# Operational chemical weather forecasting with the ECCC online Regional Air Quality Deterministic Prediction System version 023 (RAQDPS023) - Part 2: Multi-year prospective and retrospective performance evaluation

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**Abstract**

The operational online version of the Regional Air Quality Deterministic Prediction System (RAQDPS) is a chemical weather forecast system that has been employed ~~operationally~~ by Environment and Climate Change Canada (ECCC) since 2009. It is run twice daily to produce 72 hour forecasts of hourly 10 km abundance fields of three key predictands, $NO_2$, $O_3$, and $PM_{2.5}$ total mass, as well as other gas-phase chemical species, $PM_{2.5}$ chemical components, and dry and wet deposition for Canada, the contiguous U.S. and Alaska, and northern Mexico. Version 023 of the RAQDPS (RAQDPS023) went into service at ECCC in December 2021 and was replaced by the RAQDPS025 in June 2024. A companion paper by Moran et al. (2026~~5~~) describes the RAQDPS023 in detail. In this paper we present the results of a five-year performance evaluation of prospective and retrospective annual air quality (AQ) simulations made with the RAQDPS023. The annual simulations considered were the first year of operational RAQDPS023 forecasts in 2021/22 and four years of retrospective annual simulations for the 2013–2016 period that used historical, year-specific emissions. This version of the RAQDPS023, which did not include biomass burning (BB) emissions, is referred to in the text as the RAQDPS-OP023. Forecasts made by the RAQDPS-FW023, a duplicate operational system to the RAQDPS-OP023 except for the addition of ~~near real-time (NRT)~~time-dependent ~~biomass burning (~~BB~~)~~ emissions, were also evaluated for the 2021/22 period. A ~~NRT~~ near-real-time- measurement data set consisting of hourly $NO_2$, $O_3$, and $PM_{2.5}$ surface measurements for Canada and the U.S. was used for the 2021/22 evaluation, whereas a much more extensive set of air-chemistry and precipitation-chemistry measurements was used for the 2013–2016 RAQDPS-OP023 evaluations. Some evaluation results were also compared with results for the 2010–2019 period for forecasts made by earlier operational versions of the RAQDPS and with evaluation results for several peer AQ forecast models. In addition to looking at a number of highly aggregated "headline" scores, many stratified analyses were also performed, including evaluations by network, season, month, hour of day, region, and land-use type. Consideration of simulations for multiple years with the same model but~~and~~ year-specific input emissions helped to identify systematic model errors by reducing the influence of year-to-year variations in meteorology and emissions, and a comprehensive

evaluation for many additional species for 2013–2016 supported by stratified analyses provided diagnostic insights that allowed the scientific basis for the RAQDPS-OP023 forecasts to be assessed (~~i.~~e.g., were ~~"~~the right answers obtained for the right reasons?~~"~~). Although one confounding factor for this study was the sizable reduction in the emissions of some pollutants in North America that occurred from 2013 to 2021, it was found that the trends in AQ observations over this period agreed with the year-specific description of emissions used for the five annual simulations from a rank-ordered perspective.

While RAQDPS-OP023 evaluation scores for hourly $NO_2$ and $O_3$ volume mixing ratio forecasts were found to be competitive with peer models and often met suggested performance benchmarks for the five simulation years, another key finding was that the RAQDPS-OP023 forecasts consistently underpredicted hourly $PM_{2.5}$ total mass concentrations for all months in 2021/22 and for the majority of months in 2013–2016. The largest underpredictions occurred in summer and at rural stations, whereas overpredictions often occurred in the cold season at urban stations. The model also missed the observed bimodality in monthly $PM_{2.5}$ concentrations and exaggerated the observed diurnal variations in hourly $PM_{2.5}$ concentrations. Additional evaluations with daily $PM_{2.5}$ chemical composition measurements and daily gravimetric $PM_{2.5}$ total mass measurements from the U.S. $PM_{2.5}$ mass monitoring network were also examined to better understand the hourly $PM_{2.5}$ underpredictions. Consistent overpredictions of elemental carbon and sea salt concentrations and underpredictions of sulfate concentration were identified, but scores for predictions of daily gravimetric $PM_{2.5}$ total mass were better than those for hourly $PM_{2.5}$ total mass, directing attention to differences in measurement methods. $SO_2$ and $HNO_3$ levels were also found to be overpredicted in general while $NH_3$ levels were underpredicted: these three gas-phase species are all $PM_{2.5}$ precursors, which raises concerns about some process representations in the model such as those for sulfur oxidation and gas-phase dry deposition. As well, springtime $O_3$ levels were underpredicted while isoprene levels were consistently overpredicted in all seasons. The impact of BB emissions on predictions of $NO_2$, $O_3$, and $PM_{2.5}$ was also characterized in detail by comparing evaluation results for the 2021/22 RAQDPS-OP023 and RAQDPS-FW023 forecasts. Negligible impact was found for monthly $NO_2$ forecasts when BB emissions were included, but monthly $O_3$ forecast scores for the RAQDPS-FW023 were modestly improved and monthly $PM_{2.5}$ forecast scores were markedly improved from July to September 2021, as well as summer and annual scores. Taken together, the results of this comprehensive multi-year evaluation point to a number of RAQDPS023 system components where improvements are desirable. These results also provide a strong benchmark against which to compare the performance of future versions of the RAQDPS.

## 1 Introduction

The use of operational short-range air quality (AQ) forecast systems to predict tomorrow's air quality, also referred to as chemical weather, has expanded rapidly over the last two decades (e.g., Kukkonen et al., 2012; Zhang et al., 2012a,b; WMO, 2020; Kajino et al., 2018; Brasseur and Kumar, 2021; Wang et al., 2022; Williams et al., 2022; Colette et al., 2025; Li et al., 2025). Environment and Climate Change Canada (ECCC), Canada's federal environment ministry, which is responsible for operational weather forecasting in Canada, began to make operational regional AQ forecasts

in 2001. Since that time, numerous upgrades and improvements have been made to this system (see companion paper by Moran et al., 2026~~5~~). Version 23 of the ECCC Regional Air Quality Deterministic Prediction System (RAQDPS023) became the Canadian operational, continental-scale chemical weather forecast system for North America on 1 December 2021 (Moran et al., 2021b) and continued in this role until June 2024 (CMC-RAQDPS-025, 2024). One version of the RAQDPS023, which did not include biomass burning (BB) emissions, is referred to in the rest of the paper as the RAQDPS-OP023. There was also a clone of the RAQDPS-OP023 forecast system named the RAQDPS-FW023, which was identical except for the addition of time-dependent ~~near real-time (NRT) biomass burning (~~BB~~)~~ emissions (where "FW"="FireWork"; Pavlovic et al., 2016; Chen et al., 2019; Chen and Menelaou, 2021). The RAQDPS-OP023 and RAQDPS-FW023 were both run by ECCC twice per day on a 10 km continental grid to produce 72 hour forecasts of hourly surface concentration fields of ozone ($O_3$), nitrogen dioxide ($NO_2$), particulate matter with aerodynamic diameter smaller than 2.5 µm ($PM_{2.5}$), and other chemical species and compounds. These forecasts were disseminated internally to ECCC forecast offices and also directly to the public via a public ECCC website (https://weather.gc.ca/firework/index_e.html, last access: 6 April 2026). This goal of this paper is to present the results of a multi-year prospective and retrospective performance evaluation of RAQDPS-OP023 AQ predictions, which ~~both~~ quantifies predictive skill, including the impact of BB emissions, and provides an evaluation benchmark against which the performance of future RAQDPS versions can be compared.

The comparison of AQ model predictions with AQ measurements for a chosen simulation period allows the modelling system's performance to be assessed, weaknesses to be identified, and, for some more comprehensive evaluations, potential improvements to be suggested. Initially, such model performance evaluations considered retrospective simulations (or hindcasts) for individual or multiple AQ models that were being used in a regulatory environment (e.g., Dennis and Downton, 1984; Venkatram et al., 1988; Dennis et al., 1993; Tesche et al., 2006; van Loon et al., 2007; Smyth et al., 2009; Solazzo et al., 2012a,b; Yahya et al., 2014; Im et al., 2015a,b; Appel et al., 2021). In the regulatory context, however, there is typically interest in both model skill for the simulation period considered and model skill in predicting AQ changes in response to changes in input emissions or meteorological conditions (e.g., Dennis and Downton, 1984; Gilliland et al., 2008; Pun et al., 2008; Dennis et al., 2010; Foley et al., 2015; Koo et al., 2015; Colette et al., 2017). For AQ forecasting, by contrast, forecast skill under current conditions is the primary concern (Steyn and Galmarini, 2008; Dennis et al., 2010).

Zhang et al. (2012a,b) have provided a review of the history of both regional AQ forecasting in North America and Europe and global AQ forecasting, including performance evaluation approaches, up to 2012. Kukkonen et al. (2012) provided a similar overview for the same period but focussed on operational European regional-scale AQ forecast~~ing~~ models. Zhang et al. (2012a) noted that the 1998 development in the U.S. of the Aerometric Information Retrieval Now (AirNow) program (https:\\www.airnow.gov, last access: 8 April 2026), a NRT data repository and dissemination hub for North American AQ measurements supplied by more than 100 monitoring agencies in the U.S. and Canada, was revolutionary for North American AQ forecasting since it allowed forecasting teams to obtain immediate feedback on model performance (e.g., McKeen et al., 2005, 2007, 2009; Eder et al., 2006, 2009, 2010; Mathur et al., 2008;

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Chuang et al., 2011; Chai et al., 2013; Lee et al., 2017; Chen et al., 2021; Campbell et al., 2022; Williams et al., 2022). A current example of the use of AirNow data for short-term model performance evaluation is ~~in~~ an ongoing multi-model AQ forecast evaluation for North America that is led by ECCC under the umbrella of the World Meteorological Organization (WMO) Global Air quality Forecasting and Information System (GAFIS) initiative (see Sect. 4.3). AirNow data are also used for objective analyses (e.g., Robichaud and Ménard, 2014; Robichaud et al., 2016) and for chemical data assimilation (e.g., Pagowski et al., 2010; Ma et al., 2021).

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The NRT measurements available from AirNow, however, have three important disadvantages. First, measurements are only available for six chemical compounds: $NO_2$, $O_3$, CO, $SO_2$, $PM_{2.5}$, and $PM_{10}$. Second, AirNow is a "meta-network" since the multiple agencies contributing measurement data may employ different instruments and sampling techniques, each with their own biases and errors, to measure the same chemical species. As a consequence, there can be considerable heterogeneity in a combined AirNow measurement data set vs. the uniformity expected of a typical measurement network data set. And third, the AirNow measurements must be viewed as preliminary since they have not undergone the quality assurance/quality control (QA/QC) procedures normally applied by the monitoring agencies before they release new data sets.

The more traditional source of AQ measurement data is to obtain them directly from the lead agency for a monitoring network or from an AQ measurement data clearinghouse such as AQS or NAtChem (see Table S2a in the Supplement). However, these finalized network data sets suffer from the significant disadvantage of only being available anywhere from three months to years after ~~sampling~~samples were collected, since some AQ measurements require post-sampling calibration while others (e.g., from filterpacks, annular denuders, passive samplers, and precipitation samplers) must undergo laboratory analysis after collection followed by network QA/QC procedures. But such finalized data sets do have three important advantages over the NRT AirNow data. First, they include measurements of many additional chemical species, including more trace gases such as nitric acid ($HNO_3$), ammonia ($NH_3$), and some individual volatile organic compounds (VOCs), $PM_{2.5}$ chemical components, and major inorganic ions in precipitation. Second, even for the six pollutants that are reported to AirNow, not all North American stations that measure these species report to the AirNow data centre. And third, these finalized data sets have been QA/QCed before release. For example, Chai et al. (2013) compared AirNow and AQS hourly $O_3$ measurements for 2010 and showed scatterplots of differences between the two data sets for a one-month period. The issue of availability, however, means that these finalized AQ network data sets cannot be used for the immediate evaluation of AQ forecasts, that is, for prospective AQ simulations, but they are preferable for the evaluation of historical or retrospective AQ simulations since they permit more comprehensive evaluations of predictions of the atmospheric chemical environment using a broader range of QA/QCed measurement data.

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A paper by Dennis et al. (2010) proposed a framework for evaluating AQ model performance that consists of four evaluation types: operational; diagnostic; dynamic; and probabilistic. The first two evaluation types are the most relevant for evaluating deterministic AQ forecasts. Operational evaluations address the basic question of how well model predictions of chemical concentrations and deposition agree with observations of chemical concentrations and

deposition. To do this they use routine measurements of a small set of air-chemistry species, and, infrequently, additional air-chemistry, precipitation-chemistry, and meteorological parameters, to calculate standard statistical performance metric~~asure~~s (e.g., Table A2). Diagnostic evaluations, on the other hand, are less common. They ~~and~~ are used to evaluate model inputs and process representations by considering many additional relevant observations such as precursor concentrations, pollutant concentrations aloft, PM chemical composition and size distributions, and meteorological parameters that have a direct impact on pollutant concentrations such as temperature, planetary boundary layer (PBL) height, vertical wind profiles, cloud cover, and precipitation (e.g., Vautard et al., 2012). Diagnostic evaluations can address three additional important questions. First, is agreement between model predictions and observations the result of chance or of good scientific understanding and representation of atmospheric dynamics, physics, chemistry, and emissions? Put another way, is the model getting the right answers for the right reasons? Second, are differences between model predictions and observations due to errors in model input fields or to gaps or errors in model process representations or to computational ~~factors~~errors? And third, can the identification of the sources of differences between the model predictions and observations be used as a guide to improve the model? Dynamic evaluations, the third of these evaluation types, assess model skill in quantifying the impact of changes in input emissions or meteorology (e.g., Gilliland et al., 2008; Godowitch et al., 2010; Foley et al., 2015). And fourth, probabilistic evaluations examine the uncertainty and level of confidence in model predictions (e.g., Hanna et al., 2005; Mallet and Sportisse, 2006; Galmarini et al., 2010; Kioutsioukis et al., 2025).

Many operational evaluations have considered only a small number of observed species even if finalized measurement data sets were used (e.g., Chai et al., 2013; Pan et al., 2014; Marécal et al., 2015; Wagner et al., 2015; Lee et al., 2017; Campbell et al., 2022; and Williams et al., 2022). Given the complexity of atmospheric chemistry related to secondary pollutants such as $O_3$ and to the multiple chemical components of PM (e.g., Sillman, 1999; Meng et al., 1997; Bachmann, 2013), however, such limited evaluations will not provide insights into the reasons for poor model performance. A comprehensive operational evaluation, on the other hand, which makes use of the full range of available AQ measurements, can consider nearly complete mass budgets for some chemical families such as sulphur species or oxidized nitrogen species and hence may be considered closer to a diagnostic evaluation. Comprehensive operational evaluations, however, are relatively uncommon. For example, Huang et al. (2021) reviewed over 300 peer-reviewed articles that reported evaluation results from AQ modelling studies for China and found that very few considered more than seven pollutants. Nevertheless, examples of comprehensive operational evaluations include publications by Biswas et al. (2001), Hogrefe et al. (2001a,b), Zhang et al. (2006a,b), Cai et al. (2008), Yu et al. (2008), Zhang et al. (2009a), ~~Hogrefe et al. (2015),~~ Yahya et al. (2014, 2015), Hogrefe et al. (2015), Tessum et al. (2015), Zhang et al. (2016), Chen et al. (2021), and Wang et al. (2021). Lastly, examples of diagnostic evaluations include Zhang et al. (2006c, 2009b), Godowitch et al. (2011), Gan et al. (2015), Knote et al. (2015), Galmarini et al. (2021), and Clifton et al. (2023).

This paper presents the results of an operational performance evaluation of both AQ forecasts and AQ hindcasts made by the RAQDPS-OP023 chemical weather forecast system. Evaluations were performed for five simulation years:

(i) the first year of RAQDPS-OP023 (and RAQDPS-FW023) forecasts from 1 June 2021 to 31 May 2022, which used projected anthropogenic input emissions files; and (ii) four years of retrospective annual simulations for the 2013–2016 period performed with the equivalent RAQDPS-OP024 forecast system (same system but ported to a new computer; see companion paper by Moran et al., 2026~~5~~) but using historical, year-specific input emissions files. Note that from 1 June to 30 November 2021 the RAQDPS-OP023 and RAQDPS-FW023 systems were run in a parallel (i.e., pre-operational) mode beside the RAQDPS-OP022 and RAQDPS-FW022 systems that were operational at that time before ~~being~~ the former were promoted to operational status on 1 December 2021 (Moran et al., 2026). In addition, the performance of a decade of operational forecasts made by earlier RAQDPS versions from 1 January 2010 to 30 June 2019 is also examined both to show the evolution of forecast skill over this period and to allow comparison with RAQDPS-OP023 scores. Note that RAQDPS-FW023 retrospective simulations for 2013–2016 ~~have not been considered here~~were not available due to the incompatibility ~~with~~ for this earlier period of version 4.1 of the Canadian Forest Fire Emissions Prediction System (CFFEPS), which was used by the RAQDPS-FW023 to calculate BB emissions. CFFEPS v4.1, which depends on a satellite instrument launched in 2017, was not introduced until 2021 (Chen and Menelaou, 2021; Moran et al., 2026~~5~~). Given that BB emissions also have large year-to-year variations (e.g., Table A4), their neglect may complicate identification of systematic model errors, especially for the summer months. The impact of this omission for 2021/22 is examined ~~for 2021/22~~ in Sect. 4.2.

AirNow data have been used for the performance evaluation of the 2021/22 forecasts since not all finalized network measurement data sets were available for that period during the preparation of this paper. The use of AirNow data does reflect common practice for AQ forecast performance evaluations in the near term and is also consistent with evaluation results for previous RAQDPS operational versions for the 2010–2019 period, which also employed AirNow data (Sect. 4.1). On the other hand, the use of AirNow data limits the number of chemical species that can be considered, and 2021/22 analyses were only performed for $NO_2$, $O_3$, and $PM_{2.5}$ total mass. For the four years of retrospective annual runs, however, a much broader set of finalized AQ measurement data, including $PM_{2.5}$ speciation measurements and precipitation-chemistry measurements, was available and was used to carry out as broad and comprehensive an evaluation of model performance as possible. The performance evaluation results reported here include~~s~~ analyses stratified by different measurement characteristics to identify which network, species, month, hour of day, region, and land-use type resulted in the most skillful and the least skillful model predictions. Both Canadian and U.S. AQ measurement data sets were considered for all five years in order to expand the spatial coverage of the evaluation. This differs from many past evaluations of AQ model performance over North America that have only considered U.S. AQ measurements (e.g., Tessum et al., 2015; Yahya et al., 2015; Appel et al., 2017; Toro et al., 2021), although there are exceptions (e.g., Appel et al., 2021). One complicating factor for this study was that emissions of some anthropogenic pollutants decreased materially between 2013 and 2021 (see Sect. 2.2), but this factor was also ~~positive~~ helpful in that it allowed examination of the representativeness of the input model emissions that were used and constituted a dynamic evaluation of opportunity (cf. ~~e.g.,~~ Gilliland et al., 2008; Godowitch et al., 2010; Foley et al., 2015). The consideration of a total of 15 simulation years facilitated the identification of systematic model biases and

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errors by revealing common patterns across years and reducing the importance of year-to-year variations in emissions and in meteorology, including ~~its~~ the latter's impact on biogenic emissions.

The rest of this paper is organized as follows. Section 2 describes the study methodology, including the model configuration, run setup, and input emissions used to perform the 2013–2016 retrospective annual runs, the AQ measurement data sets used for the evaluation, the data processing and data filtering applied for model-measurement pairing, and the techniques and evaluation metrics used for the performance evaluation. Section 3 and the Supplement (S) present results of the RAQDPS-OP023 performance evaluation for 2021/22 and 2013–2016, where Sect. 3 focuses on aggregate annual analyses for air– and precipitation–chemistry measurements and the Supplement presents more detailed analys~~i~~es results stratified by network, season or month, hour of day, region, or land-use. Section 4 then compares RAQDPS-OP023 performance relative to 2010–2019 RAQDPS-OP forecast performance and to 2021/22 RAQDPS-FW023 performance, summarizes RAQDPS-OP023 and RAQDPS-FW023 performance vs. four peer AQ forecast systems, and discusses RAQDPS-OP023 shortcomings revealed by the evaluations. Lastly, Sect. 5 presents a summary and conclusions.

## 2 Methodology

### 2.1 Modelling system configurations and setups

The model configuration and run setup of the RAQDPS-OP023 and RAQDPS-FW023 for the 2021/22 forecasts has been summarized in CMC-RAQDPS-023 (2021) and described in detail in the companion paper by Moran et al. (2026~~5~~), so o~~. O~~nly a short overview will be given here. Some key aspects include the use of the following: (1) version 5.1.0 of the ECCC Global Environmental Multiscale (GEM) numerical weather prediction (NWP) model code and version 3.1.0.0 of the Modelling Air quality and Chemistry (MACH) chemical weather module code, which is embedded within the GEM code; (2) a ~~rotated~~ limited-area rotated latitude-longitude grid covering North America and adjacent oceans (e.g., Fig. 13) with 10-km horizontal grid discretization and 84 staggered vertical hybrid levels capped by a model lid at 0.1 hPa; (3) a two-time-level iterative-implicit time integration scheme and three-dimensional semi-Lagrangian advection scheme used with a 300 s meteorological time step and a 900 s chemistry time step; (4) imposed tracer mass conservation with an iterative, locally mass-conserving monotonicity correction and a Bermejo-Conde (2002) global mass fixer; (5) a simplified two-bin sectional representation of the $PM_{10}$ size distribution (diameter ranges of 0-2.5 µm and 2.5-10 µm); (6) PM dry chemical composition represented by eight chemical compounds [sulfate ($SO_4$), nitrate ($NO_3$), ammonium ($NH_4$), elemental carbon (EC), primary organic matter (POM), secondary organic matter (SOM), crustal material (CM), and sea salt (SS)]; (7) 42~~1~~ prognostic gas-phase chemical compounds and 16 prognostic particle-phase section-compounds (i.e., two~~2~~ size bins x eight~~8~~ compounds); (8) ADOM-2 gas-phase chemistry mechanism, ADOM aqueous-phase chemistry mechanism, HETV inorganic heterogeneous chemistry mechanism, and Instantaneous secondary organic Aerosol Yield (IAY) scheme; (9) parameterizations of aerosol particle nucleation, condensation/evaporation, coagulation, dry deposition and sedimentation, hygroscopic growth, and

activation; and (10) parameterizations of gas-phase dry deposition and in-cloud and below-cloud scavenging of particles and soluble gases.

The configuration and setup used for the 2013–2016 retrospective annual runs followed those of the RAQDPS-OP023 ~~operational~~ 2021/22 forecasts as closely as possible, but some differences could not be avoided as the retrospective simulations were performed later and outside of the operational environment. One major (and deliberate) difference was the replacement of the RAQDPS-OP023 ~~operational~~ projected anthropogenic input emissions files (see Moran et al., 2026) with year-specific anthropogenic input emissions files based on historical year-specific emissions inventories for 2013–2016 (see next section). A minor (but unavoidable) difference was the need to use an equivalent modelling system (RAQDPS-OP024) for the 2013–2016 hindcasts due to the migration with minimum changes of all ECCC operational and research computing applications to a new supercomputer in late June 2022. In addition, there were a few more minor differences related to near-surface vertical diffusion, model initialization and spin-up, meteorological piloting, and simulation run strategy, whose impacts were small.

First, the RAQDPS-OP023 ~~operational~~ runs for 2021/22 employed two code adjustments related to near-surface vertical diffusion to avoid the possibility of predicting extremely high surface concentrations due to the combina~~njunc~~tion of high surface emissions, an extremely stable PBL, and very low wind conditions such as might occur during northern winter nights under a strong anticyclone. As described in the companion paper by Moran et al. (2026~~5~~), one pre-emptive adjustment was to impose a minimum PBL height of 100 m when calculating the vertical diffusion of chemical tracers (where free-atmosphere convection applies above the PBL top); the other was to inject surface emissions into the lowest two model layers instead of the lowest layer (61 m thickness vs. 20 m thickness). For the four years of retrospective runs, however, these two adjustments were removed so that whatever PBL height ~~was~~ forecast by GEM was used in defining the vertical diffusivity profile and surface emissions were injected into the lowest (20 m thick) model layer. The reason for doing so was to test over a four-year period whether the two operational adjustments were needed.

A different approach was also used to maximize the dynamical balance between the mass and momentum fields in the meteorological initialization step. The operational forecast runs for 2021/22 employed an hourly incremental analysis update (IAU) approach from T-3 to T+3 hours, where T=0 is the run start time (e.g., Bloom et al., 1996). For the retrospective runs, on the other hand, a digital filter was employed at T=0 (Fillion et al., 1995). This difference was necessary because archived analyses for T-3 hours were not available for the 2013–2016 period.

The hourly meteorological lateral boundary conditions (LBCs) supplied by a meteorological "piloting" model for the retrospective runs also had a different source. For the 2021/22 operational forecasts these were supplied by version 8.0.0 of the operational 10-km Regional Deterministic Prediction System (RDPS), a limited-area-model configuration of GEM v5.1.0 that was run by ECCC to make meteorological forecasts for North America in advance of the RAQDPS-OP023 and RAQDPS-FW023 runs (Moran et al., 2026~~5~~). The RDPS 8.0.0 horizontal grid was a superset of the RAQDPS-OP023 horizontal grid and its vertical levels were identical with those of the RAQDPS-OP023 (CMC-

RDPS-8.0.0, 2021). For the retrospective runs, on the other hand, the meteorological LBCs were supplied from special hindcast runs of a 15-km global configuration of GEM 5.1.0. This change avoided the need to run both global and regional versions of GEM for the 2013-2016 period, and previous tests had shown that the use of a meteorological piloting model with 10 km vs. 15 km grid spacing had very little impact on RAQDPS forecasts.

Lastly, simulation run length was the source of one more difference. The twice-daily RAQDPS-OP023 and RAQDPS-FW023 operational forecast runs were 72 hours in length and were initialized at T-3 hours using the T+9 hour forecast fields from the previous RDPS 8.0.0 run launched 12 hours earlier. To save computer time the twice-daily RAQDPS-OP023 retrospective runs were only 18 hours in length and were initialized at T=0 hours using the T+12 hour forecast meteorological fields from the previous global GEM run and T+12 hour forecast chemical fields from the previous RAQDPS-OP023 run. In both cases, though, annual sequences of hourly predicted meteorological and chemical fields were prepared by concatenating hourly predictions for only the first 12 forecast hours of each RAQDPS-OP023 run (i.e., T+1 to T+12 hours).

## 2.2 Input emissions

The RAQDPS-OP023 and RAQDPS-FW023 2021/22 forecasts used the SET4.0.0 anthropogenic emissions data set described by Moran et al. (2021b, 2026~~5~~). The SET4.0.0 emissions were based on a projected 2020 Canadian national annual anthropogenic emissions inventory and projected 2023 U.S. and Mexican national annual anthropogenic emissions inventories, which were roughly in temporal alignment with the 2021/22 forecast period. Note, though, that it was not possible to modify the SET4.0.0 emissions in near-real time to account for rapidly evolving emissions changes in North America associated with large-scale public-health responses to the global COVID-19 pandemic (cf. Mashayekhi et al., 2021).

To provide year-specific anthropogenic input emissions files for each of the 2013–2016 retrospective annual runs, however, a concerted effort was made to use recently available and consistent national annual anthropogenic emissions trend data sets for Canada, the U.S., and Mexico. Each of these national emissions trend data sets provides a multi-decadal sequence of annual anthropogenic national emission inventories that were generated using largely consistent emissions estimation methodologies for all of the years considered by each data set. Four year-specific annual anthropogenic national emission inventories were extracted for the 2013–2016 period for Canada from the ECCC TREND16 emissions trend data set. Similarly, four year-specific annual anthropogenic national emission inventories were extracted for the 2013–2016 period for the U.S. from the U.S. EPA EQUATES (EPA~~pa~~'s A~~a~~ir QUAlity TimE Series) national emissions trend data set. The EQUATES data set also includes a set of annual anthropogenic national emissions inventories ~~and annual wildfire emissions inventory files~~ for 2002-2019 for Mexico, so that year-specific Mexican inventories for the 2013–2016 period were also available. More details about these emissions data sets are provided in Sect. S2.2 ~~of the Supplement~~.

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Model-ready emissions files for the years 2013–2016 were then prepared from these national inventories using version 4.8 of the Sparse Matrix Operator Kernel (SMOKE) emissions processing system (https://www.cmascenter.org/smoke/,~~;~~ last access: 6 April 2026; Zhang et al., 2018b). The use of an emissions processing system was necessary because while the national inventories of the three countries report annual emissions of seven criteria air pollutants by jurisdiction (province, county, or state), the RAQDPS-OP023 requires gridded emissions fields for each emitted model species for every hour of every day of the year (e.g., Dickson and Oliver, 1991; Houyoux et al., 2000; Matthias et al., 2018; Zhang et al., 2018b). D~~Additional d~~etails about the processing of these anthropogenic emission inventories with SMOKE are provided in Sect. S2.2.

Several types of natural emissions were also accounted for. Time-varying biogenic emissions were included in all five annual simulations, where biogenic emissions were calculated for each time step of the RAQDPS-OP023 simulations using code for a modified version of the Biogenic Emission Inventory System (BEIS) v3.09 biogenic emissions algorithms. ~~with i~~Inputs of two GEM-predicted meteorological fields were required: surface temperature and solar insolation (Moran et al., 2026~~5~~). Time-varying sea-salt emissions were also included for all five annual simulations; these emissions were calculated for each chemistry time step based on surface wind speed. Hourly BB emissions, on the other hand, were only considered in the 2021/22 RAQDPS-FW023 forecast runs (Sect. 4.2). Note that some BB emissions might be due to very large prescribed burns or grass fires as well as wildfires if these were detected by satellite (Moran et al., 2026~~5~~). Note also that neither system version considered some other types of natural emissions, namely natural wind-blown fugitive dust emissions, lightning emissions, pollen and other biological emissions, other marine emissions, and volcanic emissions, but these sources were assumed to be less important for Canada (see Moran et al., 2026).

Table 1 presents a summary of annual inventory emissions of the seven criteria pollutants for Canada, the U.S., and Mexico for the five years for which annual RAQDPS-OP023 runs were performed. The rows named "Total Anthro" and "Total Biogenic", on the other hand, are ~~, respectively,~~ the annual, SMOKE-processed anthropogenic emissions and the annual, ~~dynamically~~BEIS-calculated biogenic emissions within the model domain that the model "sees" (i.e., responds to). These domain-total "Total Anthro" and "Total Biogenic" values thus include all Canadian emissions but only U.S. emissions from the 48 contiguous U.S. states and part of Alaska and exclude emissions from the rest of Alaska, Hawaii, and some U.S. territories in the Caribbean and the Pacific Ocean and only include Mexican emissions from the 340 Mexican counties out of 2,457 that lie completely or partially within the RAQDPS-OP023 domain (e.g., Fig. 13). This spatial exclusion of emissions is the reason that the "Total Anthro" values are consistently lower than the "3 Country EIs" values for all inventory species. In addition, the domain-total SMOKE-processed values for "Total Anthro" $PM_{2.5}$ and $PM_{10}$ emissions are also considerably lower than total inventory values due to the impact of adjustments for land-use-dependent, near-source removal due to settling and impaction (i.e., transportable fraction TF), but further meteorology-dependent emission reductions due to snow cover and wet soil were only applied during the RAQDPS-OP023 simulations (Moran et al., 2026).

Some significant changes are evident in annual emissions over this nine-year time period in Table 1, first over the four-year period from 2013 to 2016, then from 2016 to 2021/22, and in total from 2013 to 2021/22. For example, North American "Total Anthro" $SO_2$ emissions decreased by 37% from 2013 to 2016 and then by a further 37% from 2016 to 2021/22, for a total decrease of 60% relative to 2013, while "Total Anthro" $NO_x$ emissions decreased by 19%, 22%, and 36% for the same three periods. "Total Anthro" VOC and CO emissions also decreased over the three periods, by 8%, 1%, and 9% for VOC emissions and by 13%, 14%, and 25% for CO emissions. "Total Anthro" $NH_3$ emissions, on the other hand, were nearly constant during the 2013–2016 period but then increased in Canada while decreasing in the U.S. and Mexico in 2021/22 for an overall domain-total decrease of 6% from 2013 to 2021/22. ~~Lastly, the domain-total SMOKE-processed values for "Total Anthro" $PM_{2.5}$ and $PM_{10}$ emissions are considerably lower than total inventory values due to the impact of adjustments for land use-dependent, near-source removal due to settling and impaction (i.e., transportable fraction), but further meteorology-dependent emission reductions due to snow cover and wet soil are only applied during the RAQDPS023 simulation (Moran et al., 2025).~~

Table 1 also compares SMOKE-processed, domain-total annual anthropogenic emissions with calculated domain-total annual biogenic emissions of $NO_x$ and VOCs. Biogenic NO emissions can be seen to have contributed 4% of domain-total $NO_x$ emissions in 2013, rising to 6% in 2021/22 as anthropogenic $NO_x$ emissions decreased and biogenic NO emissions increased. By contrast biogenic VOC emissions contributed 78% of domain-total VOC emissions in 2013, a ~~considerably~~ much larger percentage, and then rose to 81% in 2021/22 as anthropogenic VOC emissions declined and biogenic VOC emissions increased. In addition, Table S1 compares the seasonal variation of the SMOKE-processed SET4.0.0 anthropogenic emissions for 32 model species and the biogenic emissions of four model species. Seasonal variations depend strongly on both the pollutant and its source types and can have markedly different cycles. For example, $NH_3$ is primarily emitted by agricultural activities and can be seen to have a pronounced winter minimum and summer maximum, whereas seasonal emissions of two lumped model VOC species, ALD2 (acetaldehyde and higher aldehydes) and CRES (cresols and phenols), have a pronounced winter maximum and summer minimum consistent with their dominant source being residential wood combustion. For most anthropogenic species, however, seasonal variations were considerably smaller, including $SO_2$, NO, and $NO_2$. Although fossil-fuel power generation is important for both $SO_2$ and $NO_x$ emissions, this suggests that at the continental scale the increased power load for space heating in the winter in North America is roughly balanced by the increased power load for air conditioning in the summer. The seasonal variation of biogenic emitted species~~sions~~, on the other hand, was more like $NH_3$; they had strong seasonal cycles with winter minima and summer maxima. In fact, biogenic isoprene ($C_5H_8$) emissions were predicted to have the most pronounced domain-level seasonal cycle, increasing from just 2% in the winter to 65% in the summer (Table S1).

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### 2.3 Air-chemistry and precipitation-chemistry observations

Routine air-chemistry and precipitation-chemistry measurements are available from multiple measurement networks operating in Canada and the U.S. The hourly measurements of abundances of $NO_2$, $O_3$, and $PM_{2.5}$ total mass made by continuous instruments that are reported in near-real time by some agencies to the U.S. EPA's AirNow program (Dye

et al., 2004; Wayland et al., 2004; Zhang et al., 2012a) have already been mentioned. AirNow hourly measurement data for $NO_2$, $O_3$, and $PM_{2.5}$ from U.S. monitors have been combined in this study with NRT $NO_2$, $O_3$, and $PM_{2.5}$ hourly measurement data from Canadian National Air Pollution Surveillance (NAPS) monitors that report directly to ECCC's Canadian Meteorological Centre (CMC). This combined NRT data set has been used to evaluate the 2021/22 RAQDPS-OP023 and RAQDPS-FW023 operational forecasts of these species.

The use of these NRT abundance measurements, which are considered to be preliminary, for model evaluation is consistent with the operational nature of the RAQDPS-OP023 forecasts. Two automated data filters were applied by CMC to the AirNow and NAPS NRT measurements upon receipt before they were used for model evaluation or other purposes such as operational air pollutant objective analyses (e.g., Robichaud et al., 2016). The first filter flagged negative abundance values and above-threshold abundance values as suspicious ($NO_2$ levels over 200 ppbv, $O_3$ levels over 200 ppbv, $PM_{2.5}$ levels over 300 $\mu g \cdot m^{-3}$) or invalid ($NO_2$ levels over 2000 ppbv, $O_3$ levels over 500 ppbv, $PM_{2.5}$ levels over 1000 $\mu g \cdot m^{-3}$), while the second filter flagged large jumps in abundances between consecutive hours as suspicious (over 30 ppbv for $NO_2$, over 60 ppbv for $O_3$, over 90 $\mu g \cdot m^{-3}$ for $PM_{2.5}$). Values flagged as suspicious or invalid were not used for this study. Measurements from stations located near roadways were discarded as well to ensure spatial representativeness. A temporal completeness criterion was also imposed on this NRT measurement data set to ensure temporal representativeness of the evaluation data: individual station data sets were required to have at least 75% valid values out of the total possible values for a one-year evaluation period to be considered complete. More details about these representativeness constraints are provided in Sect. S2.4. Table 2 lists the number of available AirNow and NAPS stations for 2021/22 that measured hourly $O_3$, $NO_2$, and $PM_{2.5}$ abundances and the number with complete data records while Figure S2 shows the locations of ~~these~~ all stations measuring each of these three species. Note that some stations only measured one or two of these species.

