

^{14}C -based separation of fossil and non-fossil CO_2 fluxes in cities using relaxed eddy accumulation: results from tall-tower measurements in Zurich, Paris, and Munich

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Abstract. Relaxed eddy accumulation (REA) measurements for $^{14}\text{CO}_2$ enable the estimation of fossil fuel (ff) CO_2 fluxes in urban areas. This work is based on 252 REA ff CO_2 flux measurements conducted on tall towers in the cities of Zurich, Paris, and Munich. The ff CO_2 fluxes were compared to net eddy covariance CO_2 fluxes to quantify the role of non-fossil (nf) CO_2 fluxes. ~~In all three cities, winter fluxes were predominantly fossil, with mean ffecontributions of about 80%. Summer fluxes could be most clearly partitioned in Munich, where~~ While the measurements in Zurich and Paris were limited by small signal-to-noise ratios, improvements in the REA setup, the $^{14}\text{CO}_2$ measurement precision, the sampling strategy, and the source strength increased the ~~signal-to-noise ratios compared to Zurich and Paris. In Munich, the observed nffluxes were predominantly positive (~50% of net summer fluxes), demonstrating the major role of respiration, biofuels, and certain industrial processes. Particularly large~~ significance of the results in Munich. Large nf CO_2 fluxes observed in Munich from the direction of a brewery ~~suggest non-respiratory anthropogenic contributions and highlight~~ demonstrate the efficacy of the partitioning approach and illustrate the complexity of urban ~~environments. Additionally, the absolute~~ atmospheric measurement data. Excluding these

~~measurements potentially influenced by large anthropogenic nfCO_2 and concentrations of the REA samples were compared to clean background concentrations to estimate ffCO_2 excess concentrations. Across all cities, ff/CO_2 contributions to regional excess concentrations were much lower ($< 65\%$ flux ratio in Munich was approximately 47% in summer and 76% in winter, with the majority of measurements taken between 07:00 and 19:00 local time. Regional excess concentrations had much lower ffCO_2 contributions ($< 63\%$ in winter and $< 30\%$ in summer) than to local eddy covariance fluxes $< 28\%$ in summer, in all three cities), demonstrating fundamental differences between local and regional CO_2 fluxes. The combination of $^{14}\text{CO}_2$ observations and the REA method is a sophisticated approach that challenges the limits of current analytical capabilities, while providing unique opportunities for quantifying ffCO_2 and nfCO_2 fluxes.~~

20 1 Introduction

Cities are hotspots for fossil fuel (ff) CO_2 emissions and are at the heart of emission reduction efforts. To guide and monitor the pathways of cities towards climate neutrality, measuring and modeling urban ffCO_2 emissions is essential. While total CO_2 fluxes can be measured using the eddy covariance (EC) method, direct observations of fossil or non-fossil CO_2 are lacking. However, a separation of the two components is important because, in addition to ffCO_2 emissions, biospheric and human respiration fluxes play a substantial role in the urban carbon budget (e.g. Kellett et al., 2013; Miller et al., 2020; Wu et al., 2022; Stagakis et al., 2025). Several studies have attempted to separate ffCO_2 and nfCO_2 fluxes. Wu et al. (2022) combined CO_2 fluxes from EC measurements and CO fluxes from flux-gradient measurements to estimate turbulent ffCO_2 fluxes on a tower in Indianapolis 30 m above ground level, assuming a constant CO/ffCO_2 flux ratio. The latter was determined from CO and $^{14}\text{CO}_2$ concentration measurements of flask samples collected weekly at the measurement site and an upwind background station, following Levin et al. (2003). Hilland et al. (2025) proposed a linear mixing model to separate biospheric, road traffic, and stationary combustion CO_2 fluxes using simultaneous tall-tower EC measurements of CO_2 and co-emitted species (CO and NO_x), as well as sector-specific, constant flux ratios determined from a bottom-up emission inventory. Other studies used $^{14}\text{CO}_2$ observations to separate fossil and non-fossil CO_2 enhancements relative to a background concentration (e.g., Levin et al., 2003; Turnbull et al., 2015; Miller et al., 2020). In this case, surface emissions can be estimated using atmospheric transport models or the Radon-Tracer-Method, for example (Levin et al., 2003; Maier et al., 2024b). The source area thereby depends on the choice of the background station and includes a large region beyond the city boundaries if a tropospheric or continental clean air background site is used (Turnbull et al., 2015). To our knowledge, all previous studies estimating urban ffCO_2 emissions relied on bottom-up information, inverse modeling results, or assumed constant proxy/ ffCO_2 ratios, despite the fact that ratios such as CO/ffCO_2 vary significantly with fuel carbon content and combustion conditions (Turnbull et al., 2015; Maier et al., 2024a).

We overcome these limitations using $^{14}\text{CO}_2$ relaxed eddy accumulation (REA) measurements, as first described in Kunz et al. (2025a). On a tall tower over the city, air is conditionally collected ~~for during~~ one hour in ~~an updraft or downdraft reservoir~~ two separate reservoirs (an updraft and a downdraft reservoir) using fast-switching sampling valves ~~that~~. The valves respond to a 20 Hz vertical wind signal from a 3D ultrasonic anemometer. Transfer of the collected air to portable glass flasks

45 enables $^{14}\text{CO}_2$ and CO_2 measurements in a subsequent laboratory analysis, and thus the estimation of ffCO_2 concentration differences between updraft and downdraft samples. Combined with net CO_2 fluxes measured by open-path or closed-path EC, this novel approach enables the estimation of ffCO_2 fluxes for the respective, hour-long sampling periods.

In Kunz et al. (2025a), the REA flask sampling system was described and its performance was analyzed in detail. It was shown to meet high technical requirements, e.g., fast and accurate switching between updraft and downdraft sampling, while
50 maintaining a constant flow rate in sampling and non-sampling modes. For the estimation of ffCO_2 fluxes, uncertainties due to the sampling procedure were negligible compared to the analytical $^{14}\text{CO}_2$ uncertainty in the lab. Analysis of concentration differences between updraft and downdraft flask samples collected during a pilot application at a tall tower in Zurich, Switzerland, showed that separation of fossil and non-fossil components of the CO_2 concentration differences is feasible, but often limited by a low signal-to-noise ratio of the $^{14}\text{CO}_2$ difference. Since then, the REA system has been further improved and
55 operated on two tall towers in Paris, France, and Munich, Germany, for another 9 months each.

This paper presents and analyzes the ffCO_2 fluxes obtained from a total of 252 discrete hour-long $^{14}\text{CO}_2$ REA measurements conducted on three tall EC towers in Zurich, Paris, and Munich. After a brief presentation of the methods (Sect. 2) and the measurement campaigns (Sect. 3), the following questions are addressed:

1.

60 Q1) To what extent do $^{14}\text{CO}_2$ REA measurements enable the separation of local fossil and non-fossil CO_2 fluxes in an urban area? (Sect. 4.1, 4.2, 4.3)

2.

Q2) (a) What are typical ffCO_2 and nfCO_2 flux contributions? (Sect. 4.3)

(b) Do we find indications for localized fossil and non-fossil CO_2 sinks and sources, and/or observe systematic spatial
65 and temporal differences within and between the three cities? (Sect. 4.3, 4.4)

3.

Q3) How does the composition of surface fluxes in the vicinity of the tall tower compare to the composition of regional CO_2 concentration enhancements? (Sect. 4.5)

2 Methods

70 This study analyzes the contributions of fossil and non-fossil sinks and sources to net CO_2 fluxes measured successively on three different urban tall towers for about nine months each. While the net CO_2 fluxes were measured continuously by the well-established EC method (e.g., Aubinet et al., 2012b), the partitioning of individual, hour-long measurements is based on REA measurements for $^{14}\text{CO}_2$ (Kunz et al., 2025a). [Figure 1 provides an overview of the individual steps involved in the REA measurements and analysis.](#)

75 In addition to CO₂, the flask samples were also analyzed for CO, CH₄, N₂O, SF₆, H₂, δ(O₂/N₂), δ¹⁸O, and δ¹³C. Moreover, the MGA⁷ provided continuous flux measurements of CO, CH₄, NO, and NO₂. These measurements are of great value, e.g., for a future analysis of proxy/ffCO₂ flux ratios needed for estimating continuous ffCO₂ fluxes based on proxy measurements. However, a multi-species analysis is beyond the scope of this work.

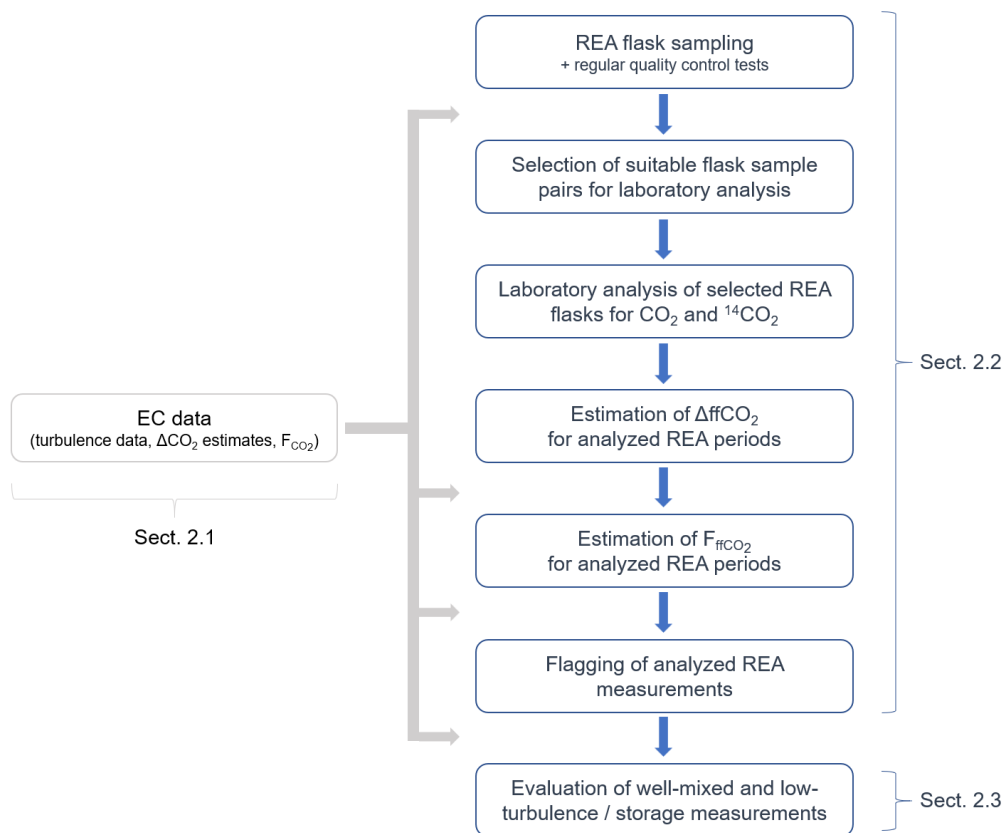


Figure 1. Overview of the REA measurement and analysis procedure. ΔCO_2 and ΔffCO_2 denote the concentration differences between updraft and downdraft REA flask samples. F_{ffCO_2} is the ¹⁴C-based ffCO₂ flux. “Well-mixed” and “low-turbulence/storage” measurements are two categories, in which the analyzed REA measurements considered in this study were divided based on several flagging criteria.

2.1 Net CO₂ fluxes from eddy covariance measurements

80 Net turbulent CO₂ fluxes were computed from high-frequency CO₂ measurements of a closed-path (MGA⁷, MIRO Analytical AG, Wallisellen, Switzerland) and an open-path gas analyzer with a co-located 3D sonic anemometer (IRGASON, Campbell Scientific, Inc., Logan, UT, USA). The To remove erroneous spikes caused by instrument malfunction or obstructions in the path of the IRGASON gas analyzer (e.g., animals, dirt, rain, snow), the 20 Hz CO₂, H₂O, and vertical wind measurements of the IRGASON were despiked by discarding using a modification of the Median Absolute Deviation (MAD) method described by Mauder et al. (2013). To this end, measurements where the median absolute deviation of less than three

85

~~consecutive observations~~ was outside the upper and lower limits defined by Mauder et al. (2013). ~~The were removed; however,~~ observations in which three or more consecutive outliers occurred were kept. The latter was necessary to retain peaks in concentrations caused by the intermittent nature of emission signals in the urban environment, which flask measurements have exemplarily proven to be real. The 10 Hz measurements of the MGA⁷ were upsampled to 20 Hz ~~and using a nearest-neighbor~~ approach with a search window of 50 ms. ~~The upsampled MGA⁷ data was then~~ synchronized with the IRGASON data ~~directly based on the high-frequency CO₂ time series of the two instruments~~ by finding the time lag of maximum correlation ~~between the high-frequency time series. For periods with low IRGASON signal strength and consequently, as in Hilland et al. (2025),~~ Erroneous time lags for periods with poor correlation between the CO₂ time series (correlation coefficient < 0.5), ~~the e.g., due to low IRGASON signal strength during a rain event, were linearly interpolated.~~ The median time lag was ~~determined via linear~~ interpolation ~~to correct clock drift and subsequent maximization of the covariance between and vertical wind velocity using a search window of ±0.5–4.15 s (Hilland et al., 2025), in Zurich, 10.45 s in Paris, and 37.30 s in Munich. See Appendix D for details.~~ The fluxes were then computed using the software EddyPro (Version 7.0.9, Licor Inc., Lincoln, NE, USA) with a 30 min averaging period, coordinate rotation via double rotation (Wilczak et al., 2001), ~~time lag compensation through covariance maximization,~~ and detrending via block average (Rebmann et al., 2012). High-pass filtering effects were corrected according to ~~Moncrieff et al. (1997) Moncrieff et al. (2004).~~ For low-pass filtering effects, the correction by ~~Moncrieff et al. (2004) Moncrieff et al. (1997)~~ was used for the IRGASON and the correction by Fratini et al. (2012) for the MGA⁷. Random errors of the turbulent flux estimates were calculated after Finkelstein and Sims (2001), and storage fluxes were estimated from concentrations and based on a single-point profile. Quality control flags of 0 (high quality), 1 (intermediate quality) or 2 (poor quality) were assigned to all flux estimates according to Mauder and Foken (2004), checking the assumptions of stationarity and well-developed turbulence. In addition, EddyPro outputs a large set of variables for each 30 min averaging period, including friction velocity u_* , standard deviation of vertical wind velocity σ_w , and molar volume of ambient air v_a . Details on the EddyPro outputs in general and the processing of the IRGASON and MGA⁷ data in particular can be found in LI-COR (2021) and Hilland et al. (2025), respectively.

To estimate the mean CO₂ fluxes during the specific, typically 60 min long REA flask sampling periods (Sect. 2.2), the 30 min EC fluxes were averaged, weighted by the fraction of the EC averaging period during which REA samples were collected. This means that each 60 min flux includes two to three 30 min fluxes (usually two, since most REA measurements were scheduled at the hour). The uncertainty of the 60 min flux was estimated by error propagation of the respective 30 min random uncertainty estimates. ~~Any additional uncertainties arising from the measurement instrument or data processing options used were not considered.~~ For quality control purposes, the maximum of the 30 min quality control flags, denoted QC in the following, was considered. Since the CO₂ concentration measurements of the MGA⁷ showed a better agreement with the ~~measured flask concentrations flask concentration differences measured between updraft and downdraft samples~~ than the IRGASON measurements (~~Sect. 4.1), the fluxes Appendix E), and since the spectral-corrected fluxes of the two instruments showed very good agreement (Appendix D), the fluxes~~ calculated from the MGA⁷ measurements were used when available, otherwise the fluxes calculated from the IRGASON were used. Information on which EC data set was used is provided for each REA measurement ~~at in Kunz et al. (2025b).~~

2.2 ¹⁴C-based separation of fossil and non-fossil CO₂ fluxes from relaxed eddy accumulation measurements

2.2.1 REA sampling and flux calculation

Fossil and non-fossil components of the CO₂ flux measurements were separated by ¹⁴CO₂ analysis of flask sample pairs conditionally collected using the REA flask sampling system described in detail in Kunz et al. (2025a). In summary, depending on the 20 Hz vertical wind measurements of the IRGASON's 3D ultrasonic anemometer (Sect. 2.1), air was collected through two co-located inlets with two fast-response valves into two separate reservoirs: one for updrafts, and one for downdrafts. After a sampling period of, e.g., 60 min, it was checked whether sufficient air has accumulated for a subsequent CO₂ and ¹⁴CO₂ analysis in the laboratory. If so, the accumulated air was transferred by an extended automated 24-port flask sampler into two 3 l glass flasks that could be analyzed in the laboratory (denoted as “successful” REA measurement in the following). Updraft and downdraft were thereby defined with respect to the mean vertical wind velocity \bar{w} , excluding a range of wind speeds centered around \bar{w} and scaled by the standard deviation of the vertical wind σ_w (scaling factor δ). This so-called deadband with half-width $\delta \cdot \sigma_w$ was intended to increase the concentration difference and to reduce the number of valve switchings (Rinne et al., 2021). \bar{w} and σ_w were either calculated from the 30 min period before sampling start (pre-set deadband) or dynamically adjusted using a 15 min backward-looking averaging interval (dynamic deadband). The latter lead to a more equally distributed sampling of updrafts and downdrafts and was therefore better suited for changes in vertical wind statistics during the sampling period. The calculated fluxes are independent of the method used to compute the deadband, since this is taken into account in the β coefficient (see below) (Pattey et al., 1993). See Kunz et al. (2025a) for technical details on the REA sampling.

Due to the costs and logistics associated with flask sample analysis, only a limited number of successful REA measurements could be analyzed. The selected REA flask samples were analyzed for CO₂ and in the ICOS (Integrated Carbon Observation System) Flask and Calibration Laboratory in Jena, Germany, and for ¹⁴CO₂ in the ICOS Central Radiocarbon Laboratory in Heidelberg, Germany. Based on these measurements, the ffCO₂ differences between updraft and downdraft samples, in the following denoted as ΔffCO_2 , were estimated (Appendix A1, Kunz et al. (2025a)). Under stationary and well-developed turbulence, the ffCO₂ flux F_{ffCO_2} can then be estimated according to Eq. (1):

$$F_{\text{ffCO}_2} = \beta \sigma_w \bar{\rho}_m \Delta\text{ffCO}_2. \quad (1)$$

$\bar{\rho}_m$ is the mean molar air density in mol m⁻³. ~~For $\delta = 0$, the~~ The proportionality factor β ~~is 0.627 if the vertical wind velocity w is normally distributed and the regression of the scalar concentration on w is linear (Wyngaard, 1992; Baker et al., 1992). For individual sampling periods, however, β can vary significantly, depending on~~ depends on the joint probability distribution of variations of the vertical wind velocity and the gas concentration (Milne et al., 1999; Fotiadi et al., 2005b; Ruppert et al., 2006; Grönholm et al., 2008). ~~In addition, β decreases with increasing deadband width (e.g., Pattey et al., 1993; Fotiadi et al., 2005b).~~ and on the deadband width (e.g., Pattey et al., 1993; Milne et al., 1999; Fotiadi et al., 2005b). Values < 0.1 or > 1 indicate non-ideal sampling conditions for REA measurements (Grönholm et al., 2008; Hensen et al., 2009; Osterwalder et al., 2016). Due to the availability of co-located ~~high-frequency~~ EC measurements of net CO₂ fluxes (Sect. 2.1), ~~the~~ these measured CO₂ fluxes were used to

calculate β for each sampling period individually:

$$155 \quad \beta = \frac{F_{\text{CO}_2}}{\sigma_w \bar{\rho}_m \Delta \text{CO}_2}. \quad (2)$$

ΔCO_2 is the CO_2 concentration difference between updraft and downdraft flask samples measured in the laboratory and F_{CO_2} is the net CO_2 flux measured by EC (Sect. 2.1). Assuming scalar similarity between CO_2 and $^{14}\text{CO}_2$, Eq. (2) can be inserted into Eq. (1):

$$F_{\text{ffCO}_2} = \frac{F_{\text{CO}_2}}{\Delta \text{CO}_2} \cdot \Delta \text{ffCO}_2 = \frac{\Delta \text{ffCO}_2}{\Delta \text{CO}_2} \cdot F_{\text{CO}_2}. \quad (3)$$

160 Accordingly, the fossil contribution to the net CO_2 flux equals the $\Delta \text{ffCO}_2 / \Delta \text{CO}_2$ ratio of the REA flask samples. The uncertainty of the ffCO_2 flux was derived according to Gauss' law of error propagation from Eq. (3). For ΔCO_2 and ΔffCO_2 , only the measurement uncertainties from the laboratory analysis were considered, as uncertainties due to the sampling process, e.g., a time lag between a change in vertical wind and a switching of the fast-response sampling valves, are negligible compared to the $^{14}\text{CO}_2$ measurement uncertainty (Kunz et al., 2025a). The uncertainty of F_{CO_2} was estimated using the random
165 uncertainty estimate from EddyPro (Sect. 2.1). Additional uncertainties, e.g., due differences between IRGASON and MGA⁷ measurements, differences between different EC data processing options, due to the assumption of scalar similarity or due to turbulent sampling error in the REA flask concentration differences, were considered less relevant and not taken into account.

It is important to note that Eq. (3) describes the turbulent fluxes at the measurement height. These fluxes only represent the surface fluxes if changes in the storage below the measurement height are negligible and there is no mean vertical advection.
170 While this is usually the case during well-mixed, convective conditions (i.e., in the afternoon), significant storage fluxes can occur, particularly in the morning hours during the transition from low-turbulence, nighttime conditions to well-developed turbulence, when the depth of the atmospheric boundary layer increases and built-up CO_2 is vented from the layer below the measurement height (e.g., Stull, 1988; Crawford and Christen, 2014). A storage correction, as it is recommended and commonly applied in EC measurements (e.g., Aubinet et al., 2012b; Crawford and Christen, 2014), would require knowledge of both the
175 storage flux $F_{\text{CO}_2, \text{strg}}$ and the $\text{ffCO}_2 / \text{CO}_2$ ratio of the storage fluxes. However, the magnitude of the storage flux in cities, especially in the morning, is associated with significant uncertainties (Crawford and Christen, 2014). The ffCO_2 contribution to the storage fluxes equals the ratio of the flux averages over the period during which CO_2 accumulated below the measurement height, ~~which does~~. For negative storage fluxes, i.e., the venting of CO_2 which accumulated prior to the measurement period, the $\text{ffCO}_2 / \text{CO}_2$ ratio will therefore not necessarily equal the surface flux ratio during the measurement period. Consequently,
180 a meaningful, observation-based storage flux correction for the REA measurements is not feasible. Thus, the presented fluxes are not corrected for changes in storage. While REA measurements during or after low-turbulence conditions therefore do not reflect the surface fluxes during the sampling period, the measured $\text{ffCO}_2 / \text{CO}_2$ ratio still provides information about the average relative contribution of fossil fuel emissions in the time period since the layer below the measurement height became decoupled prior to the start of the REA measurement – usually a nocturnal accumulation under low-wind conditions. Therefore,
185 measurements with low turbulence and/or storage fluxes are analyzed separately. The criterion used in this study to flag the corresponding measurements is described in Sect. 2.2.4.

