

Trace Organic Gas Analyzer Time-of-Flight mass spectrometer (TOGA-TOF) system for airborne observations of formaldehyde

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41 **Abstract**

42 Formaldehyde (HCHO) is a ubiquitous atmospheric constituent, originating from primary
43 emissions (natural and anthropogenic) and secondary production via the oxidation of volatile
44 organic compounds (VOCs). In addition to being a regulated pollutant, HCHO is a key species
45 used as a tracer of recent photochemical activity due to its short atmospheric lifetime and its role
46 as a source of HO_x radicals. Given its diverse sources and high spatial variability, HCHO is
47 challenging to represent accurately in chemical transport models, often resulting in significant
48 discrepancies with observations. Airborne *in situ* measurements of HCHO, especially when
49 combined with VOC precursor data, offer valuable insights into its atmospheric distributions for
50 evaluating models. Here, we present HCHO observations from the NSF NCAR Trace Organic
51 Gas Analyzer with Time-of-Flight mass spectrometer (TOGA-TOF), deployed during the 2019
52 Fire Influence on Regional to Global Environments and Air Quality (FIREX-AQ) campaign.
53 While most HCHO instruments target at most a few selected species for measurement, the TOGA-
54 TOF employs a rapid gas chromatography-mass spectrometry (GC/MS) technique and provides
55 discrete VOC measurements—including >100 C₁–C₁₀ species—at a time resolution of less than
56 2 minutes. We compare TOGA-TOF HCHO data to measurements from three 1 Hz instruments
57 aboard the NASA DC-8: the Compact Atmospheric Multi-species Spectrometer (CAMS), the In
58 Situ Airborne Formaldehyde (ISAF) instrument, and a proton-transfer-reaction time-of-flight
59 mass spectrometer (PTR-ToF-MS). The wide dynamic range of observed HCHO concentrations
60 (from < 100 ppt to ~100 ppb) during FIREX-AQ enabled a robust intercomparison. TOGA-TOF
61 HCHO agreed well with CAMS (slope = 1.1), with similar agreement with the PTR-ToF-MS,
62 while larger discrepancies were observed with ISAF (slope = 1.5), likely due to differences in
63 calibrations. Normalized excess mixing ratios (NEMRs) of HCHO relative to CO in wildfire
64 plumes exhibited consistent trends with plume age across instruments. These findings highlight
65 the TOGA-TOF's capability for highly sensitive and accurate airborne HCHO measurements.

66 **1. Introduction**

67 Formaldehyde (HCHO) is one of the most ubiquitous non-methane volatile organic
68 compounds (VOCs) in the atmosphere. In addition to being an EPA regulated pollutant (National
69 Research Council (US) Committee on Toxicology, 1980), HCHO is a key reactive compound in

70 understanding recent photochemistry in the atmosphere. HCHO is an oxidation product of a wide
71 range of VOCs, has a relatively short lifetime, and is an important source of HO_x (OH + HO₂),
72 therefore affecting NO_x-VOC-O₃ chemistry (Seinfeld and Pandis, 2006). Globally, the dominant
73 source of HCHO is secondary production from VOC oxidation (Luecken et al., 2012; Stavrakou
74 et al., 2009). Tropospheric background levels of HCHO are maintained through methane (CH₄)
75 (Stavrakou et al., 2009) and isoprene (C₅H₈) (Curci et al., 2010; Luecken et al., 2018; Palmer et
76 al., 2003; Wolfe et al., 2016) oxidation. Direct emissions of HCHO can also be important locally;
77 sources include vehicle emissions (Possanzini et al., 2002; Sagebiel et al., 1996; Viskari et al.,
78 2000), biomass burning (Lee et al., 1997; Zhang et al., 2013), and fossil fuel industries (Fried et
79 al., 2020; Green et al., 2021; Pfister et al., 2019). In the atmosphere, HCHO is mainly lost through
80 photolysis, reaction with OH, and deposition, resulting in daytime atmospheric lifetimes of 1 to 9
81 hours (1–3 hours midday).

82 As an important proxy for understanding recent photochemistry, HCHO observations have
83 been used to indicate ozone sensitivity regimes (VOC- or NO_x-limited) (Duncan et al., 2010;
84 Martin et al., 2004; Nussbaumer et al., 2021; Schroeder et al., 2017; Souri et al., 2020, 2023) as
85 well as top-down estimations of isoprene (Cao et al., 2018; Kaiser et al., 2018; Palmer et al., 2003;
86 Wolfe et al., 2016; Zhu et al., 2016) and anthropogenic VOC (Cao et al., 2018; Kwon et al., 2021)
87 emissions. While satellite retrievals of HCHO column densities have been widely used in these
88 methods, challenges remain in capturing the spatial (vertical and horizontal) representation of
89 HCHO over the wide mixing ratio (MR) range observed in the troposphere; HCHO MRs can vary
90 less from than 20 parts per trillion (ppt) in remote regions, 1–20 parts per billion (ppb) in urban
91 environments, and > 100 ppb in fire plumes as observed from aircraft. These challenges in HCHO
92 satellite observations have been driven by uncertainties in retrieval methods (Anderson et al., 2017;

93 Liao et al., 2025; Souri et al., 2023; Stavrakou et al., 2009; Zhu et al., 2016), instrument sensitivity
94 (Souri et al., 2023), and coarse pixel sizes (Kwon et al., 2021; Souri et al., 2023). Uncertainties in
95 satellite observations increase in environments with low HCHO, such as regions with weak
96 biogenic or anthropogenic sources (Souri et al., 2023). Three-dimensional chemical transport
97 models may also fail to reproduce observed HCHO levels due to inaccurate emission inventories
98 of direct HCHO emissions or its VOC precursors (Green et al., 2021; Jaeglé et al., 2018; Luecken
99 et al., 2012, 2018; Warneke et al., 2007), uncertainties in chemical mechanisms (Luecken et al.,
100 2012), and coarse grid size. High quality *in situ* HCHO observations during airborne field
101 campaigns can provide useful information to validate satellite observations and evaluate chemical
102 transport models (Anderson et al., 2017; Chan Miller et al., 2017; Schroeder et al., 2017; Souri et
103 al., 2023; Zhu et al., 2016). For example, recent studies on HCHO intercomparisons between
104 satellite and *in situ* airborne observations showed that satellite HCHO retrievals can have a
105 systematic low bias of -20 to -51% (Anderson et al., 2017; Zhu et al., 2016), which can lead to
106 underestimation of VOC emissions. These systematic biases in satellite observations can be
107 corrected using high-quality airborne *in situ* observations of HCHO (Chan Miller et al., 2017; Zhu
108 et al., 2016, 2017).

109 Biomass burning produces large amounts of HCHO both from primary emissions and
110 secondary production. During a biomass burning event, a highly complex mixture of VOCs (Akagi
111 et al., 2011; Gilman et al., 2015; Gkatzelis et al., 2024; Koss et al., 2018; Travis et al., 2023;
112 Yokelson et al., 2013a) is produced from the fire source, which subsequently undergoes complex
113 chemical (i.e., reaction with oxidants and photolysis) and physical (i.e., dilution and cooling)
114 transformations downwind of the source (Decker et al., 2021; Hornbrook et al., 2011; Robinson et
115 al., 2021; Wang et al., 2021). For the recent joint NOAA/NASA Fire Influence on Regional to

116 Global Environments and Air Quality (FIREX-AQ) mission, Wang et al. (2021) used a high-
117 resolution large eddy simulation (LES) to show the chemical complexity of a fresh wildfire plume.
118 The levels and types of predominant oxidants, which are dependent on the historic exposure to
119 light, varies with the evolution of a plume, affecting the chemistry and fate of trace gases and
120 aerosols (Decker et al., 2021; Robinson et al., 2021; Wang et al., 2021). Liao et al. (2021) showed
121 that primary emissions of HCHO dominate near the fire source but secondary production from
122 VOC oxidation becomes important (up to ~ 60% after a few hours of physical age) as the fire ages.
123 As reported by Travis et al. (2023), the MRs of different types of VOCs emitted in crop and
124 prescribed fire plumes during FIREX-AQ depends on fuel types and fire states. The physical and
125 chemical complexities in fire plumes ensure that HCHO MRs are highly dynamic presenting
126 challenges for satellite retrievals (Stavrakou et al., 2009) and models to accurately represent local
127 and regional distributions of HCHO.

128 In this study, we present the NSF NCAR Trace Organic Gas Analyzer Time-of-Flight mass
129 spectrometer (TOGA-TOF) as a fast online GC/MS (gas chromatography-mass spectrometry)
130 technique for airborne observations of HCHO. Studies on improved techniques to detect HCHO
131 have been ongoing, attesting to its analytical challenges (Gilpin et al., 1997; Hopkins et al., 2003;
132 Hunter et al., 1999; Rice and Quay, 2006). While most currently available instrumental techniques
133 are specialized in detecting HCHO, the TOGA-TOF routinely measures over a hundred VOCs
134 using a non-reactive concentration technique, with the capability to analyze untargeted
135 compounds. Thus, TOGA-TOF is a unique airborne standalone instrument for the identification
136 and quantification of a wide range of C_1 – C_{10} VOCs, including formaldehyde, in complex ambient
137 air samples. The NOAA/NASA FIREX-AQ campaign (July to September 2019) was the first
138 campaign in which the TOGA-TOF was deployed for airborne measurements. During FIREX-AQ,

139 we had the opportunity to measure alongside two state-of-the-art HCHO measurement systems
140 aboard the NASA DC-8 during the entire FIREX-AQ campaign: the NASA In Situ Airborne
141 Formaldehyde (ISAF) instrument (Cazorla et al., 2015) and the University of Colorado Compact
142 Atmospheric Multi-species Spectrometer (CAMS) (Richter et al., 2015). The University of Oslo
143 Proton-Transfer Reaction Time-of-Flight Mass Spectrometer (UiO PTR-ToF-MS) (Müller et al.,
144 2014) measured HCHO for select time periods as well. For this study, we present the TOGA-TOF
145 HCHO data and focus on intercomparisons with CAMS and ISAF but also include comparisons
146 with the UiO PTR-ToF-MS. The FIREX-AQ dataset is an excellent testbed to compare a wide
147 range of HCHO mixing ratios (below 100 ppt to ~100 ppb) in a spatially variable ambient
148 atmosphere. The technique used in the TOGA-TOF could, in principle, be implemented in many
149 existing online GC/MS systems for semi-routine measurements of HCHO.

150 **2. Methods**

151 The sample preconcentration system in the TOGA-TOF is similar to the earlier generation
152 TOGA instrument with a quadrupole mass analyzer, which has been described in the literature
153 (Apel et al., 2015). The TOGA-TOF has been extensively characterized in the lab prior to and post
154 FIREX-AQ. Here, we describe the basic operation techniques of the TOGA-TOF system (**Sect.**
155 **2.1**), summarize the HCHO measurements during FIREX-AQ (**Sect. 2.2**), calculate the weighted
156 averages for direct comparisons between instruments with different sampling time resolution
157 (**Sect. 2.3**), and derive normalized excess mixing ratios (NEMR) of HCHO from selected plume
158 transects (**Sect. 2.4**).

