



1 **Hydraulic Redistribution Decreases with Precipitation Magnitude and Frequency in a**
2 **Dryland Ecosystem: A Data-Model Fusion Approach**

3 Aneesh Kumar Chandel¹, Mitra Cattray³, Yu Zhou⁴, Hang Duong^{2,5}, Marcy Litvak², William
4 Pockman² and Yiqi Luo¹

5 ¹School of Integrative Plant Science, Cornell University, Ithaca, NY, USA

6 ²University of New Mexico, Albuquerque, NM, USA

7 ³Department of Earth and Environmental Engineering, Columbia University, USA

8 ⁴Department of Environmental Systems Science, ETH Zurich, Switzerland

9 ⁵Vietnam National University of Agriculture, Hanoi, Vietnam

10 Corresponding Authors: Aneesh Kumar Chandel (akc76@cornell.edu), Yiqi Luo
11 (y12735@cornell.edu)

12

13

14

15

16

17

18

19

20



21 **Abstract**

22 Hydraulic redistribution (HR), the movement of water via plant root systems that connect soil
23 compartments with different water potential, should influences soil moisture dynamics
24 particularly in water-limited ecosystems. Realistic representation of HR in ecosystem models is
25 essential to improve the ability of these models to predict ecosystem function in dryland regions.
26 In this study, we integrated HR into the Terrestrial ECOsystem model and employed a Bayesian
27 Markov Chain Monte Carlo technique to optimize soil hydraulic parameters and root
28 conductance using four years of soil moisture observations from a piñon-juniper woodland. We
29 found that (i) integrating HR generally improved model prediction of soil moisture during dry
30 periods, particularly in the top 30 cm of the soil profile, where more than 50% of root biomass
31 exists, mostly during dry periods; (ii) HR increased surface soil moisture by up to 60% during
32 dry periods; (iii) HR decreased with increasing precipitation magnitude and frequency, however,
33 the length of dry spells between rainfall events also influenced HR rates; and (iv) upward HR in
34 the top 60 cm soil profile became more pronounced as dry conditions progressed, with rates
35 ranging from 0.10 to 0.50 mm d⁻¹. These findings highlight that HR plays a likely role in
36 sustaining soil moisture during extended dry periods and has a limited effect during precipitation
37 events. Future research should investigate the effect of HR on other ecosystem processes, such as
38 net ecosystem exchange of carbon and evapotranspiration under varying climatic conditions.

39

40

41

42



43 1. Introduction

44 Drylands, cover over 40% of Earth's terrestrial surface and support more than 38% of the
45 global population (Právělie, 2016). Ecosystem function in these regions is likely to be limited by
46 altered precipitation in the changing climate (Beer et al., 2010; Ukkola et al., 2021).
47 Understanding the ability of plants to mitigate the potential negative impacts of alter
48 precipitation is therefore critical for predicting ecosystem stability. Hydraulic redistribution (HR)
49 is the passive movement of water through plant roots, usually at night, from wet to dry regions of
50 the plant rooting volume driven by differences in water potential. This passive process can favor
51 plant survival during droughts by tapping into deep soil layers having relatively higher water
52 potential and redistributing water to the shallow root zone (upward HR) (Barron - Gafford et al.,
53 2017; Brooks et al., 2002; Dawson, 1993; Domec et al., 2010; Nadezhdina et al., 2015; Nicola et
54 al., 2020; Nicola & Ram, 2022; Prieto et al., 2012). During wet seasons, HR can redistribute
55 water from wet surface soil into deeper, drier soil (downward HR), supplementing the infiltration
56 process in recharging deeper soil layers (Bleby et al., 2010; Fu et al., 2016; Hultine et al., 2003;
57 Scott et al., 2008). Despite its potential role in regulating plant and ecosystem productivity,
58 nutrient cycling and soil microbial activity (Grünzweig et al., 2022; Sardans & Peñuelas, 2014),
59 HR is often ignored in ecosystem models.

60 Hydraulic redistribution has been observed across diverse ecosystems and plant species
61 (Nadezhdina et al., 2010; Neumann & Cardon, 2012; Priyadarshini et al., 2016; Yu et al., 2013),
62 and has been interpreted as structuring dryland plant communities, regulating ecosystem
63 productivity, and enhancing resilience to climate extremes (Barron - Gafford et al., 2021;
64 Barron - Gafford et al., 2017; Hafner et al., 2020; Lee et al., 2018). The dynamics of HR are
65 influenced by various biotic (rooting architecture, plant capacitance, transpiration demand,



66 senescence, and dormancy), abiotic factors (soil hydraulic characteristics, soil moisture status),
67 and climatic conditions (precipitation) (Katul & Siqueira, 2010; Prieto et al., 2012; Wei et al.,
68 2022). While several studies have examined HR in deserts and semi-arid ecosystems and
69 reported upward HR during dry events (Xing - Ming Hao et al., 2013; Lee et al., 2018; Scott et
70 al., 2008; Yu et al., 2013) and downward HR following precipitation {Hultine, 2003 #2122},
71 studies focusing on fine-scale temporal variations in HR across different soil depths and multiple
72 years remain limited. Additionally, a quantitative understanding of how precipitation magnitude
73 and frequency influence HR rates, key limiting factors in dryland ecosystems, remain poorly
74 understood.

75 In this study, we explicitly test two hypotheses: (1) Direction of HR: Upward HR should be
76 the dominant form of HR in dryland ecosystem. This is due to the recharge of deeper soil layers
77 from precipitation which can retain moisture for longer periods, and during dry periods roots
78 facilitate the movement of this retained water to the drier surface soils. (2) HR-precipitation
79 relationship: upward HR should decline following precipitation events, reaching its maximum
80 rates during prolonged dry periods as the drought create steep water potential gradients between
81 deeper, moist soil layers and the drier surface layers, facilitating the upward movement of water.

82 Soil moisture dynamics are governed by a complex interplay of forces that drive water
83 movement through the soil profile. The primary drivers include matric potential (capillary and
84 adsorptive forces binding water to soil particles), gravitational potential (driving downward
85 drainage), and potential gradients that induce processes like HR (Caldwell et al., 1998). These
86 forces collectively determine water retention, redistribution, and plant availability (Hillel, 2003).
87 Isolating their individual contributions from field soil moisture data is challenging, as their
88 effects are concurrent and modulated by soil properties, root activity, and atmospheric



89 conditions. Consequently, data-model fusion approach, integrating process-based model with soil
90 moisture data, provides a robust framework to isolate and quantify HR, offering a more
91 mechanistic and quantitative understanding.

92 Several modeling studies have incorporated various HR schemes into process-based models
93 to improve understanding of hydrological and ecological processes (Amenu & Kumar, 2008; Fu
94 et al., 2016; Lee et al., 2018; Quijano & Kumar, 2015; Ryel et al., 2002; Tang et al., 2015; Wu et
95 al., 2020; Zheng & Wang, 2007). However, realistic representation and estimation of parameters
96 related to HR remains a challenge, as neither the magnitude of HR nor its associated parameters
97 can be directly observed in the soil (Quijano & Kumar, 2015; Ryel et al., 2002). As a result, most
98 models rely on default HR parameter values from Ryel et al. (2002) (Fabian et al., 2010; Zheng
99 & Wang, 2007) or estimated parameters using soil moisture data during specific periods of time
100 when upward or downward HR is assumed negligible, such as wet or dry season (Amenu &
101 Kumar, 2008; Fu et al., 2018; Fu et al., 2016; Yan & Dickinson, 2014). The challenges in direct
102 measurements, and reliance on assumed parameter values, constitute key gaps in our
103 understanding of HR dynamics.

104 To address these gaps, we focused on piñon-juniper (PJ) woodlands, the most widespread
105 semiarid ecosystem in the US. PJ woodlands are spatially widespread, ecologically important,
106 temporally dynamic, and structurally unique dryland ecosystem in the western US, spanning 10
107 US states and 40 million hectares across the American Southwest (Eastburn et al., 2024; Romme
108 et al., 2009). Despite their importance, HR has not been previously reported in PJ woodlands.
109 However, our continuous root sap flux measurements provided direct evidence of HR in both
110 piñon and juniper roots, indicated by sustained negative root sap flux during nighttime at the
111 study site (Fig. S.1).



112 In this study, we used the process-based Terrestrial ECOsystem (TECO) model to (i) develop
113 and implement a data assimilation approach to incorporate HR into the TECO model; (ii)
114 quantify and characterize the magnitude and dynamics of HR across multiple soil depths; and
115 (iii) analyze the temporal patterns of HR and its relationship with precipitation magnitude and
116 frequency. The TECO model is a well-established ecosystem model that integrates ecological
117 processes to simulate carbon, water, and energy fluxes within terrestrial ecosystems (Weng &
118 Luo, 2008). We employed data assimilation to constrain the TECO model including HR using
119 four years of soil moisture data measured at multiple soil depths, encompassing both wet and dry
120 periods.