2.2.2 REA system improvements

As the $^{14}\text{CO}_2$ differences between updraft and downdraft samples collected in Zurich and Paris were often close to or smaller than the detection limit in the laboratory analysis, the REA system was modified, as suggested in Kunz et al. (2025a). To enable the use of a larger deadband width, larger pumps were installed in the REA system before the campaign in Munich. This was necessary because a larger deadband width reduces the proportion of time during which air is collected and therefore increases the required sampling flow rate needed to collect enough air for laboratory analysis. In addition, the option for hyperbolic relaxed eddy accumulation (HREA, Bowling et al. (1999)) was added. In HREA, air is only collected if both vertical wind velocity fluctuations $w' = w - \bar{w}$ and fluctuations in the scalar concentration $c' = c - \bar{c}$ are above a certain threshold, which is characterized by the hole size H (similar to δ and a pre-set or dynamic deadband in normal REA). This maximizes the concentration differences between updraft and downdraft reservoirs, as only the eddies that contribute the most to the vertical flux are sampled, and is recommended for REA applications where sampling differences are close to the detection limit (Vogl et al., 2021).

2.2.3 Quality control of the REA system

To ensure high quality measurement data, the performance of the REA flask sampling system was tested regularly (for details, see Kunz et al. (2025a)). To examine biases between updraft and downdraft sampling, a pair of quality control flasks was sampled about once a month by continuously collecting air through both updraft and downdraft lines without switching the valves. Simultaneously, a third flask was sampled through a separate line directly into the flask sampler, bypassing the reservoirs where updrafts and downdrafts accumulate. If the system was operating as intended, the concentrations of the three quality control samples should agree within the WMO compatibility goal of 0.1 ppm for CO_2 (WMO recommendation for compatibility of measurements of greenhouse gases and related tracers (Tans and Zellweger, 2014)).

To verify the correct switching between updraft sampling, downdraft sampling, and no sampling, the measured CO_2 concentration differences between the updraft and downdraft REA flask pairs were compared to the CO_2 in situ measurements of the IRGASON and the MGA⁷. For this purpose, the high-frequency gas densities were converted to dry molar fractions and averaged over the respective actual sampling times, as described in Kunz et al. (2025a).

To detect technical problems as early as possible, automated leak and critical component tests were carried out daily in the Paris and Munich campaigns. The results of the quality control flask measurements are given in Appendix E.

2.2.4 Flagging of analyzed REA measurements

Besides technical requirements, REA is like any other turbulent flux measurement technique restricted to certain meteorological conditions, e.g., stationarity and well-developed turbulence (Rinne et al., 2021). Moreover, and in contrast to the EC technique, REA measurements cannot be processed retrospectively, e.g., cannot be corrected for changes in the mean vertical wind velocity. Therefore, additional criteria are necessary (Fotiadi et al., 2005a). Several criteria have already been considered in the selection of suitable flask samples during the campaigns (Kunz et al., 2025a). However, due to the lim-

ited number of good sampling conditions in the urban environment, the refinement of EC processing options, and the addition of further criteria after the measurement campaign, the analyzed sampling periods were not always ideal for REA measurements. For analysis of the results, the measurements were characterized based on five flagging criteria (Table ??1, see Appendix B for details). ~~The assumptions of stationarity and well-developed turbulence were validated based on the maximum of the 30 EC quality control flags according to Mauder and Foken (2004) (QC). Following Hensen et al. (2009) and Osterwalder et al. (2016), measurements for which Eq. (3) does not provide reasonable fluxes were flagged according to β (Eq. 2). Measurements with large uncertainties due to the limited resolution of the differences between updraft and downdraft samples were flagged based on the signal-to-noise ratio SNR, defined as the minimum of the relative F_{ffCO_2} and F_{ffCO_2} uncertainties. Measurements during which a decoupling between the measurement height and the surface was likely were flagged using the minimum 30 friction velocity, u_{*s} , and the maximum of the absolute 30 storage flux, $|F_{CO_2, strg}|$.~~ Based on these criteria, the analyzed REA measurements were classified into ~~four categories (Table ??)~~three categories (Fig. 2). “Well-mixed ~~”measurements~~measurements” are assumed to best represent the surface fluxes during the sampling period. These measurements are the most valuable for answering our research questions and were analyzed in the most detail. In contrast, “low-turbulence ~~”and “storage” measurements~~and storage measurements” are probably not representative of the surface fluxes during the sampling period due to insufficient turbulence or changes in storage below the measurement height (Sect. 2.2.1). However, the relative $ffCO_2$ contributions were investigated to characterize the integrated fluxes before and during the sampling period. ~~Note that the low-turbulence and storage flags are not mutually exclusive, but are closely related. Therefore, measurements with either flag were examined together.~~ Measurements with $QC = 2$, $\beta < 0.1$, $\beta > 1$ or $SNR < 100\%$ were not considered further in this study.

Table 1. Flagging of the criteria for analyzed REA measurements based on the maximum quality control flag. See Appendix B for stationarity and well-developed turbulence, (QC), the β coefficient, the signal-to-noise ratio (SNR), the minimum friction velocity (u_*), and the maximum absolute storage flux ($|F_{CO_2, strg}|$) details. Measurements were not considered further if any of the criteria was met (“/”); the other flags were only assigned if all criteria were met (“&”).

Flag Variable	Variable description
QC	The maximum of the 30 min EC quality control was used to validate the assumptions of stationarity
β	SNR $u_* F_{CO_2, strg} $ The β coefficient as measurement periods with unfavorable microm (Hensen et al., 2009; Osterwalder et al., 2016).
SNR	-% The signal-to-noise ratio (SNR) was defined uncertainties to flag measurements with large uncertainties differences between updraft and downdraft sampl
Well-mixed u_*	0, 1 & 0.1, 1 & ≥ 100 & ≥ 0.2 & ≤ 20 The minimum to identify REA measurement periods with low tu
Low turbulence $ F_{CO_2, strg} $	0, 1 & 0.1, 1 & ≥ 100 & < 0.2 The maximum of the used to identify REA measurement periods with c
Storage 0, 1 & 0.1, 1 & ≥ 100 & > 20 Not considered 2 / < 0.1 / > 1 / < 100 height	

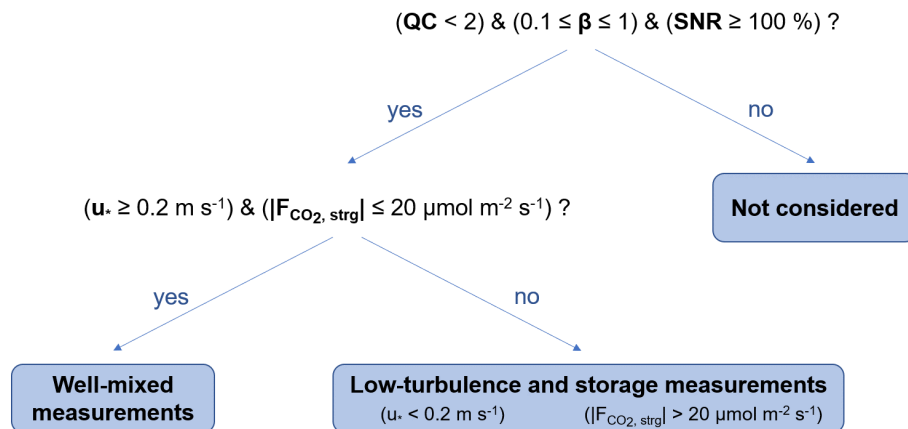


Figure 2. Illustration of the flagging of the analyzed REA measurements. Each measurement was flagged as either well-mixed measurement, low-turbulence and storage measurement, or was not considered further. The flagging criteria QC , β , SNR , u_* , and $|F_{CO_2, strg}|$ are explained in Table 1.

2.3 Analysis of well-mixed REA measurements

2.3.1 Flux footprints and mean land cover fractions

240 To analyze the flux source areas during the individual REA measurements, flux footprints were derived for each 30 min averaging interval using the flux footprint model of Kljun et al. (2015). Inputs were turbulence data from eddy covariance measurements (ICOS Ecosystem Thematic Centre et al., 2025a, b, c; Ecosystem Thematic Centre, 2025), boundary layer height ERA5 reanalysis estimates from the Copernicus Climate Change Service (Hersbach et al., 2024), measurement heights and tower coordinates, and roughness length and displacement height derived from building and vegetation height maps. Details on the flux footprints are provided in Molinier and Kljun (2024).

245 For the 30 min time periods during which well-mixed REA measurements were taken, aggregated footprints were calculated. These aggregated footprints were combined with the WorldCover product provided by the European Space Agency (<https://esa-worldcover.org>, last access: 12 December 2025) to derive comparable mean land cover fractions using the same data source for each city. The flux footprints were also used to identify measurements which were potentially influenced by emissions from a district heating plant in Zurich and a brewery in Munich by calculating the expected flux contributions from the corresponding areas of interest (Appendix F).

2.3.2 Determination of mean $ffCO_2/CO_2$ flux ratios and evaluation of the significance of average $nfCO_2$ flux components

255 To generalize and quantify the results from the individual REA measurements, the mean $ffCO_2/CO_2$ flux ratios \bar{R}_{ffCO_2} and the mean magnitude of the $nfCO_2$ fluxes \bar{F}_{nfCO_2} were determined for each city. Due to the small number of measurements,

it was not possible to fully account for the spatial and temporal variability. We distinguished between summer and winter measurements, and excluded measurements which were potentially influenced by large point-source emissions. In this work, “summer” refers to the period from July to October, and “winter” to the period from November to April (inclusive). This seasonal division of the measurement campaigns aligns roughly with the shift between European summer and winter time and with the change in local emissions due to heating degree days, and is consistent with other studies conducted in the same location during the same period (Hilland et al., 2025). If ffCO_2 and CO_2 fluxes were perfectly linearly correlated, the mean $\text{ffCO}_2/\text{CO}_2$ ratios would be best described by the slope of an error-weighted total least squares regression line (Maier et al., 2024a). Due to generally low correlations of the observed REA fluxes, however, \bar{R}_{ffCO_2} was determined as the error-weighted mean of the individual $\text{ffCO}_2/\text{CO}_2$ ratios. To minimize the uncertainty, the individual R_{ffCO_2} values were calculated directly from the flask measurements as $\Delta\text{ffCO}_2/\Delta\text{CO}_2$, i.e., completely independent of the EC flux measurements (compare Eq. 3). $R_{\text{ffCO}_2} > 100\%$ indicates a negative nfCO_2 flux, i.e., photosynthetic uptake, while $R_{\text{ffCO}_2} < 0\%$ is physically unreasonable and only observed if ΔffCO_2 is slightly negative within its measurements uncertainties. In addition, the mean and variability of the nfCO_2 fluxes were examined. A z-test was used to evaluate whether the observations were significantly different from $\bar{R}_{\text{ffCO}_2} = 100\%$ or $F_{\text{nfCO}_2} = 0 \mu\text{mol m}^{-2} \text{s}^{-1}$ (significance level of 0.05), i.e., completely fossil CO_2 fluxes, taking into account the mean measurement uncertainties. To meet the assumption of normal distribution, only measurements with relative ΔCO_2 uncertainties $\ll 1$ were considered (most, but not all, of these samples were already excluded by the consideration of the signal-to-noise ratio as defined in Sect. 2.2.4). See Appendix H for details.

2.3.3 Analysis of regional CO_2 concentration enhancements

While the REA flask measurements aimed to analyze turbulent ffCO_2 fluxes at the urban neighborhood scale, the absolute flask concentrations also contain information about the fossil and non-fossil CO_2 enhancements compared to clean background air and thus about the composition of CO_2 fluxes in a broader continental region, including other urban areas and regional emission sources. Following Levin et al. (2003), we calculated the ffCO_2 excess from the mean CO_2 and $\Delta^{14}\text{C}$ values of the up- and downdraft REA sample pairs, using the corresponding concentration measurements at the European marine background station Mace Head on the western coast of Ireland as background concentrations, and assuming that the biogenic $\Delta^{14}\text{C}$ signature equals the background concentration (see Appendix A2). Second-order effects, such as ^{14}C -enriched heterotrophic respiration and nuclear contamination (Maier et al., 2023), were not considered because the necessary concentration footprints were only available until the end of 2023, and the corrections are negligible for our analysis. For details and an evaluation of these corrections on the Zurich measurements, we refer to Maier et al. (2023) and Appendix A2. The mean $\text{ffCO}_2/\text{CO}_2$ ratios of the excess concentrations thus represent the average contributions of ffCO_2 emissions to the CO_2 fluxes on the trajectories between Mace Head and the three measurement sites.

3 Measurement campaigns

To assess the performance and to analyze the results of REA ^{14}C measurements for different urban environments, the REA system as well as the EC systems (IRGASON and MGA⁷) were successively installed and operated for nine months each on three tall towers in the cities of Zurich, Paris, and Munich. The measurements were conducted as part of the ICOS Cities project (https://www.icos-cp.eu/projects/icos-cities, last access: 01 February 2026), at the same time and place as the studies by Lan et al. (2024), Stagakis et al. (2025), and Hilland et al. (2025). At each site, the gas inlets for updraft and downdraft sampling and the inlet for the MGA⁷ measurements were mounted on a mast on top of a high-rise building or tower about 20 cm apart from the ultrasonic anemometer and the open-path CO₂ sensor of the IRGASON (Appendix C). The data logger, flask sampler, and the MGA⁷ were located in a climate controlled room. ~~Time lags due to the travel time of the sampled air from the inlets to the~~ The intake line of the MGA⁷ and the flask sampler were taken into account by synchronizing the time series of the MGA⁷ with the IRGASON (Sect. 2.1) and a site-specific rinse time, respectively (Kunz et al., 2025a). was set up in the same way as the flask sampling lines. REA samples were typically collected over 60 min, starting every other hour. Since increased stability at night is unfavorable for REA measurements (Fotiadi et al., 2005a), flasks were sampled during the day only. To ensure reliable measurements from the open-path gas analyzer, samples collected during periods of low signal strength, i.e., rain events, were discarded. With growing experimental experience, the logger program, REA system, and selection criteria were progressively updated, while the overall methodology remained consistent across the three cities. A documentation and version history of the logger program is publicly available at https://doi.org/10.5281/zenodo.13926681. Despite non-idealities in the EC measurement setups, e.g., unfavorably long intake lines for the MGA⁷, spectral analysis and comparisons between the IRGASON and MGA⁷ data showed good quality of the EC flux measurements (Appendix D). In addition, the regular quality control tests of the REA system demonstrated an overall good performance of the REA hardware (Appendix E). Figure 3 shows the locations of the three measurement sites, along with the 10 - 80 % source areas for the well-mixed REA measurements. ~~The flux footprints were derived at a 30 resolution using the flux footprint model of Kljun et al. (2015). Table 2 provides an~~ An overview of the site-specific data is provided in Table 2. For better readability, we refer to the three sites by their respective city names.

Table 2. Site-specific data from the three REA measurement campaigns in Zurich, Paris, and Munich. δ and H are the scaling factors for the deadband width (REA) and the hole size (HREA), respectively. “Pre-set” and “dynamic” indicate whether the latter was fixed at the beginning of the sampling period or continuously adjusted based on the standard deviation of the vertical wind velocity. [The numbers of successful samples and the numbers of successful samples selected and analyzed for \$^{14}\text{C}\$ are listed here.](#) “Successful” measurements refer to measurements in which enough sample air for laboratory analysis was collected in both the updraft and downdraft reservoirs. [Finally, Land cover fractions within the numbers aggregated footprints of successful samples selected and analyzed for the well-mixed REA measurements are given for the three main land cover types.](#)

	Zurich	Paris	Munich
ICOS Station ID	CH - Har	FR - Rmv	DE - Opd
Location	47° 22' 52" N 8° 30' 26" E	48° 53' 7.6" N 2° 25' 20.8" E	48° 8' 50.9" N 11° 32' 59.3" E
Measurement height [m a.g.l.]	112	103	85
Measurement period	July 2022 - April 2023	July 2023 - April 2024	July 2024 - April 2025
Wind directions with flow distortion [°N]	70 - 100	70 - 120	340 - 20
Length of REA intake lines REA and MGA⁷ intake lines:			
Length [m]	33 ± 2	27 ± 2	100
Inner diameter of REA intake lines Inner diameter [mm]	5.7	9.5	8
Deadband settings:			
Regular	Dynamic, $\delta = 0.7$	Dynamic, $\delta = 0.7, 0.9$	Dynamic, $\delta = 0.7, 0.9$
Test	Pre-set, $\delta = 0.3, 0.4, 0.7, 0.8$	Dynamic, $H = 0.6$	Dynamic, $H = 0.6$
Number of REA measurements:			
Started	709	498	601
Successful	338 (48 %)	384 (77 %)	485 (81 %)
With ^{14}C and EC data	87	65	99
Mean land cover fractions [%]:			
Built-up	64	83	77
Vegetation	33	17	23
Water	4	0.3	0.3

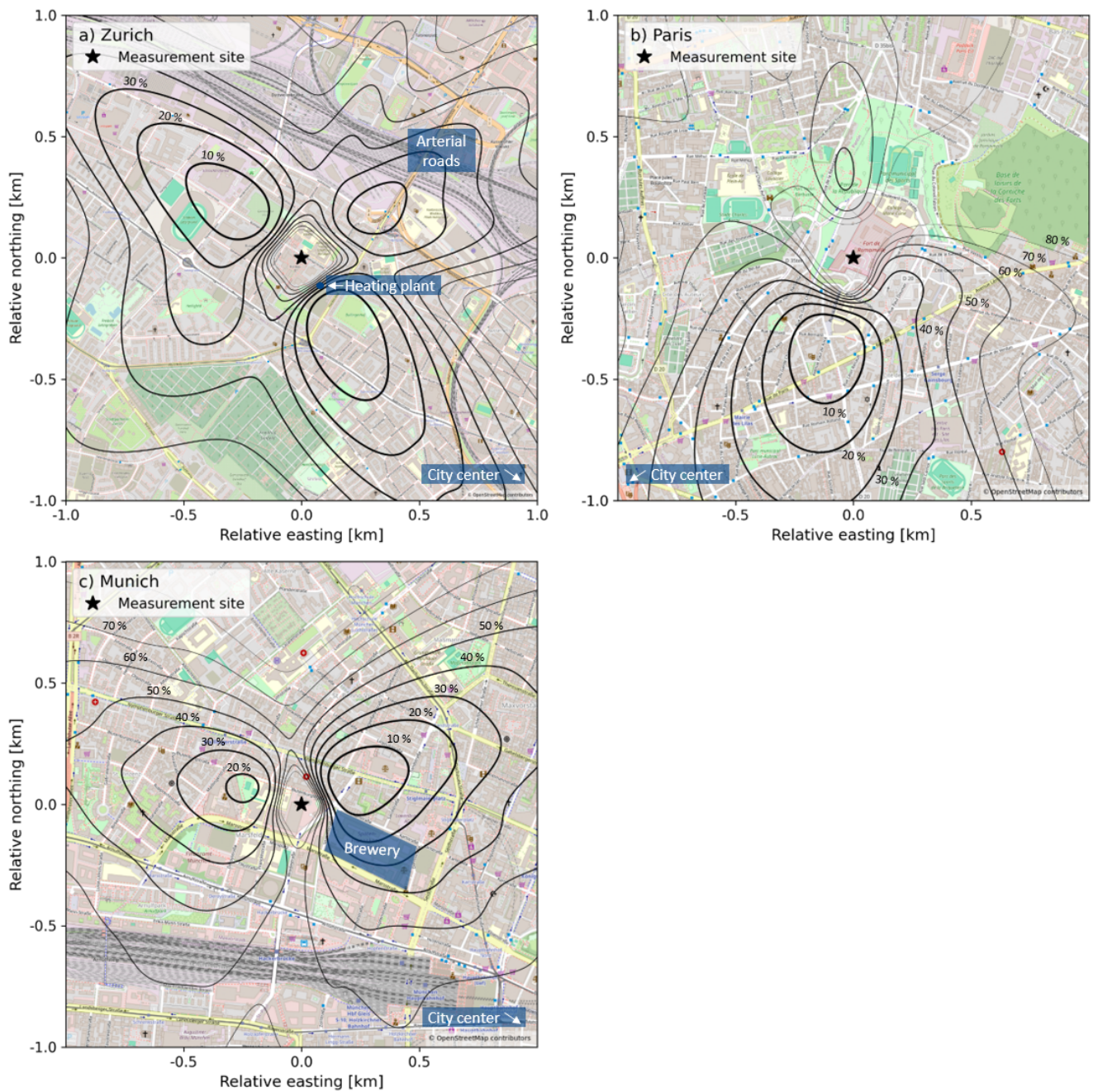


Figure 3. Locations of the measurement sites in Zurich, Paris, and Munich, and aggregated flux footprints of the well-mixed REA measurements according to Kljun et al. (2015) (black contour lines). The depicted areas contributed an average of 10 - 80 % to the fluxes observed during REA measurements under well-mixed conditions. Map data from © OpenStreetMap contributors 2025. Distributed under the Open Data Commons Open Database License (ODbL) v1.0.

310 3.1 Zurich - Hardau

In Zurich, the REA and EC measurements were conducted on an antenna of 16.5 m height on top of a 95.3 m high-rise building, i.e., approximately 112 m above ground level at the site Zurich - Hardau (ICOS Station ID 'CH-Har', Table 2, Fig. C1). The building, called Hardau II, is located roughly 1.5 km northwest of the city center of Zurich, Switzerland (Fig. 3 a). It is surrounded by three similar buildings of lower height (66 m, 76 m, and 85 m). Apart from that, the average building height
315 within a 1.5 km radius is 13.3 ± 8 m. Located to the north are an industrial sector, railway lines, and busy arterial roads, to the west is a residential, green area with a cemetery, and to the southeast is an urban sector and the city center. Within the average flux footprint of the well-mixed REA measurements, about 64 % of the surface area is covered by built-up areas, 33 % by vegetation, and 4 % by water (Lake Zurich). The largest point source in the immediate vicinity, located 145 m southeast, is a district heating plant that uses natural gas.

