four-membered structure, exhibiting a  
165 relative Gibbs free energy of  $5.6 \text{ kcal} \cdot \text{mol}^{-1}$ . The Gibbs free energy barrier for this reaction is  
166 calculated to be  $22.3 \text{ kcal} \cdot \text{mol}^{-1}$ . As indicated in Table S5, the rate coefficients for Channel LAS are  
167 extremely low, spanning from  $1.35 \times 10^{-26}$  to  $6.21 \times 10^{-25} \text{ cm}^3 \cdot \text{molecule}^{-1} \cdot \text{s}^{-1}$  across the temperature  
168 range of 230-320 K. These values suggest that this pathway is both slow and thermodynamically



169 unfavorable for LAS formation under typical atmospheric conditions.

170 H<sub>2</sub>O, highly abundant in the atmosphere with concentration around 10<sup>17</sup> molecules·cm<sup>-3</sup>  
171 (Huang et al., 2015; Tan et al., 2022), serves as both a donor and acceptor of hydrogen bonds, and  
172 is widely recognized for its ability to catalyze a wide range of proton transfer reactions. To assess  
173 its catalytic effect on the formation of LAS, we examined the SO<sub>3</sub> + LA reaction in the presence of  
174 H<sub>2</sub>O (Channel WM), as illustrated in Fig. 1. This reaction can proceed via three possible sequential  
175 bimolecular pathways: (i) SO<sub>3</sub>···LA + H<sub>2</sub>O, (ii) SO<sub>3</sub>···H<sub>2</sub>O + LA and (iii) LA···H<sub>2</sub>O + SO<sub>3</sub>.  
176 Considering typical atmospheric concentrations of SO<sub>3</sub> (10<sup>5</sup> molecules·cm<sup>-3</sup>) (Zhang et al., 2024),  
177 LA (10<sup>12</sup> molecules·cm<sup>-3</sup>) (Li et al., 2017) and H<sub>2</sub>O (10<sup>17</sup> molecules·cm<sup>-3</sup>) (Huang et al., 2015; Tan  
178 et al., 2022), the calculated concentrations of SO<sub>3</sub>···LA, SO<sub>3</sub>···H<sub>2</sub>O and LA···H<sub>2</sub>O complexes at  
179 298 K are 4.18 × 10<sup>-2</sup>, 5.80 × 10<sup>3</sup> and 2.32 × 10<sup>8</sup> molecules·cm<sup>-3</sup>, respectively (see Table S2 in  
180 the Supplement). These results suggest that Channel WM predominantly proceeds via the collision  
181 of LA···H<sub>2</sub>O with SO<sub>3</sub>.

182 The free energy barrier for Channel WM is 7.8 kcal·mol<sup>-1</sup>, which is 14.5 kcal·mol<sup>-1</sup> lower than  
183 the barrier for the uncatalyzed cycloaddition pathway. At the experimental concentration of H<sub>2</sub>O  
184 ([H<sub>2</sub>O] = 10<sup>17</sup> molecules·cm<sup>-3</sup>) (Huang et al., 2015; Tan et al., 2022), the effective rate coefficient  
185 for the H<sub>2</sub>O-catalyzed reaction is 2.00 × 10<sup>-16</sup> cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup>, which is nine orders of magnitude  
186 greater than the rate for the direct cycloaddition pathway (2.22 × 10<sup>-25</sup> cm<sup>3</sup> molecule<sup>-1</sup> s<sup>-1</sup>). These  
187 results clearly demonstrate that the H<sub>2</sub>O-catalyzed LA + SO<sub>3</sub> reaction represents a significantly more  
188 favorable route for LAS formation. Detailed effective rate coefficients for the H<sub>2</sub>O-catalyzed  
189 reaction are provided in Fig. 2(a).

190 SA is another abundant atmospheric species that efficiently donates and accepts hydrogen bond,  
191 facilitating proton transfer (Yao et al., 2018; Tan et al., 2018) and potentially catalyzing the LA +  
192 SO<sub>3</sub> reaction. As shown in Fig. 1, SA is significantly more effective than H<sub>2</sub>O in promoting LAS  
193 formation via cycloaddition. Specifically, SA increases the stabilization energy of the SO<sub>3</sub>···LA  
194 complex by 7.1 kcal·mol<sup>-1</sup>, 5.0 kcal·mol<sup>-1</sup> greater than the stabilization provided by H<sub>2</sub>O and reduces  
195 the distance between the oxygen atom of the -OH group in LA and the sulfur atom in SO<sub>3</sub> by 0.09  
196 Å in the SO<sub>3</sub>···LA···SA complex. Thermodynamically, SA lowers the Gibbs free energy barrier to  
197 3.5 kcal·mol<sup>-1</sup>, 4.3 kcal·mol<sup>-1</sup> lower than the barrier observed for the H<sub>2</sub>O-catalyzed pathway. The  
198 effective rate coefficients for the SA ([SA] = 10<sup>7</sup> molecules·cm<sup>-3</sup>)-catalyzed reaction (*k'*<sub>SA</sub>) is 4-5



orders of magnitude higher than that for the H<sub>2</sub>O-catalyzed pathway ( $k'_{WM}$ ) at 100 % relative humidity, indicating that SA is kinetically more favorable, particularly at altitudes of 5-10 km. Thus, SA predominantly catalyzes the SO<sub>3</sub> + LA reaction, significantly contributing to the gas-phase loss of SO<sub>3</sub> in LA-rich atmospheric regions.

Previous theoretical studies have indicated that atmospheric acids can catalyze the hydrolysis of SO<sub>3</sub> to form SA (Hazra and Sinha, 2011; Cheng et al., 2022; Long et al., 2013; Lv et al., 2019). In this context, the potential catalytic role of LA in SO<sub>3</sub> hydrolysis was also explored. The potential energy surface (PES) for this reaction is presented in Fig. S1, with the effective rate coefficients compared to those for SO<sub>3</sub> hydrolysis catalyzed by SA, HNO<sub>3</sub>, HCOOH, and OA. As shown in Fig. 2(b), LA predominantly catalyzes SO<sub>3</sub> hydrolysis within the temperature range of 280-320 K at a concentration of  $1.0 \times 10^{12}$  molecules·cm<sup>-3</sup>. Besides, given the current lack of atmospheric field data on gas-phase LAS and lactic acid sulfuric anhydride (LASA), thermodynamic equilibrium calculations were used to estimate their concentrations and assess their potential impacts on atmospheric NPF. Modeling results suggest LAS concentrations of 10<sup>3</sup>-10<sup>5</sup> molecules·cm<sup>-3</sup>, which is nine orders of magnitude higher than that of LASA (ranging from 10<sup>-6</sup>-10<sup>-4</sup> molecules·cm<sup>-3</sup>). This suggests that LAS has significantly more atmospheric relevance than LASA, with a correspondingly higher potential to influence NPF. Detailed calculations and further insights are provided in Table S4.

### 3.2 Enhancing effect of LAS on SA-A-driven NPF

The role of LAS in promoting SA-A-driven NPF process was thoroughly examined. Initially, potential interaction sites between LAS and SA-A clusters were identified through molecular analyses. Next, the stable structures and thermodynamic stabilities of various (LAS)<sub>x</sub>(SA)<sub>y</sub>(A)<sub>z</sub> ( $y \leq x + z \leq 3$ ) clusters were characterized, providing insight into their structural integrity. Building on these findings, the nucleation mechanism of the SA-A-LAS system was investigated, with a particular focus on the impact of temperature and precursor concentrations on LAS-mediated NPF processes. Finally, the atmospheric implications of LAS-enhanced SA-A nucleation were evaluated, especially in forested and agricultural-developed regions.

#### 3.2.1 Cluster thermodynamic data

Stable cluster formation is primarily driven by strong interactions between nucleation precursors (Lu and Chen, 2012). To assess the binding potential of LAS with the SA-A cluster, the



electrostatic potential (ESP)-mapped molecular van der Waals surface was calculated to identify key interaction sites. As shown in Fig. 3, the hydrogen atom of the  $-\text{SO}_3\text{H}$  moiety in LAS exhibits a positive ESP of  $+78.73 \text{ kcal}\cdot\text{mol}^{-1}$ , suggesting its role as a hydrogen bond donor that can interact with the double-bonded oxygen atom of SA or the nitrogen atom of A, both of which act as hydrogen bond acceptors. Additionally, the double-bonded oxygen in LAS, with a negative ESP of  $-32.51 \text{ kcal}\cdot\text{mol}^{-1}$ , can act as a hydrogen-bond acceptor, interacting with the hydroxyl hydrogen of SA ( $-\text{OH}$ ) or the hydrogen of A. These intermolecular interactions imply that LAS enhances nucleation efficiency between SA and A during aerosol nucleation, thereby stabilizing the resulting molecular clusters. Based on the ESP analysis, the most stable configurations of  $(\text{LAS})_x(\text{SA})_y(\text{A})_z$  ( $z \leq x + y \leq 3$ ) clusters were identified (Fig. S2), with the observed interaction sites in the ternary clusters corresponding well to the ESP predictions.

