



1 **Geochemical characteristics of suspended particulate matter around Piscadera Bay and its**
2 **influence on near shore ecosystems, Curaçao (Caribbean Sea)**

3
4 Virginia Sánchez Barranco^{1*}; Nienke C.J. van de Loosdrecht²; Furu Mienis¹; Jasper M. de Goeij²;
5 Rick Hennekam¹; Gert-Jan Reichart^{1,3}; Jan-Berend Stuut^{1,4}; Lennart J. de Nooijer¹

6 ¹ Royal Netherlands Institute for Sea Research (NIOZ), department of Ocean Systems, Den
7 Burg, The Netherlands.

8 ² Institute for Biodiversity and Ecosystem Dynamics (IBED), University van Amsterdam,
9 Amsterdam, The Netherlands.

10 ³ Department of Earth Sciences, Faculty of Geosciences, Utrecht University, Budapestlaan 4,
11 3584 CD Utrecht, The Netherlands.

12 ⁴ Vrije Universiteit Amsterdam (VU Amsterdam), Amsterdam, The Netherlands.

13 *Corresponding author

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16 **Abstract.** Caribbean coral reefs face rising pressure from coastal development, yet the
17 pathways by which urban pollution reaches these endangered ecosystems remain poorly
18 understood. Bays act as dynamic channels, trapping, transforming, and releasing materials
19 that can impact adjacent reef systems. We investigated the seasonal and spatial variability of
20 suspended particulate matter (SPM)—a key vector for pollutants and nutrients—coming from
21 an urbanized bay in Curaçao and determined its effect on surrounding coral reefs. Using
22 sediment traps deployed across spatial gradients (bay mouth to nearby reefs in the East and
23 West) during the dry (April-May) and wet (October-November) seasons, we measured mass,
24 carbon, and nitrogen fluxes and associated grain-size and geochemical particle composition.
25 Results were compared to environmental conditions (e.g. rain fall, current speed) and
26 revealed a clear spatial gradient of bay influence: the bay mouth showed the strongest
27 terrestrial signal, followed by the eastern reef (sheltered from currents) with elevated SPM
28 fluxes of fine particles enriched in terrigenous elements (Si, Fe, Al, and Mn), while the western
29 reef (exposed to open-ocean flow) exhibited lower fluxes of coarser particles with elevated
30 Ca/Fe, Pb, Cu and Ni. This indicates diminished bay effect and stronger marine influence mixed
31 with localized pollution. During the dry season, differences in SPM fluxes and composition
32 between reef sites were minimal, but wet season conditions amplified spatial patterns, with



33 rainfall-driven runoff locally increasing dissolved and particular matter delivery. This implies
34 that reef vulnerability to bay-derived pollution locally depends on both proximity to source
35 waters and seasonal hydrodynamic variability, with sheltered reefs experiencing the greatest
36 impacts during periods of enhanced terrestrial runoff.

37 **Key words:** SPM, Coral reefs, Terrestrial runoff, Spatial gradient, Seasonal variability, Carbon
38 flux

39 1. Introduction

40 In recent decades, land-use changes and coastal development around watersheds have
41 increased runoff of sediments, nutrients and chemical pollutants into tropical coastal waters,
42 posing significant threats to coral reefs worldwide, including those in the Caribbean (Burke
43 and Maidens, 2004; Burke et al., 2011). The accumulation of terrestrial discharge on nearshore
44 reefs cause local changes to reef communities, for example the increased cover of
45 (macro)algae (Fabricius, 2005) or suffocation of corals (Weber et al., 2006). Local stressors
46 make them also more vulnerable to global ones, such as thermal bleaching and ocean
47 acidification due to compounded energy deficits and impaired recovery pathways (Storlazzi
48 et al., 2015; Dutra et al., 2018; Anthony et al., 2008). Inland bays on Caribbean islands could
49 play an important role in the supply of sediments, nutrients and pollutants. They show strong
50 environmental variability (diel and seasonal) in physicochemical parameters and higher
51 nutrient and pollutant concentrations compared to reef waters (de Jong et al. 2025), and serve
52 as locations where land-derived substances are retained, transformed, or exported to coastal
53 waters (Heiskanen and Tallberg, 1999). In tropical island environments, discharge from bays is
54 typically episodic, occurring mainly during the rainy season and storm events leading to pulses
55 of material and runoff being released into the ocean (Ringuet and Mackenzie, 2005; De Carlo
56 et al., 2007; Sánchez Barranco et al., 2025; Larson et al., 2015).

57 Suspended particulate matter (SPM) plays a key role in the transport and cycling of minerals,
58 organic matter, nutrients, and pollutants, all of which can influence coral reef health (Jouon et
59 al., 2008). For example, sediment plumes, often dominated by fine-grained material, can
60 reduce light availability and physically stress coral reefs, especially near river mouths and
61 coastal bays (Fabricius, 2005; Ramos-Scharrón and Macdonald, 2005). Sedimentation from
62 SPM can also physically smother coral polyps, impair recruitment, and increase mortality



63 (Bhuyan et al., 2025; Erftemeijer et al., 2012). The energetic cost of shedding sediments—via
64 mucus production and ciliary action—can further compromise coral metabolism and
65 resilience (Tuttle and Donahue, 2022). Some studies also show that certain reef organisms can
66 benefit from SPM. For instance, soft corals may utilize detritus and other small SPM ($<10\ \mu\text{m}$
67 particle size) as an important food source, particularly in high-turbidity environments
68 (Anthony, 1999; Fabricius and Dommissie, 2000). De Jong et al. (2025) have shown that the
69 environmental conditions in semi-enclosed bays, where coral communities often persist, can
70 be highly variable compared to open reefs. Reduced light penetration due to suspended
71 sediments limits the photosynthetic energy available for coral growth and reproduction, with
72 finer-grained and darker-coloured sediments causing more pronounced effects than coarser,
73 lighter ones (Anthony and Fabricius, 2000; Weber et al., 2006; Storlazzi et al., 2015). In
74 addition, fine-grained SPM remains longer in suspension, leading to prolonged elevated levels
75 of turbidity. Beyond its impact on light availability, SPM also plays a crucial role in pollutant
76 dynamics. Heavy metals and organic contaminants readily bind to SPM, particularly clays and
77 silts, making them a critical vector for pollutant transport in near-shore marine environments
78 (Salomons and Förstner, 2012; Zhang et al., 2018). This is of particular concern in the
79 Caribbean, where metal pollution poses ecological risks due to its persistence in reef
80 ecosystems (Berry et al., 2013; Guzmán and Jiménez, 1992).

81 The south coast of Curaçao is characterized by several inland bays that may contribute to the
82 transport of particulate substances to nearby coastal ecosystems, including coral reefs,
83 mangroves, and seagrass beds (Debrot et al., 1998; Wagenaar Hummelinck, 1977). Despite a
84 50% loss in coral cover over the past 50 years (Sandin et al., 2022) the reefs along the south
85 coast of Curaçao (southeastern Caribbean) are still considered among the healthiest in the
86 Caribbean Sea (Sandin et al., 2008a; Rijsberman and Westmacott, 2000; Sandin et al., 2022;
87 Jackson et al., 2014). Seasonal bay-ocean exchange dynamics are influencing the water-
88 column characteristics of nearby coastal ecosystems (Sánchez Barranco et al., 2025). In
89 Curaçao, bay-ocean exchange is limited during the dry season, leading to the build-up of
90 nutrient concentrations within the bays, whereas during the wet season, increased mixing
91 between bay and offshore waters is enhanced (Sánchez Barranco et al., 2025). These dynamic
92 exchange processes potentially create a complex mechanism for sediment and nutrient
93 transfer between coastal environments. However, while it is known that bays release SPM into



94 surrounding waters (Restrepo et al., 2016; D'sa and Ko, 2008), the extent of its distribution,
 95 deposition, and seasonal variability, and effect on changes in benthic composition, remain
 96 poorly quantified.

97 Here, we aim to quantify fluxes of SPM and assess their geochemical characteristics in
 98 proximity and further away from a bay located on the Caribbean Island Curaçao. The coastal
 99 zone near Piscadera Bay, located on the south coast of Curaçao, has been studied extensively
 100 and has one of the longest benthic cover records in the world (De Bakker et al., 2017) providing
 101 a unique site to study the potential influence of bay discharge on nearby reefs. We present a
 102 study in which sediment traps were deployed for 30 days at various locations around Piscadera
 103 Bay during April/May (dry season) and October/November 2023 (wet season). Seasonal
 104 variability in SPM fluxes, sampled in the dry and wet seasons, was compared to environmental
 105 conditions (e.g. rainfall, wind, tides) to determine if they explain the observed spatio-temporal
 106 variability in SPM fluxes and composition. Collected material was analyzed for grain size,
 107 elemental, and organic matter composition.

108 **2. Methodology**

109 **2.1 Study site**

110 The island of Curaçao lies near the Venezuelan coast with a surface area of approximately 444
 111 km² and around 150.000 inhabitants and around 1.2 million (cruise) tourists visit the island
 112 annually. The south coast is bordered by fringing reefs that start at a water depth of
 113 approximately 7 m followed by a steep slope down to 30–50 m (Bak, 1977; De Bakker et al.,
 114 2016). The island is characterized by a semi-arid climate, including a wet season with highest
 115 precipitation rates in the months of October-December (ranging between 80-115 mm per
 116 month) and a dry season with lowest monthly rainfall in April-June (ranging between 8-15 mm
 117 per month) (Wit et al., 2025)

118 This study focuses on Piscadera Bay (Fig. 1), an area that serves as an important catchment
 119 area, draining 16–17 km² of the mid-west part of the island of Curaçao (Sánchez Baranco et
 120 al., 2025). The bay's wide mouth (ranging from 76 m at its narrowest point to 105 m at its
 121 widest) has a shallow entrance of less than 6 m deep. Originally, Piscadera Bay opened to the
 122 sea through a narrow, shallow channel, mostly blocked by a coral and debris bar, with a cross-