320 During the first REA measurements in July 2022, different deadband settings ($\delta = 0.3, 0.4, 0.7,$ and 0.8 with pre-set deadband) and averaging times (45, 60, 75 min) were tested (Table 2). With the pre-set deadband, in about 75 % of the REA measurements at least one of the reservoirs did not collect sufficient air to fill a flask. Therefore, a dynamic deadband with $\delta = 0.7$ was implemented and has been used since the end of August 2022. This was better suited for variable wind conditions and increased the percentage of successful measurements to 75 %. Unfortunately, all samples collected between November
325 2022 and February 2023 had to be discarded due to a leak in the REA sampler, which was detected retrospectively. More details on the Zurich measurements are given in Kunz et al. (2025a).

3.2 Paris - Romainville

In Paris, the REA and EC systems were installed on an active telecommunications tower about 5 km northeast from the city center at the site Paris-Romainville (ICOS Station ID 'FR-Rmv', Table 2). The IRGASON and the gas inlets were mounted
330 on a pylon, approximately 9 m above a wide (~ 30 m) platform (Fig. C1). Due to the massive structure of the tower, flow distortion effects were observed between 70° N and 120° N. The tower is located on a small hill in a densely urbanized area (Fig. 3 b). The average flux footprint of the well-mixed measurements was clearly dominated by built-up areas, with only 17 % of the area being vegetated.

Between July 2023 and April 2024, 66 of 384 successful and 498 scheduled REA measurements were analyzed in the
335 laboratory. One sample was rejected due to abnormal ^{12}C currents during ^{14}C analysis at the accelerator mass spectrometer (AMS), as well as implausible measurement results, leaving 65 REA measurements with ^{14}C and EC data (Table 2). ~~Due to the massive structure of the tower and the resulting~~ To minimize wind distortion effects, no samples were collected from wind directions between 70° and 120° N. For the vast majority of the analyzed samples, the mean wind direction was between 180° and 225° N. The deadband was initially scaled with $\delta = 0.7$, as in Zurich, but was increased to $\delta = 0.9$ in October 2023
340 due to very small concentration differences between updrafts and downdrafts. With a pump speed of about 7 l min^{-1} , this was the maximum possible deadband width to collect sufficient air during a 60 min sampling period. Since the concentration differences were still close to the detection limit, the option for HREA was implemented in the logger program (Sect. 2.2) at the

beginning of April 2024. To test the HREA method, nine samples were collected with $H = 0.6$. Due to technical problems with the MGA⁷ in 2023, only EC measurements of the IRGASON are available for 2023. Between November 2023 and January 2024, the MGA⁷ was dismantled for repairs and no REA measurements were conducted.

3.3 Munich - Oberpostdirektion

From July 2024 to April 2025, REA measurements were carried out on a mast of an active telecommunications tower about 1.5 km northwest of the city center of Munich at the site Munich-Oberpostdirektion (ICOS Station ID 'DE-Opd', Table 2). The tower has three platforms up to a height of 59 m and a mast on top, on which the IRGASON and the gas inlets were mounted at a height of 85 m (Fig. C1). In addition, two mid-cost sensor systems, which are based on the Non-Dispersive InfraRed CO₂ sensors GMP343, Vaisala Oyj, Vantaa, Finland, measured the CO₂ concentration at heights of 85 m and 48 m (part of the Munich mid-cost network ACROPOLIS (Aigner et al., 2026)). The tower is located in an area with many residential houses and other buildings (Fig. 3 c). To the southeast is the central railway station and behind it the historic city center. The largest point source, located approximately 200 m to the southeast, is a brewery. [Vegetation accounted for approximately 23 % of the average flux footprint area of the well-mixed REA measurements.](#)

Due to lack of space, the MGA⁷ and the REA sampler were placed in the basement of the tower, requiring inlet lines of 100 m length. During the maintenance of the REA system prior to its installation in Munich, larger flushing pumps were installed (Sect. 2.2). The sampling flow rate was increased to approximately 11 l min⁻¹. With the increased flow rate, less time was needed to collect enough air for laboratory analysis, so a larger deadband ($\delta = 1.1$) could be used. For summer afternoons with predominantly small CO₂ fluxes, a hyperbolic deadband with hole size $H = 0.8$ was used to increase the signal-to-noise ratio.

4 Results and discussion

~~Regular quality control tests of the REA system showed an overall good performance of the hardware (Appendix E). However, as the quality of the REA measurements varies depending on the micrometeorological conditions during sampling and the signal-to-noise ratios, the analyzed REA measurements were flagged as well-mixed measurements, low-turbulence and storage measurements, or were not considered further (Sect. 4.1). Section 4.2 presents an example, namely the results of six REA measurements conducted in Munich on 09 October 2024. To analyze and compare the results from all three cities, the ff and nffluxes are compared with the net fluxes, and the correlation between ff~~

4.1 [Flagging of analyzed REA measurements](#)

~~The quality of a collected REA data set strongly depends on site-specific conditions such as flux strength or micrometeorological conditions, technical settings such as the deadband, and , as well as spatial and temporal patterns, are investigated in Sect. 4.3. Furthermore, the representativeness of the results and the limitations of the methodology are examined. In Sect. 4.4, the potential of REA measurements with low turbulence or substantial storage fluxes is investigated. Finally, the absolute and -based data and knowledge available during the campaign for the selection of suitable flask samples adapted to the scientific~~

question. In our case, the largest number of high-quality ffCO_2 concentrations of the REA flask samples are compared to marine background concentrations (Sect. 4.5). flux data could be collected in Munich.

4.2 Flagging of analyzed REA measurements

In Zurich, only 30 out of 87 REA measurements with ^{14}C and EC data met the criteria finally considered, describing suitable were flagged as well-mixed conditions measurements (Table 3). Twelve samples were selected knowing that with $u_* < 0.2 \text{ m s}^{-1}$ or $|F_{\text{CO}_2, \text{strg}}| > 20 \text{ } \mu\text{mol m}^{-2} \text{ s}^{-1}$ the measurements probably do not represent the surface fluxes during the sampling period. Most of these measurements with low-turbulence or storage flag were taken in the early morning and analyzed to obtain information on the composition of the nocturnal CO_2 fluxes. As it was initially decided to relax the stationarity requirements due to the intermit-

Table 3. Number of REA well-mixed measurements with flux data, low-turbulence and storage measurements, or measurements not considered further (Fig. 2). In brackets, the percentages of the total number of REA measurements with ffCO_2 flux flagged as well-mixed, low-turbulence, storage or data are provided. For the measurements not considered further (Sect. 2.2.4) these measurements may be affected by multiple criteria. Quality control flag QC, beta coefficient β , and signal-to-noise ratio SNR are defined as in Table ??1. Measurements not considered may be affected by multiple criteria.

Number (percentage) of REA measurements ... REA measurements with ffCO_2 flux data	Zurich	Paris	Munich
... with flux data ... All REA measurements	87	65	99
... with well-mixed flag Well-mixed measurements	30 (34 %)	32 (49 %)	78 (79 %)
... with low-turbulence and/or storage flag Low-turbulence and storage measurements	12 (14 %)	4 (6 %)	13 (13 %)
... not considered ... Measurements not considered	45 (52 %)	29 (45 %)	8 (8 %)
Measurements not considered ...			
... due to with QC = 2	22 (25 %)	10 (15 %)	3 (3 %)
... due to with $\beta < 0.1$ or $\beta > 1$	13 (15 %)	10 (15 %)	1 (1 %)
... due to with SNR < 100%	28 (32 %)	21 (32 %)	5 (5 %)

tent nature of CO_2 fluxes in urban environments, 25 % of the periods did not meet the stationarity or well-developed turbulence criteria. The β criterion was not considered in the selection of the flasks, but only 15 % of the measurements were affected. Excluding measurements with $\beta < 0.1$, $\beta > 1$ and QC = 2, β was 0.44 ± 0.14 for a dynamically adjusted deadband width of $0.7\sigma_w$. This is slightly higher than the value of 0.39, which would be expected for a normally distributed timeseries with $\delta = 0.7$ (Fotiadi et al., 2005b), but in good agreement with experimental data (e.g., Pattey et al., 1993; von der Heyden et al., 2022) (see Appendix B2). The main limitation of the Zurich REA measurements was a signal-to-noise ratio of < 100 %, caused by the small $\Delta^{14}\text{C}$ differences between updraft and downdraft samples compared to the mean measurement uncertainty of the Zurich samples of 1.8 ‰ (Δ notation according to Stuiver and Polach (1977)). In Paris, low-turbulence and storage measurements

were ~~mostly discarded~~usually not selected for laboratory analysis. The β coefficient for $\delta = 0.7$ was 0.40 ± 0.20 , i.e., slightly smaller than in Zurich and in good agreement with theoretical expectations for normally distributed time series. Unfortunately, increasing δ to 0.9 did not increase the concentration differences. For the selected measurements, β was even slightly larger on average (0.46 ± 0.17 , see Appendix B2). As in Zurich, the main limitation of the measurements in Paris was a low signal-
395 to-noise ratio. In Munich, the proportion of suitable measurements was significantly improved. The concentration differences were generally increased by a larger deadband width and HREA. The β coefficient was 0.34 ± 0.07 for a deadband with $\delta = 1.1$ and 0.26 ± 0.06 in the case of HREA with $H = 0.8$, i.e., as expected much smaller than in Zurich and Paris (Appendix B2). At the same time, the $\Delta^{14}\text{C}$ measurement uncertainties were reduced by a new AMS from $2.1 \pm 0.3 \text{ ‰}$ (Zurich samples with old AMS) to $1.2 \pm 0.1 \text{ ‰}$, so that samples with $\text{SNR} > 100 \text{ ‰}$ could be selected. As in Zurich, low-turbulence and storage
400 samples collected in the morning were deliberately selected to analyze the $\text{ffCO}_2/\text{CO}_2$ ratio of nocturnal integrated fluxes. In all three cities, storage flux estimates for the well-mixed measurements were on average less than 5 % of the absolute CO_2 fluxes, which justifies the neglect of storage flux corrections for the selected measurement periods. An overview of all REA measurements and their corresponding flags can be found ~~at~~ in Kunz et al. (2025b).

~~The results presented in Table 3 illustrate that the quality of a collected REA data set strongly depends on site-specific conditions such as flux strength or micrometeorological conditions, technical settings such as the deadband, and the data and knowledge available during the campaign for the selection of suitable flask samples adapted to the scientific question. In our case, the largest number of high-quality flux data could therefore be collected in Munich.~~

405

4.2 Example diurnal course

~~To illustrate the principle and to show an example of partitioning net fluxes of collected via EC into fossil and non-fossil flux components using REA, Fig. 4 presents data collected on 09-10 October 2024 in Munich. On that day, micrometeorological conditions were suitable for REA measurements and, favorable micrometeorological conditions enabled repeated REA measurements in Munich throughout the day. All six flask pairs, sampled between 08:00 and 19:00 local time ($\text{LT} = \text{UTC}+2$), were analyzed for $^{14}\text{CO}_2$. The hour-long sampling periods are highlighted in Fig. 4 in light blue. Exemplary analysis of the measured ffCO_2 and nfCO_2 fluxes demonstrates the successful separation of net EC CO_2 fluxes using REA for $^{14}\text{CO}_2$ and its benefits.~~

410 However, the analysis also highlights the challenges and limitations in interpreting and generalizing the data, which arise from the observed variability in the CO_2 flux composition, the overall sparse data coverage, and the large measurement uncertainties.

415

The CO_2 concentration of ambient air, as measured by the two mid-cost sensor systems at 48 m above ground level (Fig. 4 a), ~~follows~~ followed the typical diurnal CO_2 cycle of a warm and sunny summer day (e.g., Stull, 1988; Lan et al., 2020).
420 During night, the CO_2 concentration ~~increases~~ increased and a vertical concentration gradient with highest values close to the surface ~~develops~~ developed. As vertical mixing ~~is~~ was suppressed ($u_* \leq 0.2 \text{ m s}^{-1}$, see Fig. 4 b), this can be attributed to surface emissions accumulating within the stable nocturnal boundary layer. After sunrise, friction velocity, temperature, and radiation ~~increase~~ increased (Fig. 4 b). As the radiative heating of the surface generates convective turbulent vertical motions, the vertical concentration gradient ~~diminishes~~ diminished. The CO_2 concentration ~~decreases~~ decreased at both heights first

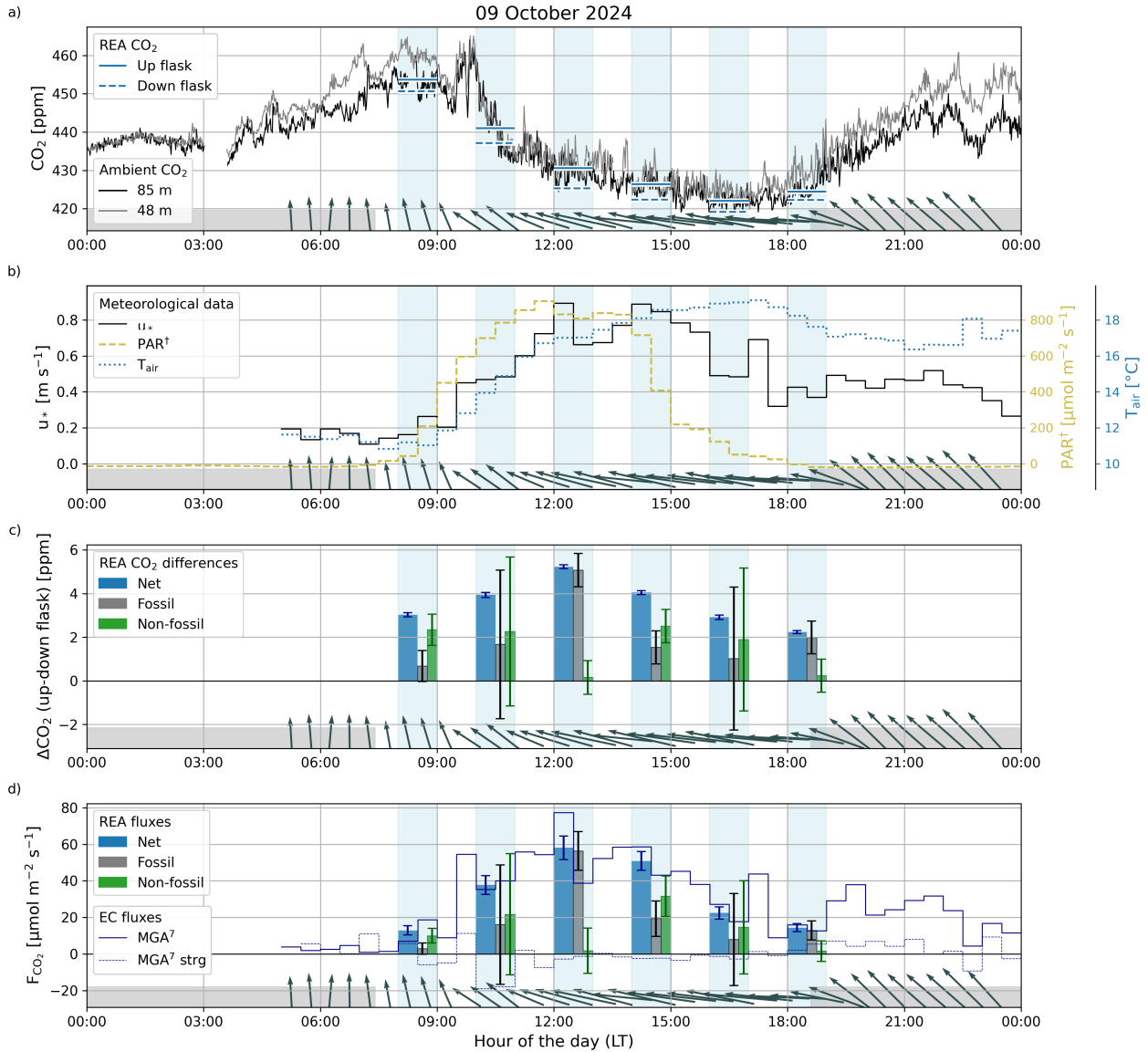


Figure 4. Visualization of EC and REA measurements on 09 October 2024 in Munich. Sampling periods of the six REA measurements are highlighted in blue. Arrows at the bottom of the plots indicate the mean horizontal wind direction and wind speed over 30 min. Day and night times are indicated by the gray bar. a) CO₂ in situ measurements of the GMP343 at 85 m (= REA sampling height) and 48 m together with CO₂ concentrations of the updraft and downdraft flask samples. b) 30 min averages of friction velocity u_* , photosynthetically active radiation PAR and air temperature T_{air} (\dagger PAR was approximated by $1.7 \mu\text{mol J}^{-1}$ times the average incoming shortwave radiation). c) CO₂ concentration differences between updraft and downdraft flask samples ΔCO_2 and their fossil and non-fossil components derived from the respective ¹⁴CO₂ measurements. d) Continuous CO₂ flux and CO₂ storage flux estimates from EC measurements of the MGA⁷ with 30 min averaging period. Blue bars indicate the mean net CO₂ fluxes during the REA sampling periods, gray and green bars the respective fossil and non-fossil components derived from the flask concentration differences.

425 rapidly due to the entrainment of fresh air from higher altitudes, then more slowly as the depth of the atmospheric boundary layer ~~stabilizes~~ stabilized and changes in CO₂ concentration ~~are~~ were primarily driven by the surface fluxes.

The ~~continuous EC measurements net CO₂ flux, on the contrary, did not follow the typical, traffic-dominated, bimodal diurnal cycle that might be expected in the urban environment. Instead, it peaked around noon~~ (Fig. 4 d) ~~show that the turbulent fluxes at 85 height are approximately 10 in the early morning, increase after sunrise, and reach a maximum of more than 60 at noon, before they decrease again. This pattern is reflected in.~~ Accordingly, the net CO₂ concentration differences between the sampled updraft and downdraft flasks ~~, which are 3 ppm in the morning, 5.2 ppm at~~ were also largest during the measurement between 12:00 , and 2.2 ppm in the evening and 13:00 (Fig. 4 c). ~~The~~ However, the ¹⁴C-based ΔffCO₂ estimates (Fig. 4 c) ~~indicate that during~~ provide additional information on the composition of the measured fluxes. At noon and in the evening, ~~these~~ the net CO₂ differences ~~are~~ , and thus also the net CO₂ fluxes, were entirely caused by fossil fuel emissions. Consequently, 435 the ffCO₂ flux ~~equals~~ was equal to the net EC-based CO₂ flux, while the nfCO₂ flux ~~is~~ was approximately zero. It should be noted that $F_{\text{nfCO}_2} \approx 0$ does not necessarily mean that there was no biospheric activity, but only that the positive fluxes (respiration + biofuels) were approximately equal to the photosynthetic uptake. In the morning and in the afternoon, on the other hand, the ΔffCO₂/ΔCO₂ ratio, and thus also the $F_{\text{ffCO}_2}/F_{\text{CO}_2}$ ratio ~~varies~~ varied between 23 % and 43 %, indicating positive nfCO₂ fluxes of about 10 to 30 μmol m⁻² s⁻¹. ~~Unfortunately, the Δff~~ For an urban environment, nfCO₂ uncertainties 440 ~~for the REA measurements at 10:00 and 16:00 LT are unusually high due to technical issues during the AMS measurements in the subsequent lab analysis.~~

~~It must be noted that $F_{\text{nfCO}_2} \approx 0$ does not necessarily mean that there is no biospheric activity, but only that the positive fluxes (respiration + biofuels) approximately equal the photosynthetic uptake. Moreover, as discussed in Sect. 2.2.1, fluxes of 30 μmol m⁻² s⁻¹ are extraordinarily large. As shown in Sect. 4.3.2, such high nfCO₂ fluxes were repeatedly observed from~~ 445 ~~the southeast, indicating emissions from a real nfCO₂ source and an appropriate measurement principle capable of detecting such signals. For the REA measurements in the early morning, it is important to recall that~~ the EC and REA data represent the turbulent fluxes and are not corrected for changes in storage below the measurement height ~~-(Sect. 2.2.1).~~ This is particularly relevant for the measurement at 08:00, where $u_* < 0.2$ m s⁻¹ and the CO₂ concentration at 48 m ~~is~~ was higher than at 85 m. Due to low turbulence, the measurement may not reflect the surface fluxes at the actual ~~sampling time~~ time of sampling. 450 Indeed, the ffCO₂/CO₂ flux ratio of 22 ± 23 % is much lower than expected during the morning rush hour. Although the measurement at 10:00 was flagged as well-mixed, the decrease in CO₂ concentration, the negative storage flux estimate, and the relatively high nfCO₂ contribution (neglecting uncertainties) also indicate a storage contribution, i.e., mixed-up near-surface accumulation from the previous night. This highlights that the 20 μmol m⁻² s⁻¹ threshold flags only the most extreme storage flux measurements, and that the flagging is not unambiguous, especially given the high uncertainty in the storage 455 flux estimates in the morning. Since storage fluxes are usually largest in the morning, the well-mixed measurements ~~are~~ were additionally analyzed for differences between measurements taken before and after 911:00 UTC LT (Sect. 4.3.3).

~~Overall, the measurements on 09 October 2024 in Munich indicate that the contribution of~~ Unfortunately, the ΔffCO₂ emissions to the measured net fluxes can vary significantly. However, ~~-based ffflux data are only available for a limited number~~

of discrete time periods and often have uncertainties $> 100\%$. uncertainties for the REA measurements at 10:00 and 16:00
460 were unusually high due to technical issues during the $^{14}\text{CO}_2$ AMS measurements in the subsequent lab analysis.