159 **2.1 Description of the NSF NCAR Trace Organic Gas Analyzer with Time-of-Flight Mass**
160 **Spectrometer (TOGA-TOF)**

161 The TOGA-TOF hardware consists of a cryogenic preconcentration system, gas
162 chromatograph (GC), time-of-flight (TOF) mass analyzer (TOFWERK, Switzerland), electronics
163 box, clean air generator and on-board calibration system (CAG), and pump box. All major
164 components except the TOF mass analyzer are custom built. Typical preconcentration and GC
165 analysis sequences for the system are shown in **Fig. 1** and the flows for each sequence are shown
166 in **Fig. 2**. The GC analysis, including temperature ramp and cool and sample preconcentration
167 steps, occur simultaneously, with the full cycle duration of both processes time synced (**Fig. 1**) to
168 less than 2 min. During FIREX-AQ, the runtime for one full cycle was 105 seconds including 33
169 seconds of sample collection (i.e., “sampling”). The preconcentration system consists of three
170 cryogenically-cooled traps as described by Apel et al. (2015): (1) a water trap for removal of water
171 in ambient sampled air, (2) a sample trap for enrichment of targeted VOCs in ambient air, and (3)
172 a cryofocusing trap to refocus the analytes prior to injection into the GC. Corresponding to each
173 temperature setpoint described in **Fig. 1**, the cooling of each trap is achieved by metering the flow
174 of gas-phase N₂ generated from the headspace of the TOGA-TOF Dewar through coils at the
175 bottom of the dewar that were immersed in liquid N₂ (Apel et al., 2015). To heat the traps, the cold
176 N₂ gas flow was first shut off and followed by PID controlled flow of electricity, via a
177 WATLOW™ (Anafaze MLS300) temperature controller, through resistive wires wrapped around
178 the traps. The sample trap is packed with glass wool to allow greater surface area for the
179 enrichment of VOCs when cooling the trap to -140°C (**Fig. 1**), while allowing bulk gases such as
180 oxygen and nitrogen to pass through unretained. Both the water trap (inner diameter (I.D.) = 0.04
181 in.) and cryofocusing trap (I.D. = 0.021 in.) are open tubes.

182 For chromatography, a Restek MXT-624 column (I.D. = 0.18 mm, length = 8 m) is used
183 with helium as the carrier gas. The GC column has two temperature set points with temperatures
184 ranging between 35°C and 130°C (**Fig. 1**), also controlled by the Watlow, allowing gas-phase C₁
185 to C₁₀ VOCs to sequentially elute from the column. This allows the separation of VOCs, including
186 many structural isomers with the same chemical formula, based on their differences in interaction
187 with the stationary phase of the column. For detection of VOCs, the GC is coupled to a high-
188 resolution electron ionization time-of-flight mass spectrometer (HR EI-TOF-MS; TOFWERK),
189 using 70 eV ionization resulting in classical electron ionization (EI) spectra of the compounds.
190 This methodology allows routine quantification of well over 100 VOCs with the ability to further
191 analyze additional C₁ to C₁₀ compounds within the detectable range. Fluctuations in sensitivity are
192 corrected by normalizing with ambient levels of tetrachloromethane (CCl₄) based on the
193 recommendations in Karbiwnyk et al. (2003). Background corrections are made by system
194 generated ultra-pure helium (“He mode”) and zero-air blanks and then subtracting average blank
195 peak areas from ambient samples. For all surfaces that the sample contacts, Restek Sulfinert[®] steel
196 tubing is used to minimize interactions. Data processing is performed using Tofware (Aerodyne
197 Research, Inc., Billerica, MA) for high resolution peak fitting and TERN-in-IGOR (Claflin et al.,
198 2021; Isaacman-VanWertz et al., 2017; Lerner et al., 2017) (Aerodyne Research, Inc.) for
199 chromatographic peak analysis.

200 **2.2 Description of HCHO and CO measurements during FIREX-AQ**

201 In this section, we describe the TOGA-TOF HCHO measurements during FIREX-AQ
202 (July–September 2019). FIREX-AQ (Warneke et al., 2023;
203 <https://asdc.larc.nasa.gov/project/FIREX-AQ>) focused on the emissions of trace gases and
204 particles from wildland, prescribed, and agricultural fires in the United States. The NASA DC-8

205 flights during FIREX-AQ were based out of Boise, ID, mainly focusing on Western U.S. fires, and
206 Salina, KS, focusing on southeastern U.S. agricultural fires (flight tracks shown in **Fig. S1**). Details
207 on the field campaign may be found in Warneke et al. (2023).

208 As described in **Sect. 2.1**, the TOGA-TOF uses electron ionization for the detection of
209 VOCs. The EI of HCHO results in a spectrum of its parent ion (m/z 30) and major fragments (m/z
210 28 and 29) (**Fig. S2a**). **Figure S2b** shows the chromatographic signals of CH_2O^+ (m/z 30; exact
211 mass 30.010016 amu) and CHO^+ (m/z 29; exact mass 29.002191 amu) eluting at ~ 20 s. The high
212 resolution (HR)-TOF allows detection of highly resolved ions and the Tofware peak fitting
213 analysis shows separate parent and dominant fragments from other adjacent peaks (**Fig. S2c**).
214 HCHO calibration was carried out using a HCHO in N_2 gas-phase mixture contained in a treated
215 aluminum cylinder (Apel-Riemer Environmental, Inc.), diluted with ultra-high purity N_2 . The
216 stated HCHO concentration of the calibration mixture cylinder from the company was 1.69 ppmv.
217 The concentration was checked using three independent methods within a two-month period using:
218 (1) Fourier transform infrared spectroscopy (FTIR – NSF NCAR ACOM direct absorption using
219 the Beer-Lambert law, 1.55 ppmv), (2) Mid-IR – University of Colorado – also direct absorption
220 employing the CAMS instrument (1.65 ppm), and (3) an EPA 2,4-dinitrophenylhydrazine (DNPH)
221 cartridge method (1.64 ppmv). For methods (1) and (2), quantification of HCHO relies on the
222 absolute IR cross section and thus does not require an external standard. Therefore, these can be
223 regarded as fully independent reference methods. For the calibration of TOGA-TOF we used the
224 averaged value of 1.61 ppmv from the three quantification methods, and the resulting calibration
225 curve is shown in **Fig. S3**. The TOGA-TOF HCHO lower limit of detection (LLOD) during
226 FIREX-AQ was 20 ppt, and the HCHO measurement had an uncertainty of 35% based on the
227 variability of repeat calibrations using a 1.6 ppm (parts per million) HCHO in N_2 standard mixture

228 and a 90 ppb HCHO in N₂ standard mixture (both Apel-Reimer, Inc.), dynamically diluted to MRs
229 ranging between 0.15 and 100 ppb using (1) the CAG (i.e., catalytically-scrubbed ambient air), (2)
230 dry N₂, and (3) humidified N₂. No humidity dependence was observed.

231 HCHO mixing ratios from the TOGA-TOF were compared to 1 Hz HCHO measurements
232 from the ISAF, CAMS, and UiO PTR-ToF-MS instruments onboard the DC-8. The ISAF
233 instrument was operated by NASA and the details on the instrumental technique are described in
234 Cazorla et al. (2015). Briefly, the ISAF instrument uses a laser-induced fluorescence technique to
235 quantify HCHO using the fluorescence resulting from excitation of HCHO at 353 nm with a
236 tunable UV laser. Backgrounds were determined from the offline position 0.005 nm away from
237 the peak. Calibrations were made using a compressed HCHO gas cylinder (Air Liquide, 584 ppbv)
238 quantified by FTIR (Liao et al., 2021). ISAF was operated at a native sampling rate of 10 Hz and
239 archived at 10 Hz and 1 Hz. We used the 1 Hz data for analysis. The reported LLOD of HCHO
240 measured by ISAF is 30 ppt for the 1 Hz data at a signal/noise = 1 and the accuracy (systematic
241 uncertainty) of the ISAF measurements is estimated to be 10% + 10 ppt. More details on the ISAF
242 HCHO measurement during FIREX-AQ can be found in Liao et al. (2021).

243 The CAMS was operated by the University of Colorado and a detailed description of the
244 system can be found in the work of Richter et al. (2015). The CAMS uses a tunable IR laser source
245 for generation of mid-IR light for the detection of HCHO based on absorption. Background spectra
246 were collected during flights using zero air from a scrubber and were subtracted from the ambient
247 spectra. Calibrations were carried out by using a 4.9 ppm HCHO standard in N₂, quantified by
248 direct absorption using the Beer-Lambert law. The reported uncertainty of the FIREX-AQ HCHO
249 CAMS measurements is 6% and the campaign-averaged LLOD is 110 ppt for 1 Hz measurements
250 at a signal/noise = 1 during FIREX-AQ. However, during previous airborne missions (e.g.,

251 KORUS-AQ; Fried et al., 2020), LLODs ranging from 28 to 80 ppt have been reported. The
252 degraded performance during FIREX-AQ resulted from dirty multipass cell mirrors.

253 The UiO PTR-ToF-MS (Müller et al., 2014) was operated by the University of Oslo with
254 a focus on measuring NH₃ during FIREX-AQ, although HCHO and a number of additional VOCs
255 were reported for select flights (i.e., 25 July and 3 August). The PTR-ToF-MS uses hydronium
256 ions (H₃O⁺) generated from water vapor in a glow discharge-drift-tube to detect VOCs through
257 nondissociative proton transfer reactions. HCHO calibrations were carried out using a gas-phase
258 HCHO in N₂ cylinder (Apel-Riemer Environmental, Inc., 9.6 ppm HCHO; ± 5%). This calibration
259 was checked post-mission by generating test atmospheres with five different HCHO mixing ratios
260 (145–1035 ppb) in a 250 L laboratory chamber. The chamber was equipped with a White cell type
261 multiple reflection mirror system with a 120 m optical path length for online detection with a
262 Bruker IFS 66v/S FTIR instrument. The two methods were in excellent agreement, with the PTR-
263 ToF-MS reporting a volume mixing ratios less than 3% higher than the FTIR. During the two flight
264 days (25 July and 3 August) the UiO PTR-ToF-MS measured HCHO; the slope of the retrieved
265 ambient HCHO relative to that from CAMS was 1.03.