The evaluation of the retrospective annual simulations for 2013–2016, on the other hand, was based on finalized AQ network data sets. The CAPMoN and NAPS networks in Canada and the AMoN, AQS, CASTNET, CSN, IMPROVE, NATTS, and PAMS networks in the U.S. provide finalized air-chemistry measurement data sets for various chemical species, including hourly $NO_2$, $O_3$, and other gas-phase species (e.g., $SO_2$, CO, $HNO_3$, ~~and~~ $NH_3$, ethene ($C_2H_4$), formaldehyde (HCHO), and $C_6H_6$), hourly $PM_{2.5}$ and $PM_{10}$ mass, daily FRM (Federal Reference Method) and FEM (Federal Equivalent Method) $PM_{2.5}$ mass (e.g., Demerjian, 2000; Noble et al., 2001; Gantt, 2022), and daily $PM_{2.5}$ chemical composition, while the CAPMoN network in Canada and the NADP network in the U.S. provide finalized precipitation-chemistry measurement data sets. Details about each of these networks are given in Table S2a, and Sect. S2.3 provides some additional information about daily $PM_{2.5}$ measurements. Note that all network data sets used in this study were accessed on 24 June 2024 (relevant because these data sets are always subject to change even years after their original release). Table 2 and Tables S2b-d list the number of stations for each network for 2013–2016 with available measurements and with complete measurements of various chemical species (see~~cf.~~ Sect. S2.4), while Figs. S3–S6 show the locations of available stations by network. It is clear from these four figures that spatial coverage varies widely by the species being measured.

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Although AQ measurements provide important chemical information about the real atmosphere, these measurements, like AQ model predictions, also have biases and errors. For example, $NO_2$ measurements made with chemiluminescence monitors frequently have positive biases due to interference from other oxidized nitrogen species (e.g., Dunlea et al., 2007; Lamsal et al., 2015; Dickerson et al., 2019), and $NH_3$ measurements made with passive monitors have negative biases (Puchalski et al., 2011). Measurements of both $PM_{2.5}$ total mass and semi-volatile $PM_{2.5}$ chemical components are known to have both positive and negative artifacts (e.g., Chow, 1995; Frank, 2006; Watson et al., 2009; Dabek-Zlotorzynska et al., 2011; Malm et al., 2011; Su et al., 2018; Gantt, 2022). Estimates of reconstructed $PM_{2.5}$ total mass based on the sum of $PM_{2.5}$ ~~species~~ chemical component mass measurements from the CSN, IMPROVE, and NAPS networks often differ from direct measurements of $PM_{2.5}$ total mass (e.g., Malm et al., 2011; Chow et al., 2015; Hand et al., 2019). As well, different networks measuring the same species often do not use the same instruments or follow the same field and laboratory protocols, which can affect comparability across networks. Examples include the use of different instruments to measure $SO_2$, CO, and $O_3$ concentrations by different agencies reporting to AQS (e.g., Demerjian, 2000; Parrish and Fehsenfeld, 2000), $SO_2$ and $HNO_3$ concentrations measured by the CAPMoN and CASTNET networks vs. the NAPS network (Dabek-Zlotorzynska et al., 2011; Feng et al., 2020), $NH_3$ concentrations measured by the AMoN and NAPS networks (Dabek-Zlotorzynska et al., 2011; Puchalski et al., 2011), sulphur and nitrogen species measured by the CASTNET and IMPROVE networks (e.g., Ames and Malm, 2001; Lavery et al., 2009), and $PM_{2.5}$ chemical components measured by the IMPROVE and CSN networks (Hand et al., 2012; Solomon et al. 2014). In order to assess measurement comparability between networks some efforts have been made to co-locate instruments used by different networks at one or more locations, including CSN and IMPROVE (Malm et al., 2011; Hand et al., 2012), CASTNET and CAPMoN (Schwede et al., 2011), and CAPMoN and NADP (Sirois et al., 2000; Wetherbee et al., 2010; Feng et al., 2023).

### 2.4 Pairing measurements with model predictions

The comparison of AQ measurements and AQ model predictions must be done with care since measurements and model predictions never represent exactly the same quantities (e.g., Seigneur and Moran, 2004; Appel et al., 2008). From a temporal perspective, reported AQ measurements may either be instantaneous values or time averages, and the time averages may in turn represent mean values for an averaging period that begins, ends, or straddles the reporting time. Model values, on the other hand, are nominally instantaneous but correspond to a time that is a discrete time step after the previous integration time and after a particular operator calculation step in a repeating sequence of numerical operators. From a spatial perspective, AQ measurements are typically made at a near-surface "point" location whereas AQ model predictions represent a volume-average value corresponding to the volume of an individual model grid cell. This spatial representativeness discrepancy is sometimes referred to as incommensurability. It is a fundamental source of model uncertainty that can be reduced by reducing model grid spacing but never removed entirely (e.g., Nappo et al., 1982; Venkatram, 1988; McNair et al., 1996; Spicer et al., 1996; Swall and Foley, 2009; Stroud et al., 2011; Schutgens et al., 2016). Estimates of population exposure to air pollutants based on ambient measurements from a small number of AQ measurement stations also suffer from the same problem (e.g., Jerrett et al., 2005; Hystad et al., 2011). And lastly, from a chemical perspective AQ measurements sometimes correspond to a combination of two or

more model variables while in other cases they correspond to more detailed chemical species than the AQ model is able to consider. For these reasons some pre-processing is generally required to pair or match AQ measurements and model predictions before they are compared while bearing in mind that some differences will still remain~~, and~~. It is thus important to document the methodology used to perform this pairing ~~should be documented~~ (Simon et al., 2012).

Temporal pairing was relatively straightforward for this study, although it was necessary to examine AQ network documentation carefully to understand the exact temporal nature of the measurements being reported. Model ~~concentration~~ abundance predictions were available every chemistry time step, or every 15 minutes in the case of the RAQDPS-OP023 (Moran et al., 2026~~5~~), so it was simple to pair model predictions with instantaneous hourly abundance ~~concentration~~ measurements. On the other hand, to pair model predictions with the mean hourly abundance~~concentration~~ measurements reported by the NAPS and AQS networks (Fig. S2), multiple consecutive sub-hourly model predictions needed to be combined using one of two approaches: an end-value approach, which averages the model values at the beginning and end of the measurement sampling period; or an integration approach, where the trapezoidal rule is used to combine the five RAQDPS-OP023 values available for the hour. The end-value approach was used in this study. To pair model predictions with mean daily air abundance ~~concentration~~ measurements from the AQS, CSN, IMPROVE, and NAPS networks (Table S2a), all hourly model values for each 24-hour sampling period were averaged. To pair with mean weekly air abundance ~~concentration~~ measurements from CASTNET, all hourly model values for each weekly sampling period were averaged; the same weekly averaging was also performed for daily mean CAPMoN and NAPS air concentrations for temporal consistency in order to be able to compare evaluation statistics between networks properly and to pool measurements for these three networks. Similarly, to pair model predictions with mean biweekly AMoN abundance measurements, hourly model $NH_3$ values were averaged for each biweekly sampling period, and the same was done for NAPS daily mean $NH_3$ values for temporal consistency. Finally, to pair model predictions with daily precipitation-chemistry measurements (CAPMoN) or weekly precipitation-chemistry measurements (NADP), hourly model wet deposition forecasts were accumulated for the appropriate period, but then weekly deposition values were also calculated for CAPMoN for temporal consistency with NADP. Weekly precipitation-weighted mean wet concentration values were then calculated from weekly precipitation and weekly deposition values.

More details about pairing related to this study, including spatial and chemical considerations, especially for PM measurements, and completeness screening, are provided in Sect. S2.4. These~~These details~~ include the nuances of evaluating VOC predictions, handling ambient vs. standard temperature and pressure, the measurement proxies used for some PM chemical components such as $NH_4$, total organic matter (TOM ~~(=~~POM+SOM), CM, and SS, the treatment of aerosol water, and the combined gas-particle phases implicit in precipitation-chemistry measurements. For example, one important consideration related to pairing with VOC measurements is that while measurements of ethene ($C_2H_4$) are available, the ADOM-2 model VOC species named ETHE is a lumped species that includes additional VOC species (Sect. S2.4). This misalignment must be considered when interpreting evaluation results for ETHE in Sect. 3.3.1 since the measurements are in effect lower bounds on the model predictions of this species rather than direct counterparts.

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## 2.5 Model performance metrics and evaluation

Once a set of paired observed and model-predicted values is available, model performance metrics can be calculated. For the present study we chose to calculate 12 statistical metrics: observation mean ($\overline{O}$); model prediction mean ($\overline{M}$); mean bias (MB); normalized mean bias (NMB); normalized mean absolute error (NMAE); root mean square error (RMSE); Pearson correlation coefficient (R); fraction of predictions within a factor of 2 of observations (FAC2); centered root mean square error (CRMSE); standard deviations of observations ($\sigma_O$ or SDO) and model predictions ($\sigma_M$ or SDM); and normalized standard deviation (NSD). Definitions of these metrics are provided in Table A2 and some background information about their selection is provided in Sect. S2.5. Inclusion of the first seven of these metrics is consistent with the recommendation of Simon et al. (2012) for a minimum set of performance evaluation statistics that should always be calculated to promote comparability across separate studies. The eighth metric, FAC2, is a dimensionless and bounded (0–1) measure of error or scatter that is not sensitive to outliers (e.g., Chang and Hanna, 2004; Borrego et al., 2008; Derwent et al., 2010; Savage et al., 2013). CRMSE, unlike RMSE, is insensitive to bias and represents the error due to differences in pattern variation or, alternately, the standard deviation of the error; it has been used in many studies (e.g., Bencala and Seinfeld, 1979; Stanski et al., 1989; Taylor, 2001; Chang and Hanna, 2004; Entekhabi et al., 2010; Sakaguchi et al., 2012; Thunis et al., 2012). NSD has been suggested as a metric by Taylor (2001), Chang and Hanna (2004), and Thunis et al. (2012), but by itself it does not provide information about the magnitudes of $\sigma_O$ or $\sigma_M$ (which were recommended by Willmott (1981) and reported by Appel et al. (2021)). Note that CRMSE, R, $\sigma_O$, and $\sigma_M$ (and sometimes NSD) are all linked by the Taylor diagram (Taylor, 2001). One other quantity that should be reported with these 12 metrics is N, the number of measurement-model pairs. Many evaluation studies fail to report this quantity, but it is used explicitly in the calculation of many metrics and it provides valuable information about sample size, representativeness, and significance (e.g., Huang et al., 2021).

Since this is an AQ forecasting evaluation, we have focused here on "native" network sampling duration: that is, hourly, daily, weekly, or biweekly samples, but based on the networks with the longest sampling duration for each species (e.g., biweekly for AMoN for $NH_3$, but weekly for CASTNET and NADP-NTN) to allow consistent comparisons between networks and pooling of network measurements. Other studies that focused on model performance for regulatory applications have looked at "constructed" predictands such as maximum daily 8-hr average (MDA8) or maximum 1-hourly values of $O_3$ volume mixing ratios (VMRs), whereas in this study we have only considered network-reported values such as hourly $O_3$ VMRs. And while we report annual domain-average statistics based on combined network measurements, we also report results for more stratified (i.e., disaggregated) analyses, including network-specific statistics, seasonal statistics, monthly statistics, diurnal statistics, regional statistics, and urban/rural statistics. To calculate urban vs. rural statistics, each measurement site was classified as urban or rural based on its grid-cell population density, where a threshold of 400 persons $km^{-2}$ was applied for Canada and 386 persons $km^{-2}$ (1000 persons per square mile) for the U.S. as the minimum urban population density. The slightly different thresholds were used for consistency with national censuses. We have also discussed our statistical results contextually in the next wo sections by referringence to the performance benchmarks proposed by Simon et al. (2012), Emery et al. (2017), Kelly

et al. (2019), Huang et al. (2021), and Zhai et al. (2024) (see Sect. S2.5). Some additional discussion related to model performance metrics ~~can also be found~~is provided in Sect. S2.5.

## 3 Results

This section presents performance evaluation ~~results~~statistics, first for one meteorological parameter important for air quality (Sect. 3.1), then for three key chemical species, $NO_2$, $O_3$, and $PM_{2.5}$ total mass (Sect. 3.2), and then for other gas-phase species, $PM_{2.5}$ chemical composition, and wet concentration and deposition of three inorganic species (Sect. 3.3). Annual, seasonal, monthly, diurnal, and regional evaluations for the one-year period from 1 June 2021 to 31 May 2022, the first year of RAQDPS-OP023 forecasts, are presented in this section along with selected evaluation results for the 2013–2016 RAQDPS-OP023 annual simulations. Many additional tables and figures related mainly to the 2013–2016 simulations, which serve to further quantify predictive skill and characterize the temporal and spatial variability of RAQDPS-OP023~~system~~ performance, can be found in Sect. S3 of the Supplement.

### 3.1 Operational evaluation of AQ-relevant meteorological predictands

Meteorological processes affect air quality through their influence on emissions, atmospheric transport and diffusion, chemistry, and wet and dry removal. Near-surface temperature, wind speed, and precipitation are three meteorological parameters important for surface air quality (e.g., Vautard et al., 2012; Gilliam et al., 2015; McNider and Pour-Biazar, 2020; Wang et al., 2021; Campbell et al., 2022). As described in the companion paper by Moran et al. (2026~~5~~), the RAQDPS-OP023 (and RAQDPS-FW023) is a regional chemical weather model configured to produce nearly identical meteorological forecasts to the RDPS 8.0.0, the ECCC regional weather forecast model that was operational at the same time (Fillion et al., 2010; Caron et al., 2015; McTaggart-Cowan et al., 2019; CMC-RDPS-8.0.0, 2021). Performance e~~E~~valuations of RDPS weather forecasts have been presented in these and other publications. For this paper we have only evaluated RAQDPS-OP023 precipitation forecasts based on precipitation measurements from ~~AQ~~ precipitation-chemistry networks, which are not usually available to or considered in NWP model performance evaluations. This choice also ensures that performance statistics for precipitation are consistent in time and space with those for pollutant concentrations in precipitation and wet deposition (Sect. 3.3.3).

Domain-average annual scores for RAQDPS-OP023 weekly precipitation forecasts at precipitation-chemistry stations for 2013 to 2016 are listed in Table 6. As noted in Sect. 2.4 w~~W~~e have chosen to consider weekly forecasts because the U.S. NADP precipitation-chemistry network only reports weekly accumulated measurements. Interestingly, this set of scores does not show much variation from year to year. ~~:~~ F~~f~~or example, annual MB values ranged from -0.1 to 2.2 mm·week$^{-1}$, NMB values from -0.01 to 0.12~~1~~, NMAE values from 0.49 to 0.54, RMSE values from 17.6 to 20.6 mm·week$^{-1}$, FAC2 values from 0.56 to 0.57, and R values from 0.71 to 0.78. Appel et al. (2011) reported a comparable NMB range but a markedly lower RMSE range for 12-km MM5 meteorological model simulations for the 2002-2006 period, but those earlier statistics were calculated for accumulated seasonal and annual precipitation predictions as opposed to weekly predictions, which have greater temporal variability.

Seasonal analyses of RAQDPS-OP023 predictions of near-surface temperature, wind speed, and precipitation can be found in Sect. S3.1 as well as individual-network annual and seasonal evaluation ~~results~~ scores and subregional annual evaluation score~~result~~s for weekly precipitation. These additional evaluations showed that model skill in predicting weekly precipitation was highest for the winter season and lowest for the summer season (e.g., Fig. S161). The probable explanation is that organized, synoptic-scale precipitation, which is more predictable, is likeliest to occur in the winter whereas small-scale convective precipitation, which is harder for NWP models to predict, is likeliest to occur in the summer (e.g., Appel et al., 2011; Gilliam et al., 2021). Many of the largest overpredictions of weekly precipitation at individual stations occurred in the Rocky Mountain region of the western U.S., where ~~subgrid-scale (SGS) topographical features and~~ station location ~~are~~ relative to subgrid-scale (SGS) topographical features is likely to be important, whereas underpredictions at stations were common in the flatter terrain of the southeastern and central U.S. (Fig. S121). In addition, the evaluation statistics for an individual network or month or subregion were sometimes qualitatively different from those for the combined networks, combined months, or combined subregions. For example, model skill in predicting annual and seasonal mean weekly precipitation was found to be higher for the Canadian CAPMoN precipitation-chemistry network than the U.S. NADP network (Tables S6A, S6S). This possibility always needs to be kept in mind when interpreting the most highly aggregated performance statistics (e.g., annual statistics and all-network, all-station statistics), which may obscure systematic network differences or be impacted by compensating errors (e.g., Makar et al., 2014).

## 3.2 Operational evaluation of three key air quality predictands

This section presents operational evaluation results for RAQDPS-OP023 predictions of hourly surface $NO_2$, $O_3$, and $PM_{2.5}$ total mass abundances for five years. Results are presented first for 2021/22 $NO_2$ and $O_3$ forecasts ~~based on~~that were compared to NRT measurements, followed by results for 2013–2016 $NO_2$ and $O_3$ hindcasts ~~based on~~that were compared to QA/QCed network measurements, 2021/22 $PM_{2.5}$ forecasts compared to ~~based on~~ NRT measurements, and 2013–2016 $PM_{2.5}$ hindcasts compared to ~~based on~~ QA/QCed network continuous and gravimetric measurements.

### 3.2.1 $NO_2$ and ozone

2021/22 operational forecasts of $NO_2$ and $O_3$

Figure 1 shows the spatial distribution over North America of annual mean $NO_2$ and $O_3$ hourly surface VMR fields predicted by the RAQDPS-OP023 for the 2021/22 period. Coloured "coffee beans" (i.e., divided dots) are superimposed on the contoured fields to show observed and predicted annual mean values at NRT measurement stations for the same period (see Fig. S2 for station locations). Generally good agreement is evident in Fig. 1 between the observed and predicted annual mean values of both pollutants (although viewing the $NO_2$ panel with higher magnification is helpful)~~;~~ ~~although~~ note that the higher $NO_2$ VMRs associated with urban centers are smaller-scale features caused by high $NO_x$ emissions over urban centers (cf. Fig. S1a)~~.~~.

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More quantitatively, Table 3 lists values of all-station annual model evaluation statistics for hourly $NO_2$ and $O_3$ VMR forecasts for 2021/22. These overall performance scores are generally less good for $NO_2$ than for $O_3$, including NMB

(-0.19 vs. -0.07), NMAE (0.56 vs. 0.28), FAC2 (0.52 vs. 0.83), and R (0.65 vs. 0.72). This difference is not surprising given that $NO_2$ is a primary pollutant whose spatial distribution is dominated by the distribution of emissions with strong spatial gradients whereas $O_3$ is a secondary pollutant with a smoother spatial pattern and smaller dynamic range. One indicator of the degree of smoothness of a pattern is the coefficient of variation (CV) or relative standard deviation (ratio of standard deviation to arithmetic mean; see Table A2), where a lower value indicates a ~~greater~~ smoother field~~ness~~ (i.e., lower variability) (e.g., Fruin et al., 2014; Lee et al., 2018). The observed and predicted annual CV values for 2021/22 calculated from Table 3 were 1.16 and 1.20, respectively, for $NO_2$ vs. 0.50 and 0.47 for $O_3$.

In order to judge the level of model skill suggested by the Table 3 scores, Zhai et al. (2024) have recommended benchmark goals for NMB, NMAE, and R of ±0.20, 0.40, and 0.60 for "good" $NO_2$ performance scores (i.e., scores above 67th percentile relative to the scores for a historical multi-model ensemble) while Emery et al. (2017) have recommended benchmark goals for NMB, NMAE, and R of ±0.05, 0.15, and 0.75 as good $O_3$ performance scores. These benchmark goals were met by RAQDPS-OP023 forecasts for 2021/22 for $NO_2$ except ~~for $NO_2$~~ annual NMAE scores ~~and~~ but not for $O_3$ annual ~~R~~ scores.

When interpreting these benchmark comparisons, however, it should be noted that Emery et al. (2017) also recommended that the evaluation period considered should be no more than one month for $O_3$ and a 40 ppbv cutoff should be used when calculating NMB and NMAE for $O_3$, whereas no cutoff was considered for Table 3. ~~I~~In addition, ~~t should also be noted that~~ Simon et al. (2012) found performance scores for retrospective model applications to be better on average than those for forecast applications due to the use by the former of year-specific emissions, meteorological reanalyses, day-specific chemical lateral boundary conditions, and other retrospective data sets that are not available to AQ forecasting applications. It was also noted in Sect. 2.3 that most network $NO_2$ monitors in North America suffer from positive biases due to interference from other oxidized nitrogen species (e.g., Dunlea et al., 2007; Dickerson et al., 2019). These biases, however, vary greatly with time and location. They are expected to be smallest when fresh $NO_2$ emissions dominate, for example, in urban areas, in the early morning hours (4-9 a.m.) when the PBL is shallow, and in the winter, but largest relatively speaking for aged air, as in rural areas, in the afternoon hours, and in the summer (e.g., Godowitch et al., 2010; Lamsal et al., 2015; Jaeglé et al., 2018; He et al., 2019; Toro et al., 2021). Thus, the annual NMB value for $NO_2$ of -0.19 would be smaller if these measurement biases were accounted for.

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Figures 2 and 3 provide an extreme disaggregation of the all-station annual scores from Table 3 by showing spatial distributions of station-specific annual values of four statistics (MB, NMB, CRMSE, and R) for hourly $NO_2$ and $O_3$ measurements, respectively, for 2021/22. Such plots can reveal regional patterns in the evaluation statistics. For example, annual NMB values for $NO_2$ tend to be negative everywhere (again consistent with a positive measurement bias) but they are more negative (i.e., worse) in general at western stations while CRMSE and R values for $NO_2$ are lower in the continental interior than ~~the~~ in coastal areas (Fig. 2). Annual MB and NMB values for $O_3$, on the other hand, are generally negative at western stations but positive at eastern stations, particularly at coastal stations (Fig. 3), while both annual CRMSE and R scores are higher overall across the continent for $O_3$ than for $NO_2$.

Figure 5 adds temporal detail to the annual analysis shown in Fig. 1. It shows the corresponding predicted spatial distributions of seasonal mean $NO_2$ and $O_3$ hourly surface VMR fields for 2021/22, again with superimposed coloured divided dots to show observed and predicted seasonal mean values at NRT measurement stations for each season. By inspection, predicted domain-scale, seasonal mean $NO_2$ levels appear to be highest in the winter season (DJF) and lowest in the summer season (JJA), whereas $O_3$ levels are predicted to be highest in the spring season (MAM) and lowest in the autumn (SON) season.

Another perspective on the temporal variation of ~~model~~ RAQDPS-OP023 performance is provided by Figs. 6 and 7, which show time series of observed and predicted all-station monthly mean VMR values of hourly $NO_2$ and $O_3$, respectively, for 2021/22 for all NRT measurement stations in the model domain ~~as well as~~. T~~t~~ime series of monthly NMB, CRMSE, and R scores are also shown in these two figures. Observed and predicted monthly mean $NO_2$ VMRs are both highest in January and lowest in June and July; observed and predicted monthly mean $O_3$ VMRs are both highest in April and lowest in November. Monthly NMB values for hourly $NO_2$ are negative for all months but are slightly worse for the winter months even though the $NO_2$ measurement bias is expected to be smaller in the winter, while~~,~~ monthly CRMSE values for $NO_2$ peak in January, and monthly R values for $NO_2$ do not vary much from month to month but are slightly higher in the winter. Monthly NMB values for hourly $O_3$ are negative for all months except December and January, monthly CRMSE values for $O_3$ peak in July, and monthly R values for $O_3$ also do not vary much but are slightly higher in the summer. It should be noted that seasonal variations in $NO_x$ emissions are very small at the domain scale (Table S1), suggesting that the observed and predicted variations in monthly mean $NO_2$ levels evident in Fig. 6 are controlled by ~~other~~ factors other than emissions, such as monthly variations in temperature, photolysis, PBL height, vegetation phenology, and dry deposition. For $O_3$, on the other hand, biogenic emissions of VOCs, its other main precursor, have very large seasonal variations (Table S1). Interestingly, although the largest predicted monthly $O_3$ values occur in April, the model still underpredicts the well-known springtime $O_3$ maximum in the Northern Hemisphere (e.g., Penkett and Brice, 1986; Monks, 2000; Liudchik et al., 2015) by about 5 ppbv in April.

It is also of interest to examine ~~model~~ RAQDPS-OP023 performance by time of day since many emission source sectors and meteorological and chemical processes vary diurnally. Figures 9 and 10 show all-station, annual-mean diurnal time series in local time (LT) of values of five statistics for hourly $NO_2$ and $O_3$ surface VMRs, respectively, for the 2021/22 period. The annual-mean diurnal time series of observed and predicted hourly VMRs for both species display a strong dependence on time of day as do the diurnal time series of annual NMB, CRMSE, and R scores. Model predictions of annual-mean hourly values of both $NO_2$ and $O_3$ surface VMRs agree well overall with observations, including the times of the observed daily maxima and minima. The annual-mean diurnal time series of hourly $NO_2$ VMR and the associated hourly evaluation statistics in Fig. 9 display extrema close to the times of morning and afternoon rush hours, suggesting that diurnal variation of on-road $NO_x$ emissions plays an important role in driving the diurnal pattern. Interestingly, hourly NMB values for $NO_2$ are most negative in the afternoon when~~,~~ the $NO_2$ measurement bias is expected to be largest. ~~while~~ By contrast, the maximum annual-mean hourly $O_3$ VMR occurs at

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14 LT and the minimum at 05 LT (Fig. 10). For $O_3$ the smallest annual-mean hourly NMB and CRMSE values and highest annual-mean hourly R values occur close to the mid-day $O_3$ peak.

Figures 2–4 showed how model performance can vary geographically. A complementary result is presented in Fig. 12, which compares regional time series of observed and RAQDPS-OP023 predicted monthly means of hourly $NO_2$ and $O_3$ VMRs for 2021/22 for the four continental quadrants shown in Fig. S7. Both observed and predicted time series exhibit a regional dependence. For $NO_2$ the agreement between observed and predicted monthly means was closest for western Canada while for $O_3$ it was closest for the eastern U.S. ~~Interestingly, p~~Peak observed monthly mean $NO_2$ VMR values were slightly higher in the west than in the east, at least for these regional sets of stations (see Fig. S2). Observed and predicted monthly mean $NO_2$ VMR peaks also occurred in January in three of the four regions and in December in the eastern U.S., in overall agreement with Fig. 6. Note too that monthly mean $NO_2$ VMRs were also underpredicted in all months in the western and eastern U.S., in agreement with Fig. 6, but some monthly overpredictions can be seen in western and eastern Canada. ~~Similarly, p~~Peak observed monthly mean $O_3$ ~~$O_3$~~ VMR values occurred in April in three of the four regions ~~and in May in the western U.S.~~, in ~~overall~~ agreement with Fig. 7, but in July in the western U.S. P~~but p~~eak predicted monthly mean $O_3$ VMR values, on the other hand, occurred in ~~both~~ March ~~and~~ or April, and b~~,~~oth underpredictions and overpredictions of monthly mean $O_3$ VMR are evident in Fig. 12 in all four regions. ~~Note too that monthly mean $NO_2$ VMRs were also underpredicted in all months in the western and eastern U.S., in agreement with Fig. 6, but some monthly overpredictions can also be seen in western and eastern Canada.~~

2013–2016 hindcasts of $NO_2$ and $O_3$

The RAQDPS-OP023 annual hindcasts for 2013–2016 can also be evaluated to look for consistencies in model performance across multiple years. Figure 13 shows plots of predicted spatial distributions of annual mean $NO_2$ and $O_3$ surface VMR fields for 2013–2016 and 2021/22. For both species the broad spatial patterns are very similar over land for the five years despite year-to-year variations in meteorology and the monotonic decrease of 18% in domain-total $NO_x$ emissions over the 2013–2016 period and the further decrease of 20% from 2016 to 2021/22 (Table 1). Nevertheless, year-to-year decreases in annual $NO_2$ levels are visible over this near-decadal period, including in Texas, the Ohio Valley, and the Washington, D.C.–Boston corridor. Latitudinal gradients in the spatial distributions of $O_3$ surface VMR can be seen over land in Fig. 13 for all five years, with a east–west band of elevated values stretching across the continental U.S. and peaking in the elevated terrain of the U.S. Rocky Mountain and Great Basin regions. A recent analysis of winter and spring surface $O_3$ observations over North America for 2010–2014 showed a similar pattern (Gaudel et al., 2018). However, trends in $O_3$ from 2013 to 2021/22 are not obvious in ~~this f~~Fig. 13~~ure~~, unlike those for $NO_2$.

Table 3 lists values of observed and RAQDPS-OP023-predicted all-station annual mean $NO_2$ and $O_3$ surface VMRs for 2013–2016 as well as 2021/22. Note that the statistics for the 2013–2016 period are based on quality-assured, retrospective observation data sets released by individual agencies rather than the NRT measurements used to evaluate

2021/22 forecasts (see Fig. S3 for station locations). Consistent with Fig. 13, observed and predicted all-station annual mean $NO_2$ VMR values both exhibit a monotonic decrease from 2013 to 2021/22, though for a smaller set of U.S. stations in 2021/22 (Table 2), whereas observed all-station annual mean $O_3$ VMR values exhibit little change over this period vs. a small upward trend for predicted all-station annual mean $O_3$ VMR. Table 3 also lists ~~domain-wide~~all-station annual values of 10 other model performance statistics for the five years. Statistics for the hindcasts might be expected to be better than the forecasts due to the use of year-specific emissions. In fact, the scores are mixed and are comparable overall for the five years. For example, annual NMB values for $NO_2$ VMR are negative for 2021/22 but positive and smaller in magnitude for the other four years, annual RMSE and NMAE values for $NO_2$ VMR are better for 2021/22 than for 2013–2016, but FAC2 and R values for $NO_2$ VMR are better for 2013–2016 than for 2021/22. The annual NMB and R values for $NO_2$ VMR for 2013–2016 all meet~~exceed~~ the benchmark goals of ±0.20 and 0.60 for good performance recommended by Zhai et al. (2024), similar to 2021/22, but annual NMAE scores for 2013–2016 do not meet either of their recommended thresholds (0.40 or 0.55). Annual NMB values for $O_3$ VMR, on the other hand, are more negative for 2013–2016 than for 2021/22 and FAC2 scores are also lower while NMAE and R scores are comparable. In addition, the annual NMB and R scores for $O_3$ VMR for 2013–2016 ~~are~~ all ~~above~~ meet the acceptable benchmarks of ±0.15 and 0.50 for these statistics recommended by Emery et al. (2017), similar to 2021/22, but fall below the more stringent benchmark goals of ±0.05 and 0.75, while NMAE scores for 2013–2016 do not meet either recommended threshold (0.15 or 0.25).

Additional analyses

Results from additional data analyses for $NO_2$ and $O_3$ with a focus on the 2013–2016 RAQDPS-OP023 hindcasts can be found in Sect. S3.2.1. These results include tables of separate annual and seasonal scores for the AQS and NAPS networks as well as seasonal and regional diurnal analyses for 2021/22, regional scores for all five years, spatial plots of both annual station scores and predicted seasonal mean surface VMR fields for 2013–2016, ~~seasonal and regional diurnal analyses,~~ and monthly time series, monthly density scatterplots, and urban vs. rural monthly time series for 2013–2016. One finding from these supplemental analyses is the high level of consistency between the aggregated annual statistical scores across years for both species for the AQS and NAPS networks individually (Table S3A) and annual scores at the individual station level (e.g., Fig. S42). Another is the clear consistency across the 2013–2016 hindcasts of the seasonal variations in the $NO_2$ and $O_3$ seasonal mean VMR fields, statistical scores, ~~diurnal time series,~~ and monthly mean time series (e.g., Table S3S, Fig. S1140), which helps to identify systematic model errors. Third, the overall agreement in observed and predicted temporal trends also provides support for the representativeness of the year-specific emissions used for these hindcasts (e.g., Fig. S140~~1~~). Fourth, there are some striking differences between the time series of monthly $NO_2$ and $O_3$ statistics for urban stations vs. rural stations that underline the importance of emissions forcing, including higher $NO_2$ levels and lower $O_3$ levels in urban areas (e.g., Fig. S209 vs. Fig. S210). Fifth, ~~the larger annual, seasonal, and monthly negative biases for $NO_2$ that were found in the 2021/22 forecasts for U.S. regions vs. Canadian regions (e.g., Fig. S168) suggest that the U.S. $NO_x$ emissions used for these forecasts may have been too low whereas the Canadian $NO_x$ emissions that were used were more representative of 2021/22 conditions. The same pattern was not seen in the 2013–2016 hindcasts, which used different emissions. To detect this possible~~

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~~issue, however, statistics for the individual networks had to be computed and then compared (Table S3A). Sixth,~~ the annual regional analysis found that the two western regions had the largest negative NMB values for $O_3$, consistent with Fig. 12 and pointing to possible issues with the $O_3$ lateral boundary conditions for the western boundary (Table S7). ~~And seven~~Sixth, the monthly density scatterplots reveal an obvious precision limitation (only whole numbers) in the reported hourly $NO_2$ measurements (Fig. S169).

A seventh finding was that annual, seasonal, and monthly negative NMB values were considerably larger for $NO_2$ in the 2021/22 forecasts for the two U.S. regions (-0.31, -0.25) vs. the two Canadian regions (-0.06, -0.03). The same pattern was not seen in the 2013-2016 hindcasts, which used retrospective rather than projected emissions (Table S7). This is one example of the value of computing and then comparing statistics for the individual networks as well as the combined networks (e.g., Tables S3A, S3S). The differences in biases between countries found for the 2021/22 forecasts might suggest that the U.S. $NO_x$ emissions used for these forecasts were too low, whereas the Canadian $NO_x$ emissions that were used were more representative of 2021/22 conditions. Since some time has passed since 2021 when the RAQDPS-OP023 became operational, retrospective national emissions inventories are now available for 2021 for both Canada and the U.S. (ECCC, 2025a; U.S. EPA, 2025). Interestingly, a comparison of these inventories with Table 1 showed that the projected annual $NO_x$ inventory emissions considered for both countries were in fact higher than the actual inventory values for 2021: 8% higher for the U.S. and 13% higher for Canada. Other explanations must thus be sought.

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### 3.2.2 $PM_{2.5}$ total mass

2021/22 operational forecasts of $PM_{2.5}$ total mass

Figure 1 also shows the spatial distribution over North America of the annual mean $PM_{2.5}$ hourly surface concentration field (without sea–salt component) predicted by the RAQDPS-OP023 for 2021/22. Coloured divided dots are again superimposed to show observed and predicted annual values at NRT hourly $PM_{2.5}$ measurement stations for the same period (see Fig. S2 for station locations). It is clear from this figure that the RAQDPS-OP023 underpredicts annual $PM_{2.5}$ levels for 2021/22 at a majority of measurement stations.

Table 3 lists values of observed and RAQDPS-OP023-predicted all-station annual mean $PM_{2.5}$ surface concentrations (including the sea–salt component) and 10 other annual evaluation statistics for hourly $PM_{2.5}$ forecasts for all stations measuring hourly $PM_{2.5}$ (including both Class III FEM and non-FRM/FEM monitors: see Sect. S2.3 for more detaiils) for 2021/22. All-station annual MB and NMB values for forecast hourly $PM_{2.5}$ concentrations were -2.5 $\mu g \cdot m^{-3}$ and -0.31, respectively, consistent with Fig. 1. Other statistics in Table 3 for $PM_{2.5}$ were also poorer (e.g., NMAE=0.66, FAC2 =0.46, R=0.24, NSD=0.69) than corresponding scores for $NO_2$ and $O_3$. Emery et al. (2017) proposed "acceptable" score benchmarks (i.e., above 33rd percentile for a historical multi-model ensemble) for predicted $PM_{2.5}$ total mass for NMB, NMAE, and R scores of ±0.30, 0.50, and 0.40, but none of these benchmarks were met for the 2021/22 hourly $PM_{2.5}$ forecasts.