4.3 Partitioning of net CO_2 fluxes under well-mixed conditions

4.3.1 Overview of sampling times and ffCO_2 vs. CO_2 fluxes from all three cities

As general overview over the REA measurements from Zurich, Paris, and Munich. A qualitative analysis of the observed ffCO_2
465 fluxes from the well-mixed REA measurements, in relation to net CO_2 fluxes and the time of day (Fig. 5 shows for each city the
net), shows that due to the selection of samples with large SNR the dataset is biased towards high fluxes. While the large CO_2
fluxes during the selected REA sampling intervals with well-mixed conditions over the hour of the day and over the respective
-based ff flux estimates. In this way, the ff/flux ratios (P_{ffCO_2}) as well as their temporal variability and representativeness can
be classified and differences or similarities fluxes clearly differed in composition between the three cities and measurement
campaigns can be analyzed qualitatively. A more quantitative analysis and dependencies on other parameters such as wind
470 direction are discussed in Sect. 4.3.2 and Sect. 4.3.3. sites, the separation of smaller fluxes was limited by measurement
uncertainties, particularly in Zurich and Paris.

In the right panels of Fig. 5, the 1:1 line marks the case when the net

The continuous EC measurements showed that the CO_2 flux equals the ff flux and the nf flux is fluxes were, on average, as
475 expected higher in winter than in summer, especially during the day, pointing at increased emissions and reduced photosynthetic
uptake. The differences were most pronounced in Paris, where in summer the median turbulent flux between 12:00 and 17:00
LT was approximately zero. Accordingly, measurements above the 1:1 line have a net positive nf . This means that in 50 % of the
considered time periods, negative CO_2 flux component, while measurements below the line have a negative nf component fluxes,
i.e., photosynthesis has dominated. The magnitude of the nf , were larger than positive CO_2 flux is indicated by the parallel
dashed lines and the axes on the right, and the F_{nfCO_2} uncertainties by the vertical error bars. Due to error propagation of
480 the fluxes through respiration and anthropogenic emissions. Seasonal differences are also evident in the REA ffCO_2 flux
uncertainties, the nf measurements. In Zurich and Paris, ffCO_2 flux uncertainties are usually much larger than the uncertainties
of the net fluxes, which are shown in the left panels. Negative fluxes were, on average, much smaller in summer than in
winter. However, the representativity of the results is limited by the small number of well-mixed measurements, particularly in
summer. It should also be noted that negative ffCO_2 surface fluxes are unreasonable and are attributed to the limited resolution
485 of small $^{14}\text{CO}_2$ differences between updraft and downdraft samples (the error bars indicate the 1σ uncertainties). Nevertheless,
the measurements are shown here, because they have a significant nfCO_2 component ($\text{SNR} > 100\%$).

For comparison, the median fluxes of the continuous EC measurements are shown in the left panels of Fig. 5 for both summer
and winter. In this work, “summer” refers to the period from 15 July to 31 October, and “winter” to the period from 1 November
to 15 April. This seasonal division of the measurement campaigns aligns roughly with the shift between European. The fact
490 that far fewer negative ffCO_2 flux data points were measured in Munich than in Zurich and Paris highlights the improvement
in the REA measurements in Munich. In Munich, the difference between summer and winter time and with the change in local

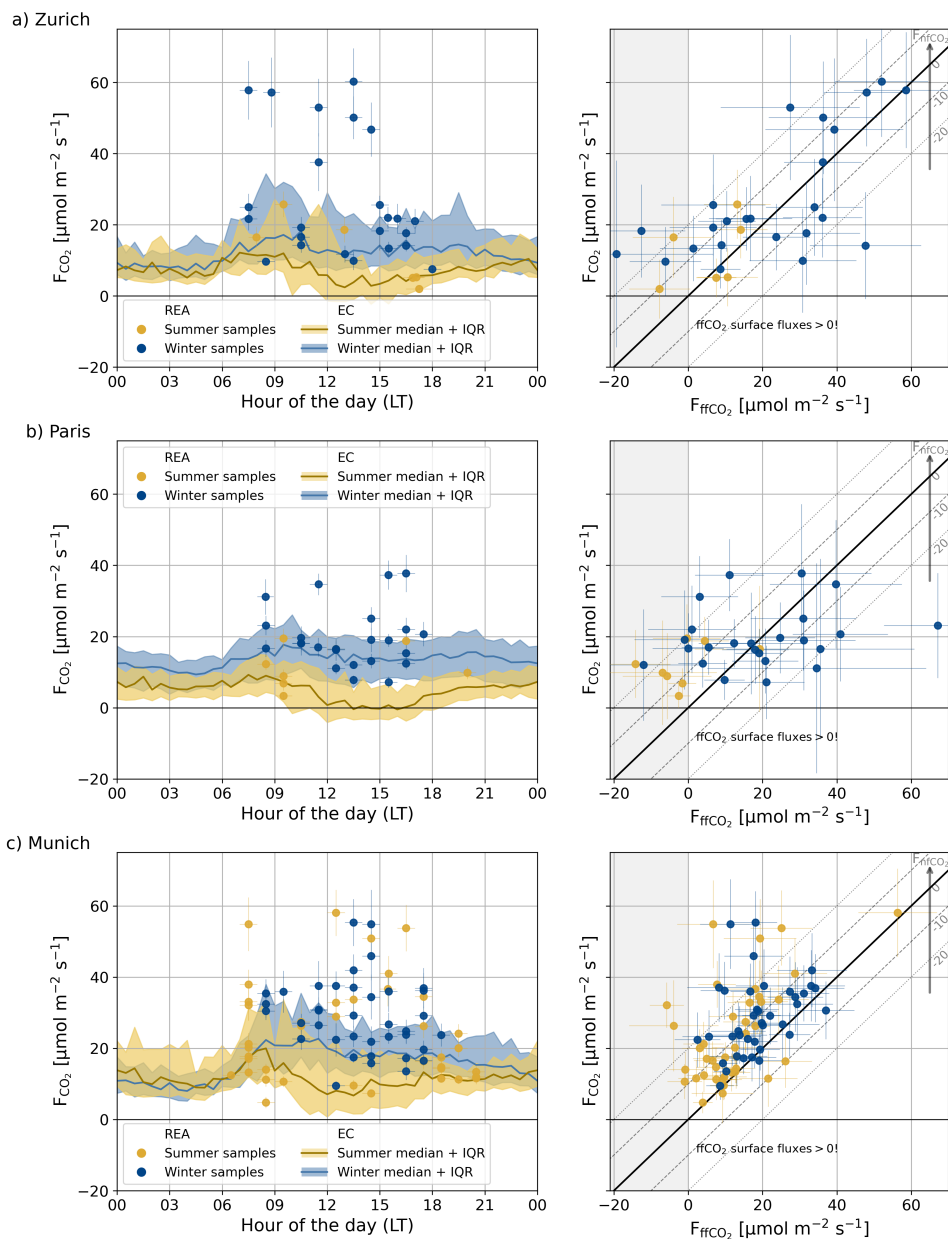


Figure 5. Overview of all REA measurements in Zurich, Paris, and Munich with well-mixed conditions. **The color distinguishes summer measurements (15 July–31 October) and winter measurements (1 November–15 April).** Left: Net CO₂ flux F_{CO_2} during the REA sampling periods over the hour of the day. Error bars in x-direction indicate the length of the REA sampling period (mostly 60 min), error bars in y-direction the uncertainty of F_{CO_2} **derived from the random uncertainty estimates of the EC measurements.** The yellow and blue lines and shaded areas represent the medians and the interquartile ranges (IQR) of the continuous **EC-IRGASON** CO₂ fluxes (see Appendix D for details). Right: CO₂ fluxes during the REA sampling periods compared to the ¹⁴C-based $f\text{fCO}_2$ fluxes. The areas with $F_{\text{ffCO}_2} < 0$ are shaded gray because the physical $f\text{fCO}_2$ fluxes at the surface are positive. The magnitude of the $f\text{fCO}_2$ flux is indicated by the parallel dashed lines and the axes on the right. Error bars in **x-direction represent F_{ffCO_2} uncertainties, error bars in y-direction represent F_{nfCO_2} uncertainties.**

emissions due to heating degree days, and is consistent with other studies conducted in the same location during the same period (Hilland et al., 2025). Due to gaps in the MGA⁷ data (Sect. 2.1), the IRGASON measurements are used, considering only high quality data with $QC < 2$, $u_* \geq 0.2$, signal strength $\geq 90\%$, and wind directions without flow distortion effects (compare Table 2). The shaded areas represent the interquartile ranges. As expected, the fluxes are on average higher in winter than in summer, especially during the day, pointing at increased emissions and reduced photosynthetic uptake. The differences are most pronounced in Paris, where in summer the median turbulent flux between 10:00 and 15:00 UTC is approximately zero. This means that in 50 % of the considered time periods, negative fluxes, i.e., photosynthesis, were larger than positive fluxes through respiration and anthropogenic emissions. Large fluxes observed between 6:00 and 9:00 could be attributed to increased emissions during the morning rush hour and/or venting of nocturnally accumulated. In general, the median flux is highest in Munich, with largest differences compared to Zurich and Paris during summer daytime measurements was much smaller, despite comparable mean air temperatures during the REA measurements to those in Zurich and Paris (Table 4). This could be related to the relatively large proportion of buildings in the footprint of the Munich tower that are heated by district heating.

Compared to the median CO₂ fluxes, the fluxes during the selected REA sampling periods ~~are were~~ often exceptionally high. This ~~is~~ indicates that the dataset is biased towards high fluxes, caused by the systematic selection of flask pairs with large CO₂ concentration differences to increase the potential ffCO₂ signal. In Zurich, all of the analyzed fluxes that exceeded the 75th percentile of the continuous EC fluxes (denoted as P0.75 in the following) were measured in winter and ~~are were~~ almost entirely due to fossil fuel emissions. In Paris, there were only five REA measurements with $F_{CO_2} > P0.75$. As in Zurich, they were measured in winter, but they ~~are were~~ not as clearly dominated by fossil fuel emissions as the large winter fluxes measured in Zurich. In Munich, turbulent fluxes $> P0.75$ were analyzed in both summer and winter, and most ~~have had~~ a significant positive nfCO₂ component. Thus, while the large fluxes represent relatively rare conditions, the high signal-to-noise ratio (which was the main reason for analyzing them) allows observation of differences in the composition of the fluxes between the three cities (cf. Sect. 4.3.2 and Sect. 4.3.3).

REA measurements conducted in Zurich and Paris when CO₂ fluxes were below P0.75 ~~show~~ showed positive and negative nfCO₂ components of up to $\pm 45 \mu\text{mol m}^{-2} \text{s}^{-1}$. However, the uncertainties ~~are were~~ large and there ~~are were~~ very few summer measurements, as most of the measurements were flagged because of $\text{SNR} < 100\%$, $\beta < 0.1$ or $\beta > 1$. In Munich, on the contrary, the uncertainties ~~are were~~ much smaller (see Table 4) and, except for a few measurements, all measurements ~~show~~ showed positive nfCO₂ components. This means that respiration and biofuel emissions were generally larger than photosynthetic uptake. The latter is consistent with the observations from the continuous EC measurements that the net CO₂ fluxes ~~are were~~ highest in Munich and mostly positive throughout the year.

The correlation between the ffCO₂ and CO₂ fluxes ~~is was~~ largest (0.68) for the Zurich winter measurements (Table 4). However, no clear correlation ~~is was~~ observed when only the measurements with $F_{CO_2} < P0.75$ ~~are were~~ considered. This could be caused by a large biospheric signal, a large temporal or spatial variability and/or an insufficient signal-to-noise ratio. To investigate the cause more closely, spatial patterns and expected effects of measurement uncertainties are discussed in Sect. 4.3.2 and Sect. 4.3.3.

Table 4. Mean uncertainties of the ffCO₂ fluxes F_{ffCO_2} and the ffCO₂/CO₂ flux ratios R_{ffCO_2} of the REA measurements under well-mixed conditions in Zurich, Paris, and Munich. In addition, the Pearson correlation coefficients of the ffCO₂ and CO₂ fluxes and the mean air temperatures during the sampling periods are given. P0.75 denotes the 75th percentile of the continuous EC CO₂ fluxes.

Variable	Unit	Zurich		Paris		Munich	
		Summer	Winter	Summer	Winter	Summer	Winter
Number of REA measurements	-	6	24	8	24	40	38
Mean F_{ffCO_2} uncertainty	$\mu\text{mol m}^{-2} \text{s}^{-1}$	8	12	9	13	6	7
Mean R_{ffCO_2} uncertainty	%	126	59	86	71	31	23
Mean air temperature	°C	18	9	20	10	19	6
Correlation($F_{\text{ffCO}_2}, F_{\text{CO}_2}$)...	-	0.47	0.68	0.43	0.22	0.54	0.34
... for $F_{\text{CO}_2} < \text{P0.75}$	-	-0.19	0.25	-0.86	0.31	-0.02	0.63
		($N = 4$)	($N = 15$)	($N = 5$)	($N = 19$)	($N = 26$)	($N = 27$)

4.3.2 Spatial flux patterns and influence from ~~point sources~~large point-source emissions

~~This section examines the spatial distribution of the observed~~Extraordinarily large CO₂ ~~ff- and fluxes analyzed in Zurich and Munich showed different compositions (Sect. 4.3.1). While the Zurich fluxes were mostly fossil, the Munich fluxes contained~~
530 ~~a surprisingly large~~nfCO₂ fluxes-based component. Based on the mean ~~horizontal wind directions during the well-mixed REA sampling periods. The aim is a first-order analysis of spatial patterns, which, if attributed to a specific land cover type, for example, is an important step towards generalizing the discrete flux measurements~~ wind directions and modeled flux
~~footprints, we found that these measurements were likely affected by fossil fuel emissions from a district heating plant in Zurich and non-fossil emissions from a fermentation process in a brewery in Munich, respectively. The conclusive results~~
535 ~~indicate high quality of the REA and EC measurements, and the footprint analysis. Moreover, these point-source influenced measurements highlight the challenges involved in interpreting atmospheric observations in a complex urban environment.~~

In Zurich, the net CO₂ fluxes observed with wind from the west ~~are~~were generally smaller than those with wind from the east (Fig. 6 a). The CO₂ fluxes > P0.75, which ~~are~~were clearly dominated by fossil fuel emissions (Fig. 6 b), ~~are~~were observed from about 70° N and 135° N. This is consistent with the high proportion of vegetated areas in ~~the~~west, in contrast to
540 ~~the city center, a district heating plant, and arterial roads in the east (Sect. 3.1). As emissions from point sources are generally not representative of the average fluxes in a city, and the comparison of measured and modeled point source emissions on an hourly basis is limited by uncertainties in the emissions inventory and transport models, we attempted to identify the REA measurements which were potentially influenced by emissions from the district heating plant. For this purpose, we considered the individual flux footprints according to Kljun et al. (2015) Based on analysis of the flux footprints and the operating times~~
545 ~~of the district heating plant, which are known with a temporal resolution of 5 . Based on these data, the measurements we identified 10 measurements from the southeast that were~~ potentially influenced by emissions from the district heating plant~~were identified (represented in Fig. 6 in red, see . See Appendix F for details). Due to uncertainties and limitations in the footprint~~

modeling, we also attempted to investigate the potential flux contribution from the district heating plant based on observations. generally enables a distinction between from natural gas, which is used in the district heating plant and has a signature (in δ notation) of about -40 (Tans, 1981; Widory and Javoy, 2003), and that from liquid and solid fuel or biogenic fluxes with $\delta \approx -25$ (Tans, 1981; Widory and Javoy, 2003; Bakwin et al., 1998). However, the Zurich REA flasks were not analyzed for by the high-precision ICOS Flask and Calibration Laboratory, but as a by-product of the extraction at the ICOS CRL, with an order-of-magnitude lower precision of about 0.2 . Thus, although the potentially influenced samples showed influence from an isotopically lighter source, this was not significant within the measurement uncertainties, and an unambiguous gas source attribution was not possible. A contribution from a gas source is therefore likely, but cannot be clearly attributed to individual measurements. Flagging of measurements with a potential contribution from emissions from the district heating plant was therefore based on footprint data alone.

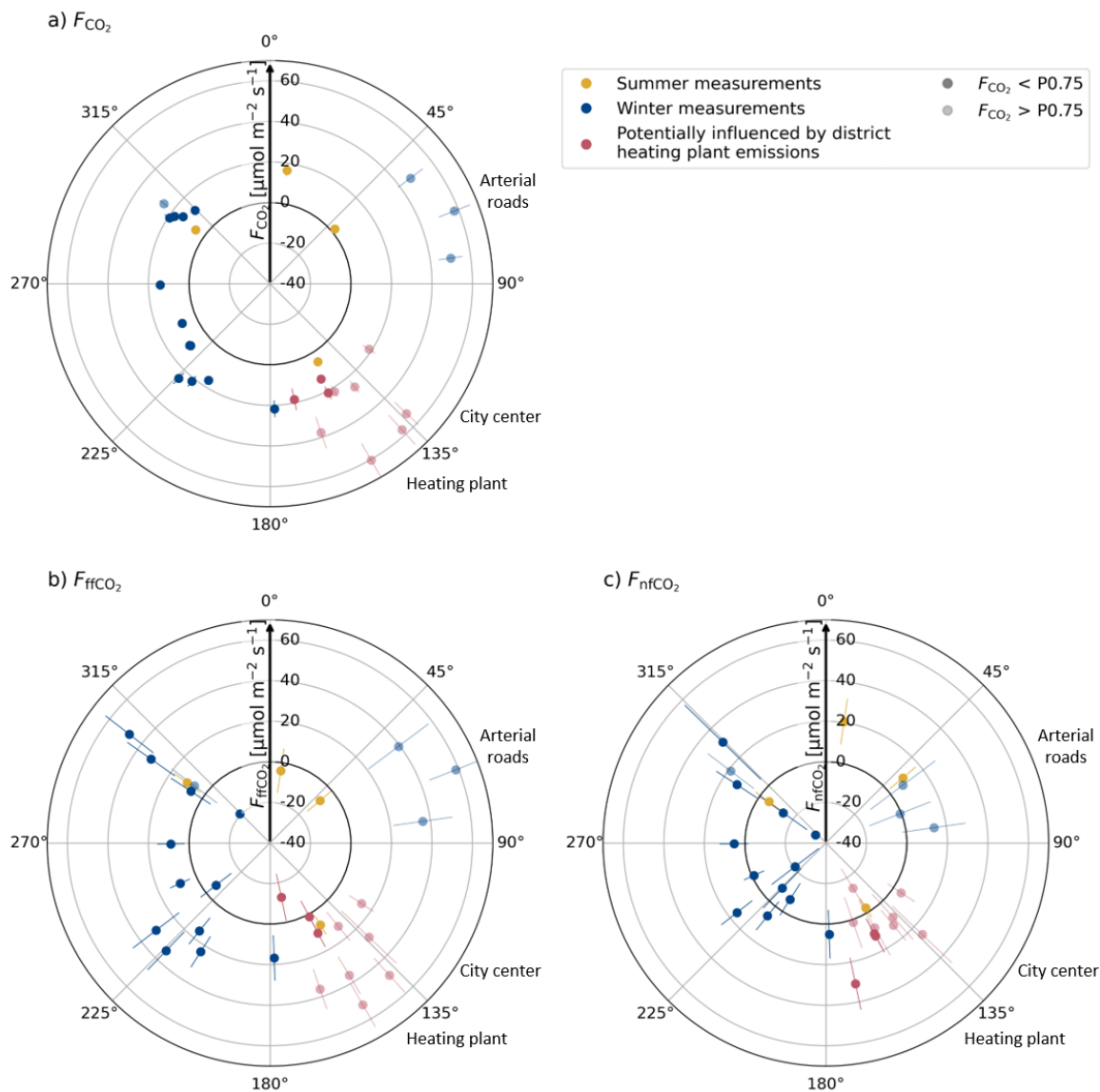


Figure 6. Net CO₂ fluxes (a), ffCO₂ fluxes (b), and nfCO₂ fluxes (c) with respect to the mean wind directions during the measurement intervals in Zurich with well-mixed conditions. The error bars represent the respective flux uncertainties. Measurements potentially influenced by emissions from a district heating plant to the southeast are indicated in red. P0.75 denotes the 75th percentile of the continuous EC CO₂ fluxes at the respective hour of the day of the respective season. Indicated are also the directions of the arterial roads, the city center, and the district heating plant.

In Paris, measurements were primarily taken during south-southwesterly wind. Due to the sparse data coverage, no spatial patterns can could be investigated. There is no evidence of any distinct point-source emissions that could have affected the REA measurements. For completeness, the corresponding directional figures for Paris are shown in Appendix G.

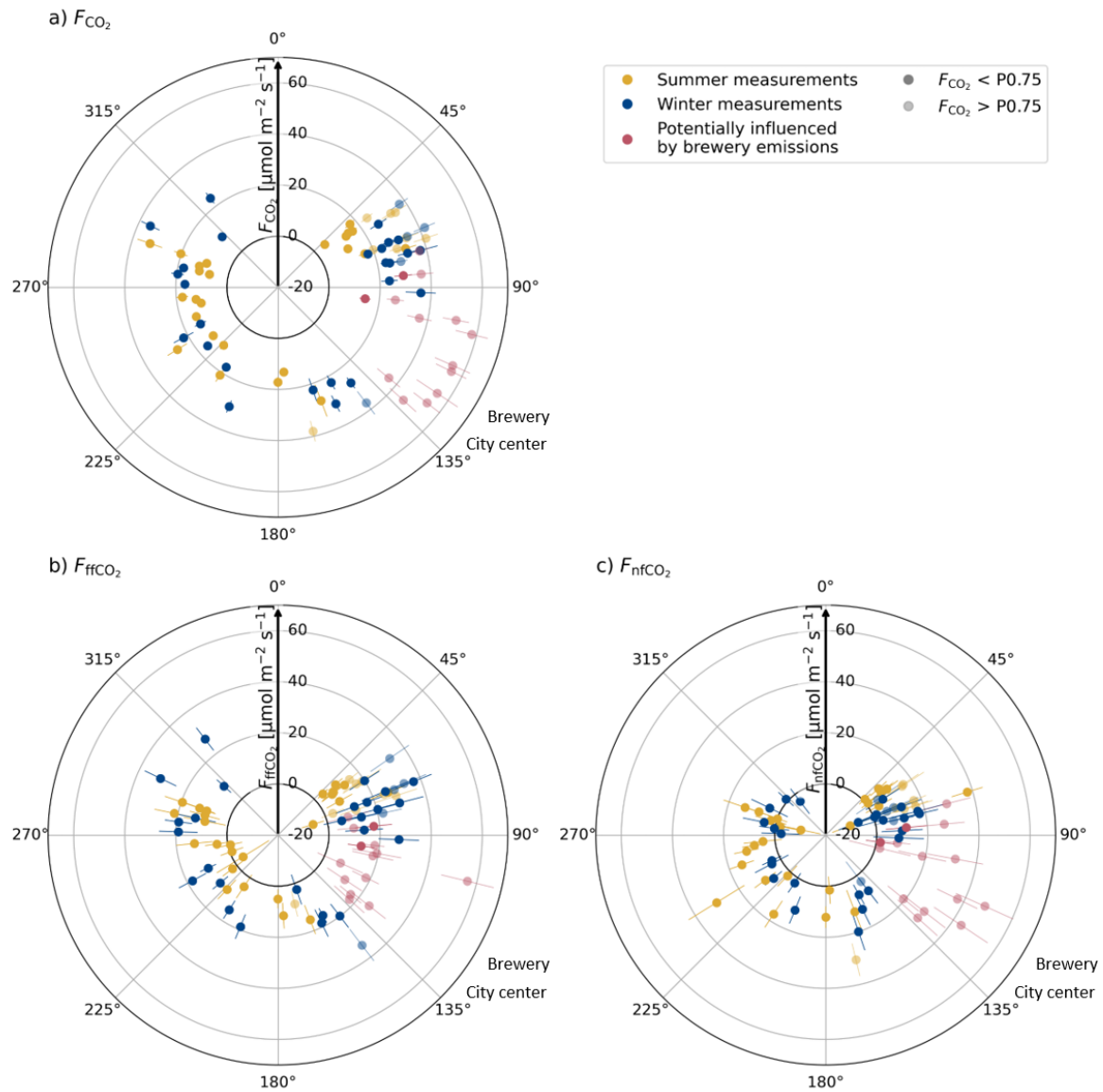


Figure 7. Net CO₂ fluxes (a), ffCO₂ fluxes (b), and nfCO₂ fluxes (c) with respect to the mean wind directions during the measurement intervals in Munich with well-mixed conditions. The error bars represent the respective flux uncertainties. Measurements potentially influenced by emissions from a district heating plant to the southeast are indicated in red. P0.75 denotes the 75th percentile of the continuous EC CO₂ fluxes at the respective hour of the day. Indicated are also the directions of the brewery and the city center.