To quantitatively evaluate the binding strength of LAS within binary SA-A-based clusters, the Gibbs free energies ( $\Delta G$ ,  $\text{kcal}\cdot\text{mol}^{-1}$ , Table S7) for the  $(\text{LAS})_x(\text{SA})_y(\text{A})_z$  ( $z \leq x + y \leq 3$ ) clusters were calculated at temperatures of 238.15 K, 258.15 K, 278.15 K and 298.15 K. All clusters exhibited negative  $\Delta G$  values, confirming thermodynamic favorability. Importantly, ternary SA-A-LAS clusters consistently demonstrated lower  $\Delta G$  values compared to their binary counterparts, suggesting that the presence of LAS reinforces the stability of SA-A clusters. Further analysis of stability at 278.15 K was carried out by examining total evaporation rates ( $\sum \gamma$ ), derived from cluster  $\Delta G$  values (Table S7) and collision rates ( $\beta$ , Table S8), as summarized in Fig. 4. Previous research indicates that lower  $\sum \gamma$  are indicative of greater cluster stability (Li et al., 2024; Zu et al., 2024a). At 278.15 K, clusters incorporating LAS exhibit a lower  $\sum \gamma$  compared to those composed solely of SA and A molecules. For example, the  $\sum \gamma$  values for the  $(\text{A})_1(\text{LAS})_1$  ( $1.19 \times 10^4 \text{ s}^{-1}$ ) and  $(\text{A})_3(\text{LAS})_3$  ( $8.64 \times 10^{-8} \text{ s}^{-1}$ ) clusters were  $3.1\text{--}10^8$  times lower than those for the  $(\text{SA})_1(\text{A})_1$  ( $3.73 \times 10^4 \text{ s}^{-1}$ ) and  $(\text{SA})_3(\text{A})_3$  ( $3.28 \times 10^1 \text{ s}^{-1}$ ) clusters. Similarly, the  $\sum \gamma$  values of the  $(\text{SA})_1(\text{A})_3(\text{LAS})_2$  ( $1.99 \times 10^0 \text{ s}^{-1}$ ) and  $(\text{SA})_2(\text{A})_3(\text{LAS})_1$  ( $2.29 \times 10^{-4} \text{ s}^{-1}$ ) clusters at 278.15 K were found to be  $10^1\text{--}10^5$  times lower than the most stable binary cluster,  $(\text{SA})_3(\text{A})_3$  ( $3.28 \times 10^1 \text{ s}^{-1}$ ). Moreover, these clusters exhibited  $\beta C/\sum \gamma$  ratios greater than 1 (Table S11), suggesting a favorable balance between cluster growth and evaporation. Similar trends in  $\Delta G$  and  $\sum \gamma$  were observed across the other temperatures studied, including 238.15 K, 258.15 K and 298.15 K. Taken together, the  $\Delta G$  and  $\sum \gamma$  analyses provide strong evidence that LAS incorporation enhances SA-A cluster stability, thereby



259 increasing their likelihood of participating in nucleation events.

### 260 **3.2.2 Cluster formation pathways**

261 To investigate the detailed nucleation pathways of LAS in the formation of SA-A clusters,  
 262 ACDC simulation were conducted at 278.15 K, with the concentrations of [SA] ( $10^6$  molecules·cm<sup>-3</sup>)  
 263 <sup>3</sup>, [A] ( $10^9$  molecules·cm<sup>-3</sup>) and [LAS] ( $10^5$  molecules·cm<sup>-3</sup>). The results are presented in Fig. 5(a),  
 264 illustrating two distinct mechanisms for cluster growth. The first pathway (depicted by black arrows)  
 265 corresponds to pure SA-A clustering, starting from the (SA)<sub>1</sub>·(A)<sub>1</sub> dimer. Subsequent stepwise  
 266 addition of SA or A monomers drives the assembly of progressively larger and more stable clusters  
 267 such as (SA)<sub>3</sub>·(A)<sub>3</sub>, which eventually exits the system. The second pathway (depicted by blue arrows)  
 268 includes clusters containing LAS, in which LAS performs two distinct roles: one as a “catalyst” and  
 269 the other as a “participant”. When LAS acts as a “catalyst”, the (SA)<sub>1</sub>·(A)<sub>2</sub>·(LAS)<sub>1</sub> trimer collides  
 270 with the SA monomer, forming the (SA)<sub>2</sub>·(A)<sub>2</sub>·(LAS)<sub>1</sub> cluster. Subsequently, LAS evaporates from  
 271 the cluster, leaving behind the (SA)<sub>2</sub>·(A)<sub>2</sub> cluster. Meanwhile, when LAS acts as a “participant”,  
 272 collisions between the (SA)<sub>1</sub>·(A)<sub>1</sub> dimer and LAS monomers lead to the assembly of the  
 273 (SA)<sub>1</sub>·(A)<sub>1</sub>·(LAS)<sub>1</sub> cluster. This trimer then undergoes further collisions with either an SA or A  
 274 monomer, producing the (SA)<sub>2</sub>·(A)<sub>3</sub>·(LAS)<sub>1</sub> cluster, which ultimately grows out of the system.  
 275 These dual roles of LAS in SA-A clusters are observed across other temperatures of 298.15 K,  
 276 238.15 K and 258.15 K; however, at lower temperatures, such as 238.15 K, the LAS-involved  
 277 pathway simplifies (as shown in Figs. S8, S9 and S10).

278 As shown in Fig. 5(b), the contributions of LAS to the SA-A nucleation process was examined  
 279 across a range of temperatures, with a focus on the nucleation mechanism that involves LAS  
 280 participation. As temperature increases, the influence of LAS-involved pathways becomes  
 281 progressively more dominant. At lower temperatures (238.15 and 258.15 K), SA-A clustering  
 282 remains the dominant process, accounting for 73% of nucleation events, while LAS-involved  
 283 pathways contribute a modest 21%. However, as the temperature rises to 278.15 K, LAS  
 284 participation increases to 33%, signaling a more prominent role in cluster growth. At 298.15 K, this  
 285 contribution further rises to 49%, nearly double that observed at the lower temperatures. These  
 286 results highlight the crucial role of elevated temperatures in enhancing LAS’s contribution to SA-A  
 287 nucleation, emphasizing the temperature-dependent amplification of LAS-driven cluster formation.

### 288 **3.2.3 Atmospheric implications of LAS**



289 In addition to temperature, the concentrations of precursors play a pivotal role in SA-A aerosol  
290 nucleation. Atmospheric LAS concentrations exhibit considerable variability across different global  
291 environments (Tan et al., 2022; Mochizuki et al., 2017; Ristovski et al., 2010; Hettiyadura et al.,  
292 2017; Kanellopoulos et al., 2022). For example, lower LAS concentrations, ranging from  $1.00 \times 10^4$   
293 to  $8.34 \times 10^5$  molecules·cm<sup>-3</sup>, are found in regions such as eucalypt forest (Ristovski et al., 2010),  
294 Mt. Tai (China) (Mochizuki et al., 2017) and Athens (Kanellopoulos et al., 2022). In contrast, higher  
295 LAS concentrations have been recorded in Centreville, Alabama ( $1.77 \times 10^6$  molecules·cm<sup>-3</sup>)  
296 (Hettiyadura et al., 2017), with peak levels in Patra (Kanellopoulos et al., 2022), reaching up to  
297  $1.70 \times 10^7$  molecules·cm<sup>-3</sup>. Similarly, the concentrations of SA and A vary, with SA ranging from  
298  $10^4$ - $10^7$  molecules·cm<sup>-3</sup> (Zhang et al., 2024; Ding et al., 2019), and A ranging from  $10^7$ - $10^{11}$   
299 molecules·cm<sup>-3</sup> (Wu et al., 2017; Luo et al., 2014). Elevated concentrations of these species are  
300 particularly prominent in regions such as northern China, the Midwestern United States, and  
301 agricultural areas in Europe. Based on field observations of LAS, SA and A concentrations, the  
302 contribution of LAS to SA-A nucleation was systematically assessed. As illustrated in Fig. S11, the  
303 impact of LAS on the SA-A system is primarily governed by the concentrations of LAS and SA,  
304 with minimal dependence on [A]. Consequently, Fig. 6 illustrates how the contribution ratio of LAS  
305 varies with different concentrations of SA and LAS, under the previously identified favorable high-  
306 temperature condition of 278.15 K.

307 The three pie charts in the upper map illustrate the changing contribution of LAS to SA-A  
308 aerosol nucleation as SA concentration increases from  $3.00 \times 10^4$  to  $6.00 \times 10^4$  molecules·cm<sup>-3</sup>, with  
309 a corresponding decrease in LAS contribution as [SA] rises. In regions characterized by low SA  
310 concentrations ( $3.00 \times 10^4$  molecules·cm<sup>-3</sup>), such as Hyytiälä, nucleation is predominantly driven  
311 by the LAS-SA-A pathway, contributing approximately 93%. However, at higher SA concentrations  
312 (up to  $2.00 \times 10^6$  molecules·cm<sup>-3</sup>), such as on the west coast of Ireland (O'dowd et al., 2002), the  
313 LAS contribution drops from 93% to 33%. At even higher SA levels (up to  $1.00 \times 10^7$  molecules·cm<sup>-3</sup>),  
314 LAS-involved pathways account for only 18% of the total nucleation flux, as observed in Beijing,  
315 China (Wang et al., 2011). These findings highlight that lower [SA] levels substantially amplify the  
316 contribution of LAS contribution to SA-A aerosol nucleation.

317 The contribution of LAS to SA-A aerosol nucleation increases with LAS concentration,  
318 ranging from  $1.00 \times 10^4$  to  $1.77 \times 10^6$  molecules·cm<sup>-3</sup>, as shown in the pie chart below the map.



319 This pattern indicates a positive correlation between LAS concentration and its contribution to  
 320 nucleation. In regions with low LAS concentrations ( $1.00 \times 10^4$  molecules·cm<sup>-3</sup>), such as eucalypt  
 321 forests (Ristovski et al., 2010), LAS-mediated pathways account for only 15% of the total nucleation  
 322 flux. In areas with moderate LAS concentrations, such as Athens ( $8.34 \times 10^5$  molecules·cm<sup>-3</sup>)  
 323 (Kanellopoulos et al., 2022) and Mt. Tai ( $1.00 \times 10^5$  molecules·cm<sup>-3</sup>) (Mochizuki et al., 2017), LAS  
 324 contribution increases substantially, rising from 15% to 73%. At high [LAS], as observed in the  
 325 Centreville, Alabama ( $1.77 \times 10^6$  molecules·cm<sup>-3</sup>) (Hettiyadura et al., 2017), LAS-driven nucleation  
 326 becomes dominate, contributing up to 97 % of the total nucleation rate. These findings underscore  
 327 that elevated LAS concentrations significantly enhance SA-A nucleation. Thus, in regions  
 328 characterized by high *T*, low [SA], high [A] and high [LAS], especially in agricultural-developed  
 329 areas and forested areas, the LAS contribution to SA-A aerosol nucleation can be substantial.