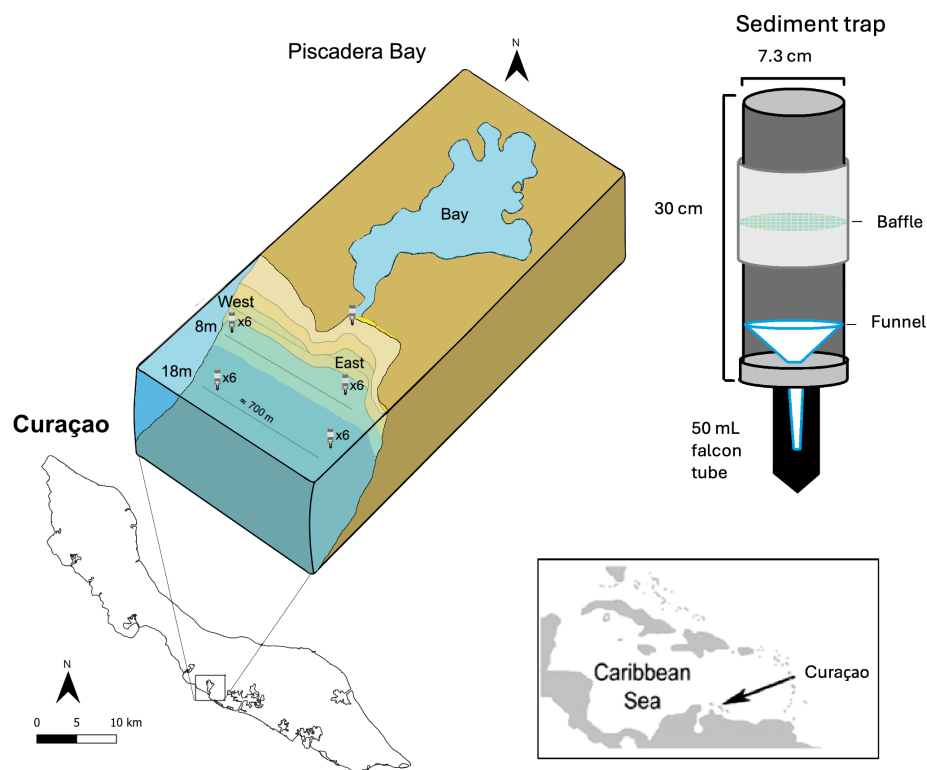
123 sectional area of only 7 m². In 1962, the entrance was widened to 45 m², and in the early
124 seventies it was further enlarged and deepened (Hofker, 1971).

125 The reef sites surrounding the bay reflect significant ecological changes over time (Sandin et
126 al., 2022; De Bakker et al., 2017). The reef west of the bay's mouth, which has experienced
127 substantial coral degradation, showed 8.2% hard coral cover in 2015, compared to an
128 estimated 40% in 1982, with 29.7% turf cover and 2.7% cyanobacterial mats in 2015 (Sandin
129 et al., 2022; Van Duyl, 1985). The reef directly east of Piscadera's mouth also experienced an
130 ecological transformation, with hard coral cover declining from an estimated 5% in 1982 to
131 merely 0.9% in 2015, accompanied by 51.5% turf cover and 11.5% cyanobacterial mats (De
132 Bakker et al., 2016; Van Duyl, 1985; Sandin et al., 2022)

133 Fieldwork was conducted in April–May (dry season) and October–November (wet season)
134 2023. Sediment traps were deployed for 30 days at the mouth of the bay (3–5 m water depth),
135 at the east reef side at 8 and 18 m water depth and at the west reef side of the bay at 8 and
136 18 m water depth (Fig. 1). At each site, 6 traps were deployed at least 1 m apart from each
137 other following Storlazzi et al. (2011).



138



139

140 **Figure 1.** Map of Curaçao showing the studied location of Piscadera Bay on the west side of
141 Willemstad. The inset provides a detailed sketch of the study area around Piscadera Bay, with
142 sediment traps deployed at the bay mouth and at eastern and western reef at depths of 8 m
143 and 18 m. The sediment traps used, follow the design recommendations of Storlazzi et al.
144 (2011) and is shown on the right side for reference.

145 **2.2 Current sensors and environmental data**

146 In October–November 2023, two Acoustic Doppler Current Profilers (ADCP) were deployed at
147 a depth of 10 m at the eastern and western reef of Piscadera Bay for a 30-d period. In April–
148 May 2023, the ADCPs were also deployed for shorter periods of time at the mouth (7 d),
149 eastern (4 d), and western reef (20 d). The ADCPs were used to measure current speed and
150 direction, water temperature, and pressure, which was used to determine tidal cycles. Water
151 speed and direction data used in this study were taken from ADCP bin 3 located 2 m above
152 the seafloor, which corresponded to 8 m water depth at the reef sites (with a total water depth



153 of 10 m) and 3 m water depth near the mouth of the bay (with a total water depth of 5 m).
 154 Precipitation and wind data during the sampling periods were retrieved from the
 155 Meteorological Department of Curaçao (Meteorological Department Curaçao, n.d.).

156 **2.3 Sediment traps and mass fluxes**

157 Sediment traps were built following Storlazzi et al. (2011), with a height of 30 cm, and a surface
 158 area of 41.85 cm². Traps ($n = 6$ per site and depth) were secured with a steel rebar into the
 159 sediments with their mouths at 0.7–0.8 m above the seafloor. Inside the trap a baffle was
 160 placed to avoid organisms or large objects entering the funnel. At the bottom of the trap, a
 161 funnel connected to a 50 mL dark falcon tube was used to collect suspended particulate
 162 matter. In April–May 2023, traps at the mouth of the bay were sampled every 5 d and the ones
 163 in the eastern and western reef every 10 d. In October–November 2023, Falcon tubes were
 164 replaced every 5 d at each location. After collection, samples were centrifuged and excess
 165 water was removed, frozen in a -20 freezer and freeze-dried to be transported to the
 166 Netherlands for further analysis. To remove marine salts, the freeze-dried samples were
 167 resuspended in Milli-Q water, centrifuged again, and the supernatant was carefully pipetted
 168 out, after which samples were freeze-dried again. Freeze-dried samples were weighed with a
 169 high-precision microbalance (± 0.1 mg resolution), and an uncertainty of ± 0.0001 g was applied
 170 to all mass measurements. Carbon and nitrogen percentages were measured with a precision
 171 of $\pm 0.02\%$. These uncertainties were propagated through all derived fluxes, resulting in mass
 172 flux values reported to two decimal places, and C_{org} and total N flux values reported to two
 173 and three decimal places, respectively. Grain-size measurements and silt/clay percentages
 174 were obtained using laser diffraction, with instrument precision better than 1 μ m and 0.1%,
 175 respectively. The final calculated grain-size measurements and silt/clay percentages are
 176 reported to two decimal places.

177 **2.4 Organic carbon and nitrogen content**

178 Organic carbon and total nitrogen contents were measured using an Elementar Vario El Cube
 179 elemental analyzer (EA; Elementar Analysensysteme GmbH, Germany) coupled to an Isoprime
 180 vision isotopic ratio mass spectrometer (IRMS; Elementar Analysensysteme, GmbH, Germany)
 181 at the University of Amsterdam. Prior to analysis, dried suspended particulate matter from
 182 the traps was ground using an agate mortar and pestle to homogenize. Subsequently,



183 samples were transferred into silver cups and weighed. The day before measuring, inorganic
 184 carbonates were removed by repeated treatment with 1 mol L⁻¹ hydrochloric acid until
 185 effervescence stopped. The silver cups were dried overnight on a heating plate and closed the
 186 following day with tweezers. For carbon and nitrogen content analyses, three out of the six
 187 replicates collected per location and time point were randomly selected for measurements.

188 **2.5 Grain-size analysis**

189 Grain-size analyses were conducted using a Beckman Coulter laser diffraction particle sizer
 190 LS13 320. Prior to analysis, 2 mL of sodium pyrophosphate (Na₄ P₂ O₇ · 10 H₂O) was added to
 191 each sample, followed by brief sonication for 10 min to ensure complete disaggregation of
 192 any remaining particle aggregates. Degassed water was used during the analysis to reduce the
 193 impact of gas bubbles, and a magnetic stirrer was used to keep the sample homogenized. The
 194 particle-size distributions were categorized into 92 size classes, ranging from 0.375 to 2,000
 195 µm. Testing with internal glass-bead standards demonstrated a reproducibility of better than
 196 0.7 µm for mean particle size and 0.6 µm for median particle size (1σ). The average standard
 197 deviation across all size classes was within 4 vol.%. following Van Der Does et al. (2016).

198 **2.6 XRF elemental analysis**

199 X-ray Fluorescence (XRF) was employed as a non-destructive method to assess elemental
 200 composition in SPM samples, performed with an Avaatech scanner at NIOZ. Sub-samples
 201 (triplicate) were taken from the Milli-Q-washed, freeze-dried sediment trap samples (see
 202 Section 2.3) and homogenized and ground with a pestle and mortar to a fine particle size. The
 203 samples were placed in sample holders with a 6 mm radius, following the procedure of Korte
 204 et al. (2017). The XRF analysis was carried out using two energy settings: 10 kV for aluminum
 205 (Al), silicon (Si), sulfur (S), iron (Fe), and manganese (Mn), and 30 kV for nickel (Ni), bromine
 206 (Br), and lead (Pb). For the 10 kV measurements, a current of 1500 µA was applied for 20
 207 seconds without a filter, while the 30 kV measurements used a 500 µA current for 20 seconds
 208 with a palladium filter to optimize detection. To check precision of our analyses, one external
 209 reference standards (SARM 2) was measured repeatedly and we calculated the coefficient of
 210 variation (CV%) per element for one of the standards (SARM 2), which was consistently below
 211 4%. Each sample was analyzed three times to assess internal variability. Data was processed
 212 using bAxil spectrum analysis software (developed by Brightspec).



213 **2.7 Statistical analysis**

214 First, outliers of mass, carbon, and nitrogen fluxes were identified and removed following
 215 robust statistical methods for outlier detection, using the median absolute deviation (MAD)
 216 approach as described by Rousseeuw and Hubert (2011). This resulted in the removal of 15
 217 mass flux values out of 274 datapoints (5.5 %). These outliers corresponded to exceptionally
 218 low fluxes, which were caused by physical obstruction, specifically, shells clogging the
 219 sediment trap funnels and preventing material from entering the collection tubes, as observed
 220 during sample retrieval. We used the two-sample t-test (independent samples t-test) to
 221 compare the means of mass, carbon and nitrogen fluxes, grain size and silt and clay % between
 222 wet and dry seasons for each location separately. Prior to analysis, data were checked for
 223 normality using the Shapiro–Wilk test and visually assessed using Q-Q plots. Most variables
 224 were not normally distributed. For non-normally distributed variables, we applied the non-
 225 parametric Kruskal–Wallis test followed by pairwise Wilcoxon rank-sum tests with Bonferroni
 226 correction. For normally distributed variables, we used one-way ANOVAs followed by Tukey’s
 227 Honest Significant Difference (HSD) post-hoc tests to assess differences between locations in
 228 the dry and wet seasons. These analyses were conducted separately for each season and
 229 applied to grain size, as well as mass, C_{org} , and total N fluxes. Statistical significance was
 230 determined using a threshold of $p < 0.05$.