In Munich (Fig. 7), the highest CO₂ fluxes were measured when the wind came from southeast-east. Located in this direction are a brewery, the central railway station, and the historic city center (~ 0.3 , 1, and 2 km horizontal distance, respectively, see Sect. 3.3). Striking are the large nfCO₂ fluxes of up to $50 \mu\text{mol m}^{-2} \text{s}^{-1}$. The fact that biospheric and human respiration fluxes are typically much smaller (e.g., Wu et al., 2022; Stagakis et al., 2023b, 2025) indicates a non-respiratory anthropogenic

565 nfCO₂ source. Footprint analyses of the respective measurements, using the model of Kljun et al. (2015), ~~show~~showed that the brewery was within the peak area of the flux footprint (Appendix F). Therefore, we assume that the large nfCO₂ emissions from the southeast-east ~~result~~resulted from a fermentation process (Elshani et al., 2018; Olajire, 2020). As there is no information available regarding operating times or the temporal emission profile of the brewery, all measurements with a substantial flux contribution from the brewery area, as estimated from the flux footprints, were considered to be potentially influenced by ~~these~~these 570 ~~large non-fossil~~large non-fossil point-source emissions (see Appendix F). Apart from measurements from the southeast, all nfCO₂ fluxes > 20 μmol m⁻² s⁻¹ were measured in the early morning. As discussed in Sect. 4.2, this could indicate an unaccounted contribution from storage fluxes, further supported by the uncertainties in the distinction between low-turbulence ~~and~~and storage and well-mixed conditions (Sect. 4.4).

~~The results thus emphasize that tall-tower measurements in urban environments can often be affected by individual point sources. In Zurich and Munich, most REA measurements with $F_{\text{CO}_2} > P0.75$ are likely attributable to fossil emissions from a district heating plant and non-fossil emissions from a brewery, respectively. The measurements~~The measurements provide 575 confidence in the EC and REA measurements as well as in the footprint analysis, and could be used to validate or refine bottom-up emission estimates of the respective point sources. Non-fossil CO₂ emissions from fermentation processes in breweries, for example, are usually not included in emission inventories. For the characterization of the usually smaller CO₂ fluxes and the 580 analysis of the biospheric nfCO₂ fluxes, however, these measurements need to be excluded.

4.3.3 Mean ffCO₂/CO₂ ratios and mean nfCO₂ fluxes

~~To generalize and quantify the results from~~When the measurements potentially influenced by the large CO₂ emissions from a district heating plant and a brewery near the measurement sites in Zurich and Munich were excluded, the ~~individual REA measurements, we analyze the error-weighted~~individual REA measurements, we analyze the error-weighted mean ffCO₂/CO₂ flux ratios \bar{R}_{ffCO_2} and the mean magnitude of the nffluxes 585 \bar{F}_{nfCO_2} for each city. ~~Due to the small number of measurements, it is not possible to fully account for the spatial and temporal variability. Based on the analyses in Sect. 4.3.1 and Sect. 4.3.2, we distinguish between summer and winter measurements, and exclude measurements which were potentially influenced by identified point-source emissions. Despite the flagging, it should be noted that due to the sampling method, the results may not be representative of the mean continuous fluxes. This especially limits comparisons between the three cities.~~

590 ~~If ffand fluxes were perfectly linearly correlated, the mean ff/ratios would be best described by the slope of an error-weighted total least squares regression line (Maier et al., 2024a). Due to the generally low correlations of the observed REA fluxes (Table 4), however, \bar{R}_{ffCO_2} is determined as the error-weighted mean of the individual ff/ratios. To minimize the uncertainty, the individual R_{ffCO_2} values are calculated directly from the flask measurements as $\Delta\text{ff}/\Delta$, i.e., completely independent of the EC flux measurements (compare Eq. 3). $R_{\text{ffCO}_2} > 100\%$ indicates a negative nfflux, i.e., photosynthetic uptake, while 595 $R_{\text{ffCO}_2} < 0\%$ is physically unreasonable and only observed if Δff is slightly negative within its measurements uncertainties. ~~In addition, of the mean and variability of the well-mixed measurements were approximately 50 % or less in summer and 80 to 90 % in winter. In Zurich and Paris, however, the significance and representativeness of the results were limited by the small number of measurements. In Munich, on the contrary, average nfCO₂ fluxes are examined. A z-test is used to evaluate~~~~

600 whether the observations are significantly different from $\bar{R}_{\text{ffCO}_2} = 100\%$ or $F_{\text{nfCO}_2} = 0$ (significance level of 0.05), i.e., completely fossil contributions were significantly greater than zero, particularly in the early morning in summer. This highlights the improvements achieved in the REA measurements and the importance of considering nfCO_2 fluxes, taking into account the mean measurement uncertainties (Appendix H). To meet the assumption of normal distribution, only measurements with relative Δ uncertainties $\ll 1$ are considered (most, but not all, of these samples are already excluded by the consideration of the signal-to-noise ratio as defined in Sect. 2.2.4). fluxes in cities.

Table 5. Error-weighted mean $\text{ffCO}_2/\text{CO}_2$ flux ratio \bar{R}_{ffCO_2} and error-weighted mean nfCO_2 flux F_{nfCO_2} of the well-mixed REA measurements, excluding measurements in Zurich and Munich, which were potentially influenced by identified point-source emissions, and four measurements with $\Delta\text{CO}_2 < 0.4$ ppm. N is the number of samples. Stars indicate that, given the number of measurements and mean measurement uncertainties, the results are significantly different from $\bar{R}_{\text{ffCO}_2} = 100\%$ or $F_{\text{nfCO}_2} = 0 \mu\text{mol m}^{-2} \text{s}^{-1}$, respectively (* $p < 0.05$, ** $p < 0.01$, *** $p < 0.001$).

	\bar{R}_{ffCO_2} [-]			F_{nfCO_2} [$\mu\text{mol m}^{-2} \text{s}^{-1}$]		
	Zurich	Paris	Munich	Zurich	Paris	Munich
Summer measurements	$48 \pm 52\%$ ($N = 3$)	$-7 \pm 22\%$ *** ($N = 8$)	$47 \pm 4\%$ *** ($N = 33$)	0 ± 4 ($N = 3$)	9.7 ± 2.2 *** ($N = 8$)	7.8 ± 1.0 *** ($N = 33$)
Winter measurements	$92 \pm 11\%$ ($N = 16$)	$80 \pm 10\%$ * ($N = 23$)	$76 \pm 4\%$ *** ($N = 31$)	1.5 ± 2.7 ($N = 16$)	2.7 ± 2.1 ($N = 23$)	5.3 ± 1.1 *** ($N = 31$)

605 In Zurich, no significant average nfCO_2 signal (p -values > 0.05) was observed (Table 5). In summer, the mean $\text{ffCO}_2/\text{CO}_2$ flux ratio was $48 \pm 52\%$ and the mean absolute nfCO_2 flux was $0 \pm 4 \mu\text{mol m}^{-2} \text{s}^{-1}$. The significance of the results was mainly limited by the small number of well-mixed measurements ($N = 3$). In winter, the mean ffCO_2 contribution of the Zurich samples was $92 \pm 11\%$. To resolve the presumably small mean nfCO_2 component, many more measurements and/or smaller measurement uncertainties would have been necessary (see Appendix H).

610 In Paris, the eight selected summer samples showed mostly non-fossil CO_2 contributions. The negative mean ffCO_2 ratio can-could be explained by the ffCO_2 flux uncertainties (compare Fig. 5), but a larger ffCO_2 contribution was expected. Note that most of the measurements were conducted in the early morning. Therefore, storage fluxes cannot be ruled out. However, due to the small number of samples, a further subdivision of the measurements into morning and afternoon measurements, for example, was not feasible. Similar to the Zurich measurements, the Paris measurements were generally more successful in
615 winter than in summer due to larger signals. The mean ffCO_2 contribution in winter was $80 \pm 10\%$, meaning that, on average, about 20% of the observed CO_2 emissions were due to positive nfCO_2 fluxes.

In Munich, the higher data quality and greater number of measurements enabled the detection of significant nfCO_2 contributions in both summer and winter. The larger nfCO_2 fluxes observed in summer compared to winter are-were primarily attributed to the measurements taken in the early morning during summer. When only 18 summer measurements taken after 911:00 UTC
620 are-LT were considered, \bar{R}_{ffCO_2} is-was $64 \pm 6\%$ and F_{nfCO_2} is-was $5.6 \pm 1.3 \mu\text{mol m}^{-2} \text{s}^{-1}$, which is much smaller than

for the early-morning measurements and comparable to the winter measurements. In winter, no significant differences were observed between measurements taken before and after 9:00 UTC. This could be explained by larger respiratory fluxes and nfCO_2 dominated storage fluxes in the morning (Sect. 4.4) and larger photosynthetic uptake, i.e., negative nfCO_2 fluxes in the afternoon. This temporal variability is larger in summer than in winter. With the reduced measurement uncertainties, about 50 to 100 REA measurements are sufficient to identify average nfCO_2 fluxes of the order of 10 % or $3 \mu\text{mol m}^{-2} \text{s}^{-1}$ at a 0.05 significance level (Appendix H).

Overall, it is remarkable that the mean nfCO_2 contributions are positive in all three cities, both in summer and in winter. Only a few measurements show a significant negative nfCO_2 flux. This contrasts with various studies that estimated negative nfCO_2 fluxes in urban areas, particularly during the warm growing season but also during the cold dormant season (e.g., Wu et al., 2022; Miller et al., 2020). The positive nfCO_2 fluxes in our study could be explained, for example, by the low proportion of vegetated area within the flux footprints (Fig. 1). However, the differences in \bar{F}_{nfCO_2} between the three cities could not be attributed to the respective average vegetated area, which was slightly larger in Zurich than in Paris and Munich (Table 2). It should also be noted that the observed nfCO_2 fluxes include human respiration. According to bottom-up estimates for 2022, the mean annual human respiration fluxes within a $2 \times 2 \text{ km}^2$ square around the measurement sites are about $2.5 \mu\text{mol m}^{-2} \text{s}^{-1}$, or 10 % of the net CO_2 fluxes in all three cities (Dröge et al., 2024). For comparison, the estimated human respiration flux in the footprint of the study by Wu et al. (2022) was only $0.22 \mu\text{mol m}^{-2} \text{s}^{-1}$. Human respiration could therefore account for a significant proportion of the observed nfCO_2 fluxes. Moreover, due to the small number of analyzed samples and the systematic selection of samples with presumably large concentration differences, the results may be biased toward periods with positive nfCO_2 fluxes. As a further analysis, a 1:1 comparison of the REA fluxes with the emission inventories and biospheric models, taking into account the respective flux footprints, could be useful. As the example of the high nfCO_2 fluxes from the direction of a brewery in Munich shows, the measurements could also be influenced by other anthropogenic nfCO_2 point sources. In Munich, for instance, there are also other, more distant breweries. Based on our flux footprint analysis, we excluded all measurements where one of these breweries could have impacted the measured flux. However, excluding these measurements had no significant impact on the results (not shown here). Consistent with the aforementioned studies, our measurements underscore the importance of nfCO_2 fluxes in urban areas.

4.4 Low-turbulence and storage measurements

~~This section presents the f_f /flux ratios of the low-turbulence and storage measurements conducted before 11:00 UTC. While turbulent fluxes measured when $u_* < 0.2$ and/or $|F_{\text{CO}_2, \text{strg}}| > 20$ are unlikely to represent the instantaneous surface fluxes, their composition contains information about the relative strength of individual sources during the time of suppressed turbulence (Sect. 2.2.1). Therefore, measurements taken in the morning are assumed to contain information about the mean nocturnal emissions. As there is no photosynthetic activity at night, the nfCO_2 component represents the sum of soil respiration, plant respiration, human respiration, and biofuel burning. Since most low-turbulence and storage samples in Zurich were collected in winter and most low-turbulence and storage samples in Munich were collected in summer, the results are analyzed with respect to the mean night air temperature (Fig. 8).~~

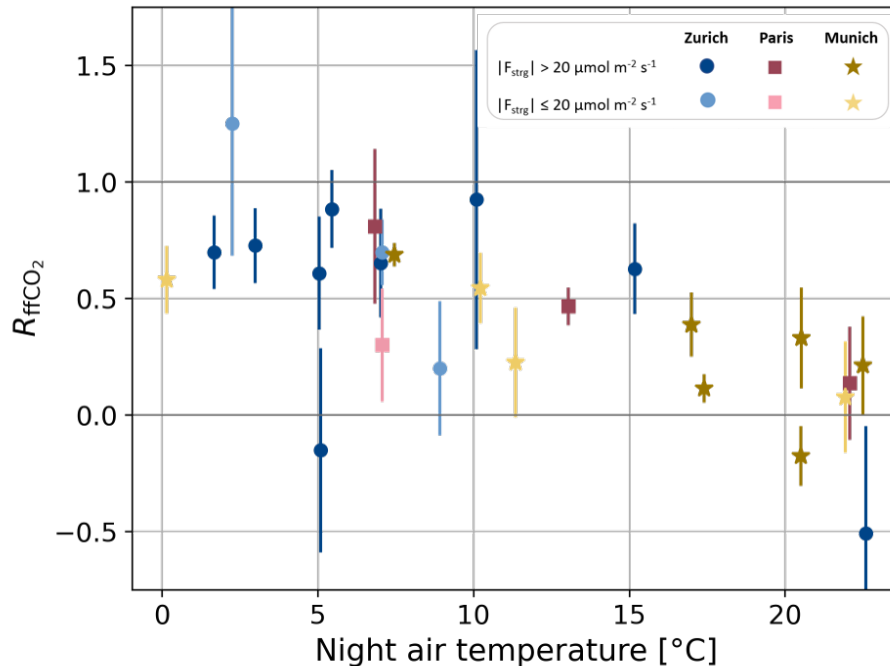


Figure 8. $\text{ffCO}_2/\text{CO}_2$ flux ratios (R_{ffCO_2}) of the low-turbulence and storage measurements taken before 11:00 UTC. The colors indicate whether $|F_{\text{CO}_2,\text{strg}}| > 20 \mu\text{mol m}^{-2} \text{s}^{-1}$ (storage flag) or $|F_{\text{CO}_2,\text{strg}}| \leq 20 \mu\text{mol m}^{-2} \text{s}^{-1}$ (low-turbulence flag only). The error bars represent the measurement uncertainties. The x-axis shows the mean air temperature between 0:00 and 6:00 UTC on the respective days.

655 Although the low-turbulence and storage samples were collected under very different conditions, i.e., in different cities, with variable contributions from surface and storage fluxes, at different times of the year, etc., the $\text{ffCO}_2/\text{CO}_2$ flux ratios are were mostly $< 70 \%$ and larger during cold temperatures than during warm temperatures -(Fig. 8). As the low-turbulence and storage measurements are assumed to contain information about the fluxes prior to the measurement period, the increased ffCO_2 contribution in the morning measurements is most likely due to reduced traffic emissions at night, as well as no heating emissions and increased biospheric respiration, particularly in summer.

660 In Zurich, the error-weighted mean $\text{ffCO}_2/\text{CO}_2$ flux ratio of the samples with night-time temperatures $< 10 \text{ }^\circ\text{C}$ was $68 \pm 7 \%$. This indicates that the surface fluxes, as well as the accumulation of CO_2 in the stable nocturnal boundary layer, were primarily caused by fossil fuel emissions, e.g., due to building emissions, traffic, or industrial processes. However, there was also a substantial ffCO_2 contribution of about 30 % or more in winter. The samples collected in Munich with night temperatures

665 $> 10 \text{ }^\circ\text{C}$ show a showed a much lower mean $\text{ffCO}_2/\text{CO}_2$ ratio of $16 \pm 4 \%$. The increased ffCO_2 contribution is most likely due to reduced traffic emissions at night, as well as no heating emissions and increased biospheric respiration in summer.

The results are in good agreement with other studies. Moriwaki et al. (2006) attributed the nocturnal build-up of CO_2 in a suburban canopy layer in winter to the subsidence of (fossil) building emissions. Wu et al. (2022) observed nocturnal

670 $\text{ffCO}_2/\text{CO}_2$ flux ratios in Indianapolis of $\sim 66\%$ in winter and $\sim 33\%$ in summer. In general, nocturnal net ecosystem exchange is found to be much larger in summer than in winter (e.g., Crawford and Christen, 2015; Stagakis et al., 2025).

4.5 Comparison with regional CO_2 enhancements

675 ~~While the REA flask measurements aimed to analyze turbulent fluxes at the urban neighborhood scale, the absolute flask concentrations also contain information about the fossil and non-fossil enhancements compared to clean background air and thus about the composition of fluxes in a broader continental region, including other urban areas and regional emission sources. Following Levin et al. (2003), we calculate the excess from the mean and Δ values of the up- and downdraft REA sample pairs, using the corresponding concentration measurements at the European marine background station Mace Head on the western coast of Ireland as background concentrations, and assuming that the biogenic $\Delta^{14}\text{C}$ signature equals the background concentration (see Appendix A2). Second-order effects, such as enriched heterotrophic respiration and nuclear contamination (Maier et al., 2023), were not considered because the necessary concentration footprints are only available until~~
680 ~~the end of 2023, and the corrections are negligible for our analysis. For details and an evaluation of these corrections on the Zurich measurements, we refer to Maier et al. (2023) and Appendix A2. The mean ratios of the excess concentrations thus represent the average contributions of emissions to the fluxes on the trajectories between Mace Head and the three measurement sites. The results from all REA flask samples (the micrometeorological flagging criteria do not have to be applied for concentration measurements) are shown in Fig. 9. For clarity, the uncertainties of about 1 ppm are omitted, but~~
685 ~~are considered in the orthogonal regression.~~

While the concentration differences between updraft and downdraft samples, which were used to calculate the turbulent ffCO_2 fluxes (Eq. 3), ~~are were~~ typically about 1 ppm, with a maximum of 14 ppm, the CO_2 and ffCO_2 enhancements compared to the background concentrations ~~are were~~ significantly larger, especially in Zurich (median / maximum CO_2 enhancement of 14 / 123 ppm). Moreover, the regional CO_2 and ffCO_2 enhancements ~~are were~~ much more correlated than the local
690 turbulent fluxes and ~~show showed~~ a clear difference between summer and winter ~~-(Fig. 9)~~. For the summer samples, the mean $\text{ffCO}_2/\text{CO}_2$ ratio obtained from orthogonal regression ~~is was~~ 28 % for Zurich, 19 % for Paris, and 21 % for Munich, indicating that about 80 % of the net CO_2 enhancements in summer ~~are were~~ due to non-fossil CO_2 emissions. For the winter samples, the average ratio ~~is was~~ 63 % for Zurich, 51 % for Paris and 51 % for Munich, i.e., still much lower than the typical ffCO_2 flux contributions in the flux footprints (compare Sect. 4.3.3).

695 The results illustrate that the absolute CO_2 concentrations at the measurement site ~~are were~~ primarily driven by the background concentration (between 413 ppm and 435 ppm) and the regional CO_2 fluxes integrated along the path from the marine background station to the urban area. In comparison to the local CO_2 emissions, the regional fluxes ~~are were~~ much more dominated by non-fossil CO_2 emissions, in this case presumably biospheric respiration. The results agree well with those of Turnbull et al. (2015), who found that the ffCO_2 enhancements measured in the city of Indianapolis with respect to a
700 continental background station were two to three times higher than when a local background station directly upwind of the city was used. With a continental background, the ffCO_2 enhancements accounted for only about 50 % of the net CO_2 enhancement, whereas the local CO_2 enhancement could be almost entirely explained by the ffCO_2 contribution. Therefore, the CO_2 fluxes

analyzed in this paper represent only the local urban emissions and differ significantly from the net emissions in the surrounding area. When analyzing CO₂ concentrations, the choice of the background station is of great importance and must be adapted to the scientific question.