### 330 **3.3 The comparison of enhancement effect between LAS and LA**

331 To evaluate the relative enhancing effects of LA versus LAS in the typical SA-A-driven  
 332 nucleation. The  $\Delta G$  (pink histograms) and  $\sum \gamma$  (red points) of the (LAS)<sub>*x*</sub>(SA)<sub>*y*</sub>(A)<sub>3</sub> and  
 333 (LA)<sub>*x*</sub>(SA)<sub>*y*</sub>(A)<sub>3</sub> (*x* = 0 - 3, *x* + *y* = 3) clusters at 278.15 K are presented in Fig. 7(a) as a comparison.  
 334 The (SA)<sub>3</sub>(A)<sub>3</sub> cluster, the thermodynamic minimum of the SA-A system (Chen et al., 2025; Li et  
 335 al., 2020a), was chosen as a reference for comparison. Relative to this baseline, (LA)<sub>1-3</sub>(SA)<sub>0-2</sub>(A)<sub>3</sub>  
 336 clusters consistently exhibited higher  $\Delta G$  values, elevated by roughly 18.36-41.94 kcal·mol<sup>-1</sup>. In  
 337 contrast, (LAS)<sub>1-3</sub>(SA)<sub>0-2</sub>(A)<sub>3</sub> clusters were slightly more stable, differing from the reference by only  
 338 0.09-5.80 kcal·mol<sup>-1</sup>. This suggests that LAS incorporation leads to a slight stabilization of the  
 339 cluster relative to LA.

340 Moreover, the evaporation rate ( $\sum \gamma$ ) of the (Org)<sub>*x*</sub>(SA)<sub>*y*</sub>(A)<sub>3</sub> (Org = LA and LAS; *x* = 1-3, *x* +  
 341 *y* = 3) clusters do not exhibit a simple relationship with the proportion of organic components within  
 342 the clusters. The highest  $\sum \gamma$  was observed for the (Org)<sub>2</sub>·(SA)<sub>1</sub>·(A)<sub>3</sub> (Org = LA and LAS) clusters,  
 343 regardless of whether LA or LAS was used. For the (LAS)<sub>1</sub>·(SA)<sub>2</sub>·(A)<sub>3</sub> and (LAS)<sub>3</sub>·(A)<sub>3</sub> clusters,  
 344 the  $\sum \gamma$  ranged from  $10^{-4}$  to  $10^{-1}$  s<sup>-1</sup>, lower than that of the (SA)<sub>3</sub>·(A)<sub>3</sub> cluster, indicating that replacing  
 345 one or three SA molecules with LAS enhances the thermodynamic stability of the clusters. In  
 346 contrast, the  $\sum \gamma$  of the LA-SA-A clusters were found to be higher than those of the corresponding  
 347 LAS-SA-A and (SA)<sub>3</sub>·(A)<sub>3</sub> clusters, as displayed in Fig. 7(a). The LAS-SA-A clusters exhibit more  
 348 negative negative  $\Delta G$  values and lower  $\sum \gamma$ , suggesting that their formation is thermodynamically



349 more favorable than that of the LA-SA-A system. This enhanced stability can be attributed to  
 350 stronger interactions between LAS and SA-A systems relative to those between LA and SA-A.  
 351 Based on these results, we can conclude that LAS, produced through the LA + SO<sub>3</sub> reaction, more  
 352 effectively stabilizes the SA-A system than LA itself.

353 Fig. 7(b) illustrates the variation in the cluster formation rate ( $J$ ) and enhancement strength ( $R$ )  
 354 as a function of [LAS] and [LA] at 278.15 K, under the condition of [SA] = 10<sup>6</sup> molecules·cm<sup>-3</sup> and  
 355 [A]=10<sup>9</sup> molecules·cm<sup>-3</sup>. In the LAS-SA-A system,  $J$  increases sharply with rising [LAS],  
 356 particularly when [LAS] exceeds 10<sup>5</sup> molecules·cm<sup>-3</sup>. As [LAS] grows from 10<sup>5</sup> to 10<sup>6</sup>  
 357 molecules·cm<sup>-3</sup>,  $J$  for the LAS-SA-A system rises by three orders of magnitude, whereas in the LAS-  
 358 SA-A system,  $J$  exhibits only a modest increase from  $3.36 \times 10^{-9}$  to  $1.12 \times 10^{-8}$  cm<sup>3</sup> s<sup>-1</sup>, consistent  
 359 with the corresponding increase in [LA] (Fig. 7b). Although LAS concentrations are typically three  
 360 orders of magnitude lower than LA (Tan et al., 2022), LAS exerts a substantially stronger  
 361 enhancement effect in SA-A-driven nucleation. These contrasting trends are primarily due to the  
 362 combined influence of cluster thermodynamic properties  $\Delta G$  and  $\Sigma\gamma$ , and the concentrations of  
 363 organic species within the respective systems. The sharp increase in  $J$  for the LAS-SA-A system  
 364 stems from the favorable  $\Delta G$  and low  $\Sigma\gamma$  of the (LAS)<sub>x</sub>(SA)<sub>y</sub>(A)<sub>z</sub> clusters, along with the relatively  
 365 high non-equilibrium concentration of LAS. In contrast, the less favorable  $\Delta G$  and higher  $\Sigma\gamma$  of the  
 366 (LA)<sub>x</sub>(SA)<sub>y</sub>(A)<sub>z</sub> clusters limit the kinetic efficiency of the LA-SA-A system, even at elevated [LA].

367 This study reveals that the reaction of LA and SO<sub>3</sub> generates LAS which acts as an effective  
 368 atmospheric nucleation precursor and significantly accelerates SA-A nucleation. Consequently,  
 369 atmospheric LA can react with part of SO<sub>3</sub>, potentially accounting for the relatively low observed  
 370 low SA concentration, while the generated LAS markedly promotes SA-A-driven NPF under such  
 371 conditions. To date, the effects of hydroxy acids and their derivatives on atmospheric NPF have not  
 372 been comprehensively investigated. The mechanism proposed here offers a general approach to  
 373 evaluate the roles of these acids, like 2-methylglyceric acid, aromatic acids and their derivatives,  
 374 influence atmospheric nucleation processes. Incorporating this novel OSs pathways into  
 375 contemporary atmospheric models will advance the quantitative understanding of OSs'  
 376 contributions to aerosol formation. Furthermore, OSs originating from secondary processes, such as  
 377 gas-phase chemical reactions, deserve further observation and evaluation.



## 4 Conclusions

In this study, quantum chemical calculations, master equation analysis, and the ACDC kinetic model were employed to investigate the cycloaddition reaction between  $\text{SO}_3$  and LA, the role of LAS in SA-A nucleation, and its impact on NPF.

Quantum chemical results in the gas phase indicate that SA and  $\text{H}_2\text{O}$  effectively lower the reaction barriers for LAS formation from the LA- $\text{SO}_3$  reaction, functioning as catalysts and even enabling a barrierless reaction. The effective rate coefficient for the  $\text{SO}_3$ -LA reaction catalyzed by SA ( $10^7 \text{ molecules}\cdot\text{cm}^{-3}$ ) is 4-5 times higher than the pathway catalyzed by  $\text{H}_2\text{O}$  ( $10^{17} \text{ molecules}\cdot\text{cm}^{-3}$ ), making it more effective, particularly at altitudes of 5-10 km. Additionally, the effective rate coefficients for LA ( $10^{12} \text{ molecules}\cdot\text{cm}^{-3}$ ) catalyzing the  $\text{SO}_3 + \text{H}_2\text{O} \rightarrow \text{SA}$  reaction is about  $10^1$ - $10^4$  times larger than the corresponding values for  $\text{SO}_3$  hydrolysis catalyzed by  $\text{H}_2\text{SO}_4$  ( $10^7 \text{ molecules}\cdot\text{cm}^{-3}$ ),  $\text{HNO}_3$  ( $10^9 \text{ molecules}\cdot\text{cm}^{-3}$ ),  $\text{HCOOH}$  ( $10^{11} \text{ molecules}\cdot\text{cm}^{-3}$ ), and OA ( $10^9 \text{ molecules}\cdot\text{cm}^{-3}$ ), indicating that LA primarily catalyzes  $\text{SO}_3$  hydrolysis within the temperature range of 280-320 K.

LAS, functioning as both a hydrogen-bond donor and acceptor, participates in SA-A-driven ternary nucleation, directly interacting with SA and A. Gibbs free energy analysis demonstrates that ternary SA-A-LAS clusters consistently exhibit lower  $\Delta G$  values than their binary counterparts, suggesting that LAS incorporation stabilizes the SA-A clusters. ACDC kinetic simulations further demonstrate that LAS significantly enhances NPF, especially at low temperatures, low SA concentration, and high A and LAS concentrations. In regions with elevated LAS concentrations, such as Centreville, Alabama, particle formation rates can increase by up to  $10^8$ -fold, with SA-A-LAS clusters contributing up to 97% of the overall cluster formation pathways. It is noteworthy that LAS not only acts as a catalyst in enhancing SA-A cluster stability but also directly participates in nucleation. Moreover, LAS exerts a stronger enhancement effect than LA, making it a more effective stabilizing agent for atmospheric NPF. These findings suggest that LAS plays a critical role in enhancing SA-A-driven NPF in forested and agriculturally developed regions, providing insights into previously unaccounted NPF sources and refining nucleation models.

This study deepens the understanding of OSs formation in organic acid-polluted regions and underscores the potential contribution of other OSs to NPF. Neglecting the contribution of OSs in



the SA-A aerosol nucleation, particularly in forested and agricultural regions, may lead to an underestimation of organic aerosol nucleation risks.

## Acknowledgments

This work was supported by the National Natural Science Foundation of China (No: 22203052; 22073059) and the Funds of Graduate Innovation of Shaanxi University of Technology (No: SLGYCX2506).

## Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

## Reference

- Anderson, J. O., Thundiyil, J. G., and Stolbach, A.: Clearing the air: a review of the effects of particulate matter air pollution on human health, *J. Med. Toxicol.*, 8, 166-175, 2012.
- Brean, J., Beddows, D. C. S., Shi, Z., Temime-Roussel, B., Marchand, N., Querol, X., Alastuey, A., Minguillón, M. C., and Harrison, R. M.: Molecular insights into new particle formation in Barcelona, Spain, *Atmos. Chem. Phys.*, 20, 10029-10045, 2020.
- Brüggemann, M., Poulain, L., Held, A., Stelzer, T., Zuth, C., Richters, S., Mutzel, A., van Pinxteren, D., Iinuma, Y., Katkevica, S., Rabe, R., Herrmann, H., and Hoffmann, T.: Real-time detection of highly oxidized organosulfates and BSOA marker compounds during the F-BEACH 2014 field study, *Atmos. Chem. Phys.*, 17, 1453-1469, 2017.
- Cai, R., Yan, C., Yang, D., Yin, R., Lu, Y., Deng, C., Fu, Y., Ruan, J., Li, X., Kontkanen, J., Zhang, Q., Kangasluoma, J., Ma, Y., Hao, J., Worsnop, D. R., Bianchi, F., Paasonen, P., Kerminen, V. M., Liu, Y., Wang, L., Zheng, J., Kulmala, M., and Jiang, J.: Sulfuric acid-amine nucleation in urban Beijing, *Atmos. Chem. Phys.*, 21, 2457-2468, 2021.
- Chen, S. S., Li, R. R., Zhang, C. Y., Wei, S. Q., Wang, R., Chu, B. W., Zhang, X. M., Li, H., and Zhang, T. L.: The enhanced role of formic acid on sulfuric acid-ammonia-driven nucleation in forest regions and polluted city areas, *J. Environ. Sci.*, 621-628, 2025.
- Cheng, Y., Wang, R., Chen, Y., Tian, S., Gao, N., Zhang, Z., and Zhang, T.: Hydrolysis of SO<sub>3</sub> in small clusters of sulfuric acid: mechanistic and kinetic study, *ACS Earth and Space Chemistry*, 6, 3078-3089, 2022.
- Dai, L., Zhao, Y., Zhang, L., Chen, D., and Wu, R.: Particle number size distributions and formation and growth rates of different new particle formation types of a megacity in China, *J. Environ. Sci.*, 131, 11-25, 2023.
- Darar, A. I., Cole-Filipiak, N. C., O'Connor, A. E., and Elrod, M. J.: Formation and stability of atmospherically relevant isoprene-derived organosulfates and organonitrates, *Environ. Sci. Technol.*, 45, 1895-1902, 2011.
- Ding, J., Zhao, P., Su, J., Dong, Q., Du, X., and Zhang, Y.: Aerosol pH and its driving factors in Beijing, *Atmos. Chem. Phys.*, 19, 7939-7954, 2019.
- Faloona, I., Conley, S. A., Blomquist, B., Clarke, A. D., Kapustin, V., Howell, S., Lenschow, D. H.,



- and Bandy, A. R.: Sulfur dioxide in the tropical marine boundary layer: dry deposition and heterogeneous oxidation observed during the pacific atmospheric sulfur experiment, *J. Atmos. Chem.*, 63, 13-32, 2009.
- Froyd, K. D., Murphy, S. M., Murphy, D. M., de Gouw, J. A., Eddingsaas, N. C., and Wennberg, P. O.: Contribution of isoprene-derived organosulfates to free tropospheric aerosol mass, *Proc. Natl. Acad. Sci. U.S.A.*, 107, 21360-21365, 2010a.
- Glasius, M., Hansen, A. M. K., Claeys, M., Henzing, B., Jedyńska, A., Kasper-Giebl, A., Kistler, M., Kristensen, K., Martinsson, J., Maenhaut, W., Nøjgaard, J., Spindler, G., Stenström, K., Swietlicki, E., Szidat, S., Simpson, D., and Yttri, K. E.: Composition and sources of carbonaceous aerosols in Northern Europe during winter, *Atmos. Environ.*, 173, 127-141, 2017.
- Glowacki, D. R., Liang, C. H., Morley, C., Pilling, M. J., and Robertson, S. H.: MESMER: an open-source master equation solver for multi-energy well reactions, *J. Phys. Chem. A*, 116, 9545-9560, 2012.
- Hazra, M. K. and Sinha, A.: Formic acid catalyzed hydrolysis of  $\text{SO}_3$  in the gas phase: a barrierless mechanism for sulfuric acid production of potential atmospheric importance, *J. Am. Chem. Soc.*, 133, 17444-17453, 2011.
- Hettiyadura, A. P. S., Jayarathne, T., Baumann, K., Goldstein, A. H., Gouw, J. A., Koss, A., Keutsch, F. N., Skog, K., and Stone, E. A.: Qualitative and quantitative analysis of atmospheric organosulfates in Centreville, Alabama, *Atmos. Chem. Phys.*, 17, 1343-1359, 2017.
- Hirsikko, A., Nieminen, T., Gagné, S., Lehtipalo, K., Manninen, H. E., Ehn, M., Hörrak, U., Kerminen, V. M., Laakso, L., and McMurry, P.: Atmospheric ions and nucleation: a review of observations, *Atmos. Chem. Phys.*, 11, 767-798, 2011.
- Hodshire, A. L., Campuzano J. P., Kodros, J. K., Croft, B., Nault, B. A., Schroder, J. C., Jimenez, J. L., and Pierce, J. R.: The potential role of methanesulfonic acid (MSA) in aerosol formation and growth and the associated radiative forcings, *Atmos. Chem. Phys.*, 19, 3137-3160, 2019.
- Horváth, G., Horváth, I., Almousa, S. A. D., and Telek, M.: Numerical inverse laplace transformation using concentrated matrix exponential distributions, *Performance Evaluation*, 137, 102067, 2020.
- Hratchian, H. P. and Schlegel, H. B.: Using hessian updating to increase the efficiency of a hessian based predictor-corrector reaction path following method, *J. Chem. Theory Comput.* 1, 61-69, 2005.
- Huang, H. L., Chao, W., and Lin, J. J. M.: Kinetics of a Criegee intermediate that would survive high humidity and may oxidize atmospheric  $\text{SO}_2$ , *Proc. Natl. Acad. Sci. U. S. A.*, 112, 10857-10862, 2015.
- Kanellopoulos, P. G., Kotsaki, S. P., Chrysoschou, E., Koukoulakis, K., Zacharopoulos, N., Philippopoulos, A., and Bakeas, E.: PM<sub>2.5</sub>-bound organosulfates in two Eastern Mediterranean cities: The dominance of isoprene organosulfates, *Chemosphere*, 297, 134103, 2022.
- Kirkby, J., Curtius, J., Almeida, J., Dunne, E., Duplissy, J., Ehrhart, S., Franchin, A., Gagné, S., Ickes, L., Kürten, A., Kupc, A., Metzger, A., Riccobono, F., Rondo, L., Schobesberger, S., Tsagkogeorgas, G., Wimmer, D., Amorim, A., Bianchi, F., Breitenlechner, M., David, A., Dommen, J., Downard, A., Ehn, M., Flagan, R. C., Haider, S., Hansel, A., Hauser, D., Jud, W., Junninen, H., Kreissl, F., Kvashin, A., Laaksonen, A., Lehtipalo, K., Lima, J., Lovejoy, E. R., Makhmutov, V., Mathot, S., Mikkilä, J., Minginette, P., Mogo, S., Nieminen, T., Onnela, A., Pereira, P., Petäjä, T., Schnitzhofer, R., Seinfeld, J. H., Sipilä, M., Stozhkov, Y., Stratmann, F., Tomé, A., Vanhanen, J., Viisanen, Y., Vrtala, A., Wagner, P. E., Walther, H., Weingartner, E., Wex, H., Winkler, P. M., Carslaw,



- 489 K. S., Worsnop, D. R., Baltensperger, U., and Kulmala, M.: Role of sulphuric acid, ammonia and  
490 galactic cosmic rays in atmospheric aerosol nucleation, *Nature*, 476, 429-433, 2011.
- 491 Kirkby, J., Duplissy, J., Sengupta, K., Frege, C., Gordon, H., Williamson, C., Heinritzi, M., Simon,  
492 M., Yan, C., Almeida, J., Tröstl, J., Nieminen, T., Ortega, I. K., Wagner, R., Adamov, A., Amorim,  
493 A., Bernhammer, A.-K., Bianchi, F., Breitenlechner, M., Brilke, S., Chen, X., Craven, J., Dias, A.,  
494 Ehrhart, S., Flagan, R. C., Franchin, A., Fuchs, C., Guida, R., Hakala, J., Hoyle, C. R., Jokinen, T.,  
495 Junninen, H., Kangasluoma, J., Kim, J., Krapf, M., Kürten, A., Laaksonen, A., Lehtipalo, K.,  
496 Makhmutov, V., Mathot, S., Molteni, U., Onnela, A., Peräkylä, O., Piel, F., Petäjä, T., Praplan, A. P.,  
497 Pringle, K., Rap, A., Richards, N. A. D., Riipinen, I., Rissanen, M. P., Rondo, L., Sarnela, N.,  
498 Schobesberger, S., Scott, C. E., Seinfeld, J. H., Sipilä, M., Steiner, G., Stozhkov, Y., Stratmann, F.,  
499 Tomé, A., Virtanen, A., Vogel, A. L., Wagner, A. C., Wagner, P. E., Weingartner, E., Wimmer, D.,  
500 Winkler, P. M., Ye, P., Zhang, X., Hansel, A., Dommen, J., Donahue, N. M., Worsnop, D. R.,  
501 Baltensperger, U., Kulmala, M., Carslaw, K. S., and Curtius, J.: Ion-induced nucleation of pure  
502 biogenic particles, *Nature*, 533, 521-526, 2016.
- 503 Klippenstein, S. J. and Marcus, R. A.: Unimolecular reaction rate theory for highly flexible  
504 transition states. 2. Conventional coordinate formulas for the various possible fragment  
505 combinations: miscellaneous topics, *J. Phys. Chem. C.*, 92, 5412-5417, 1988.
- 506 Kulmala, M., Vehkamäki, H., Petäjä, T., Dal Maso, M., Lauri, A., Kerminen, V. M., Birmili, W., and  
507 McMurry, P. H.: Formation and growth rates of ultrafine atmospheric particles: a review of  
508 observations, *Aerosol Sci.*, 35, 143-176, 2004.
- 509 Kundu, S., Quraishi, T. A., Yu, G., Suarez, C., Keutsch, F. N., and Stone, E. A.: Evidence and  
510 quantitation of aromatic organosulfates in ambient aerosols in Lahore, Pakistan, *Atmos. Chem.*  
511 *Phys.*, 13, 4865-4875, 2013.
- 512 Kurtén, T., Loukonen, V., Vehkamäki, H., and Kulmala, M.: Amines are likely to enhance neutral  
513 and ion-induced sulfuric acid-water nucleation in the atmosphere more effectively than ammonia,  
514 *Atmos. Chem. Phys.*, 8, 4095-4103, 2008.
- 515 Lee, S. H., Gordon, H., Yu, H., Lehtipalo, K., Haley, R., Li, Y., and Zhang, R.: New particle  
516 formation in the atmosphere: from molecular clusters to global climate, *Journal of Geophysical*  
517 *Research: Atmospheres*, 124, 7098-7146, 2019.
- 518 Li, D., Chen, D., Liu, F., and Wang, W.: Role of glycine on sulfuric acid-ammonia clusters formation:  
519 Transporter or participator, *J. Environ. Sci.*, 89, 125-135, 2020a.
- 520 Li, H., Kupiainen-Määttä, O., Zhang, H., Zhang, X., and Ge, M.: A molecular-scale study on the  
521 role of lactic acid in new particle formation: Influence of relative humidity and temperature, *Atmos.*  
522 *Environ.*, 166, 479-487, 2017.
- 523 Li, H., Ning, A., Zhong, J., Zhang, H., Liu, L., Zhang, Y., Zhang, X., Zeng, X. C., and He, H.:  
524 Influence of atmospheric conditions on sulfuric acid-dimethylamine-ammonia-based new particle  
525 formation, *Chemosphere*, 245, 125554-125584, 2020b.
- 526 Li, H., Zhong, J., Vehkamäki, H., Kurtén, T., Wang, W., Ge, M., Zhang, S., Li, Z., Zhang, X.,  
527 Francisco, J. S., and Zeng, X. C.: Self-catalytic reaction of SO<sub>3</sub> and NH<sub>3</sub> to produce sulfamic acid  
528 and its implication to atmospheric particle formation, *J. Am. Chem. Soc.*, 140, 11020-11028, 2018a.
- 529 Li, J., Tsou, N. T., and Du, L.: Effect of a single water molecule on the HO<sub>2</sub> + ClO reaction, *Phys.*  
530 *Chem. Chem. Phys.*, 20, 10650-10659, 2018b.
- 531 Li, J., Ning, A., Liu, L., and Zhang, X.: Atmospheric bases enhanced iodic acid nucleation: altitude-  
532 dependent characteristics and molecular mechanisms, *Environ. Sci. Technol.*, 58, 16962-16973,