266 Normalized excess mixing ratios (NEMRs) of HCHO to CO are calculated as described in
267 **Sect. 2.4**. CO measurements were made onboard the DC-8 using the NOAA Los Gatos Research
268 (LGR) CO instrument and the NASA Differential Absorption Carbon monOxide Measurement
269 (DACOM) instruments. The LGR instrument (Bourgeois et al., 2022) uses an infrared laser off-
270 axis integrated-cavity-output spectroscopy technique with a time response of 1 Hz. The DACOM
271 instrument uses a differential absorption technique with an infrared tunable diode laser detecting
272 CO at the 4.7 μm wavelength, giving a data frequency of 5 Hz and measurement uncertainty of
273 2% (Sachse et al., 1987). The two CO instruments were in excellent agreement with the coefficient

274 of determination $r^2 = 0.97$, and a slope of 0.97 during the entirety of the campaign (**Fig. S4**), with
275 the reported LGR CO consistently 2.2% below that of DACOM.

276 **2.3 Averaging methods for comparison of HCHO between different instruments**

277 Data averaging across various sampling time periods is often carried out for direct
278 comparison between instrumental techniques with different sample integration times. **Figure 3**
279 shows HCHO measured by TOGA-TOF and CAMS with LGR CO measurements during plume
280 transects from the Castle Fire (**Fig. 3a**) and the Williams Flats Fire (**Fig. 3b**). As described in **Sect.**
281 **2.1**, during FIREX-AQ, the TOGA-TOF took discrete samples for 33 seconds in a 105 second
282 cycle period; therefore, subsections of each plume transect were sampled for analysis (**Fig. 3**). To
283 directly compare the discrete TOGA-TOF HCHO sampling data, CAMS, ISAF, and UiO PTR-
284 ToF-MS 1 Hz HCHO data are averaged over the time during which the TOGA-TOF sampling
285 occurred. Here we use two averaging methods for HCHO comparison: (1) normal averaging and
286 (2) volume-weighted averaging. The normal averaging is performed by taking the arithmetic mean
287 of the 1 Hz source data measured between the start and stop times of TOGA-TOF sampling. Below
288 LLOD values of the source data are included in the averaging. A minimum of at least one valid
289 (i.e., measured value or below LLOD) 1-Hz source data point during the TOGA-TOF sampling
290 period is allowed for the analysis to produce a non-missing merged value. This algorithm is how
291 the NASA online merging tool generates the TOGA.DC8_MRG and is publicly available on the
292 NASA Langley Research Center (LaRC) Airborne Science Data repository at [https://www-](https://www-air.larc.nasa.gov/cgi-bin/ArcView/firexaq?MERGE=1#TOGA.DC8_MRG/)
293 [air.larc.nasa.gov/cgi-bin/ArcView/firexaq?MERGE=1#TOGA.DC8_MRG/](https://www-air.larc.nasa.gov/cgi-bin/ArcView/firexaq?MERGE=1#TOGA.DC8_MRG/). More details on the
294 NASA online merging tool can be found in [https://www-](https://www-air.larc.nasa.gov/missions/etc/onlinemergedoc.pdf)
295 [air.larc.nasa.gov/missions/etc/onlinemergedoc.pdf](https://www-air.larc.nasa.gov/missions/etc/onlinemergedoc.pdf).

296 The second averaging method is the volume-weighted averaging method based on the
297 volume of air sampled. A post-mission laboratory analysis of the normalized volume vs. sample
298 collection duration demonstrated that the flow rate through the sample trap was not constant during
299 the sampling time period, despite the mass flow controller (MFC) being set to a fixed value.
300 Specifically, relative flow rates were higher during the initial seconds of sampling before
301 decreasing to the MFC setpoint. Through a series of experiments, we quantified the total sample
302 volume as a function of collection time and derived an empirical relationship to determine the
303 sample-volume midpoint, defined as the time at which 50% of the total collected volume had been
304 accumulated during the sampling cycle (**Fig. 3**). Midpoints were calculated for each sample using
305 this method. The midpoints determined through this method are reported in the TOGA-TOF
306 FIREX-AQ data files archived at the NASA Atmospheric Science Data Center (ASDC)
307 (https://doi.org/10.5067/ASDC/FIREXAQ_TraceGas_AircraftInSitu_DC8_Data_1). The
308 arithmetic mean of the 1 Hz data is thus calculated for the time periods between the start to
309 midpoint (T_1) and midpoint to stop times (T_2), and these two means are averaged to weight the
310 merge value on the sampled volume.

311 When sampling near strong VOC sources (e.g., wildfires or urban emissions), air masses
312 often have significant heterogeneity of HCHO levels from varying direct emissions and
313 photochemical production from rapidly changing VOC precursors. In those cases, using the non-
314 weighted averages of 1 Hz measurements can lead to significant biases when they are directly
315 compared to the TOGA-TOF data. Therefore, using volume-weighted averages of higher-
316 frequency data that factor in the true sample volume midpoints for comparisons with TOGA-TOF
317 data allows a more accurate comparison of the sampled air masses.

318 Using the above methodology, volume-weighted averages of LGR and DACOM 1 Hz CO
319 data are carried out, during the times when TOGA was sampling. In our study, we use the volume-
320 weighted averaged LGR CO data when it is available, and for times when it is not available, we
321 use the volume-weighted averaged DACOM CO data multiplied by 0.978 to account for the
322 relative differences in the two CO datasets. We normalized to the LGR CO data as it is a little
323 lower to be conservative in our analyses, although the two CO datasets are in excellent agreement
324 (**Fig. S4**) and within the combined measurement uncertainties (LGR = 2 ppb + 2% and DACOM
325 = $\pm 2\%$).

326 **2.4 Formaldehyde Normalized Excess Mixing Ratios**

327 We compare HCHO normalized excess mixing ratios (NEMRs) for Western U.S. wildland
328 fires calculated using HCHO reported by TOGA-TOF, ISAF, and CAMS. There are several ways
329 to derive NEMRs (Yokelson et al., 2013a) and here we used the slope method, which involves
330 determining the slope of HCHO in ppt against CO in ppb for a single plume transect and the
331 background air outside the plume. CO was used as the reference species to normalize for dilution
332 as it is a conserved biomass burning tracer in wildfire plumes (Müller et al., 2016; Selimovic et
333 al., 2019; Yokelson et al., 2013a). For the NEMRs from CAMS and ISAF, we use the 1 Hz HCHO
334 and CO data for the entire plume transect. For the TOGA-TOF NEMRs, we use the discrete HCHO
335 sampling points, which range between two and ten samples for each plume transect — including
336 a background sample and plotted these against the volume-weighted average (described in **Sect.**
337 **2.3**) TOGA merged CO for corresponding sampling periods.

338 The NEMR analysis was carried out for the plumes selected based on the following criteria:
339 a) plume transects perpendicular to the fire source, b) availability of at least one full TOGA-TOF
340 sampling cycle within the plume, c) stable background HCHO prior to and after the plume transect,

341 and d) availability of both CAMS and ISAF 1 Hz HCHO data sets and CO for both background
342 and in-plume sampling. Fire plumes were identified via plume flags using above background CO
343 in addition to visual inspection of the data (FIRE_FLAG_TABULAR_DATA.dat, [https://www-](https://www-air.larc.nasa.gov/cgi-bin/ArcView/firexaq#SCHWARZ.JOSHUA/)
344 [air.larc.nasa.gov/cgi-bin/ArcView/firexaq#SCHWARZ.JOSHUA/](https://www-air.larc.nasa.gov/cgi-bin/ArcView/firexaq#SCHWARZ.JOSHUA/)). Based on these criteria,
345 NEMRs were calculated for a total of 86 plume transects from 7 fire sources in the Western U.S.
346 and these are summarized in the supporting information (**Table S1**).

347 **3. Results and Discussion**

348 **3.1 Comparison of HCHO between different instrumental techniques**

349 Direct intercomparisons of measured HCHO between TOGA-TOF and the 1 Hz
350 instruments were carried out for data collected during all flights with available HCHO observations
351 onboard the DC-8. This included 13 flights sampling Western U.S. fires, 7 flights in the
352 Southeastern U.S., and 2 transit flights (**Fig. S1**). The dataset allowed comparisons of the widest
353 possible range of HCHO observed in the troposphere (below 100 ppt to ~ 100 ppb). **Figure 4**
354 shows correlations between TOGA-TOF, CAMS, and ISAF HCHO, averaged during the
355 corresponding TOGA-TOF sampling periods. Comparisons with normal averaged 1 Hz HCHO
356 (**Fig. 4a** and **b**) and volumed-weighted averaged 1 Hz HCHO (**Fig. 4c** and **b**), described in **Sect.**
357 **2.3**, are both shown. The data points are colored according to the relative standard deviation of
358 HCHO, calculated as the standard deviation of 1 Hz HCHO data during the TOGA-TOF sampling
359 period divided by the HCHO measured by TOGA-TOF. This relative standard deviation represents
360 the variability of the 1 Hz HCHO data during the TOGA-TOF sampling period.

361 Both non-weighted (**Fig. 4a**) and volume-weighted (**Fig. 4c**) averages showed that HCHO
362 measured by CAMS was within the 35% measurement uncertainty of the TOGA-TOF while the
363 comparisons with ISAF (**Fig. 4b** and **d**) were outside of the range of measurement uncertainties.

364 When using the non-weighted averaging method (**Fig. 4a** and **b**), the bivariate fitting (orthogonal
365 distance regression, ODR) of the TOGA-TOF and the 1 Hz HCHO gave a slope of 1.2 for CAMS
366 (y -intercept -177.4 ppt, **Fig. 4a**) and 1.6 for ISAF (y -intercept -210.2 ppt, **Fig. 4b**). The y -axis
367 residuals from the ODR fit were largest for data points with higher relative standard deviations
368 (i.e., > 0.7) of HCHO. For the non-weighted averages, the absolute values of the y -axis residuals
369 ranged from 0.6 ppt to 59.0 ppb for CAMS (mean \pm standard deviation = 1.0 ± 3.0 ppb) and 0.1
370 ppt to 68.1 ppb for ISAF (0.9 ± 3.2 ppb). To differentiate the dependence on HCHO variations
371 during each TOGA-TOF sampling, a relative standard deviation of 0.7 was used as a cut-off
372 threshold. For data points with relative standard deviations > 0.7 , residuals ranged from 2.6 ppt to
373 59.0 ppb (2.7 ± 6.8 ppb) for CAMS and 52.1 ppt to 46.8 ppb (3.7 ± 7.8 ppb) for ISAF. In
374 comparison, for less heterogeneous air masses (i.e., TOGA-merge 1 Hz HCHO with relative
375 standard deviations < 0.7), residuals ranged from 0.6 ppt to 44.1 ppb (0.7 ± 1.8 ppb) for CAMS
376 and from 0.1 ppt to 68.1 ppb (0.6 ± 2.5 ppb) for ISAF. Therefore, the data points with less variable
377 HCHO (i.e., HCHO relative standard deviation < 0.7) showed on average a factor of 4–6 lower
378 residual values, showing that the increased heterogeneity of HCHO in the sampled air masses lead
379 to poorer correlations between TOGA-TOF and the 1 Hz HCHO instruments.