121 **2. Data and Methods**

122 **2.1 Study site and data**

123 Our modeling study utilized data from a PJ woodland plot (Lat. 35.642, Long. -104.607,
124 elevation 1925 m) located in New Mexico, USA, and previously described in Schwinning et al.
125 (2020). The site is a private ranch covering an area of over 6800 hectares that was ungrazed
126 from 2012 through the measurement period used for this study and is characterized by a semi-
127 arid climate. Mean annual precipitation of the site is approximately 460 mm, with the majority
128 falling between May and October, and a mean annual temperature of 10.5 °C. The soil texture at
129 the site varies with depth, ranging from loam to clay loam. The vegetation consists of distinct
130 tree clusters dominated by piñon pine (*Pinus edulis* (Englem.)) and juniper (*Juniperus*
131 *monosperma* (Englem.) Sarg.) separated by open areas of bare soil and herbaceous cover.

132 SWC was continuously monitored using multi-sensor frequency domain capacitance probes
133 (Decagon EC-5) installed at four depths (5, 15, 30 and 60 cm), in four soil pits under the tree
134 canopies. All sensors were monitored every minute by a datalogger (model CR6, Campbell



Scientific), storing 15-minute averages were stored by a data logger. For this study, we used the average SWC across all four pits. Each sensor was calibrated in the lab before installation for both air and water frequency. Significant shifts in soil temperature can affect both soil permittivity and the response of capacitance sensors, potentially confounding the small fluctuations in VWC caused by HR. Therefore, temperature correction factors were applied to the measured VWC at each depth, using the nearest measured temperature, following the method described by Saito et al. (2009).

2.2 Modeling framework

TECO is a process-based ecosystem model (Hou et al., 2021; Jiang et al., 2018; Weng & Luo, 2008), and has evolved from the TCS model (Luo & Reynolds, 1999). The model has four major components: canopy photosynthesis, plant growth, soil water dynamics, and soil carbon transfers. The canopy photosynthesis and soil water dynamics submodels run at the hourly time step whereas the plant growth and soil carbon submodels run at the daily time step. The model is driven by seven environmental variables, including precipitation (mm), wind speed (m s^{-1}), solar radiation (W m^{-2}), air and soil temperature (C), relative humidity (%), and vapor pressure deficit (kPa). The detailed description of TECO model is available (Weng & Luo, 2008) and only the brief description of soil water dynamics is provided here.

The soil profile is divided into 10 layers, with varying thickness: 5 cm for the first layer, 10, 15, and 30 cm for the second, third, and fourth layers respectively, and 20 cm for the fifth to tenth layers. SWC in each layer results from the mass balance between influx and efflux, with changes primarily attributed to vertical unsaturated flow, transpiration, precipitation, runoff, and drainage. Evaporation depletes water from the first two soil layers, while transpiration depletes water from all soil layers containing roots, allocated based on root fraction in each layer (Eq. 8).



158 Given the predominantly arid conditions of the study site, runoff and drainage were found
 159 negligible. Thus, water movement between soil layers is simulated as follows:

$$160 \quad \frac{dW_i}{dt} = \frac{dF_i}{dz} - E_i - T_i \quad (1)$$

161 where W_i is the water storage (cm) in layer i , t is time (h), F_i is net unsaturated flow of water into
 162 layer i (cm h^{-1}), z is vertical thickness, E_i and T_i are evaporation and transpiration water loss from
 163 layer i (cm h^{-1}).

164 The unsaturated soil water movement is simulated vertically according to modified form of
 165 Buckingham-Darcy's law (Campbell, 1985) (Eq. 2), with Brooks (1965) equation (Eq. 4)
 166 estimating hydraulic conductivity and soil water retention curve (SWRC) to simulate soil water
 167 potential (Ψ).

$$168 \quad \frac{dF_i}{dz} = K(\theta_i) \left(\frac{d\Psi_i}{dz} + 1 \right) \quad (2)$$

169 where $K(\theta_i)$ is the unsaturated soil hydraulic conductivity (cm h^{-1}) for SWC θ ($\text{cm}^3 \text{cm}^{-3}$) in layer
 170 i , Ψ_i is soil water matric potential (MPa) in layer i , and z is the vertical thickness (cm) of the soil.

$$171 \quad K(\theta_i) = K_s \left[\frac{\theta_i - \theta_r}{\theta_s - \theta_r} \right]^{(2m+3)} \quad (3)$$

172 where, K_s is the soil saturated hydraulic conductivity (cm h^{-1}), m is the pore size distribution
 173 index, θ_s and θ_r are saturated and residual SWC ($\text{cm}^3 \text{cm}^{-3}$)

$$174 \quad \frac{\theta - \theta_r}{\theta_s - \theta_r} = \left(\frac{\Psi}{\Psi_b} \right)^{-1/m} \quad (4)$$

175 Ψ_b is the soil air entry water potential.

176 To quantify the direction and magnitude of HR, we integrated the HR model by Ryel et al.
 177 (2002) into equation 1 of TECO model (presented in equation 5). This HR model empirically
 178 describes HR flux based on the soil water potential gradient between two soil layers (Eq. 6). HR
 179 was assumed to occur only at night, with its occurrence controlled by solar radiation instead of



fixed day and night hours. Daytime starts as solar radiation exceeds 10 W m^{-2} , thereby inhibiting HR since the water potential gradient typically favors water movement from roots to canopy to meet transpiration demand during the day. This pattern is evident in Fig. S1, where under low or zero solar radiation, root sap flux was found to be negative, indicating water movement away from the root zone which is an indicator of occurrence of HR at the study site. Using these assumptions, the net water movement into soil layer i from other soil layers can be expressed as:

$$\frac{dW_i}{dt} = \frac{dF_i}{dz} - E_i - T_i + H_i \quad (5)$$

$$H_i = C_{RT} \sum (\Psi_j - \Psi_i) \max(c_i, c_j) \frac{R_i R_j}{1 - R_x} D_{tran} \quad (6)$$

$$c_i = \frac{1}{1 + \left(\frac{\Psi_i}{\Psi_{50}}\right)^b} \quad (7)$$

$$R_i = \frac{R_0}{1 + \left(\frac{d}{d_{50}}\right)^a} \quad (8)$$

Where in Eq 6, H_i is the net water redistributed by roots into layer i (cm h^{-1}), C_{RT} is the maximum radial soil-root conductance of the entire active root system for water ($\text{cm MPa}^{-1} \text{ h}^{-1}$), Ψ is soil matric potential (MPa), c_i is a factor reducing soil-root conductance based on Ψ_i , R_i is the fraction of active roots in layer i , R_0 is the average vertically summed root dry mass from the bottom to the root zone to the soil surface, and D_{tran} is a factor reducing water movement among layers by roots while plant is transpiring and is assumed to be 1.0 during the night when transpiration is minimal and 0.0 during day. $R_x = R_i$ when $\theta_i > \theta_j$ or $R_x = R_j$ when $\theta_j > \theta_i$. In Eq 7, Ψ_{50} is the soil water potential (MPa) where conductance is reduced by 50% and b is an empirical constant. In Eq 8, d is soil depth (cm), and d_{50} is the soil depth at the median of the root distribution and a is a shape parameter (Table 1). The Brooks (1965) model for SWRC was utilized to simulate soil water potential (Ψ), facilitating the development of soil water potential



201 gradients necessary for HR by tree roots (Eq. 4). Due to lack of site-specific parameters, the
202 default values of b and Ψ_{50} were used as 3.22 and -1 MPa, respectively (Ryel et al., 2002).

203 **2.3 Data assimilation for parameters estimation**

204 We used Bayesian probabilistic inversion to calibrate parameters associated with soil
205 hydraulics, where posterior probability density functions of parameters are obtained from prior
206 knowledge about the parameters and the error between model and observations. According to
207 Mosegaard and Sambridge (2002), Bayesian inversion can be summarized by the following
208 equation:

$$209 \quad p(c|Z) \propto p(Z|c) p(c) \quad (9)$$

210 where $p(c|Z)$ is posterior probability density function of model parameters c ; $p(Z|c)$ is a
211 likelihood function of parameters c ; $p(c)$ is prior probability density function of parameters c . We
212 assumed that the prediction errors were normally distributed and uncorrelated, hence, the
213 likelihood function, $p(Z|c)$, was calculated as follows:

$$214 \quad p(Z|c) \propto \exp\left\{-\sum_{i=1}^k \frac{(Z_i - X_i)^2}{2\sigma_i^2}\right\} \quad (10)$$

215 where Z_i is observed VWC at i^{th} soil layer, X_i is VWC simulated by TECO at a corresponding
216 soil depth; σ_i^2 is the variance of a measurement at a soil layer; k is the total number of soil layers.