- 2024.
- Liu, J. R., Liu, L., Rong, H., and Zhang, X.: The potential mechanism of atmospheric new particle formation involving amino acids with multiple functional groups, *Phys. Chem. Chem. Phys.*, 23, 10184-10195, 2021a.
- Liu, L., Yu, F., Du, L., Yang, Z., Francisco, J. S., and Zhang, X.: Rapid sulfuric acid-dimethylamine nucleation enhanced by nitric acid in polluted regions, *Proc. Natl. Acad. Sci. U.S.A.*, 118, e2108384118, 2021b.
- Liu, L., Zhong, J., Vehkamäki, H., Kurtén, T., Du, L., Zhang, X., Francisco, J. S., and Zeng, X. C.: Unexpected quenching effect on new particle formation from the atmospheric reaction of methanol with SO<sub>3</sub>, *Proc. Natl. Acad. Sci. U.S.A.*, 116, 24966-24971, 2019a.
- Liu, L., Kupiainen M. O., Zhang, H., Li, H., Zhong, J., Kurtén, T., Vehkamäki, H., Zhang, S., Zhang, Y., Ge, M., Zhang, X., and Li, Z.: Clustering mechanism of oxocarboxylic acids involving hydration reaction: Implications for the atmospheric models, *J. Chem. Phys.*, 148, 214303, 2018.
- Long, B., Chang, C. R., Long, Z. W., Wang, Y. B., Tan, X. F., and Zhang, W.J.: Nitric acid catalyzed hydrolysis of SO<sub>3</sub> in the formation of sulfuric acid: a theoretical study, *Chem. Phys. Lett.*, 581, 26-29, 2013.
- Lu, T. and Chen, F.: Multiwfn: a multifunctional wavefunction analyzer, *J. Comput. Chem.*, 33, 580-592, 2012.
- Lu, Y., Liu, L., Ning, A., Gan, Y., Liu, Y., Kurtén, T., Vehkamäki, H., and Wang, L.: Atmospheric sulfuric acid-dimethylamine nucleation enhanced by trifluoroacetic acid, *Geophys. Res. Lett.*, 47, 2020.
- Lund, M. T., Myhre, G., and Samset, B. H.: Anthropogenic aerosol forcing under the shared socioeconomic pathways, *Atmos. Chem. Phys.*, 19, 13827-13839, 2019.
- Luo, X. S., Tang, A. H., Shi, K., Wu, L. H., Li, W. Q., Shi, W. Q., Shi, X. K., Erismann, J. W., Zhang, F. S., and Liu, X. J.: Chinese coastal seas are facing heavy atmospheric nitrogen deposition, *Environ. Res. Lett.*, 9, 2014.
- Lv, G., Sun, X., Zhang, C., and Li, M.: Understanding the catalytic role of oxalic acid in SO<sub>3</sub> hydration to form H<sub>2</sub>SO<sub>4</sub> in the atmosphere, *Atmos. Chem. Phys.*, 19, 2833-2844, 2019.
- Ma, F., Xie, H. B., Elm, J., Shen, J., Chen, J., and Vehkamäki, H.: Piperazine enhancing sulfuric acid-based new particle formation: implications for the atmospheric fate of piperazine, *Environ. Sci. Technol.*, 53, 8785-8795, 2019.
- Mai, T. V., Duong, M. V., Nguyen, H. T., and Huynh, L. K.: Ab initio kinetics of the HOSO<sub>2</sub> + <sup>3</sup>O<sub>2</sub> → SO<sub>3</sub> + HO<sub>2</sub> reaction, *Phys. Chem. Chem. Phys.*, 20, 6677-6687, 2018.
- Mochizuki, T., Kawamura, K., Miyazaki, Y., Kunwar, B., and Boreddy, S. K. R.: Distributions and sources of low-molecular-weight monocarboxylic acids in gas and particles from a deciduous broadleaf forest in northern Japan, *Atmos. Chem. Phys.*, 19, 2421-2432, 2019.
- Mochizuki, T., Kawamura, K., Nakamura, S., Kanaya, Y., and Wang, Z.: Enhanced levels of atmospheric low-molecular weight monocarboxylic acids in gas and particulates over Mt. Tai, North China, during field burning of agricultural wastes, *Atmos. Environ.*, 171, 237-247, 2017.
- Mutzel, A., Poulain, L., Berndt, T., Iinuma, Y., Rodigast, M., Böge, O., Richters, S., Spindler, G., Sipilä, M., Jokinen, T., Kulmala, M., and Herrmann, H.: Highly oxidized multifunctional organic compounds observed in tropospheric particles: a field and laboratory study, *Environ. Sci. Technol.*, 49, 7754-7761, 2015.
- Neese, F.: The ORCA program system, *WIREs computational molecular science*, 2, 73-78, 2012.



- 577 Ning, A. and Zhang, X.: The synergistic effects of methanesulfonic acid (MSA) and methanesulfinic  
578 acid (MSIA) on marine new particle formation, *Atmos. Environ.*, 269, 118826-118834, 2022.
- 579 O'Dowd, C. D., Hämeri, K., Mäkelä, J. M., Pirjola, L., Kulmala, M., Jennings, S. G., Berresheim,  
580 H., Hansson, H. C., de Leeuw, G., Kunz, G. J., Allen, A. G., Hewitt, C. N., Jackson, A., Viisanen,  
581 Y., and Hoffmann, T.: A dedicated study of New Particle Formation and Fate in the Coastal  
582 Environment (PARFORCE): Overview of objectives and achievements, *J. Geophys. Res. Atmos.*,  
583 107, 2002.
- 584 Olson, C. N., Galloway, M. M., Yu, G., Hedman, C. J., Lockett, M. R., Yoon, T., Stone, E. A., Smith,  
585 L. M., and Keutsch, F. N.: Hydroxycarboxylic acid-derived organosulfates: synthesis, stability, and  
586 quantification in ambient aerosol, *Environ. Sci. Technol.*, 45, 6468-6474, 2011.
- 587 Partanen, L., Vehkamäki, H., Hansen, K., Elm, J., Henschel, H., Kurtén, T., Halonen, R., and  
588 Zapadinsky, E.: Effect of conformers on free energies of atmospheric complexes, *J. Phys. Chem. A*,  
589 120, 8613-8624, 2016.
- 590 Ristovski, Z. D., Suni, T., Kulmala, M., Boy, M., Meyer, N. K., Duplissy, J., Turnipseed, A.,  
591 Morawska, L., and Baltensperger, U.: The role of sulphates and organic vapours in growth of newly  
592 formed particles in a eucalypt forest, *Atmos. Chem. Phys.*, 10, 2919-2926, 2010.
- 593 Riva, M., Tomaz, S., Cui, T., Lin, Y.-H., Perraudin, E., Gold, A., Stone, E. A., Villenave, E., and  
594 Surratt, J. D.: Evidence for an unrecognized secondary anthropogenic source of organosulfates and  
595 sulfonates: gas-phase oxidation of polycyclic aromatic hydrocarbons in the presence of sulfate  
596 aerosol, *Environ. Sci. Technol.*, 49, 6654-6664, 2015.
- 597 Rong, H., Liu, L., Liu, J., and Zhang, X.: Glyoxylic sulfuric anhydride from the gas-phase reaction  
598 between Glyoxylic Acid and SO<sub>3</sub>: A potential nucleation precursor, *J. Phys. Chem. A*, 124, 3261-  
599 3268, 2020a.
- 600 Rong, H., Liu, J. R., Zhang, Y. J., Du, L., Zhang, X., and Li, Z.: Nucleation mechanisms of iodic  
601 acid in clean and polluted coastal regions, *Chemosphere*, 253, 126743-126752, 2020b.
- 602 Shampine, L. F. and Reichelt, M. W.: The MATLAB ode suite, *SIAM journal on scientific*  
603 *Computing*, 18, 1-22, 1997.
- 604 Shen, J., Elm, J., Xie, H. B., Chen, J., Niu, J., and Vehkamäki, H.: Structural effects of amines in  
605 enhancing methanesulfonic acid-driven new particle formation, *Environ. Sci. Technol.*, 54, 13498-  
606 13508, 2020.
- 607 Shen, J., Xie, H.-B., Elm, J., Ma, F., Chen, J., and Vehkamäki, H.: Methanesulfonic Acid-driven  
608 new particle formation enhanced by monoethanolamine: A computational study, *Environ. Sci.*  
609 *Technol.*, 53, 14387-14397, 2019.
- 610 Sipilä, M., Sarnela, N., Jokinen, T., Henschel, H., Junninen, H., Kontkanen, J., Richters, S.,  
611 Kangasluoma, J., Franchin, A., Peräkylä, O., Rissanen, M. P., Ehn, M., Vehkamäki, H., Kurten, T.,  
612 Berndt, T., Petäjä, T., Worsnop, D., Ceburnis, D., Kerminen, V.-M., Kulmala, M., and O'Dowd, C.:  
613 Molecular-scale evidence of aerosol particle formation via sequential addition of HIO<sub>3</sub>, *Nature*, 537,  
614 532-534, 2016.
- 615 Smith, C. J., Huff, A. K., Ward, R. M., and Leopold, K. R.: Carboxylic sulfuric anhydrides, *J. Phys.*  
616 *Chem. A*, 124, 601-612, 2020.
- 617 Stewart, J. J.: Optimization of parameters for semiempirical methods V: modification of NDDO  
618 approximations and application to 70 elements, *J. Mol. Model.*, 13, 1173-1213, 2007.
- 619 Sun, H., Liu, Y., Nie, W., Li, Y., Ge, D., Xu, T., Yin, J., Liu, C., Fu, Z., Qi, X., Liu, T., Zha, Q., Yan,  
620 C., Wang, Z., Chi, X., and Ding, A.: Unexpected gas-Phase formation of Glycolic Acid Sulfate in