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Figure 4 shows the spatial distribution of station-specific annual values of MB, NMB, CRMSE, and R for hourly $PM_{2.5}$ total mass for 2021/22 based on NRT hourly measurements at AirNow and NAPS stations. Consistent with Fig. 1 and Table 3, annual MB was negative for most stations, but values were small at a minority of stations and even positive at a handful of stations. Annual NMB values, on the other hand, were greater than -0.2~~4~~0 at the majority of stations, especially in the west. Annual CRMSE values were highest in the western U.S away from the coast; as discussed in Sect. 4.2 these western scores were influenced by the lack of BB emissions in the RAQDPS-OP023 runs. Lastly, annual R scores were highest in the northeast and along the U.S. west coast and we~~a~~re lowest for many Rocky Mountain, Great Plains (central U.S.), and Prairie (central Canada) stations.

Figure 5 shows the ~~predicted~~ spatial distributions of seasonal mean $PM_{2.5}$ hourly surface concentration fields (without sea-~~-~~salt component) for 2021/22 predicted by the RAQDPS-OP023, again with superimposed divided dots that show observed and predicted seasonal mean $PM_{2.5}$ concentration values at NRT hourly measurement stations. Similar to Fig. 1, observed seasonal mean values were higher in general than predicted seasonal mean values, especially in the summer and in the west (see Sect. 4.2 for a discussion of the impact of the inclusion of ~~wildfire~~ BB emissions).

Figure 8 adds further temporal detail by showing time series of observed and RAQDPS-OP023-predicted all-station monthly mean $PM_{2.5}$ concentration values for 2021/22 as well as time series of monthly NMB, CRMSE, and R values based on all NRT measurements of hourly $PM_{2.5}$ surface concentration in the model domain. The observed all-station monthly mean $PM_{2.5}$ concentration time series has a large August peak and a lower January peak but the reverse is true for the predicted all-station monthly mean $PM_{2.5}$ time series. The RAQDPS-OP023 also underpredicted monthly mean $PM_{2.5}$ concentrations in all months, with the largest underpredictions occurring in the summer and the smallest in the winter. While monthly NMB was negative for all months, it was most negative for spring and summer, with an extreme value close to -0.6 in August. Monthly CRMSE was quite variable, but with a pronounced primary peak in August and a secondary peak in January. Monthly R values did not vary much for winter and spring, but they were considerably lower for summer and autumn and were close to zero for July and August. The poor summer scores were in large part due to the neglect of BB emissions (Sect. 4.2).

Figure 11 shows all-station annual-mean diurnal time series of five statistics for hourly $PM_{2.5}$ surface concentration for 2021/22. RAQDPS-OP023~~Model~~ performance for $PM_{2.5}$ clearly varied with time of day. Predicted annual-mean hourly concentrations were biased low at all hours of the day, and they displayed more diurnal variability than the measurements. The largest negative annual-mean hourly NMB value occurred in the early afternoon (14 LT) while the smallest values occurred at the beginning of the morning rush hour (06 LT) and towards the end of the evening rush hour (20 LT). Annual-mean hourly CRMSE and R values, on the other hand, exhibited relatively small diurnal variations.

It is also of interest to examine how RAQDPS-OP023~~model~~ performance varied with geography. Figure 12 compares regional time series of observed and predicted regional monthly means of hourly $PM_{2.5}$ surface concentration for 2021/22 for the four continental quadrants (Fig. S7). Peak observed monthly mean $PM_{2.5}$ concentrations occurred in

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July or August and were higher in the west than in the east, but as in Fig. 8 there was also a secondary peak in the cold season in December or January for all four regions (cf. Fig. 5). In contrast the predicted monthly mean $PM_{2.5}$ concentrations had a primary cold-season peak in December or January and a weak warm-season secondary peak in July or August, again consistent with Fig. 8. Note that the RAQDPS-OP023-predicted warm-season peak occurred without any contribution from BB emissions ~~(see Sect. 4.2)~~, suggesting a second warm-season emissions source such as biogenic secondary organic aerosol (SOA). Another regional difference is that predicted monthly mean $PM_{2.5}$ concentrations were higher than the observed values in early winter for eastern Canada but were lower for all months for the other three regions.

2013–2016 hindcasts of $PM_{2.5}$ total mass

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Figure 13 shows the RAQDPS-OP023-predicted spatial distributions of annual mean $PM_{2.5}$ surface concentration fields (including the sea–salt component) over ~~North America~~the model domani for 2013–2016 and 2021/22. The spatial patterns of annual mean $PM_{2.5}$ concentration for these five years are broadly similar, although some minor ~~interannual~~ variations between years can be seen over the ocean regions due to variations in sea-salt emissions ~~due to~~that result from interannual differences in near-surface wind speed (cf. Fig. S9). Over North America, a~~A~~nnual mean $PM_{2.5}$ surface concentrations, like annual mean $NO_2$ surface VMRs, are highest over the eastern U.S. and California and lower over most of Canada and the rest of the western U.S., with the exception of isolated urban areas in the western U.S. and a tongue of elevated $PM_{2.5}$ levels over the Canadian province of Alberta. The impact of the large decreases in domain-total anthropogenic emissions of two $PM_{2.5}$ precursors, $SO_2$ and $NO_x$, of 60% and 36~~1~~% from 2013 to 2021 (Table 1) is also reflected in this figure by a decrease in $PM_{2.5}$ levels with time over North America (see also the multi-year plots of annual mean $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$ concentration fields in Fig. 14).

Table 3 lists annual values of 12 evaluation statistics for $PM_{2.5}$ hourly surface predictions for all stations measuring hourly $PM_{2.5}$ for the 2013–2016 RAQDPS-OP023 hindcasts in addition to the 2021/22 forecasts. The values of predicted annual mean $PM_{2.5}$ surface concentrations show a noticeable downward trend across the five simulation years: the observed annual mean surface concentrations also show a downward trend but it is much weaker. This difference might be due in part to the impact of BB emissions of $PM_{2.5}$ on observed ambient $PM_{2.5}$ values that is missing from the predicted ambient $PM_{2.5}$ values. Overall, the 2013–2016 scores for each statistic were very similar amongst themselves, suggesting consistent behaviour in RAQDPS-OP023 $PM_{2.5}$ predictions from year to year, but some differences are evident between the 2013–2016 scores and the 2021/22 scores. Annual NMB values were negative for all five years, but the range for 2013–2016 was -0.06 to -0.09, considerably better than the 2021/22 value of -0.31. Annual NMAE scores, on the other hand, were slightly worse for 2013–2016 (0.71–0.73) than 2021/22 (0.66), while annual FAC2 scores were slightly better (0.48–0.50 vs. 0.46) and annual R scores were comparable (0.17–0.29 vs. 0.24). Note that all of the annual NMB scores for 2013–2016 (unlike 2021/22), but none of the NMAE and R scores, met the $PM_{2.5}$ benchmarks for acceptable performance recommended by Emery et al. (2017).

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In addition to hourly continuous $PM_{2.5}$ total mass concentration measurements, there are also roughly 900 daily $PM_{2.5}$ monitors operating in North America that make 24-hour $PM_{2.5}$ total mass measurements using a filter-based, gravimetric approach (see Sect. S2.3 and Table S2c). These daily gravimetric $PM_{2.5}$ measurements mainly support regulatory and human health applications and are not available in near-real time, but they are notable because they represent an independent data source for model evaluation that can supplement the Class III FEM and non-FRM/FEM hourly continuous $PM_{2.5}$ concentration measurements used for NRT evaluations and the Table 3 statistics. Daily gravimetric $PM_{2.5}$ measurements also have different uncertainties than the hourly continuous $PM_{2.5}$ measurements, including negative artefacts due to the volatilization of semi-volatile species such as ammonium, nitrate, and particle water during transport to and inside the controlled laboratory environments where filter analyses are performed (e.g., Frank, 2006; Dabek-Zlotorzynska et al., 2011; Malm et al., 2011; Chow et al., 2015; Hand et al., 2019). Gravimetric $PM_{2.5}$ monitor locations from the AQS, IMPROVE, and NAPS networks are shown in Fig. S6a (vs. Fig. S3c for continuous $PM_{2.5}$ monitors). Note that these combined networks include monitors with every-day, one-day-in-three, and one-day-in-six sampling frequencies, but roughly 65% of the daily gravimetric $PM_{2.5}$ mass measurements in the U.S. are made by non-speciation, mass-only monitors (Tables 5 and S2c; Malm et al., 2011).

Table 5 lists all-station annual evaluation statistics for gravimetric measurements of daily $PM_{2.5}$ mass for the 2013–2016 hindcasts. First, note that observed annual $PM_{2.5}$ total mass concentration values for the gravimetric data set in Table 5 have a range from 7.2 to 8.2 $\mu g \cdot m^{-3}$, similar to but smaller than the corresponding range of 7.4 to 8.7 $\mu g \cdot m^{-3}$ for the hourly continuous $PM_{2.5}$ data set from Table 3. The range of annual NMB scores for the gravimetric data set, however, is 0.0 to 0.07, slightly better and of opposite sign to the range of -0.09 to -0.06 ~~range~~ for the continuous $PM_{2.5}$ measurements in Table 3. Other annual scores for the 2013–2016 hindcasts are also better for the gravimetric measurement data set. The range of annual NMAE scores for the gravimetric $PM_{2.5}$ measurements is 0.51 to 0.54 (vs. 0.71 to 0.73 for the continuous $PM_{2.5}$ measurements), the range of annual FAC2 scores is 0.68 to 0.70 (vs. 0.48 to 0.50), and the range of annual R scores is 0.43 to 0.47 (vs. 0.17 to 0.29). Note the narrow range of each of these annual statistics for the four years, suggesting a high level of consistency in model performance. Note too that the range of annual NSD scores for the gravimetric measurements is 1.41 to 1.61, higher than the range of 0.78 to 1.36 for the continuous $PM_{2.5}$ measurements.

Some differences in scores for these two independent data sets are to be expected due to differences in monitor locations, in temporal aggregation (daily vs. hourly), and in measurement technologies. One reason for some of the better scores for the gravimetric $PM_{2.5}$ mass measurements may be their daily sampling period, which will reduce the influence of short-term model errors compared to the hourly continuous measurements (e.g., Appel et al., 2008, 2021). The sampling period would not, however, affect annual means or MB and NMB scores. One technical difference between the daily gravimetric $PM_{2.5}$ mass measurements and the continuous mass measurements is that the filter analysis for the former is performed under constant-temperature, low-humidity conditions in a laboratory after transport from the field and storage prior to analysis whereas the hourly continuous mass measurements are made under ambient conditions where temperature and humidity can vary widely. This means that daily gravimetric $PM_{2.5}$ mass values are

likely to be lower than daily continuous $PM_{2.5}$ mass values due to loss of some semi-volatile mass from ammonium nitrate, organic matter, or particle water from the sample filter, especially for filters exposed under high-humidity or low-temperature ambient conditions. The fact that annual NMB scores for gravimetric $PM_{2.5}$ mass measurements are more positive than those for continuous $PM_{2.5}$ measurements is thus somewhat surprising.

Additional analyses

The results of additional analyses for hourly continuous and daily gravimetric $PM_{2.5}$ total mass measurements with a focus on the 2013–2016 hindcasts are presented in Sect. S3.2.2. These results include tables of separate annual and seasonal scores for the individual AQS, IMPROVE, and NAPS networks as well as regional scores for all five years, spatial plots of both annual station scores and predicted seasonal mean $PM_{2.5}$ surface concentration fields for 2013–2016, seasonal and regional diurnal analyses for 2021/22, and monthly time series, monthly density scatterplots, and urban vs. rural monthly time series for 2013–2016. One insight from these additional analyses are the considerable variations between seasons that are evident across all five years in the seasonal mean $PM_{2.5}$ mass fields, in the seasonal statistical scores, and in monthly mean time series (e.g., Figs. S13, S142). Another is the high level of consistency between seasonal scores for $PM_{2.5}$ for the 2013–2016 hindcasts (and in many cases for the 2021/22 forecasts) at the aggregated all-station level and annual scores at the individual station level, which allows systematic model errors to be identified. For example, all-station monthly NMB scores were most negative in summer for all four years (Figs. S142 and S198). Third, some of the analyses suggest that the neglect of BB emissions by the RAQDPS-OP023 was an important contributing factor to its overall underpredictions of $PM_{2.5}$ total mass. Fourth, there were some striking differences for 2013-2016 between the time series of monthly $PM_{2.5}$ statistics for urban stations (Fig. S213) vs. rural stations (Fig. S214) that underline the importance of emissions forcing. For example, monthly NMB values for the urban stations were positive for some months whereas they were uniformly negative for the rural stations, suggesting that $PM_{2.5}$ underprediction is mainly a rural issue. Fifth, additional differences are shown between evaluation statistics for 2013-2016 for daily gravimetric $PM_{2.5}$ total mass measurements vs. hourly continuous $PM_{2.5}$ measurements. These differences include positive monthly NMB scores for the cold-season months for the gravimetric measurements (Fig. S198) vs. lower scores for the continuous measurements (Fig. S142). And sixth, some of the evaluation scores between individual networks were also very different, and certain differences in the overall characteristics of the individual networks, in particular the dominance of either urban stations or rural stations, can help to explain why ~~network~~ the scores between network were different. For example, for the daily gravimetric $PM_{2.5}$ measurements the seasonal MB and NMB scores for one measurement network (NAPS) were positive for all seasons and for a second network (AQS) they were positive for most seasons (Table S5S). For the hourly continuous $PM_{2.5}$ measurements, on the other hand, the seasonal MB and NMB scores were negative for both networks for all seasons. The fact that some scores point to model overpredictions of $PM_{2.5}$ make it clear that the negative all-station MB and NMB scores presented in Table 3 and Figs. 5, 8, and 11 do not tell the whole story. Multiple factors must be considered in addition to model formulation to explain these different scores, including the magnitude and distribution of emissions, seasonal and regional variations in meteorology, differences in network composition, and differences in measurement instrument characteristics.

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### 3.3 Expanded evaluation for additional species and processes

As described in the companion paper by Moran et al. (2026~~5~~), the RAQDPS023 predicts abundances of 47 individual and lumped gas-phase chemical species and 16 size bin-chemical components. While direct atmospheric measurements are not available for all of these species and components, this section describes evaluation results based on 2013–2016 QA/QCed measurements of nine atmospheric gases in addition to $NO_2$ and $O_3$, seven $PM_{2.5}$ chemical components, and three aqueous-phase inorganic ions.

#### 3.3.1 Other gases

Table 4 extends the all-station annual statistical scores reported in Table 3 for two gas-phase species, $NO_2$ and $O_3$, to nine other individual or lumped ADOM-2 gas-phase species predicted by the RAQDPS-OP023 for 2013–2016: NO, $NO_x$, $HNO_3$, $NH_3$, $SO_2$, CO, ETHE, HCHO, and ISOP, where ETHE is a lumped species (Sect. S2.4). The measurements considered were provided by five networks: AMoN, AQS, CAPMoN, CASTNET, and NAPS. As discussed in Sect. 2.3 and illustrated by Figs. S4 and S5, the available sets of measurements for these other gas-phase species are different from those for $NO_2$ and $O_3$ in terms of the numbers and the locations of surface monitors, and also, in some cases, the sampling period, which ranged from hourly to biweekly (see also Tables S2a and S2b). Looking at values of N, the number of complete measurements per year, in Table 4, we see that NO, $NO_x$, and CO have the most measurements by far, followed by ETHE and ISOP, then $SO_2$, HCHO and $HNO_3$, and lastly $NH_3$ with the fewest measurements. However, due to the different sampling period lengths, the number of measurements does not necessarily reflect the number of measurement stations. For example, there are more monitors measuring $HNO_3$, $NH_3$, and HCHO than there are ~~for~~ measuring ETHE and ISOP (Table S2b) even though N is smaller for the first three species. In fact, only 8 to 13 stations have complete annual measurements for ETHE and only 6 to 13 stations have complete annual measurements for ISOP for 2013–2016. Note that the evaluation scores for $HNO_3$, $NH_3$, and $SO_2$ will also be referred to in the next section, since these three species are precursors to three $PM_{2.5}$ chemical components.

Compared to the annual model performance for $NO_2$ and $O_3$ predictions for 2013–2016 summarized in Table 3, the overall model skill for these other gas-phase species presented in Table 4 is more varied. For example, all-station annual mean NO VMRs, like those for $NO_2$, were overpredicted for all four years (by 9% to 32%). All-station annual NMB, NMAE, FAC2, and R scores for hourly NO VMR, however, were less good than those for hourly $NO_2$ VMR for all four years, but this difference is at least partly due to the limited precision at which NO VMR measurements were reported (see Sect. S3.3.1). All-station annual NMB, NMAE, FAC2, and R scores for hourly CO and daily HCHO VMRs, on the other hand, were comparable to those for $NO_2$. All-station annual mean ~~$SO_2$,~~ $HNO_3$, $SO_2$, and ISOP VMRs were overpredicted for all four years (and also all months: see Sect. S3.3.1), by 16% to 33%, 43% to 60%, and 227% to 271%, respectively, whereas all-station annual mean $NH_3$ VMR was underpredicted for all four years (and all months: Fig. S145), by 42% to 47%. All-station annual ~~NMB scores for hourly~~mean ETHE ~~predictions vs. hourly ethene measurements~~was also overpredicted for all four years, by ~~ranged from 0.~~68% to 1~~.~~19% ~~(i.e., overpredictions)~~, but these scores were confounded by the ~~model's~~ inclusion of isoprene oxidation products in addition to ethene in this lumped VOC species (Sect. S2.4~~Moran et al., 2025~~), which suggest~~sing~~ that overpredictions should be expected. The

impact of decreasing $SO_2$ and $NO_x$ emissions in 2013–2016 (Table 1) can also be clearly seen in corresponding decreases in Table 4 in both observed and predicted all-station annual mean $SO_2$ and $HNO_3$ VMR values.

Additional analyses

More evaluation results for 2013–2016 for these other gas-phase species are provided in Sect. S3.3.1, including tables of annual and seasonal scores for individual Canadian and U.S. networks as well as all-station regional scores, spatial plots of both annual station scores and predicted seasonal mean surface VMR fields, time series of all-station monthly statistics, and monthly density scatterplots. It is clear from these additional analyses that all nine species exhibit strong seasonal variations: four species ($HNO_3$, $NH_3$, HCHO, ISOP) were observed and predicted to have summer maxima, four (NO, $NO_x$, $SO_2$, CO) were observed and predicted to have winter maxima, and ETHE was observed to have a winter maximum but predicted to have a summer maximum (Table S4S). As noted above, the disagreement for ETHE was not surprising ~~due~~ given to the inconsistency between measured ethene ~~($C_2H_4$)~~ and lumped model ETHE. ~~, and~~ However, winter ETHE scores, when isoprene emissions were low (Table S1), were better than those for the other three seasons, when isoprene emissions were higher (Table S4S). For example, winter NMB values ranged from -6% to 13%, suggesting that predictions of pure ethene VMR driven by pure ethene emissions demonstrated skill. As was the case for $NO_2$ and $O_3$, there were also marked similarities in the scores for these other gas-phase species across the four annual hindcast simulations, which supports identification of systematic model errors. For example, ~~$HNO_3$,~~ $SO_2$ and ISOP were consistently overpredicted and $NH_3$ was consistently underpredicted in all seasons and months (Table S4S; Figs. S146, S150, and S145).

The consistency in annual, seasonal, and monthly scores for the four years also suggests that the year-specific emissions used for these hindcasts were representative. However, some scores discussed in Sect. S3.3.1 raised concerns that Canadian $SO_2$ emissions and biogenic ISOP emissions may have been too high and North American $NH_3$ emissions too low for the 2013–2016 period. The decreases in $SO_2$ and $NO_x$ emissions from 2013 to 2016 were also reflected in time series of both observed and predicted monthly mean VMRs for NO, $HNO_3$, and $SO_2$ (Figs. S143, S144, S146) as well as $NO_2$ (Fig. S140). Performance benchmarks for CO and $SO_2$ proposed by Zhai et al. (2024) are also discussed in Sect. S3.3.1, and more CO scores than $SO_2$ scores met these benchmarks. Scores for individual networks can also be quite different owing to different network characteristics. $SO_2$ provides a good illustration of this behaviour as $SO_2$ measurements were available from four networks: scores for the CAPMoN and CASTNET networks were significantly better overall than those for the AQS and NAPS networks for both annual and seasonal evaluations (Tables S4A, S4S). It was noted that evaluation scores also tended to be similar for species with similar characteristics, such as primary species vs. secondary species and species having similar emissions sources (e.g., combustion). Observed and predicted annual and seasonal CV values for the other gas-phase species were also considered. Based on CV values, they fell naturally into three groups, depending on whether the pollutants were primary, secondary, or mixed primary-secondary in nature. Lastly, it was noted that a few scores for 2016 appeared to be outliers and that monthly density scatterplots for hourly NO and CO revealed obvious precision problems with reported measurements for these two species (Figs. S172, S176).

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### 3.3.2 $PM_{2.5}$ chemical composition~~nents~~

As emphasized by Bachmann (2013), $PM_{2.5}$ is a multi-pollutant. This means that accurate forecasts of $PM_{2.5}$ total mass require accurate forecasts of its underlying chemical components, each of which has different emissions sources and different formation pathways. Forecast errors for $PM_{2.5}$ total mass can thus be better understood by examining forecast errors for its underlying chemical components. Fortunately, three North American networks (CSN, IMPROVE, NAPS) make measurements of $PM_{2.5}$ composition (Sect. 2.3 and Table S2a). Station locations for these three networks are plotted in Fig. S6b, while Table S2c summarizes the number of stations for 2013–2016 for which $PM_{2.5}$ speciation measurements were available and the smaller number for which temporally representative seasonal and annual measurements were available (see Sect. S2.4).

Figure 14 shows plots of the spatial distributions of annual mean surface concentrations of nine $PM_{2.5}$ chemical components over North America for 2013–2016 and 2021/22 predicted by the RAQDPS-OP023. Note that the sets of contour intervals used vary by component, with $PM_{2.5}$-EC, $PM_{2.5}$-$NH_4$, and $PM_{2.5}$-CM having the smallest ranges and $PM_{2.5}$-SS and $PM_{2.5}$-TOM (= $PM_{2.5}$-POM + $PM_{2.5}$-SOM) having the largest ranges. It is clear from this figure that the predicted spatial distributions vary markedly between $PM_{2.5}$ chemical components. It is also clear that the predicted annual spatial distributions of each of these chemical components for these five years are broadly similar, pointing to the anchoring effect of relatively constant emissions to the atmosphere of primary $PM_{2.5}$ chemical components and $PM_{2.5}$ gas-phase precursors (e.g., Fig. S1), which for most pollutants changed relatively little from year to year (Table 1). The annual mean surface concentrations of $PM_{2.5}$ total mass shown in Fig. 13 ~~also displayed this~~were also broadly similar from year to year~~ity~~. However, a comparison of the spatial distributions of annual mean $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$ surface concentrations from 2013 to 2021/22 in Fig. 14 does suggest decreasing concentrations of these two chemical components over this period, consistent with the large decreases in domain-total anthropogenic emissions of $SO_2$ and $NO_x$ from 2013 to 2021/22 (Table 1). Interestingly, annual mean $PM_{2.5}$-$NH_4$ surface concentration fields can also be seen to decrease from 2013 to 2021/22 despite nearly constant $NH_3$ emissions over this period. This downward trend is due instead to the reduced availability of gaseous $H_2SO_4$ and $HNO_3$ in the atmosphere, which form sulfate and nitrate particulate salts with $NH_3$. Domain-wide primary $PM_{2.5}$ emissions also decreased by 8% from 2013 to 2016 (Table 1), and some decline is evident from 2013 to 2016 for $PM_{2.5}$-EC and $PM_{2.5}$-POM in Fig. 14. The annual mean spatial distributions of other $PM_{2.5}$ components display only small year-to-year variations with the exception of sea salt, for which interannual variations in mean surface concentration are driven by interannual variations in surface wind speed (see Fig. S9).

Although the above discussion suggests the close connection between the spatial distributions of emissions of primary $PM_{2.5}$ components and $PM_{2.5}$ gas-phase precursors and the resulting spatial distributions of $PM_{2.5}$ chemical components in the atmosphere, it is not a simple relationship since many chemical and physical processes in the atmosphere modulate the chemical transformation and removal pathways of these multiple chemical components. For example, the annual spatial distributions of $PM_{2.5}$-$SO_4$ shown in Fig. ~~ure~~ 14 are quite smooth, which reflect its origin as a secondary pollutant resulting from gas-phase or aqueous-phase oxidation of North American $SO_2$ emissions, even

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though the $SO_2$ emissions shown in Fig. S1d are largely emitted by isolated point sources (e.g., ECCC, 2018; Foley et al., 2023). The spatial distribution in Fig. 14 of $PM_{2.5}$-SOM, another secondary pollutant, is also smooth, but its maximum is located further south over the southeastern U.S., where biogenic VOC emissions are high, especially in the summer season (cf. Fig. S27). It is also clear from Fig. S1 that there are marked differences in the spatial distributions of some of the anthropogenic emissions associated with different $PM_{2.5}$ chemical components. For example, the majority of $NO_x$ emissions are located in the eastern half of North America, but the locations of some major highways and large urban areas in western North America are visible in the $PM_{2.5}$-$NO_3$ surface concentration panels in Fig. 14, which suggests the important contributions of on-road mobile sources and population centres to $NO_x$ emissions. Emissions of primary $PM_{2.5}$ and of $NH_3$ gas, the precursor to $PM_{2.5}$-$NH_4$, on the other hand, are stronger over the North American interior (Figs. S1e,f). Elevated $NH_3$ emissions from agricultural activities in the midwestern U.S. are visible in Fig. S1e as well as fertilizer application in the San Joaquin Valley of California, large animal feedlot operations in Texas and Oklahoma, and extensive swine production in North Carolina. While $NH_3$ emissions are dominated by agricultural activities, some $NH_3$ emissions are also associated with population centres due to on-road mobile emissions (e.g., Toro et al., 2024). In addition, the chemical composition of primary $PM_{2.5}$ emissions depends on the emissions source type. Combustion sources dominate $PM_{2.5}$-EC and $PM_{2.5}$-POM emissions, as ~~can be seen~~suggested in Fig. 14 by their association with major highways and population centres, while the $PM_{2.5}$-TOM surface concentration field displays characteristics of both the $PM_{2.5}$-POM and $PM_{2.5}$-SOM fields. Fugitive dust emissions from paved and unpaved roads are the main source of $PM_{2.5}$-CM, and $PM_{2.5}$-CM surface concentrations can be seen in Fig. 14 to be elevated in both urban centres and rural areas. Lastly, the spatial distribution of $PM_{2.5}$-SS is dominated by its oceanic sources, but the limited transport of sea salt from the oceans inland over most of North America that is evident in Fig. 14 should also be noted.

Table 5 presents all-station annual scores for seven $PM_{2.5}$ chemical components for 2013–2016 for the three $PM_{2.5}$ speciation networks combined. The scores for each component tend to be similar from year to year, but ~~these~~ scores can also vary considerably between components. For example, all-station annual NMB scores for $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$ were negative for all four years whereas all-station annual NMB scores for $PM_{2.5}$-$NH_4$, EC, CM, and SS were positive for all four years. Only $PM_{2.5}$-TOM had small all-station annual NMB values of both signs for this period. Note that the $PM_{2.5}$-$NH_4$ overpredictions are inconsistent with the underpredictions of both $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$ given the inorganic salts formed by these three components. One possible explanation is that this is an artefact due to the lack of available IMPROVE $PM_{2.5}$-$NH_4$ measurements (e.g., Solomon et al., 2014) so that only CSN and NAPS measurements were considered for $PM_{2.5}$-$NH_4$ in Table 5. However, the same inconsistency can be seen in Table S5A for just CSN measurements. Another possible explanation is that the RAQDPS-OP023 does not consider the neutralization of $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$ by base cations, which would reduce $PM_{2.5}$-$NH_4$ concentrations and increase $NH_3$ VMR (e.g., Vasilakos et al., 2018; Miller et al., 2024; Semeniuk et al., 2025). All-station annual NMAE scores in Table 5 for 2013–2016 were lowest for $PM_{2.5}$-$SO_4$ (~0.47), followed in ~~rank~~ ascending order by annual NMAE scores for $PM_{2.5}$-$NO_3$, TOM, EC, $NH_4$, CM, and SS (~1.45). All-station annual FAC2 scores were highest for $PM_{2.5}$-$SO_4$ (~0.64), followed in ~~decreasing~~ descending order by those for $PM_{2.5}$-EC, TOM, $NH_4$, $NO_3$, SS, and CM (~0.31).

All-station annual R scores were also highest for $PM_{2.5}$-$SO_4$ (~0.66), followed in ~~decreasing~~ descending order by those for $PM_{2.5}$-$NO_3$, EC and SS, $NH_4$, TOM, and CM (~0.17). Based on the annual NMAE, FAC2, and R scores taken together, the RAQDPS-OP023 showed the most skill for $PM_{2.5}$-$SO_4$ followed in descending order by $PM_{2.5}$-$NO_3$, EC, TOM, $NH_4$, SS, and CM.

Note also that all-station annual NSD scores fell into two groups: values for $PM_{2.5}$-$SO_4$, $NO_3$, and $NH_4$ were less than ~~one~~ 1.0 whereas values for $PM_{2.5}$-EC, TOM, CM, and SS were greater than ~~one~~1.0. The three components in the first group are all secondary components whereas those in the second group are all primary components or a mixed primary-secondary component in the case of $PM_{2.5}$-TOM. While this difference might appear to suggest that the model is overemphasizing the contribution of temporal variations due to emissions, the $PM_{2.5}$ speciation measurements are 24-hour samples so that the observed and predicted temporal variation at measurement locations can only be due to interday and longer variations. For anthropogenic emissions this would point to the day-of-week and month-of-year temporal profiles that have been assumed by the emissions processing system for different source sectors (Sect. S2.2), but emissions for the two primary components that had the largest NSD values, $PM_{2.5}$-CM and SS, are also the ones most affected by meteorology.

It is worth noting that downward trends can be seen in Table 5 in the values of both observed and predicted annual mean concentrations for $PM_{2.5}$-$SO_4$, $NO_3$, and $NH_4$, consistent with the monotonic decreases in $SO_2$ and $NO_x$ emissions that occurred over this period and with Fig. 14. For the $PM_{2.5}$-$SO_4$ evaluation statistics, the values of annual RMSE and R also decreased from 2013 to 2016. This is similar to decreases in these two statistics reported by Kelly et al. (2019) for the CMAQ model over the 2007 to 2015 period, which they attributed to decreasing $SO_2$ emissions and lower summertime $PM_{2.5}$-$SO_4$ peaks, which in turn reduced the $PM_{2.5}$-$SO_4$ "signal" (e.g., Chan et al., 2018). Note also that 2016 annual $\overline{O}$, $\overline{M}$, MB, NMB, RMSE, R, $\sigma_O$, and $\sigma_M$ scores for $PM_{2.5}$-$SO_4$, $NO_3$, EC, and OC for two recent versions of the CMAQ model (with wildfire emissions) were ~~given~~ reported in Appel et al. (2021). Although minor methodological differences such as inclusion or exclusion of NAPS measurements are suggested by comparisons of the 2016 $\overline{O}$ and $\sigma_O$ scores ~~for the two models~~in Table 5 and in Appel et al. (2021), there is rough agreement between the 2016 RAQDPS-OP023 scores and CMAQ scores ~~and the RAQDPS023 2016 scores in Table 5~~ for $PM_{2.5}$-$SO_4$, $NO_3$, and EC while RAQDPS-OP023 $PM_{2.5}$-TOM scores ~~CMAQ $PM_{2.5}$ OC scores~~ cannot be compared directly with CMAQ $PM_{2.5}$-OC scores ~~RAQDPS023 $PM_{2.5}$ TOM scores~~.

$PM_{2.5}$ reconstructed mass

$PM_{2.5}$ speciation measurements can also be used to calculate $PM_{2.5}$ reconstructed dry mass (i.e., excluding aerosol water) as a weighted sum of the individual $PM_{2.5}$ chemical ~~species~~ components (e.g., Malm et al., 2011; Chow et al. 2015; Hand et al., 2019). This is often done ~~using~~ with the IMPROVE formula, which uses some measured components~~species~~ directly and some measured components as proxies for the TOM, CM, and SS components, which are not measured directly. The RAQDPS-OP023 $PM_{2.5}$ total mass forecasts, on the other hand, which are the sum of the seven predicted $PM_{2.5}$ ~~speciated~~ chemical components shown in Table 5 (including SS but excluding aerosol water),

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are thus also a reconstructed dry mass but do not require the use of any proxies. A slightly modified version of the IMPROVE formula has been used in this study to calculate $PM_{2.5}$ reconstructed mass from speciation measurements (see Sect. S3.3.2 for more details).

Figure 15 compares observed and RAQDPS-OP023-predicted all-station seasonal mean $PM_{2.5}$ reconstructed dry mass and chemical composition for 2013–2016 based on combined CSN, IMPROVE, and NAPS measurements that were mass-complete (Sect. S3.3.2). There is good agreement overall between the seasonal means of observed and predicted $PM_{2.5}$ reconstructed dry mass ~~seasonal means~~. Predicted $PM_{2.5}$ total mass was greater than observed $PM_{2.5}$ reconstructed total mass for all four winters, while the opposite was true for all four summers, with closer agreement for spring and autumn. This good agreement might seem surprising given the consistent model underpredictions of hourly $PM_{2.5}$ mass measurements described in Sect. 3.2.2, but it is more consistent with the better evaluation results for daily gravimetric $PM_{2.5}$ mass measurements also discussed in that section. Similar comparisons for the CMAQ model against CSN and IMPROVE measurements have been presented for 2011 and 2016 simulations by Appel et al. (2017, 2021). Interestingly, both models overpredicted $PM_{2.5}$ total mass in winter 2016 and underpredicted it in summer 2016.

In addition, the stars plotted in Fig. 15 indicate the seasonal means of observed gravimetric $PM_{2.5}$ total mass for 2013–2016. It is natural to compare gravimetric $PM_{2.5}$ total mass and $PM_{2.5}$ reconstructed mass since they both attempt to measure of the same quantity. Good agreement can be seen in the observations between these two measurements for three seasons but not for the summer, for which the gravimetric total mass was larger. This is~~,~~ consistent with previous findings by Malm et al. (2011) for the CSN and IMPROVE networks. Predicted seasonal means of $PM_{2.5}$ total mass, on the other hand, were greater than the gravimetric seasonal means in winter and autumn but were even smaller than the observed $PM_{2.5}$ reconstructed mass in summer. A closely related quantity, the $PM_{2.5}$ residual mass, is defined to be the difference between gravimetric $PM_{2.5}$ mass and reconstructed $PM_{2.5}$ mass (e.g., Hand et al., 2019). Observed seasonal mean $PM_{2.5}$ residual mass was greater than 0.5 $\mu g \cdot m^{-3}$ for summer but was small otherwise, whereas predicted seasonal mean $PM_{2.5}$ residual mass was greater than 1 $\mu g \cdot m^{-3}$ for summer but was negative for winter, with values ranging from -1.47 to -0.90 $\mu g \cdot m^{-3}$ (see Table S5S-mr). Hand et al. (2019) found observed seasonal $PM_{2.5}$ residual mass to be mostly positive after 2011 with a strong summer peak, consistent with Fig. 15. Given the marked ~~model~~ RAQDPS-OP023 underpredictions of summer mean gravimetric $PM_{2.5}$ total mass but overpredictions of winter mean gravimetric $PM_{2.5}$ total mass (Sect. S3.2.2), the factors that contribute to the non-negligible observed summer $PM_{2.5}$ residual mass may be of interest because they may also be relevant to the model performance. Malm et al. (2011), Chow et al. (2015), and Hand et al. (2019) have suggested that some of these factors may be the neglect of particle-bound water in calculating $PM_{2.5}$ reconstructed mass, ammonium and nitrate volatilization under laboratory conditions, and seasonal variations ~~in the OM:OC scaling ratio~~ (lower in winter, higher in summer) in the OM:OC scaling ratio needed to scale observed OC concentration to ambient TOM concentration. Note too that the volatilization of $PM_{2.5}$-$NO_3$ and $PM_{2.5}$-$NH_4$ during the transport, storage, and analysis of filter samples, a negative measurement artefact, will create spurious positive model biases for these two $PM_{2.5}$ components (cf. Table 5).