217 To generate the posterior distributions, we first specified the priors of the parameters to be
218 uniformly distributed over the intervals specified in Table 1. We put constraints on parameters
219 based on the literature. The initial set of parameters was randomly selected within the prior
220 parameter ranges. Once we specified parameter ranges, we used the Metropolis-Hastings (M-H)
221 algorithm (Hastings, 1970; Metropolis et al., 1953), a Markov chain Monte Carlo method, to



sample from the posterior parameter distribution. To generate a parameter set, we ran M-H algorithm in two steps: proposing step and a moving step. In the proposing step, a new parameter set c^{new} was generated from a previously accepted parameter set c^{k-l} through a proposal distribution ($c^{new}|c^{k-l}$):

$$c^{new} = c^{k-l} + r \times \frac{c^{max} - c^{min}}{D} \quad (11)$$

The value of $P(c^{k-l}|c^{new})$ was then compared with a random number U from 0 to 1. Parameter set c^{new} was accepted if $P(c^{k-l}|c^{new}) \geq U$, otherwise c^k was set to c^{k-l} . In the moving step, a probability of acceptance $P(c^{k-l}|c^{new})$ was calculated as in the following (Marshall et al., 2004):

$$P(c^{k-l}|c^{new}) = \min \left\{ 1, \frac{p(Z|c^{new})p(c^{new})}{p(Z|c^{k-1})p(c^{k-1})} \right\} \quad (12)$$

The M-H algorithm was repeated for 50,000 simulations, and then all accepted parameters values were used to generate the probability distribution functions (Xu et al., 2006). Finally, before each model simulation with optimized parameters, we ran the model for 200 years, a spin-up period that was long enough to obtain stable carbon stock as an initial condition for these simulations.

To evaluate the impact of HR on soil moisture dynamics in a PJ woodland, we conducted two multi-year simulations using two configurations of the TECO model: TECO+HR (with HR) and default TECO (HR turned off). To distinguish the influence of HR from soil hydraulic properties, we adopted a data assimilation approach focused on calibrating only the TECO+HR model. We calibrated TECO+HR model using soil moisture data measured at 5, 15, 30, and 60 cm depths over a four-year period. The range of prior parameter values was informed by available soil texture data for the study site (Brooks, 1965; Carsel & Parrish, 1988). Within this



range, we optimized depth-specific soil hydraulic parameters to achieve a close match between modeled and observed soil moisture (Table 1).

After calibrating the TECO+HR model, we deactivated the HR process and ran simulations with the same optimized parameters to generate the default TECO scenario. This approach allowed us to ensure that differences in soil moisture dynamics between TECO+HR and default TECO simulations were attributable solely to the presence or absence of HR. The motivation to calibrate only TECO+HR model, rather than the default TECO is to avoid parameter compensation for unresolved processes (Luo & Schuur, 2020), in which the absence of HR could lead to unrealistic adjustments of soil hydraulic parameters to indirectly capture its effects.

2.4 Statistical analyses

Model performance was assessed by comparing simulated outputs with observed data during full simulation periods (2018- 2021), dry, and wet periods, defined as days without and with rainfall events, respectively. Evaluation was conducted using statistical metrics, including root mean square error (RMSE), and absolute mean error (MAE).

$$RMSE = \sqrt{\frac{1}{n} \sum_{i=1}^n (m_i - o_i)^2} \quad (13)$$

$$MAE = \frac{1}{n} \sum_{i=1}^n |m_i - o_i| \quad (14)$$

Where: o_i represents observed values, m_i represents modeled values, and n represents the number of data points.



261
262

Table 1: Parameters constrained using data assimilation in the TECO model from soil moisture data from 2018 to 2021.

Parameters	Symbols	Constrained values	Range	Units	References
Saturated water content	θ_s	0.34/0.38/0.36/0.33	[0.3, 0.4]	$\text{cm}^3 \text{cm}^{-3}$	Calibrated
Residual water content	θ_r	0.05/0.07/0.06/0.03	[0, 0.08]	$\text{cm}^3 \text{cm}^{-3}$	Calibrated
Saturated hydraulic conductivity	K_s	0.14/0.29/0.30/0.70	[0.1, 2]	cm h^{-1}	Calibrated
Pore size distribution	m	0.89/0.66/0.88/0.84	[0, 1]	-	Calibrated
Air entry water potential	ψ_b	96/60/50/40	[0, 100]	cm	Calibrated
Maximum radial soil-root conductance	C_{RT}	0.022	[0, 1]	$\text{cm MPa}^{-1} \text{h}^{-1}$	Calibrated
Soil Ψ where root conductivity reduced by 50%	ψ_{50}	-1.0	-	MPa	(Ryel et al., 2002)
Empirical constant	b	3.22	-	-	(Ryel et al., 2002)
Average vertically summed root dry mass	R_0	0.90	-	kg m^{-2}	(Schwinning et al., 2020)
Soil depth at the median of the root distribution	D_{50}	25	-	cm	(Schwinning et al., 2020)
Root distribution shape parameter	a	2.2	-	-	(Schwinning et al., 2020)

Four values represent parameters in the four modeled SWC at depths of 5, 15, 30, and 60 cm, respectively.



263 3. Results

264 3.1 Parameter estimation via data assimilation and water mass balance

265 The data assimilation approach, using observational SWC data to constrain the model,
266 yielded well-constrained soil hydraulic parameters (Table 1; Fig. S2 and S3). The resulting
267 posterior probability density functions, characterized by sharp peaks, narrow spread, and
268 consistency across soil depth support the reliability and accuracy of these calibrated parameter
269 values. Additionally, soil water mass balance of soil profile was conserved before and after
270 incorporating the HR process into the TECO model (Fig. S4). The key components of the water
271 budget: precipitation, evapotranspiration, and changes in soil water content remained balanced,
272 ensuring that the model accounted for all water fluxes. Furthermore, the sum of HR across all
273 soil layers (10 layers) was consistently equal to zero, further ensuring that no water was
274 artificially introduced or lost from the system.

275

276

277

278

279

280

281

282

283



3.2 Observed and simulated soil moisture

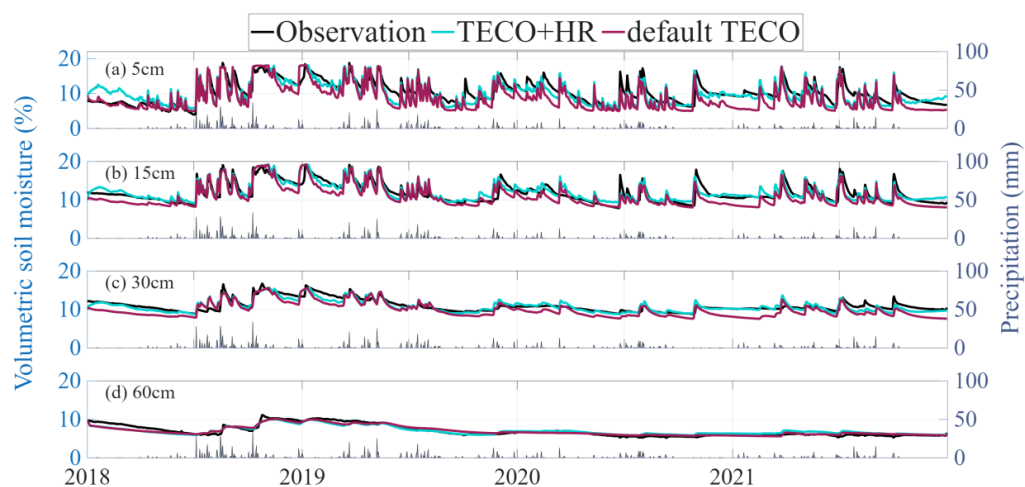


Figure 1a-d: Observed and simulated soil volumetric water content for the year 2019 (January 1, 2018 to December 31, 2021) at soil depths of 5 cm (a), 15 cm (b), 30 cm (c), and 60 cm (d). Vertical bars indicate

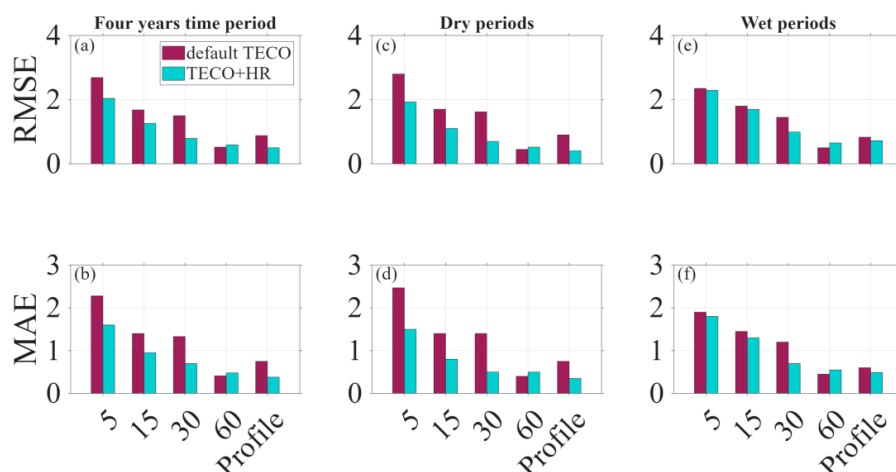


Figure 2: Model performance for soil moisture across different depths (5, 15, 30, and 60 cm, and 0-60 cm integrated soil profile), considering temporal variations in soil moisture conditions. Root Mean Square Error (RMSE) and Mean Absolute Error (MAE) are presented for the complete time series (a, b), dry periods (c, d), and wet periods (e, f).