- 621 the atmosphere, *Environ. Sci. Technol.*, 59, 16556-16566, 2025.
- 622 Tan, S., Zhang, X., Lian, Y., Chen, X., Yin, S., Du, L., and Ge, M.: OH group orientation leads to  
623 organosulfate formation at the liquid aerosol surface, *J. Am. Chem. Soc.*, 144, 16953-16964, 2022.
- 624 Tan, X. F., Long, B., Ren, D. S., Zhang, W. J., Long, Z. W., and Mitchell, E.: Atmospheric chemistry  
625 of CH<sub>3</sub>CHO: the hydrolysis of CH<sub>3</sub>CHO catalyzed by H<sub>2</sub>SO<sub>4</sub>, *Phys. Chem. Chem. Phys.*, 20, 7701-  
626 7709, 2018.
- 627 Tan, X. F., Zhang, L., and Long, B.: New mechanistic pathways for the formation of organosulfates  
628 catalyzed by ammonia and carbinolamine formation catalyzed by sulfuric acid in the atmosphere,  
629 *Phys. Chem. Chem. Phys.*, 22, 8800-8807, 2020.
- 630 Walker, M., Harvey, A. J. A., Sen, A., and Dessent, C. E. H.: Performance of M06, M06-2X, and  
631 M06-HF density functionals for conformationally flexible anionic clusters: M06 functionals  
632 perform better than B3LYP for a model system with dispersion and ionic hydrogen-bonding  
633 interactions, *J. Phys. Chem. A*, 117, 12590-12600, 2013.
- 634 Wang, R., Li, R., Chen, S., Mu, R., Zhang, C., Ma, X., Khan, M., and Zhang, T.: Enhancing SO<sub>3</sub>  
635 hydrolysis and nucleation: the role of formic sulfuric anhydride, *Atmos. Chem. Phys.*, 25, 5695-  
636 5709, 2025.
- 637 Wang, Y., Hu, M., Guo, S., Wang, Y., Zheng, J., Yang, Y., Zhu, W., Tang, R., Li, X., Liu, Y., Le  
638 Breton, M., Du, Z., Shang, D., Wu, Y., Wu, Z., Song, Y., Lou, S., Hallquist, M., and Yu, J.: The  
639 secondary formation of organosulfates under interactions between biogenic emissions and  
640 anthropogenic pollutants in summer in Beijing, *Atmos. Chem. Phys.*, 18, 10693-10713, 2018.
- 641 Wang, Z. B., Hu, M., Yue, D. L., Zheng, J., Zhang, R. Y., Wiedensohler, A., Wu, Z. J., Nieminen, T.,  
642 and Boy, M.: Evaluation on the role of sulfuric acid in the mechanisms of new particle formation  
643 for Beijing case, *Atmos. Chem. Phys.*, 11, 12663-12671, 2011.
- 644 Wardlaw, D. M. and Marcus, R. A.: RRKM reaction rate theory for transition states of any looseness,  
645 *Chem. Phys. Lett.*, 110, 230-234, 1984.
- 646 Wu, S., Dai, L. H., Wei, Y., Zhu, H., Zhang, Y. J., Schwab, J., and Yuan, C. S.: Atmospheric ammonia  
647 measurements along the coastal lines of Southeastern China: Implications for inorganic nitrogen  
648 deposition to coastal waters, *Atmos. Environ.*, 177, 1-11, 2017.
- 649 Xing, Y. F., Xu, Y. H., Shi, M. H., and Lian, Y. X.: The impact of PM<sub>2.5</sub> on the human respiratory  
650 system, *J. Thorac. Dis.*, 8, 69-74, 2016.
- 651 Yang, Y., Liu, L., Wang, H., and Zhang, X.: Molecular-scale mechanism of sequential reaction of  
652 oxalic acid with SO<sub>3</sub>: potential participator in atmospheric aerosol nucleation, *J. Phys. Chem. A*,  
653 125, 4200-4208, 2021.
- 654 Yao, L., Garmash, O., Bianchi, F., Zheng, J., Yan, C., Kontkanen, J., Junninen, H., Mazon, S. B.,  
655 Ehn, M., Paasonen, P., Sipilä, M., Wang, M., Wang, X., Xiao, S., Chen, H., Lu, Y., Zhang, B., Wang,  
656 D., Fu, Q., Geng, F., Li, L., Wang, H., Qiao, L., Yang, X., Chen, J., Kerminen, V. M., Petäjä, T.,  
657 Worsnop, D. R., Kulmala, M., and Wang, L.: Atmospheric new particle formation from sulfuric acid  
658 and amines in a Chinese megacity, *Science*, 361, 278-281, 2018.
- 659 Yao, L., Fan, X., Yan, C., Kurtén, T., Daellenbach, K. R., Li, C., Wang, Y., Guo, Y., Dada, L.,  
660 Rissanen, M. P., Cai, J., Tham, Y. J., Zha, Q., Zhang, S., Du, W., Yu, M., Zheng, F., Zhou, Y.,  
661 Kontkanen, J., Chan, T., Shen, J., Kujansuu, J. T., Kangasluoma, J., Jiang, J., Wang, L., Worsnop,  
662 D. R., Petäjä, T., Kerminen, V. M., Liu, Y., Chu, B., He, H., Kulmala, M., and Bianchi, F.:  
663 Unprecedented ambient sulfur trioxide (SO<sub>3</sub>) detection: possible formation mechanism and  
664 atmospheric implications, *Environ. Sci. Technol.*, 7, 809-818, 2020.