380 For more accurate comparisons, the CAMS and ISAF HCHO data averages were weighted
381 by the TOGA-TOF sampling volume, factoring in the 50% volume sampling midpoint. Weighted
382 averages of the CAMS and ISAF measured HCHO improved the correlations with TOGA-TOF to
383 give an r^2 of 0.97 (compared to 0.93 for non-weighted averages) for CAMS and 0.96 (compared
384 to 0.91 for non-weighted averages) for ISAF (**Fig. 4c** and **d**). The residuals of the volume-weighted
385 averages were 30% lower for CAMS (range 0.2 ppt to 32.2 ppb and average 0.7 ± 2.0 ppb) and
386 33% lower for ISAF (range 0.03 ppt to 32 ppb and average 0.6 ± 2.0 ppb). Biases at low mixing

387 ratios significantly improved when using volume-weighted averages with absolute values of the
388 y -intercepts much closer to 0 than the non-weighted averages, at -14.4 ± 42.9 ppt for CAMS
389 (versus -177.4 ± 56.4 ppt) and 9.7 ± 36.8 ppt for ISAF (versus -210.2 ± 59.3 ppt). Comparisons
390 between TOGA-TOF and the UiO PTR-ToF-MS (**Fig. S5**), showed similar results with better
391 agreement when using volume-weighted averages.

392 The discrepancies of HCHO between different measurement techniques during FIREX-
393 AQ are similar to those shown by Liao et al. (2021), in which a comparison of CAMS vs. ISAF
394 HCHO data had a slope of 1.27. In our study, a comparison of the volume-weighted TOGA-merge
395 CAMS vs. ISAF HCHO from FIREX-AQ has a slope of 1.28, similar to Liao et al. (2021). The
396 minor difference in the slopes is most likely due to the different periods of data selected for analysis
397 as we only included CAMS and ISAF data during TOGA sampling in our analysis. Liao et al.
398 (2021) suggested that the main discrepancies between CAMS and ISAF were driven by the
399 uncertainties in the absolute calibration of the HCHO standards. As mentioned in **Sect. 2.2**, each
400 HCHO measurement technique has been carefully calibrated with HCHO standards. The HCHO
401 calibration standards used by each group were gravimetrically prepared and the absolute
402 concentrations were quantified using optical absorption techniques. As the main purpose of this
403 paper is to validate the TOGA-TOF as a unique technique for measuring HCHO, we will not
404 further explore the reasons for the discrepancies between the instruments.

405 **3.2 Comparison of HCHO NEMR between different instrumental techniques**

406 Normalized excess mixing ratios (NEMR) are used to derive the enhancement of a trace
407 gas within a smoke plume (Yokelson et al., 2013a). **Figure 5** shows HCHO NEMRs derived from
408 Western U.S. fires using the TOGA-TOF, CAMS, and ISAF instrumental techniques with respect
409 to smoke plume physical age (<https://www-air.larc.nasa.gov/cgi->

410 [bin/ArcView/firexaq#SCHWARZ.JOSHUA/](#)). Based on the instrument uncertainties of both CO
411 and the individual HCHO measurements, the uncertainties of the NEMRs are 35%, 6%, and 10%
412 for TOGA-TOF, CAMS, and ISAF, respectively. The HCHO NEMRs observed during FIREX-
413 AQ from TOGA-TOF were in the range of 12 to 30 ppt ppb⁻¹ CO. These NEMRs were within the
414 wide range observed from previous wildfire biomass burning studies from 2.2 to 46 ppt ppb⁻¹ CO
415 (e.g., Hornbrook et al., 2011 and references therein) and during prescribed burning events in the
416 Southeastern U.S. during FIREX-AQ that ranged from 16 to 29 ppt ppb⁻¹ CO (Travis et al., 2023).
417 The quadratic polynomial fitting of HCHO NEMRs from all three instruments showed that the
418 NEMRs increased up to ~ 5 h of plume physical age then declined. The age at which the HCHO
419 NEMRs from each instrument begin to decrease (i.e., the change in the NEMRs with respect to
420 plume age = 0) are all within 20 minutes of 5 h. This trend with respect to plume physical age
421 agrees with what Liao et al. (2021) reported for HCHO NEMRs measured by ISAF from individual
422 fire sources and is driven by the balance of production of HCHO, mainly from VOC oxidation by
423 OH, and loss of HCHO, mainly from photolysis, throughout the aging of plume (Liao et al., 2021).

424 While the trend with respect to plume physical age was similar using HCHO from the
425 different instruments, systematic differences were observed. To further investigate the possibility
426 of systematic differences of the NEMRs derived from the three HCHO instruments, paired *t*-tests
427 were used as a statistical method to compare two different analytical methods (Yen et al., 2020).
428 Here we carried out a paired *t*-test on the HCHO NEMRs calculated from the HCHO
429 measurements from the TOGA-TOF against CAMS and ISAF data. The sample number of the
430 analysis was equivalent to the number of selected transects for NEMR analysis (86 pairs). The null
431 hypothesis was whether the HCHO NEMRs from a pair of instruments are the same ($H_0: \mu_1 = \mu_2$),
432 where the absolute value of the *t*-value (*t*) higher than the critical *t* (T_c) rejects the hypothesis.

433 Between the TOGA-TOF and CAMS NEMRs, the analysis yielded $t = -0.42$, $T_c = 1.99$, and p -
434 value = 0.68, retaining the hypothesis (at the $p = 0.05$ (5%) level). Between the TOGA-TOF and
435 ISAF NEMRs, the analysis yielded $t = 11.53$, $T_c = 1.99$, and p -value = 4.68×10^{-19} , rejecting the
436 hypothesis. Between the CAMS and ISAF NEMRs, the analysis yielded $t = 29.84$, $T_c = 1.99$, and
437 p -value = 8.21×10^{-47} , again rejecting the hypothesis. Therefore, both TOGA-TOF and CAMS
438 derived NEMRs had a systematic difference from the ISAF derived NEMRs with a significance at
439 the $p = 0.05$ (5%) level.

440 As mentioned in **Sect. 3.1**, systematic biases exist between the HCHO measurement
441 techniques which are likely due to calibration standard differences. As illustrated in **Fig. 5**, these
442 biases are similarly present in the derived HCHO NEMRs. Nonetheless, comparisons between the
443 HCHO NEMRs derived from TOGA-TOF and CAMS (**Fig. 6a**) and TOGA-TOF and ISAF (**Fig.**
444 **6b**) indicate that the correlations are within the uncertainty bounds of the NEMR.

445 **3.3 Discussion on TOGA-TOF HCHO NEMRs and their applicability**

446 The applicability of TOGA-TOF-derived HCHO NEMRs is further explored for different
447 smoke plume conditions. **Figure 7** shows the difference of HCHO NEMRs between TOGA-TOF
448 and CAMS with respect to the averages from the two methods. Similar to **Fig. 5**, **Fig. 7b** shows
449 that the averaged HCHO NEMRs increase with plume physical age. With greater plume physical
450 age, the bias between TOGA-TOF driven HCHO NEMRs to CAMS becomes lower (**Fig. 7b**) and
451 this lower bias corresponds to more TOGA-TOF samples at each plume transect (**Fig. 7a**). With
452 increasing physical age, the smoke plumes studied were generally more dispersed and spatially
453 wider, therefore allowing more TOGA-TOF samples per transect.

454 The reduced bias between the two instrumental techniques with plume physical age can be
455 explained from the reduced plume heterogeneity with aging, as indicated by the improved r^2 of
456 each plume sampling with plume physical age (**Fig. 7c**). Modeling studies of fire plumes
457 (Trentmann et al., 2003a, 2003b; Wang et al., 2021) have shown significant spatial heterogeneity,
458 both vertically and horizontally, of fire plumes with different levels of oxidants (i.e., OH, O₃, NO₃).
459 This plume heterogeneity has also been observed in power plant plumes (Brock et al., 2002). As
460 discussed in Wang et al. (2021), at the early stage of a fire plume, OH levels are severely dampened
461 at the core of the plume, mainly due to attenuation of solar radiation from aerosols, but enhanced
462 at the edges. With aging of the plume, OH levels become more homogeneous (Wang et al., 2021).
463 Liao et al. (2021) showed that HCHO levels observed in fire plumes as they age during FIREX-AQ
464 were primary emissions coupled with subsequent photolytic loss and secondary production,
465 primarily from the oxidation of VOCs with OH. Therefore, with plume physical age, as oxidant
466 levels and photochemistry become more homogeneous, we can expect HCHO NEMRs to be more
467 constant throughout a transect perpendicular to the direction of the smoke transport. As such, the
468 r^2 value of the slope of CAMS HCHO and CO, from which the NEMRs were calculated, is closer
469 to 1 for larger NEMRs (and larger plume physical ages) (**Fig. 7c**). These results demonstrate that
470 the TOGA-TOF data are sufficient for assessing HCHO NEMRs in plumes with physical ages >
471 2–3 h and may result in a small bias in calculated NEMRs for younger, narrower plumes in
472 comparison to those derived using fast HCHO measurement techniques.

473 **4. Conclusions**

474 In this study, we demonstrated the TOGA-TOF as a fast response (< 2 min) GC/MS
475 instrument that can measure HCHO along with > 100 additional C₁ to C₁₀ VOCs. During FIREX-
476 AQ, the TOGA-TOF measured a wide range of HCHO mixing ratios with an LLOD of 20 ppt

477 aboard the NASA DC-8. To our knowledge, this technique has not been previously reported for
478 any ambient measurements. The discrete TOGA-TOF HCHO measurements during FIREX-AQ
479 are comparable to the averaged non-discrete 1 Hz HCHO data measured by the CAMS and UiO
480 PTR-ToF-MS instruments while the discrepancies are beyond the combined instrumental
481 uncertainties for the TOGA-TOF and ISAF instruments. The discrepancies between HCHO
482 instrumental techniques during FIREX-AQ are currently believed to be due to differences in
483 calibration standards as quantified by different optical techniques. Using volume-weighted TOGA
484 merge averages of HCHO measured by CAMS, ISAF, and UiO PTR-ToF-MS improved the
485 comparison agreement with TOGA-TOF HCHO observations. Therefore, we strongly encourage
486 the community to use the reported Time_Start, Time_Mid, and Time_Stop values in the TOGA-
487 TOF archived data files to determine volume-weighted averages of concurrent measurements.