298 The data assimilation-constrained models, generally captured both the magnitude and
 299 dynamics of observational data, reproducing seasonal variations in soil moisture across four soil
 300 depths. While TECO+HR simulation showed an improvement in the overall model performance,
 301 the impact of HR was mostly pronounced during dry periods (Fig. 1 and 2). We further examined
 302 diurnal soil moisture fluctuations (Fig. S5) and found that TECO+HR successfully reproduced
 303 the observed diurnal cycles, whereas the default TECO failed to capture this pattern, suggesting
 304 that the observed diurnal variability was likely driven by HR. Additionally, we compared min-
 305 max normalized soil matric potential at 15, 30, and 60 cm with simulations derived from Eq. (4)
 306 (Fig. S6). Both models reproduced the general trends of the observations, suggesting that the
 307 simulated soil water potential gradients were consistent with measurement.

308 Moreover, during periods of limited precipitation, the TECO+HR (blue lines) consistently
 309 maintained higher soil moisture compared to default TECO (red lines), aligning closer to
 310 observation particularly in the topsoil layers (Fig. 1a-c). Following precipitation events, the



311 default TECO and TECO+HR simulations converged, suggesting the minimal influence of HR
312 under wet conditions at the study site. However, as surface soil moisture decreased following
313 precipitation, the two simulations diverged again, with TECO+HR maintaining higher moisture
314 levels in the topsoil layers, highlighting the role of HR in maintaining soil moisture during
315 prolonged drought.

316 The incorporation of HR into TECO resulted in reductions in model errors. During dry
317 periods, the RMSE decreased by 25, 43, and 52% at 5, 15, and 30 cm soil depths, respectively.
318 However, limited improvement was observed at 60 cm soil depth. Correspondingly, the MAE
319 was reduced by 30, 53, and 60% at 5, 15, and 30 cm, respectively. Over the entire study period,
320 RMSE decreased by 24, 25, and 47% at 5, 15, and 30 cm, with MAE reductions were 29, 34, and
321 55% at the same depths (Fig. 2a-d). Overall soil profile performance improved as well, with
322 RMSE and MAE reductions over 40% for both the four-year simulation and dry periods. These
323 improvements during dry periods are especially important, as roots are most vulnerable to
324 drought. By mitigating soil water deficits in surface layers, HR could reduce the risk of hydraulic
325 failure, thereby supporting plant species survival and it could enable better prediction of
326 ecosystem responses to water stress, such as carbon uptake (Domec et al., 2010), and
327 evapotranspiration (Zhu et al., 2017). In contrast, during wet periods, HR had minimal influence
328 on soil moisture (Fig. 2e, f).

329

330



3.3 HR simulations

Model simulation revealed distinct patterns of HR dynamics across soil depths and temporal scales. Fig. 3 illustrates these patterns over two timescales: a short-term, diurnal pattern (Fig. 3a), and a long-term perspective from 2018 to 2021 (Fig. 3b). HR is a process with both a source and a sink for water movement. In the Fig. 3, positive HR suggest that a soil layer is gaining water (sink), whereas negative HR values suggest that the layer is losing water (source).

The short-term modeling analysis highlights diurnal pattern of HR during dry conditions and a precipitation event (Fig. 3a). For instance, on July 23, 2018, during a dry period, upward HR occurred, moving water from deeper (> 100 cm) to shallower (0-30 cm) soil layers. However, following a precipitation event on July 24, 2018 (12 mm), this pattern shifted. The top 5 and 15 cm layers showed negative HR and a decrease in the upward HR rate, respectively, acting as a water source for deeper layers. At the same time, deeper soil layers showed a decline in negative HR rates, suggesting signs of receiving water likely from the topsoil layers. The sum of HR across all soil layers remained

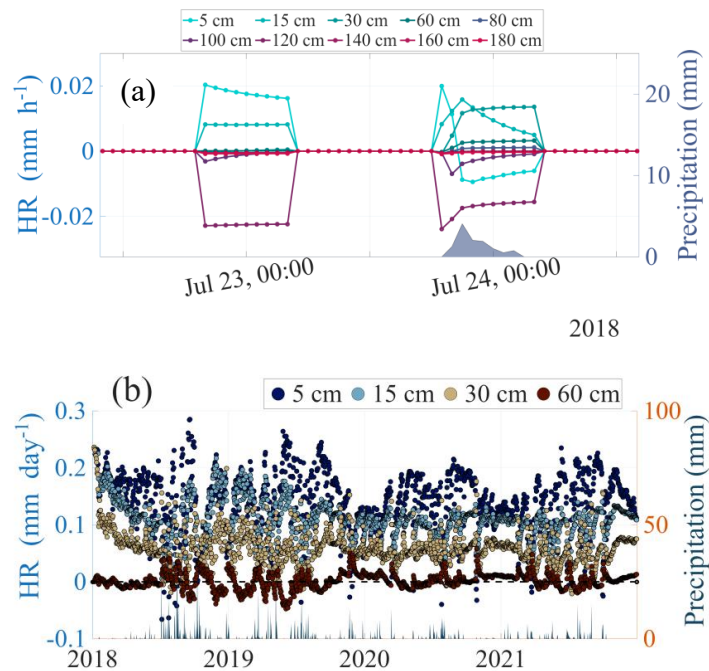


Figure 3: Temporal dynamics of hydraulic redistribution (HR). (a) diurnal pattern of modeled hydraulic redistribution across soil depths from July 22-24, 2018. The graph illustrates HR patterns during a dry period followed by a precipitation event. Colored lines represent different soil depths. (b) long-term daily HR trends and precipitation from January 2018 to December 2021. The blue shaded area represents precipitation (right y-axis).



354 zero, confirming that HR redistributed water rather than adding to the system. Consequently,
 355 downward HR from the topsoil supplemented infiltration, enhancing water movement into
 356 deeper soil layers, reflected by a decrease in the negative HR rates at depths and an increase in
 357 the positive HR rate at 30 cm (Fig. 3a).

358 While our model simulates HR across 10 soil layers, we present long-term results for only
 359 the top four soil layers (5, 15, 30, and 60 cm) to enable direct comparison with the available
 360 observed soil moisture data. A clear seasonal pattern emerged, with HR generally intensifying
 361 during dry periods (Fig. 3b).

362 Our model showed that upward HR was predominantly occurring in up to top 30 cm of soil
 363 profile, with values ranging from -0.066 to 0.29 mm d^{-1} in each soil layer and an average of 0.30
 364 mm d^{-1} across the top 30 throughout the study period. Downward HR, while less pronounced,
 365 moved water only from the 5 cm soil layer during monsoon seasons and large precipitation
 366 events (e.g., July 2018, 2019, 2020, and 2021; Fig. 3b). In contrast, 60 cm soil layer typically
 367 exhibits a negative HR during dry periods, acting as a water source for upper layers, and positive
 368 HR during wet periods, suggesting occasional water input from surface layers ranging from -
 369 0.096 to 0.059 mm d^{-1} (mean 0.0015 mm d^{-1}). Moreover, integrated soil profile (top 60 cm of
 370 soil profile), showed that upward HR was the dominant form of HR throughout the year, ranging
 371 from 0.10 to 0.53 mm d^{-1} with a mean value 0.31 mm d^{-1} (Fig. S7).