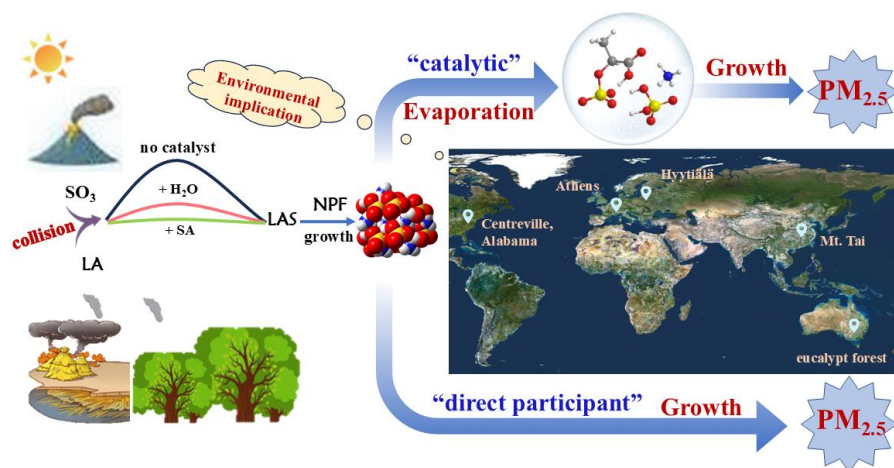


- 665 Yin, R., Yan, C., Cai, R., Li, X., Shen, J., Lu, Y., Schobesberger, S., Fu, Y., Deng, C., Wang, L., Liu,  
666 Y., Zheng, J., Xie, H., Bianchi, F., Worsnop, D. R., Kulmala, M., and Jiang, J.: Acid-base clusters  
667 during atmospheric new particle formation in urban Beijing, *Environ. Sci. Technol.*, 55, 10994-  
668 11005, 2021a.
- 669 Yin, R. J., Yan, C., Cai, R. L., Li, X., Shen, J., Lu, Y., Schobesberger, S., Fu, Y., Deng, C., Wang, L.,  
670 Liu, Y., Zheng, J., Xie, H., Bianchi, F., Worsnop, D. R., Kulmala, M., and Jiang, J.: Acid-base  
671 clusters during atmospheric new particle formation in urban Beijing, *Environ. Sci. Technol.*, 55,  
672 10994-11005, 2021b.
- 673 Zhang, H., Gao, R., Li, H., Li, Y., Xu, Y., and Chai, F.: Formation mechanism of typical aromatic  
674 sulfuric anhydrides and their potential role in atmospheric nucleation process, *J. Environ. Sci.*, 123,  
675 54-64, 2023a.
- 676 Zhang, H., Li, H., Liu, L., Zhang, Y., Zhang, X., and Li, Z.: The potential role of malonic acid in the  
677 atmospheric sulfuric acid-ammonia clusters formation, *Chemosphere*, 203, 26-33, 2018.
- 678 Zhang, H., Wang, W., Fan, L., Li, J., Ren, Y., Li, H., Gao, R., and Xu, Y.: The role of sulfur cycle in  
679 new particle formation: Cycloaddition reaction of SO<sub>3</sub> to H<sub>2</sub>S, *J. Environ. Sci.*, 148, 489-501, 2025.
- 680 Zhang, H., Worton, D. R., Lewandowski, M., Ortega, J., Rubitschun, C. L., Park, J.-H., Kristensen,  
681 K., Campuzano-Jost, P., Day, D. A., Jimenez, J. L., Jaoui, M., Offenberg, J. H., Kleindienst, T. E.,  
682 Gilman, J., Kuster, W. C., de Gouw, J., Park, C., Schade, G. W., Frossard, A. A., Russell, L., Kaser,  
683 L., Jud, W., Hansel, A., Cappellin, L., Karl, T., Glasius, M., Guenther, A., Goldstein, A. H., Seinfeld,  
684 J. H., Gold, A., Kamens, R. M., and Surratt, J. D.: Organosulfates as tracers for secondary organic  
685 aerosol (SOA) formation from 2-Methyl-3-Buten-2-ol (MBO) in the atmosphere, *Environ. Sci.*  
686 *Technol.*, 46, 9437-9446, 2012a.
- 687 Zhang, J. and Dolg, M.: Global optimization of clusters of rigid molecules using the artificial bee  
688 colony algorithm, *Phys. Chem. Chem. Phys.*, 18, 3003-3010, 2016.
- 689 Zhang, Q., Wang, Y., Liu, M., Zheng, M., Yuan, L., Liu, J., Tao, S., and Wang, X.: Wintertime  
690 formation of large sulfate particles in China and implications for human health, *Environ. Sci.*  
691 *Technol.*, 57, 20010-20023, 2023b.
- 692 Zhang, R., Khalizov, A., Wang, L., Hu, M., and Xu, W.: Nucleation and growth of nanoparticles in  
693 the atmosphere, *Chem. Rev.*, 112, 1957-2011, 2012b.
- 694 Zhang, R., Wang, G., Guo, S., Zamora, M. L., Ying, Q., Lin, Y., Wang, W., Hu, M., and Wang, Y.:  
695 Formation of urban fine particulate matter, *Chem. Rev.*, 115, 3803-3855, 2015.
- 696 Zhang, X. M., Lian, Y. J., Tan, S. D., and Yin, S.: Organosulfate produced from consumption of SO<sub>3</sub>  
697 speeds up sulfuric acid-dimethylamine atmospheric nucleation, *Atmos. Chem. Phys.*, 24, 3593-3612,  
698 2024.
- 699 Zhao, B., Donahue, N. M., Zhang, K., Mao, L., Shrivastava, M., Ma, P.-L., Shen, J., Wang, S., Sun,  
700 J., Gordon, H., Tang, S., Fast, J., Wang, M., Gao, Y., Yan, C., Singh, B., Li, Z., Huang, L., Lou, S.,  
701 Lin, G., Wang, H., Jiang, J., Ding, A., Nie, W., Qi, X., Chi, X., and Wang, L.: Global variability in  
702 atmospheric new particle formation mechanisms, *Nature*, 631, 98-105, 2024.
- 703 Zheng, B., Tong, D., Li, M., Liu, F., Hong, C., Geng, G., Li, H., Li, X., Peng, L., Qi, J., Yan, L.,  
704 Zhang, Y., Zhao, H., Zheng, Y., He, K., and Zhang, Q.: Trends in China's anthropogenic emissions  
705 since 2010 as the consequence of clean air actions, *Atmos. Chem. Phys.*, 18, 14095-14111, 2018.
- 706 Zu, H., Zhang, S., and Liu, L.: The vital role of sulfuric acid in iodine oxoacids nucleation: impacts  
707 of urban pollutants on marine atmosphere, *Environ. Res. Lett.*, 19, 2024a.
- 708 Zu, H., Zhang, S., Li, S., Liu, L., and Zhang, X.: The synergistic nucleation of iodous acid and



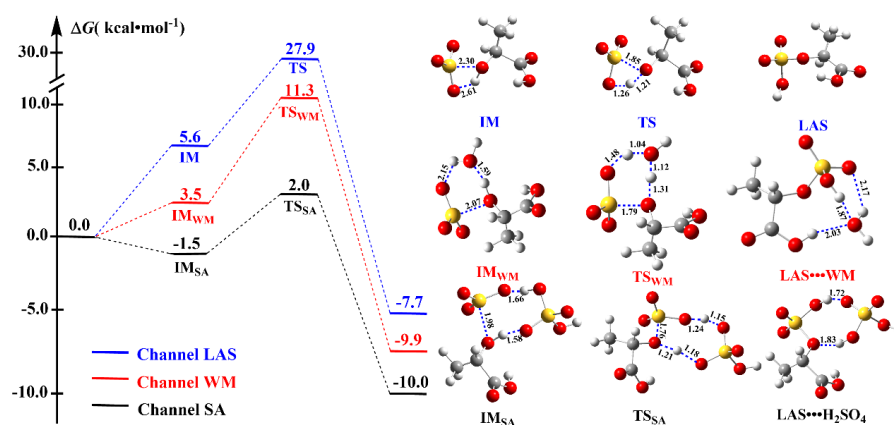
709 sulfuric acid: A vital mechanism in polluted marine regions, Atmos. Environ., 318, 120266, 2024b.

710

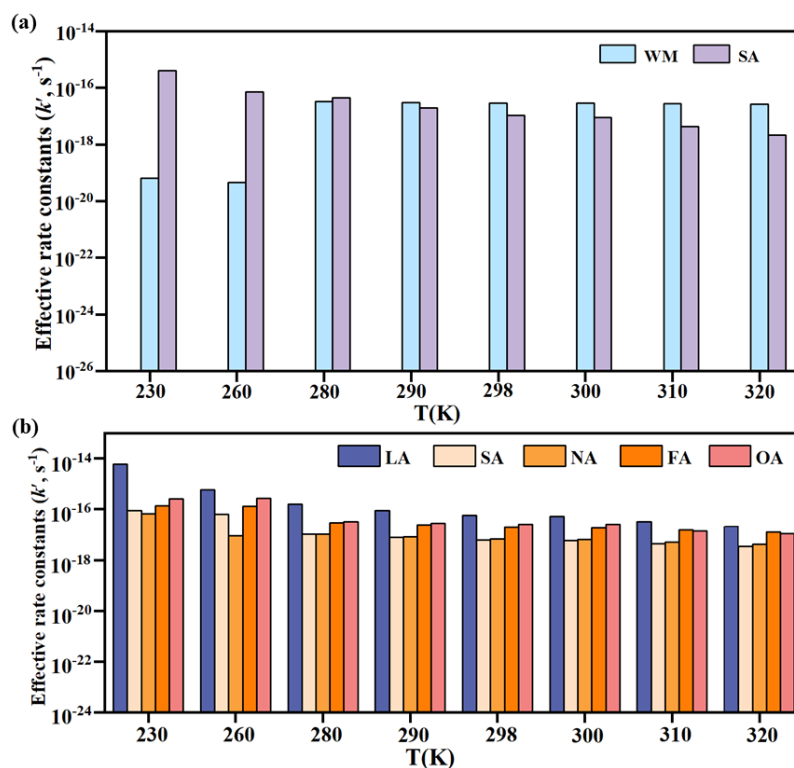


Map source: ©Google Maps (<https://www.google.com/maps>)

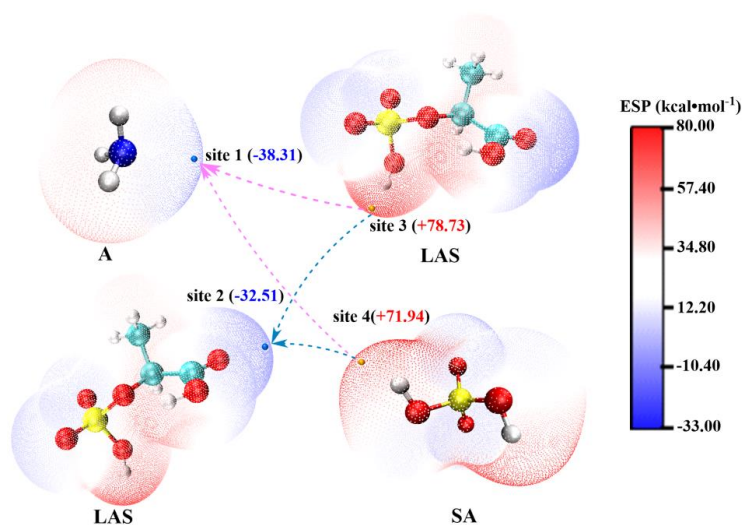
**Graphic abstract**



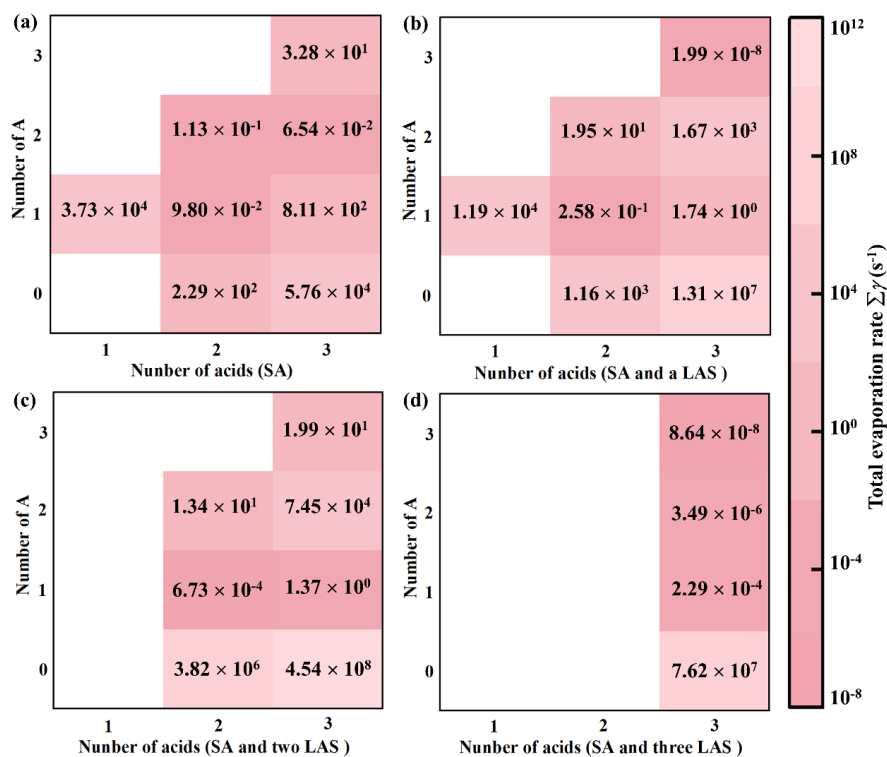
**Fig. 1** Potential energy profiles and corresponding molecular structures for the  $\text{LA} + \text{SO}_3 \rightarrow \text{LAS}$  reaction in the absence and presence of  $\text{H}_2\text{O}$  and  $\text{H}_2\text{SO}_4$  investigated at the CCSD(T)-F12/cc-pVDZ-F12/M06-2X/6-311++G(2df, 2pd) level