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Figure 15 also shows good agreement overall between observed and RAQDPS-OP023-predicted seasonal mean $PM_{2.5}$ chemical composition. Looking component by component, t~~T~~he dominant contribution of the $PM_{2.5}$-EC and TOM carbonaceous components to $PM_{2.5}$ total mass is evident in both the observed and predicted stacked bar graphs as are the anticorrelated seasonal variations in $PM_{2.5}$-$SO_4$ and $PM_{2.5}$-$NO_3$. Overpredictions of $PM_{2.5}$-TOM concentration in the winter but underpredictions in the summer can be seen for all four years. The decrease in the observed and predicted contribution of the three major inorganic ions ($SO_4$, $NO_3$, $NH_4$) to $PM_{2.5}$ total mass from 2013 to 2016 due to decreases in annual $SO_2$ and $NO_x$ emissions can also be seen. Some of the annual mean biases for the $PM_{2.5}$ chemical components noted in Table 5 are also reflected in almost all seasons in Fig. 15, including the underpredictions of $PM_{2.5}$-$SO_4$ and $NO_3$ and overpredictions of $PM_{2.5}$-EC and SS. In fact, the reduction of bias for predictions of $PM_{2.5}$ total mass that is associated with the summation of these components, with their individual underpredictions and overpredictions is an example of the positive impact that compensating errors can have on model skill, and it demonstrates the value of a more comprehensive evaluation of model performance, in this case evaluating the prediction of $PM_{2.5}$ chemical components in addition to $PM_{2.5}$ total mass.

Since Fig. 15 is based on measurements from sampling sites located mainly over the continental U.S. (see Fig. S6b~~a~~), it does not account for $PM_{2.5}$ composition over northern Mexico and most of Canada. Figure 16, by contrast, shows a monthly time series of the RAQDPS-OP023-predicted mean $PM_{2.5}$ chemical composition averaged over the land portion of the domain and the 2013–2016 simulations, with $PM_{2.5}$-TOM separated into POM and SOM (not possible for the measurements). Seasonal variations of $PM_{2.5}$-$SO_4$, $NO_3$, POM, SOM, and SS can be clearly seen, whereas seasonal variations of $PM_{2.5}$-$NH_4$, EC, and CM are less pronounced. $PM_{2.5}$-$NO_3$ and POM are predicted to have winter maxima and summer minima, whereas $PM_{2.5}$-$SO_4$ and SOM are predicted to have winter minima and summer maxima. Note that the total inorganic component (sum of $PM_{2.5}$-$SO_4$, $NO_3$, and $NH_4$) only has a relatively small monthly variation with a minimum in November. The predicted peak monthly mean $PM_{2.5}$ total dry mass occurs in August, driven by monthly maximum values of $PM_{2.5}$-$SO_4$ and SOM. This is different from Figs. 8 and15, where the highest $PM_{2.5}$ total mass was predicted to occur in the winter, but note that predicted $PM_{2.5}$ total mass in Figs. 8 and 15 is roughly 6~~5~~ $\mu g \cdot m^{-3}$ vs. 1.5 $\mu g \cdot m^{-3}$ in Fig. 16, which suggest~~s~~ing that the former overweight~~s~~ the influence of urban areas. One other interesting feature in Fig. 16 is the SS maximum in August, which is in apparent contradiction to the SS wintertime maximum evident in Fig. S30. However, while Fig. S30 showed that inland penetration of sea salt over North America is limited, some seasonal variations are evident near California, ~~Florida and~~ the U.S. Gulf coast, and Florida that may be associated with the occurrence of sea-land breezes in the warm season (and observed seasonal-mean SS values in Table S5S are largest in the spring and summer). Lastly, Fig. S203 presents a similar analysis to Fig. 16 but for averaging over the full domain. Unlike Fig. 16 t~~T~~he SS component clearly dominates $PM_{2.5}$ ~~bulk~~ total dry mass in this figure, ~~unlike Fig. 16,~~ reflecting predicted high levels of sea salt over the Pacific and Atlantic Oceans. In addition, the largest monthly mean SS concentrations in Fig. S203 occur in the cold season, consistent with Fig. S30.

It is also informative to look at the diurnal variation of the $PM_{2.5}$ chemical components. Figure 17 shows the predicted diurnal variation of eight $PM_{2.5}$ chemical components for each season after averaging over the 2013–2016 simulations

and all North American continental grid cells. It is clear from this figure that both $PM_{2.5}$ total mass and chemical composition are predicted to vary with time of day. $PM_{2.5}$ total mass has a maximum for three seasons at 12 UTC (=7 EST), near sunrise and morning rush hour, and a minimum in all seasons at 21 UTC (=16 EST). The wintertime maximum, on the other hand, occurs at 04 UTC (i.e., near local midnight), pointing to a different balance between surface emissions and vertical stability and mixing. Note too that the individual $PM_{2.5}$ chemical components display different diurnal behaviours. Hourly $PM_{2.5}$-$SO_4$ concentration is predicted to be effectively constant, consistent with a nonvolatile~~,~~ secondary ~~regional~~ pollutant. Hourly $PM_{2.5}$-$NO_3$ and $NH_4$ concentrations, by contrast, are lowest in the afternoon when near-surface temperature is highest, and highest at night, especially before sunrise when near-surface temperature is often lowest. This behaviour is consistent with the semi-volatile nature of ammonium nitrate, for which lower temperatures favour the particle phase (e.g., Malm et al., 2004; Yu et al., 2005). Hourly $PM_{2.5}$-EC, POM, and CM concentrations are also predicted to be highest during the night and lowest in the afternoon, but this behaviou is likely due to greater vertical mixing of local emissions during the day, which reduces the near-surface buildup of these three primary species. Hourly $PM_{2.5}$-SOM, which is assumed to be nonvolatile in the RAQDPS-OP023 (see companion paper by Moran et al., 2026~~5~~), behaves like $PM_{2.5}$-$SO_4$ and displays little diurnal variation. Finally, sea-salt concentrations tend to be higher at night and lower during the day, likely due to diurnal variations in surface wind speed and hence in sea-salt emissions.

Note that the daily measurements made by the $PM_{2.5}$ speciation networks are not able to confirm the~~se~~ diurnal variations of $PM_{2.5}$ chemical components in Fig. 17 predicted by the RAQDPS-OP023~~model~~. However, the all-station, annual-mean diurnal analyses of observed and predicted hourly $PM_{2.5}$ total mass shown in Fig. 11 for North America for 2021/22, in Fig. S167 for four seasons, and in Fig. S168 for four sub-continental regions, all suggest that measured diurnal variations were smaller than the predicted variations. Two contributing factors to this difference might be the diurnal allocation of primary $PM_{2.5}$ emissions used by the RAQDPS-OP023 and the parameterization of $PM_{2.5}$ chemical volatility, including ammonium, nitrate, and water components. Interestingly, Fig. S167 shows that the observed diurnal variation was largest in the winter and smallest in the summer, which is consistent with Fig. 17. In addition, both the observed and predicted all-station, seasonal-mean diurnal curves of $PM_{2.5}$ total mass in Fig. S167 have one peak at about the time of morning rush hour and sunrise and a second peak at about the time of evening rush hour and sunset. By contrast the seasonal continent-wide diurnal time series in Fig. 17 only have one peak, in the morning near sunrise, but the majority of grid cells sampled for this figure will contain little vehicular activity, unlike the urban areas in which many monitors are located.

Additional analyses

More evaluation results for the daily $PM_{2.5}$ speciation measurements and gravimetric $PM_{2.5}$ total mass measurements for 2013–2016 can be found in Sect. S3.3.2. These results include tables of annual and seasonal scores for the individual CSN, IMPROVE, and NAPS networks as well as regional scores, spatial plots of annual MB, NMB, CRMSE, and R station scores for each $PM_{2.5}$ chemical component, spatial plots of predicted seasonal mean $PM_{2.5}$ component concentration fields (cf. Fig. 14), monthly time series of $PM_{2.5}$ component statistics, monthly density

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scatterplots, and additional stacked bar graphs stratified by network and by region. The discussion of Table 5 noted consistent annual underpredictions or overpredictions for some of the seven $PM_{2.5}$ chemical components for the 2013–2016 hindcasts. Similarly consistent biases were found throughout the year by season or month for the combined networks for three $PM_{2.5}$ components, namely underpredictions for $PM_{2.5}$-$SO_4$ (Fig. S151) and overpredictions for $PM_{2.5}$-EC (Fig. S154) and $PM_{2.5}$-SS (Fig. S157). In addition, $PM_{2.5}$-TOM, the component found to have the smallest annual ~~bias~~NMB values, was shown to have pronounced seasonal biases consistent with Fig. 15, with marked overpredictions in winter and underpredictions in summer (Fig. S155). Nevertheless, performance benchmarks for five $PM_{2.5}$ chemical components ($SO_4$, $NO_3$, $NH_4$, EC, TOM) proposed by Emery et al. (2017) were also compared with both annual and seasonal scores. Nearly all $PM_{2.5}$-$SO_4$, $NO_3$, $NH_4$, and TOM annual NMB scores and many seasonal scores met acceptable NMB benchmarks, all $PM_{2.5}$-$SO_4$ and $NO_3$ annual and most seasonal NMAE scores met acceptable NMAE benchmarks, and all $PM_{2.5}$-$SO_4$, $NO_3$, $NH_4$, and EC annual and seasonal R scores met acceptable R benchmarks. Some annual and seasonal biases were also shown to be network-dependent. Annual NMB values for 2013–2016 for $PM_{2.5}$-TOM ranged from 0.20 to 0.31 for CSN and from 0.52 to 0.80 for NAPS vs. -0.46 to -0.30 for IMPROVE (Table S5A). For $PM_{2.5}$-CM the network-dependent biases were even more pronounced: annual NMB values ranged from 1.23 to 1.59 for CSN and from 1.96 to 2.34 for NAPS vs. -0.59 to -0.50 for IMPROVE. Seasonal $PM_{2.5}$-$SO_4$ was underpredicted and $PM_{2.5}$-SS was overpredicted throughout the year for all three speciation networks, but $PM_{2.5}$-EC was overpredicted for the urban-focused CSN and NAPS networks in all seasons but not for the rural-focused IMPROVE network (Table S5S). Since ~~Similarly, annual NMB values for $PM_{2.5}$ TOM ranged from 0.20 to 0.31 for CSN and from 0.52 to 0.80 for NAPS vs. -0.46 to -0.30 for IMPROVE (Table S5A). And for $PM_{2.5}$ CM the network-dependent biases were even more pronounced: annual NMB values ranged from 1.23 to 1.59 for CSN and from 1.96 to 2.34 for NAPS vs. -0.59 to -0.50 for IMPROVE.~~ $PM_{2.5}$-EC, POM, and CM are all primary pollutants, ~~suggesting that~~ some of these model biases ~~may~~ might be connected to the representation of primary $PM_{2.5}$ emissions.

The systematic biases by network that were identified for some $PM_{2.5}$ chemical components also affect predictions of $PM_{2.5}$ composition and total mass. The discussion of Fig. 15 noted that the best agreement was for spring and autumn. However, when the same analysis was applied to only CSN measurements and to only IMPROVE measurements, the agreement was less good for these two seasons, with the ~~model~~ RAQDPS-OP023 overpredicting $PM_{2.5}$ total mass for CSN for spring and autumn (Fig. S200) and underpredicting $PM_{2.5}$ total mass for IMPROVE (Fig. S201). The findings were similar for a regional analysis, where predicted $PM_{2.5}$ composition and total mass were in very good agreement with the combined measurements for the WUS and EUS (Fig. S205), but the same regional analysis for CSN-only measurements (Fig. S206) and IMPROVE-only measurements (Fig. S207) revealed errors of opposite sign, with overpredictions of $PM_{2.5}$ total mass for CSN for the WUS and underpredictions of $PM_{2.5}$ total mass for IMPROVE for both the WUS and EUS. The presence of these compensating errors for urban-focused and rural-focused stations again points to the benefits of performing a more disaggregated analysis. These network-dependent differences are also consistent ~~y also agree~~ with the finding in Sect. 3.2.2 that underprediction of $PM_{2.5}$ total mass by the RAQDPS-OP023 ~~underprediction~~ appears to be mainly a rural issue. Predicted peak seasonal concentrations for $PM_{2.5}$-$NO_3$, $NH_4$, POM, and SS occurred in the winter (Figs. S23, S24, S26, and S30) vs. summer peaks for $PM_{2.5}$-$SO_4$, SOM, and TOM (Figs.

S22, S27, and S28). This complementarity seasonal phase shift can help to explain the observed bimodality that was observed in monthly $PM_{2.5}$ total mass (Figs. 8 and S198).

Additional cClose links wearе also evident between the inorganic $PM_{2.5}$ components and their gas-phase precursors. For example, the monthly mean concentration time series for $PM_{2.5}$-$SO_4$ (Fig. S151) was seasonally anticorrelated with and the monthly time series for its gas-phase precursor $SO_2$ (Fig. -S146), which were seasonally anticorrelated:, that is, the -$PM_{2.5}$-$SO_4$ time series had an observed and predicted summer maximum and winter minimum and monthly NMB values werеas underpredictednegative, while the $SO_2$, its precursor, time series had an observed and predicted winter maximum and summer minimum and monthly NMB values werеas overpredictedpositive. The monthly mean concentration time series for A similar anticorrelation was evident for $PM_{2.5}$-$NO_3$ (Fig. S152) was seasonally anti-correlated with that for and its gas-phase precursor $HNO_3$ (Fig. S144) (and also $PM_{2.5}$-$SO_4$), which were also closely related; that is, the $PM_{2.5}$-$NO_3$ time series had an observed and predicted summer minimum and winter maximum and was mostly underpredicted while the $HNO_3$ time series had an observed and predicted winter minimum and summer maximum and was overpredicted. The monthly mean concentration time series for $PM_{2.5}$-$NH_4$, the third inorganic component (Fig. S153), was also seasonally anticorrelated with that for its gas-phase precursor $NH_3$ (Fig. S145) (and also $HNO_3$): that is, the $PM_{2.5}$-$NH_4$ time series had an observed and predicted summer minimum and winter maximum and was mostly overpredicted while the $NH_3$ time series had an observed and predicted winter minimum and summer maximum and was underpredicted. Lastly, downward time trends for 2013-2016 were also visible in observed monthly surface concentrations of $PM_{2.5}$-$SO_4$ and $NH_4$ (Figs. S151, S153) due to the $SO_2$ and $NO_x$ emissions decreases that took place over this period.

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### 3.3.3 Precipitation chemistry

Precipitation-chemistry measurements are valuable for model evaluation because they are quite different from the air-chemistry measurements considered in the previous sections. One difference is that they quantify a removal process, wet deposition, as opposed to near-surface ambient concentrations determined by dispersion and chemistry. In addition, they represent removal through a column extending from the Earth's surface to cloud top rather than a pure surface-level quantity. They are also an integrated measurement in terms of both gas-particle partitioning and aerosol particle size distribution because they can include contributions from both gas and particle phases and from all aerosol particle sizes. Precipitation-chemistry measurements can thus provide some important insights into removal processes and atmospheric mass budgets for sulfur, oxidized nitrogen, reduced nitrogen, and base cations removal processes and atmospheric mass budgets. As described in Sect. 2.3, two large national networks are responsible for North American precipitation-chemistry measurements, CAPMoN in Canada and NADP in the U.S.

Figure 18 shows the RAQDPS-OP023-predicted spatial distributions of annual mean concentration in precipitation over North America of the three major inorganic ions, $SO_4^=$, $NO_3^-$, and $NH_4^+$, for 2013–2016 and 2021/22. Several features are evident. First, the spatial distributions are different for each of these ions, largely in response to the different spatial distributions of their respective precursor emissions (cf. Fig. S1). Annual mean $SO_4^=$ concentration in

precipitation tends to be larger over the eastern U.S., whereas annual mean $NO_3^-$ and $NH_4^+$ concentrations in precipitation are larger over the western and central U.S. Another feature is that predicted $SO_4^=$ and $NO_3^-$ annual mean concentrations in precipitation appear to decrease from 2013 to 2021/22~~16~~, consistent with the 6037% and 3618% decreases in North American $SO_2$ and $NO_x$ anthropogenic emissions, respectively, over this period (Table 1). No time trend is evident for this period, however, for annual mean $NH_4^+$ concentration in precipitation, which differs from the downward trend for annual mean $PM_{2.5}$-$NH_4$ concentration visible in Table 5 and Fig. 14. Two reasons are that annual $NH_3$ emissions only decreased slightly (67%) from 2013 to 2021/22~~16~~ and both $NH_3$ and PM-$NH_4$ are removed by precipitation scavenging regardless of the exact balance for gas-particle partitioning, ~~wh~~whereas annual mean $NH_3$ VMR displayed a slight upward trend from 2013 to 2021/22~~16~~ (cf. Table 4, Fig. S16). Figure 18 also agrees qualitatively with comparable plots of observed annual mean $SO_4^=$, $NO_3^-$, and $NH_4^+$ concentrations in precipitation for these four years produced by the U.S. NADP network (e.g., National Acid Deposition Program, 2014, 2017), except for the elevated values of annual mean $NO_3^-$ and $NH_4^+$ concentrations in precipitation predicted over California in 2013. Note ~~too~~ that the corresponding annual wet deposition fields predicted by the RAQDPS-OP023 for 2014–2016 (see Figs. ~~-~~S34–S36) have also been used in a study by Cathcart et al. (2025) to calculate total deposition and acidity and nutrient nitrogen critical-load exceedance fields for Canada.

Table 6 lists all-station annual evaluation statistics for 2013–2016 for RAQDPS-OP023-predicted weekly concentrations in precipitation of $SO_4^=$, $NO_3^-$, and $NH_4^+$, for weekly precipitation, and for the corresponding predicted weekly wet deposition of these ions. The weekly precipitation scores were already discussed in Sect. 3.1. The reductions in concentration~~s~~ in precipitation of $SO_4^=$ and $NO_3^-$ from 2013 to 2016 suggested visually by Fig. 18 are~~is~~ confirmed by Table 6. Observed and predicted annual-~~-~~mean $SO_4^=$ weekly concentration in precipitation for the combined networks decreased from 0.81 to 0.62 $mg \cdot L^{-1}$ and from 0.82 to 0.61 $mg \cdot L^{-1}$, respectively, over this period, while observed and predicted annual-~~-~~mean $SO_4^=$ weekly wet deposition decreased from 15.6 to 10.7 $mg \cdot m^{-2} \cdot week^{-1}$ and from 15.9 to 10.1 $mg \cdot m^{-2} \cdot week^{-1}$, respectively. Observed and ~~measured~~ predicted annual-~~-~~mean $NO_3^-$ weekly concentration in precipitation and weekly wet deposition also declined from 2013 to 2016, though by smaller percentages, from 0.96 to 0.90 $mg \cdot L^{-1}$ and 0.94 to 0.78 $mg \cdot L^{-1}$ and from 16.5 to 14.1 $mg \cdot m^{-2} \cdot week^{-1}$ and 16.2 to 11.6 $mg \cdot m^{-2} \cdot week^{-1}$, respectively. By contrast, observed and predicted annual-~~-~~mean $NH_4^+$ weekly concentration in precipitation increased slightly while observed and predicted annual-~~-~~mean $NH_4^+$ weekly wet deposition decreased slightly.

Table 6 also reveals that model performance can vary considerably by ionic species. For example, all-station annual NMB values for $SO_4^=$ and $NO_3^-$ weekly concentrations in precipitation ranged from -0.01 to 0.14 and from -0.14 to ~~-~~0.01, respectively, but corresponding all-station annual NMB values for $NH_4^+$ weekly concentration in precipitation ranged from 0.23 to 0.41, pointing to a considerable and consistent overprediction for this quantity. All-station annual NMB scores for $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly wet deposition ranged from -0.06 to 0.05, -0.18 to -0.02, and 0.13 to 0.16, respectively, suggesting small biases for $SO_4^=$ wet deposition, a small but consistent underprediction for $NO_3^-$ wet deposition, and a consistent overprediction for $NH_4^+$ wet deposition. In comparison, Simon et al. (2012) reported

(25~~10~~%, 75~~90~~%) quantile values for NMB scores for accumulated seasonal and annual $SO_4^=$, $NO_3^-$, and $NH_4^+$ wet deposition of (-0.01~~9~~, 0.23~~38~~), (-0.37~~45~~, 0.10~~9~~), and (-0.21~~33~~, 0.0~~28~~), respectively, in a review of North American AQ model performance evaluations. The RAQDPS-OP023 annual NMB scores fit comfortably with~~in~~ these reported ranges ~~despite their much lower level of temporal aggregation (i.e., weekly vs. annual)~~. Zhang et al. (2018c) reported ~~large negative annual~~mean NMB values for $SO_4^=$, $NO_3^-$, and $NH_4^+$ accumulated annual wet deposition of -0.05, -0.32, and -0.31, respectively, for the 1990–2010 period using the CMAQ model without post-processing adjustments, and Benish et al. (2022) obtained ~~annual~~ mean NMB values for $SO_4^=$, $NO_3^-$, and $NH_4^+$ accumulated annual wet deposition of -0.12, -0.10, and -0.20 for the 2002–2017 period using a newer version of the CMAQ model ~~and~~ but also without post-processing adjustments. Note that the negative bias in $NO_3^-$ concentration in precipitation and wet deposition is consistent with the neglect of lightning NO emissions in both ~~models~~ CMAQ and the RAQDPS-OP023 (Zhang et al., 2018c).

Corresponding all-station annual NMAE values from Table 6 for $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly concentrations in precipitation for 2013–2016 ranged from 0.59 to 0.68, 0.52 to 0.57, and 0.78 to 0.95, respectively, and annual NMAE values for $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly wet deposition ranged from 0.61 to 0.65, 0.52 to 0.55, and 0.68 to 0.72. The NMAE scores for weekly wet deposition compare favourably with top-quartile NMAE scores for ~~accumulated seasonal and annual~~monthly $SO_4^=$, $NO_3^-$, and $NH_4^+$ wet deposition of 0.61, 0.51, and 0.57 compiled by Simon et al. (2012) despite their lower level of temporal aggregation. All-station annual FAC2 values for weekly concentration in precipitation were slightly higher (better) for $NO_3^-$ (0.68–0.70) than for $SO_4^=$ or $NH_4^+$ (0.63–0.65; 0.57–0.60), and the same was true for weekly wet deposition (0.62–0.64 vs. 0.53–0.56 and 0.55–0.57). All-station annual R scores for all stations were also slightly higher for $NO_3^-$ weekly concentration in precipitation (0.46-0.52) than for $SO_4^=$ or $NH_4^+$ weekly concentration in precipitation (0.25–0.41, 0.38–0.47), whereas all-station annual R scores for $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly wet deposition were comparable (0.49–0.58, 0.46–0.59, 0.47–0.59). By comparison, Zhang et al. (2018c) obtained higher ~~annual~~ mean R scores for accumulated $SO_4^=$, $NO_3^-$, and $NH_4^+$ annual wet deposition for the 1990–2010 period of 0.92, 0.89, and 0.77, respectively, and Benish et al. (2022) obtained ~~annual~~ mean R scores for accumulated $SO_4^=$, $NO_3^-$, and $NH_4^+$ annual wet deposition for the 2002–2017 period of 0.7~~8~~8, 0.77~~88~~, and 0.61~~78~~, but these higher scores may again be explained at least in part by the much greater temporal aggregation and averaging. Note also that the values of the statistics in Table 6 were fairly constant across the four years in spite of the large $SO_2$ and $NO_x$ emission reductions, again suggesting that the year-specific input emissions that were used for the retrospective simulations were representative of each year.

Additional analyses

More evaluation results for predicted $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly concentrations in precipitation and weekly wet deposition can be found in Sect. S3.3.3. These include tables of separate annual and seasonal scores for the CAPMoN and NADP networks for 2013–2016 as well as regional scores, spatial plots of annual MB, NMB, CRMSE, and R scores at individual stations, spatial plots of predicted seasonal concentration in precipitation and wet deposition fields for 2013–2016 and 2021/22, monthly time series of statistics for weekly concentrations in precipitation and wet

deposition, and monthly density scatterplots. A number of additional insights can be found in these supplemental analyses. For example, the separate annual and seasonal scores for the CAPMoN and NADP networks presented in Tables S6A and S6S revealed ~~some differences that tended to favour~~that RAQDPS-OP023 scores were better overall for CAPMoN than NADP. For example, annual R scores were consistently higher for CAPMoN than NADP for $SO_4^=$, $NO_3^-$, and $NH_4^+$ weekly concentrations in precipitation and weekly wet deposition. Some regional differences were also evident in many of the spatial distributions of station-specific annual values of MB, NMB, CRMSE, and R. For example, the station-level annual MB scores for $SO_4^=$ weekly concentration in precipitation tended to be positive in eastern Canada and the northeastern U.S. but negative elsewhere (Fig. S109). Annual R values for $SO_4^=$ and $NO_3^-$ weekly wet deposition also tended to be lower in the west and higher in the east (Figs. S128, ~~and~~ S132). It is also clear from Table S6S and Figs. S158–S16~~43~~ that seasonal model skill was consistent overall from year to year but sometimes varied markedly between seasons and species. For example, $NH_4^+$ weekly wet deposition had the largest variations in observed and predicted seasonal mean values of the three major ions, with maximum seasonal values ~~two to three~~about four times higher than minimum seasonal values (Fig. S164). Scores for $SO_4^=$ weekly concentration in precipitation and weekly wet deposition were worse overall in the winter (Figs. S158, S162) whereas scores for $NO_3^-$ and $NH_4^+$ weekly concentration in precipitation and weekly wet deposition were worse overall in the summer (Figs. S159–S160, S163–S164). Predicted monthly mean concentrations in precipitation for $SO_4^=$, $NO_3$, and $NH_4^+$ all had both negative and positive biases, but the majority of monthly NMB values for $SO_4^=$ and $NO_3^-$ weekly concentrations in precipitation fell in the ±0.3~~2~~0 range for 2013–2016 (Figs. S158–S159), whereas monthly NMB values for weekly $NH_4^+$ concentration in precipitation had peak values from 0.70 to 1.70 in July (Fig. S160). M~~L~~~~ower m~~onthly NMB values ~~were found~~ for weekly wet deposition were comparable. The predicted seasonal ~~al mean~~ precipitation fields were shown to link ~~but geographically shift~~ seasonal concentration in precipitation fields and the seasonal wet deposition fields but to shift them geographically. Lastly, observed and predicted monthly mean values of $SO_4^=$ weekly concentration in precipitation and $SO_4^=$ weekly wet deposition all declined from 2013 to 2016 (Figs. S158, ~~and~~ S162), providing observational support together with Figs. S146 and S151 for the representativeness of the year-specifi~~ed~~ decreases in $SO_2$ annual emissions used for the hindcast annual simulations.

# 4 Discussion

## 4.1 Comparison with RAQDPS forecast performance for 2010–2019

The RAQDPS operational AQ forecast system underwent 22 upgrades between 2010 and 2021 (Moran et al., 2026~~5~~). These included eight major upgrades: four that implemented significant changes to the modelling system code and configuration, two that involved major changes to the input emissions files, and two that included changes to both (Table A3). RAQDPS operational performance was routinely monitored over this period using hourly surface measurements of a small number of chemical species, in particular $NO_2$, $O_3$, and $PM_{2.5}$, from North American air-chemistry stations that report their measurements in near real time to AirNow or NAPS (Sect. 2.3). In order to assess the impact of this succession of system upgrades on forecast performance, Moran et al. (2021a) used hourly operational forecast outputs and NRT hourly AQ measurements archived at CMC to evaluate 9.5 years of RAQDPS operational

forecasts of hourly surface $NO_2$, $O_3$, and $PM_{2.5}$, starting from RAQDPS-OP001 forecasts at the beginning of 2010 through to July 2019, when the RAQDPS-OP021 became operational (and before ~~2020~~ COVID-19 pandemic impacts in 2020: ~~see~~ e.g., Mashayekhi et al., 2021). This analysis aligned with suggestions made by Kelly et al. (2019) by providing as much consistency as possible in the calculation of evaluation statistics, in the model domain and grid resolution, and in the weighting of different time periods.

While some data filtering is performed routinely on the NRT measurements received at CMC before they are used (Sect. 2.3), additional filtering was applied to this 2010–2019 AQ measurement data set (Moran et al., 2021a). More stringent lower and upper cutoff thresholds of 0 and 150 ppbv, 0 and 150 ppbv, and 0 and 200 $\mu g \cdot m^{-3}$ were imposed on the $NO_2$, $O_3$, and $PM_{2.5}$ hourly observations, respectively. This additional filtering removed less than 0.1% of the available hourly measurements but avoided the impacts on the evaluation statistics of some extreme outliers. Measurement-model pairing was then performed for this 9.5-year period (cf. Sect. 2.4) followed by the calculation of seasonal performance statistics. Seasonal performance statistics for the 2013–2016 annual hindcast runs and 2021/22 forecast runs made with the RAQDPS-OP023 (Table S3S) could then be compared with these earlier results to examine the impact of using the new version of the forecast system along with year-specific emissions for 2013–2016 and projected emissions for 2021/22.

Figure 19 shows decade-long time series of all-station seasonal R scores for hourly $NO_2$, $O_3$, and $PM_{2.5}$ abundances for the sequence of operational systems from the RAQDPS-OP001 to the RAQDPS-OP020 that were in use for the 2010–2019 period. The time series of 2010–2019 seasonal R scores for $NO_2$ and $O_3$ shown in this figure are broadly comparable: both time series fall in a numerical band from 0.50 to 0.70, both exhibit cyclical seasonal variations, and both exhibit some overall improvement with time. These positive trends in R scores were anticipated since proposed RAQDPS upgrades are only accepted for operational implementation at CMC if they can demonstrate at least equal or better expected forecast skill relative to the existing operational version (Moran et al., 2026~~5~~). There is also a suggestion of some anticorrelation to the seasonal R scores for $NO_2$ and $O_3$, with R scores for $NO_2$ tending to be highest in the winter and lowest in the summer whereas R scores for $O_3$ tended to be highest in the summer but lowest in the spring. The 2010–2019 seasonal R scores for $PM_{2.5}$, on the other hand, are lower than those for $NO_2$ and $O_3$, are less cyclical seasonally, and do not show any improvement with time.

Figure 19 also shows seasonal R scores for the five RAQDPS-OP023 annual simulations for 2013–2016 and 2021/22 that are the focus of this paper. The RAQDPS-OP023 seasonal R scores for 2013–2016 a~~w~~ere higher for both $NO_2$ and $O_3$ but lower for $PM_{2.5}$ compared to the previous operational RAQDPS versions. The RAQDPS-OP023 seasonal scores for 2021/22 forecasts a~~w~~ere slightly lower for $NO_2$, comparable for $O_3$, and higher for $PM_{2.5}$ compared to the hindcast scores. These differences could be due either to changes to the treatments of atmospheric chemistry in GEM-MACH or to the use of more representative emissions in the RAQDPS-OP023 simulations, but of course neither change was intended to produce poorer $PM_{2.5}$ forecasts. Each panel of Fig. 19 also shows the species-specific values of "acceptable" and "good" benchmarks for R scores recommended by Emery et al. (2017) and Zhai et al. (2024). Seasonal R scores for $NO_2$ surpassed the more stringent "good" benchmark of 0.60 for later operational versions of the

RAQDPS, seasonal R scores for $O_3$ for the RAQDPS-OP023 surpassed the "acceptable" threshold of 0.50 but fe~~a~~ll below the more stringent benchmark of 0.75, ~~but~~ and seasonal R scores for $PM_{2.5}$ did not meet even the less stringent "acceptable" threshold of 0.40 for most forecast system versions or time periods.

Seasonal time series plots of four other statistics in the same format as Fig. 19 are shown in Figs. S215 to S219 for $NO_2$, $O_3$, and $PM_{2.5}$. It was found that RAQDPS-OP023 seasonal MB and NMB scores for 2013–2016 were better than the historical operational scores for $NO_2$, slightly worse for $O_3$, and worse for $PM_{2.5}$, seasonal NMAE scores were better for $NO_2$ and $O_3$ and neutral for $PM_{2.5}$, seasonal RMSE scores were better for $NO_2$ and $O_3$ and worse for $PM_{2.5}$, and seasonal FAC2 scores were better for $NO_2$, neutral for $O_3$, and ~~worse~~ better for $PM_{2.5}$. The RAQDPS-OP023 seasonal MB and NMB scores for 2021/22 compared to 2013–2016 were worse for $NO_2$ and $PM_{2.5}$ but slightly better for $O_3$, NMAE scores were slightly better for all three species, RMSE scores were comparable for all three species, and FAC2 scores were worse for $NO_2$ ~~and comparable~~better for $O_3$ and slightly worse for $PM_{2.5}$.

### 4.2 Impact of biomass burning emissions

As noted in Sect. 2.2, near-real-time BB emissions were included in the 2021/22 RAQDPS-FW023 forecasts but not in the 2021/22 RAQDPS-OP023 forecasts or the four annual RAQDPS-OP023 hindcasts considered in this paper. Such episodic but often large emissions can have a significant impact on both local and regional air quality (e.g., Jaffe et al., 2004; Liu et al., 2015; Rappold et al., 2017; Hand et al., 2024), and discussions in Sect. 3 of ~~a number of figures (~~Figs. 4, 5, 8, and 12~~)~~ noted a drop in RAQDPS-OP023 skill for predicting $PM_{2.5}$ concentration in the summer, the time of year when BB emissions are generally highest (e.g., Munoz-Alpizar et al., 2017; Chen et al., 2019). Table A4 summarizes annual wildfire statistics for Canada and the U.S. for the 2013–2022 period. Large year-to-year variations can be seen for both countries in the number of wildfires and the land area burned each year. At the continental scale, 2015, 2017, and 2021 stand out as high wildfire years in terms of land area burned whereas 2016, 2019, and 2020 were lower wildfire years. The 2021/22 RAQDPS-OP023 forecasts thus corresponded to a high BB emissions year in 2021, in fact, a record-breaking year (Jain et al., 2024). Note, however, that years with high levels of land area burned in the U.S. (e.g., 2015, 2017, 2018, 2020) may have a greater impact in terms of overall North American population exposure to wildfire smoke since many Canadian wildfires occur far from population centres. Note ~~too~~ that summer 2015 ~~was identified as an outlier~~stands out in Fig. S201 ~~the discussion~~as an outlier ~~of Fig. S201 in Sect. S3.3.2~~ due to the elevated seasonal-mean $PM_{2.5}$ ~~EC and~~ TOM concentrations that were observed at IMPROVE stations, most of which are located in the western U.S. (Fig. S6b).

To assess the impact of the inclusion of BB emissions on forecast scores we can compare the RAQDPS-OP023 and RAQDPS-FW023 forecasts for the 12-month 2021/22 period. The only difference between these two forecast system versions was the inclusion of BB emissions in the RAQDPS-FW023 runs but not the RAQDPS-OP023 runs. Table 7 compares seasonal evaluation statistics for hourly $NO_2$, $O_3$, and $PM_{2.5}$ forecasts for the two versions. For $NO_2$ only the summer statistics were slightly different (e.g., MB, FAC2). For $O_3$ there were slight differences in a few scores for the winter and spring and larger differences in the summer and autumn, including a reduction in seasonal MB in the

summer of 0.8 ppbv for the RAQDPS-FW023. Differences in $O_3$ seasonal CRMSE, FAC2, and R scores in the summer and autumn were also small improvements. And for $PM_{2.5}$ there were larger differences in all four seasons, but especially in the summer and autumn. The summer mean $PM_{2.5}$ concentration predicted by the RAQDPS-FW023 was slightly more than double that of the RAQDPS-OP023, which improved the summer NMB score from -0.51 to 0.04. Summer NMAE, FAC2, and R scores were also considerably better for the wildfire version (e.g., R increased from 0.09 to 0.55), although the RMSE and CRMSE values for the wildfire version were worse and the NSD value ~~is now~~was too high (1.87) vs. too low (0.35). In addition, for the $PM_{2.5}$ benchmarks recommended by Emery et al. (2017) for acceptable performance (NMB=±0.30, NMAE=0.50, R=0.40), seasonal scores for the RAQDPS-OP023 did not meet the thresholds for summer NMB ~~o~~and~~r~~ R ~~and~~ scores or autumn R but the RAQDPS-FW023 scores did meet these thresholds. These large improvements in ~~most~~ many evaluation metrics suggest that BB ~~is~~ emissions are an emissions source for ~~that is important for~~ part of the year that should be considered, but at the same time there is a deterioration in a few scores, suggesting that there is room to improve the estimation of ~~wildfire~~ BB emissions in the RAQDPS-FW023.