372

373

374

375



376 3.4 Precipitation influences on HR

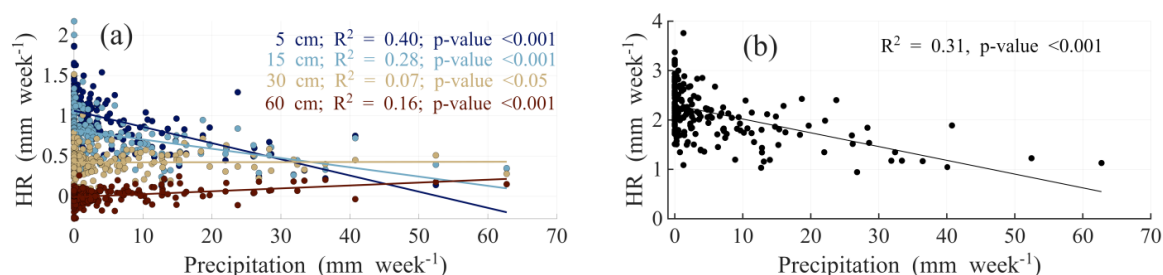


Figure 4: Relationships between HR and precipitation. (a) Weekly HR rates versus weekly precipitation amounts at different soil depths (5, 15, 30, and 60 cm). Trend lines and R² values are shown for each depth. (b) Weekly HR (0–60 cm) versus weekly precipitation, with trend line and R² value.

377 The model results showed a significant linear relationship between weekly HR and
 378 precipitation (mm week⁻¹) (Fig. 4a-b). In the topsoil layers (5 cm and 15 cm), negative
 379 correlations were observed (R² = 0.40 and 0.28 respectively, both p-values < 0.001). At the depth
 380 of 30 cm, a weak negative correlation exists (R² = 0.07, p-value < 0.05), suggesting a less
 381 pronounced reduction in HR rates with increasing precipitation. Interestingly, at 60 cm depth,
 382 HR was positively correlated (R² = 0.16, p-value < 0.001) with precipitation.

383 Additionally, when considering total HR across the soil profile (0-60 cm), the modeled data
 384 showed a significant negative correlation with precipitation was observed (R² = 0.31, p-values <
 385 0.001, Fig. 4b). This suggests that overall HR activity decreased as precipitation increases,
 386 highlighting the stronger potential impact of HR on soil moisture during drier conditions.

387 However, we observed a considerable variability in HR was observed across the range of
 388 precipitation (Fig. 4a-b), which could be attributed to the rainfall frequency, event size, and the
 389 duration of dry periods between rainfall events (Fig. 5). These results suggest that HR may
 390 increase not only with reduced precipitation frequency but also as the interval between



consecutive rainfall events lengthen (Fig. 5).
HR was lowest under conditions of high
rainfall frequency and shorter dry spells,
progressively increasing to its peak in the
absence of rainfall. However, as the drought
period extended beyond 30 days, HR

declined, suggesting potential limitation on
availability of deeper water to sustain HR.
This variability is further illustrated through
three scenarios (Fig. S7): 1) Following a

rainfall event (28 mm on July 5, 2018), HR in the top 60 cm of soil profile was minimal at 0.13
mm d⁻¹, indicating limited driving force for water redistribution when soil moisture was
abundant. 2) During a transition period between rainfall events (July 5-10, 2018), HR gradually
increased but remained moderate, ranging from 0.13 to 0.20 mm d⁻¹, suggesting a progressive
activation of the redistribution process as soil began to dry. 3) During a prolonged dry period
(November 23-30, 2018), HR peaked at 0.20-0.52 mm d⁻¹, demonstrating enhanced
redistribution activity in response to the development of soil moisture gradients.

4. Discussion

4.1 Patterns of hydraulic redistribution

Our findings support the hypothesis that upward HR is the dominant form of HR in dryland
ecosystems due to limited precipitation amount and sporadic rainfall events (Fig. S7). This
prevalence of upward water movement is characteristic of semi-arid regions, where deep-rooted
plants often redistribute water from moist deeper layers to drier surface soils during periods of

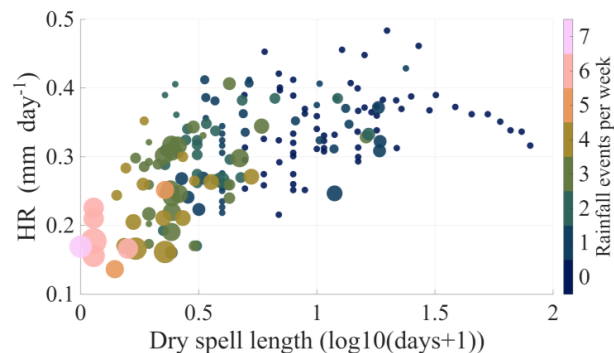


Figure 5: Relationship between weekly mean hydraulic redistribution (mm d⁻¹), dry spell length between two rainfall events (log₁₀(days+1)). The color scale indicates the number of rainfall events per week, while marker size represents the weekly precipitation amount (mm week⁻¹). Dry spell length denotes the number of rainless days between two consecutive precipitation events.



414 water stress (Caldwell et al., 1998; Ryel et al., 2002). Notably, the most pronounced HR
 415 occurred in the topsoil layer (5, 15, and 30 cm), (Fig. 3b), which can be attributed to vertical root
 416 distribution, with over 50% of root biomass concentrated in the top 30 cm ($D_{50} = 25$ cm) of the
 417 soil profile (Fig. S8). This pattern aligns with findings of Xing Ming Hao et al. (2013), which
 418 suggest that deeper root distributions extend HR to deeper soil layers, while shallower root
 419 systems enhance HR in the topsoil layers due to higher root density and activity (Fig. S8).

420 Our model simulations estimated HR rates in range 0.10-0.53 mm d⁻¹ for top 60 cm soil
 421 depth (Fig. S7), values that fall within the broader range of 0.04 to 3.2 mm d⁻¹ reported in the
 422 comprehensive review by Neumann and Cardon (2012) but exceed the upper limit of the 95%
 423 confidence interval of 0.014-0.475 mm d⁻¹ reported by Yang et al. (2022) for desert or sparsely
 424 vegetated ecosystems, which synthesized empirical observations and modeling estimates of
 425 average water movement attributed to HR.

426 The direct impact of HR on hydrological processes should be evident in the soil profile water
 427 content. We tested this by comparing SWC model simulations in TECO with and without HR
 428 processes, to observed SWC time series at four depths (Fig. 6). We found that cumulative effects
 429 of HR on soil moisture vary with depth, primarily due to the non-uniform root biomass
 430 distribution throughout the soil profile (Fig. S8). The most pronounced effects of HR were
 431 observed in the topsoil layers (5, 15, and 30 cm), where average daily water content increased by
 432 up to 60% compared to simulation without HR. This increase was driven by upward HR,
 433 especially during dry-down periods (Fig. 3b, and 6).

434

435

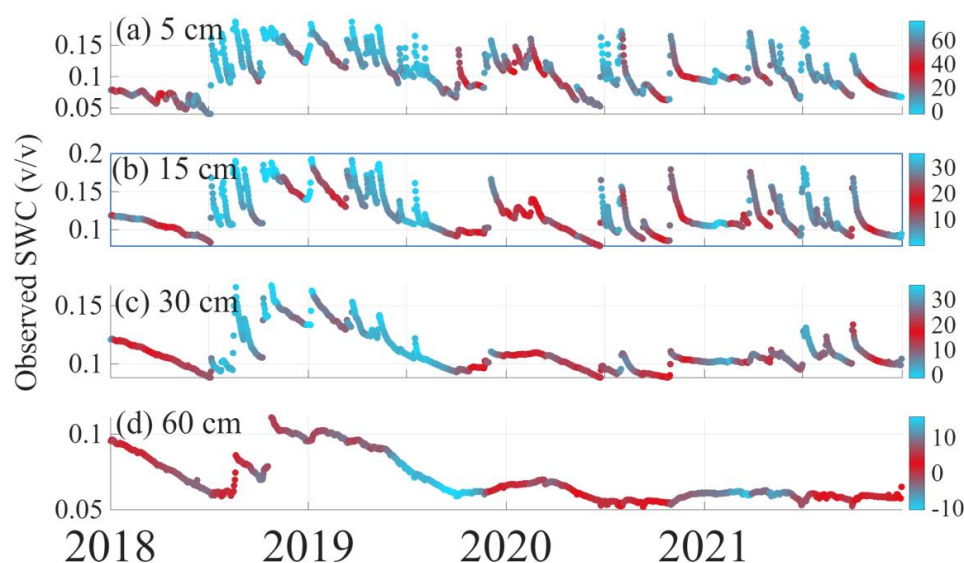


Figure 6: Relative change in soil water content (SWC) (%) compared to observed SWC at 5, 15, 30, and 60 cm depths. The color gradient represents the magnitude of relative change in SWC, calculated as $(HR - \text{No HR}) / \text{No HR} \times 100$, with HR and No HR indicating simulations with and without hydraulic redistribution."