231 Elemental counts obtained from XRF analysis were normalized to account for variability in
 232 physical material properties (e.g., density) and measurement conditions (e.g., detector live
 233 time). Normalization was performed by calculating centered log-ratios (CLR) of elemental
 234 intensities, following the method described by Bertrand et al. (2024). This approach addresses
 235 non-linear matrix effects and ensures comparability between measurements. The CLR
 236 transformation involves dividing the intensity of each element by the geometric mean of
 237 intensities of all selected elements from the same measurement. The geometric mean is
 238 calculated as:

$$239 \quad g(z) = \sqrt[n]{Counts_{Element A(z)} * Counts_{Element B(z)} * ... * Counts_{Element n(z)}}$$

240 where n is the total number of selected elements. The CLR value for a specific element is then
 241 computed as:



$$CLR A(z) = \ln \left[\frac{Counts\ Element\ A(z)}{g(z)} \right]$$

Only well-measured elements free of noise and zero values were included in the centered log-ratio (CLR) transformation. Principal Component Analysis (PCA) was then performed on the CLR-transformed elemental data to identify the dominant geochemical processes contributing to differences in sediment composition. Element loadings were used to identify groups of elements with high contributions to the total variance along the first two principal components (PC1 and PC2) using a scree plot. To account for the potential impact of differences in dilution due to a variable terrigenous input, trace metals (e.g., Pb, Ni, Y, Rb, and Cu) were not CLR-transformed, but instead expressed as the Ln of ratios to Al (e.g., Ln (Pb/Al), Ln (Ni/Al)) (Weltje and Tjallingii, 2008). For visualization, average cluster values were calculated per location and season.

To explore the relationship between SPM characteristics and environmental processes across locations and seasons, a Redundancy Analysis (RDA) was performed. In this analysis, response variables (i.e. Mass, C_{org} and total N fluxes, grain size, silt and clay % and CLR of elemental intensities) were modeled as a function of explanatory variables (e.g., rainfall, wind speed, wind direction, distance from land and from the bay mouth). For this analysis, the locations East 8m and 18 m were combined as East, and West 8 m and 18 m were combined as West, as there were no clear differences between the different depths when examining the RDA results. Before performing the RDA, correlations among explanatory variables were assessed, and highly correlated variables were removed to reduce multicollinearity. All variables were auto-scaled and mean-centered, and as suggested by Shrestha and Kazama (2007), only RDAs with eigenvalues greater than 1 were considered significant. Additionally, a k-means clustering algorithm was applied to the site scores, revealing distinct groups within the dataset.

In addition, we determined the relationships between various environmental and SPM variables, by conducting a Spearman correlation analysis for three locations (East, West, and Mouth) within the study area using data of the wet and dry season. Spearman correlations were chosen as most of the variables were not normally distributed. The variables included CLR of elemental intensities, SPM properties (i.e. grain size, silt and clay ratio, mass, C_{org} and total N fluxes), and environmental factors (i.e. rainfall, wind speed and direction and depth). We categorized correlation (Spearman's rho, r_s) strengths as strong ($r_s > 0.6$) and very strong



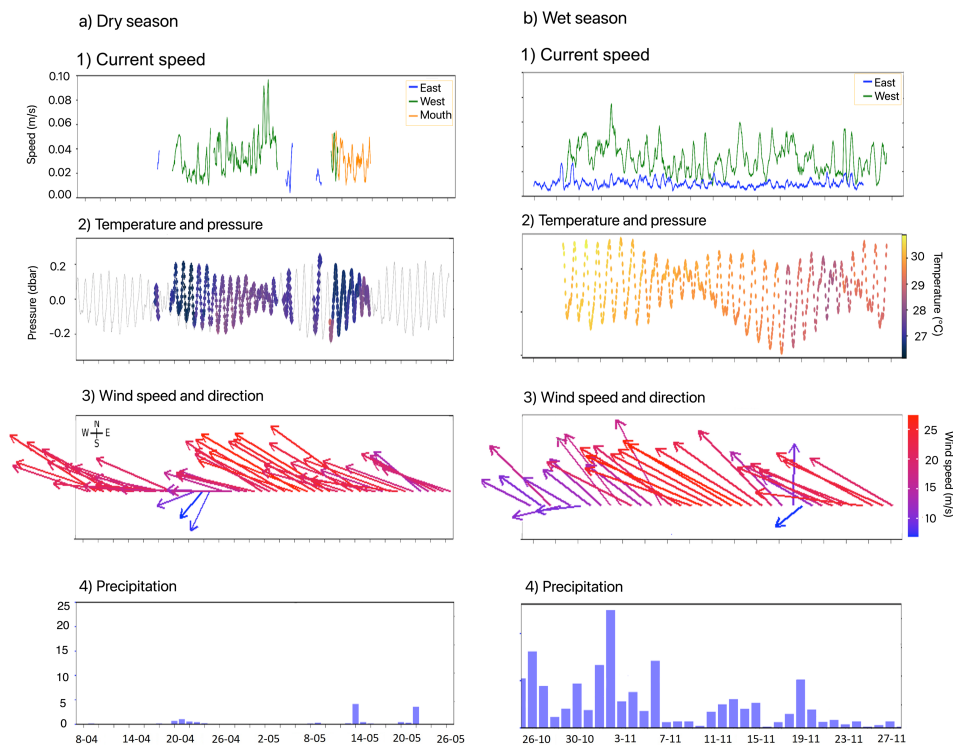
272 ($r_s > 0.8$), while weak to moderate correlations were excluded from the visualizations. All
 273 statistical analyses and graphing were performed in R (R-CoreTeam 2012) using RStudio
 274 (RStudioTeam 2015).

275 **3. Results**

276 **3.1 Environmental parameters: seasonal and spatial differences**

277 Environmental data revealed distinct differences between the dry and wet seasons (Fig. 2).
 278 During the dry season (April–May 2023), recorded water temperatures at 10 m water depth
 279 of the eastern and western reef varied from 26.5–28.8°C, with the highest temperatures
 280 observed between April 24 and May 2, coinciding with the transition from spring tide to neap
 281 tide. Rainfall during this period was minimal, with only two notable events: 4 mm on May 14
 282 and 2 mm on May 22. In contrast, the wet season (October–November 2023) showed higher
 283 water temperatures, exceeding 30°C from October 28 to November 11 and averaging
 284 approximately 29°C throughout the rest of November. Rainfall was substantially higher than
 285 in the dry season, with precipitation recorded on nearly every sampling day. The most intense
 286 rainfall occurred between October 25 and November 6, peaking at 24 mm on November 2.
 287 Wind also showed distinct seasonal patterns. The dry season was characterized by consistent
 288 southeasterly winds with high average speeds of approximately 20 m s⁻¹, reaching peaks of
 289 27.4 m s⁻¹. The wet season exhibited more variable wind directions, predominantly from the
 290 southeast as well, with occasional extreme shifts (e.g., the change in direction recorded on
 291 November 20). Wet season wind speeds were notably lower, averaging around 15 m s⁻¹,
 292 particularly in early October, though they increased by mid-November. The wet season also
 293 showed a relationship between precipitation events and wind patterns, with lower wind
 294 speeds often coinciding with heavier rainfall.

295 In addition to seasonal environmental differences, the ADCP data revealed spatial variations
 296 in water flow between the eastern and western reefs (Fig. 2). Current velocities were notably
 297 higher at the western reef, ranging from 0.01 to 0.07 m s⁻¹, compared to the eastern reef,
 298 where velocities ranged mostly from 0.005 to 0.02 m s⁻¹. Although limited dry season data
 299 prevented robust seasonal comparisons, water velocities remained within similar ranges
 300 across both seasons



301

302 **Figure 2.** Environmental variables during the sampling periods in the dry and wet seasons.
303 Panels a) shows current speed over time at the western reef at 8 m (green), eastern reef at 8
304 m (blue) and mouth at 3 m (only in the dry season; orange). Panels b) show water pressure at
305 10 m in the reef in dbar (reflecting tidal cycles) and water temperature in °C, the latter
306 represented by a color scale. Panels c) show wind speed (color of the vectors) and direction.
307 Panels d) show rainfall (mm).

308

3.2 Seasonal and spatial averages of fluxes and grain size

309

Properties of SPM showed clear spatial and seasonal patterns (Table 1). In the dry season,
310 fluxes (total mass, organic carbon (C_{org}), and total N) were significantly higher at the mouth
311 than at the reef locations (Tables 1, S1). In the western reef, SPM had a larger grain size (with
312 a significantly lower proportion of silt and clay) compared to the east and the mouth (Tables
313 1, S1). In the wet season, the mouth exhibited the highest fluxes, with higher silt and clay
314 contents compared to the material found at the other locations (Table S1). At most sites, mass,
315 C_{org} and total N fluxes were higher in the wet season compared to the dry season, except for
316 west 18 (Table S2). The mouth showed the strongest seasonal signal: fluxes were 3–4 times



317 higher in the wet season and grain size was significantly smaller than in the dry season (Table
 318 1, S2).

319 At the east reef, seasonal differences were also evident, especially at 8 m, where mass fluxes
 320 doubled, and C_{org} and total N fluxes increased fourfold in the wet season. At 18 m, this
 321 seasonal increase in fluxes was more modest, but still clear (Table S2).

322 Western reefs showed a mixed pattern. At 8 m, fluxes increased in the wet season, while at
 323 18 m, they were higher in the dry season, contrasting with the other locations. Nonetheless,
 324 average grain size was smaller, and silt and clay contents were higher in the wet season across
 325 all reef sites (i.e. eastern and western reef; Table 1, S1, S2). In particular, during the dry season,
 326 the western reef had the lowest portion of silt and clay (28–38%), which was significantly
 327 lower than at the eastern reef (51–53%).

328 **Table 1.** Summary of the mean (highest between seasons per site in bold), maximum and
 329 minimum values of mass, C_{org} and total N fluxes ($\text{g m}^{-2} \text{d}^{-1}$), grain size (μm), silt and clay %
 330 measured at the mouth east 8 and 18 m, and west 8 and 18 m during the dry and wet season.
 331 Asterisks show significant differences between seasons (See Tables S1 and S2 for full statistical
 332 analyses between locations and seasons, respectively).