488 TOGA-TOF derived NEMRs showed similar trends with plume physical age to what has
489 been reported from ISAF (Liao et al., 2021), with NEMRs increasing with plume ages up to ~ 5 h.
490 Due to its nature of taking discrete subsamples of air masses, as shown in Fig. 3a, highly
491 heterogeneous features may not be captured by the TOGA-TOF system in comparison to fast
492 HCHO measurement techniques, especially in chemically dynamic environments near strong
493 sources with highly heterogeneous emissions and rapidly varying HCHO production and loss often
494 detected during airborne measurements. Nevertheless, the technique used in the TOGA-TOF offers
495 a powerful HCHO-measuring method to the scientific community, with the capability to
496 simultaneously measure > 100 other C₁-C₁₀ VOC species. The wide range of gas-phase VOCs
497 measured with the TOGA-TOF can be used to provide significant insights into understanding the
498 photochemical state of air masses in rural to polluted environments. In addition, the time resolution

499 and inherent averaging is well-suited for verifying satellite observations, three-dimensional
500 modeling, and to compare with other instrumental techniques.

501 **Data availability.** All data used in this paper are available at [https://www-](https://www-air.larc.nasa.gov/missions/firex-aq/index.html)
502 [air.larc.nasa.gov/missions/firex-aq/index.html](https://www-air.larc.nasa.gov/missions/firex-aq/index.html)

503
504 **Author contributions.** ECA conceptualized the project. DJ performed the data analyses and wrote
505 the manuscript with input from ECA and RSH, RSH, AJH, GD, HSH, JPD, AF, DR, JW, PW,
506 TFH, GMW, JSC, JP, AW, TM, JBN, FP, LT, JD, CW, JS, ECA performed the measurements,
507 CDH analyzed the fire smoke ages, AS and EG contributed to the smoke and fuel identification
508 data product, JHC was the project PI and provided guidance on the analyses.

509
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524 525 **References**

526 Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., Crounse, J.
527 D., and Wennberg, P. O.: Emission factors for open and domestic biomass burning for use in
528 atmospheric models, *Atmos. Chem. Phys.*, 11, 4039–4072, [https://doi.org/10.5194/acp-11-4039-](https://doi.org/10.5194/acp-11-4039-2011)
529 2011, 2011.

530 Anderson, D. C., Nicely, J. M., Wolfe, G. M., Hanisco, T. F., Salawitch, R. J., Canty, T. P.,
531 Dickerson, R. R., Apel, E. C., Baidar, S., Bannan, T. J., Blake, N. J., Chen, D., Dix, B.,
532 Fernandez, R. P., Hall, S. R., Hornbrook, R. S., Gregory Huey, L., Josse, B., Jöckel, P.,
533 Kinnison, D. E., Koenig, T. K., Le Breton, M., Marécal, V., Morgenstern, O., Oman, L. D., Pan,
534 L. L., Percival, C., Plummer, D., Revell, L. E., Rozanov, E., Saiz-Lopez, A., Stenke, A., Sudo,
535 K., Tilmes, S., Ullmann, K., Volkamer, R., Weinheimer, A. J., and Zeng, G.: Formaldehyde in
536 the Tropical Western Pacific: Chemical Sources and Sinks, Convective Transport, and
537 Representation in CAM-Chem and the CCM1 Models, *J. Geophys. Res.-Atmos.*, 122, 11201–
538 11226, <https://doi.org/10.1002/2016JD026121>, 2017.

539 Apel, E. C., Hornbrook, R. S., Hills, A. J., Blake, N. J., Barth, M. C., Weinheimer, A., Cantrell,
540 C., Rutledge, S. A., Basarab, B., Crawford, J., Diskin, G., Homeyer, C. R., Campos, T., Flocke,
541 F., Fried, A., Blake, D. R., Brune, W., Pollack, I., Peischl, J., Ryerson, T., Wennberg, P. O.,

542 Crounse, J. D., Wisthaler, A., Mikoviny, T., Huey, G., Heikes, B., O'Sullivan, D., and Riemer,
543 D. D.: Upper tropospheric ozone production from lightning NO_x-impacted convection: Smoke
544 ingestion case study from the DC3 campaign, *J. Geophys. Res.-Atmos.*, 120, 2505–2523,
545 <https://doi.org/10.1002/2014JD022121>, 2015.

546 Bourgeois, I., Peischl, J., Neuman, J. A., Brown, S. S., Allen, H. M., Campuzano-Jost, P.,
547 Coggon, M. M., DiGangi, J. P., Diskin, G. S., Gilman, J. B., Gkatzelis, G. I., Guo, H., Halliday,
548 H. A., Hanisco, T. F., Holmes, C. D., Huey, L. G., Jimenez, J. L., Lamplugh, A. D., Lee, Y. R.,
549 Lindaas, J., Moore, R. H., Nault, B. A., Nowak, J. B., Pagonis, D., Rickly, P. S., Robinson, M.
550 A., Rollins, A. W., Selimovic, V., St. Clair, J. M., Tanner, D., Vasquez, K. T., Veres, P. R.,
551 Warneke, C., Wennberg, P. O., Washenfelder, R. A., Wiggins, E. B., Womack, C. C., Xu, L.,
552 Zarzana, K. J., and Ryerson, T. B.: Comparison of Airborne Measurements of NO, NO₂, HONO,
553 NO_y, and CO during FIREX-AQ, *Atmos. Meas. Tech.*, 15, 4901–4930,
554 <https://doi.org/10.5194/amt-15-4901-2022>, 2022.

555 Brock, C. A., Washenfelder, R. A., Trainer, M., Ryerson, T. B., Wilson, J. C., Reeves, J. M.,
556 Huey, L. G., Holloway, J. S., Parrish, D. D., Hübler, G., and Fehsenfeld, F. C.: Particle growth in
557 the plumes of coal-fired power plants, *J. Geophys. Res.-Atmos.*, 107,
558 <https://doi.org/10.1029/2001jd001062>, 2002.

559 Cao, H., Fu, T. M., Zhang, L., Henze, D. K., Miller, C. C., Lerot, C., Abad, G. G., De Smedt, I.,
560 Zhang, Q., Van Roozendaal, M., Hendrick, F., Chance, K., Li, J., Zheng, J., and Zhao, Y.:
561 Adjoint inversion of Chinese non-methane volatile organic compound emissions using space-
562 based observations of formaldehyde and glyoxal, *Atmos. Chem. Phys.*, 18, 15017–15046,
563 <https://doi.org/10.5194/acp-18-15017-2018>, 2018.

564 Cazorla, M., Wolfe, G. M., Bailey, S. A., Swanson, A. K., Arkinson, H. L., and Hanisco, T. F.: A
565 new airborne laser-induced fluorescence instrument for in situ detection of formaldehyde
566 throughout the troposphere and lower stratosphere, *Atmos. Meas. Tech.*, 8, 541–552,
567 <https://doi.org/10.5194/amt-8-541-2015>, 2015.

568 Chan Miller, C., Jacob, D. J., Marais, E. A., Yu, K., Travis, K. R., Kim, P. S., Fisher, J. A., Zhu,
569 L., Wolfe, G. M., Hanisco, T. F., Keutsch, F. N., Kaiser, J., Min, K. E., Brown, S. S.,
570 Washenfelder, R. A., González Abad, G., and Chance, K.: Glyoxal yield from isoprene oxidation
571 and relation to formaldehyde: Chemical mechanism, constraints from SENEX aircraft
572 observations, and interpretation of OMI satellite data, *Atmos. Chem. Phys.*, 17, 8725–8738,
573 <https://doi.org/10.5194/acp-17-8725-2017>, 2017.

574 Claflin, M. S., Pagonis, D., Finewax, Z., Handschy, A. V., Day, D. A., Brown, W. L., Jayne, J.
575 T., Worsnop, D. R., Jimenez, J. L., Ziemann, P. J., de Gouw, J., and Lerner, B. M.: An in situ gas
576 chromatograph with automatic detector switching between PTR-and EI-TOF-MS: Isomer-
577 resolved measurements of indoor air, *Atmos. Meas. Tech.*, 14, 133–152,
578 <https://doi.org/10.5194/amt-14-133-2021>, 2021.

579 Curci, G., Palmer, P. I., Kurosu, T. P., Chance, K., and Visconti, G.: Estimating European
580 volatile organic compound emissions using satellite observations of formaldehyde from the
581 Ozone Monitoring Instrument, *Atmos. Chem. Phys.*, 10, 11501–11517,
582 <https://doi.org/10.5194/acp-10-11501-2010>, 2010.

583 Decker, Z. C. J., Robinson, M. A., Barsanti, K. C., Bourgeois, I., Coggon, M. M., Digangi, J. P.,
584 Diskin, G. S., Flocke, F. M., Franchin, A., Fredrickson, C. D., Gkatzelis, G. I., Hall, S. R.,
585 Halliday, H., Holmes, C. D., Huey, L. G., Lee, Y. R., Lindaas, J., Middlebrook, A. M., Montzka,
586 D. D., Moore, R., Neuman, J. A., Nowak, J. B., Palm, B. B., Peischl, J., Piel, F., Rickly, P. S.,
587 Rollins, A. W., Ryerson, T. B., Schwantes, R. H., Sekimoto, K., Thornhill, L., Thornton, J. A.,
588 Tyndall, G. S., Ullmann, K., Van Rooy, P., Veres, P. R., Warneke, C., Washenfelder, R. A.,
589 Weinheimer, A. J., Wiggins, E., Winstead, E., Wisthaler, A., Womack, C., and Brown, S. S.:
590 Nighttime and daytime dark oxidation chemistry in wildfire plumes: An observation and model
591 analysis of FIREX-AQ aircraft data, *Atmos. Chem. Phys.*, 21, 16293–16317,
592 <https://doi.org/10.5194/acp-21-16293-2021>, 2021.

593 Duncan, B. N., Yoshida, Y., Olson, J. R., Sillman, S., Martin, R. V., Lamsal, L., Hu, Y.,
594 Pickering, K. E., Retscher, C., Allen, D. J., and Crawford, J. H.: Application of OMI
595 observations to a space-based indicator of NO_x and VOC controls on surface ozone formation,
596 *Atmos. Environ.*, 44, 2213–2223, <https://doi.org/10.1016/j.atmosenv.2010.03.010>, 2010.

597 Fried, A., Walega, J., Weibring, P., Richter, D., Simpson, I. J., Blake, D. R., Blake, N. J.,
598 Meinardi, S., Barletta, B., Hughes, S. C., Crawford, J. H., Diskin, G., Barrick, J., Hair, J., Fenn,
599 M., Wisthaler, A., Mikoviny, T., Woo, J.-H., Park, M., Kim, J., Min, K.-E., Jeong, S., Wennberg,
600 P. O., Kim, M. J., Crouse, J. D., Teng, A. P., Bennett, R., Yang-Martin, M., Shook, M. A.,
601 Huey, G., Tanner, D., Knote, C., Kim, J., Park, R., and Brune, W.: Airborne formaldehyde and
602 volatile organic compound measurements over the Daesan petrochemical complex on Korea's
603 northwest coast during the Korea-United States Air Quality study, *Elem. Sci. Anth.*, 8, 121,
604 <https://doi.org/10.1525/elementa.2020.121>, 2020.