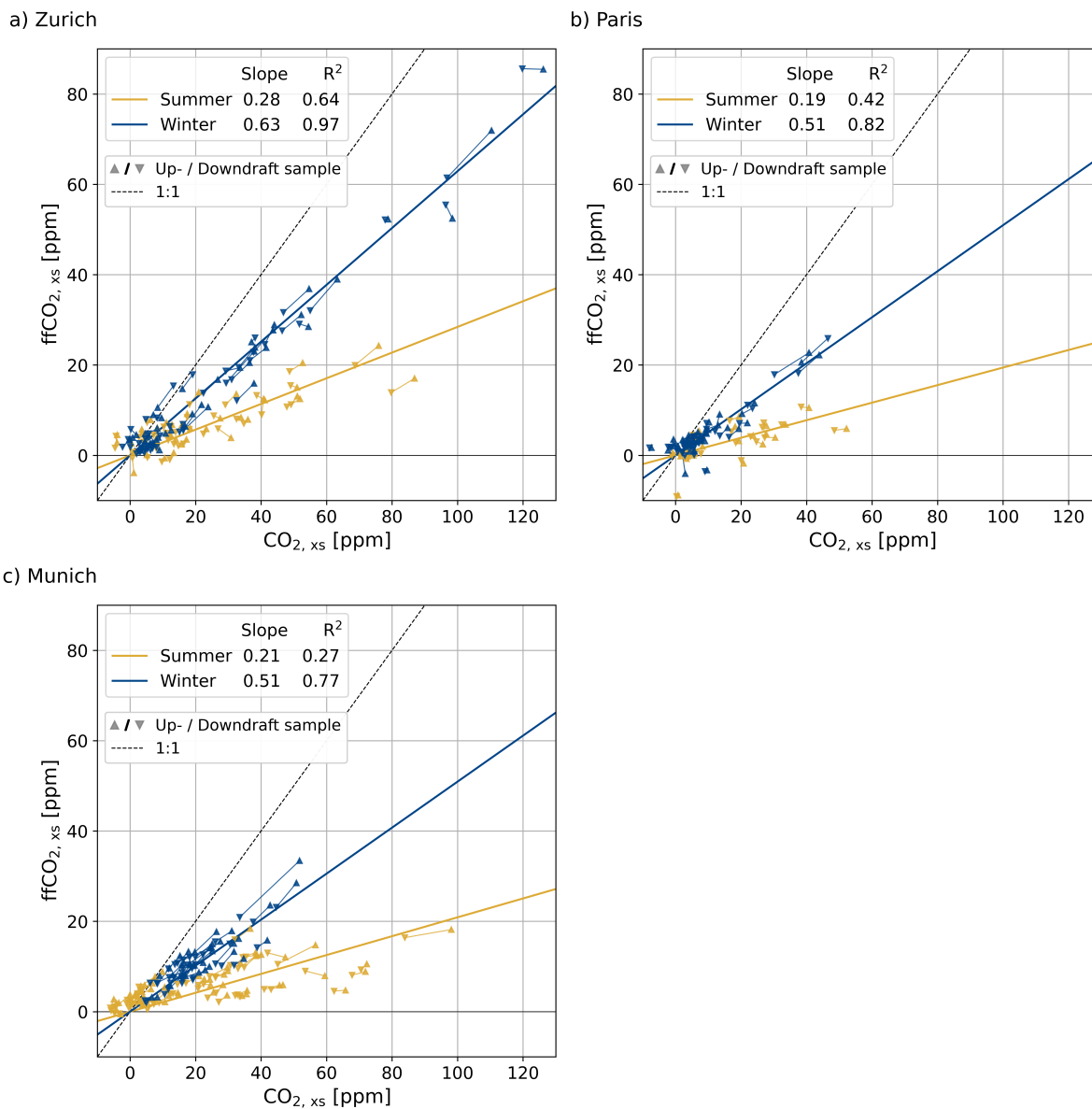


Figure 9. CO₂ and ffCO₂ excess concentrations (“xs”) of the REA flask samples compared to concentration measurements at the European marine background station Mace Head. The pairs of updraft and downdraft measurements are connected by a line. For each site, the slope and the coefficient of determination R^2 of a linear regression through the origin for the summer and winter measurements are given. For clarity, the uncertainties of about 1 ppm are omitted, but are considered in the orthogonal regression.

5 Summary

~~A REA flask sampling system for based estimation of ff fluxes was operated alongside continuous EC flux measurements on three urban tall towers in the cities of Zurich, Paris, and Munich for about nine months each. The analysis of 252 REA measurements was presented with regard to three research questions.~~

710 ~~1. Potentials and limitations of REA measurements for flux partitioning in cities~~

5.1 Potentials and limitations of $^{14}\text{CO}_2$ REA measurements for CO_2 flux partitioning in cities (Q1)

This study demonstrates the successful implementation of the REA method for $^{14}\text{CO}_2$ measurements as a powerful technique for a purely observation-based separation of fossil and non-fossil CO_2 fluxes. The ~~Munich efficacy of the partitioning approach is demonstrated by observations of extraordinarily large nf CO_2 fluxes in Munich, which could be attributed to non-fossil anthropogenic emissions from a brewery. Moreover, the Munich~~ measurements show that with an improved technical setup and an adapted flask sampling and selection strategy, average nf CO_2 fluxes of the order of 10 % or $3 \mu\text{mol m}^{-2} \text{s}^{-1}$ can be identified with a reasonable number of measurements (50 to 100). ~~Assuming scalar similarity between and, the~~ The primary contributor to the overall flux partitioning uncertainty was the current $^{14}\text{CO}_2$ measurement precision in the laboratory. At the given CO_2 source strengths within the flux footprints of the chosen measurement sites, the signal-to-noise ratios were often
715 below 100 %. Situations with large fluxes are therefore favorable for the uncertainty-limited REA measurements and were preferentially selected for sample analysis. This systematic sample selection can introduce biases in the retrieved flux partitioning compared to the mean CO_2 fluxes. Due to the complex, heterogeneous nature of urban environments, the micrometeorological requirements, and the costs and logistics associated with $^{14}\text{CO}_2$ analyses, the ^{14}C -based separation of ff CO_2 and nf CO_2 fluxes is limited to a small number of time periods and cannot be easily generalized.
720

725 ~~2. Indications for point sources and typical fossil and non-fossil flux compositions~~

5.2 Indications for large point source emissions and typical fossil and non-fossil CO_2 flux compositions (Q2)

In Zurich and Munich, sectorial high ff CO_2 or nf CO_2 fluxes indicated significant fossil and non-fossil anthropogenic CO_2 sources. Based on the respective flux footprints, these observations were potentially influenced by ~~point-source~~ emissions from a district heating plant in Zurich and a brewery in Munich, respectively. Excluding the measurements potentially influenced
730 by ~~the identified large~~ point-source emissions, the mean ff CO_2/CO_2 flux ratios of the analyzed winter measurements from the remaining urban emission mix were about 80 to 90 % at each of the three measurement sites, with average nf CO_2 fluxes of about $2 \mu\text{mol m}^{-2} \text{s}^{-1}$ in Zurich and Paris and $5 \mu\text{mol m}^{-2} \text{s}^{-1}$ in Munich. In Zurich and Paris, however, the average nf CO_2 components were within the uncertainties of the partitioning approach. In Munich, on the contrary, average nf CO_2 contributions were significantly larger than zero, especially in summer in the early morning and during conditions of low
735 turbulence and/or changes in storage below the measurement height.

~~The conclusive results from the measurements potential influence by point-source emissions can be regarded as proof-of-concept for the flux estimation based on EC measurements, flux partitioning based on REA measurements, and flux footprint modeling.~~

At the same time, these measurements highlight the challenge of potential impact from point sources on tall-tower measurements in urban areas. Thus, to upscale tall-tower measurements to the city scale, compare them with bottom-up estimates, or to integrate them into inversion models, emission inventories and footprint models must represent point-source emissions (both fossil and non-fossil) and their temporal emission characteristics with high temporal and spatial resolution. The footprint models must also be capable of accounting for different emission heights and potential plume rise. The Munich measurements indicate the importance of plant and soil respiration, human respiration, and non-respiratory anthropogenic emissions, especially during night. However, the representativeness of the selected REA measurements must be further analyzed. This again highlights the fundamental challenges of extrapolating local observations to derive emissions at the scale of an entire city.

3. Compositions of local vs. regional fluxes

5.3 Compositions of local vs. regional CO₂ fluxes (Q3)

While the mean ffCO₂/CO₂ flux ratios were about 80 % in winter and 50 % in summer, the CO₂ concentration enhancements compared to marine background concentrations were in all three cities on average < 63 % in winter and < 28 % fossil in summer. This illustrates the locality of the urban flux footprint characterized by ffCO₂ emissions compared to the significantly larger continental concentration footprint, where biogenic fluxes dominate. A thorough selection of background stations is of great importance for the interpretation of urban CO₂ concentration enhancements.

Outlook

In the future, the REA measurements could be

6 Conclusions

Despite the limited representativity and comparatively large measurement uncertainties, the observation of substantial non-fossil CO₂ fluxes underlines the necessity of separating fossil and non-fossil CO₂ fluxes in cities. To maximize the number of high quality REA measurements, we recommend to clearly define the flagging criteria and research questions prior to the measurement campaign, specifying time periods, spatial directions, and micrometeorological conditions of interest. For this purpose, a near real-time metric for identifying measurements affected by storage fluxes could enable a more targeted selection or avoidance of such samples, depending on the scientific question at hand. To further increase the concentration differences between updraft and downdraft samples, and thereby the SNR, increasing the deadband width or measuring at lower heights below the inertial sublayer, closer to a particular source, could be considered. However, this would require a thorough evaluation of the statistical significance of the resulting shorter sample periods and the representativeness of the smaller footprints (see also Kunz et al., 2025a). For an independent validation of emission inventories, the REA measurements could be used for a 1:1 comparison with hourly bottom-up estimates or as input (with uncertainties) to inversion models. As shown by Stagakakis et al. (2023a), the assimilation of CO₂ flux observations from urban EC towers with very high spatiotemporal resolution information from urban bottom-up surface flux models has great potential to improve high-resolution bottom-up surface flux

770 ~~model estimates. In future REA campaigns, a near real-time metric for identifying cases of atmospheric decoupling in cities under unstable conditions during the day could enable a more targeted selection or avoidance of samples influenced by storage fluxes based on the scientific question at hand. A multispecies for model optimization. A multi-species analysis, including observations-MGA⁷ and flask measurements~~ of co-emitted species such as CO, could allow for further attribution of emission sources and estimation of a continuous ffCO₂ flux record (e.g., Maier et al., 2024b; Hilland et al., 2025; Juchem et al., 2025).

775 The extraordinarily large ffCO₂ and nfCO₂ fluxes observed from the directions of a district heating plant in Zurich and from a brewery in Munich show that to compare tall-tower measurements with bottom-up estimates or to integrate them into inversion models, inventory approaches should be able to represent large point-source emissions (both fossil and non-fossil) and their emission characteristics with high temporal resolution and three-dimensional spatial accuracy. It should also be noted that the EC method and flux footprint models rely on the assumption of stationary and horizontally homogeneous turbulent mixing.

780 However, large point-source emissions are often associated with buoyancy fluxes and plume rise. These inhomogeneities in the turbulent mixing limit the applicability of the EC method for adequately quantifying large point-source emissions. Since emissions from large power plants and industrial facilities are generally better known than those from residential buildings, traffic, and human respiration, for example, (Super et al., 2020), it should generally be attempted to exclude atmospheric measurements affected by large point sources by analyzing wind direction, times of day, or other proxies. To this end,

785 knowledge of the location and operating times of large emitters is essential. If the general urban mix is to be analyzed, a location without large point sources within the tower footprint should be selected, if possible.

Data availability. The raw data, the processed, quality-controlled fluxes, and the footprints used in this analysis are available from the ICOS Cities carbon portal <https://citydata.icoscp.eu/portal/>. Flags and comments on the individual REA measurements are provided in Kunz et al. (2025b).

790 **Appendix A: ffCO₂ estimates**

To estimate ffCO₂ concentrations, measured atmospheric $\Delta^{14}\text{C}$ (Δ notation according to Stuiver and Polach (1977)) and CO₂ concentrations are considered as the sum of a background (bg), a fossil fuel (ff), a biofuel (bf), a nuclear (nuc), a stratospheric (strato), a respiratory (resp), a photosynthetic (photo), and an oceanic (oc) component (Turnbull et al., 2016; Maier et al., 2023):

$$c_{\text{meas}} = \sum_i c_i \quad (\text{A1})$$

$$795 \quad c_{\text{meas}} \Delta^{14} = \sum_i c_i \Delta^{14}_i. \quad (\text{A2})$$

Here, $\Delta^{14}\text{C}$ has been abbreviated by Δ^{14} and $i = \text{bg, ff, bf, nuc, strato, resp, photo, oc}$. Although not all components from Eq. (A1) and Eq. (A2) are known, the budget equations allow, under certain assumptions, the calculation of ffCO₂ differences between updraft samples and downdraft samples from REA measurements as well as between individual measurements and

a background concentration. This section shows the equations and values used in this study, while detailed derivations and justifications of the assumptions can be found in the relevant literature.

A1 Concentration differences between updraft and downdraft REA samples

Combining Eq. (A1) and Eq. (A2) and assuming that REA sample pairs differ only in their fossil fuel, non-fossil emissions (biofuel and respiration), and photosynthesis components, the difference in c_{ff} between updraft and downdraft sample can be estimated via:

$$c_{\text{ff}}^{\uparrow} - c_{\text{ff}}^{\downarrow} = \frac{1}{\Delta_{\text{photo}}^{14} - \Delta_{\text{ff}}^{14}} \left[c_{\text{meas}}^{\uparrow} (\Delta_{\text{photo}}^{14} - \overline{\Delta_{\text{meas}}^{14 \uparrow}}) - c_{\text{meas}}^{\downarrow} (\Delta_{\text{photo}}^{14} - \overline{\Delta_{\text{meas}}^{14 \downarrow}}) + (c_{\text{nf}}^{\uparrow} - c_{\text{nf}}^{\downarrow}) (\Delta_{\text{nf}}^{14} - \Delta_{\text{photo}}^{14}) \right]. \quad (\text{A3})$$

We follow Maier et al. (2023) to account for the second-order effects of non-fossil $^{14}\text{CO}_2$ fluxes and assume that a) the $^{14}\text{CO}_2$ signature of photosynthetic fluxes equals the mean of the updraft and downdraft flasks, b) respiration fluxes are enriched by $25 \pm 12 \%$ compared to the mean atmospheric signature in the respective summer (July–September), and c) that the CO_2 concentration difference between updraft and downdraft flasks resulting from respiration and biofuels can be roughly accounted for with 5 ± 5 ppm as an upper limit. Table A1 shows the assumptions and values for $\Delta_{\text{photo}}^{14}$, Δ_{nf}^{14} , and Δc_{nf} used for the Zurich, Paris, and Munich measurements. Details and an analysis of the corresponding uncertainties can be found in Kunz et al. (2025a).

Table A1. Variables used to estimate $c_{\text{ff}}^{\uparrow} - c_{\text{ff}}^{\downarrow} = \Delta \text{ffCO}_2$. Δ_{i}^{14} denote the $\Delta^{14}\text{C}$ values of fossil fuels (ff), photosynthetic (photo) and non-fossil emissions (nf) CO_2 , and flask measurements (meas). $\overline{\Delta_{\text{meas}}^{14}} = 0.5 \cdot (\Delta_{\text{meas}}^{14 \uparrow} + \Delta_{\text{meas}}^{14 \downarrow})$ denotes the mean of the updraft and downdraft samples, which is different for each REA sampling. The atmospheric signature during CO_2 uptake of the biosphere $\overline{\Delta_{\text{atmo}}^{14}}$ is estimated by the mean $\Delta_{\text{meas}}^{14}$ value in summer (July to September 2022/2023/2024 in the case of the Zurich/Paris/Munich campaign). Also given are the specific values derived for the measurement campaigns in each city.

Variable	Unit	Approximation	Zurich value	Paris value	Munich value
Δ_{ff}^{14}	$\%$	-1000	-1000	-1000	-1000
$\Delta_{\text{photo}}^{14}$	$\%$	$\overline{\Delta_{\text{meas}}^{14}}$	$\overline{\Delta_{\text{meas}}^{14}} \pm 10$	$\overline{\Delta_{\text{meas}}^{14}} \pm 10$	$\overline{\Delta_{\text{meas}}^{14}} \pm 10$
Δ_{nf}^{14}	$\%$	$\overline{\Delta_{\text{atmo}}^{14}} + 25$	9 ± 16	14 ± 14	5 ± 14
$c_{\text{nf}}^{\uparrow} - c_{\text{nf}}^{\downarrow}$	ppm	$\sim \overline{\Delta \text{CO}_2}$	5 ± 5	5 ± 5	5 ± 5

A2 Concentration differences between REA flasks and a marine background station

Approximating $\Delta_{\text{photo}}^{14}$ by Δ_{bg}^{14} , the ffCO_2 concentration compared to clean background air can be calculated from Eq. (A1) and Eq. (A2) according to Maier et al. (2023):

$$c_{\text{ff}} = c_{\text{meas}} \cdot \frac{\Delta_{\text{bg}}^{14} - \Delta_{\text{meas}}^{14}}{\Delta_{\text{bg}}^{14} - \Delta_{\text{ff}}^{14}} + c_{\text{meas}} \cdot \frac{\Delta_{\text{nuc}}^{14}}{\Delta_{\text{bg}}^{14} - \Delta_{\text{ff}}^{14}} + c_{\text{resp}} \cdot \frac{\Delta_{\text{resp}}^{14} - \Delta_{\text{bg}}^{14}}{\Delta_{\text{bg}}^{14} - \Delta_{\text{ff}}^{14}} \quad (\text{A4})$$

As described in detail in Maier et al. (2023), the background concentrations can be estimated from measurements at the ICOS station Mace Head (MHD) on the western coast of Ireland. Nuclear contributions can be modeled using a dedicated Jupyter notebook from the ICOS Carbon Portal (<https://www.icos-cp.eu/data-services/tools/jupyter-notebook>, last access 20
 820 September 2025). Respiratory concentrations can be obtained using the Vegetation Photosynthesis and Respiration Model (VPRM, Mahadevan et al. (2008)) in combination with the Stochastic Time-Inverted Lagrangian Transport model (STILT, Lin et al. (2003)). However, STILT simulations require meteorological input fields, which are to date only available until the end of 2023. Therefore, the nuclear and respiratory corrections (last two terms in Eq. A4) were neglected in our analysis (Levin et al., 2003). Figure A1 compares the Zurich results with and without the corrections. The slopes of the linear regressions differ
 825 about 4 %. Part of this difference is due to the exclusion of one summer and one winter sample that could have been affected by a revision of a nuclear facility. For a qualitative comparison of local ffCO_2 REA fluxes and regional ffCO_2 concentration enhancements, however, the nuclear and respiratory corrections are considered negligible.

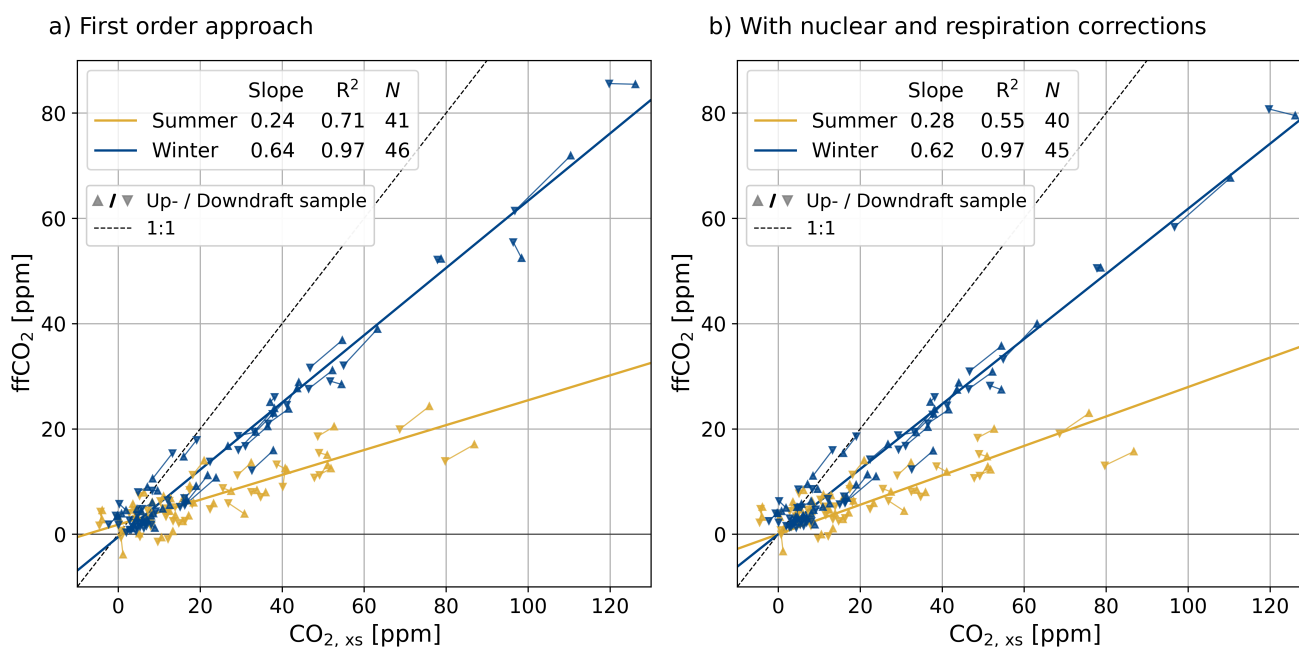


Figure A1. Comparison of concentration enhancements of the Zurich REA samples with respect to MHD (a) without corrections and (b) with corrections for nuclear contamination and ^{14}C -enriched respiration in the ffCO_2 estimation. R^2 is the coefficient of determination of the orthogonal regression, N the number of samples considered.

Appendix B: Flagging criteria for analyzed REA measurements

B1 Stationarity and well-developed turbulence

830 As with any turbulent trace gas flux measurement method, stationarity and well-developed turbulence are prerequisites for taking REA measurements (Rinne et al., 2021). We use the 0-1-2 quality control flagging scheme according to Mauder and Foken (2004), which labels "0" as high quality fluxes, "1" as medium quality fluxes, and "2" as poor quality fluxes, based on the steady state test and the developed turbulence test (Foken and Wichura, 1996). For the usually 60 min long REA measurements, the maximum of the 30 min EC averaging periods is considered.