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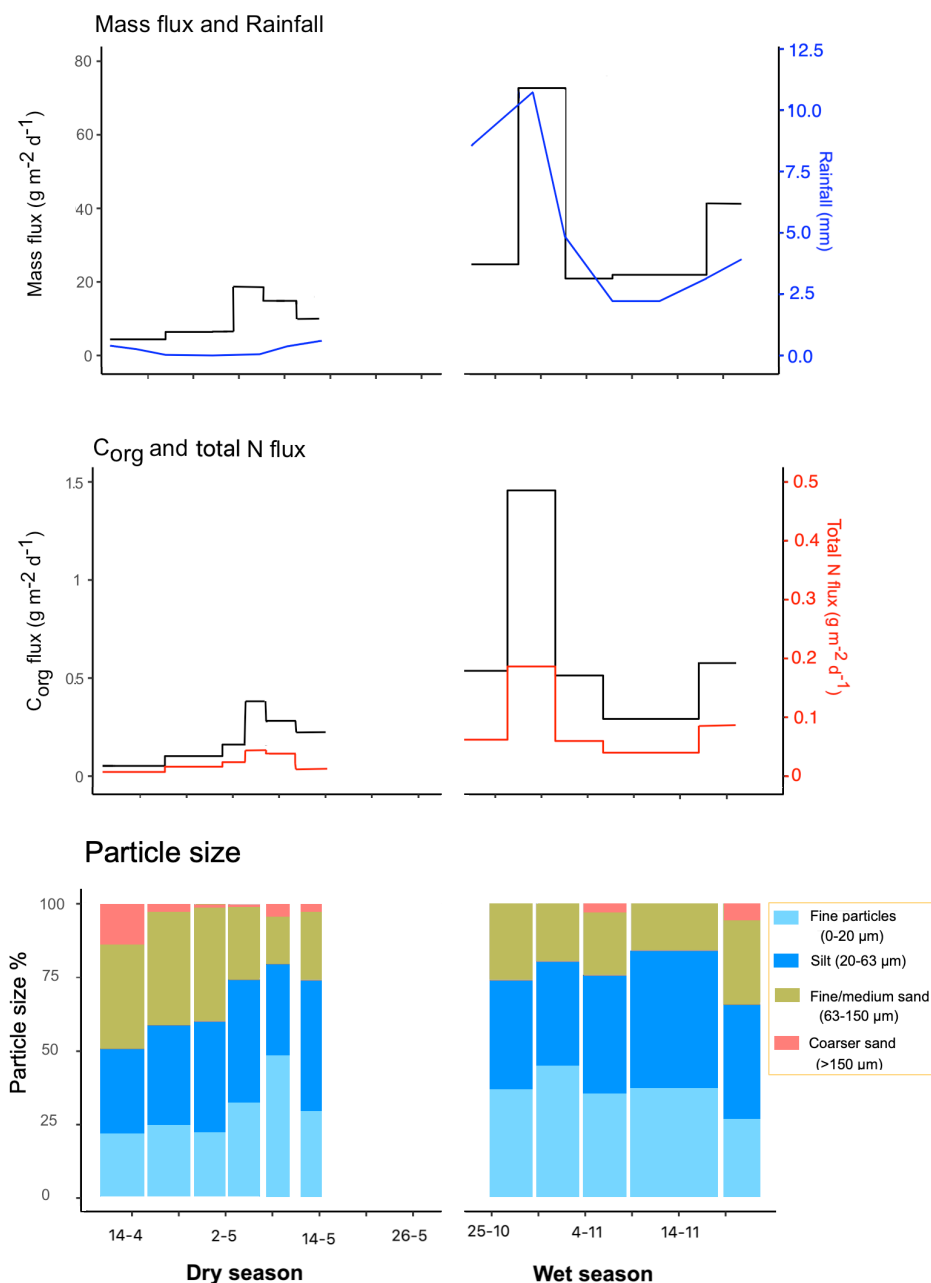
		Mouth		East 8 m		East 18 m		West 8 m		West 18 m	
		Dry	Wet	Dry	Wet	Dry	Wet	Dry	Wet	Dry	Wet
Mass flux ($\text{g m}^{-2} \text{d}^{-1}$)	Mean	8.47	36.32*	5.35	9.52*	6.60	8.99	5.23	10.61	7.48	5.23
	Max	24.11	84.43	6.54	18.74	11.75	36.61	13.28	39.45	12.93	15.12
	Min	1.56	16.65	2.27	1.16	3.53	0.97	1.31	2.58	3.31	2.07
C_{org} flux ($\text{g m}^{-2} \text{d}^{-1}$)	Mean	0.18	0.61*	0.05	0.17	0.09	0.21	0.06	0.11	0.12	0.09
	Max	0.53	0.17	0.09	0.49	0.18	0.55	0.20	0.63	0.24	0.16
	Min	0.03	0.26	0.01	0.03	0.05	0.06	0.04	0.04	0.04	0.04
Total N flux ($\text{g m}^{-2} \text{d}^{-1}$)	Mean	0.031	0.082*	0.007	0.031*	0.012	0.021	0.022	0.033	0.022	0.012
	Max	0.081	0.233	0.011	0.074	0.024	0.083	0.085	0.112	0.043	0.023
	Min	0.004	0.031	0.001	0.007	0.007	0.012	0.005	0.005	0.008	0.004
Grain Size (μm)	Mean	45.53	31.46	62.62	51.17	85.96	35.41	86.46*	42.88	95.34*	43.75
	Max	74.85	52.05	279.46	81.10	273.09	54.79	125.05	60.74	159.49	62.28
	Min	36.09	23.52	27.35	27.81	26.16	11.65	36.09	25.08	25.62	15.96
Silt and Clay %	Mean	61.67	75.06	52.96	55.30	51.69	61.84	38.16	62.41*	28.65	57.00*
	Max	79.93	96.54	72.03	77.18	77.40	88.79	58.06	79.41	40.54	81.92
	Min	49.45	50.40	25.14	35.59	37.27	47.63	25.52	43.40	22.21	39.82



334 **3.3 Seasonal trends of mass fluxes and SPM properties**

335 **3.3.1 Mouth**

336 Fluxes at the mouth exhibited temporal variability within each season. During the dry season
337 mass, C_{org} , and total N fluxes remained relatively stable over time. In contrast, during the wet
338 season, fluxes were significantly higher and associated with a finer grain-size distribution (Fig.
339 3, Table 1). The proportion of fine particles (0–20 μm) and silt (20–63 μm) increased in the
340 wet season, while the proportion of fine-to-medium sand (63–150 μm) and coarser sand
341 (>150 μm) decreased compared to the dry season (Fig. 3). The highest fluxes observed at the
342 mouth ($\sim 70 \text{ g m}^{-2} \text{ d}^{-1}$) directly followed a heavy rainfall event on November 2nd. At this time,
343 the average grain size of settled material was particularly small ($\sim 24 \mu\text{m}$), with the fine
344 particles and silt making up $\sim 80\%$ of the material.



345

346 **Figure 3.** SPM fluxes and grain-size characteristics of the mouth of the dry season and wet
 347 season. The first panel shows mass ($\text{g m}^{-2} \text{d}^{-1}$) and average rainfall (mm). Second panel shows
 348 C_{org} and total total N fluxes ($\text{m}^{-2} \text{d}^{-1}$). Note that mass, organic carbon, and total nitrogen flux



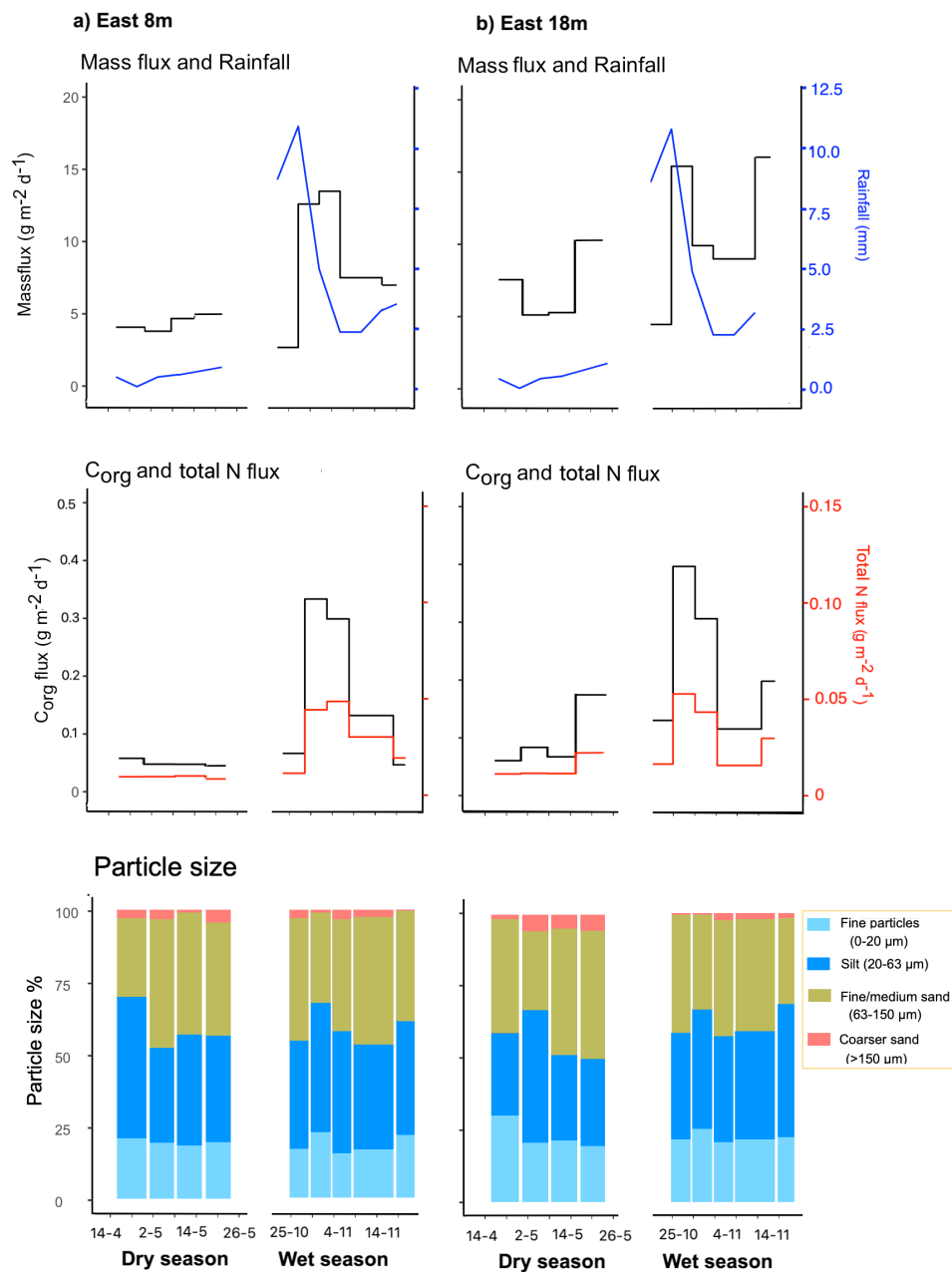
scales are different than in Figure 4 and 5. Bottom panel shows the % of fine particles (0-20 μm) silt (20-63 μm), fine-medium sand (20-150 μm) and coarser sand ($>150 \mu\text{m}$).