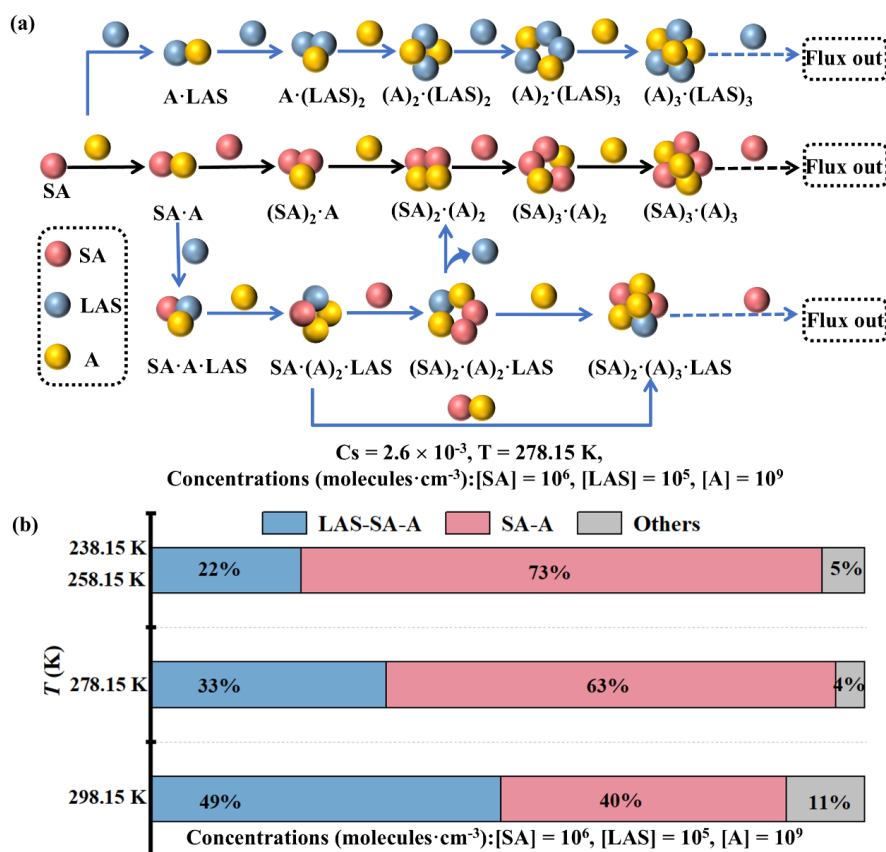
**Fig. 2**(a) Effective rate constants for the  $LA + SO_3 \rightarrow LAS$  reaction in the presence of  $H_2O$  ( $k'_{WM}$ ,  $cm^3 \cdot molecule^{-1} \cdot s^{-1}$ ) and  $H_2SO_4$  ( $k'_{SA}$ ,  $cm^3 \cdot molecule^{-1} \cdot s^{-1}$ ) calculated using the master equation over the temperature range of 230-320 K; (b) Effective rate constants ( $k'$ ,  $s^{-1}$ ) for the hydrolysis of  $SO_3$  with various species  $X$  ( $X = LA, SA, NA, FA$  and  $OA$ ) within the temperature range of 230-320 K, where  $SA, NA, FA$  and  $OA$  are denoted as  $H_2SO_4, HNO_3, HCOOH$  and  $H_2C_2O_4$ , respectively.



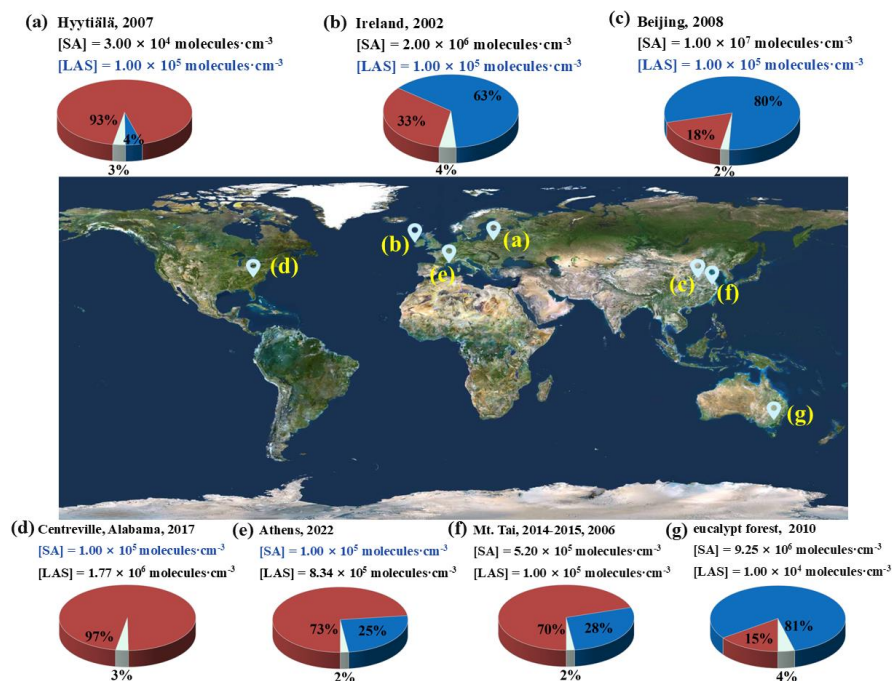
725  
 726 **Fig. 3** Electrostatic potential (ESP)-mapped van der Waals surfaces of A, LAS and SA molecules.  
 727 ESP minima and maxima for different functional groups are shown as blue and yellow spheres,  
 728 respectively, with their corresponding values ( $\text{kcal}\cdot\text{mol}^{-1}$ ) indicated in parentheses. Red arrows  
 729 denote preferred directions for hydrogen bond formation, while blue arrows illustrate likely  
 730 pathways for proton transfer.



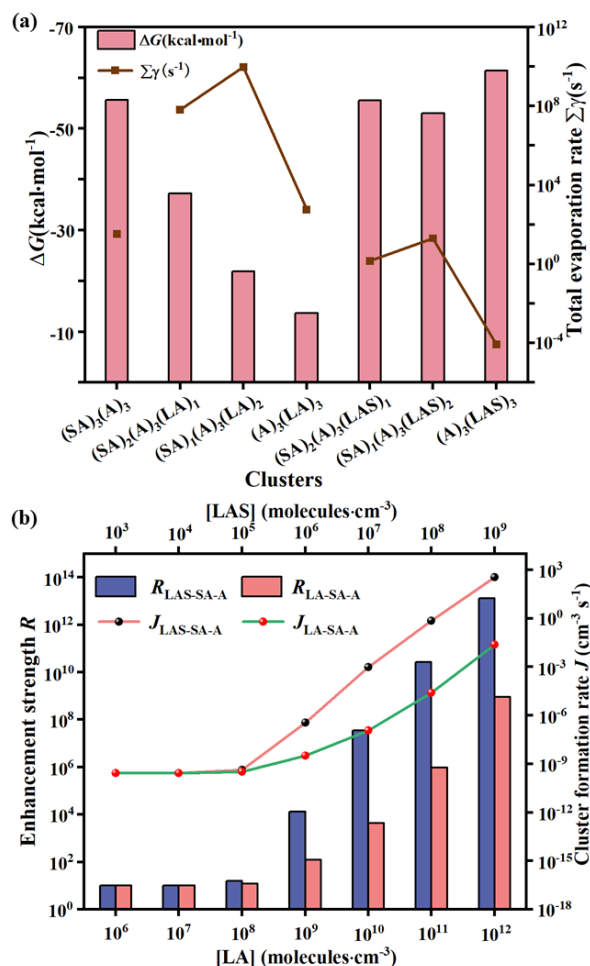
**Fig. 4** The total evaporation rates ( $\Sigma\gamma$ ) ( $\text{s}^{-1}$ ) of  $(\text{SA})_x(\text{A})_y(\text{LAS})_z$  ( $y \leq x + z \leq 3$ ) clusters at 278.15 K and 1 atm calculated at the M06-2X/6-311++G(2df, 2pd) level of theory. (a) without LAS monomer, (b) containing 1 LAS monomer, (c) containing 2 LAS monomers, and (d) containing 3 LAS monomers



**Fig. 5** Nucleation mechanism of the LAS-SA-A system. (a) Cluster formation pathway at 278.15 K, with concentrations of [SA] = 10<sup>6</sup>, [A] = 10<sup>9</sup> and [LAS] = 10<sup>5</sup> molecules·cm<sup>-3</sup>; (b) the branch ratio of outward flux at different temperatures. Only net fluxes contributing more than 5% to cluster growth are depicted.



**Fig. 6** Branching ratios of SA-A-LAS (red) and SA-A (blue) cluster growth pathways in regions with varying  $[LAS]$  concentrations. Black data points indicate field observations, while blue points represent the median values used in this study. Ammonia concentration is fixed at  $10^9 \text{ molecules} \cdot \text{cm}^{-3}$ . Map source: ©Google Maps (<https://www.google.com/maps>)



**Fig. 7** (a) Gibbs free energies  $\Delta G$  ( $\text{kcal}\cdot\text{mol}^{-1}$ ) and total evaporation rates  $\Sigma\gamma$  ( $\text{s}^{-1}$ ) for  $(LA)_x(SA)_y(A)_3$  and  $(LAS)_x(SA)_y(A)_3$  ( $x = 0-3$ ,  $x + y = 3$ ) clusters calculated at the M06-2X/6-311++G(2df, 2pd) level of theory and 278.15 K; (b) Cluster formation rate ( $J$ ) and enhancement strength ( $R$ ) for LAS as a function of monomer concentrations ( $[LA]$  and  $[LAS]$ ) at 278.15 K, with  $[SA]$  fixed at  $10^5 \text{ molecules}\cdot\text{cm}^{-3}$  and  $[A]$  at  $10^9 \text{ molecules}\cdot\text{cm}^{-3}$ .