605 Gilman, J. B., Lerner, B. M., Kuster, W. C., Goldan, P. D., Warneke, C., Veres, P. R., Roberts, J.
606 M., de Gouw, J. A., Burling, I. R., and Yokelson, R. J.: Biomass burning emissions and potential
607 air quality impacts of volatile organic compounds and other trace gases from fuels common in
608 the US, *Atmos. Chem. Phys.*, 15, 13915–13938, <https://doi.org/10.5194/acp-15-13915-2015>,
609 2015.

610 Gilpin, T., Apel, E., Fried, A., Wert, B., Calvert, J., Genfa, Z., Dasgupta, P., Harder, J. W.,
611 Heikes, B., Hopkins, B., Westberg, H., Kleindienst, T., Lee, Y. N., Zhou, X., Lonneman, W., and
612 Sewell, S.: Intercomparison of six ambient [CH₂O] measurement techniques, *J. Geophys. Res.-*
613 *Atmos.*, 102, 21161–21188, <https://doi.org/10.1029/97jd01314>, 1997.

614 Gkatzelis, G. I., Coggon, M. M., Stockwell, C. E., Hornbrook, R. S., Allen, H., Apel, E. C., Bela,
615 M. M., Blake, D. R., Bourgeois, I., Brown, S. S., Campuzano-Jost, P., Clair, J. M. S., Crawford,
616 J. H., Crouse, J. D., Day, D. A., Digangi, J. P., Diskin, G. S., Fried, A., Gilman, J. B., Guo, H.,
617 Hair, J. W., Halliday, H. S., Hanisco, T. F., Hannun, R., Hills, A., Huey, L. G., Jimenez, J. L.,
618 Katich, J. M., Lamplugh, A., Lee, Y. R., Liao, J., Lindaas, J., Mckeen, S. A., Mikoviny, T.,
619 Nault, B. A., Neuman, J. A., Nowak, J. B., Pagonis, D., Peischl, J., Perring, A. E., Piel, F.,
620 Rickly, P. S., Robinson, M. A., Rollins, A. W., Ryerson, T. B., Schueneman, M. K., Schwantes,
621 R. H., Schwarz, J. P., Sekimoto, K., Selimovic, V., Shingler, T., Tanner, D. J., Tomsche, L.,
622 Vasquez, K. T., Veres, P. R., Washenfelder, R., Weibring, P., Wennberg, P. O., Wisthaler, A.,
623 Wolfe, G. M., Womack, C. C., Xu, L., Ball, K., Yokelson, R. J., and Warneke, C.:
624 Parameterizations of US Wildfire and Prescribed Fire Emission Ratios and Emission Factors
625 Based on FIREX-AQ Aircraft Measurements, *Atmos. Chem. Phys.*, 24, 929–956,
626 <https://doi.org/10.5194/acp-24-929-2024>, 2024.

627 Green, J. R., Fiddler, M. N., Fibiger, D. L., McDuffie, E. E., Aquino, J., Campos, T., Shah, V.,
628 Jaeglé, L., Thornton, J. A., DiGangi, J. P., Wolfe, G. M., Bililign, S., and Brown, S. S.:
629 Wintertime Formaldehyde: Airborne Observations and Source Apportionment Over the Eastern
630 United States, *J. Geophys. Res.-Atmos.*, 126, e2020JD033518,
631 <https://doi.org/10.1029/2020JD033518>, 2021.

632 Hopkins, J. R., Still, T., Al-Haider, S., Fisher, I. R., Lewis, A. C., and Seakins, P. W.: A
633 simplified apparatus for ambient formaldehyde detection via GC-pHID, *Atmos. Environ.*, 37,
634 2557–2565, [https://doi.org/10.1016/S1352-2310\(03\)00178-X](https://doi.org/10.1016/S1352-2310(03)00178-X), 2003.

635 Hornbrook, R. S., Blake, D. R., Diskin, G. S., Fried, A., Fuelberg, H. E., Meinardi, S., Mikoviny,
636 T., Richter, D., Sachse, G. W., Vay, S. A., Walega, J., Weibring, P., Weinheimer, A. J.,
637 Wiedinmyer, C., Wisthaler, A., Hills, A., Riemer, D. D., and Apel, E. C.: Observations of
638 nonmethane organic compounds during ARCTAS-Part 1: Biomass burning emissions and plume
639 enhancements, *Atmos. Chem. Phys.*, 11, 11103–11130, [https://doi.org/10.5194/acp-11-11103-](https://doi.org/10.5194/acp-11-11103-2011)
640 2011, 2011.

641 Hunter, M. C., Bartle, K. D., Seakins, P. W., and Lewis, A. C.: Direct measurement of
642 atmospheric formaldehyde using gas chromatography-pulsed discharge ionisation detection,
643 *Anal. Commun.*, 36, 101–104, <https://doi.org/10.1039/A809762C>, 1999.

644 Isaacman-VanWertz, G., Sueper, D. T., Aikin, K. C., Lerner, B. M., Gilman, J. B., de Gouw, J.
645 A., Worsnop, D. R., and Goldstein, A. H.: Automated single-ion peak fitting as an efficient
646 approach for analyzing complex chromatographic data, *J. Chromatogr. A*, 1529, 81–92,
647 <https://doi.org/10.1016/j.chroma.2017.11.005>, 2017.

648 Jaeglé, L., Shah, V., Thornton, J. A., Lopez-Hilfiker, F. D., Lee, B. H., McDuffie, E. E., Fibiger,
649 D., Brown, S. S., Veres, P., Sparks, T. L., Ebben, C. J., Wooldridge, P. J., Kenagy, H. S., Cohen,
650 R. C., Weinheimer, A. J., Campos, T. L., Montzka, D. D., Digangi, J. P., Wolfe, G. M., Hanisco,
651 T., Schroder, J. C., Campuzano-Jost, P., Day, D. A., Jimenez, J. L., Sullivan, A. P., Guo, H., and
652 Weber, R. J.: Nitrogen Oxides Emissions, Chemistry, Deposition, and Export Over the Northeast
653 United States During the WINTER Aircraft Campaign, *J. Geophys. Res.-Atmos.*, 123, 12368–
654 12393, <https://doi.org/10.1029/2018JD029133>, 2018.

655 Kaiser, J., Jacob, D. J., Zhu, L., Travis, K. R., Fisher, J. A., González Abad, G., Zhang, L.,
656 Zhang, X., Fried, A., Crouse, J. D., Clair, J. M. S., and Wisthaler, A.: High-resolution inversion
657 of OMI formaldehyde columns to quantify isoprene emission on ecosystem-relevant scales:
658 Application to the southeast US, *Atmos. Chem. Phys.*, 18, 5483–5497,
659 <https://doi.org/10.5194/acp-18-5483-2018>, 2018.

660 Karbiwnyk, C. M., Mills, C. S., Helmig, D., and Birks, J. W.: Use of Chlorofluorocarbons as
661 Internal Standards for the Measurement of Atmospheric Non-Methane Volatile Organic
662 Compounds Sampled onto Solid Adsorbent Cartridges, *Environ. Sci. Technol.*, 37, 1002–1007,
663 <https://doi.org/10.1021/es025910q>, 2003.

664 Koss, A. R., Sekimoto, K., Gilman, J. B., Selimovic, V., Coggon, M. M., Zarzana, K. J., Yuan,
665 B., Lerner, B. M., Brown, S. S., Jimenez, J. L., Krechmer, J., Roberts, J. M., Warneke, C.,
666 Yokelson, R. J., and de Gouw, J.: Non-methane organic gas emissions from biomass burning:
667 Identification, quantification, and emission factors from PTR-ToF during the FIREX 2016

668 laboratory experiment, *Atmos. Chem. Phys.*, 18, 3299–3319, [https://doi.org/10.5194/acp-18-](https://doi.org/10.5194/acp-18-3299-2018)
669 3299–2018, 2018.

670 Kwon, H. A., Park, R. J., Oak, Y. J., Nowlan, C. R., Janz, S. J., Kowalewski, M. G., Fried, A.,
671 Walega, J., Bates, K. H., Choi, J., Blake, D. R., Wisthaler, A., and Woo, J. H.: Top-down
672 estimates of anthropogenic VOC emissions in South Korea using formaldehyde vertical column
673 densities from aircraft during the KORUS-AQ campaign, *Elem. Sci. Anth.*, 9, 00109,
674 <https://doi.org/10.1525/elementa.2021.00109>, 2021.

675 Lee, M., Heikes, B. G., Jacob, D. J., Sachse, G., and Anderson, B.: Hydrogen peroxide, organic
676 hydroperoxide, and formaldehyde as primary pollutants from biomass burning, *J. Geophys. Res.-*
677 *Atmos.*, 102, 1301–1309, <https://doi.org/10.1029/96JD01709>, 1997.

678 Lerner, B. M., Gilman, J. B., Aikin, K. C., Atlas, E. L., Goldan, P. D., Graus, M., Hendershot,
679 R., Isaacman-VanWertz, G. A., Koss, A., Kuster, W. C., Lueb, R. A., McLaughlin, R. J., Peischl,
680 J., Sueper, D., Ryerson, T. B., Tokarek, T. W., Warneke, C., Yuan, B., and de Gouw, J. A.: An
681 improved, automated whole air sampler and gas chromatography mass spectrometry analysis
682 system for volatile organic compounds in the atmosphere, *Atmos. Meas. Tech.*, 10, 291–313,
683 <https://doi.org/10.5194/amt-10-291-2017>, 2017.

684 Liao, J., Wolfe, G. M., Hannun, R. A., St. Clair, J. M., Hanisco, T. F., Gilman, J. B., Lamplugh,
685 A., Selimovic, V., Diskin, G. S., Nowak, J. B., Halliday, H. S., Digangi, J. P., Hall, S. R.,
686 Ullmann, K., Holmes, C. D., Fite, C. H., Agastra, A., Ryerson, T. B., Peischl, J., Bourgeois, I.,
687 Warneke, C., Coggon, M. M., Gkatzelis, G. I., Sekimoto, K., Fried, A., Richter, D., Weibring, P.,
688 Apel, E. C., Hornbrook, R. S., Brown, S. S., Womack, C. C., Robinson, M. A., Washenfelder, R.
689 A., Veres, P. R., and Neuman, J. A.: Formaldehyde evolution in US wildfire plumes during the
690 Fire Influence on Regional to Global Environments and Air Quality experiment (FIREX-AQ),
691 *Atmos. Chem. Phys.*, 21, 18319–18331, <https://doi.org/10.5194/acp-21-18319-2021>, 2021.

692 Liao, J., Wolfe, G. M., Kotsakis, A. E., Nicely, J. M., St. Clair, J. M., Hanisco, T. F., Abad, G.
693 G., Nowlan, C. R., Ayazpour, Z., De Smedt, I., Apel, E. C., Hornbrook, R. S.: Validation of
694 formaldehyde products from three satellite retrievals (OMI SAO, OMPS-NPP SAO, and OMI
695 BIRA) in the marine atmosphere with four seasons of Atmospheric Tomography Mission
696 (ATom) aircraft observations, *Atmos. Meas. Tech.*, 18, 1–16, [https://doi.org/10.5194/amt-18-1-](https://doi.org/10.5194/amt-18-1-697)
697 2025, 2025.