351

352 **3.3.2 East 18 and 8 m depth**

353 At the eastern reef, and a depth of 8 m fluxes remained relatively constant throughout most
 354 of the sampling period in the dry season (Figure 4.a). Here, mass flux increased during May 2–
 355 16, although C_{org} and total total N fluxes did not show a corresponding increase. At 18 m water
 356 depth there was higher variation during the sampling periods. Also at this depth, the highest
 357 mass, C_{org} and total total N fluxes were recorded towards the end of the sampling period (May
 358 16–24). Grain-size composition was constant during the whole sampling period at both 8 and
 359 18 m.

360 Over the wet season sampling period, the lowest mass, C_{org} and total N fluxes were recorded
 361 from October 25–31 (mass flux: $2\text{--}5 \text{ g m}^{-2} \text{ d}^{-1}$, C_{org} flux: $0.06\text{--}0.1 \text{ g m}^{-2} \text{ d}^{-1}$ and total N flux:
 362 $0.004\text{--}0.01 \text{ g m}^{-2} \text{ d}^{-1}$), while relatively high fluxes occurred from October 31–November 5
 363 (mass flux: $12\text{--}15 \text{ g m}^{-2} \text{ d}^{-1}$, C_{org} flux: $0.3\text{--}0.4 \text{ g m}^{-2} \text{ d}^{-1}$ and total N flux: $0.04\text{--}0.05 \text{ g m}^{-2} \text{ d}^{-1}$),
 364 coinciding with the highest rainfall event of the sampling period (November 2). During this
 365 high-rainfall period, SPM exhibited smaller grain size, and a higher percentage of silt and clay
 366 similar to what was observed at the mouth. At both depths, the proportion of coarser sand
 367 ($>150 \mu\text{m}$) was lower in the wet season compared to the dry season.



368

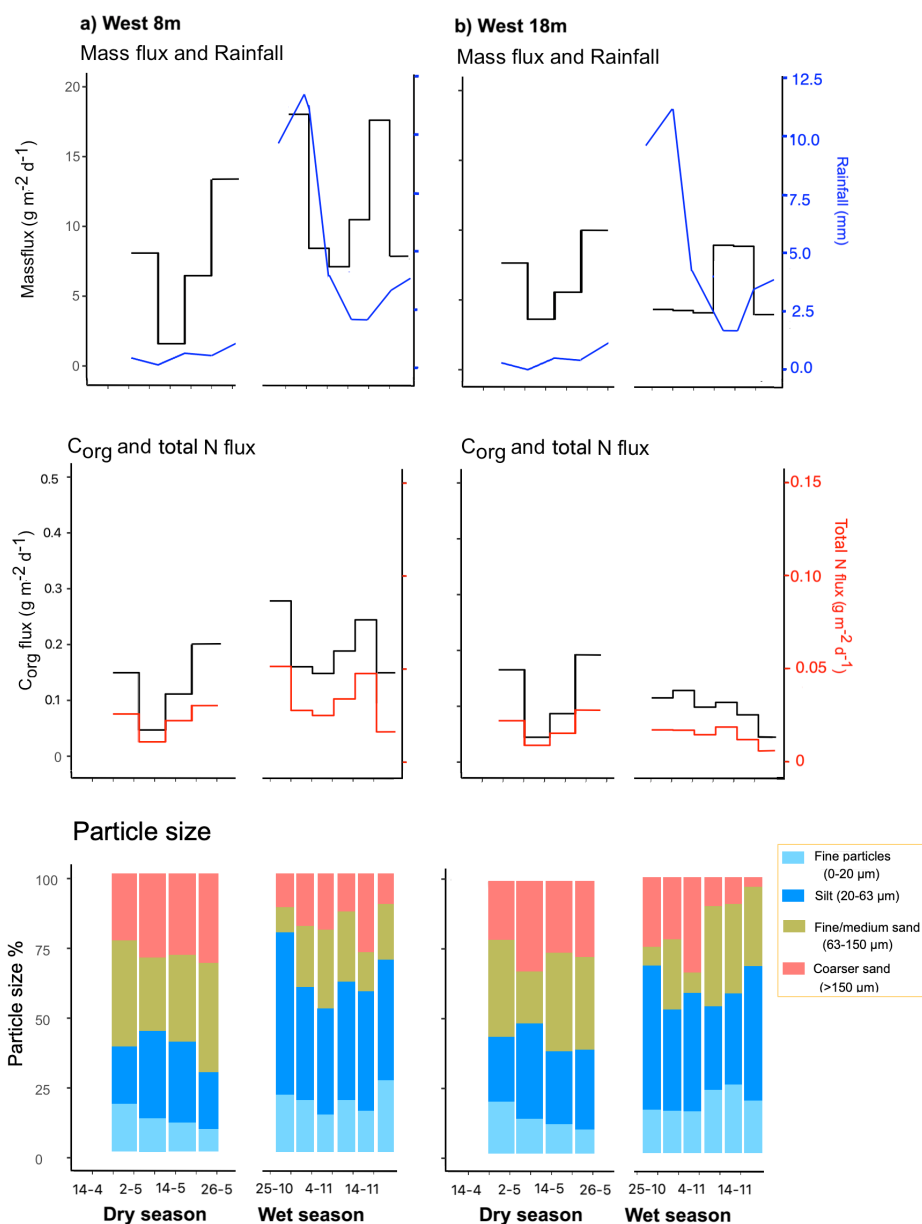
369 **Figure 4.** SPM fluxes and grain size of the a) East 8m and b) East 18m of the dry season and
 370 wet season. The upper panels show mass fluxes ($\text{g m}^{-2} \text{d}^{-1}$) and average rainfall (mm). Middle
 371 panels show C_{org} and total total N fluxes ($\text{g m}^{-2} \text{d}^{-1}$). Bottom panel shows the proportion of fine



372 particles (0–20 μm) silt (20–63 μm), fine-medium sand (20–150 μm) and coarser sand (>150
 373 μm).

374 **3.3.3 West 18 m and 8 m depth**

375 Seasonal differences at the western reef were less pronounced compared to those observed
 376 at the other locations. At 18 m, mass, C_{org} , and total total N fluxes remained relatively stable
 377 both within and between seasons (Fig. 5). However, the material was coarser at this site in
 378 both seasons, with a higher proportion of coarse sand (>150 μm) than at the mouth and
 379 eastern reef. This coarser fraction was particularly dominant in the dry season. At 8 m, mass,
 380 C_{org} , and total total N fluxes showed a clear seasonal difference. During the dry season, highest
 381 mass flux was recorded between May 8 and 18 ($\sim 13 \text{ g m}^{-2} \text{ d}^{-1}$), coinciding with one of the few
 382 rainfall events in that period (May 14). In the wet season, fluxes at 8 m displayed greater
 383 temporal variation, with peak values around the intense rainfall on November 2 (mass flux: 18
 384 $\text{g m}^{-2} \text{ d}^{-1}$; C_{org} flux: 0.3 $\text{g m}^{-2} \text{ d}^{-1}$; total N flux: 0.05 $\text{g m}^{-2} \text{ d}^{-1}$). Grain-size analysis showed that
 385 the percentage of fine particles (0–20 μm) and silt (20–63 μm) was higher in the wet than in
 386 the dry season, particularly at 8 m. In contrast, the proportion of coarser sand was lower in
 387 the wet season and generally higher at both 8 and 18 m compared to the mouth and eastern
 388 reef. At 18 m, mass fluxes were lower than at all other locations (i.e., West 8 m, East 8 m and
 389 18 m, and the mouth).



390

391 **Figure 5.** SPM fluxes and grain size of the a) West 8m and b) West 18m of the dry season and
 392 wet season. Upper panels show mass fluxes ($\text{g m}^{-2} \text{d}^{-1}$) and average rainfall (mm). Middle
 393 panels show C_{org} and total total N fluxes ($\text{g m}^{-2} \text{d}^{-1}$). Bottom panel shows the % of fine particles
 394 (0-20 μm) silt (20-63 μm), fine-medium sand (63-150 μm) and coarse sand (>150 μm).