Some additional analyses to understand the impact of BB emissions on 2021/22 RAQDPS-FW023 forecasts are presented in Sect. S4.2. These include plots of spatial distributions of annual and seasonal mean $NO_2$, $O_3$, and $PM_{2.5}$ abundance fields and station-specific annual statistics, monthly time series of $\overline{O}$, median O, $\overline{M}$, median M, MB, NMB, CRMSE, FAC2, R, and NSD scores, and diurnal time series of five statistics by season and by region. These analyses confirmed that the inclusion of BB emissions improved some summer and~~,~~ autumn~~, and also annual~~ scores for $PM_{2.5}$ in 2021/22 (e.g., Fig. S227 vs. Fig. 8) along with minor improvements for $O_3$ scores and no impact for $NO_2$ scores. Winter and spring scores, on the other hand, were virtually unchanged, which meant that $PM_{2.5}$ mass was underpredicted in these seasons by both system versions for other reasons.~~-~~ Interestingly, annual-mean $PM_{2.5}$ scores also showed significant improvement when BB emissions were included, even though the BB emissions were strongly seasonal (e.g., Fig. S230 vs. Fig. 11). In addition, Fig. S237 showed that the 2021 wildfire season was an outlier even relative to 2015. Monthly time series of predicted median $PM_{2.5}$ suggested that anthropogenic $PM_{2.5}$ emissions may have been too low in the cold-season months (Fig. S234). And regional analyses suggested that wildfire smoke affected all of North America in 2021/22 (Fig. S231 vs. Fig. 12).

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### 4.3 Comparison with other AQ forecast systems

It is also of interest to compare RAQDPS-OP023 and RAQDPS-FW023 forecast performance with that of peer AQ forecast systems operated by other agencies since such peer systems face similar constraints and limitations, including lack of access to year-specific input emissions and to meteorological and/or chemical data assimilation. To perform such a comparison, however, Simon et al. (2012) noted the challenge posed by the use of different measurement networks, ~~evaluation metrics,~~ sampling periods, sampling durations,~~-~~ evaluation metrics, ~~measurement networks,~~ and spatial domains in publications by different forecasting teams.

One such peer AQ forecast system is the U.S. National Air Quality Forecast Capability (NAQFC), which went operational in 2004 followed by many upgrades (see https://www.emc.ncep.noaa.gov/mmb/aq/AQChangelog.html,~~;~~ last access:~~ed~~ 6 ~~Aug.~~April 202~~5~~6). A number of papers that discuss NAQFC performance evaluations have been published, including Eder et al. (2006, 2009) Saylor and Stein (2012), Chai et al. (2013), Pan et al. (2014), Huang et al. (2017), Lee et al. (2017), Chen et al. (2021), Campbell et al. (2022), and Li et al. (2025). As suggested by Simon et al. (2012), only a limited number of evaluation results for surface $PM_{2.5}$ predictions from these papers, including those summarized in Table A5, ~~can~~ could be used to compare NAQFC and RAQDPS performance over the parallel evolution of these two North American regional operational forecast systems. The overall conclusion based on these limited ~~published~~ comparisons for 2010, 2014, and 2015 is that performance was comparable for the two systems over the past decade (see Sect. S4.3 for details).

An ongoing co-operative North American AQ forecast intercomparison being conducted under the WMO GAFIS initiative (see Sect. S4.2) has allowed~~s~~ direct comparisons of more recent NAQFC and RAQDPS023 quarterly performance ~~to be compared directly~~ for 2021/22 ~~(~~and subsequent years~~)~~. The 2022 second-quarter GAFIS report (Manseau et al., 2022) compared monthly MFB (see Table A2 and Sect. S2.5), FAC2, and R scores for daily maximum forecasts of $NO_2$, $O_3$, and $PM_{2.5}$ abundances for three North American regional AQ forecast systems, the RAQDPS-OP023, RAQDPS-FW023, and NAQFC, and three global AQ forecast systems for the 12-month 2021/22 forecast period, with a focus on spring 2022. For the monthly~~—~~mean diurnal time series of daily maximum surface concentrations of $NO_2$, $O_3$, and $PM_{2.5}$ presented in this report, there were marked variations evident between the six ~~different~~ AQ forecast systems. There was also considerable variation between the systems in monthly combined performance scores. Although no one system dominated, the NAQFC tended to be the top performer for $O_3$ forecasts, the RAQDPS-OP023 and RAQDPS-FW023 for $NO_2$ forecasts, and IFS-CAMS for $PM_{2.5}$ forecasts. T~~, and t~~he combined performance scores also tended to be highest overall for $O_3$ forecasts and lowest for $PM_{2.5}$ forecasts. The monthly R scores can also be compared to available benchmarks. Most of the $NO_2$ R scores for five of the systems ($NO_2$ forecasts were not available for NAQFC) ~~were above~~met the acceptable benchmark of 0.50 recommended by Zhai et al. (2024) and some RAQDPS-OP023, RAQDPS-FW023, and IFS-SILAM scores also ~~exceeded~~ met their benchmark goal of 0.60. R scores for $O_3$ for all six systems ~~exceeded~~ met the acceptable benchmark of 0.50 recommended by Emery et al. (2017) and in a few cases also ~~exceeded~~ met their benchmark goal of 0.75. And for $PM_{2.5}$ R scores none of the systems ~~attained~~ met the acceptable benchmark of 0.60 recommended by Huang et al. (2021), but in July and August 2021, when many wildfires were burning, the lower benchmark of 0.40 recommended by Emery et al. (2017) was ~~reached~~ met by most of the models, including the RAQDPS-FW023 but not the RAQDPS-OP023. Section S4.3 also describes an ongoing evaluation and intercomparison of 11 operational regional AQ forecast models for Europe that is similar to the GAFIS regional intercomparison for North America and ~~compares~~ a few evaluation scores from this European intercomparison are compared with ~~the~~ RAQDPS-OP023 scores.

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## 4.4 Forecast system shortcomings, opportunities, and priorities for further development

Many of the evaluation results for RAQDPS-OP023 forecasts and hindcasts versus AQ measurements, previous system versions, and peer forecast systems discussed in previous sections were positive. For example, Figs. 6 and 7 for the 2021/2022 forecasts ~~(~~and Figs. S140 and S141 for the 2013–2016 hindcasts~~)~~ show model skill for predicting~~ed~~ $NO_2$ and $O_3$ surface VMR monthly means at the continental scale as does Fig. 12 for four North American regions. Good agreement for all-station, annual-mean diurnal time series for $NO_2$ and $O_3$ is evident in Figs. 9 and 10, in Fig. S168 for four regions, and for seasonal-mean diurnal time series in Figs. S165 and S166. And Fig. 15 shows good model performance in predicting all-station, seasonal-mean $PM_{2.5}$ chemical composition and gravimetric $PM_{2.5}$ total mass. Evaluation results presented in Sect. S3 also suggested that the year-specific ~~2013–2016~~ annual emissions that were used for 2013–2016 for the RAQDPS-OP023 retrospective runs broadly represented emissions changes over that period (Table 1) given the year-ordered agreement between measurements and model predictions (e.g., Figs. S144, S146, S151, S158, and S1625~~8~~ for $HNO_3$, $SO_2$, $PM_{2.5}$-$SO_4$, $SO_4^=$ concentration in precipitation, and $SO_4^=$ ~~concentration in precipita~~wet deposition, respectively). Results presented in Sect. 4.1 showed some improvements in forecast skill over a 10-year period due to a series of operational RAQDPS upgrades, including increasing seasonal R scores for hourly $NO_2$ and $O_3$ forecasts (Fig. 19), decreasing seasonal NMAE scores for hourly $O_3$ and $PM_{2.5}$ forecasts (Fig. S217), decreasing seasonal RMSE scores for hourly $PM_{2.5}$ forecasts (Fig. S218), and increasing seasonal FAC2 scores for hourly $NO_2$ and $O_3$ forecasts (Fig. S219). Relative to suggested benchmark values, seasonal NMB scores for $NO_2$, $O_3$, and $PM_{2.5}$ for most RAQDPS operational versions met the acceptable benchmark (Fig. S216) as did seasonal R scores for $NO_2$ and $O_3$ (Fig. 19). Lastly, the ongoing GAFIS quarterly evaluation of RAQDPS-OP and RAQDPS-FW forecasts for North America along with those made by four operational AQ forecast peer models described in Sect. 4.3 found that RAQDPS-OP023 and RAQDPS-FW023 performance in 2021/22 was competitive with their peer models for surface $NO_2$ and $O_3$ VMRs, though less so for surface $PM_{2.5}$ total mass (for which all of the models performed least well).

### 4.4.1 $PM_{2.5}$ total mass and chemical composition

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It is also evident that many evaluation scores were similar for the five years simulated by the RAQDPS-OP023, and some scores point to systematic errors in the model itself rather than model inputs since the anthropogenic input emissions used were tailored to be year-specific and the similarities in results we~~a~~re present despite interannual variations in meteorology. The most concerning systematic error may be consistent underpredictions of hourly $PM_{2.5}$ surface concentrations. These underpredictions were evident visually in Figs. 1 and 5, in which observed annual and seasonal $PM_{2.5}$ mean values at station locations stood out from the predicted mean $PM_{2.5}$ concentration fields across the continent due to their higher values, and in Table 3 where all-station annual NMB values for hourly $PM_{2.5}$ concentrations were negative for all five years, ranging from -0.09 to -0.06 for 2013–2016 and -0.31 for 2021/22. However, these underpredictions also varied strongly by season and were smallest in winter and largest in summer (e.g., Fig. 8, Table S3S). In addition, ~~Figure~~ Fig. 11 showed a large underprediction of all-station annual-mean diurnal values for hourly $PM_{2.5}$ concentration at all hours but especially in early afternoon, and Fig. S167 showed

corresponding underpredictions for seasonal-mean diurnal values, with the largest differences in the early afternoon and the summer.

Some other $PM_{2.5}$ evaluation results, however, tell a more nuanced story. First, the results just summarized are largely for aggregated all-station ~~aggregated~~ statistics. When monthly statistics were calculated separately for urban stations only and for rural stations only, as shown in Figs. S213 and S214, a key difference was that hourly $PM_{2.5}$ underpredictions for 2013–2016 were limited to rural stations in all months and to urban stations in the summer. For most of the year, however, hourly $PM_{2.5}$ mass was overpredicted at urban stations. RAQDPS-OP023~~Model~~ predictions also showed better agreement with daily gravimetric $PM_{2.5}$ total mass measurements for 2013–2016 than with hourly continuous $PM_{2.5}$ measurements for those years (~~cf.~~ Fig. S198 vs. Fig. S142). Interestingly, monthly scores for the daily gravimetric $PM_{2.5}$ total mass measurements were similar to those for urban hourly $PM_{2.5}$ total mass measurements (cf. Fig. S198 vs. Fig. S142), which is consistent with the gravimetric $PM_{2.5}$ total mass measurements being urban-focused (Demerjian, 2000). Negative biases also tended to be more common and larger at western stations than eastern stations (e.g., Fig. S45, S46). Results from the evaluation of $PM_{2.5}$ chemical composition and $PM_{2.5}$ reconstructed mass for 2013–2016 add further complications. Two $PM_{2.5}$ chemical components (EC, SS) were found to be consistently overpredicted in all months (Figs. S154, S157), one component ($SO_4$) was consistently underpredicted in all months (Fig. S151), and another component (CM) had large overpredictions for most months (Fig. S156). As a consequence, predicted annual mean $PM_{2.5}$ composition was incorrect on a rank-ordered basis since annual mean $PM_{2.5}$-CM was predicted to be the second-largest $PM_{2.5}$ component after annual mean $PM_{2.5}$-TOM but was observed to be the fourth-largest annual mean $PM_{2.5}$ component after $PM_{2.5}$-TOM, $SO_4$, and $NO_3$ (Tables 5 and S5A). Despite these errors in predicting $PM_{2.5}$ composition, however, predictions of $PM_{2.5}$ reconstructed mass were surprisingly good because of partial cancellation of overpredictions and underpredictions of the mass of individual chemical components. In fact, Fig. 15 showed that observed seasonal reconstructed $PM_{2.5}$ dry mass for the combined CSN, IMPROVE, and NAPS $PM_{2.5}$ speciation networks for 2013–2016 was lower than predicted seasonal $PM_{2.5}$ dry mass for three of the four seasons in all years, summer being the exception. For the corresponding analysis based only on CSN measurements, however, observed seasonal reconstructed $PM_{2.5}$ dry mass was lower than predicted seasonal $PM_{2.5}$ dry mass for 15 out of 16 seasons (summer 2015, a high wildfire period, was the exception), with the largest model overpredictions occurring in the winter (Fig. S200). But for the same analysis based only on IMPROVE measurements, observed seasonal reconstructed $PM_{2.5}$ mass was higher than predicted seasonal $PM_{2.5}$ mass for all 16 seasons, with the largest model underpredictions occurring in the summer (Fig. S201). Since the CSN network consists largely of urban stations while the IMPROVE network consists largely of rural stations, these differences are consistent with Figs. S213 and S214.

Taken together the above findings offer five clues to~~n~~ possible improvements of~~to~~ RAQDPS-OP023 hourly $PM_{2.5}$ total mass predictions:

(i) model underpredictions were largest for the summer months and are consistent across multiple years;

(ii) underpredictions were larger in western North America than eastern North America;

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(iii) rural stations ~~underpredictions~~ were strongly associated with underpredictions ~~rural stations~~ while urban stations were often~~sometimes~~ associated with overpredictions;

(iv) predictions of some $PM_{2.5}$ chemical components had systematic biases, both negative and positive; and

(v) underpredictions were greater for the combined ~~continuous~~ hourly continuous $PM_{2.5}$ measurement networks than the combined daily gravimetric $PM_{2.5}$ measurement networks.

The RAQDPS-OP023 ~~maximum summertime~~ negative biases for $PM_{2.5}$ total mass ~~and~~ in the summer and in western North America~~bias~~, which are both consistent with the timing and location of the majority of wildfires in North America (e.g., Holden et al., 2011; Mao et al., 2011; Hand et al., 2013; Schichtel et al., 2017), strongly suggest the importance of including wildfire emissions. The comparison of RAQDPS-OP023 and RAQDPS-FW023 scores for 2021/22 discussed in Sect. 4.2 strongly supports this conclusion. However, Figs. S245 and S246 showed that while mean $PM_{2.5}$ underpredictions were much improved in the summer months at both urban and rural stations for the RAQDPS-FW023, this was not true for the rest of the year, especially at rural stations, so that other improvements to hourly $PM_{2.5}$ total mass predictions are also needed.

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It is also clear that multiple actions will be required to address the shortcomings identified for predictions of the multiple chemical components that make up $PM_{2.5}$. For example, the overprediction of monthly mean $PM_{2.5}$ total mass in urban areas vs. underprediction in rural areas suggests that the spatial allocation of anthropogenic primary $PM_{2.5}$ emissions might need to be modified to re-allocate some of these emissions from urban to rural areas. This change might also help to address persistent $PM_{2.5}$-EC overpredictions (Fig. S154), whose emissions vary directly with population (e.g., Fig. 14). Another approach to reduce urban $PM_{2.5}$ overpredictions might be to add an urban heat island mixing parameterization (e.g., Ren et al., 2020) or a SGS on-road mobile mixing parameterization (Makar et al., 2021). Figures 12 and S231 showed that $PM_{2.5}$ total mass was underpredicted in spring 2022 in the eastern U.S., where biogenic emissions are high (e.g., Fig. S21), and Fig. S207 showed that $PM_{2.5}$-TOM was underpredicted at IMPROVE measurement sites in the eastern U.S. from 2013 to 2016. These results suggest that biogenic SOA levels might be underpredicted there, so the parameterization of biogenic SOA should be reviewed to see whether its contribution to $PM_{2.5}$-TOM in rural areas might be increased (e.g., Schichtel et al., 2017; Zhang et al., 2018a). Process representations related to $PM_{2.5}$-$SO_4$ production and $PM_{2.5}$-CM emissions should also be reviewed to see whether the contributions of these processes in rural areas could be increased (Fig. S201), ~~and~~ the contribution of $PM_{2.5}$-CM emissions in urban areas reduced, and the poor monthly representation of $PM_{2.5}$-CM concentration evident in Fig. S156 addressed~~)~~. One possible way to increase rural $PM_{2.5}$-CM levels would be to add a parameterization for wind-blown dust emissions from natural sources, which is not part of the RAQDPS-OP023 system, as another source of $PM_{2.5}$ emissions (e.g., Park et al., 2010; Appel et al., 2013; Foroutan et al., 2017). Another avenue to investigate given the overpredictions of $PM_{2.5}$-CM in the winter is to review the meteorological modulation scheme that was used by the RAQDPS-OP023 to reduce fugitive dust emissions when the ground was predicted to be wet or snow-covered (Moran et al., 2026~~5~~). In addition, the $PM_{2.5}$–SS component~~, which~~ should be included in the calculation of predicted $PM_{2.5}$ total mass. This component~~,~~ was used in the calculation of predicted $PM_{2.5}$ total mass in Fig. 15, where it made a non-negligible contribution, but it was not included in the disseminated RAQDPS-OP023 operational forecasts of $PM_{2.5}$ concentration

due to the extreme $PM_{2.5}$-SS overpredictions that were encountered in the earliest RAQDPS versions. This omission should be addressed, but at the same time an effort should be made to reduce the remaining $PM_{2.5}$-SS overprediction (e.g., Table 5, Figs. S105, S157, ~~S105,~~ S186) while keeping in mind the higher uncertainties of these scores due to the use of a measured proxy species to estimate $PM_{2.5}$-SS mass (see Sect. S2.4).

The RAQDPS-OP023 predictions of hourly $PM_{2.5}$ total mass also only considered $PM_{2.5}$ dry mass. The better agreement of predicted $PM_{2.5}$ total mass with gravimetric $PM_{2.5}$ total mass measurements, which are analyzed under low-humidity laboratory conditions and only contain particle-bound water, than with non-FRM continuous $PM_{2.5}$ total mass measurements, which are made regardless of humidity levels and include both ~~include~~ semi-volatile and particle-bound water, suggests that predicted hourly $PM_{2.5}$ total mass should include an aerosol water component (e.g., Frank, 2006; Malm et al., 2011; Nguyen et al., 2016; Pye et al., 2017; Widziewicz-Rzońca and Tytła, 2020). The gravimetric $PM_{2.5}$ filter analysis can also result~~s~~ in the loss of some $PM_{2.5}$-$NO_3$ and $NH_4$ as well as some aerosol water due to volatilization (e.g., Frank, 2006; Malm et al., 2011; Chow et al., 2015; Nguyen et al., 2016; Hand et al., 2019). In addition, predicted $PM_{2.5}$-$NO_3$ is biased low when compared to laboratory-analyzed speciation measurements (e.g., Table S5S-mr, Fig. S152). The representation of inorganic heterogeneous chemistry used by the RAQDPS-OP023 (Moran et al., 2026~~5~~) should be examined critically. For example, one process missing from the RAQDPS-OP023 inorganic hetereogeneous chemistry parameterization was an explicit treatment of the role of base cations (~~e.g.,~~ Miller et al., 2024).

#### 4.4.2 Ozone

Turning to gas-phase species, o~~O~~ne shortcoming revealed by the evaluation of $O_3$ forecasts was the underprediction of the well-known Northern Hemisphere spring $O_3$ peak (e.g., Penkett and Brice, 1986; Monks, 2000; Liudchik et al., 2015). All-station spring NMB values for $O_3$ surface VMR were the lowest of the four seasons for 2013–2016 and ranged from -0.16 to -0.20 (Table S3S). The largest negative all-station monthly NMB values for $O_3$ occurred in April and May in 2013–2016 (Fig. S141), and~~.~~ ~~Other~~ Fig. S166 provides another example~~s~~ of negative $O_3$ bias in spring and early summer months. Other analyses found larger $O_3$ underpredictions at western stations ~~are evident in~~(e.g., Table S7~~ and; in~~ Figs. ~~-~~3, 12, and S42~~, and S166~~). Different explanations for these~~is~~ systematic springtime and western underpredictions are possible, but one obvious candidate for investigation is the model's treatment of $O_3$ chemical lateral boundary conditions (e.g., Pendlebury et al., 2018; Moran et al., 2026~~5~~), where $O_3$ levels at the western boundary in the spring and summer may be too low. Two more aspects of the $O_3$ predictions that could be improved are also worth noting. First, annual mean $O_3$ surface VMRs showed a latitudinal step jump over the southeastern U.S. for all five years (Fig. 13). This artefact was likely caused by the parameterization of $O_3$ dry deposition used by the RAQDPS-OP023, which considered five coarse "seasons" that were defined based only on broad latitudinal bands with no dependence on longitude or elevation (e.g., Zhang et al., 2002; Makar et al., 2018; Moran et al., 2026~~5~~). And second, Fig. S42 showed that many coastal stations along the U.S. Gulf of Mexico and Florida peninsula had positive annual NMB values for 2013–2016. These $O_3$ overpredictions at coastal stations are consistent with the ~~model's~~ lack in the

RAQDPS-OP023 of any treatment of marine halogen chemistry, which can reduce surface $O_3$ concentrations (e.g., Sarwar et al., 2015; Li et al., 2019).

### 4.4.3 $PM_{2.5}$ gas-phase precursors

The evaluation of predictions of other gas-phase species besides $NO_2$ and $O_3$ also pointed to shortcomings in modelling some $PM_{2.5}$ precursors. For example, all-station, annual-mean $SO_2$, $HNO_3$, and ISOP VMRs were consistently overpredicted and ~~annual mean~~ $NH_3$ VMRs were consistently underpredicted for 2013–2016 (Table 4). All-station seasonal-mean $SO_2$ VMRs were also overpredicted for all 16 seasons in 2013–2016 (Table S4S), which was consistent with the underprediction of all-station, seasonal-mean ~~mean~~ $PM_{2.5}$-$SO_4$ air concentrations for the same 16 seasons (Table S5S). Figure S146 showed the largest monthly $SO_2$ NMB values to be associated with the winter and summer seasons. Table S8 showed that the highest annual-mean regional $SO_2$ VMRs were predicted to occur in eastern Canada whereas measurements put the highest annual-mean regional $SO_2$ VMRs in either the eastern U.S. or western Canada. The majority of measurement stations with annual $SO_2$ overpredictions were located in eastern North America or Alberta, whereas many stations in the western U.S. exhibited underpredictions (Figs. S61, S62). Two $SO_2$ removal processes were found to be missing from the RAQDPS-OP023: the soil-wetness and cuticle-wetness gas-phase dry deposition pathways (Moran et al., 2026~~5~~). Adding treatments for these two pathways is an obvious first step to reduce $SO_2$ overpredictions. In addition, the RAQDPS dry deposition scheme was examined in detail as part of the AQMEII-4 deposition intercomparison study (Clifton et al., 2023; Hogrefe et al., 2025; Kioutsioukis et al., 2025), and the results from this study may lead to further improvements to the RAQDPS scheme. Note that other gas-phase species such as $HNO_3$ and $NH_3$, which use $SO_2$ as an archetypal species~~e~~ for modelling dry deposition (Zhang et al., 2002), would also ~~be impacted by~~have reduced levels due to this change: note that monthly mean $HNO_3$ VMR was also consistently overpredicted by the RAQDPS-OP023 for all months (Fig. S144) although monthly mean $NH_3$ VMR was ~~not~~ actually underpredicted (Fig. S145). Reduced $HNO_3$ levels, however, would reduce $NH_3$ removal to the particle phase. In addition, the facts that monthly mean $SO_2$ VMR was consistently overpredicted while monthly mean $PM_{2.5}$-$SO_4$ air concentration was consistently underpredicted (Fig. S151) suggests that predicted $SO_2$-to-$SO_4$ conversion was too low. The two chemical pathways for $SO_2$-to-$SO_4$ conversion considered by the RAQDPS-OP023 were gas-phase oxidation and aqueous-phase oxidation. It is shown in Fig. S158 that monthly NMB values for weekly $SO_4^=$ concentration in precipitation for 2013–2016 were positive for most of the year. This result makes it more likely that the RAQDPS-OP023 representation of gas-phase oxidation of $SO_2$ might~~ay~~ be the main reason for the year-round underprediction of $PM_{2.5}$-$SO_4$ air concentration.

All-station, seasonal-mean $NH_3$ VMR was underpredicted for all seasons and months in 2013–2016, with the largest seasonal NMB values occurring in the winter (Table S4S; Fig. S145). In addition, Puchalski et al. (2011) found the Radiello passive samplers used by the AMoN network, which provided most $NH_3$ measurements, to be biased low, thus suggesting an even larger model underprediction. This consistent underprediction for $NH_3$ VMR occurred at the same time as overpredictions for most of the year ~~o~~for (i) $PM_{2.5}$-$NH_4$ air concentration (Fig. S153), (ii) $NH_4^+$ concentration in precipitation (Fig. S160), and (iii) $NH_4^+$ wet deposition (Fig. S164). Table S8 showed that the highest

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annual-~~-~~mean $NH_3$ VMRs were predicted by the RAQDPS-OP023 to occur in eastern Canada, whereas the western U.S. had the highest observed annual-~~-~~mean VMRs. ~~Figure S145 shows time series of monthly mean $NH_3$ VMRs and related evaluation statistics for 2013–2016, for which monthly NMB values were negative for all months and years but were largest in the winter.~~ The accuracy of emissions is always a potential factor to explain either model over- or underpredictions, and the large negative NMB values for $NH_3$ VMR occurring ~~in the west and~~ in the winter and in the west ~~could~~ might point to an underestimate of ~~western and~~ wintertime and western $NH_3$ emissions (e.g., Momeni et al., 2025). The level of autumn $NH_3$ emissions could also be underestimated (Fig. S145). In addition, Figs. S160 and S164 show peak overpredictions of monthly $NH_4^+$ wet concentration and wet deposition in the summer at the same time as monthly ambient $NH_3$ was underpredicted (Fig. S145) but ambient $PM_{2.5}$-$NH_4$ was overpredicted (Fig. S153). This ~~could~~ might mean that too much $NH_3$ gas was being removed in the RAQDPS-OP023~~model~~ by wet deposition in the summer when inorganic aerosol thermodynamics favours ambient $NH_3$ over $PM_{2.5}$-$NH_4$ and ambient $HNO_3$ over $PM_{2.5}$-$NO_3$ (e.g., Ansari and Pandis, 1998).

Improvements to the prediction of $HNO_3$ and $PM_{2.5}$-$NO_3$ could also improve the prediction of $PM_{2.5}$-$NH_4$. For example, monthly NMB scores for $PM_{2.5}$-$NH_4$ peak in September and October (Fig. S153) when monthly NMB scores for $HNO_3$ also peak (Fig. S144), ~~while~~ whereas monthly $PM_{2.5}$-$NO_3$ was underpredicted in all months except September and October (Fig. S152), and monthly $NO_3^-$ ~~wet~~ concentration in precipitation was also overpredicted in September and October (Fig. S159). $HNO_3$ precursors $NO_2$ and NO were overpredicted in almost all months (Figs. S140 and S143), so reducing their levels by decreases to $NO_x$ emissions (if justifiable) ~~decreasing $NO_x$ emissions~~ would decrease $HNO_3$ levels. Increases to available $NH_3$ levels in the cold season through temporal reallocation of $NH_3$ emissions would result in increased $PM_{2.5}$-$NO_3$ levels and decreased $HNO_3$ levels in those months. And implementation of the missing dry deposition pathways for $SO_2$ to wet surfaces noted above would~~ill~~ also increase $HNO_3$ dry deposition . It is clear, though, from this discussion how intertwined the sulfur, oxidized nitrogen, and reduced nitrogen budgets are via emissions, chemistry, gas-phase partitioning, and wet and dry removal.

Lastly, seasonal-mean ISOP was overpredicted for all 16 seasons in the 2013–2016 period (Table S4S). Annual overpredictions also occurred at all available PAMS measurement stations (Figs. S77, S78). Figure S150 showed time series of monthly ISOP prediction errors, which included very high NMB scores, poor FAC2 values, and near-zero R scores in the cold season. Although the ISOP measurements considered here were obtained from a small number of U.S. stations, an earlier study by Stroud et al. (2008) that compared Canadian ISOP measurements against predictions by the AURAMS chemical transport model, which used the same gas-phase chemistry mechanism and a similar treatment of biogenic emissions as the RAQDPS-OP023, also found ISOP overpredictions, especially in eastern Canada. When considered together these findings suggest that further examination of the RAQDPS-OP023 biogenic emissions scheme (Moran et al., 2026~~5~~) is warranted for both magnitude and timing, including the level of isoprene emissions as well as other biogenic emitted species such as monoterpenes and sesquiterpenes and also the spatiotemporal specification of vegetation phenology.

This section has discussed possible improvements to the RAQDPS-OP023 as suggested by an assessment and analysis of the full set of evaluation results from Sects. 3 and S3. Note that it has not been possible, however, to consider all aspects of the model. For example, due to limitations on the availability of measurement data, performance scores were examined for only one meteorological quantity, precipitation, even though, as noted in Sect. 3.1, other meteorological factors such as horizontal transport and vertical mixing are also very important in AQ modelling and forecasting. The same comment applies to the vertical distributions of pollutants, which are also of interest when assessing model performance but for which relatively few studies have been published (e.g., Solazzo et al., 2013; Astitha et al., 2018) given the scarcity of available measurements in the vertical. While acknowledging these limitations, it is still possible to extend the model performance assessment through inference. For example, it was noted in the discussion of Fig. 17 in Sect. 3.3.2 that the times of day when the maxima of primary pollutants occur is determined by the balance between surface emissions and vertical stability and mixing. Figures 9, S165, S168, S228 S238, and S241 show all-station annual-mean, seasonal-mean, and annual-mean regional diurnal time series for $NO_2$ VMR. In all of these figures, the timing of the twice-daily maxima in $NO_2$ VMR is well represented by the model. This suggests that both the diurnal variations in emissions and in vertical stability and mixing have been captured well by the RAQDPS-OP023. As a second example, it was noted in Sect. 2.1 that the injection of surface emissions into the lowest layer of the atmosphere is handled differently by the RAQDPS-OP023 forecasts for 2021/22 versus the hindcasts for 2013–2016, leaving open the possibility that unrealistically high surface pollutant abundances might have been predicted in the hindcasts for primary pollutants under conditions of strong near-surface vertical stability and low wind speeds. However, examination of the five years of mid-season monthly density scatterplots for $NO_2$ in Fig. S169 show little difference in the January scatterplots for the highest $NO_2$ monthly values between January 2022 and the other four Januarys. This comparison thus provides at least an indication that unrealistically high $NO_2$ surface VMR values were very infrequent even if they did occur.

## 5 Summary and conclusions

This paper presents results from a comprehensive, five-year performance evaluation and analysis of both prospective and retrospective annual simulations made with version 023 of the ECCC Regional Air Quality Deterministic Prediction System (RAQDPS023), an operational online chemical weather forecast system for North America. A companion paper by Moran et al. (2026~~5~~) provides a comprehensive and detailed description of this forecast system. The performance evaluation described in this paper consists of three parts. In the first part, near-real-time (NRT) hourly measurements of three pollutant species, $NO_2$, $O_3$, and $PM_{2.5}$ total mass, were used to perform an operational evaluation of the first year of ~~RAQDPS023~~ forecasts from July 2021 to June 2022 made by the RAQDPS-OP023, the system version without biomass burning (BB) emissions. The Canadian Air Quality Health Index is based on these three species, so they are considered to be the key forecast species out of the dozens that were predicted by the RAQDPS-OP023. The anthropogenic input emission files used for these 2021/22 forecasts were based on a projected 2020 Canadian national emission inventory and projected 2023 U.S. and Mexican national emission inventories (i.e., emission inventories forecasted from retrospective base-year inventories). The performance of the RAQDPS-FW023,

a second version of the ECCC operational AQ forecast system that is a duplicate of the RAQDPS-OP023 except for the addition of NRT ~~biomass burning (~~BB~~)~~ emissions, was also evaluated for 2021/22 using the same measurement data set.

The 2021/22 NRT measurement data set spanned much of North America and included measurements from roughly 200 surface sites in Canada and 1100 sites in the U.S., although only one or two of the three pollutants were measured at some sites. Before being used for the evaluation, the AQ measurements underwent a two-step station-level screening process at ECCC. The first step was to apply validity checks to discard negative concentrations and above-threshold concentration values flagged as suspicious or invalid. Each valid measurement ~~(or scaled or combination of measurements for some $PM_{2.5}$ chemical components)~~ was then paired with a forecast value, after which the second screening step, period-specific completeness checking, was performed to ensure that at least a minimum number of valid hourly measurements were available ~~at a site~~for a station for the evaluation period being considered to ensure temporal representativeness. ~~No calculation was performed I~~if there were not enough valid measurements for the period, then none of the measurements for that station were included in the statistical calculations for that period. Annual, seasonal, monthly, and hour-of-day values of 10 statistical metrics were then calculated for both individual measurement networks and combined networks, for both the entire continent and four continental quadrants, and for urban sites only and rural sites only. In addition to summary tables, some of these statistical scores were presented visually in multiple ways, including site-specific "dot" statistics maps, monthly and diurnal time series, and density scatterplots,.

In the second part of the performance evaluation, an expanded and more detailed analysis was performed on four annual hindcasts for 2013–2016 made with the RAQDPS-OP0244, which was ~~(~~algorithmically equivalent to the RAQDPS-OP023 but run on a newer computer system ~~(;~~see companion paper by Moran et al., 2026~~5~~). The anthropogenic input emission files used for each of these four annual simulations were year-specific since they were generated for each year based on multi-year retrospective data sets of Canadian, U.S., and Mexican annual emission inventories. The evaluation of retrospective simulations makes it possible to access a much larger set of AQ measurement data for North America, including more AQ measurement networks (AMoN, CAPMoN, CASTNET, CSN, IMPROVE, NADP, NATTS, PAMS), more chemical species (23 vs. 3), and more measurement sites (roughly 2000 vs. 1300). These historical data sets have also undergone quality assurance and quality control checks by the individual networks before being released publicly. The consideration of multiple simulation years reduces the confounding influence of interannual meteorological variability, which helps with the identification of systematic prediction errors. However, differences between measurement network infrastructure and procedures, in particular, sampling methodology and duration but also sampling frequency and instrument type, mean that pre-processing of both measurements and model predictions is often required before measured values and model predictions can be combined or compared across networks. Most importantly, after station-level screening for validity and completeness, the measurements from some networks had to be averaged or accumulated in time to match the longest sampling duration employed by any network for that same species in order to calculate consistent all-station (i.e., multiple-network or combined) statistics. For

example, the U.S. CASTNET and NADP networks collect weekly samples and the U.S. AMoN network collects biweekly samples. ~~In a few cases~~For the same reason, predicted hourly ~~model~~ RAQDPS-OP023 species values ~~also~~ frequently had to be averaged or summed before being paired with measurements.

Of the three key AQHI forecast species, evaluation scores for $O_3$ hourly forecasts made by the RAQDPS-OP023 were generally the highest for all five years, followed by $NO_2$ scores and then $PM_{2.5}$ scores. The finding that $PM_{2.5}$ total mass was the most difficult AQHI pollutant to predict was not surprising given its inherent complexity as a hybrid, primary-secondary multi-pollutant that spans a wide size range and has numerous sources. Another important finding was that monthly mean hourly $PM_{2.5}$ total mass predicted by the RAQDPS-OP023 was biased low in all months in 2021/22 and in most months, especially in the summer, in 2013–2016. Relative to the NMB, NMAE, and R "acceptability" benchmarks recommended by Zhai et al. (2024) for $NO_2$ VMR predictions, all-station seasonal RAQDPS-OP023 predictions met the NMB benchmark for 19 out of 20 seasons for these five years, met the NMAE benchmark for the five winter seasons, and met the R benchmark for all 20 seasons. Similarly, for the NMB, NMAE, and R acceptability benchmarks recommended by Emery et al. (2017) for $O_3$ VMR predictions, all-station seasonal RAQDPS-OP023 predictions met the NMB benchmark for 15 seasons (but none of the springs), met the NMAE benchmark for only one season, but met the R benchmark for all 20 seasons. And for the NMB, NMAE, and R acceptability benchmarks recommended by Emery et al. (2017) for $PM_{2.5}$ concentration predictions, all-station seasonal RAQDPS-OP023 predictions met the NMB benchmark for 16 seasons (but not four summers), did not meet the NMAE benchmark for any season, but met the R benchmark for all 20 seasons. The ongoing WMO GAFIS multi-model comparison of operational AQ forecast systems for North America also found RAQDPS-OP023 forecasts to be competitive with four peer forecast systems for $NO_2$ and $O_3$ for all months in 2021/22 and for $PM_{2.5}$ total mass for cold-season months.