4.2 Effects of precipitation variability on HR

Our findings support the hypothesis that the precipitation pattern significantly ($p\text{-value} < 0.001$) influences the magnitude and variability of HR (Figs. 3a-b, and S7). The rate of HR in the topsoil profile (< 60 cm) exhibited a consistent response pattern to precipitation events, characterized by sharp decreases following large rainfall, and gradual recovery to pre-rain levels. These dynamics were particularly evident during years with frequent large precipitation events (2018-2019), where HR rates oscillated between 0.04- and 0.20-mm d^{-1} . For instance, after an 18 mm rainfall event on July 24, 2018, HR rates dropped below 0.15 mm d^{-1} before recovering to 0.30 mm d^{-1} within 10 days. This pattern suggests a recharging effect: initially, infiltrating rainwater increases soil water potential in both shallow and deep layers, reducing the gradient between shallow and deep layers and temporarily suppressing HR (Xing Ming Hao et al., 2013).



447 However, as water redistributes through the soil profile, new hydraulic gradients develop,
448 leading to enhanced HR activity. In this phase, roots actively redistribute water from newly
449 moistened deep layers to drier shallow layers, consistent with findings from Yu and D'Odorico
450 (2014) and (Ryel et al., 2002).

451 Our model predicted that HR rates were generally higher during rainless periods compared to
452 rainfall periods within a given year. For instance, during the prolonged dry period in 2020 (driest
453 year with few small precipitation events), HR rates remained consistently high, 0.17-0.40 mm d⁻¹,
454 with minimal fluctuations. The consistent high HR rates, likely arises from more pronounced
455 soil water potential gradients derived from sustained plant water demand and surface evaporation
456 in the absence of frequent precipitation (Fu et al., 2016; Meinzer et al., 2004).

457 Seasonal trends include higher HR rates (0.12-0.53 mm d⁻¹) during the drier periods
458 (typically November to May) and lower rates (0.10-0.30 mm d⁻¹) during the monsoon season
459 (usually June to October) (Fig. S7). This seasonality underscores the influence of both
460 precipitation patterns and potential evapotranspiration on HR dynamics, highlighting that HR is
461 likely more pronounced during drier seasons when soil moisture gradients are likely to be more
462 substantial due to reduced precipitation and potentially higher evaporative demand (Fu et al.,
463 2016; Scott et al., 2008; Yu & D'Odorico, 2014).

464 **4.3 Limitation and future perspectives**

465 While our modeling study provides valuable insights into HR dynamics in PJ woodlands,
466 several limitations should be noted: (1) The model does not account for inter-annual changes in
467 vegetation cover or species composition. Variations in plant functional types and leaf area index
468 may influence soil moisture and HR, and incorporating these dynamics could improve long-term
469 simulations. (2) We focused on dominant tree species at our study site, but other plants may



470 benefit from water redistributed by these trees, affecting ecosystem water dynamics. (3) We did
471 not include stem water refilling or nighttime transpiration reported by Howard et al. (2009);
472 Neumann et al. (2014) which could influence the magnitude of HR. (4) Finally, future studies
473 could explore the role of HR in regulating ecosystem functions, such as carbon exchange and
474 evapotranspiration, and examine whether incorporating HR improves predictions of ecosystem
475 carbon dynamics proportionally to its effects on soil water.

476 **5. Conclusions**

477 This study demonstrates the role of hydraulic redistribution (HR) in soil water dynamics in
478 piñon-juniper woodlands. By integrating HR processes and observations into the Terrestrial
479 Ecosystem Model (TECO) via data assimilation, we successfully constrained model soil
480 hydraulics parameters and improved simulations of soil water content across multiple depths,
481 particularly in shallow soil layers (0–30 cm) and during dry periods. Our model results showed
482 that HR rates vary with the length of dry spells between rainfall events, generally decreasing
483 with increasing precipitation magnitude and frequency, with HR rates ranging from 0.10 to 0.50
484 mm d⁻¹ as conditions transitioned from wet to dry. Consequently, HR increased soil moisture in
485 topsoil layers by up to 60% during dry periods, with upward HR emerging as the dominant flux,
486 especially in the top 30 cm. These findings underscore the potential influence of HR during dry
487 periods and highlight its role in sustaining soil water availability for vegetation. Future research
488 should explore how HR-mediated water redistribution affects ecosystem functions including
489 carbon exchange, and evapotranspiration.



490 **6. Acknowledgments**

491 This research is supported by US Department of Energy’s Office of Biological and
492 Environmental Research, Environmental System Science (ESS) Program (DE-SC0023514 to
493 WTP, MEL and YL) and the modeling effort used data collected with funding from the US
494 National Science Foundation (IOS [1557176](#) to MEL, WTP, and Susan Schwinning). Additional
495 support from US NSF grants (DEB 2242034, DEB 2406930, and DEB 2425290), the US
496 Department of Energy’s Terrestrial Ecosystem Sciences Grant DE-SC0023514, subcontract
497 CW55561 from Oak Ridge National Laboratory to Cornell University, and the “NYS Connects:
498 Climate Smart Farms & Forestry” project, funded by the USDA, the New York State Department
499 of Environmental Conservation, and the New York State Department of Agriculture and
500 Markets. This research is also part of AI-CLIMATE: “AI Institute for Climate-Land Interactions,
501 Mitigation, Adaptation, Tradeoffs and Economy”, funded by the USDA National Institute of
502 Food and Agriculture (NIFA), the NSF National AI Research Institutes Competitive Award (No.
503 2023-67021-39829) and Swiss National Foundation, Award#P500PN_206603 for their financial
504 support of this collaboration.

505 **7. Competing interests**

506 The authors declare no competing interests.

507 **8. Author contributions**

508 **AKC:** Conceptualization, Methodology, Data curation, Formal analysis, Writing - original draft.
509 **YZ:** Conceptualization, Writing - review & editing. **MC:** Writing - review & editing. **HD:**
510 Writing - review & editing, Data curation. **ML:** Conceptualization, Data curation, Supervision,
511 Writing - review & editing, Funding acquisition, Project administration. **WP:** Conceptualization,
512 Data curation, Supervision, Funding acquisition, Writing - review & editing. **YL:**



513 Conceptualization, Supervision, Project administration, Funding acquisition, Writing - review &
 514 editing.

515 **9. Code availability**

516 The Terrestrial ECOSystem (TECO) model used in this paper is available on GitHub at
 517 https://github.com/aneeshchandel/TECO_HR.

518 **10. Data availability**

519 The data supporting the findings of this study are available within the manuscript. Additional
 520 data may be available upon request from the corresponding author, subject to compliance with
 521 relevant data protection and privacy regulations.

522 **11. Supporting Information**

523 Supporting information accompanying this manuscript is available as a separate Word file. It
 524 includes supplementary figures referenced in the main text.

525 **References**

- 526 Amenu, G. G., & Kumar, P. (2008). A model for hydraulic redistribution incorporating coupled
 527 soil-root moisture transport. *Hydrology and Earth System Sciences*, 12(1), 55-74.
 528 <https://doi.org/DOI.10.5194/hess-12-55-2008>
 529 Barron - Gafford, G. A., Knowles, J. F., Sanchez - Cañete, E. P., Minor, R. L., Lee, E., Sutter,
 530 L., Tran, N., Murphy, P., Hamerlynck, E. P., & Kumar, P. (2021). Hydraulic
 531 redistribution buffers climate variability and regulates grass - tree interactions in a
 532 semiarid riparian savanna. *Ecohydrology*, 14(3), e2271.
 533 Barron - Gafford, G. A., Sanchez - Cañete, E. P., Minor, R. L., Hendryx, S. M., Lee, E., Sutter,
 534 L. F., Tran, N., Parra, E., Colella, T., & Murphy, P. C. (2017). Impacts of hydraulic
 535 redistribution on grass - tree competition vs facilitation in a semi - arid savanna. *New*
 536 *Phytologist*, 215(4), 1451-1461. <https://doi.org/10.1111/nph.14693>
 537 Beer, C., Reichstein, M., Tomelleri, E., Ciais, P., Jung, M., Carvalhais, N., Rödenbeck, C.,
 538 Arain, M. A., Baldocchi, D., & Bonan, G. B. (2010). Terrestrial gross carbon dioxide
 539 uptake: global distribution and covariation with climate. *Science*, 329(5993), 834-838.
 540 Bleby, T. M., Mcelrone, A. J., & Jackson, R. B. (2010). Water uptake and hydraulic
 541 redistribution across large woody root systems to 20 m depth. *Plant, Cell & Environment*,
 542 33(12), 2132-2148.