395

3.4 Elemental composition of the suspended matter

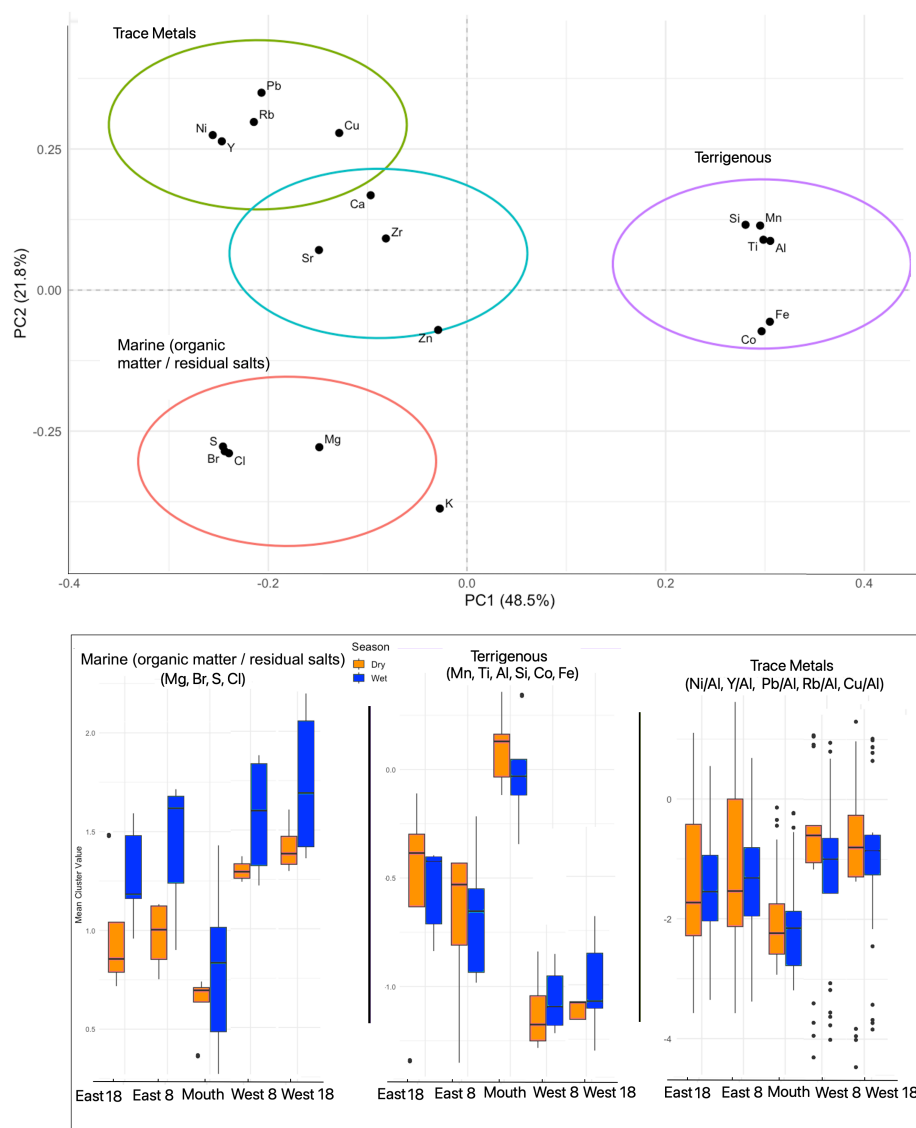


396 A principal component analysis (PCA) was performed to identify the largest patterns in the
397 settled material elemental composition (Fig. 6). This revealed four distinct clusters of
398 elements: the first cluster, comprising Si, Mn, Ti, Al, Co and Fe (purple circle Fig. 6), reflects a
399 dominant terrigenous origin probably related to runoff, with their distribution being most
400 pronounced at the bay mouth and east reef locations. A second cluster, consisting of Ni, Y, Rb,
401 Pb, and Cu (green circle Fig. 6), consists of trace elements generally associated with pollution,
402 which were found in elevated concentrations particularly at the reef sites. A third cluster,
403 defined by Br, S, Cl, and Mg (red circle Fig. 6), consists of elements typical of marine organic
404 matter and/or salts indicating a strong seawater influence, likely reflecting regions with higher
405 marine impact, such as at the western reef. A fourth, less well-defined cluster included Sr, Ca,
406 Zn, and Zr (blue circle Fig. 6), elements associated with marine carbonates (Ca and Sr) as well
407 as terrigenous (Zr) and pollution (Zn) sources. As this cluster is located central in the PCA space
408 it is likely not specifically related to any of these factors.

409

410

411



412

413 **Figure 6.** Principal Component Analysis (PCA) cross plot of PC1 (48.5%) and PC2 (21.8%)
 414 showing the loadings of XRF-derived elemental ratios. Four distinct element clusters are
 415 visible: (1) a terrigenous cluster (Si, Mn, Ti, Al, Co, Fe) loading positively on PC1; (2) a trace
 416 metal cluster (Ni, Y, Rb, Pb, Cu) with positive loadings on PC2 and negative on PC1; (3) a marine
 417 cluster typically associated to organic matter and/or residual salts (Br, S, Cl, Mg); and (4) a
 418 fourth group (Sr, Ca, Zn, Zr) with unclear affiliation. Spatially, the terrigenous elements were
 419 more prominent at the bay mouth and east reef, trace metals were highest at the east reef,
 420 and marine salt elements dominated the western reef.

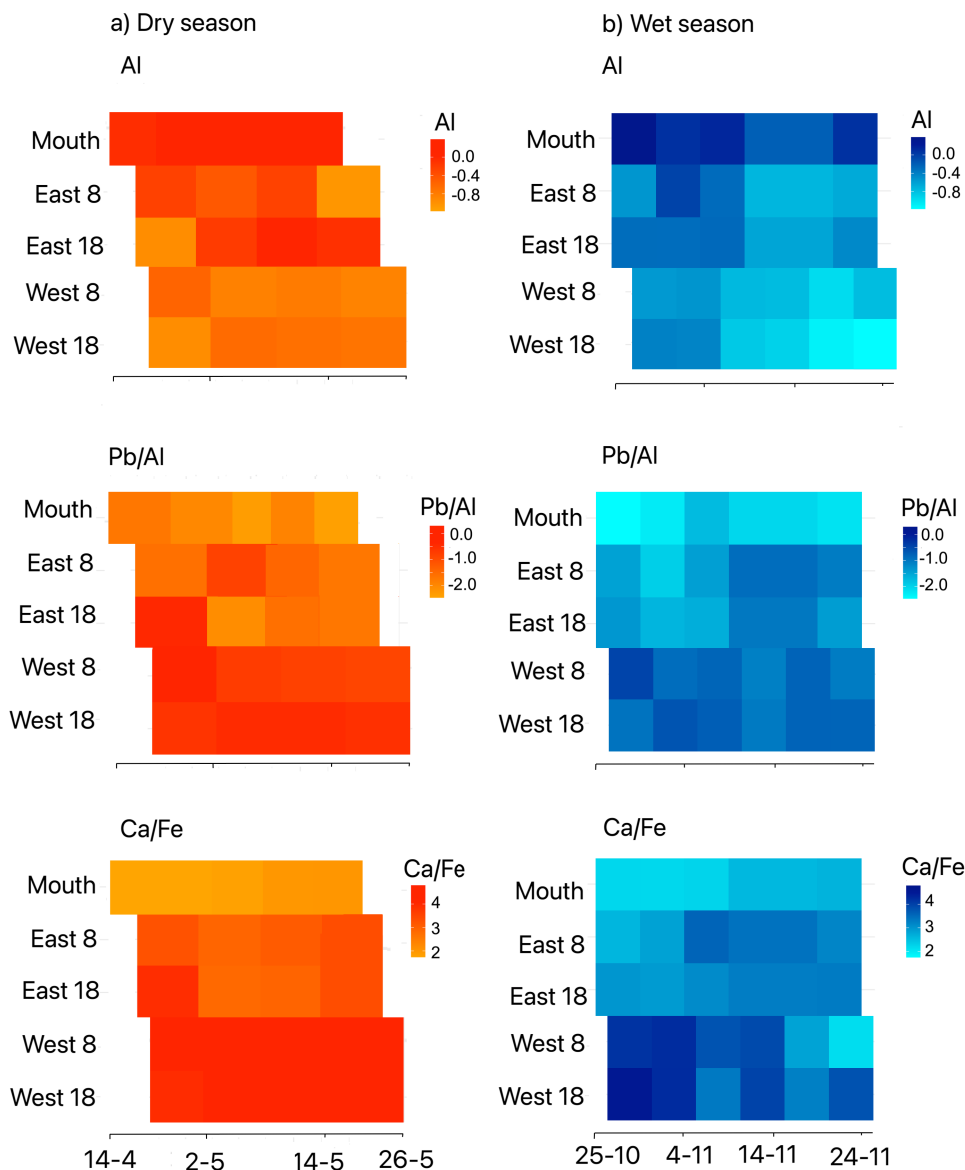


421 **3.4.1 Seasonal and spatial trends of elemental composition and correlations of SPM** 422 **characteristics with environmental parameters**

423 XRF analysis revealed distinct relative spatial differences in the composition of SPM (Fig. 7).
 424 Aluminum (Al), an element associated to terrigenous input, displayed the highest intensities
 425 at the mouth, and eastern reef, and the lowest at the western reef. This spatial pattern was
 426 consistent across both seasons. During the wet season Al values increased following periods
 427 of intense rainfall (25 October–8 November, Fig. 2). Other terrigenous elements showed
 428 similar trends (Fig. S1). In contrast, the Pb/Al ratio, indicative of anthropogenic pollution, was
 429 highest in the western reef, lowest at the mouth and showed intermediate values in the east.
 430 This pattern held for both seasons and was consistent across other trace metal ratios (Fig. S2).
 431 The Ca/Fe ratio, reflecting the relative contribution of marine carbonate versus terrigenous
 432 inputs, exhibited a similar spatial pattern, with the highest values found in the western reef,
 433 intermediate values in the east, and the lowest at the mouth. Additionally, Ca/Fe values were
 434 higher during the dry season, showing a greater influence of marine carbonate inputs during
 435 periods of reduced runoff.

436 Correlations between SPM properties and environmental drivers confirmed these spatial
 437 contrasts (Fig. S3). At the mouth, mass, C_{org} , and total N fluxes were strongly intercorrelated
 438 (>0.9) and closely linked to Al, Si, Mn, and Fe, indicating a dominance of fine-grained,
 439 terrigenous material. Rainfall and wind also correlated with fluxes during the wet season. At
 440 the eastern reef, fluxes were again intercorrelated, but showed weaker associations with
 441 rainfall. Correlations with Fe and Mn point to a stronger influence of silt and clay fractions. In
 442 contrast, at the western reef, fluxes correlated less with environmental drivers, but grain size
 443 showed a stronger role. Here, coarser particles were occasionally linked to higher mass and
 444 organic matter fluxes in the dry season, whereas in the wet season fluxes were more
 445 influenced by fine-grained material and terrigenous elements such as Al and Fe.

446



447

448 **Figure 7.** Heat maps of elemental intensities of Al, Pb/AL and Ca/Fe in sediment trap samples
449 during the a) dry season and b) wet season. The x-axis represents sampling dates, and the y-
450 axis displays the five sampled locations (Mouth, East 8, East 18, West 8 and West 18 m). The
451 color scale indicates the centered log-ratios (CLR) of elemental intensities for Al and log-ratios
452 for Pb/Al and Ca/Fe.

453

454 **4. Discussion**



455 Near the bay's mouth, the highest fluxes of suspended matter (SPM) were found and this
 456 material was finer with elevated values of terrestrial elements (Si, Fe, Al, and Mn) than in
 457 locations further away from the bay. In contrast, the location west of the bay's mouth showed
 458 lower concentrations of SPM and the settled material was coarser grained with higher values
 459 of Pb, Ni, and Cu. These spatial differences show that the suspended matter near the bay's
 460 mouth and eastern reef is primarily influenced by material derived from terrestrial runoff,
 461 whereas the western reef reflects a stronger open-ocean and more marine signature. Notably,
 462 despite receiving less terrestrial input, the reef sites—particularly the western reef—showed
 463 the strongest signals of anthropogenic pollution (e.g., Pb, Rb, Cu, Ni), suggesting that fluxes
 464 and deposition are decoupled and that contaminants accumulate preferentially at these sites.
 465 Differences between seasons are caused by changes in runoff due to variations in rainfall and
 466 wind regimes.