Biomass burning is an important seasonal source of emissions of both primary $PM_{2.5}$ mass and some of its gas-phase precursors. A comparison of RAQDPS-FW023 and RAQDPS-OP023 performance for 2021/22 found much improved $PM_{2.5}$ evaluation scores for the RAQDPS-FW023 for the summer months, when BB emissions peak. It was also shown that the inclusion of BB emissions that occur mainly in the summer and early autumn affected annual evaluation statistics significantly. This comparison has quantified the impact of BB emissions on AQ forecasts for one year and has provided strong evidence for the importance of including BB emissions, which mainly affect $PM_{2.5}$ levels in the summer and, to a lesser degree, autumn. One further insight came from separate evaluations with hourly $PM_{2.5}$ measurements that had been divided into urban and rural subsets, which was that once BB emissions were included, remaining model underpredictions of hourly $PM_{2.5}$ concentration were found to occur mainly in rural areas from October to June for both system versions~~, whereas model overpredictions of $PM_{2.5}$ concentration were sometimes a second issue in urban areas~~. Other explanations are thus still needed for hourly $PM_{2.5}$ underpredictions outside of the wildfire season, especially in rural areas and in the eastern U.S.

The evaluation of $PM_{2.5}$ predictions for the 2013–2016 annual hindcasts was expanded by considering two additional types of $PM_{2.5}$ measurements available from multiple networks, namely (i) daily gravimetric $PM_{2.5}$ total mass

measurements from the U.S. $PM_{2.5}$ mass monitoring network and (ii) daily $PM_{2.5}$ chemical composition measurements from the CSN, IMPROVE, and NAPS networks. These additional $PM_{2.5}$ data sets ~~augmented~~ included measurements from nearly as many stations as the hourly $PM_{2.5}$ total mass data sets but employed very different measurement technologies with different error characteristics. By comparing and contrasting evaluation results for these multiple networks, ~~and provided~~ insights could be drawn into causes of the hourly $PM_{2.5}$ mass underpredictions. One finding was that RAQDPS-OP0234 predictions of daily gravimetric $PM_{2.5}$ total mass were less negatively biased than those for hourly $PM_{2.5}$ mass, which immediately draws attention to instrument and analysis differences between the two types of measurements. Unlike hourly continuous $PM_{2.5}$ total mass measurements, the daily gravimetric $PM_{2.5}$ total mass measurements are known to have negative artefacts due to the volatilization of semi-volatile species such as ammonium, nitrate, and particle water. This difference suggests that hourly continuous $PM_{2.5}$ measurements might be higher than gravimetric measurements if those measurements were taken at the same time and location, thus resulting in a larger negative bias for the former. Monthly evaluation scores for the daily gravimetric $PM_{2.5}$ total mass measurements were also closer to those for the CSN and NAPS networks than the IMPROVE network, which is consistent with the first three networks being urban-focused whereas the IMPROVE network is rural-focused. A second finding was that daily $PM_{2.5}$ mass reconstruction results based on observed daily $PM_{2.5}$ speciation measurements agreed well both with observed daily gravimetric $PM_{2.5}$ total mass measurements for 2013–2016 and with RAQDPS-OP0234 predictions of daily gravimetric $PM_{2.5}$ total mass, which ~~was~~ again was somewhat surprising in ~~view~~ light of the consistent model underpredictions of hourly $PM_{2.5}$ total mass. However, $PM_{2.5}$ mass reconstruction values did not include aerosol water and t~~T~~he RAQDPS-OP0234 calculation of hourly $PM_{2.5}$ total mass~~, however,~~ did not include either aerosol water or SS components, but comparisons with observed gravimetric $PM_{2.5}$ total mass, which would still include a particle-bound water component, suggested that both $PM_{2.5}$ chemical components should be included in the model calculation of $PM_{2.5}$ total mass. ~~And~~ Moreover, while RAQDPS-OP0234 predictions of all-station $PM_{2.5}$ chemical composition were reasonably good, it was also shown that the good agreement between observed and predicted daily gravimetric $PM_{2.5}$ total mass was partly due to compensating RAQDPS-OP023~~model~~ errors in the prediction of individual $PM_{2.5}$ chemical components. In particular, ~~because~~ the model was found to overpredict $PM_{2.5}$-EC and SS but underpredict $SO_4$ in all seasons and to overpredict $PM_{2.5}$-TOM and CM at urban stations but underpredict these components at rural stations. Lastly, both observed and predicted seasonal $PM_{2.5}$ residual mass (the difference between gravimetric mass and reconstructed mass) were found to be largest in the summer while annual $PM_{2.5}$ residual mass was largest in eastern North America. Some possible explanations include incorrect ~~partitioning~~ spatial allocation of primary $PM_{2.5}$ emissions between urban and rural areas and underpredictions of ammonium nitrate, crustal material, and biogenic SOA in the summer.

In the third part of the evaluation, trends in seasonal performance were shown for the first decade of operational forecasts by the online GEM-MACH-based version of the RAQDPS from 2010 to 2019. Overall skill in predicting hourly $NO_2$ and $O_3$ VMRs improved modestly over this period due to a series of modelling system upgrades, including code improvements and updated input emissions files (see companion paper by Moran et al., 2026~~5~~). Predictions of hourly $PM_{2.5}$ total mass, on the other hand, improved initially but then declined after 2017. These seasonal forecast

scores for earlier operational RAQDPS versions were then compared to seasonal scores for the RAQDPS-OP023~~3~~ hindcast simulations for 2013–2016 and 2021/22 forecasts. RAQDPS-OP023~~3~~ seasonal scores were better overall for $NO_2$ and $O_3$ but showed little improvement for $PM_{2.5}$.

Meteorology and climate can also affect ~~model~~ RAQDPS performance through the direct influence of some meteorological variables, such as wind speed, PBL height, temperature, precipitation, solar radiation, and cloud cover, on pollutant abundances and removal, and indirectly through related factors such as vegetation phenology, snow cover, and emissions affected by meteorology. As a consequence, there were seasonal or diurnal cycles in objective scores for many pollutants that were as large or larger than year-to-year fluctuations or trends in model performance. These temporal variations in model performance can provide further clues and guidance to identify those components of the modelling system where further improvements may be needed. Examples include maximum underprediction of surface $O_3$ VMR in the spring, ~~peak~~ maximum underprediction of $NH_3$ VMR in winter, and consistent underpredictions of $PM_{2.5}$ ~~–~~$SO_4$ for all seasons. Some model errors are also correlated in time for chemically coupled species, such as those for $SO_2$ and $PM_{2.5}$-$SO_4$, for $HNO_3$ and $PM_{2.5}$-$NO_3$, and for $NH_3$ and $PM_{2.5}$-$NH_4$, and may be in phase or out of phase.

One confounding factor faced by this study was the changing chemical environment in North America caused by significant changes in regional and national emissions that have occurred over the past two decades, including the four-year period from 2013 to 2016 and the six-year period from 2016 to 2021. For example, the impact of large monotonic reductions in Canadian and U.S. $SO_2$ annual emissions of 16% and 50%, respectively, from 2013 to 2016 can be seen in both observed and predicted monthly time series of $SO_2$ VMR, $PM_{2.5}$-$SO_4$ concentration, and $SO_4^=$ concentration in precipitation and wet deposition. This agreement in temporal trends between measurements and ~~model~~ RAQDPS-OP023 predictions affirms the representativeness of the retrospective, multi-year emission inventory data sets used ~~for~~ in this study for 2013–2016 and hence represents a positive result from a dynamic evaluation perspective. However, having successfully represented these emissions changes in the 2013–2016 RAQDPS-OP023~~model~~ simulations, monthly and seasonal scores were found to vary in a similar manner from year to year, which points to systematic errors in the RAQDPS-OP023~~model~~ itself (although 2016 seemed to be an outlier for some species). In addition, these results clearly show the value of using year-specific emissions, although this is not possible in the case of AQ forecast models because current emissions cannot be known beforehand.

This study has demonstrated the value of a comprehensive, quasi-diagnostic model performance evaluation. Consideration of a wide range of pollutant species allows some pollutant mass budgets to be examined and consistency to be checked for coupled pollutant species. Different statistical analyses can also provide different information. For example, spatial plots of ~~site-level~~station statistics can show regional clustering of similar station scores and regional differences in station scores, including compensating ~~site-level~~ errors between regions or between urban and rural stations that can be hidden by~~in~~ and improve network-level aggregated statistics. Density scatterplots can show different levels of scatter between species or seasons at the station level and reveal low measurement precision by making quantization of observed values visible (e.g., NO and $NO_2$). Time series of monthly or hourly mean values are more aggregated than the scatterplots but can help to understand temporal variations in performance and, when collated,

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mass budgets. Other s~~S~~tratified analyses (e.g., by individual network, season, month~~ly~~, hour of day, region, or urban/rural) have also revealed model behaviour, including compensating errors, that was not visible in highly aggregated (e.g., all-station, annual) "headline" statistics. This comprehensive, quasi-diagnostic model performance evaluation has also provided a baseline set of scores against which future system versions can be compared.

Results from the performance evaluation described in this paper have thus pointed to aspects of the RAQDPS-OP023 that warrant further investigation and improvement. First, the anthropogenic input emissions files and the parameterizations of natural emissions used by the forecast system should be one important focus, especially as anthropogenic emissions continue to change. In addition to strong evidence to support the inclusion of BB emissions, other results point to the likely overallocation of $PM_{2.5}$ primary emissions to urban areas and underallocation to rural areas, overallocation of $NH_3$ emissions to summer and underallocation to winter, overallocation of fugitive dust emissions to winter, excessive sea-salt emissions in all seasons, and several missing natural emissions sources, including wind-blown dust from natural surfaces ~~and lightning NO emissions~~. Evaluation results for ISOP VMR also showed large positive biases, arguing for the biogenic emissions scheme to be revisited. Second, $O_3$ forecasts might be improved by (i) revising $O_3$ western lateral boundary conditions for the spring and summer and (ii) introducing a parameterization of halogen chemistry impacts ~~near oceans~~ on $O_3$ near oceans. Third, the RAQDPS-OP023 was missing two gas-phase dry deposition processes for $SO_2$, and the addition of these processes should decrease $SO_2$ and $HNO_3$ levels, which were both overpredicted. Seasonal surface properties in the gas-phase dry deposition scheme also need to have a more detailed representation of seasonal variations and to include dependence on longitude and elevation, which was lacking in the RAQDPS-OP023. Fourth, precipitation-chemistry measurements suggested that too much $NH_3$ gas was being removed by wet deposition. Fifth, the RAQDPS-OP023 was found to underpredict $PM_{2.5}$-$SO_4$ and overpredict $PM_{2.5}$-EC and $PM_{2.5}$-SS concentrations in all seasons, and the treatments of the lifecycles of each of these components will need to be examined. Particulate organic matter, especially $PM_{2.5}$-SOM from biogenic emissions, may be underpredicted in the summer, which will require examination of the representation of biogenic emissions and of SOA formation. $PM_{2.5}$-$NH_4$ and $PM_{2.5}$-$NO_3$ may also be underpredicted in the summer, which will require the representation of inorganic heterogeneous thermodynamics to be assessed, including the addition of an explicit treatment of base cations, a known gap in the RAQDPS-OP023. Lastly, the calculation of hourly $PM_{2.5}$ total mass in the RAQDPS by summation of $PM_{2.5}$ chemical components should include sea salt and both volatile and particle-bound aerosol water.

As a final comment it should be noted that the RAQDPS-OP025, which was introduced operationally at ECCC in June 2024, did implement some of the changes recommended above (see companion paper by Moran et al., 2026). One very significant change was to merge the two operational system versions, the RAQDPS-OP024 and RAQDPS-FW024, into a single system version that included NRT biomass burning emissions. This particular change had the multiple advantages of improving forecast scores, simplifying maintenance of the operational AQ forecast system, and reducing computer usage.

**Appendix**

Table A1. List of acronyms~~, and~~ abbreviations, and symbols.~~:~~

| | |
|---|---|
| ADOM | Acid Deposition and Oxidant Model (Canada) |
| ADOM-2 | Acid Deposition and Oxidant Model gas-phase chemical mechanism version 2 |
| AirNow | Aerometric Information Retrieval Now (U.S.) |
| ALD2 | acetaldehyde and higher aldehydes (ADOM-2 lumped VOC species) |
| AMoN | Ammonia Monitoring Network (U.S.) |
| APEI | Air Pollutant Emissions Inventory (Canada) |
| AQ | air quality |
| AQHI | Air Quality Health Index (Canada) |
| AQS | Air Quality System (U.S.) |
| BB | biomass burning |
| BDL | below detection limit |
| BEIS | Biogenic Emission Inventory System (U.S.) |
| CAMS | Copernicus Atmosphere Monitoring Service (EU) |
| CAPMoN | Canadian Acid Precipitation Monitoring Network |
| CASTNET | Clean Air Status and Trends Network (U.S.) |
| CEDS | Community Emissions Data System |
| CFFEPS | Canadian Forest Fire Emissions Prediction System |
| CFR | Code of Federal Regulations (U.S.) |
| CM | crustal material |
| CMC | Canadian Meteorological Centre |
| CRES | cresols and phenols (ADOM-2 lumped VOC species) |
| CRMSE | centered root mean square error |
| CSN | Chemical Speciation Network (U.S.) |
| CV | coefficient of variation (or relative standard deviation) |
| CVM | coefficient of variation (model) |
| CVO | coefficient of variation (observations) |
| EC | elemental carbon |
| ECA | eastern Canada |
| ECCC | Environment and Climate Change Canada |
| ECMWF | European Centre for Medium-range Weather Forecasts |
| EEA | European Environment Agency |
| EI | emission inventory |
| EMEP | European Monitoring and Evaluation Programme |
| EOTH | total non-reactive or low-reactivity VOC species (ADOM-2 lumped VOC species) |
| EPA | Environmental Protection Agency (U.S.) |
| EQUATES | EPA's air QUAlity TimE Series |
| EST | Eastern Standard Time |
| ETHE | ethene and some isoprene oxidation products (ADOM-2 lumped VOC species) |
| EUS | eastern U.S. |
| FAC2 | factor-of-two metric |
| FB | fractional bias |

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| FE | fractional error |
| FEM | Federal Equivalent Method (U.S.) |
| FRM | Federal Reference Method (U.S.) |
| FW | FireWork |
| GAFIS | Global Air quality Forecasting and Information System (WMO) |
| GAW | Global Atmospheric Watch (WMO) |
| GEM | Global Environmental Multiscale (model) (ECCC) |
| GEM-MACH | Global Environmental Multiscale–Modelling Atmospheric CHemistry (model) (ECCC) |
| GEMS | Global and regional Earth-system Monitoring using Satellite and in-situ data (EU) |
| GEOS-CF | Goddard Earth Observing System - Composition Forecasting (NASA) |
| GIS | geographic information system |
| HETV | HETerogeneous Vectorized scheme |
| IAU | incremental analysis update |
| IAY | Instantaneous secondary organic Aerosol Yield |
| IFS-CAMS | Integrated Forecasting System – Copernicus Atmosphere Monitoring Service (ECMWF) |
| IFS-SILAM | Integrated Forecasting System – SILAM (ECMWF/Finnish Meteorological Institute) |
| IOA | index of agreement |
| IMPROVE | Interagency Monitoring of Protected Visual Environments (U.S.) |
| ISOP | isoprene |
| LBC | lateral boundary condition |
| LT | local time |
| MAE | mean absolute error |
| MB | mean bias |
| MDA8 | maximum daily 8-hr average |
| MDL | minimum detection limit |
| MFB | mean fractional bias |
| MFE | mean fractional error |
| MMR | mass mixing ratio |
| MOVES | MOtor Vehicle Emission Simulator (U.S.) |
| N | number of measurement-model pairs (i.e., sample size) |
| NAAQS | National Ambient Air Quality Standards (U.S.) |
| NACC | NOAA-EPA Atmosphere-Chemistry Coupler |
| NADP | National Atmospheric Deposition Program (U.S.) |
| NAPS | National Air Pollution Surveillance system (Canada) |
| NAQFC | National Air Quality Forecast Capability (U.S.) |
| NASA | National Aeronautics and Space Administration (U.S.) |
| NAtChem | National Atmospheric Chemistry database (Canada) |
| NATTS | National Air Toxics Trends Sites (U.S.) |
| NEI | National Emission Inventory (U.S.) |
| NH4 | particle ammonium |
| NMAE | normalized MAE |
| NMB | normalized mean bias |
| NME | normalized mean error (assumed equivalent to NMAE) |
| NMSE | normalized mean square error |

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|---|---|
| NO3 | particle nitrate |
| NOAA | National Oceanic and Atmospheric Administration (U.S.) |
| NPRI | National Pollutant Release Inventory (Canada) |
| NRT | near real time |
| NSD | normalized standard deviation |
| NTN | National Trends Network (U.S. NADP) |
| NWP | numerical weather prediction |
| OC | organic carbon |
| OM | organic matter |
| PAMS | Photochemical Assessment Monitoring Stations (U.S.) |
| PBL | planetary boundary layer |
| PCL | Precipitation Coverage Length |
| PM | particulate matter |
| $PM_{2.5}$ | particulate matter with aerodynamic diameter smaller than 2.5 µm |
| $PM_{2.5}$-noSS | $PM_{2.5}$ without sea-salt component |
| POM | primary organic matter |
| PR | precipitation |
| QA/QC | quality assurance/quality control |
| R | Pearson correlation coefficient |
| RAQDPS | Regional Air Quality Deterministic Prediction System (ECCC) |
| RAQDPS-FW | RAQDPS-FireWork (with BB emissions) |
| RAQDPS-OP | RAQDPS-Operational (standard, no BB emissions) |
| RDPS | Regional Deterministic Prediction System (ECCC) |
| RH | relative humidity |
| RMSE | root mean square error |
| SCC | Source Classification Code |
| SDM | standard deviation of model predictions |
| SDO | standard deviation of observations |
| SEMARNAT | Secretariat of Environment and Natural Resources (Mexico) |
| SGS | subgrid-scale |
| SILAM | System for Integrated modeLling of Atmospheric composition (Finnish Meteorological Institute) |
| S/L | state/local |
| SMOKE | Sparse Matrix Operator Kernel Emissions (modeling system) |
| SO4 | particle sulfate |
| SOA | secondary organic aerosol |
| SOM | secondary organic matter |
| SS | sea salt |
| STP | standard temperature and pressure |
| TEOM | tapered element oscillating microbalance |
| TF | transportable fraction |
| TOC | total organic carbon |
| TOM | total organic matter |
| TP | total precipitation |
| UTC | Coordinated Universal Time |

| | |
|---|---|
| VMR | volume mixing ratio |
| VOC | volatile organic compound |
| WCA | western Canada |
| WMO | World Meteorological Organization |
| WRF | Weather Research and Forecasting NWP model (U.S.) |
| WUS | western U.S. |
| ADOM | Acid Deposition and Oxidant Model (Canada) |
| AirNow | Aerometric Information Retrieval Now (U.S.) |
| ALD2 | acetaldehyde and higher aldehydes (ADOM-2 lumped VOC species) |
| AMoN | Ammonia Monitoring Network (U.S.) |
| APEI | Air Pollutant Emissions Inventory (Canada) |
| AQ | air quality |
| AQHI | Air Quality Health Index (Canada) |
| AQS | Air Quality System (U.S.) |
| BB | biomass burning |
| BDL | below detection limit |
| BEIS | Biogenic Emission Inventory System (U.S.) |
| CAMS | Copernicus Atmosphere Monitoring Service (EU) |
| CAPMoN | Canadian Acid Precipitation Monitoring Network |
| CASTNET | Clean Air Status and Trends Network (U.S.) |
| CEDS | Community Emissions Data System |
| CFFEPS | Canadian Forest Fire Emissions Prediction System |
| CFR | Code of Federal Regulations (U.S.) |
| CM | crustal material |
| CMC | Canadian Meteorological Centre |
| CRES | cresols and phenols (ADOM-2 lumped VOC species) |
| CRMSE | centered root mean square error |
| CSN | Chemical Speciation Network (U.S.) |
| CV | coefficient of variation (or relative standard deviation) |
| EC | elemental carbon |
| ECA | eastern Canada |
| ECCC | Environment and Climate Change Canada |
| ECMWF | European Centre for Medium-range Weather Forecasts |
| EEA | European Environment Agency |
| EI | emission inventory |
| EMEP | European Monitoring and Evaluation Programme |
| EPA | Environmental Protection Agency (U.S.) |
| EQUATES | EPA's Air QUAlity TimE Series |
| EST | Eastern Standard Time |
| ETHE | ethene and some isoprene oxidation products (ADOM-2 lumped VOC species) |
| EUS | eastern U.S. |
| FAC2 | factor-of-two metric |
| FB | fractional bias |
| FE | fractional error |

| | | |
|---|---|---|
| | FEM | Federal Equivalent Method (U.S.) |
| | FRM | Federal Reference Method (U.S.) |
| 1945 | FW | FireWork |
| | GAFIS | Global Air quality Forecasting and Information System (WMO) |
| | GAW | Global Atmospheric Watch (WMO) |
| | GEM | Global Environmental Multiscale (model) (ECCC) |
| | GEM-MACH | Global Environmental Multiscale Modelling Atmospheric CHemistry (model) (ECCC) |
| 1950 | GEMS | Global and regional Earth-system Monitoring using Satellite and in-situ data (EU) |
| | GEOS-CF | Goddard Earth Observing System – Composition Forecasting (NASA) |
| | GIS | geographic information system |
| | HETV | HETerogeneous Vectorized scheme |
| | IAU | incremental analysis update |
| 1955 | IAY | Instantaneous secondary organic Aerosol Yield |
| | IFS-CAMS | Integrated Forecasting System – Copernicus Atmosphere Monitoring Service (ECMWF) |
| | IFS-SILAM | Integrated Forecasting System – SILAM (ECMWF/Finnish Meteorological Institute) |
| | IOA | index of agreement |
| | IMPROVE | Interagency Monitoring of Protected Visual Environments (U.S.) |
| 1960 | ISOP | isoprene |
| | LBC | lateral boundary condition |
| | LT | local time |
| | MAE | mean absolute error |
| | MB | mean bias |
| 1965 | MDA8 | maximum daily 8-hr average |
| | MDL | minimum detection limit |
| | MFB | mean fractional bias |
| | MFE | mean fractional error |
| | MMR | mass mixing ratio |
| 1970 | MOVES | MOtor Vehicle Emission Simulator (U.S.) |
| | NAAQS | National Ambient Air Quality Standards (U.S.) |
| | NACC | NOAA-EPA Atmosphere-Chemistry Coupler |
| | NADP | National Atmospheric Deposition Program (U.S.) |
| | NAPS | National Air Pollution Surveillance system (Canada) |
| 1975 | NAQFC | National Air Quality Forecast Capability (U.S.) |
| | NASA | National Aeronauties and Space Administration (U.S.) |
| | NAtChem | National Atmospheric Chemistry database (Canada) |
| | NATTS | National Air Toxics Trends Sites (U.S.) |
| | NEI | National Emission Inventory (U.S.) |
| 1980 | NH4 | particle ammonium |
| | NMAE | normalized MAE |
| | NMB | normalized mean bias |
| | NME | normalized mean absolute error |
| | NMSE | normalized mean square error |
| 1985 | NO3 | particle nitrate |
| | NOAA | National Oceanic and Atmospheric Administration (U.S.) |

| | |
|---|---|
| NPRI | National Pollutant Release Inventory (Canada) |
| NRT | near real time |
| NSD | normalized standard deviation |
| NTN | National Trends Network (U.S. NADP) |
| NWP | numerical weather prediction |
| OC | organic carbon |
| OM | organic matter |
| PAMS | Photochemical Assessment Monitoring Stations (U.S.) |
| PBL | planetary boundary layer |
| PCL | Precipitation Coverage Length |
| PM | particulate matter |
| $PM_{2.5}$ | particulate matter with aerodynamic diameter smaller than 2.5 µm |
| POM | primary organic matter |
| PR | precipitation |
| QA/QC | quality assurance/quality control |
| RAQDPS | Regional Air Quality Deterministic Prediction System (ECCC) |
| RAQDPS-FW | RAQDPS-FireWork |
| RDPS | Regional Deterministic Prediction System (ECCC) |
| RH | relative humidity |
| RMSE | root mean square error |
| SCC | Source Classification Code |
| SDM | standard deviation of model predictions |
| SDO | standard deviation of observations |
| SEMARNAT | Secretariat of Environment and Natural Resources (Mexico) |
| SGS | subgrid-scale |
| SILAM | System for Integrated modeLling of Atmospheric composition (Finnish Meteorological Institute) |
| S/L | state/local |
| SMOKE | Sparse Matrix Operator Kernel Emissions (modeling system) |
| SO4 | particle sulfate |
| SOA | secondary organic aerosol |
| SOM | secondary organic matter |
| SS | sea salt |
| STP | standard temperature and pressure |
| TEOM | tapered element oscillating microbalance |
| TF | transportable fraction |
| TOC | total organic carbon |
| TOM | total organic matter |
| TP | total precipitation |
| UTC | Coordinated Universal Time |
| VMR | volume mixing ratio |
| VOC | volatile organic compound |
| WCA | western Canada |
| WMO | World Meteorological Organization |
| WRF | Weather Research and Forecasting NWP model (U.S.) |

WUS western U.S.

Table A2. Statistical metric~~easure~~s used in this study for model performance evaluation plus coefficient of variation (or relative standard deviation). MFB and MFE have been used in some related studies (e.g., Manseau et al., 2022)

| Metric Name | Abbreviation | Definition |
|---|---|---|
| Observed mean | $\bar{O}$ | $\frac{1}{N}\sum_{1}^{N} O_i$ |
| Model mean | $\bar{M}$ | $\frac{1}{N}\sum_{1}^{N} M_i$ |
| Mean bias | MB | $\frac{1}{N}\sum_{1}^{N} (M_i - O_i)$ |
| Root mean square error | RMSE | $\left(\frac{1}{N}\sum_{1}^{N} (M_i - O_i)^2\right)^{1/2}$ |
| Normalized mean bias | NMB | $\frac{\sum_1^N (M_i - O_i)}{\sum_1^N O_i}$ |
| Normalized mean absolute error | NMAE | $\frac{\sum_1^N \lvert M_i - O_i \rvert}{\sum_1^N \lvert O_i \rvert}$ |
| Pearson correlation coefficient | R | $\frac{\sum[(M_i - \bar{M}) \times (O_i - \bar{O})]}{\sqrt{\sum(M_i - \bar{M})^2 \times \sum(O_i - \bar{O})^2}}$ |
| Centred RMSE | CRMSE | $\left(\frac{1}{N}\sum_{1}^{N} [(M_i - \bar{M}) - (O_i - \bar{O})]^2\right)^{1/2}$ |
| Standard deviation (observations) | $\sigma_O$ (or SDO) | $\left(\frac{1}{N}\sum_{1}^{N} (O_i - \bar{O})^2\right)^{1/2}$ |

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|---|---|---|
| Standard deviation (model) | $\sigma_M$ (or SDM) | $\left(\frac{1}{N}\sum_1^N (M_i - \bar{M})^2\right)^{1/2}$ |
| Normalized standard deviation | NSD | $\sigma_M / \sigma_O$ |
| Coefficient of variation (observations) | CVO | $\sigma_O / \overline{O}$ |
| Coefficient of variation (model) | CVM | $\sigma_M / \overline{M}$ |
| Mean fractional bias | MFB | $\frac{2}{N}\sum_1^N \left(\frac{M_i - O_i}{M_i + O_i}\right)$ |
| Mean fractional error | MFE | $\frac{2}{N}\sum_1^N \left\|\frac{M_i - O_i}{M_i + O_i}\right\|$ |

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| Metric Name | Abbreviation | Definition |
|---|---|---|
| Observed mean | $\bar{O}$ | $\frac{1}{N}\sum_1^N O_t$ |
| Model mean | $\bar{M}$ | $\frac{1}{N}\sum_1^N M_t$ |
| Mean bias | MB | $\frac{1}{N}\sum_1^N (M_t - O_t)$ |
| Root mean square error | RMSE | $\left(\frac{1}{N}\sum_1^N (M_t - O_t)^2\right)^{1/2}$ |
| Normalized mean bias | NMB | $\frac{\sum_1^N (M_t - O_t)}{\sum_1^N O_t}$ |
| Normalized mean absolute error | NME | $\frac{\sum_1^N \lvert M_t - O_t\rvert}{\sum_1^N \lvert O_t\rvert}$ |
| Pearson correlation coefficient | R | $\frac{\sum[(M_t - \bar{M}) \times (O_t - \bar{O})]}{\sqrt{\sum (M_t - \bar{M})^2 \times \sum (O_t - \bar{O})^2}}$ |
| Centred RMSE | CRMSE | $\left(\frac{1}{N}\sum_1^N [(M_t - \bar{M}) - (O_t - \bar{O})]^2\right)^{1/2}$ |
| Standard deviation (observations) | $\sigma_O$ (or SDO) | $\left(\frac{1}{N}\sum_1^N (O_t - \bar{O})^2\right)^{1/2}$ |

| | | |
|---|---|---|
| Standard deviation (model) | $\sigma_M$ (or SDM) | $\left(\frac{1}{N}\sum_1^N (M_i - \overline{M})^2\right)^{1/2}$ |
| Normalized standard deviation | NSD | $\sigma_M / \sigma_O$ |
| Coefficient of variation (observations) | CVO | $\sigma_O / \overline{O}$ |
| Coefficient of variation (model) | CVM | $\sigma_M / \overline{M}$ |
| Mean fractional bias | MFB | $\frac{2}{N}\sum_1^N \left(\frac{M_i - O_i}{M_i + O_i}\right)$ |
| Mean fractional error | MFE | $\frac{2}{N}\sum_1^N \left\|\frac{M_i - O_i}{M_i + O_i}\right\|$ |

Table A3. Major upgrades to the operational RAQDPS from 2009-2021. See Moran et al. (2026) for more details.

| Version | Release Date | Short Description |
|---|---|---|
| 001 | Nov. 2009 | First version (Emission Inventories: 2006 CA, 2005 US, 1999 MX) |
| 004 | Oct. 2011 | New emissions (EIs: 2006 CA, projected 2012 US, 1999 MX) |
| 007 | Oct. 2012 | New model code, new grid (15 km → 10 km, 58 → 80 levels) |
| 009 | Feb. 2013 | New model code with 3 bug fixes, including one to near-surface vertical diffusion |
| 013 | Jun. 2015 | New emissions (EIs: 2010 CA, 2011 US, 1999 MX) |
| 016 | Sep. 2016 | New model code, new vertical discretization (non-staggered → staggered) |
| 020 | Sep. 2018 | New model code, new emissions (EIs: 2013 CA, projected 2017 US, 2008 MX) |
| 021 | Jul. 2019 | New model code, new vertical discretization (80 → 84 levels), longer forecast (2–>3 days) |
| 023 | Nov. 2021 | New model code, new emissions (EIs: projected 2020 CA, projected 2023 US & 2023 MX) |

Table A4. Summary of Canadian and U.S. national annual wildfire statistics for the 2013–2022 period. Data sources: Canadian Interagency Forest Fire Centre (2023); NOAA National Centers for Environmental Information (2023).

| | Number of Fires | | | Hectares Burned | | |
|---|---|---|---|---|---|---|
| **Year** | **Canada** | **U.S.** | **Total** | **Canada** | **U.S.** | **Total** |
| 2013 | 6,246 | 46,615 | 52,861 | 4,203,867 | 1,743,054 | 5,946,921 |
| 2014 | 5,126 | 63,345 | 68,471 | 4,563,847 | 1,451,836 | 6,015,683 |
| 2015 | 7,068 | 61,922 | 68,990 | 3,903,277 | 4,097,506 | 8,000,783 |
| 2016 | 5,173 | 65,575 | 70,748 | 1,532,440 | 2,204,130 | 3,736,570 |
| 2017 | 5,611 | 66,131 | 71,742 | 3,371,825 | 3,958,259 | 7,330,084 |
| 2018 | 7,068 | 55,911 | 62,979 | 2,272,269 | 3,473,262 | 5,745,531 |
| 2019 | 3,933 | 49,786 | 53,719 | 1,787,793 | 1,873,699 | 3,661,492 |
| 2020 | 3,916 | 58,258 | 62,174 | 227,389 | 4,158,019 | 4,385,408 |
| 2021 | 6,596 | 58,733 | 65,329 | 4,307,520 | 2,889,342 | 7,196,862 |
| 2022 | 5,726 | 66,255 | 71,981 | 1,656,504 | 3,049,067 | 4,705,571 |

Table A5. Comparison of selected NAQFC domain-average monthly statistics for daily mean surface $PM_{2.5}$ predictions ($\mu g \cdot m^{-3}$) for 2014 and 2015 (from Lee et al., 2017) with domain-average seasonal statistics for RAQDPS010 forecasts for hourly mean surface $PM_{2.5}$ predictions ($\mu g \cdot m^{-3}$) for spring (MAM) and summer (JJA) 2014 and RAQDPS011 forecasts for winter (DJF) 2015 (from Moran et al., 2021a).

| **Period** | **System** | **O** | **M** | **MB** | **NMB** | **RMSE** | **R** |
|---|---|---|---|---|---|---|---|
| May 2014 | NAQFC | 7.76 | 6.20 | -1.56 | -0.20 | 4.46 | 0.32 |
| Spring 2014 | RAQDPS010 | | | -0.88 | -0.12 | 8.02 | 0.37 |
| July 2014 | NAQFC | 9.93 | 6.62 | -2.71 | -0.28 | 5.36 | 0.23 |
| Summer 2014 | RAQDPS010 | | | 0.64 | 0.07 | 11.21 | 0.29 |
| January 2015 | NAQFC | 9.83 | 11.16 | 1.33 | 0.13 | 6.46 | 0.38 |
| Winter 2015 | RAQDPS011 | | | -1.04 | -0.12 | 10.15 | 0.38 |

***Code and data availability***

Version 5.1 of the GEM numerical weather prediction model code used by the RAQDPS023 is free software which can be redistributed and/or modified under the terms of version 2.1 of the GNU Lesser General Public License as published by the Free Software Foundation. The GEM source code has been developed by the Meteorological Research Division of ECCC. This code is available for download from https://zenodo.org/records/17782580 (~~Environment and Climate Change Canada~~ECCC, 2025b).

MACH, the atmospheric chemistry library for the GEM model (©2007–2021, Air Quality Research Division and National Prediction Operations Division, Environment and Climate Change Canada), is free software that can be redistributed and/or modified under the terms of the GNU Lesser General Public License as published by the Free

Software Foundation – either version 2.1 of the license or any later version. The GEM-MACH version 3.1.0.0 code used by the RAQDPS023 can be downloaded from website https://doi.org/10.5281/zenodo.15330612 (Savic-Jovcic et al., 2025). Related documentation is also available on that website, including information about key input and configuration files and copies of several relevant reports. The GEM-MACH v3.1.1.2 source code for the RAQDPS024, an equivalent version to the RAQDPS023 that went into operation after a migration to a new ECCC high-performance computer system in June 2022, is available at https://zenodo.org/records/13952893 (GEM-MACH development team, 2022).

The CFFEPS version 4.1 code that was used by the RAQDPS-FW023 and RAQDPS-FW024 is free software that can be redistributed and/or modified under the terms of the GNU Lesser General Public License, either version 2.1 or any later version, as published by the Free Software Foundation. It is available to download from website https://doi.org/10.5281/zenodo.15305591 (Anderson and Chen, 2021).

Data sources for the AQ measurement data sets used in this study are listed in Table S2a of the Supplement. Multiple sets of files containing (i) the "raw" measurements that were used in this study, (ii) intermediate measurement files after necessary unit conversions and proxy calculations, (iii) filtered measurement files after application of validity and temporal completeness checks, and (iv) final, evaluation-ready, paired model-measurement files after temporal aggregation for data pooling and filtering for reconstructed mass completeness are available from https://doi.org/10.5281/zenodo.16944371 (Lupu and Moran, 2025). This data repository also contains a complete set of the output files of evaluation scores that were the basis for all of the evaluation-related tables and figures presented in this paper.

A package of seasonal and annual model-predicted gridded surface concentration fields and dry, wet, and total acidic deposition fields for the 2013-2016 simulations is available from the website https://doi.org/10.5281/zenodo.16970403 (Moran and Savic-Jovcic, 2025). Some of the archived model hourly output fields used to create this package were also used by Cathcart et al. (2025) and Robichaud et al. (2025, 2026) in their recent papers.

All other data sets used in this work are available upon request from the authors. Please contact one of the corresponding authors to make a request.

***Supplement***

The supplement related to this article is available on-line at https://doi.org/10.5281/zenodo.19489036 ~~6929949~~(Moran and Lupu, 2026).