- 543 Brooks, J. R., Frederick, C. M., Rob, C., & Jillian, W. G. (2002). Hydraulic redistribution of soil
 544 water during summer drought in two contrasting Pacific Northwest coniferous forests.
 545 *Tree physiology*, 22(15-16), 1107-1117. <https://doi.org/10.1093/treephys/22.15-16.1107>
 546 Brooks, R. H. (1965). *Hydraulic properties of porous media*. Colorado State University.
 547 Caldwell, M. M., Dawson, T. E., & Richards, J. H. (1998). Hydraulic lift: consequences of water
 548 efflux from the roots of plants. *Oecologia*, 113(2), 151-161.
 549 <https://doi.org/10.1007/s004420050363>
 550 Campbell, G. S. (1985). *Soil physics with BASIC: transport models for soil-plant systems*.
 551 Elsevier.
 552 Carsel, R. F., & Parrish, R. S. (1988). Developing joint probability distributions of soil water
 553 retention characteristics. *Water Resources Research*, 24(5), 755-769.
 554 Dawson, T. E. (1993). Hydraulic lift and water use by plants: implications for water balance,
 555 performance and plant-plant interactions. *Oecologia*, 95(4), 565-574.
 556 <https://doi.org/10.1007/BF00317442>
 557 Domec, J. C., King, J. S., Noormets, A., Treasure, E., Gavazzi, M. J., Sun, G., & McNulty, S. G.
 558 (2010). Hydraulic redistribution of soil water by roots affects whole-stand
 559 evapotranspiration and net ecosystem carbon exchange. *New Phytol*, 187(1), 171-183.
 560 <https://doi.org/10.1111/j.1469-8137.2010.03245.x>
 561 Eastburn, J. F., Campbell, M. J., Dennison, P. E., Anderegg, W. R., Barrett, K. J., Fekety, P. A.,
 562 Flake, S. W., Huffman, D. W., Kannenberg, S. A., & Kerr, K. L. (2024). Ecological and
 563 climatic transferability of airborne lidar-driven aboveground biomass models in Piñon-
 564 Juniper woodlands. *GIScience & Remote Sensing*, 61(1), 2363577.
 565 Fabian, G. S., Sandra, J. B., William, A. H., Frederick, C. M., & Guillermo, G. (2010). Hydraulic
 566 lift in a Neotropical savanna: Experimental manipulation and model simulations.
 567 *Agricultural and Forest Meteorology*, 150(4), 629-639.
 568 <https://doi.org/10.1016/j.agrformet.2010.02.001>
 569 Fu, C., Wang, G., Bible, K., Goulden, M. L., Saleska, S. R., Scott, R. L., & Cardon, Z. G. (2018).
 570 Hydraulic redistribution affects modeled carbon cycling via soil microbial activity and
 571 suppressed fire. *Glob Chang Biol*, 24(8), 3472-3485. <https://doi.org/10.1111/gcb.14164>
 572 Fu, C. S., Wang, G. L., Goulden, M. L., Scott, R. L., Bible, K., & Cardon, Z. G. (2016).
 573 Combined measurement and modeling of the hydrological impact of hydraulic
 574 redistribution using CLM4.5 at eight AmeriFlux sites. *Hydrology and Earth System*
 575 *Sciences*, 20(5), 2001-2018. <https://doi.org/10.5194/hess-20-2001-2016>
 576 Grünzweig, J. M., De Boeck, H. J., Rey, A., Santos, M. J., Adam, O., Bahn, M., Belnap, J.,
 577 Deckmyn, G., Dekker, S. C., & Flores, O. (2022). Dryland mechanisms could widely
 578 control ecosystem functioning in a drier and warmer world. *Nature ecology & evolution*,
 579 6(8), 1064-1076.
 580 Hafner, B. D., Hesse, B. D., Bauerle, T. L., & Grams, T. E. (2020). Water potential gradient, root
 581 conduit size and root xylem hydraulic conductivity determine the extent of hydraulic
 582 redistribution in temperate trees. *Functional Ecology*, 34(3), 561-574.
 583 Hao, X. M., Chen, Y. N., Guo, B., & Ma, J. X. (2013). Hydraulic redistribution of soil water in
 584 *Populus euphratica* Oliv. in a central Asian desert riparian forest. *Ecohydrology*, 6(6),
 585 974-983. <https://doi.org/10.1002/eco.1338>
 586 Hao, X. M., Li, W. H., Guo, B., & Ma, J. X. (2013). Simulation of the effect of root distribution
 587 on hydraulic redistribution in a desert riparian forest. *Ecological research*, 28, 653-662.



- 588 Hastings, W. K. (1970). Monte Carlo sampling methods using Markov chains and their
 589 applications.
- 590 Hillel, D. (2003). *Introduction to environmental soil physics*. Elsevier.
- 591 Hou, E., Litvak, M. E., Rudgers, J. A., Jiang, L., Collins, S. L., Pockman, W. T., Hui, D., Niu, S.,
 592 & Luo, Y. (2021). Divergent responses of primary production to increasing precipitation
 593 variability in global drylands. *Glob Chang Biol*, 27(20), 5225-5237.
 594 <https://doi.org/10.1111/gcb.15801>
- 595 Howard, A. R., Van Iersel, M. W., Richards, J. H., & Donovan, L. A. (2009). Night - time
 596 transpiration can decrease hydraulic redistribution. *Plant, Cell & Environment*, 32(8),
 597 1060-1070.
- 598 Hultine, K. R., Cable, W. L., Burgess, S. S. O., & Williams, D. G. (2003). Hydraulic
 599 redistribution by deep roots of a Chihuahuan Desert phreatophyte. *Tree physiology*,
 600 23(5), 353-360. <https://doi.org/DOI> 10.1093/treephys/23.5.353
- 601 Jiang, J., Huang, Y., Ma, S., Stacy, M., Shi, Z., Ricciuto, D. M., Hanson, P. J., & Luo, Y. (2018).
 602 Forecasting responses of a northern peatland carbon cycle to elevated CO₂ and a gradient
 603 of experimental warming. *Journal of Geophysical Research: Biogeosciences*, 123(3),
 604 1057-1071. <https://doi.org/10.1002/2017jg004040>
- 605 Katul, G. G., & Siqueira, M. B. (2010). Biotic and abiotic factors act in coordination to amplify
 606 hydraulic redistribution and lift. *The New Phytologist*, 187(1), 3-6.
- 607 Lee, E., Kumar, P., Barron-Gafford, G. A., Hendryx, S. M., Sanchez-Cañete, E. P., Minor, R. L.,
 608 Colella, T., & Scott, R. L. (2018). Impact of hydraulic redistribution on multispecies
 609 vegetation water use in a semiarid savanna ecosystem: An experimental and modeling
 610 synthesis. *Water Resources Research*, 54(6), 4009-4027.
 611 <https://doi.org/10.1029/2017wr021006>
- 612 Luo, Y., & Reynolds, J. F. (1999). Validity of extrapolating field CO₂ experiments to predict
 613 carbon sequestration in natural ecosystems. *Ecology*, 80(5), 1568-1583.
 614 <https://doi.org/DOI> 10.1890/0012-9658(1999)080[1568:Voefce]2.0.Co;2
- 615 Luo, Y., & Schuur, E. A. G. (2020). Model parameterization to represent processes at unresolved
 616 scales and changing properties of evolving systems. *Glob Chang Biol*, 26(3), 1109-1117.
 617 <https://doi.org/10.1111/gcb.14939>
- 618 Marshall, L., Nott, D., & Sharma, A. (2004). A comparative study of Markov chain Monte Carlo
 619 methods for conceptual rainfall - runoff modeling. *Water Resources Research*, 40(2).
- 620 Meinzer, F. C., Brooks, J. R., Bucci, S., Goldstein, G., Scholz, F. G., & Warren, J. M. (2004).
 621 Converging patterns of uptake and hydraulic redistribution of soil water in contrasting
 622 woody vegetation types. *Tree physiology*, 24(8), 919-928. <https://doi.org/DOI>
 623 10.1093/treephys/24.8.919
- 624 Metropolis, N., Rosenbluth, A. W., Rosenbluth, M. N., Teller, A. H., & Teller, E. (1953).
 625 Equation of state calculations by fast computing machines. *The journal of chemical*
 626 *physics*, 21(6), 1087-1092.
- 627 Mosegaard, K., & Sambridge, M. (2002). Monte Carlo analysis of inverse problems. *Inverse*
 628 *problems*, 18(3), R29.
- 629 Nadezhkina, N., David, T. S., David, J. S., Ferreira, M. I., Dohnal, M., Tesar, M., Gartner, K.,
 630 Leitgeb, E., Nadezhdin, V., & Cermak, J. (2010). Trees never rest: the multiple facets of
 631 hydraulic redistribution. *Ecophysiology*, 3(4), 431-444.