698 Luecken, D. J., Hutzell, W. T., Strum, M. L., and Pouliot, G. A.: Regional sources of
699 atmospheric formaldehyde and acetaldehyde, and implications for atmospheric modeling, *Atmos.*
700 *Environ.*, 47, 477–490, <https://doi.org/10.1016/j.atmosenv.2011.10.005>, 2012.

701 Luecken, D. J., Napelenok, S. L., Strum, M., Scheffe, R., and Phillips, S.: Sensitivity of Ambient
702 Atmospheric Formaldehyde and Ozone to Precursor Species and Source Types Across the
703 United States, *Environ. Sci. Technol.*, 52, 4668–4675, <https://doi.org/10.1021/acs.est.7b05509>,
704 2018.

705 Martin, R. V., Fiore, A. M., and Van Donkelaar, A.: Space-based diagnosis of surface ozone
706 sensitivity to anthropogenic emissions, *Geophys. Res. Lett.*, 31, L06120,
707 <https://doi.org/10.1029/2004GL019416>, 2004.

708 Müller, M., Mikoviny, T., Feil, S., Haidacher, S., Hanel, G., Hartungen, E., Jordan, A., Märk, L.,
709 Mutschlechner, P., Schottkowsky, R., Sulzer, P., Crawford, J. H., and Wisthaler, A.: A compact

710 PTR-ToF-MS instrument for airborne measurements of volatile organic compounds at high
711 spatiotemporal resolution, *Atmos. Meas. Tech.*, 7, 3763–3772, [https://doi.org/10.5194/amt-7-](https://doi.org/10.5194/amt-7-3763-2014)
712 3763-2014, 2014.

713 Müller, M., Anderson, B. E., Beyersdorf, A. J., Crawford, J. H., Diskin, G. S., Eichler, P., Fried,
714 A., Keutsch, F. N., Mikoviny, T., Thornhill, K. L., Walega, J. G., Weinheimer, A. J., Yang, M.,
715 Yokelson, R. J., and Wisthaler, A.: In situ measurements and modeling of reactive trace gases in
716 a small biomass burning plume, *Atmos. Chem. Phys.*, 16, 3813–3824,
717 <https://doi.org/10.5194/acp-16-3813-2016>, 2016.

718 National Research Council (US) Committee on Toxicology: Formaldehyde - An Assessment of
719 Its Health Effects, Washington (DC), 1980.

720 Nussbaumer, C. M., Crowley, J. N., Schuladen, J., Williams, J., Hafermann, S., Reiffs, A.,
721 Axinte, R., Harder, H., Ernest, C., Novelli, A., Sala, K., Martinez, M., Mallik, C., Tomsche, L.,
722 Plass-Dülmer, C., Bohn, B., Lelieveld, J., and Fischer, H.: Measurement report: Photochemical
723 production and loss rates of formaldehyde and ozone across Europe, *Atmos. Chem. Phys.*, 21,
724 18413–18432, <https://doi.org/10.5194/acp-21-18413-2021>, 2021.

725 Palmer, P. I., Jacob, D. J., Fiore, A. M., Martin, R. V., Chance, K., and Kurosu, T. P.: Mapping
726 isoprene emissions over North America using formaldehyde column observations from space, *J.*
727 *Geophys. Res.-Atmos.*, 108, 4180, <https://doi.org/10.1029/2002jd002153>, 2003.

728 Pfister, G., Wang, C. T., Barth, M., Flocke, F., Vizuete, W., and Walters, S.: Chemical
729 Characteristics and Ozone Production in the Northern Colorado Front Range, *J. Geophys. Res. -*
730 *Atmos.*, 124, 13397–13419, <https://doi.org/10.1029/2019JD030544>, 2019.

731 Possanzini, M., Palo, V. di, and Cecinato, A.: Sources and photodecomposition of formaldehyde
732 and acetaldehyde in Rome ambient air, *Atmos. Environ.*, 36, 3195–3201,
733 [https://doi.org/10.1016/S1352-2310\(02\)00192-9](https://doi.org/10.1016/S1352-2310(02)00192-9), 2002.

734 Rice, A. L. and Quay, P. D.: Isotopic analysis of atmospheric formaldehyde by gas
735 chromatography isotope ratio mass spectrometry, *Anal. Chem.*, 78, 6320–6326,
736 <https://doi.org/10.1021/ac0602367>, 2006.

737 Richter, D., Weibring, P., Walega, J. G., Fried, A., Spuler, S. M., and Taubman, M. S.: Compact
738 highly sensitive multi-species airborne mid-IR spectrometer, *Appl. Phys. B*, 119, 119–131,
739 <https://doi.org/10.1007/s00340-015-6038-8>, 2015.

740 Robinson, M. A., Decker, Z. C. J., Barsanti, K. C., Coggon, M. M., Flocke, F. M., Franchin, A.,
741 Fredrickson, C. D., Gilman, J. B., Gkatzelis, G. I., Holmes, C. D., Lamplugh, A., Lavi, A.,
742 Middlebrook, A. M., Montzka, D. M., Palm, B. B., Peischl, J., Pierce, B., Schwantes, R. H.,
743 Sekimoto, K., Selimovic, V., Tyndall, G. S., Thornton, J. A., Van Rooy, P., Warneke, C.,
744 Weinheimer, A. J., and Brown, S. S.: Variability and Time of Day Dependence of Ozone
745 Photochemistry in Western Wildfire Plumes, *Environ. Sci. Technol.*, 55, 10280–10290,
746 <https://doi.org/10.1021/acs.est.1c01963>, 2021.

747 Sachse, G. W., Hill, G. F., Wade, L. O., and Perry, M. G.: Fast-response, high-precision carbon
748 monoxide sensor using a tunable diode laser absorption technique, *J. Geophys. Res.-Atmos.*, 92,
749 2071–2081, <https://doi.org/10.1029/JD092iD02p02071>, 1987.

750 Sagebiel, J. C., Zielinska, B., Pierson, W. R., and Gertler, A. W.: Real-world emissions and
751 calculated reactivities of organic species from motor vehicles, *Atmos. Environ.*, 30, 2287–2296,
752 [https://doi.org/10.1016/1352-2310\(95\)00117-4](https://doi.org/10.1016/1352-2310(95)00117-4), 1996.

753 Schroeder, J. R., Crawford, J. H., Fried, A., Walega, J., Weinheimer, A., Wisthaler, A., Müller,
754 M., Mikoviny, T., Chen, G., Shook, M., Blake, D. R., and Tonnesen, G. S.: New insights into the
755 column CH₂O/NO₂ ratio as an indicator of near-surface ozone sensitivity, *J. Geophys. Res.-*
756 *Atmos.*, 122, 8885–8907, <https://doi.org/10.1002/2017JD026781>, 2017.

757 Seinfeld, J. H. and Pandis, S. N.: *Atmospheric Chemistry and Physics: From Air Pollution to*
758 *Climate Change*, 2nd ed., Wiley-Interscience, Hoboken, NJ, 2006.

759 Selimovic, V., Yokelson, R. J., McMeeking, G. R., and Coe field, S.: In situ measurements of
760 trace gases, PM, and aerosol optical properties during the 2017 NW US wildfire smoke event,
761 *Atmos. Chem. Phys.*, 19, 3905–3926, <https://doi.org/10.5194/acp-19-3905-2019>, 2019.

762 Souri, A. H., Nowlan, C. R., Wolfe, G. M., Lamsal, L. N., Miller, C., González Abad, G., Janz,
763 S. J., Fried, A., Blake, D. R., Weinheimer, A. J., Diskin, G. S., Liu, X., and Chance, K.:
764 Revisiting the effectiveness of HCHO/NO₂ ratios for inferring ozone sensitivity to its precursors
765 using high resolution airborne remote sensing observations in a high ozone episode during the
766 KORUS-AQ campaign, *Atmos. Environ.*, 224, 117341,
767 <https://doi.org/10.1016/j.atmosenv.2020.117341>, 2020.

768 Souri, A. H., Johnson, M. S., Wolfe, G. M., Crawford, J. H., Fried, A., Wisthaler, A., Brune, W.
769 H., Blake, D. R., Weinheimer, A. J., Verhoelst, T., Compernelle, S., Pinardi, G., Vigouroux, C.,
770 Langerock, B., Choi, S., Lamsal, L., Zhu, L., Sun, S., Cohen, R. C., Min, K. E., Cho, C., Philip,
771 S., Liu, X., and Chance, K.: Characterization of errors in satellite-based HCHO/NO₂
772 tropospheric column ratios with respect to chemistry, column-to-PBL translation, spatial
773 representation, and retrieval uncertainties, *Atmos. Chem. Phys.*, 23, 1963–1986,
774 <https://doi.org/10.5194/acp-23-1963-2023>, 2023.

775 Stavrou, T., Müller, J.-F., De Smedt, I., Van Roozendaal, M., van der Werf, G. R., Giglio, L.,
776 and Guenther, A.: Evaluating the performance of pyrogenic and biogenic emission inventories
777 against one decade of space-based formaldehyde columns, *Atmos. Chem. Phys.*, 1037–1060,
778 <https://doi.org/10.5194/acp-9-1037-2009>, 2009.

779 Travis, K. R., Crawford, James. H., Soja, A. J., Gargulinski, E. M., Moore, R. H., Wiggins, E.
780 B., Diskin, G. S., DiGangi, J. P., Nowak, J. B., Halliday, H., Yokelson, R. J., McCarty, J. L.,
781 Simpson, I. J., Blake, D. R., Neinardi, S., Hornbrook, R. S., Apel, E. C., Hills, A. J., Warneke,
782 C., Coggon, M. M., Rollins, A. W., Gilman, J. B., Womack, C. C., Robinson, M. A., Katich, J.
783 M., Peischl, J., Gkatzelis, G. I., Bourgeois, I., Rickly, P. S., Lamplugh, A., Dibb, J. E., Jimenez,
784 J. L., Campuzano-Jost, P., Day, D. A., Guo, H., Pagonis, D., Wennberg, P. O., Crouse, J. D.,
785 Xu, L., Hanisco, T. F., Wolfe, G. M., Liao, J., St. Clair, J. M., Nault, B. A., Fried, A., and
786 Perring, A. E.: Emission Factors for Crop Residue and Prescribed Fires in the Eastern US during
787 FIREX-AQ, *J. Geophys. Res.-Atmos.*, 128, e2023JD039309,
788 <https://doi.org/10.1029/2023JD039309>, 2023.