***Author contributions***

MDM was the science lead for the development of the online RAQDPS from the RAQDPS001 up to the RAQDPS023 and was the co-supervisor for all operational deliveries from 2009 to 2021. He conceived the objectives and scope of

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this study, oversaw the evaluations of the RAQDPS023, and prepared the initial and final versions of this paper. AL assisted with the 2013-2016 simulations, obtained all AQ measurement data sets, developed all measurement data processing and evaluation scripts, and generated all evaluation tables and data-related figures for the 2013–2016 and 2021/22 annual simulations, VSJ developed and maintained RAQDPS code and scripts, performed the 2013–2016 annual simulations, and prepared model-related figures and analyses. JZ performed the 2010–2019 operational evaluation, and JZ, QZ, EIB, and RM generated the 2013–2016 and 2021/22 anthropogenic emissions files and performed the analyses to construct Tables 1 and S1 and Figure S1. CAS led the migration from the RAQDPS023 to the RAQDPS024 on the new CMC supercomputers in June 2022 and the development of the RAQDPS025, which became operational in June 2024. SM has been the co-supervisor for all operational deliveries of the RAQDPS with the assistance of VSJ, JC, KM, RMA, and DK. JC led the development and delivery of the RAQDPS-FW023 and all CFFEPS versions with the assistance of KM and RMA. PMM oversaw operational evaluation of North American NRT AQ forecasts at ECCC for GAFIS. Lastly, AL, VSJ, JZ, RM, CAS, JC, QZ, EIB, SM, and RMA reviewed the manuscript.

***Competing interests***

The authors declare that they have no conflicts of interest.

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deficiency-friendly, perceptually-uniform colour maps designed by Crameri et al. (2020) for some figures in this paper.

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## Tables

Table 1. Comparison of Canadian, U.S., and Mexican annual anthropogenic and biogenic criteria-air-contaminant inventory emissions (tonnes/year) and model-ready anthropogenic and biogenic emissions for five years: 2013–2016 and 2021/22. Note that the U.S. inventory emissions are for all 50 U.S. states and other territories and the Mexican inventory emissions are for all 32 Mexican states. The "3 Country EIs" rows provide the sum of the three national emissions inventory amounts, while the "Total Anthro" rows correspond to the ~~grid~~ limited-area, domain-total emissions that are input by the model after SMOKE emissions processing and TF scaling for fugitive PM emissions. The "Total Anthro" VOC emissions are the sum of 12 model VOC species, including EOTH (all unreactive or low-reactivity VOC species that are not considered by the gas-phase chemistry mechanism; see Moran et al, 2026~~5~~). The "Total Biogenic" emissions depend on hourly meteorology and are accumulated hourly predicted fields of soil NO emissions (in $NO_2$ units) and biogenic VOC emissions (as model VOC species) saved during each annual simulation.

| | | Year | | | | | Relative Difference (%) | | |
|---|---|---|---|---|---|---|---|---|---|
| **Species** | **Region** | **2013** | **2014** | **2015** | **2016** | **2021/22** | **2016-to-2013** | **2021/22-to-2016** | **2021/22-to-2013** |
| $SO_2$ | Canada | 1,245,629 | 1,196,925 | 1,067,133 | 1,050,048 | 720,937 | -15.7 | -31.3 | -42.1 |
| | U.S.A. | 4,855,796 | 4,632,986 | 3,232,955 | 2,453,837 | 1,614,783 | -49.5 | -34.2 | -66.7 |
| | Mexico | 1,907,119 | 1,955,040 | 1,955,789 | 1,956,538 | 1,963,766 | 2.6 | 0.4 | 3.0 |
| | 3 Country EIs | 8,008,544 | 7,784,951 | 6,255,877 | 5,460,422 | 4,299,487 | -31.8 | -21.3 | -46.3 |
| | Total Anthro | 6,788,380 | 6,566,599 | 5,008,720 | 4,248,067 | 2,683,021 | -37.4 | -36.8 | -60.5 |
| $NO_x$ | Canada | 1,859,213 | 1,812,458 | 1,749,885 | 1,689,466 | 1,534,067 | -9.1 | -9.2 | -17.5 |
| | U.S.A. | 12,010,496 | 11,318,521 | 10,317,943 | 9,281,137 | 7,462,553 | -22.7 | -19.6 | -37.9 |
| | Mexico | 2,633,018 | 2,613,843 | 2,628,180 | 2,642,516 | 2,722,692 | 0.4 | 3.0 | 3.4 |
| | 3 Country EIs | 16,502,727 | 15,744,822 | 14,696,007 | 13,613,119 | 11,719,312 | -17.5 | -13.9 | -29.0 |
| | Total Anthro | 15,286,803 | 14,565,303 | 13,489,658 | 12,425,458 | 9,759,000 | -18.7 | -21.5 | -36.2 |
| | Total Biogenic | 613,443 | 613,133 | 637,436 | 649,609 | 653,066 | 5.9 | 0.5 | 6.5 |
| | Total Anthr+Bio | 15,900,246 | 15,178,436 | 14,127,094 | 13,075,067 | 10,412,066 | -17.8 | -20.4 | -34.5 |
| VOC | Canada | 1,639,307 | 1,676,389 | 1,624,563 | 1,530,449 | 1,517,407 | -6.6 | -0.9 | -7.4 |
| | U.S.A. | 11,060,472 | 11,016,641 | 10,803,994 | 9,906,147 | 9,693,686 | -10.4 | -2.1 | -12.4 |
| | Mexico | 4,223,883 | 4,246,882 | 4,247,239 | 4,247,596 | 4,619,106 | 0.6 | 8.7 | 9.4 |
| | 3 Country EIs | 16,923,662 | 16,939,913 | 16,675,796 | 15,684,193 | 15,830,199 | -7.3 | 0.9 | -6.5 |
| | Total Anthro | 13,528,480 | 13,532,330 | 13,282,897 | 12,374,426 | 12,292,800 | -8.5 | -0.7 | -9.1 |
| | Total Biogenic | 47,373,454 | 47,372,280 | 49,734,073 | 51,315,450 | 52,655,466 | 8.3 | 2.6 | 11.1 |
| | Total Anthr+Bio | 60,901,934 | 60,904,610 | 63,016,970 | 63,689,876 | 64,948,266 | 4.6 | 2.0 | 6.6 |
| CO | Canada | 5,414,122 | 5,310,926 | 5,227,752 | 5,173,958 | 4,322,475 | -4.4 | -16.5 | -20.2 |
| | U.S.A. | 40,865,160 | 39,613,006 | 37,758,763 | 34,539,607 | 29,945,990 | -15.5 | -13.3 | -26.7 |
| | Mexico | 8,738,655 | 8,863,635 | 8,867,142 | 8,870,649 | 9,015,675 | 1.5 | 1.6 | 3.2 |
| | 3 Country EIs | 55,017,936 | 53,787,568 | 51,853,657 | 48,584,214 | 43,284,139 | -11.7 | -10.9 | -21.3 |
| | Total Anthro | 48,394,003 | 47,088,320 | 45,142,220 | 41,929,513 | 36,118,000 | -13.4 | -13.9 | -25.4 |
| $NH_3$ | Canada | 496,282 | 490,436 | 491,656 | 490,675 | 543,154 | -1.1 | 10.7 | 9.4 |
| | U.S.A. | 3,814,960 | 3,723,552 | 3,839,459 | 3,853,365 | 3,481,215 | 1.0 | -9.7 | -8.7 |
| | Mexico | 840,629 | 844,080 | 844,297 | 844,514 | 841,397 | 0.5 | -0.4 | 0.1 |
| | 3 Country EIs | 5,151,871 | 5,058,068 | 5,175,412 | 5,188,554 | 4,865,766 | 0.7 | -6.2 | -5.6 |
| | Total Anthro | 4,521,443 | 4,424,069 | 4,541,247 | 4,553,502 | 4,234,533 | 0.7 | -7.0 | -6.3 |
| $PM_{2.5}$ | Canada | 817,212 | 814,398 | 793,598 | 780,950 | 804,186 | -4.4 | 3.0 | -1.6 |
| | U.S.A. | 3,421,551 | 3,580,536 | 3,287,999 | 3,359,590 | 3,614,926 | -1.8 | 7.6 | 5.7 |
| | Mexico | 603,696 | 591,783 | 593,799 | 595,814 | 661,387 | -1.3 | 11.0 | 9.6 |
| | 3 Country EIs | 4,842,459 | 4,986,718 | 4,675,396 | 4,736,354 | 5,080,499 | -2.2 | 7.3 | 4.9 |
| | Total Anthro | 3,101,173 | 3,076,508 | 2,932,475 | 2,855,582 | 2,889,000 | -7.9 | 1.2 | -6.8 |
| $PM_{10}$ | Canada | 3,548,631 | 3,543,931 | 3,515,237 | 3,495,289 | 3,746,529 | -1.5 | 7.2 | 5.6 |
| | U.S.A. | 15,427,692 | 15,426,307 | 15,438,707 | 15,532,333 | 18,729,161 | 0.7 | 20.6 | 21.4 |
| | Mexico | 839,386 | 822,182 | 824,358 | 826,535 | 923,082 | -1.5 | 11.7 | 10.0 |
| | 3 Country EIs | 19,815,709 | 19,792,420 | 19,778,302 | 19,854,158 | 23,398,772 | 0.2 | 17.9 | 18.1 |
| | Total Anthro | 10,122,137 | 10,094,335 | 9,989,971 | 10,001,543 | 10,826,000 | -1.2 | 8.2 | 7.0 |

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Table 2. Total n~~N~~umber of U.S. and Canadian measurement stations with complete hourly $NO_2$, $O_3$, and $PM_{2.5}$ measurements for an annual evaluation vs. available measurements by network and year for 2013–2016 (AQS, NAPS) and 2021/22 (AirNow, NAPS).

| *Year* | 2013 | | 2014 | | 2015 | | 2016 | | 2021/22 | |
|---|---|---|---|---|---|---|---|---|---|---|
| | Complete | Available | Complete | Available | Complete | Available | Complete | Available | Complete | Available |
| *Variable* | | | | | | | | | | |
| | | | | AQS | | | | | | AirNow US |
| $NO_2$ | 327 | 405 | 338 | 412 | 341 | 412 | 331 | 402 | 162 | 193 |
| $O_3$ | 720 | 1310 | 733 | 1298 | 716 | 1285 | 755 | 1274 | 580 | 1085 |
| $PM_{2.5}$ | 601 | 777 | 623 | 800 | 621 | 832 | 614 | 826 | 603 | 770 |
| | | | | NAPS | | | | | | NAPS |
| $NO_2$ | 132 | 141 | 131 | 146 | 138 | 157 | 142 | 159 | 143 | 175 |
| $O_3$ | 171 | 187 | 175 | 192 | 174 | 193 | 180 | 200 | 175 | 209 |
| $PM_{2.5}$ | 158 | 189 | 179 | 196 | 165 | 198 | 174 | 198 | 173 | 203 |

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Table 3. Summary table of all-station annual statistics for the RAQDPS-OP023 for hourly $NO_2$, $O_3$, and $PM_{2.5}$ surface measurements for 2021/22 and 2013–2016. For dimensional statistics, units are ppbv for $NO_2$ and $O_3$ and $\mu g \cdot m^{-3}$ for $PM_{2.5}$.

| | | AirNow + NAPS Combined | AQS + NAPS Combined | | | |
|---|---|---|---|---|---|---|
| Variable | Statistic | 2021/22 | 2013 | 2014 | 2015 | 2016 |
| $NO_2$ | N | 2,509,716 | 3,776,292 | 3,856,650 | 3,933,509 | 3,910,073 |
| | Obs Mean | 6.57 | 7.90 | 7.74 | 7.60 | 7.18 |
| | Model Mean | 5.30 | 8.80 | 8.53 | 8.32 | 7.74 |
| | MB | -1.27 | 0.90 | 0.78 | 0.73 | 0.57 |
| | NMB | -0.19 | 0.11 | 0.10 | 0.10 | 0.08 |
| | RMSE | 6.09 | 7.21 | 7.11 | 7.00 | 6.66 |
| | CRMSE | 5.96 | 7.15 | 7.07 | 6.96 | 6.63 |
| | NMAE | 0.56 | 0.58 | 0.58 | 0.58 | 0.58 |
| | FAC~~a~~2 | 0.52 | 0.59 | 0.58 | 0.58 | 0.57 |
| | R | 0.65 | 0.68 | 0.68 | 0.68 | 0.68 |
| | NSD | 0.84 | 1.05 | 1.05 | 1.06 | 1.05 |
| | Obs SD | 7.61 | 8.63 | 8.59 | 8.40 | 8.08 |
| | Model SD | 6.37 | 9.10 | 9.00 | 8.93 | 8.48 |
| $O_3$ | N | 6,175,254 | 7,468,605 | 7,617,067 | 7,472,330 | 7,871,784 |
| | Obs Mean | 29.64 | 29.50 | 29.14 | 28.95 | 29.42 |
| | Model Mean | 27.60 | 25.87 | 26.03 | 26.34 | 26.68 |
| | MB | -2.04 | -3.62 | -3.11 | -2.61 | -2.75 |
| | NMB | -0.07 | -0.12 | -0.11 | -0.09 | -0.09 |
| | RMSE | 10.77 | 12.15 | 11.66 | 11.60 | 11.35 |
| | CRMSE | 10.57 | 11.59 | 11.24 | 11.30 | 11.01 |
| | NMAE | 0.28 | 0.31 | 0.31 | 0.30 | 0.29 |
| | FAC~~a~~2 | 0.83 | 0.78 | 0.79 | 0.79 | 0.80 |
| | R | 0.72 | 0.71 | 0.71 | 0.71 | 0.72 |
| | NSD | 0.88 | 0.99 | 0.99 | 0.98 | 0.97 |
| | Obs SD | 14.76 | 15.34 | 14.97 | 15.03 | 14.89 |
| | Model SD | 13.06 | 15.18 | 14.76 | 14.80 | 14.51 |
| $PM_{2.5}$ | N | 6,374,875 | 6,280,280 | 6,635,531 | 6,471,206 | 6,471,336 |
| | Obs Mean | 8.02 | 8.67 | 8.34 | 8.37 | 7.44 |
| | Model Mean | 5.51 | 8.17 | 7.87 | 7.59 | 6.98 |
| | MB | -2.51 | -0.50 | -0.46 | -0.78 | -0.46 |
| | NMB | -0.31 | -0.06 | -0.06 | -0.09 | -0.06 |
| | RMSE | 9.73 | 11.62 | 11.33 | 11.39 | 13.43 |
| | CRMSE | 9.41 | 11.61 | 11.32 | 11.36 | 13.42 |
| | NMAE | 0.66 | 0.71 | 0.72 | 0.71 | 0.73 |
| | FAC~~a~~2 | 0.46 | 0.50 | 0.49 | 0.49 | 0.48 |
| | R | 0.24 | 0.29 | 0.27 | 0.26 | 0.17 |
| | NSD | 0.69 | 1.36 | 1.35 | 1.13 | 0.78 |
| | Obs SD | 8.78 | 8.11 | 7.83 | 8.73 | 11.54 |
| | Model SD | 6.06 | 11.02 | 10.60 | 9.89 | 9.05 |

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Table 4. Summary table of all-station annual statistics for the RAQDPS-OP023 for nine other ambient gas-phase chemistry measurements (NO, $NO_x$, CO, $HNO_3$, $NH_3$, $SO_2$, ETHE, ISOP, HCHO) for 2013–2016. For dimensional statistics, units are ppmv for CO, $\mu g \cdot m^{-3}$ for $NH_3$, and ppbv for other species. Sample duration used for all networks is hourly for NO, $NO_x$, CO, ETHE, and ISOP, daily for HCHO, weekly for $HNO_3$ and $SO_2$, and biweekly for $NH_3$.

| Variable | Statistic | 2013 | 2014 | 2015 | 2016 |
|---|---|---|---|---|---|
| NO | N | 3,562,348 | 3,721,353 | 3,680,583 | 3,564,269 |
| | Obs Mean | 4.09 | 3.89 | 4.01 | 3.61 |
| | Model Mean | 5.39 | 4.88 | 4.68 | 3.93 |
| | MB | 1.30 | 0.99 | 0.68 | 0.32 |
| | NMB | 0.32 | 0.25 | 0.17 | 0.09 |
| | RMSE | 16.44 | 15.21 | 14.65 | 12.90 |
| | CRMSE | 16.39 | 15.18 | 14.63 | 12.90 |
| | NMAE | 1.34 | 1.29 | 1.26 | 1.21 |
| | FAC2 | 0.28 | 0.29 | 0.28 | 0.28 |
| | R | 0.41 | 0.42 | 0.41 | 0.41 |
| | NSD | 1.23 | 1.21 | 1.13 | 1.07 |
| | Obs SD | 13.36 | 12.61 | 12.60 | 11.46 |
| | Model SD | 16.49 | 15.30 | 14.20 | 12.26 |
| $NO_x$ | N | 3,408,562 | 3,507,613 | 3,246,050 | 3,476,220 |
| | Obs Mean | 12.03 | 11.72 | 11.86 | 10.86 |
| | Model Mean | 14.38 | 13.40 | 12.69 | 11.67 |
| | MB | 2.35 | 1.68 | 0.83 | 0.82 |
| | NMB | 0.20 | 0.14 | 0.07 | 0.08 |
| | RMSE | 21.12 | 19.72 | 18.58 | 17.13 |
| | CRMSE | 20.98 | 19.65 | 18.56 | 17.11 |
| | NMAE | 0.79 | 0.77 | 0.73 | 0.74 |
| | FAC2 | 0.53 | 0.53 | 0.53 | 0.53 |
| | R | 0.54 | 0.54 | 0.55 | 0.55 |
| | NSD | 1.19 | 1.15 | 1.04 | 1.05 |
| | Obs SD | 19.81 | 19.00 | 19.16 | 17.60 |
| | Model SD | 23.60 | 21.84 | 19.91 | 18.54 |
| $HNO_3$ | N | 4,766 | 4,946 | 4,874 | 4,788 |
| | Obs Mean | 0.25 | 0.24 | 0.23 | 0.22 |
| | Model Mean | 0.33 | 0.32 | 0.29 | 0.25 |
| | MB | 0.08 | 0.08 | 0.06 | 0.04 |
| | NMB | 0.33 | 0.32 | 0.25 | 0.16 |
| | RMSE | 0.19 | 0.19 | 0.18 | 0.14 |
| | CRMSE | 0.18 | 0.18 | 0.17 | 0.14 |
| | NMAE | 0.52 | 0.53 | 0.48 | 0.42 |
| | FAC2 | 0.76 | 0.75 | 0.78 | 0.81 |
| | R | 0.69 | 0.64 | 0.66 | 0.69 |
| | NSD | 1.28 | 1.24 | 1.30 | 1.12 |
| | Obs SD | 0.19 | 0.18 | 0.17 | 0.16 |
| | Model SD | 0.24 | 0.22 | 0.22 | 0.18 |
| $NH_3$ | N | 1,705 | 1,731 | 2,417 | 2,464 |
| | Obs Mean | 1.54 | 1.71 | 1.62 | 1.73 |
| | Model Mean | 0.87 | 0.91 | 0.94 | 1.00 |
| | MB | -0.67 | -0.81 | -0.68 | -0.73 |
| | NMB | -0.44 | -0.47 | -0.42 | -0.42 |
| | RMSE | 2.54 | 2.86 | 3.10 | 2.66 |
| | CRMSE | 2.45 | 2.75 | 3.02 | 2.56 |
| | NMAE | 0.60 | 0.62 | 0.60 | 0.57 |

| | | | | | |
|---|---|---|---|---|---|
| | FAC2 | 0.50 | 0.47 | 0.54 | 0.57 |
| | R | 0.34 | 0.51 | 0.44 | 0.51 |
| | NSD | 0.44 | 0.41 | 0.36 | 0.44 |
| | Obs SD | 2.59 | 3.17 | 3.36 | 2.97 |
| | Model SD | 1.13 | 1.30 | 1.21 | 1.29 |
| $SO_2$ | N | 27,172 | 28,170 | 28,311 | 27,279 |
| | Obs Mean | 1.23 | 1.18 | 0.99 | 0.83 |
| | Model Mean | 1.96 | 1.78 | 1.42 | 1.22 |
| | MB | 0.74 | 0.60 | 0.43 | 0.39 |
| | NMB | 0.60 | 0.51 | 0.43 | 0.47 |
| | RMSE | 2.60 | 2.36 | 1.97 | 1.98 |
| | CRMSE | 2.49 | 2.28 | 1.92 | 1.94 |
| | NMAE | 1.11 | 1.00 | 0.99 | 1.10 |
| | FAC2 | 0.45 | 0.49 | 0.47 | 0.45 |
| | R | 0.35 | 0.39 | 0.40 | 0.38 |
| | NSD | 1.43 | 1.35 | 1.29 | 1.37 |
| | Obs SD | 1.75 | 1.72 | 1.51 | 1.43 |
| | Model SD | 2.50 | 2.31 | 1.94 | 1.97 |
| CO | N | 2,413,310 | 2,313,412 | 2,263,157 | 2,207,060 |
| | Obs Mean | 0.27 | 0.27 | 0.27 | 0.28 |
| | Model Mean | 0.31 | 0.30 | 0.29 | 0.27 |
| | MB | 0.03 | 0.03 | 0.01 | -0.01 |
| | NMB | 0.12 | 0.09 | 0.05 | -0.05 |
| | RMSE | 0.29 | 0.27 | 0.25 | 0.56 |
| | CRMSE | 0.28 | 0.26 | 0.25 | 0.56 |
| | NMAE | 0.61 | 0.57 | 0.55 | 0.55 |
| | FAC2 | 0.69 | 0.72 | 0.73 | 0.74 |
| | R | 0.42 | 0.43 | 0.43 | 0.14 |
| | NSD | 1.04 | 1.09 | 1.04 | 0.38 |
| | Obs SD | 0.26 | 0.24 | 0.23 | 0.55 |
| | Model SD | 0.27 | 0.26 | 0.24 | 0.21 |
| ETHE | N | 95,968 | 94,497 | 73,794 | 57,832 |
| | Obs Mean | 1.05 | 1.03 | 1.06 | 0.95 |
| | Model Mean | 1.77 | 1.80 | 2.01 | 2.08 |
| | MB | 0.72 | 0.77 | 0.95 | 1.13 |
| | NMB | 0.68 | 0.75 | 0.89 | 1.19 |
| | RMSE | 2.25 | 2.34 | 2.55 | 2.82 |
| | CRMSE | 2.13 | 2.21 | 2.36 | 2.58 |
| | NMAE | 1.24 | 1.27 | 1.42 | 1.66 |
| | FAC2 | 0.41 | 0.42 | 0.39 | 0.35 |
| | R | 0.32 | 0.26 | 0.20 | 0.18 |
| | NSD | 1.05 | 1.08 | 1.12 | 1.13 |
| | Obs SD | 1.78 | 1.75 | 1.75 | 1.89 |
| | Model SD | 1.87 | 1.89 | 1.97 | 2.13 |
| HCHO | N | 4,909 | 5,063 | 4,889 | 4,826 |
| | Obs Mean | 2.38 | 2.07 | 2.27 | 2.37 |
| | Model Mean | 2.47 | 2.32 | 2.43 | 2.56 |
| | MB | 0.09 | 0.25 | 0.16 | 0.19 |
| | NMB | 0.04 | 0.12 | 0.07 | 0.08 |
| | RMSE | 3.71 | 2.13 | 2.09 | 2.34 |
| | CRMSE | 3.71 | 2.11 | 2.08 | 2.33 |
| | NMAE | 0.66 | 0.61 | 0.56 | 0.61 |

| | | | | | |
|---|---|---|---|---|---|
| | FAC2 | 0.65 | 0.66 | 0.71 | 0.66 |
| | R | 0.26 | 0.43 | 0.46 | 0.41 |
| | NSD | 0.64 | 1.24 | 1.23 | 1.31 |
| | Obs SD | 3.59 | 1.75 | 1.78 | 1.83 |
| | Model SD | 2.28 | 2.17 | 2.19 | 2.39 |
| ISOP | N | 79,679 | 94,818 | 66,013 | 42,845 |
| | Obs Mean | 0.12 | 0.11 | 0.15 | 0.13 |
| | Model Mean | 0.39 | 0.38 | 0.56 | 0.48 |
| | MB | 0.27 | 0.27 | 0.41 | 0.35 |
| | NMB | 2.27 | 2.36 | 2.71 | 2.60 |
| | RMSE | 0.77 | 0.78 | 0.99 | 0.99 |
| | CRMSE | 0.73 | 0.73 | 0.90 | 0.93 |
| | NMAE | 2.81 | 2.85 | 3.12 | 3.20 |
| | FAC2 | 0.18 | 0.19 | 0.18 | 0.19 |
| | R | 0.45 | 0.46 | 0.50 | 0.37 |
| | NSD | 3.37 | 3.25 | 3.13 | 3.69 |
| | Obs SD | 0.24 | 0.25 | 0.32 | 0.27 |
| | Model SD | 0.80 | 0.82 | 1.02 | 0.99 |

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Table 5. Summary table of all-station annual statistics for the RAQDPS-OP023 for daily gravimetric and speciated $PM_{2.5}$ measurements for 2013–2016. For dimensional statistics, units are $\mu g \cdot m^{-3}$.

| Variable | Statistic | 2013 | 2014 | 2015 | 2016 |
|---|---|---|---|---|---|
| $SO_4$ | N | 32,101 | 32,027 | 30,103 | 28,972 |
| | Obs Mean | 1.17 | 1.15 | 0.97 | 0.75 |
| | Model Mean | 0.84 | 0.82 | 0.73 | 0.63 |
| | MB | -0.32 | -0.33 | -0.25 | -0.12 |
| | NMB | -0.28 | -0.29 | -0.25 | -0.16 |
| | RMSE | 0.82 | 0.83 | 0.73 | 0.58 |
| | CRMSE | 0.75 | 0.76 | 0.69 | 0.57 |
| | NMAE | 0.46 | 0.46 | 0.47 | 0.49 |
| | FAC~~ae~~2 | 0.61 | 0.64 | 0.65 | 0.64 |
| | R | 0.71 | 0.69 | 0.65 | 0.61 |
| | NSD | 0.83 | 0.75 | 0.74 | 0.84 |
| | Obs SD | 1.05 | 1.05 | 0.90 | 0.69 |
| | Model SD | 0.87 | 0.79 | 0.67 | 0.57 |
| $NO_3$ | N | 31,992 | 31,983 | 30,046 | 29,029 |
| | Obs Mean | 0.75 | 0.79 | 0.68 | 0.54 |
| | Model Mean | 0.67 | 0.61 | 0.58 | 0.49 |
| | MB | -0.08 | -0.19 | -0.10 | -0.05 |
| | NMB | -0.11 | -0.23 | -0.15 | -0.09 |
| | RMSE | 1.40 | 1.25 | 1.00 | 0.91 |
| | CRMSE | 1.40 | 1.23 | 1.00 | 0.91 |
| | NMAE | 0.67 | 0.63 | 0.64 | 0.71 |
| | FAC~~ae~~2 | 0.37 | 0.35 | 0.34 | 0.32 |
| | R | 0.57 | 0.66 | 0.70 | 0.66 |
| | NSD | 0.80 | 0.71 | 0.85 | 0.88 |
| | Obs SD | 1.64 | 1.63 | 1.36 | 1.17 |
| | Model SD | 1.32 | 1.16 | 1.16 | 1.02 |
| $NH_4$ | N | 14,828 | 14,711 | 13,283 | 12,757 |
| | Obs Mean | 0.64 | 0.69 | 0.54 | 0.33 |
| | Model Mean | 0.78 | 0.73 | 0.68 | 0.60 |
| | MB | 0.14 | 0.04 | 0.14 | 0.28 |
| | NMB | 0.22 | 0.06 | 0.26 | 0.85 |
| | RMSE | 0.75 | 0.69 | 0.61 | 0.60 |
| | CRMSE | 0.73 | 0.69 | 0.60 | 0.54 |
| | NMAE | 0.66 | 0.59 | 0.72 | 1.23 |
| | FAC~~ae~~2 | 0.58 | 0.60 | 0.52 | 0.34 |
| | R | 0.52 | 0.57 | 0.58 | 0.51 |
| | NSD | 0.81 | 0.71 | 0.82 | 0.92 |
| | Obs SD | 0.81 | 0.83 | 0.71 | 0.56 |
| | Model SD | 0.66 | 0.59 | 0.58 | 0.52 |
| EC | N | 30,770 | 31,862 | 28,866 | 28,114 |
| | Obs Mean | 0.34 | 0.33 | 0.32 | 0.30 |
| | Model Mean | 0.52 | 0.50 | 0.47 | 0.40 |
| | MB | 0.18 | 0.17 | 0.15 | 0.09 |
| | NMB | 0.53 | 0.51 | 0.48 | 0.30 |
| | RMSE | 0.74 | 0.66 | 0.67 | 0.55 |
| | CRMSE | 0.71 | 0.64 | 0.65 | 0.55 |
| | NMAE | 0.91 | 0.91 | 0.90 | 0.77 |
| | FAC~~ae~~2 | 0.57 | 0.58 | 0.56 | 0.56 |
| | R | 0.63 | 0.60 | 0.58 | 0.60 |
| | NSD | 2.11 | 1.94 | 1.91 | 1.68 |
| | Obs SD | 0.43 | 0.41 | 0.42 | 0.41 |
| | Model SD | 0.90 | 0.80 | 0.80 | 0.69 |

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|---|---|---|---|---|---|
| TOM | N | 30,609 | 31,747 | 28,688 | 27,508 |
| | Obs Mean | 2.48 | 2.43 | 2.76 | 2.46 |
| | Model Mean | 2.74 | 2.54 | 2.61 | 2.41 |
| | MB | 0.26 | 0.11 | -0.15 | -0.05 |
| | NMB | 0.11 | 0.04 | -0.05 | -0.02 |
| | RMSE | 6.49 | 3.54 | 6.24 | 10.33 |
| | CRMSE | 6.48 | 3.54 | 6.24 | 10.33 |
| | NMAE | 0.75 | 0.68 | 0.72 | 0.74 |
| | FAC~~a~~2 | 0.54 | 0.55 | 0.51 | 0.50 |
| | R | 0.36 | 0.45 | 0.25 | 0.10 |
| | NSD | 2.52 | 1.44 | 1.54 | 0.54 |
| | Obs SD | 2.76 | 2.65 | 3.86 | 9.50 |
| | Model SD | 6.95 | 3.82 | 5.95 | 5.14 |
| CM | N | 31,429 | 31,675 | 29,871 | 28,902 |
| | Obs Mean | 0.58 | 0.63 | 0.61 | 0.56 |
| | Model Mean | 0.83 | 0.85 | 0.83 | 0.83 |
| | MB | 0.25 | 0.22 | 0.22 | 0.27 |
| | NMB | 0.42 | 0.35 | 0.36 | 0.48 |
| | RMSE | 1.60 | 1.67 | 1.53 | 1.46 |
| | CRMSE | 1.58 | 1.65 | 1.52 | 1.43 |
| | NMAE | 1.40 | 1.37 | 1.30 | 1.36 |
| | FAC~~a~~2 | 0.31 | 0.30 | 0.32 | 0.32 |
| | R | 0.12 | 0.13 | 0.17 | 0.22 |
| | NSD | 1.30 | 1.19 | 1.27 | 1.59 |
| | Obs SD | 1.02 | 1.14 | 1.03 | 0.85 |
| | Model SD | 1.33 | 1.36 | 1.30 | 1.35 |
| SS | N | 31,860 | 31,852 | 30,185 | 29,104 |
| | Obs Mean | 0.26 | 0.25 | 0.29 | 0.23 |
| | Model Mean | 0.45 | 0.44 | 0.49 | 0.51 |
| | MB | 0.19 | 0.19 | 0.20 | 0.29 |
| | NMB | 0.74 | 0.78 | 0.71 | 1.25 |
| | RMSE | 1.02 | 1.12 | 1.09 | 1.16 |
| | CRMSE | 1.00 | 1.10 | 1.07 | 1.12 |
| | NMAE | 1.36 | 1.42 | 1.35 | 1.70 |
| | FAC~~a~~2 | 0.31 | 0.32 | 0.33 | 0.33 |
| | R | 0.62 | 0.58 | 0.61 | 0.60 |
| | NSD | 2.55 | 2.74 | 2.66 | 3.03 |
| | Obs SD | 0.48 | 0.48 | 0.48 | 0.44 |
| | Model SD | 1.23 | 1.31 | 1.29 | 1.33 |
| $PM_{2.5}$ | N | 104,944 | 104,095 | 111,150 | 102,120 |
| | Obs Mean | 8.19 | 8.12 | 7.94 | 7.24 |
| | Model Mean | 8.72 | 8.10 | 8.10 | 7.47 |
| | MB | 0.54 | -0.02 | 0.16 | 0.23 |
| | NMB | 0.07 | 0.00 | 0.02 | 0.03 |
| | RMSE | 8.24 | 7.86 | 7.65 | 7.15 |
| | CRMSE | 8.22 | 7.86 | 7.65 | 7.14 |
| | NMAE | 0.54 | 0.51 | 0.53 | 0.54 |
| | FAC~~a~~2 | 0.70 | 0.70 | 0.70 | 0.68 |
| | R | 0.47 | 0.44 | 0.45 | 0.43 |
| | NSD | 1.61 | 1.50 | 1.41 | 1.53 |
| | Obs SD | 5.69 | 5.67 | 5.81 | 5.04 |
| | Model SD | 9.19 | 8.49 | 8.20 | 7.70 |

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Table 6. Summary table of all-station annual statistics for the RAQDPS-OP023 for weekly precipitation-chemistry measurements for 2013–2016. For dimensional statistics, units are $mg \cdot L^{-1}$ for concentration in precipitation, $mm \cdot week^{-1}$ for precipitation, and $mg \cdot m^{-2} \cdot week^{-1}$ for wet deposition. Note that daily CAPMoN measurements have been aggregated to weekly values for consistency with NADP measurements.

| Variable | Statistic | 2013 | 2014 | 2015 | 2016 |
|---|---|---|---|---|---|
| $SO_4^{=}$ conc | N | 9,115 | 8,997 | 9,759 | 9,193 |
| | Obs Mean | 0.81 | 0.74 | 0.65 | 0.62 |
| | Model Mean | 0.82 | 0.84 | 0.74 | 0.61 |
| | MB | 0.02 | 0.10 | 0.09 | -0.01 |
| | NMB | 0.02 | 0.14 | 0.14 | -0.01 |
| | RMSE | 0.91 | 0.91 | 0.87 | 1.19 |
| | CRMSE | 0.91 | 0.90 | 0.87 | 1.19 |
| | NMAE | 0.59 | 0.65 | 0.68 | 0.61 |
| | FAC~~a~~2 | 0.64 | 0.65 | 0.63 | 0.65 |
| | R | 0.41 | 0.38 | 0.37 | 0.25 |
| | NSD | 0.89 | 1.00 | 1.09 | 0.50 |
| | Obs SD | 0.88 | 0.81 | 0.74 | 1.19 |
| | Model SD | 0.78 | 0.81 | 0.80 | 0.60 |
| $NO_3^{-}$ conc | N | 9,115 | 8,995 | 9,752 | 9,190 |
| | Obs Mean | 0.96 | 0.93 | 0.91 | 0.90 |
| | Model Mean | 0.94 | 0.91 | 0.90 | 0.78 |
| | MB | -0.02 | -0.01 | -0.01 | -0.12 |
| | NMB | -0.02 | -0.02 | -0.01 | -0.14 |
| | RMSE | 0.97 | 0.93 | 1.02 | 0.94 |
| | CRMSE | 0.97 | 0.93 | 1.02 | 0.93 |
| | NMAE | 0.55 | 0.54 | 0.57 | 0.52 |
| | FAC~~a~~2 | 0.69 | 0.70 | 0.68 | 0.69 |
| | R | 0.49 | 0.52 | 0.48 | 0.46 |
| | NSD | 1.06 | 1.03 | 1.06 | 0.80 |
| | Obs SD | 0.92 | 0.94 | 0.97 | 0.98 |
| | Model SD | 0.98 | 0.97 | 1.03 | 0.78 |
| $NH_4^{+}$ conc | N | 9,103 | 8,980 | 9,727 | 9,181 |
| | Obs Mean | 0.37 | 0.36 | 0.39 | 0.39 |
| | Model Mean | 0.46 | 0.46 | 0.54 | 0.52 |
| | MB | 0.09 | 0.09 | 0.16 | 0.13 |
| | NMB | 0.23 | 0.25 | 0.41 | 0.35 |
| | RMSE | 0.65 | 0.65 | 0.91 | 0.79 |
| | CRMSE | 0.65 | 0.65 | 0.90 | 0.78 |
| | NMAE | 0.78 | 0.80 | 0.95 | 0.89 |
| | FAC~~a~~2 | 0.60 | 0.59 | 0.57 | 0.58 |
| | R | 0.47 | 0.41 | 0.38 | 0.39 |
| | NSD | 1.62 | 1.54 | 1.98 | 1.75 |
| | Obs SD | 0.45 | 0.44 | 0.49 | 0.47 |
| | Model SD | 0.72 | 0.68 | 0.96 | 0.83 |
| PR | N | 13,732 | 13,874 | 14,050 | 13,403 |
| | Obs Mean | 19.48 | 20.27 | 19.77 | 19.38 |
| | Model Mean | 21.72 | 21.64 | 19.63 | 20.11 |
| | MB | 2.25 | 1.37 | -0.14 | 0.73 |
| | NMB | 0.12 | 0.07 | -0.01 | 0.04 |
| | RMSE | 19.62 | 20.63 | 17.60 | 18.91 |
| | CRMSE | 19.49 | 20.58 | 17.60 | 18.90 |
| | NMAE | 0.54 | 0.53 | 0.49 | 0.52 |
| | FAC~~a~~2 | 0.57 | 0.57 | 0.57 | 0.56 |
| | R | 0.72 | 0.71 | 0.78 | 0.75 |
| | NSD | 1.07 | 1.06 | 0.99 | 1.03 |