467 **4.1 Bay influence and marine processes**

468 The influence of Piscadera Bay on SPM content in nearshore waters reveals a clear spatial and
 469 temporal pattern. At the bay mouth, highest mass fluxes, elevated organic content, and finest
 470 particles were found. While comprehensive studies using sediment traps to examine spatial
 471 gradients in suspended particulate matter characteristics in bay-reef systems remain scarce,
 472 this pattern is consistent with findings from other Caribbean coastal systems, where sediment
 473 accumulation rates show strong spatial gradients related to proximity to terrestrial sources
 474 (Toro et al., 2021). Similarly, the Magdalena River system in Colombia, which discharges
 475 exceptionally high sediment loads directly into the Caribbean Sea, illustrates how major
 476 terrestrial inputs create pronounced sediment flux gradients in Caribbean coastal waters, with
 477 higher amounts and finer material found close to the river mouth (Restrepo et al., 2014). The
 478 elemental composition, particularly the high values of Si, Al, Mn, Co and Fe in this study, point
 479 to a terrestrial origin of SPM as these lithogenic elements are fundamental markers of soil
 480 erosion and terrestrial runoff (Larson et al., 2015). Together these findings show a transition
 481 from mainly terrestrial input to more marine influenced SPM dynamics (indicated by fluxes
 482 and XRF elements) with increased distance from the bay. This pattern aligns with several
 483 studies showing that SPM concentrations and their negative effects on coral reefs decrease
 484 with distance from sources, such as rivers or bays (Fabricius, 2005; Rogers and Ramos-
 485 Scharrón, 2022; Storlazzi et al., 2011).



486 East of the bay mouth, the suspended matter shows more impact of the bay system compared
 487 to the western reef, however less pronounced than at the mouth. This aligns with Sánchez
 488 Baranco et al. (2025), showing that bay-derived characteristics, such as dissolved-nutrient
 489 concentrations, salinity, and temperature attenuate with increasing distance from bays. The
 490 eastern reef, particularly at 8 m depth, presents a recognizable bay signature through elevated
 491 carbon and nitrogen content and relatively high values of elements, such as Al, Si or Fe,
 492 indicating terrestrial runoff. The low current speeds measured at the eastern reef (8 m depth;
 493 Fig. 2) can explain this accumulation of finer material with higher organic content, creating a
 494 zone where bay influence lasts longer due to longer residence times. Safak et al. (2015)
 495 demonstrated through modeling that in areas surrounding shallow coastal bays—particularly
 496 near inlets where water circulation is restricted—tidal forces can enhance material
 497 accumulation. This pattern is supported by our data, showing elevated concentrations of
 498 terrigenous elements (Fig. 7) at the mouth and eastern reef, and strong correlations between
 499 water pressure (used here as a proxy for tidal variation, Fig. S3) and both mass fluxes and
 500 coarser material, especially at the bay mouth. This likely indicates that during periods of higher
 501 water pressure, such as high tide or spring tide, larger quantities of suspended matter,
 502 including coarser particles, are transported near the mouth of the bay.

503 The western reef, positioned furthest from Piscadera Bay, reveals different patterns compared
 504 to the mouth and eastern reef. This site is characterized by less variability in fluxes of mass,
 505 C_{org} , and total N. The consistently higher Ca/Fe found here further suggest a stronger marine
 506 signal at this site compared to the mouth and eastern reef, reflecting greater contributions of
 507 carbonate-rich marine particles relative to terrigenous inputs (Nizou et al., 2011). However,
 508 this location shows relatively high values of Br, S, Pb, and Ni. Trace metals were found at both
 509 reefs, but values were consistently higher at the western reef, particularly for Pb and other
 510 pollution indicators (Figs. 7, S2). These trace elements suggest a complex interplay of marine
 511 biological activity and potential anthropogenic influences, with origins ranging from marine
 512 organism decomposition to maritime and industrial processes (Caballero-Gallardo et al., 2020;
 513 Fernandez et al., 2007). The presence of Pb and Ni indicates potential anthropogenic input,
 514 such as maritime activities, or coastal infrastructure (Nalley et al., 2021; Fernandez et al.,
 515 2007) and corroborates with high anthropogenic input found related to the close proximity of
 516 Piscadera Bay to urbanized areas around Willemstad, the largest city on Curaçao (Van de



517 Loosdrecht, submitted). The presence of pesticides, or industrial processes and persistence
518 and accumulation of these elements at the western reef may be partly explained by the site's
519 three-dimensional reef structure, which can act as an efficient trap for suspended and settling
520 particles. Greater structural complexity increases the capacity of the reef to retain material,
521 allowing for longer particle residence times and the potential for local enrichment of both
522 natural and anthropogenic elements (Ng et al., 2022; Bothner et al., 2006).

523 Hydrodynamics also shape the conditions for suspension and settling of particulate matter at
524 this site, with current speeds higher than those in the eastern reef (Fig. 2). The effect of current
525 speed appears to be depth-dependent: at 8 m water depth we still observe bay or land
526 influence, as evidenced by fluctuating fluxes during the wet season that correspond with
527 rainfall events (Fig. 5). However, at 18 m water depth, this effect is not noticeable, suggesting
528 diminished influence from surface runoff at greater depths. The higher current velocities
529 affect SPM dynamics by reducing particle settling time and increasing resuspension, resulting
530 in winnowing of fine material from the seabed and increased turbidity in the water column
531 (De Jong et al., 2025; Prochnow et al., 1970; Cuttler et al., 2017). During 2023, the model by
532 (Bertoncelj et al., 2025) identified a high presence of mesoscale eddies in the region, which
533 may have contributed to the transport of particulate material from areas beyond the
534 immediate vicinity of the study site. These spatial and seasonal patterns are further supported
535 by a redundancy analysis (Fig. 8), revealing that approximately 49% of the total variance in
536 SPM characteristics is explained by spatial factors including depth, distance to land, and
537 proximity to the mouth of Pescadara Bay. In line with our results, Jouon et al. (2008) showed
538 that fluxes of suspended particles have higher variability over space than over time. This
539 analysis clearly differentiated the western reef (characterized by marine-associated elements)
540 from the bay mouth (dominated by terrestrial markers), with the eastern reef showing
541 intermediate characteristics. A portion of variance (7%) is attributable to seasonal factors,
542 primarily driven by rainfall, wind speed, and wind direction. This multivariate approach
543 confirms that the bay's influence follows both spatial gradients and responds to seasonal
544 changes in environmental conditions.



545

546 **Figure 8.** Redundancy analysis (RDA) biplot of sediment trap data collected from three
547 locations in the bay (mouth, East, and West) during dry and wet seasons. Points are colored
548 by season and shaped by location, with ellipses representing clusters at 95% confidence. RDA1
549 explains 48.57% of the variance and is primarily determined by depth, distance to land, and
550 distance to the mouth, separating the locations. RDA2 accounts for 7.05% of the variance and
551 is driven by rainfall, wind speed, and wind direction, distinguishing between seasons

552

553 4.2 Seasonal modulation of processes

554 Mass fluxes recorded at both reef sites during the dry season fall within the ranges reported
555 for reefs worldwide that are relatively uninfluenced by human activities (i.e., $1\text{--}10\text{ g m}^{-2}\text{ d}^{-1}$)
556 (Rogers, 1990). However, during the wet season, this range is greatly exceeded at both reef
557 sites (except for the western reef at 18 m water depth), with mass fluxes reaching $18\text{--}40\text{ g m}^{-2}$
558 d^{-1} . In addition, organic matter fluxes measured in reefs near other bays in Curaçao (e.g. Santa
559 Martha) are half of the fluxes we measured in both the east (8 and 18 m) and west (8 m)
560 during the wet season (De Jong et al., 2025).

561 The bay mouth exhibits the most pronounced correlations (Fig. S3), particularly during the
562 wet season, where fluxes and elemental concentrations strongly align with rainfall and wind



563 dynamics. This suggests that the mouth functions as a conduit for terrestrial and marine
 564 material exchange, with runoff and hydrodynamics controlling SPM composition. Similar
 565 patterns have been observed on other Caribbean islands, with increased presence of elements
 566 indicating runoff such as Fe, Si, and Al following rain events (Larson et al., 2015). Although Si
 567 can also indicate biogenic silica (i.e. diatoms) (Lobo et al., 2016; Palmer and Abbott, 1986), the
 568 alignment of Si with Fe and Al trends suggests it primarily reflects terrestrial runoff in our case.
 569 In addition, groundwater in Curaçao has been shown to contain elevated nitrogen and trace
 570 metal concentration in urban areas, largely due to wastewater leakage, and this contaminated
 571 groundwater discharges into coastal zones and bays, posing a risk for nutrient enrichment in
 572 these areas (Wit et al., 2025).

573 At the eastern reef, a similar but less pronounced trend is observed, with environmental
 574 influences becoming more evident during the wet season. In contrast, the western reef shows
 575 minimal correlations in the dry season, indicating that local sedimentary processes such as
 576 resuspension likely determine SPM characteristics. Still, correlations increase during the wet
 577 season, indicating an increased influence of rainfall-driven processes and enhanced SPM
 578 transport in this area as well.

579 East of the bay's mouth, the seasonal variability in suspended matter, both its abundance and
 580 composition, is less pronounced than at the mouth. At 8 m water depth, mass fluxes in the
 581 wet season were highest. At 18 m however, mass fluxes show no significant seasonal
 582 differences although elemental compositions vary consistently between 8 and 18 m water
 583 depths, suggesting a depth-dependent gradient in the bay's influence. This vertical variability
 584 may be driven by several processes: fine-grained particles originating from the bay or from
 585 runoff may settle closer to shore before reaching deeper waters (McCave, 1975); differences
 586 in current speeds or directions between depths may influence the transport and resuspension
 587 of particles (Gardner et al., 2018); and other environmental conditions such as stratification
 588 or wave-induced mixing may act to retain suspended material at shallow depths.

589 The seasonal and spatial correlation patterns highlight the important role of environmental
 590 conditions, particularly precipitation and runoff, in modulating SPM transport and
 591 composition (Fig. S3). Recent studies have documented enhanced bay–ocean exchange during
 592 the wet season in Curaçao, resulting in increased mobilization and transport of SPM and
 593 associated elements along the southwest coast reefs (Sánchez Barranco et al., 2025; Van de



Loosdrecht et al., submitted). This exchange favors the suspension of finer particles (higher silt and clay), which provide greater surface area for adsorption of metals and organic matter, enriching particle-bound elements such as Cu, Zn, Pb, and Ni (Lakhan et al., 2003; Malarkey et al., 2015; Bergamaschi et al., 1997).

Flux–environment correlations were stronger in the wet season, with reef fluxes 2–3 times higher during peak rainfall (Figures 4, 5, 7), consistent with reported seasonal increases (Ismail et al., 2005; Li et al., 2013) and even extreme storm-driven surges (>1000-fold; Risk, 2014). In contrast, reduced exchange in the dry season, as warmer and saltier bay waters accumulate, traps nutrients and pollutants (Sánchez Barranco et al., 2025). Our current-speed data (Figure 2) confirm this seasonal contrast. In the wet season, however, mixing with offshore waters intensifies material export, as reflected in the stronger correlations between fluxes and environmental drivers at the mouth and eastern reef.