789 Trentmann, J., Andreae, M. O., and Graf, H. F.: Chemical processes in a young biomass-burning
790 plume, *J. Geophys. Res.-Atmos.*, 108, 4705, <https://doi.org/10.1029/2003jd003732>, 2003a.

791 Trentmann, J., Früh, B., Boucher, O., Trautmann, T., and Andreae, M. O.: Three-dimensional
792 solar radiation effects on the actinic flux field in a biomass-burning plume, *J. Geophys. Res.-*
793 *Atmos.*, 108, 4558, <https://doi.org/10.1029/2003jd003422>, 2003b.

794 Viskari, E.-L., Vartiainen, M., and Pasanen, P.: Seasonal and diurnal variation in formaldehyde
795 and acetaldehyde concentrations along a highway in Eastern Finland, *Atmos. Environ.*, 34, 917–
796 923, [https://doi.org/10.1016/S1352-2310\(99\)00307-6](https://doi.org/10.1016/S1352-2310(99)00307-6), 2000.

797 Wang, S., Coggon, M. M., Gkatzelis, G. I., Warneke, C., Bourgeois, I., Ryerson, T., Peischl, J.,
798 Veres, P. R., Neuman, J. A., Hair, J., Shingler, T., Fenn, M., Diskin, G., Huey, L. G., Lee, Y. R.,
799 Apel, E. C., Hornbrook, R. S., Hills, A. J., Hall, S. R., Ullmann, K., Bela, M. M., Trainer, M. K.,
800 Kumar, R., Orlando, J. J., Flocke, F. M., and Emmons, L. K.: Chemical Tomography in a Fresh
801 Wildland Fire Plume: A Large Eddy Simulation (LES) Study, *J. Geophys. Res.-Atmos.*, 126,
802 e2021JD035203, <https://doi.org/10.1029/2021JD035203>, 2021.

803 Warneke, C., McKeen, S. A., de Gouw, J. A., Goldan, P. D., Kuster, W. C., Holloway, J. S.,
804 Williams, E. J., Lerner, B. M., Parrish, D. D., Trainer, M., Fehsenfeld, F. C., Kato, S., Atlas, E.
805 L., Baker, A., and Blake, D. R.: Determination of urban volatile organic compound emission
806 ratios and comparison with an emissions database, *J. Geophys. Res.-Atmos.*, 112, D10S47,
807 <https://doi.org/10.1029/2006JD007930>, 2007.

808 Warneke, C., Schwarz, J. P., Dibb, J., Kalashnikova, O., Frost, G., Al-Saad, J., Brown, S. S.,
809 Brewer, Wm. A., Soja, A., Seidel, F. C., Washenfelder, R. A., Wiggins, E. B., Moore, R. H.,
810 Anderson, B. E., Jordan, C., Yacovitch, T. I., Herndon, S. C., Liu, S., Kuwayama, T., Jaffe, D.,
811 Johnston, N., Selimovic, V., Yokelson, R., Giles, D. M., Holben, B. N., Goloub, P., Popovici, I.,
812 Trainer, M., Kumar, A., Pierce, R. B., Fahey, D., Roberts, J., Gargulinski, E. M., Peterson, D. A.,
813 Ye, X., Thapa, L. H., Saide, P. E., Fite, C. H., Holmes, C. D., Wang, S., Coggon, M. M., Decker,
814 Z. C. J., Stockwell, C. E., Xu, L., Gkatzelis, G., Aikin, K., Lefer, B., Kaspari, J., Griffin, D.,
815 Zeng, L., Weber, R., Hastings, M., Chai, J., Wolfe, G. M., Hanisco, T. F., Liao, J., Campuzano
816 Jost, P., Guo, H., Jimenez, J. L., Crawford, J., and Team, T. F.-A. S.: Fire Influence on Regional
817 to Global Environments and Air Quality (FIREX-AQ), *J. Geophys. Res.-Atmos.*, 128,
818 e2022JD037758, <https://doi.org/https://doi.org/10.1029/2022JD037758>, 2023.

819 Wolfe, G. M., Kaiser, J., Hanisco, T. F., Keutsch, F. N., de Gouw, J. A., Gilman, J. B., Graus,
820 M., Hatch, C. D., Holloway, J., Horowitz, L. W., Lee, B. H., Lerner, B. M., Lopez-Hilifiker, F.,
821 Mao, J., Marvin, M. R., Peischl, J., Pollack, I. B., Roberts, J. M., Ryerson, T. B., Thornton, J. A.,
822 Veres, P. R., and Warneke, C.: Formaldehyde production from isoprene oxidation across NO_x
823 regimes, *Atmos. Chem. Phys.*, 16, 2597–2610, <https://doi.org/10.5194/acp-16-2597-2016>, 2016.

824 Yen, J., Leber, D., and Pibida, L.: Comparing Instruments, Gaithersburg, MD,
825 <https://doi.org/10.6028/NIST.TN.2106>, 2020.

826 Yokelson, R. J., Andreae, M. O., and Akagi, S. K.: Pitfalls with the use of enhancement ratios or
827 normalized excess mixing ratios measured in plumes to characterize pollution sources and aging,
828 *Atmos. Meas. Tech.*, 6, 2155–2158, <https://doi.org/10.5194/amt-6-2155-2013>, 2013a.

829 Yokelson, R. J., Burling, I. R., Gilman, J. B., Warneke, C., Stockwell, C. E., de Gouw, J., Akagi,
830 S. K., Urbanski, S. P., Veres, P., Roberts, J. M., Kuster, W. C., Reardon, J., Griffith, D. W. T.,
831 Johnson, T. J., Hosseini, S., Miller, J. W., Cocker, D. R., Jung, H., and Weise, D. R.: Coupling
832 field and laboratory measurements to estimate the emission factors of identified and unidentified

833 trace gases for prescribed fires, *Atmos. Chem. Phys.*, 13, 89–116, [https://doi.org/10.5194/acp-13-](https://doi.org/10.5194/acp-13-89-2013)
834 89-2013, 2013b.

835 Zhang, H., Li, J., Ying, Q., Guven, B. B., and Olaguer, E. P.: Source apportionment of
836 formaldehyde during TexAQS 2006 using a source-oriented chemical transport model, *J.*
837 *Geophys. Res.-Atmos.*, 118, 1525–1535, <https://doi.org/10.1002/jgrd.50197>, 2013.

838 Zhu, L., Jacob, D. J., Kim, P. S., Fisher, J. A., Yu, K., Travis, K. R., Mickley, L. J., Yantosca, R.
839 M., Sulprizio, M. P., De Smedt, I., Abad, G. G., Chance, K., Li, C., Ferrare, R., Fried, A., Hair, J.
840 W., Hanisco, T. F., Richter, D., Scarino, A. J., Walega, J., Weibring, P., and Wolfe, G. M.:
841 Observing atmospheric formaldehyde (HCHO) from space: Validation and intercomparison of
842 six retrievals from four satellites (OMI, GOME2A, GOME2B, OMPS) with SEAC⁴RS aircraft
843 observations over the southeast US, *Atmos. Chem. Phys.*, 16, 13477–13490,
844 <https://doi.org/10.5194/acp-16-13477-2016>, 2016.

845 Zhu, L., Jacob, D. J., Keutsch, F. N., Mickley, L. J., Scheffe, R., Strum, M., González Abad, G.,
846 Chance, K., Yang, K., Rappenglück, B., Millet, D. B., Baasandorj, M., Jaeglé, L., and Shah, V.:
847 Formaldehyde (HCHO) As a Hazardous Air Pollutant: Mapping Surface Air Concentrations
848 from Satellite and Inferring Cancer Risks in the United States, *Environ. Sci. Technol.*, 51, 5650–
849 5657, <https://doi.org/10.1021/acs.est.7b01356>, 2017.

850

851 **Figure captions**

852

853 **Figure 1.** TOGA-TOF operational sequences for the gas chromatograph temperature program
854 (red) and the preconcentration steps (blue). The position of the valves (10-port valve and stream
855 selector valve; SSV) and temperatures of each trap are shown for each step. The duration of each
856 sequence corresponds to the normal sampling mode timing during FIREX-AQ.

857

858 **Figure 2.** Simplified flow diagram of TOGA-TOF showing flows and temperatures of each trap
859 during (a) trap cleaning mode, (b) sampling mode, (c) chasing mode, and (d) cryofocusing mode,
860 corresponding to the sequence in **Fig. 1**.

861

862 **Figure 3.** Example time series of CAMS and TOGA-TOF HCHO and LGR CO observations
863 during FIREX-AQ transverse (i.e., perpendicular to the transport of the fire plume) plume
864 transects with relatively (a) high and (b) low variability in the observed 1 Hz HCHO. The pink
865 shading shows when the DC-8 was inside the smoke plume based on the relative changes in CO.
866 Vertical error bars on the TOGA-TOF HCHO data points show the measurement uncertainty (\pm
867 35%), centered on the TOGA-TOF sampling time midpoint, and the horizontal error bars span
868 the start to stop times of each TOGA-TOF sample. For each of the start-to-midpoint and
869 midpoint-to-stop times of the TOGA-TOF sampling periods, 50% of the total sample volume
870 occurred. These times (start/midpoint/stop) were used for deriving weighted averages of the 1 Hz
871 HCHO and CO data.

872

873 **Figure 4.** Correlation of measured HCHO (a,c) between TOGA-TOF and CAMS and (b,d)
874 between TOGA-TOF and ISAF. CAMS and ISAF 1 Hz HCHO data were averaged over the
875 TOGA-TOF sampling start and stop times using the (a,b) non-weighted and (c,d) volume-
876 weighted averaging methods. Measurement uncertainties of each measurement technique are

877 shown as error bars. The grey shading around the 1:1 line is the HCHO measurement
878 uncertainties of the two corresponding instruments added in quadrature. Each data point is
879 colored by the relative standard deviation of the 1 Hz HCHO data (standard deviation of 1 Hz
880 averaged HCHO divided by TOGA-TOF HCHO) during the TOGA-TOF sampling period.

881
882 **Figure 5.** HCHO NEMRs of Western U.S. fires with respect to plume physical age during
883 FIREX-AQ. Quadratic polynomial fittings of NEMRs derived using TOGA-TOF, CAMS, and
884 ISAF HCHO data are shown.

885
886 **Figure 6.** Correlation plots between (a) TOGA-TOF and CAMS and (b) TOGA-TOF and ISAF
887 derived HCHO NEMRs. Orthogonal distance regression fittings of the correlations are shown.
888 The grey shading around the dashed 1:1 line shows the measurement uncertainty (35%) of
889 TOGA-TOF HCHO.

890
891 **Figure 7.** Plots of HCHO NEMR bias between TOGA-TOF and CAMS measurements with
892 respect to the magnitude of the mean values from both instruments. Each data point is color
893 coded with (a) number of TOGA-TOF sampling in each plume transect, (b) plume physical age,
894 and (c) r^2 of the CAMS measured HCHO NEMR to CO.

895