835 B2 β coefficients

Figure B1 shows the CO₂ flux F_{CO_2} with respect to the product of the air density ρ , the standard deviation of the vertical wind velocity σ_w , and the CO₂ difference between updraft and downdraft flasks of all REA flask samples collected in Zurich, Paris, and Munich. The high correlation between the EC-based F_{CO_2} and the REA-based $\sigma_w \overline{\rho_m} \Delta \text{CO}_2$ shows the high quality of both measurement methods. According to Eq. (2), the slope of a linear fit corresponds to the β coefficient. If the vertical wind
840 velocity w were normally distributed and the regression on the CO₂ concentration were linear, β would depend only on the deadband width δ . Then all data points with the same δ would fall on a line with a slope of $\beta = 0.627$ for $\delta = 0$ and smaller slopes for larger δ (Grönholm et al., 2008). Deviations from this line indicate deviations from a Gaussian distribution. Since differences between individual measurements were found to be larger than differences between different scalars (Grönholm et al., 2008; Pattey et al., 1993), this is taken into account by calculating β for each sampling period individually according to
845 Eq. (2). However, Eq. (2) is unstable for ΔCO_2 close to zero, and $\beta < 0.1$ or $\beta > 1$ indicate non-ideal sampling conditions for REA measurements, e.g., due to skewness and kurtosis of the w time series or a linear drift leading to an unequal distribution of sampling times into the updraft and the downdraft reservoirs (Fotiadi et al., 2005a; Grönholm et al., 2008). Following Hensen et al. (2009) and Osterwalder et al. (2016), we only analyze measurements with $0.1 \leq \beta \leq 1$.

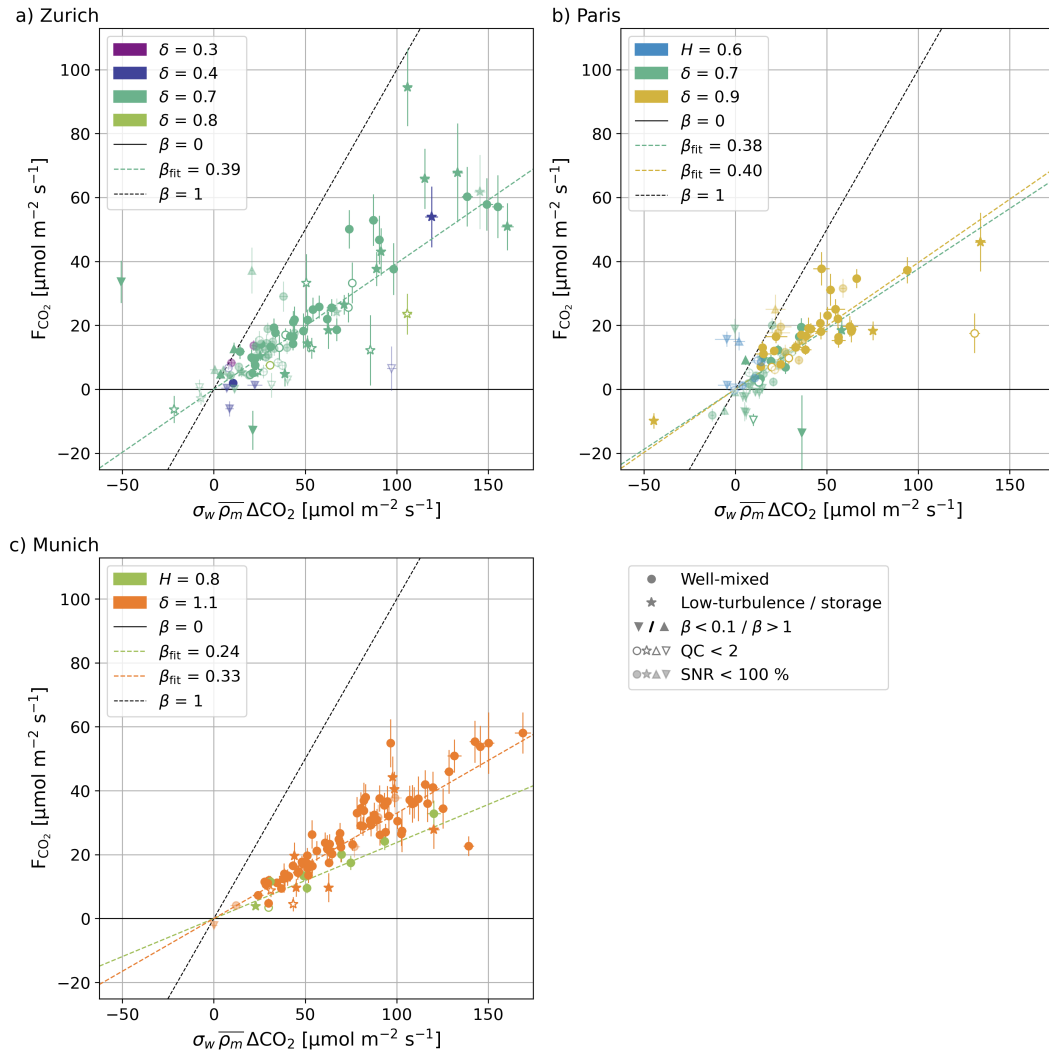


Figure B1. CO_2 flux F_{CO_2} vs. the standard deviation of the vertical wind velocity σ_w times the mean molar air density $\overline{\rho_m}$ and the CO_2 concentration difference between updraft and downdraft flasks of all REA flask samples collected in Zurich, Paris, and Munich. The colored dashed lines correspond to a linear regression of the well-mixed measurements (including measurements with $\text{SNR} < 100\%$, only if $N > 5$) with slope β_{fit} .

Table B1. β coefficients determined from the well-mixed measurements (including measurements with SNR < 100 %) from a linear regression of F_{CO_2} and ΔCO_2 (“Fit”, see Fig. B1) compared to the mean and standard deviation of the individually calculated β values (Eq. 2). In addition, the expected values for a normally distributed w and CO_2 timeseries are given (Fotiadi et al. (2005b), no value for $H = 0.8$ found in the literature). N denotes the number of samples considered.

Deadband width	City	N	Mean \pm std	Fit	Gauss
Linear (δ)					
0.7	Zurich	62	0.44 ± 0.14	0.39 ± 0.01	0.39
0.7	Paris	20	0.40 ± 0.20	0.38 ± 0.04	0.39
0.9	Paris	36	0.46 ± 0.17	0.40 ± 0.02	0.34
1.1	Munich	88	0.34 ± 0.07	0.33 ± 0.01	0.30
Hyperbolic (H)					
0.8	Munich	8	0.26 ± 0.06	0.24 ± 0.02	?

B3 Signal-to-noise ratio

850 The calculation of fluxes based on REA measurements from Eq. (1) requires that the concentration difference between updraft and downdraft samples is greater than the measurement uncertainty. Otherwise, it is unclear whether the flux was actually small or whether it was a measurement error (Fotiadi et al., 2005a). In our case of separating net CO_2 fluxes into fossil and non-fossil components, we consider the relative uncertainties of both ffCO_2 and nfCO_2 fluxes and discard samples only if both are > 100 %, otherwise the results would be biased toward large ffCO_2 fluxes. For this purpose, we define the signal-to-noise ratio (SNR) as the minimum of the relative uncertainties of the ffCO_2 and the nfCO_2 fluxes. Examples are shown in Fig. B2.

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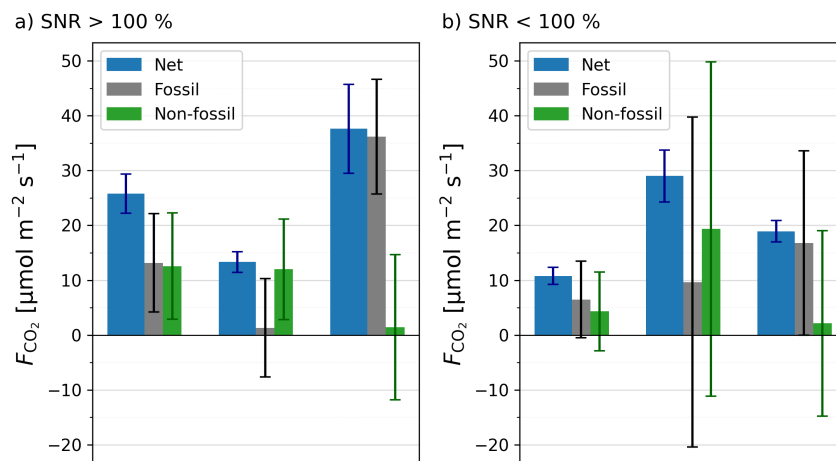


Figure B2. Three examples of REA measurements with signal-to-noise ratio SNR > 100 % (a) and SNR < 100 % (b). SNR is defined as the minimum of the relative uncertainties of the ffCO_2 and the nfCO_2 fluxes.

B4 Friction velocity and storage fluxes

During or after time periods of low turbulence, the measurement system may be decoupled from the surface so that the eddy flux is no longer representative of the local surface flux (Aubinet et al., 2012a). Instead, the measured flux will also contain non-turbulent flux components. These components can be caused by changes in storage below the measurement height or
860 by turbulence generated at elevated layers by high wind shear, for example (low-level jets (e.g., Prabha et al., 2007)). The composition of these non-turbulent fluxes will be largely determined by the surface fluxes prior to the measurement period. For example, the $\text{ffCO}_2/\text{CO}_2$ ratio of a storage flux during the break up of the nocturnal boundary layer in the morning will approximately reflect the $\text{ffCO}_2/\text{CO}_2$ ratio of the integrated nocturnal CO_2 emissions. Due to reduced anthropogenic emissions at night, this nocturnal ratio is assumed to be lower than the $\text{ffCO}_2/\text{CO}_2$ ratio of the surface fluxes during the measurement
865 period (e.g., morning rush hour). Consequently, the mean $\text{ffCO}_2/\text{CO}_2$ ratio of the integrated nocturnal CO_2 emissions is assumed to be smaller than the measured $\text{ffCO}_2/\text{CO}_2$ ratio. To identify the measurement periods of low turbulence and/or changes in the storage below the measurement height, we consider two quantities: the friction velocity u_* and the storage flux estimated from the EC measurements.

As the assumption of well-developed turbulence for EC is often not fulfilled during periods of low friction velocity, u_* is
870 commonly used as criterion to filter EC fluxes (Aubinet et al., 2012a). Although friction velocities tend to be greater in cities due to enhanced mechanical forcing, for example, the use of a u_* filter has also proved useful in many urban studies (e.g., Wu et al., 2022; Vogel et al., 2024; Hilland et al., 2025). As the continuous EC CO_2 fluxes showed a systematic decrease in flux magnitude at $u_* < 0.2 \text{ m s}^{-1}$ (Hilland et al., 2025), these periods were flagged. Since u_* also becomes small during strong convective events, the u_* criterion was only applied to periods with stability parameter $\zeta > -15.5$ (this threshold was chosen
875 based on the stability range where the footprint model by Kljun et al. (2015) is applicable).

Although storage fluxes, i.e., changes in the mean CO_2 concentrations within the air volume below the measurement height, are mostly negligible at higher u_* , a storage flux correction is usually applied and recommended for EC flux measurements (Crawford and Christen, 2014). Since we cannot apply such a storage correction to the REA ffCO_2 fluxes (see Sect. 2.2), measurements with large storage fluxes were flagged. The threshold was set to $|F_{\text{CO}_2, \text{strg}}| > 20 \text{ } \mu\text{mol m}^{-2} \text{ s}^{-1}$ (Fig. B3),
880 which is relatively large compared to the median value of the continuous EC measurements of about $3 \text{ } \mu\text{mol m}^{-2} \text{ s}^{-1}$. Due to the limited number of analyzed REA samples and the large uncertainties in storage flux estimation, only the most extreme measurements were flagged. The resulting uncertainties are discussed in the text, and the results are analyzed with respect to differences between measurements taken before and after 911:00 UTC-LT (for Munich only, due to the small number of measurements in Zurich and Paris, see Sect. 4.3.3).

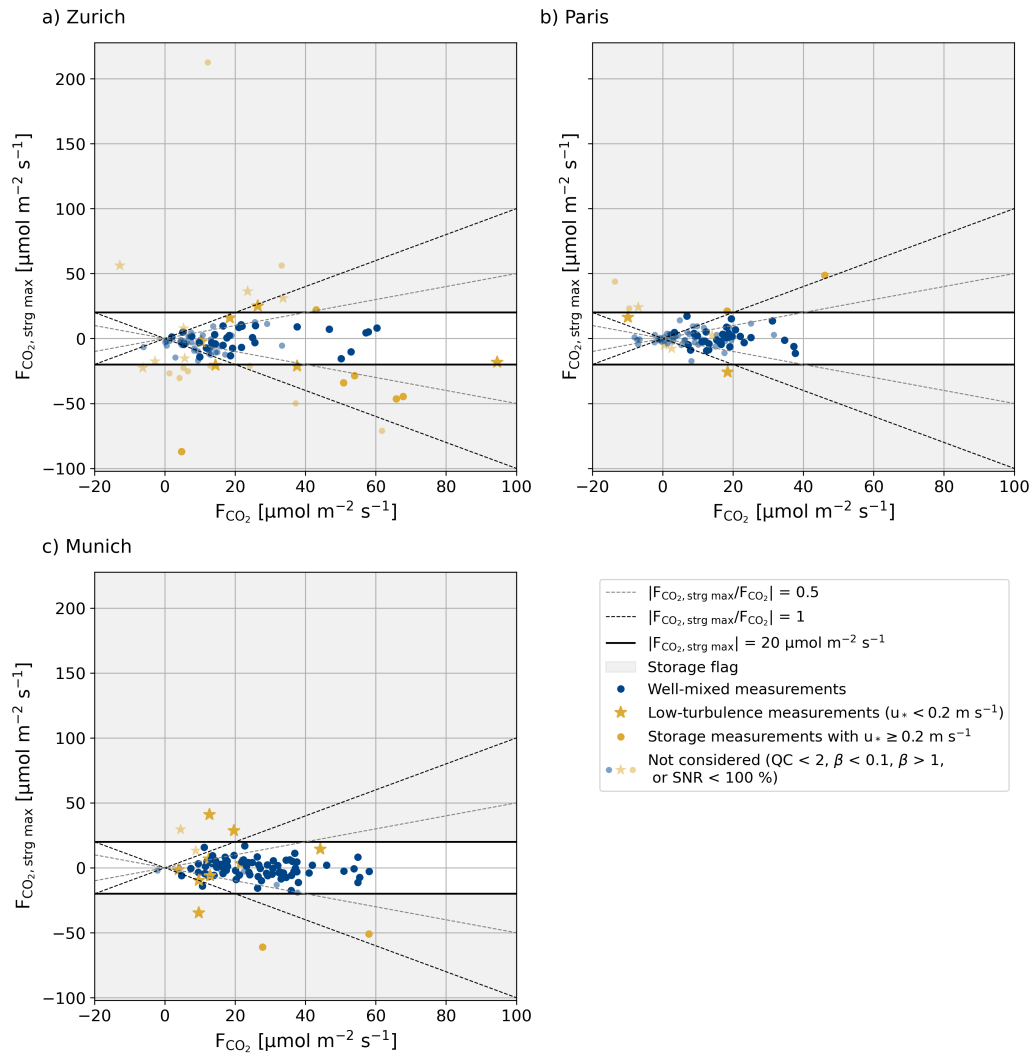


Figure B3. Maximum of the 30 min storage flux estimates with respect to the mean absolute CO_2 fluxes during the REA measurement periods. Measurements with $|F_{CO_2, \text{strg}}| > 20 \mu\text{mol m}^{-2} \text{s}^{-1}$ were flagged as storage measurements and analyzed together with the low-turbulence measurements ($u_* < 0.2 \text{ m s}^{-1}$).

a) Zurich



b) Paris



c) Munich



Figure C1. Photos of the measurement sites in Zurich, Paris, and Munich. The black arrows indicate the height at which the IRGASON and the gas inlets (two REA inlets with fast-response valves for updrafts and downdrafts, one inlet for REA quality control tests, one inlet for MGA⁷ measurements) were mounted. Pictures from [Roland Vogt \(University of Basel\)](#), Pekka Pelkonen (ICOS RI), Pedro Henrique Herig Coimbra (INRAE), and Reiter Antennenbau-Energietechnik GmbH.

Appendix D: EC measurements

890 Table D1 shows the time periods covered by IRGASON measurements, used for the presentation of median net CO₂ fluxes in Fig. 5. Considered were only the high and medium quality data with QC < 2, $u_* > 0.2 \text{ m s}^{-1}$, CO₂ signal strength > 90 %, and wind directions without flow distortion effects (compare Table 2). Due to later deployments and outages, the overall data coverage from MGA⁷ measurements is poorer. However, for most of the REA measurements, MGA⁷ data is available and used for the calculation of ffCO₂ fluxes (Sect. 2.2).

Table D1. Measurement periods and number (percentages) of retained measurements after quality control and filtering of the continuous CO₂ flux measurements from the IRGASON instrument.

	Summer measurements	Winter measurements
Zurich	14 July 2022 – 31 October 2022 $N = 2607$ (48.4 %)	1 November 2022 – 1 April 2023 $N = 3757$ (50.8 %)
Paris	1 July 2023 – 31 October 2023 $N = 3276$ (54.4 %)	1 November 2023 – 09 April 2024 $N = 4059$ (51.6 %)
Munich	1 July 2024 – 30 October 2024 $N = 2706$ (45.4 %)	1 November – 31 March 2025 $N = 2751$ (37.6 %)

895 Figure D1 shows the average normalized spectra and co-spectra of the IRGASON and the MGA⁷ against normalized frequency. For reference, the theoretical slopes according to Kaimal et al. (1972) are shown. In each city, the low-frequency, i.e., large energy-carrying eddies, are well captured by the MGA⁷. The slopes of the (co-) spectra agree well with the theoretical expectations of $-2/3$ and $-4/3$, respectively. For high normalized frequencies, the impact of spectral attenuation in the intake line of the MGA⁷ is visible. However, these losses were corrected by applying high frequency spectral corrections according to Fratini et al. (2012). The large contribution of high frequencies observed in the IRGASON data in Paris may be related to the proximity of a strong electromagnetic source (antenna), which may be the reason for the increase the observed IRGASON noise (personal correspondence with the manufacturer). Fortunately, the IRGASON noise was not correlated to the sonic velocity, 900 and was hence mostly filtered out by the covariance computation itself, which showed comparable $w'T'$ and $w'CO_2'$ spectra. Applying high frequency corrections according to Moncrieff et al. (1997), the CO₂ fluxes of the IRGASON and the MGA⁷ agreed well in all three cities ($r^2 > 0.89$, $RMSE < 2.5 \mu\text{mol m}^{-2} \text{ s}^{-1}$, mean bias $< 1.2 \mu\text{mol m}^{-2} \text{ s}^{-1}$). The median time lags between the raw high-frequency CO₂ measurements of the IRGASON and the MGA⁷ caused by clock drift and the travel time of the sample air to the MGA⁷ instrument (Sect. 2.1) was 4.15 s in Zurich, 10.45 s in Paris, and 37.30 s in Munich. The mean

905 [correlation between the two time lag corrected CO₂ concentration time series was 0.83 in Zurich, 0.76 in Paris, and 0.83 in Munich.](#)

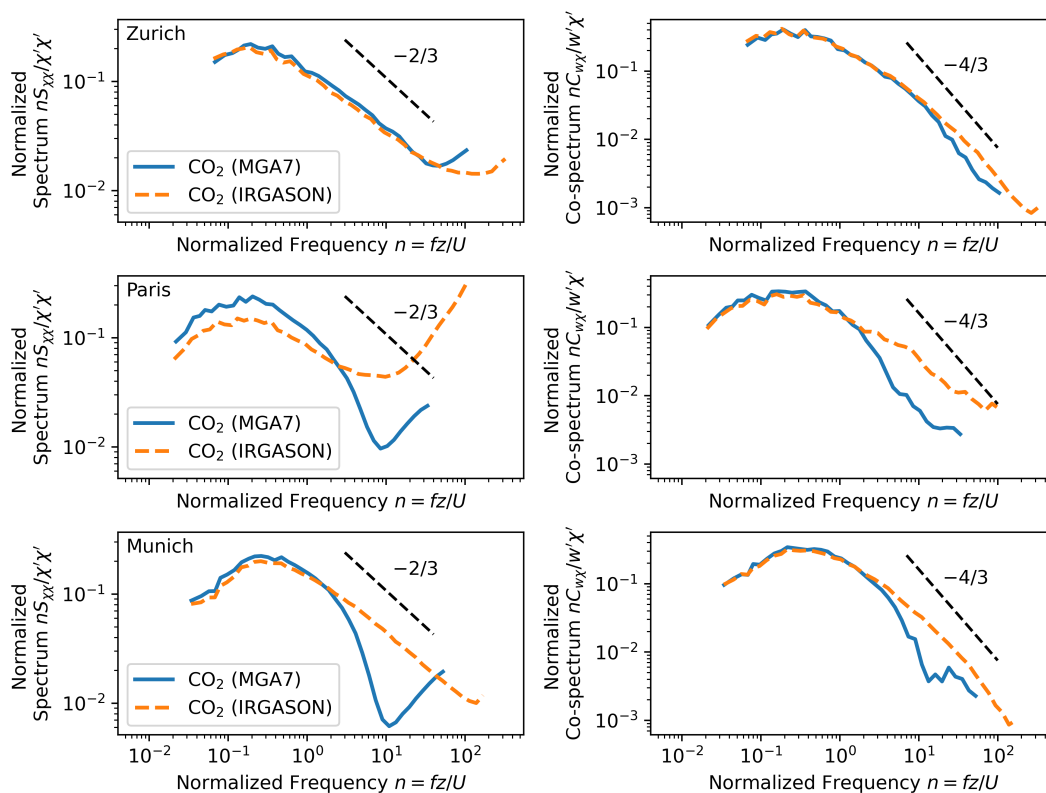


Figure D1. [Average normalized spectra and co-spectra against normalized frequency and theoretical slopes according to Kaimal et al. \(1972\)](#)

Appendix E: Quality control of the REA system

In all three cities, quality control tests conducted approximately once a month showed an overall good agreement between the CO₂ concentration of flasks sampled through the updraft and, in parallel, through the downdraft lines without switching of the valves (Table E1). The CO₂ difference between these quality control flask pairs and air samples collected simultaneously through a third intake line directly into the flask sampler was slightly higher (± 0.1 ppm on average). This difference can be partly attributed to the fact that with direct sampling, the weighting of the CO₂ concentration over the sampling period is not completely homogeneous, leading to larger deviations if the CO₂ concentration has a large variability or a trend (Levin et al., 2020; Kunz et al., 2025a). It can therefore be assumed that biases between updraft and downdraft sampling are negligible.