- 632 Nadezhdina, N., Ferreira, M. I., Conceição, N., Pacheco, C. A., Häusler, M., & David, T. S.
 633 (2015). Water uptake and hydraulic redistribution under a seasonal climate: long - term
 634 study in a rainfed olive orchard. *Ecohydrology*, 8(3), 387-397.
- 635 Neumann, R. B., & Cardon, Z. G. (2012). The magnitude of hydraulic redistribution by plant
 636 roots: a review and synthesis of empirical and modeling studies. *New Phytol*, 194(2),
 637 337-352. <https://doi.org/10.1111/j.1469-8137.2012.04088.x>
- 638 Neumann, R. B., Cardon, Z. G., TESHERA - LEVYE, J., Rockwell, F. E., Zwieniecki, M. A., &
 639 Holbrook, N. M. (2014). Modelled hydraulic redistribution by sunflower (*Helianthus*
 640 *annuus* L.) matches observed data only after including night - time transpiration. *Plant*,
 641 *Cell & Environment*, 37(4), 899-910.
- 642 Nicola, M., Nicola, M., Matteo, C., Roberto, C., & Ram, O. (2020). Fixed and variable
 643 components of evapotranspiration in a Mediterranean wild-olive - grass landscape
 644 mosaic. *Agricultural and Forest Meteorology*, 280.
 645 <https://doi.org/10.1016/j.agrformet.2019.107769>
- 646 Nicola, M., & Ram, O. (2022). Rhizosphere water content drives hydraulic redistribution:
 647 Implications of pore-scale heterogeneity to modeling diurnal transpiration in water-
 648 limited ecosystems. *Agricultural and Forest Meteorology*, 312.
 649 <https://doi.org/10.1016/j.agrformet.2021.108720>
- 650 Prăvălie, R. (2016). Drylands extent and environmental issues. A global approach. *Earth-Science*
 651 *Reviews*, 161, 259-278.
- 652 Prieto, I., Armas, C., & Pugnaire, F. I. (2012). Water release through plant roots: new insights
 653 into its consequences at the plant and ecosystem level. *New Phytol*, 193(4), 830-841.
 654 <https://doi.org/10.1111/j.1469-8137.2011.04039.x>
- 655 Priyadarshini, K. V. R., Prins, H. H. T., de Bie, S., Heitkönig, I. M. A., Woodborne, S., Gort, G.,
 656 Kirkman, K., Ludwig, F., Dawson, T. E., & de Kroon, H. (2016). Seasonality of
 657 hydraulic redistribution by trees to grasses and changes in their water - source use that
 658 change tree–grass interactions. *Ecohydrology*, 9(2), 218-228.
 659 <https://doi.org/10.1002/eco.1624>
- 660 Quijano, J. C., & Kumar, P. (2015). Numerical simulations of hydraulic redistribution across
 661 climates: The role of the root hydraulic conductivities. *Water Resources Research*,
 662 51(10), 8529-8550. <https://doi.org/10.1002/2014wr016509>
- 663 Romme, W. H., Allen, C. D., Bailey, J. D., Baker, W. L., Bestelmeyer, B. T., Brown, P. M.,
 664 Eisenhart, K. S., Floyd, M. L., Huffman, D. W., & Jacobs, B. F. (2009). Historical and
 665 modern disturbance regimes, stand structures, and landscape dynamics in pinon–juniper
 666 vegetation of the western United States. *Rangeland Ecology & Management*, 62(3), 203-
 667 222.
- 668 Ryel, R., Caldwell, M., Yoder, C., Or, D., & Leffler, A. (2002). Hydraulic redistribution in a
 669 stand of *Artemisia tridentata*: evaluation of benefits to transpiration assessed with a
 670 simulation model. *Oecologia*, 130(2), 173-184. <https://doi.org/10.1007/s004420100794>
- 671 Saito, T., Fujimaki, H., Yasuda, H., & Inoue, M. (2009). Empirical temperature calibration of
 672 capacitance probes to measure soil water. *Soil Science Society of America Journal*, 73(6),
 673 1931-1937.
- 674 Sardans, J., & Peñuelas, J. (2014). Hydraulic redistribution by plants and nutrient stoichiometry:
 675 Shifts under global change. *Ecohydrology*, 7(1), 1-20.
- 676 Schwinning, S., Litvak, M. E., Pockman, W. T., Pangle, R. E., Fox, A. M., Huang, C. W., &
 677 McIntire, C. D. (2020). A 3-dimensional model of *Pinus edulis* and *Juniperus*



- monosperma root distributions in New Mexico: implications for soil water dynamics.
Plant and Soil, 450(1-2), 337-355. <https://doi.org/10.1007/s11104-020-04446-y>
- Scott, R. L., Cable, W. L., & Hultine, K. R. (2008). The ecohydrologic significance of hydraulic redistribution in a semiarid savanna. *Water Resources Research*, 44(2).
<https://doi.org/Artn> W02440
- 10.1029/2007wr006149
- Tang, J., Riley, W. J., & Niu, J. (2015). Incorporating root hydraulic redistribution in CLM 4.5: Effects on predicted site and global evapotranspiration, soil moisture, and water storage. *Journal of Advances in Modeling Earth Systems*, 7(4), 1828-1848.
- Ukkola, A. M., De Kauwe, M. G., Roderick, M. L., Burrell, A., Lehmann, P., & Pitman, A. J. (2021). Annual precipitation explains variability in dryland vegetation greenness globally but not locally. *Global Change Biology*, 27(18), 4367-4380.
- Wei, L., Qiu, Z., Zhou, G., Zuecco, G., Liu, Y., & Wen, Y. (2022). Soil water hydraulic redistribution in a subtropical monsoon evergreen forest. *Science of The Total Environment*, 835, 155437. <https://doi.org/10.1016/j.scitotenv.2022.155437>
- Weng, E., & Luo, Y. (2008). Soil hydrological properties regulate grassland ecosystem responses to multifactor global change: A modeling analysis. *Journal of Geophysical Research: Biogeosciences*, 113(G3). <https://doi.org/Artn> G03003
- 10.1029/2007jg000539
- Wu, H., Fu, C., Wu, H., & Zhang, L. (2020). Influence of the dry event induced hydraulic redistribution on water and carbon cycles at five AsiaFlux forest sites: A site study combining measurements and modeling. *Journal of Hydrology*, 587, 124979.
<https://doi.org/ARTN> 124979
- 10.1016/j.jhydrol.2020.124979
- Xu, T., White, L., Hui, D., & Luo, Y. (2006). Probabilistic inversion of a terrestrial ecosystem model: Analysis of uncertainty in parameter estimation and model prediction. *Global Biogeochemical Cycles*, 20(2). <https://doi.org/10.1029/2005gb002468>
- Yan, B., & Dickinson, R. E. (2014). Modeling hydraulic redistribution and ecosystem response to droughts over the Amazon basin using Community Land Model 4.0 (CLM4). *Journal of Geophysical Research: Biogeosciences*, 119(11), 2130-2143.
<https://doi.org/10.1002/2014jg002694>
- Yang, G., Huang, L., & Shi, Y. (2022). Magnitude and determinants of plant root hydraulic redistribution: A global synthesis analysis. *Frontiers in Plant Science*, 13, 918585.
- Yu, K. L., & D'Odorico, P. (2014). Climate, vegetation, and soil controls on hydraulic redistribution in shallow tree roots. *Advances in Water Resources*, 66, 70-80.
<https://doi.org/10.1016/j.advwatres.2014.02.003>
- Yu, T., Feng, Q., Si, J., Xi, H., Li, Z., & Chen, A. (2013). Hydraulic redistribution of soil water by roots of two desert riparian phreatophytes in northwest China's extremely arid region. *Plant and Soil*, 372(1), 297-308. <https://doi.org/10.1007/s11104-013-1727-8>
- Zheng, Z., & Wang, G. (2007). Modeling the dynamic root water uptake and its hydrological impact at the Reserva Jaru site in Amazonia. *Journal of Geophysical Research: Biogeosciences*, 112(G4).
- Zhu, S., Chen, H., Zhang, X., Wei, N., Shangguan, W., Yuan, H., Zhang, S., Wang, L., Zhou, L., & Dai, Y. (2017). Incorporating root hydraulic redistribution and compensatory water



722 uptake in the Common Land Model: Effects on site level and global land modeling.
723 *Journal of Geophysical Research: Atmospheres*, 122(14), 7308-7322.
724