High rainfall during the 2023 El Niño amplified wet-season dynamics by increasing runoff and wastewater inputs, leading to elevated fluxes (Table 1) and higher Si, Al, Fe, and Mn at the bay mouth (Figure 6). Wastewater management on the island is very limited, with only ~20% of the population connected to sewage systems (Wit et al., 2025). Additional anthropogenic inputs, including the emergency sewage overflow (Govers et al., 2014), deliver organic-rich wastewater to reefs during heavy rainfall, enhancing C_{org} and total N fluxes and strengthening their correlations with Fe and Mn (Figure 8). Resuspension of polluted bay sediments during the wet season further exposes reef organisms to toxic pollutants (De Jong et al., 2025).

4.3 Ecological implications

The observed SPM flux patterns reveal a clear spatial gradient of terrestrial material input, with potentially significant ecological consequences. Increased eutrophication, sedimentation, and decreased herbivory have been identified as key factors reducing coral resilience (Littler and Littler 1998; Kuffner et al. 2008). Dutra et al. (2006) showed that SPM fluxes higher than $10 \text{ g m}^{-2} \text{ d}^{-1}$, which in our study were recorded in both reefs during the wet season, correlate with lower coral larvae survival and settlement and negatively affect coral growth. The eastern reef, more directly connected to the bay, receives higher concentrations of organic carbon, nitrogen, and lithogenic material compare to the western reef. Fine-grained material can have detrimental effects on reef ecosystems, primarily by altering benthic



624 conditions and preventing coral larvae settlement (Risk, 2014; Hodgson, 1990). In addition,
 625 increased water turbidity diminishes photosynthetically available radiation, which is crucial
 626 for coral and algal photosynthesis (Rogers and Ramos-Scharrón, 2022; Weber et al., 2012;
 627 Fabricius, 2005). Although elevated organic content may initially appear beneficial, chronic
 628 nutrient enrichment has been shown to exacerbate coral disease prevalence and bleaching
 629 (Aronson et al., 2003; Vega Thurber et al., 2014), while also promoting abundances of
 630 opportunistic organisms such as turf algae, macroalgae, and bacterial communities that can
 631 rapidly colonize available substrate and outcompete corals for space and light resources,
 632 ultimately leading to phase shifts from coral-dominated to algae-dominated reef systems
 633 (Gorgula and Connell, 2004; Vermeij et al., 2010).

634 In the last decades, studies have shown the effect of these bays in nearby reefs and the
 635 properties of the water column. For instance, Govers et al. (2014) demonstrated that other
 636 ecosystems, such as seagrass beds, are threatened by eutrophication resulting from
 637 emergency wastewater overflow and coastal housing near these bays. Also, recent studies
 638 have proven the presence of organic chemical pollution around bays in Curaçao as indicators
 639 of waste water and industrial pollution (De Jong et al., 2025). The SPM dynamics observed in
 640 this study provide important insight into the environmental conditions affecting the reefs
 641 surrounding Piscadera Bay, which have undergone significant benthic cover transformations
 642 in recent decades (Sandin et al., 2022; De Bakker et al., 2016; Van Duyl, 1985). In-situ studies
 643 conducted around Piscadera Bay have demonstrated that turf algae and microalgae are more
 644 efficient than corals in taking up nutrients episodically released by plumes from the bay during
 645 rainfall events, providing these algae with a competitive advantage over corals (Den Haan et
 646 al., 2016; Vermeij et al., 2010).

647 Our study also shows relatively high levels of elements such as Fe, Al, and Mn, which
 648 corroborates with the land-derived anthropogenic signatures found along the southwest
 649 coast of Curaçao (Van der Loosdrecht, submitted). Reefs close to land areas with increasing
 650 altered surface (i.e., the proportion of the onshore watershed that was occupied by human-
 651 modified surfaces ,including habitations, commercial developments, and roads) that can be
 652 associated with increased risk for urban runoff and pollution) showed higher loading of
 653 pollutants, and this correlation was enhanced with higher precepitation (Van de Loosdrecht,
 654 submitted). While XRF measurements are semi-quantitative and primarily reflect relative



655 abundances, they are useful to indicate terrestrial runoff. In elevated concentrations, this
 656 material can be harmful to corals and promote growth of cyanobacteria and turf algae (Kelly,
 657 2013; Esslemont, 1999). Turf algae, which can rapidly colonize newly available substratum,
 658 have become dominant on many coral reefs worldwide (Sandin et al., 2008b), further
 659 indicating ecosystem shifts in response to environmental changes. This pattern has also been
 660 seen over the past 40 years in the reefs of this study. Both the western and eastern reefs have
 661 experienced noticeable decrease in coral cover, now predominantly featuring turf algae and
 662 increased abundance of cyanobacterial mats, being more pronounced in the eastern reef,
 663 which is more directly subjected to bay-derived material (Sandin et al., 2022; De Bakker et al.,
 664 2016; Van Duyl, 1985).

665 The western reef faces different conditions, characterized by trace-metal values that suggest
 666 alternative environmental pressures. While the semi-quantitative XRF measurements of Ni,
 667 Cu, Pb cannot be definitively classified as toxic, their presence indicates enhanced
 668 anthropogenic-related matter fluxes (Fernandez et al., 2007; Caballero-Gallardo et al., 2020).
 669 For example, Ni can limit calcium available for growth, affects respiration, and can reduce coral
 670 larvae settlement, coral reproduction and negatively affect growth and reproduction of other
 671 key reef organisms such as anemones or sea urchins (Méndez et al., 2021; Reichelt-Brushett
 672 and Hudspeth, 2016; Reichelt-Brushett and Harrison, 2005; Nalley et al., 2021). Other
 673 elements such as Pb can also affect growth and photosynthesis of corals and macroalgae
 674 (Baumann et al., 2009). Our findings align with studies suggesting that chronic exposure to
 675 certain pollutants may contribute to long-term reef degradation (Jackson et al., 2014).

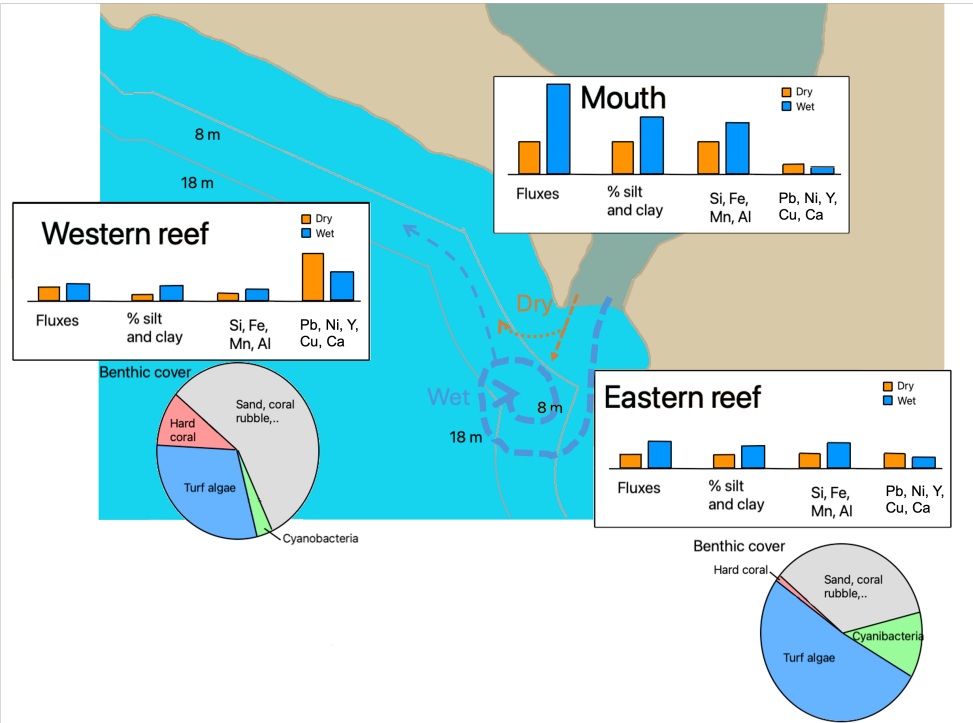
676 The rapid northwest surface current appears to play a crucial role in dispersing bay-derived
 677 materials away from nearshore areas, reducing residence time and potentially preventing the
 678 accumulation of high concentrations in localized zones. However, this same dispersal
 679 mechanism also facilitates the broader distribution of potential contaminants across the reef
 680 system. The contrasting hydrodynamic conditions between sites may explain observed
 681 differences in reef community composition, where the eastern reef—characterized by lower
 682 current velocities (Fig.2) and thus longer residence times—shows greater ecological
 683 degradation compared to the western reef (De Bakker et al., 2017; Sandin et al., 2022), which
 684 experiences higher current velocities (Fig.2) and more effective flushing by the main current



685 system. This differential exposure to terrestrial materials represents a critical factor in
686 understanding current reef composition and predicting future development patterns (Fig. 9).

687

688



689
690 **Figure 9.** Summary of the spatial and seasonal variability in SPM characteristics in reefs nearby
691 Piscadera Bay. (Graphical abstract)

692 **5. Conclusions**

693 This study highlights the spatial and seasonal dynamics of suspended particulate matter near
694 Piscadera Bay, one of over 35 bays and water inlets along Curaçao's south coast that are
695 subject to varying degrees of human impact. Seasonal differences show that dry-season SPM
696 dynamics are controlled by localized factors (tides, resuspension, currents), while wet-season
697 conditions are dominated by rainfall, wind, and enhanced hydrodynamics.

698 Coral reefs near bay mouths are particularly influenced by bay-derived inputs during wet-
699 season runoff, with strengthening west-to-east correlations highlighting the role of bay-ocean
700 exchange in transport processes. The eastern reef experiences prolonged exposure to



701 terrestrial material due to proximity and slower dispersal, potentially intensifying ecological
702 stress, while the western reef benefits from rapid material dispersal.

703 The threat to reefs stems from increased pollutant loads and persistent fine particles that
704 reduce light availability and transport harmful elements. In sheltered areas, slower
705 hydrodynamic exchange prolongs these effects, altering material quality and quantity
706 reaching reefs with important implications for reef health and resilience. Future research
707 should compare SPM characteristics across multiple bays and include ecotoxicological assays
708 to identify elements posing the greatest risk—essential for informing targeted mitigation
709 strategies and evidence-based coastal management.

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721 **Data**

722 The data supporting this study are openly available and can be accessed via the DOI:
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724 **Author Contributions**

725 Virginia Sánchez Barranco: conceptualization, sampling, laboratory analysis, statistical
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731 **Competing Interests**

732 The authors declare no competing interests.

733 **References**

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