915 For the analyzed REA sample pairs, the measured CO₂ differences between updraft and downdraft samples agree well with the CO₂ difference estimates from both concurrent and continuous open-path IRGASON and closed-path MGA⁷ measure-

ments (Table E1). As discussed in Kunz et al. (2025a), a 0.2 ± 0.3 ppm difference between flask and IRGASON measurements in Zurich could be partly attributed to the fact that the IRGASON CO_2 dry molar fractions were derived from a CO_2 density output that does not properly account for high-frequency fluctuations in air temperature in the sensing path, because the ambient temperature measured by an EC100 slow-response temperature probe was used in the conversion of the absorption measurements to CO_2 density. Since 13 April 2024 (end of Paris measurements), an updated logger program records the CO_2 measurements using a fast-response temperature of the ultrasonic anemometer. In addition, as an open-path gas analyzer, the IRGASON is much more susceptible to weather conditions and obstructions in the path than the MGA⁷. The slightly smaller ΔCO_2 estimates from the MGA⁷ in Munich may be due to high-frequency attenuation caused by the long intake lines affecting the MGA⁷ (100 m vs. approximately 30 m in Zurich and Paris). Nevertheless, the overall good agreement between flask and in situ measurements indicates that the system was operating as intended and that uncertainties due to the sampling process are negligible. Due to the slightly better agreement between high-precision CO_2 flask measurements and the MGA⁷ than with the IRGASON, the MGA⁷ flux measurements were used when available; otherwise the fluxes calculated from the IRGASON data were used. Overall, the CO_2 flux estimates from both instruments agreed well (see Appendix D). As shown in Kunz et al. (2025a) for the Zurich measurements, the ffCO_2 flux uncertainties are dominated by the ^{14}C measurement precision.

Table E1. Means and standard deviations of the CO_2 differences between quality control flasks sampled without switching of the valves (all-valves-open tests) through the updraft ($\text{CO}_{2,\text{qc}\uparrow}$), the downdraft ($\text{CO}_{2,\text{qc}\downarrow}$) and a direct line ($\text{CO}_{2,\text{qc direct}}$). Furthermore, the CO_2 concentration ΔCO_2 between updraft and downdraft flasks collected during the actual REA measurements are compared to estimates from the 20 Hz in situ measurements of the IRGASON and the MGA⁷. For the latter, only IRGASON measurements with CO_2 signal strength > 90 % and only MGA⁷ measurements with good spectral fit of the CO_2 laser are considered.

	Zurich	Paris	Munich
All-valves-open tests			
$\text{CO}_{2,\text{qc}\uparrow} - \text{CO}_{2,\text{qc}\downarrow}$ [ppm]	-0.007 ± 0.023 ($N = 6$)	0.016 ± 0.026 ($N = 7$)	-0.016 ± 0.044 ($N = 11$)
$\text{CO}_{2,\text{qc}\downarrow} - \text{CO}_{2,\text{qc direct}}$ [ppm]	0.12 ± 0.14 ($N = 6$)	0.13 ± 0.37 ($N = 7$)	-0.14 ± 0.18 ($N = 11$)
Flask - in situ comparison			
$\Delta\text{CO}_{2,\text{flasks}} - \Delta\text{CO}_{2,\text{IRGASON}}$ [ppm]	0.21 ± 0.3 ($N = 85$)	0.07 ± 0.44 ($N = 55$)	0.23 ± 0.36 ($N = 92$)
$\Delta\text{CO}_{2,\text{flasks}} - \Delta\text{CO}_{2,\text{MGA}^7}$ [ppm]	0.01 ± 0.20 ($N = 64$)	-0.03 ± 0.26 ($N = 31$)	0.07 ± 0.24 ($N = 86$)

Appendix F: Point-source Identification of measurements potentially influenced by emissions from a district heating plant in Zurich and a brewery Munich

In Zurich, emissions from a district heating plant (natural gas) are likely to have influenced the REA measurements when the district heating plant was operating and within the peak area of the flux footprint (Sect. 3.1 and Sect. 4.3.2). To identify the potentially affected REA measurements, the flux contributions from a $40 \times 40 \text{ m}^2$ area centered around the chimney of the district heating plant were estimated based on the footprint model by Kljun et al. (2015) (Fig. F1 a, left). For this purpose, two 30 min footprints were averaged for each REA measurement. There were three measurements in which the modeled footprint was in the direction of the district heating plant, but the contribution from the considered area was zero due to the finite distance of the peak contribution from the measurement site and the immediate proximity of the district heating plant to the tower ($\sim 150 \text{ m}$) (Fig. F1 a, right). Since the footprint model does not account for the explicit height of the point emissions (chimney of $\sim 30 \text{ m}$) nor plume rise due to hot emissions, and since CO_2 spikes observed in the continuous concentration measurements indicate an influence from the point source, we assume that these three measurements could nevertheless have been influenced by the district heating plant. The operating times of the three burners of the district heating plant are known with a temporal resolution of 5 min.

Due to uncertainties and limitations in the footprint modeling, we also attempted to investigate the potential flux contribution from the district heating plant based on $^{13}\text{CO}_2$ observations. ^{13}C generally enables a distinction between CO_2 from natural gas, which is used in the district heating plant and has a ^{13}C signature (in δ notation) of about -40‰ (Tans, 1981; Widory and Javoy, 2003), and that from liquid and solid fuel or biogenic fluxes with $\delta^{13}\text{C} \approx -25\text{‰}$ (Tans, 1981; Widory and Javoy, 2003; Bakwin et al., 1998). However, the Zurich REA flasks were not analyzed for $^{13}\text{CO}_2$ by the high-precision ICOS Flask and Calibration Laboratory, but as a by-product of the ^{14}C extraction at the ICOS CRL, with an order of magnitude lower precision of about 0.2‰ . Thus, although the potentially influenced samples showed influence from an isotopically lighter source, this was not significant within the measurement uncertainties, and an unambiguous gas source attribution was not possible. A contribution from a gas source is therefore likely, but cannot be clearly attributed to individual measurements. Flagging of measurements with a potential contribution from emissions from the district heating plant was therefore based on footprint data alone.

Analogously, all Munich REA measurements in which the flux footprint contribution from the area where the brewery is located was $> 3.5\%$, were considered to be potentially influenced by emissions from the brewery (Fig. F1 b). In Munich, neither the operating times nor the exact location of the emission source is known.

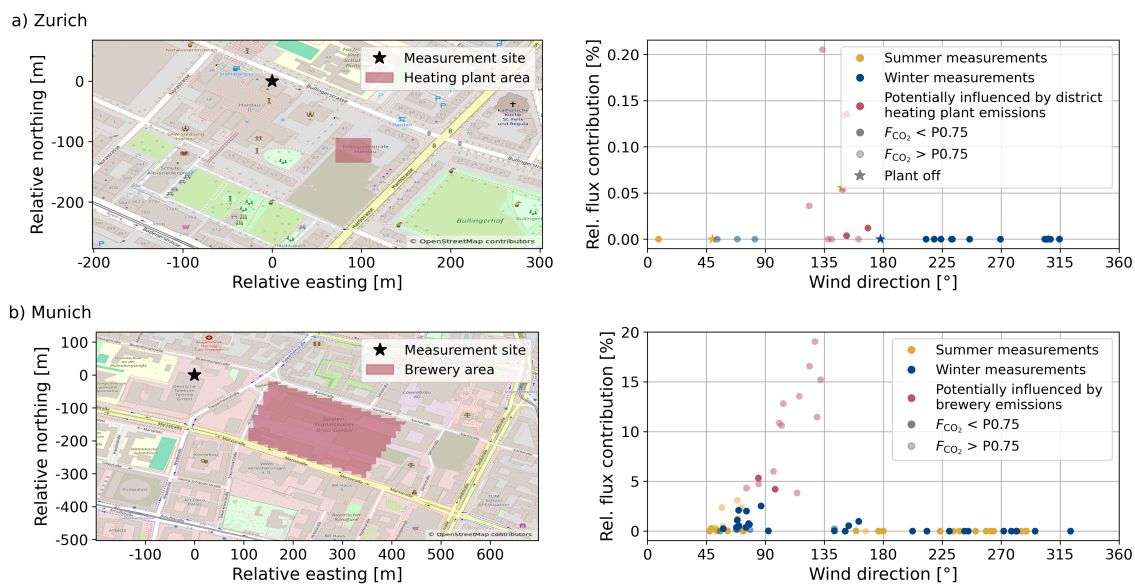


Figure F1. Relative flux contributions from the areas where the district heating plant (Zurich) and the brewery (Munich) are located based on the flux footprints of the well-mixed REA measurements. Map data from © OpenStreetMap contributors 2025. Distributed under the Open Data Commons Open Database License (ODbL) v1.0.

Appendix G: Spatial flux patterns in Paris

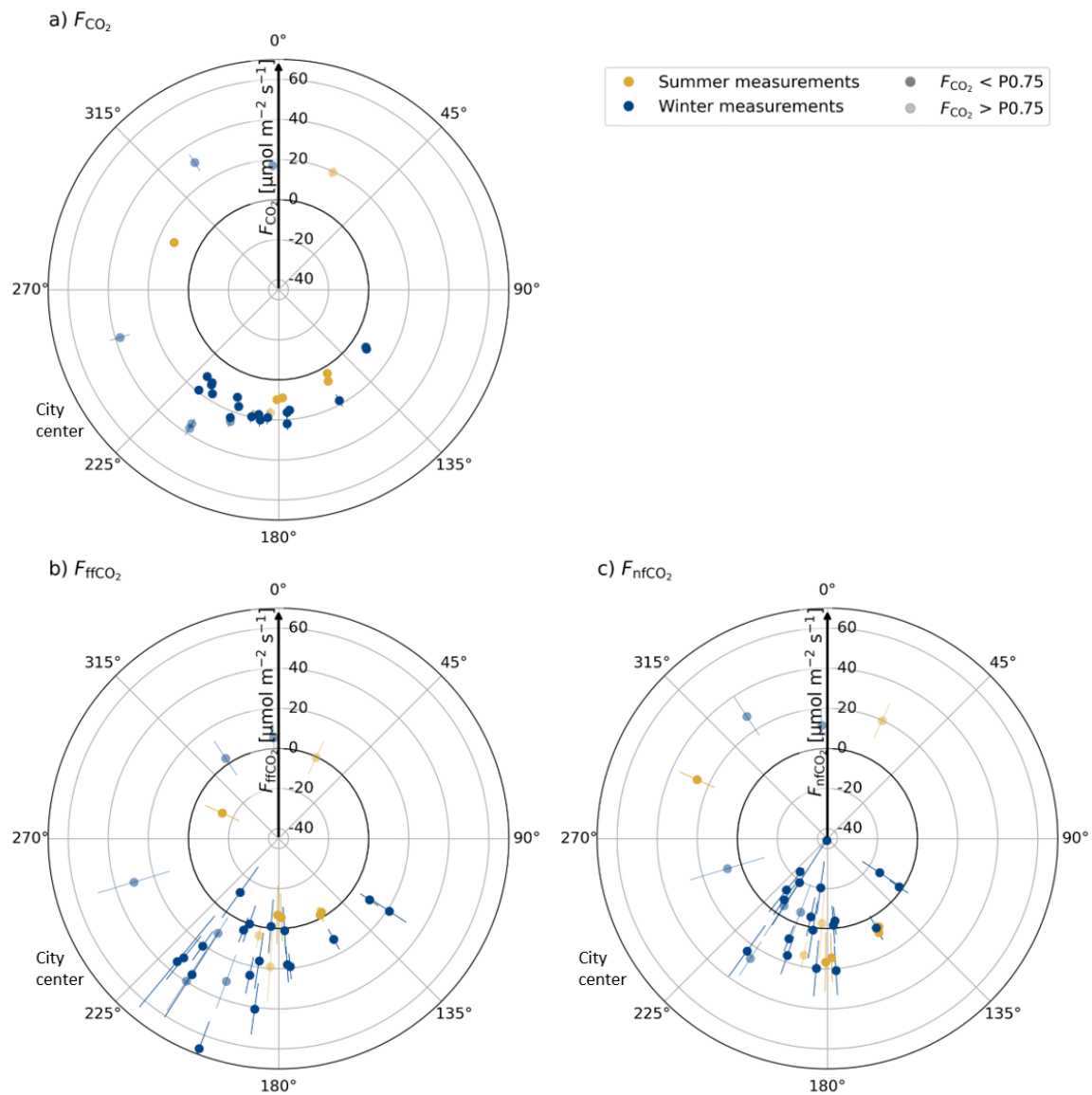


Figure G1. Net CO_2 fluxes (a), ffCO_2 fluxes (b), and nfCO_2 fluxes (c) with respect to the mean wind directions during the measurement intervals in Paris with well-mixed conditions. The error bars represent the respective flux uncertainties. P0.75 denotes the 75th percentile of the continuous EC CO_2 fluxes at the respective hour of the day. Indicated is also the direction of the city center.

Appendix H: Z-tests

960 Since the relatively large positive and negative nfCO_2 fluxes observed in Zurich and Paris for fluxes $< 30 \mu\text{mol m}^{-2} \text{s}^{-1}$ could not be sufficiently explained by temperature, radiation or other variables, it was investigated to what extent the results could be caused by measurement uncertainties alone and whether the available data sets show a significant difference to the naive assumption of purely fossil fluxes in the city. For this purpose, a z-test was used to calculate the probability of measuring the observed error-weighted mean $\text{ffCO}_2/\text{CO}_2$ ratios and mean nfCO_2 flux under the null hypotheses of entirely fossil fluxes, i.e.,

965 $\bar{R}_{\text{ffCO}_2} = 100 \%$ or $\bar{F}_{\text{nfCO}_2} = 0 \mu\text{mol m}^{-2} \text{s}^{-1}$, given the mean measurement uncertainties (Table 4). The null hypothesis was rejected if the p-value was less than the significance level of 0.05. Since the z-test assumes a normal distribution of the observed variables, measurements with ΔCO_2 less than the measurement uncertainty of about 0.04 ppm were excluded to avoid extreme values in the $\text{ffCO}_2/\text{CO}_2$ ratio. In addition, we determined the minimum effect, i.e., the minimum deviation from the null hypothesis that would be required to correctly reject the null hypothesis at a 0.05 significance level and 80 % power. Smaller

970 deviations from $\bar{R}_{\text{ffCO}_2} = 100 \%$ or $\bar{F}_{\text{nfCO}_2} = 0 \mu\text{mol m}^{-2} \text{s}^{-1}$ could not be detected with the given number of samples and measurement uncertainties. The number of samples required to detect an assumed difference in the mean $\text{ffCO}_2/\text{CO}_2$ ratio of 10 % or an assumed mean nfCO_2 flux of $3 \mu\text{mol m}^{-2} \text{s}^{-1}$ was also determined. Note that a constant $\text{ffCO}_2/\text{CO}_2$ ratio is not compatible with a constant nfCO_2 flux. However, both are possible conceptual models that are analyzed here. Table H1 and Table H2 show the results for the well-mixed measurements, divided into summer and winter measurements.

Table H1. Analysis of the $\text{ffCO}_2/\text{CO}_2$ ratios of the well-mixed measurements, excluded those likely influenced by [large](#) point source emissions and four measurements with $\Delta\text{CO}_2 < 0.4$ ppm. N denotes the number of measurements, \bar{R}_{ffCO_2} the error-weighted mean $\text{ffCO}_2/\text{CO}_2$ ratio, and $\delta\bar{R}_{\text{ffCO}_2}$ the mean measurement uncertainty of the ratios. The p-values describe the probabilities of observing the measured mean ratio under the assumption (null hypothesis) that $\bar{R}_{\text{ffCO}_2} = 100 \%$ and that deviations are solely due to measurement uncertainty. In addition, the minimum deviation from $\bar{R}_{\text{ffCO}_2} = 100 \%$ required to reject the null hypothesis at a significance level of 0.05 (minimum effect) and the number of samples required to detect a deviation from the null hypothesis of 10 % at significance level of 0.05 and a power of 80 % (N_{10}) is given.

Variable	Zurich		Paris		Munich	
	Summer	Winter	Summer	Winter	Summer	Winter
N	3	16	8	23	33	31
$\bar{R}_{\text{ffCO}_2} [\%]$	48 ± 52	92 ± 11	-7 ± 22	80 ± 10	47 ± 4	76 ± 4
$\delta\bar{R}_{\text{ffCO}_2} [\%]$	106	70	86	63	33	25
p-value	0.3	0.4	< 0.001	0.04	< 0.001	< 0.001
Minimum effect [%]	103	21	42	20	8	7
N_{10}	885	384	575	308	88	49

Table H2. Analysis of the nfCO₂ fluxes of the well-mixed measurements, \pm -excluded those likely influenced by [large](#) point source emissions and four measurements with $\Delta\text{CO}_2 < 0.4$ ppm. N denotes the number of measurements, \bar{F}_{nfCO_2} the error-weighted mean nfCO₂ flux, and $\overline{\delta F}_{\text{nfCO}_2}$ the mean measurement uncertainty of the nfCO₂ fluxes. The p-values describe the probabilities of observing the measured mean flux under the assumption (null hypothesis) that $\bar{F}_{\text{nfCO}_2} = 0$ and that deviations are solely due to measurement uncertainty. In addition, the minimum deviation from $F_{\text{nfCO}_2} = 0$ required to reject the null hypothesis at a significance level of 0.05 (minimum effect) and the number of samples required to detect a deviation from the null hypothesis of $3 \mu\text{mol m}^{-2} \text{s}^{-1}$ at significance level of 0.05 and a power of 80 % (N_3) is given.

Variable	Zurich		Paris		Munich	
	Summer	Winter	Summer	Winter	Summer	Winter
N	3	16	8	23	33	31
\bar{F}_{nfCO_2} [$\mu\text{mol m}^{-2} \text{s}^{-1}$]	0 ± 4	1.5 ± 2.7	9.7 ± 2.2	2.7 ± 2.1	7.8 ± 1.0	5.3 ± 1.1
$\overline{\delta F}_{\text{nfCO}_2}$ [$\mu\text{mol m}^{-2} \text{s}^{-1}$]	8.3	13.4	9.7	12.2	6.5	7.6
p-value	1.0	0.6	< 0.001	0.2	< 0.001	< 0.001
Minimum effect [$\mu\text{mol m}^{-2} \text{s}^{-1}$]	8.3	5.3	4.3	4.1	1.9	2.3
N_3	61	158	83	129	37	50

975 The results show that for the Paris summer measurements and for the Munich measurements the mean ffCO₂/CO₂ flux ratios were significantly different from 100 % (p-values < 0.05), with about 20 % non-fossil contribution in winter and 50 % (Munich) and 100 % (Paris) non-fossil contribution in summer. The small fossil component in Paris is surprising and not yet fully understood. In Zurich, no significant average nfCO₂ component was observed. While the small number of samples and the large measurement uncertainties in Zurich and Paris required a minimum non-fossil contribution of more than 20 % in winter and more than 40 % in summer to reject the null hypothesis / more than 300 measurements to detect a mean nfCO₂ contribution of 10 % at a power of 80 %, the minimum effect was reduced to 8 % in Munich and the required number of samples to about 90 summer measurements and 50 winter measurements, respectively. Similarly, the mean nfCO₂ fluxes were significantly different from zero for the Paris summer samples and the Munich samples. With the current setup, i.e., as in Munich, mean nfCO₂ fluxes of $3 \mu\text{mol m}^{-2} \text{s}^{-1}$ can be determined with about 40 to 50 measurements, which is close to the number of samples collected in this study.

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Author contributions. LBo, ME, RK, and VL designed and built the REA flask sampler, and helped with its installation, maintenance, and setup improvements during the three measurement campaigns. AC wrote the logger programs. AC and SH acquired funding and managed the project. LE, BL, MR, JC, CH, and MM managed the installations of the three measurement sites. SS, LBi, and CL oversaw the IRGASON measurements in Zurich, Paris, and Munich, respectively, and helped, e.g., with updates of the logger program. RH operated the MGA⁷ in all three cities, processed the EC data for the REA sample selection, and provided the final EC flux data used in this study. PA provided data from the midcost-sensors and helped with the installation in Munich. XG and JDC were responsible for the flask measurements at the ICOS Flask

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and Calibration Laboratory and the ICOS Central Radiocarbon Laboratory, AJ and SP for the corresponding data processing and quality control. BM produced the footprint and surface-cover analysis based on the model of NK. AK operated the REA system, performed the analysis, and wrote the manuscript, with conceptual and methodological input and supervision from AC, SH, and NK. All authors reviewed and contributed to the manuscript.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. The authors have received funding from ICOS Cities, a.k.a. the Pilot Applications in Urban Landscapes – Towards integrated city observatories for greenhouse gases (PAUL) project, from the European Union’s Horizon 2020 research and innovation program under grant agreement no. 101037319. [The TUM authors are partly supported by the ERC Consolidator Grant CoSense4Climate \(grant no. 101089203\).](#) Additional support was provided by internal funds and staff at the Universities of Heidelberg, Freiburg, and the Max Planck Institute for Biogeochemistry in Jena, [as well as by the Swedish Strategic Research Area “Modelling the Regional and Global Earth system”, MERGE, funded by the Swedish government.](#) Financial support from ICOS Switzerland (ICOS-CH) Phase 3 and Phase 4 (Swiss National Science Foundation, grants 20FI20_198227, 20FI-0_229655) is also acknowledged. We thank the following people for their contributions to this work: Felix Baab and Dirk Redepenning (University of Freiburg, Germany) for building the hardware at the REA inlet and logistics; Roland Vogt (University of Basel, Switzerland) and Carsten Jahn (KIT, Germany) for negotiations and installations at the Zurich and Munich sites; Pascal Rubli and Andrea Fischer (EMPA, Switzerland), Sophie Emberger (ETHZ, Switzerland), Sophie Bevini, Laura Bouillon, Ingrid Chanca, Lorna Foliot, Cécile Gaudry, and Guillaume Nief (LSCE, France), and Christian Becker and Klaus Kürzinger (TUM, Germany) for regular maintenance and logistics related to the REA flask sampler in Zurich, Paris, and Munich, respectively; Steffen Knabe and the entire staff of the ICOS Flask and Calibration Laboratory in Jena and the ICOS Central Radiocarbon Laboratory in Heidelberg for measuring the test and REA flasks; Hannes Juchem (University of Heidelberg) for providing CO₂ background concentration data from Mace Head; Matthias Zeeman (University of Freiburg, Germany) for managing the data infrastructure. [We would also like to thank Kenneth Davis and the anonymous reviewer for their comments and suggestions, which helped to improve the paper.](#) DeepL was used for grammar and spell checking.

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