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|---|---|---|---|---|---|
| | Obs SD | 25.24 | 26.33 | 26.65 | 26.40 |
| | Model SD | 26.91 | 27.90 | 26.52 | 27.18 |
| $SO_4^{=}$ dep | N | 9,115 | 8,997 | 9,759 | 9,193 |
| | Obs Mean | 15.65 | 14.41 | 11.41 | 10.71 |
| | Model Mean | 15.91 | 15.18 | 11.36 | 10.08 |
| | MB | 0.26 | 0.78 | -0.06 | -0.63 |
| | NMB | 0.02 | 0.05 | 0.00 | -0.06 |
| | RMSE | 16.29 | 16.85 | 12.91 | 12.58 |
| | CRMSE | 16.29 | 16.83 | 12.91 | 12.56 |
| | NMAE | 0.61 | 0.65 | 0.65 | 0.62 |
| | FAC~~ac~~2 | 0.55 | 0.56 | 0.53 | 0.56 |
| | R | 0.58 | 0.49 | 0.53 | 0.58 |
| | NSD | 0.90 | 0.85 | 0.83 | 0.76 |
| | Obs SD | 18.55 | 17.85 | 14.33 | 15.04 |
| | Model SD | 16.72 | 15.09 | 11.84 | 11.38 |
| $NO_3^{-}$ dep | N | 9,115 | 8,995 | 9,752 | 9,190 |
| | Obs Mean | 16.48 | 15.64 | 13.71 | 14.12 |
| | Model Mean | 16.16 | 14.76 | 12.35 | 11.65 |
| | MB | -0.33 | -0.88 | -1.35 | -2.47 |
| | NMB | -0.02 | -0.06 | -0.10 | -0.18 |
| | RMSE | 14.46 | 16.49 | 12.16 | 13.00 |
| | CRMSE | 14.46 | 16.47 | 12.09 | 12.76 |
| | NMAE | 0.54 | 0.54 | 0.55 | 0.52 |
| | FAC~~ac~~2 | 0.63 | 0.64 | 0.62 | 0.62 |
| | R | 0.59 | 0.46 | 0.56 | 0.56 |
| | NSD | 0.89 | 0.72 | 0.85 | 0.72 |
| | Obs SD | 16.84 | 17.81 | 13.76 | 15.13 |
| | Model SD | 14.97 | 12.86 | 11.68 | 10.93 |
| $NH_4^{+}$ dep | N | 9,103 | 8,980 | 9,727 | 9,181 |
| | Obs Mean | 6.60 | 6.26 | 5.85 | 6.14 |
| | Model Mean | 7.65 | 7.28 | 6.75 | 6.94 |
| | MB | 1.05 | 1.02 | 0.90 | 0.80 |
| | NMB | 0.16 | 0.16 | 0.15 | 0.13 |
| | RMSE | 8.88 | 9.54 | 7.78 | 8.33 |
| | CRMSE | 8.82 | 9.48 | 7.73 | 8.29 |
| | NMAE | 0.69 | 0.72 | 0.72 | 0.68 |
| | FAC~~ac~~2 | 0.56 | 0.57 | 0.55 | 0.57 |
| | R | 0.59 | 0.47 | 0.55 | 0.56 |
| | NSD | 1.08 | 1.01 | 1.05 | 1.06 |
| | Obs SD | 9.32 | 9.20 | 7.90 | 8.53 |
| | Model SD | 10.06 | 9.29 | 8.33 | 9.03 |

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Table 7. Comparison of RAQDPS-OP023 and RAQDPS-FW023 all-station seasonal scores for predicted hourly surface $NO_2$, $O_3$, and $PM_{2.5}$ abundances for 2021/22. For dimensional statistics, units are ppbv for $NO_2$ and $O_3$ and $\mu g \cdot m^{-3}$ for $PM_{2.5}$.

| | | Winter | | Spring | | Summer | | Autumn | |
|---|---|---|---|---|---|---|---|---|---|
| Variable | Statistic | OP023 | FW023 | OP023 | FW023 | OP023 | FW023 | OP023 | FW023 |
| $NO_2$ | N | 618,274 | 618,274 | 635,685 | 635,685 | 633,977 | 633,977 | 621,780 | 621,780 |
| | Obs Mean | 9.42 | 9.42 | 5.53 | 5.53 | 4.57 | 4.57 | 6.82 | 6.82 |
| | Model | 7.32 | 7.32 | 4.50 | 4.50 | 3.98 | 4.02 | 5.45 | 5.46 |
| | MB | -2.10 | -2.10 | -1.03 | -1.03 | -0.59 | -0.55 | -1.38 | -1.37 |
| | NMB | -0.22 | -0.22 | -0.19 | -0.19 | -0.13 | -0.12 | -0.20 | -0.20 |
| | RMSE | 7.69 | 7.69 | 5.69 | 5.69 | 4.62 | 4.64 | 6.00 | 6.00 |
| | CRMSE | 7.40 | 7.40 | 5.60 | 5.60 | 4.59 | 4.60 | 5.84 | 5.84 |
| | NMAE | 0.52 | 0.52 | 0.60 | 0.60 | 0.60 | 0.61 | 0.55 | 0.55 |
| | FAC~~ae~~2 | 0.56 | 0.56 | 0.48 | 0.48 | 0.49 | 0.50 | 0.54 | 0.54 |
| | R | 0.66 | 0.66 | 0.62 | 0.62 | 0.56 | 0.56 | 0.64 | 0.64 |
| | NSD | 0.85 | 0.85 | 0.87 | 0.87 | 0.89 | 0.90 | 0.78 | 0.79 |
| | Obs SD | 9.54 | 9.54 | 6.76 | 6.76 | 5.14 | 5.14 | 7.49 | 7.49 |
| | Model SD | 8.14 | 8.14 | 5.89 | 5.89 | 4.57 | 4.62 | 5.88 | 5.89 |
| $O_3$ | N | 1,523,37 | 1,523,37 | 1,567,17 | 1,567,17 | 1,570,21 | 1,570,21 | 1,514,48 | 1,514,48 |
| | Obs Mean | 26.37 | 26.37 | 34.78 | 34.78 | 31.34 | 31.34 | 25.83 | 25.83 |
| | Model | 27.56 | 27.57 | 30.59 | 30.61 | 27.09 | 27.93 | 25.07 | 25.31 |
| | MB | 1.19 | 1.19 | -4.19 | -4.17 | -4.25 | -3.41 | -0.76 | -0.52 |
| | NMB | 0.05 | 0.05 | -0.12 | -0.12 | -0.14 | -0.11 | -0.03 | -0.02 |
| | RMSE | 9.65 | 9.65 | 11.02 | 11.01 | 12.36 | 11.95 | 9.76 | 9.67 |
| | CRMSE | 9.57 | 9.57 | 10.19 | 10.19 | 11.60 | 11.46 | 9.73 | 9.66 |
| | NMAE | 0.28 | 0.28 | 0.25 | 0.25 | 0.31 | 0.30 | 0.28 | 0.28 |
| | FAC~~ae~~2 | 0.82 | 0.82 | 0.88 | 0.88 | 0.82 | 0.83 | 0.81 | 0.82 |
| | R | 0.68 | 0.68 | 0.66 | 0.66 | 0.75 | 0.76 | 0.75 | 0.76 |
| | NSD | 0.88 | 0.88 | 0.93 | 0.93 | 0.90 | 0.93 | 0.89 | 0.91 |
| | Obs SD | 12.50 | 12.50 | 12.82 | 12.82 | 16.94 | 16.94 | 14.42 | 14.42 |
| | Model SD | 11.06 | 11.06 | 11.92 | 11.93 | 15.32 | 15.79 | 12.88 | 13.12 |
| $PM_{2.5}$ | N | 1,574,82 | 1,574,82 | 1,616,88 | 1,616,88 | 1,622,57 | 1,622,57 | 1,560,59 | 1,560,59 |
| | Obs Mean | 8.03 | 8.03 | 6.20 | 6.20 | 10.39 | 10.39 | 7.41 | 7.41 |
| | Model | 7.00 | 7.04 | 4.56 | 4.69 | 5.06 | 10.82 | 5.47 | 7.06 |
| | MB | -1.04 | -0.99 | -1.64 | -1.52 | -5.34 | 0.42 | -1.94 | -0.35 |
| | NMB | -0.13 | -0.12 | -0.26 | -0.24 | -0.51 | 0.04 | -0.26 | -0.05 |
| | RMSE | 8.38 | 8.37 | 5.93 | 5.88 | 14.14 | 19.92 | 8.48 | 10.26 |
| | CRMSE | 8.32 | 8.31 | 5.70 | 5.68 | 13.09 | 19.92 | 8.25 | 10.25 |
| | NMAE | 0.66 | 0.66 | 0.63 | 0.62 | 0.69 | 0.65 | 0.65 | 0.64 |
| | FAC~~ae~~2 | 0.49 | 0.49 | 0.46 | 0.47 | 0.43 | 0.56 | 0.48 | 0.54 |
| | R | 0.40 | 0.41 | 0.37 | 0.38 | 0.09 | 0.55 | 0.28 | 0.49 |
| | NSD | 1.16 | 1.17 | 0.94 | 0.96 | 0.35 | 1.87 | 0.76 | 1.50 |
| | Obs SD | 7.00 | 7.00 | 5.21 | 5.21 | 12.74 | 12.74 | 7.68 | 7.68 |
| | Model SD | 8.14 | 8.17 | 4.91 | 5.01 | 4.46 | 23.79 | 5.83 | 11.54 |

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**Figures**

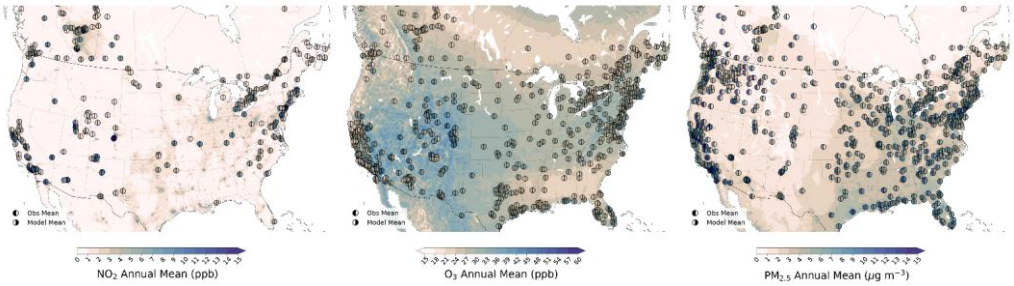


Figure 1. Spatial distribution of RAQDPS-OP023-predicted 2021/22 ~~mean~~ annual-mean abundance fields of hourly (a) $NO_2$ VMR, (b) $O_3$ VMR, and (c) $PM_{2.5}$-noSS concentration with NRT observed and predicted ~~NRT~~ station annual abundances superimposed (shown as filled-in divided circles, same colour bar). Units are ppbv, ppbv, and $\mu g \cdot m^{-3}$, respectively.

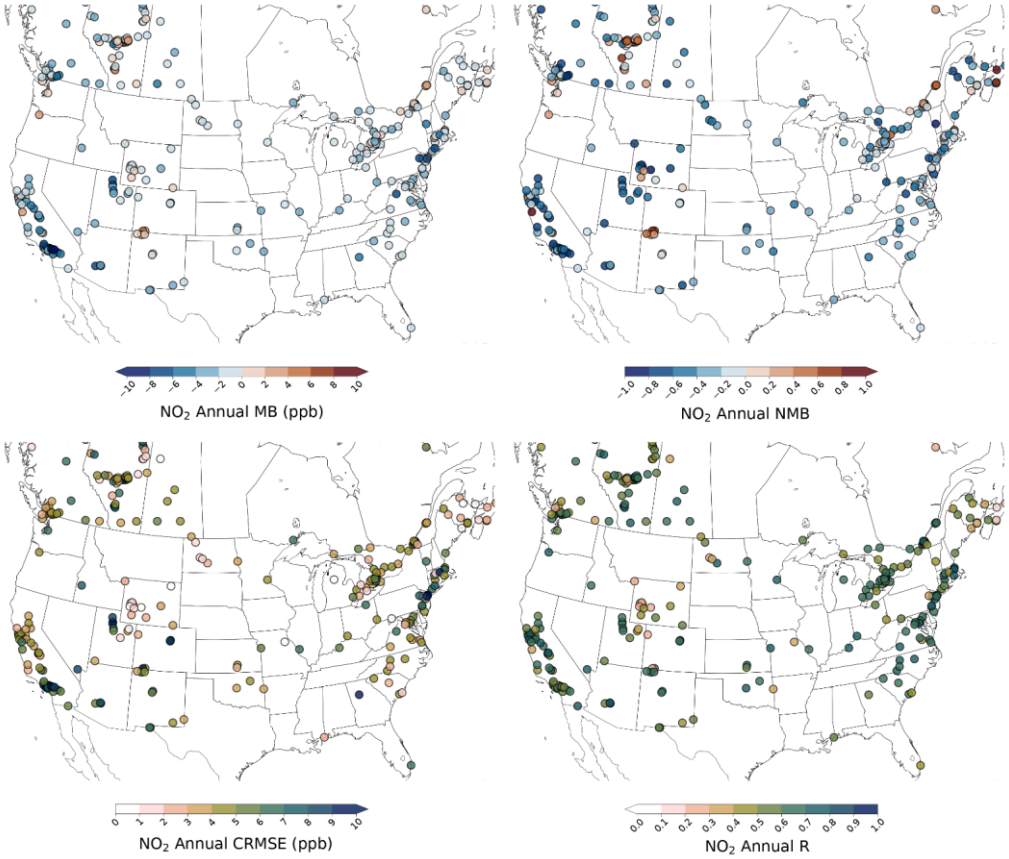


Figure 2. Spatial distribution of 2021/22 annual (a) MB, (b) NMB, (c) CRMSE, and (d) R scores for the RAQDPS-OP023 at all NRT stations for hourly $NO_2$ VMR measurements (ppbv).

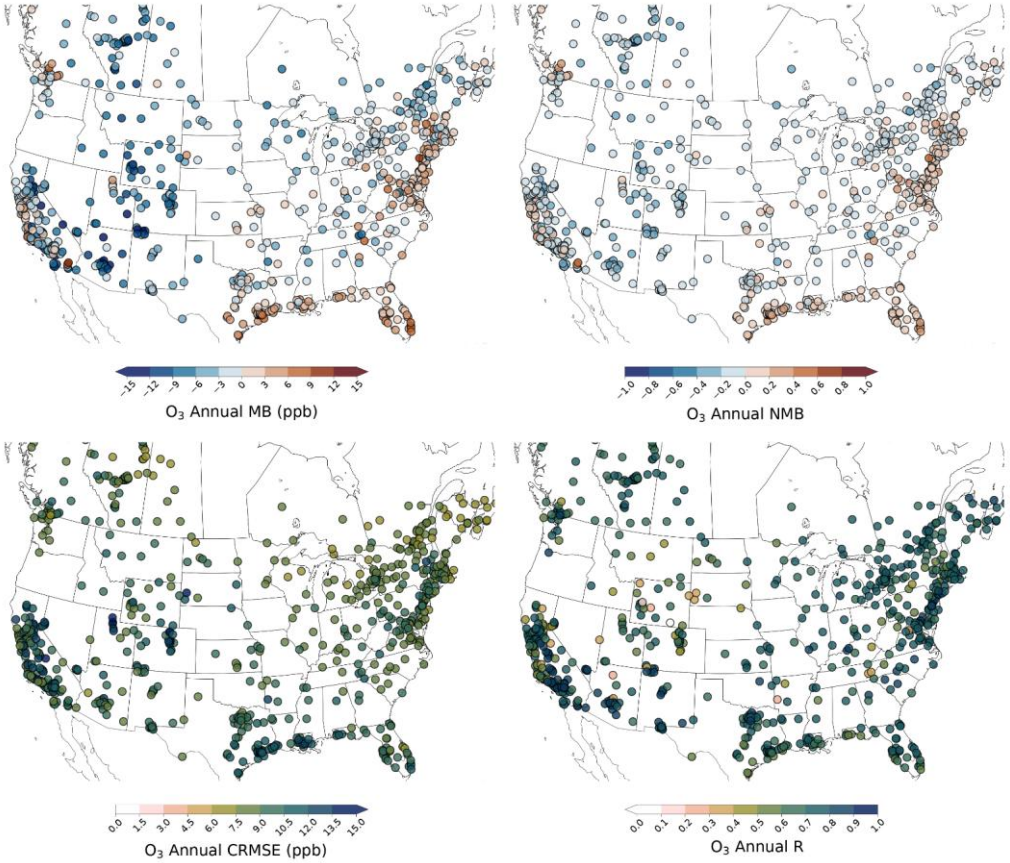


Figure 3. Spatial distribution of 2021/22 annual (a) MB, (b) NMB, (c) CRMSE, and (d) R scores for the RAQDPS-OP023 at all NRT stations for hourly $O_3$ VMR measurements (ppbv).

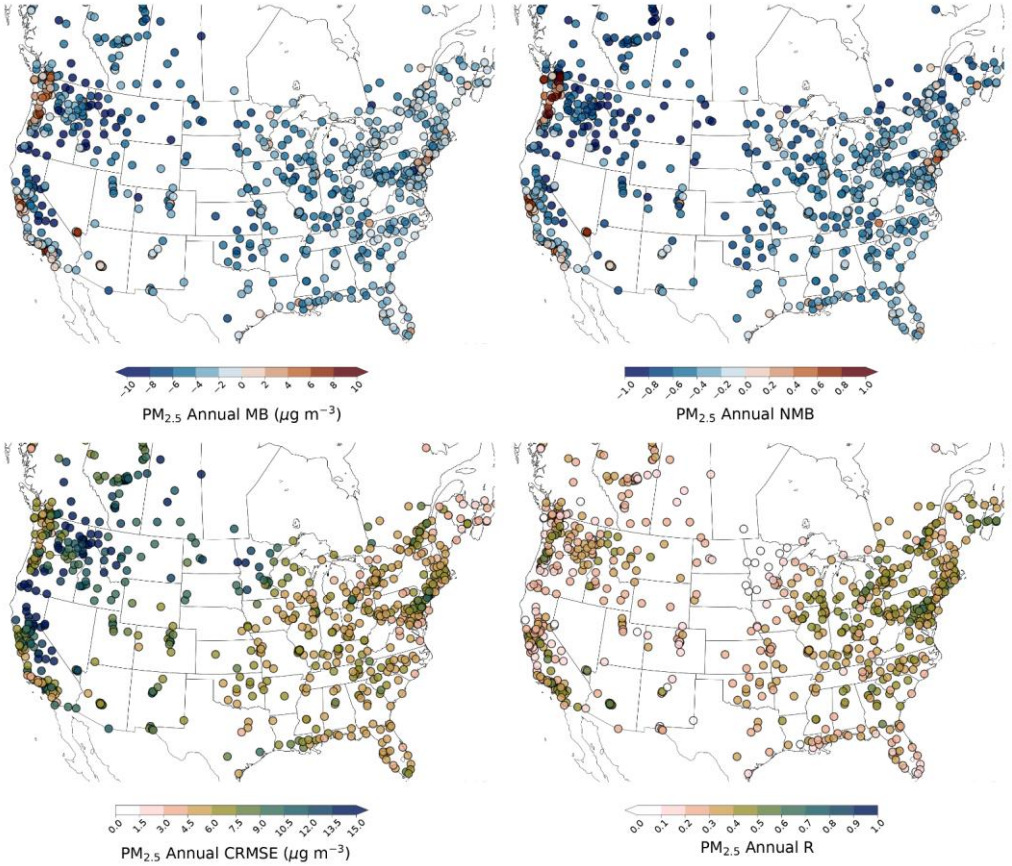


Figure 4. Spatial distribution of 2021/22 annual (a) MB, (b) NMB, (c) CRMSE, and (d) R scores for the RAQDPS-OP023 at all NRT stations for hourly $PM_{2.5}$ concentration measurements ($\mu g \cdot m^{-3}$).

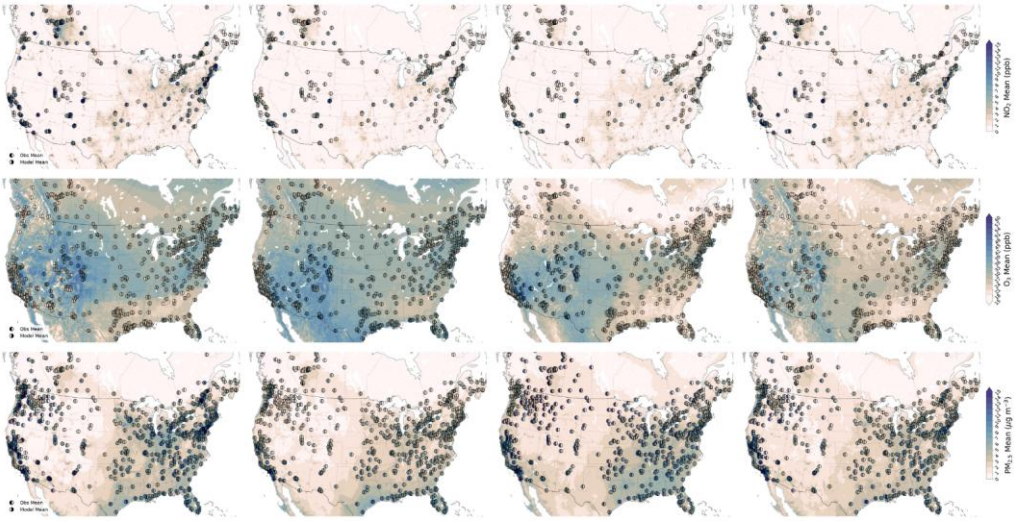

Figure 5. Spatial distribution of RAQDPS-OP023-predicted 2021/22 ~~mean~~ seasonal-mean abundance fields (from left to right: winter [DJF], spring [MAM], summer [JJA], autumn [SON]) of hourly (top row) $NO_2$ VMR, (middle row) $O_3$ VMR, and (bottom row) $PM_{2.5}$-noSS concentration with NRT observed and predicted station seasonal abundances superimposed (shown as filled-in divided circles, same colour bar). Units are ppbv, ppbv, and $\mu g \cdot m^{-3}$, respectively.

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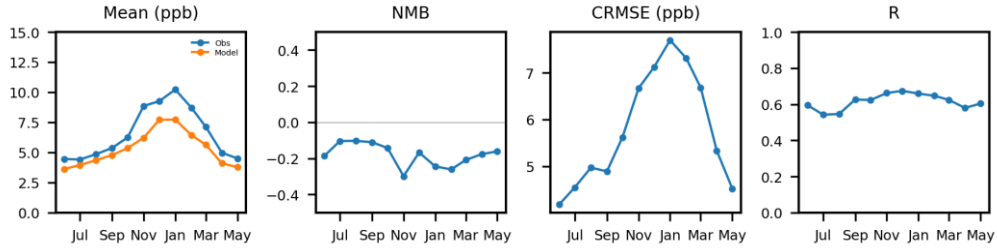


Figure 6. Time series of (a) observed and RAQDPS-OP023-predicted monthly means of hourly $NO_2$ ~~volume mixing ratio~~VMR (ppbv) and monthly (b) NMB, (c) CRMSE, and (d) R scores for 2021/22 for all NRT measurement stations.

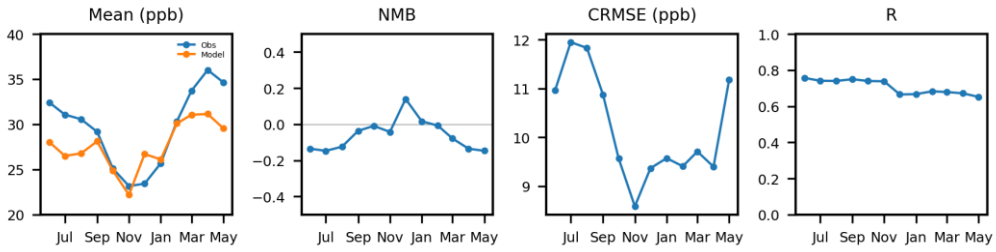


Figure 7. Time series of (a) observed and RAQDPS-OP023-predicted monthly means of hourly $O_3$ ~~volume mixing ratio~~VMR (ppbv) and monthly (b) NMB, (c) CRMSE, and (d) R scores for 2021/22 for all NRT measurement stations.

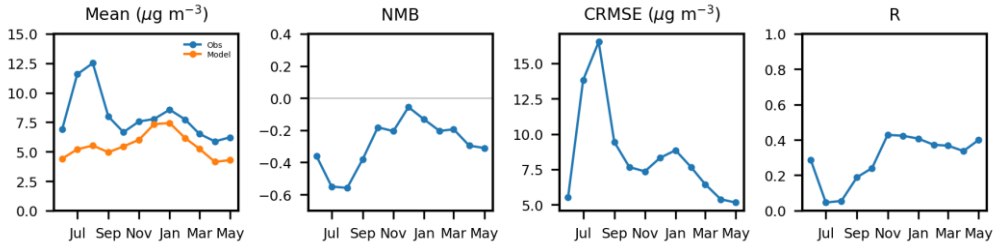


Figure 8. Time series of monthly (a) observed and RAQDPS-OP023-predicted mean hourly $PM_{2.5}$ concentration ($\mu g \cdot m^{-3}$) and (b) NMB, (c) CRMSE, and (d) R scores for 2021/22 for all NRT measurement stations.

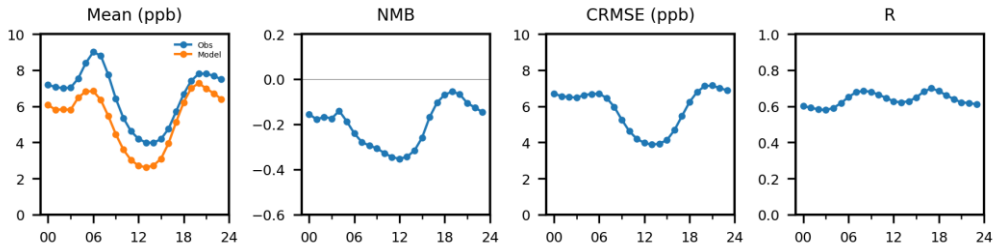


Figure 9. Time series of diurnal (a) observed and RAQDPS-OP023-predicted annual-mean hourly $NO_2$ ~~volume mixing ratio~~VMR (ppbv) and (b) NMB, (c) CRMSE, and (d) R scores for all of 2021/22 for all NRT stations. Time is in hours LT.

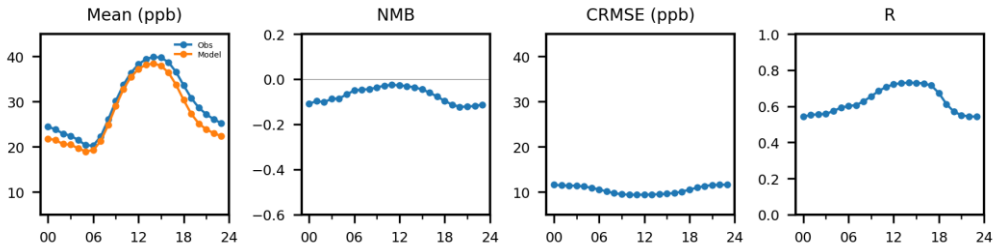


Figure 10. Time series of diurnal (a) observed and RAQDPS-OP023-predicted annual-mean hourly $O_3$ ~~volume mixing ratio~~VMR (ppbv) and (b) NMB, (c) CRMSE, and (d) R scores for all of 2021/22 for all NRT stations. Time is in hours LT.

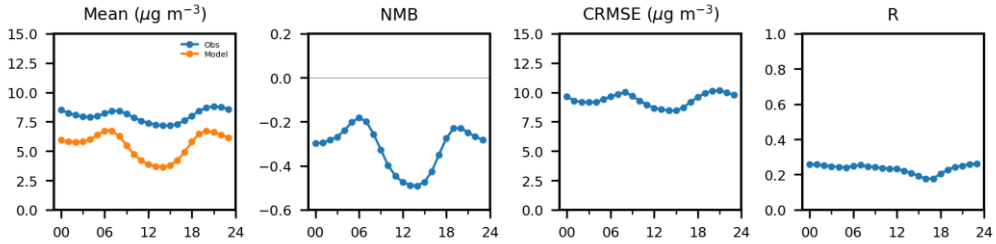


Figure 11. Time series of diurnal (a) observed and RAQDPS-OP023-predicted annual-mean hourly $PM_{2.5}$ concentration ($\mu g \cdot m^{-3}$) and (b) NMB, (c) CRMSE, and (d) R scores for all of 2021/22 for all NRT stations. Time is in hours LT.

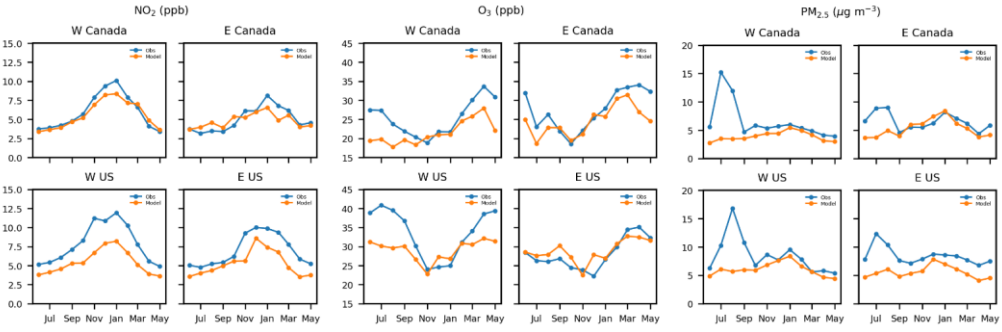


Figure 12. Time series of monthly observed and RAQDPS-OP023-predicted annual-mean hourly (left) $NO_2$ ~~volume mixing ratio~~VMR (ppbv), (centre) $O_3$ ~~volume mixing ratio~~VMR (ppbv), and (right) $PM_{2.5}$ concentration ($\mu g \cdot m^{-3}$) for all NRT stations in LT but stratified into four regions [see Fig. S7] for 2021/22. Orange curves denote predicted monthly values and blue curves denote observed monthly values.

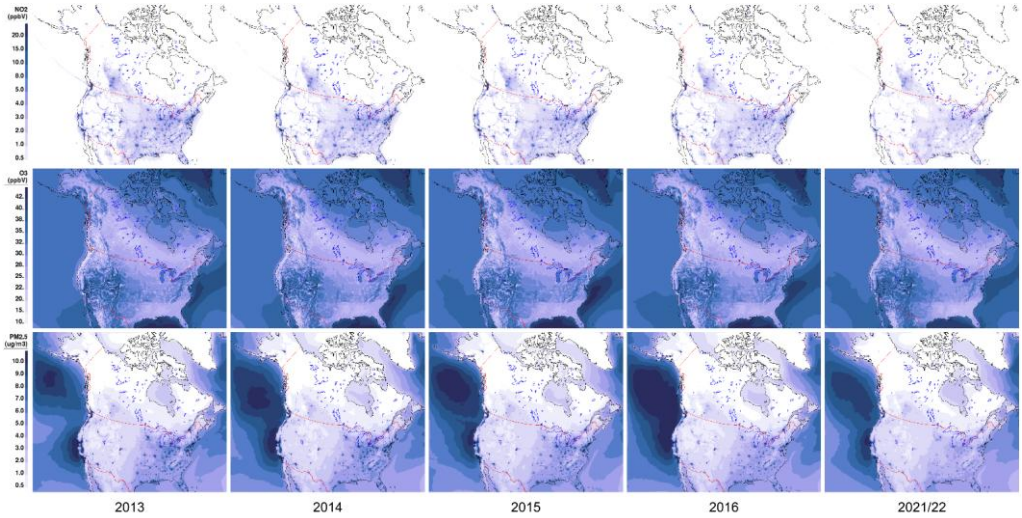


Figure 13. Spatial distribution of RAQDPS-OP023-predicted ~~mean~~ annual-mean abundances of hourly (top row) $NO_2$, (middle row) $O_3$, and (bottom row) $PM_{2.5}$ total mass for five years (from left to right, 2013–2016 and 2021/22). Units are ppbv, ppbv, and $\mu g \cdot m^{-3}$, respectively.

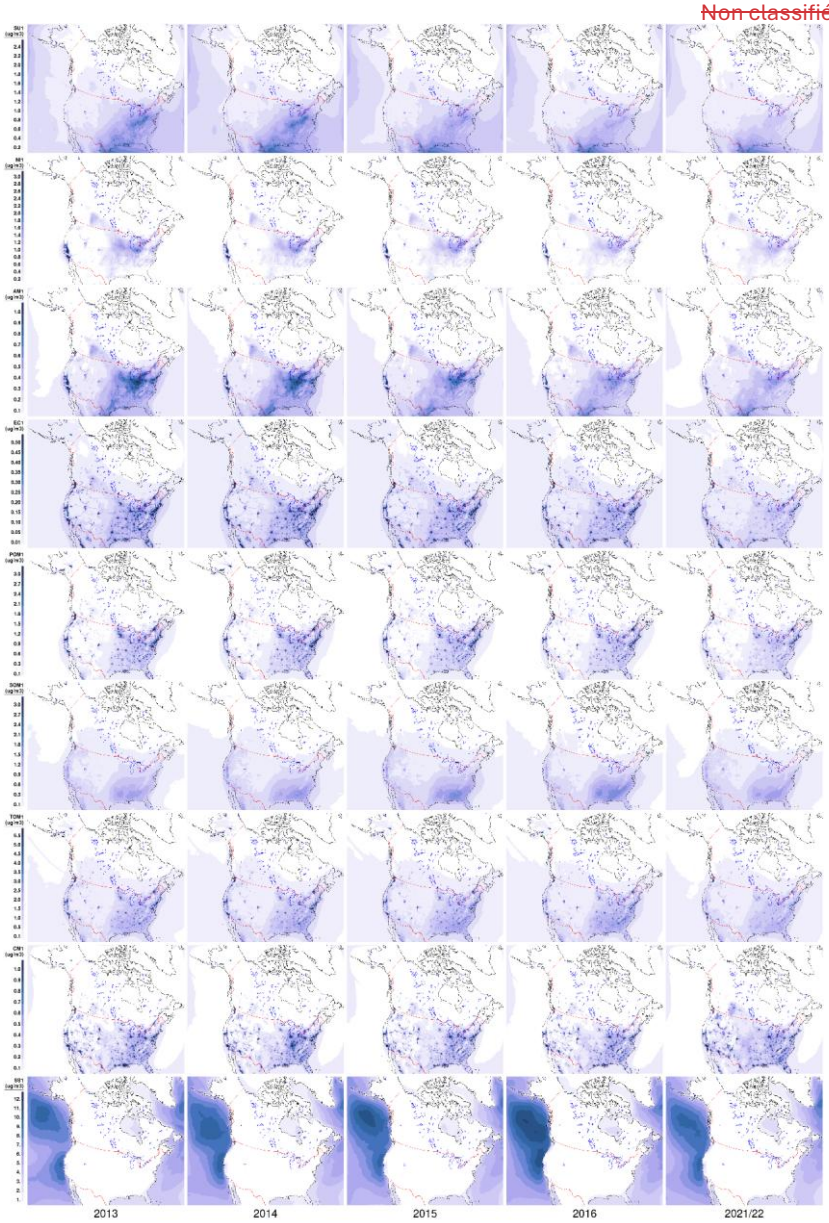


Figure 14. Spatial distribution of RAQDPS-OP023-predicted ~~mean~~ annual-mean ambient surface concentrations of nine daily $PM_{2.5}$ chemical components ($\mu g \cdot m^{-3}$) for five years (from left to right, 2013–2016 and 2021/22). These components are $SO_4$, $NO_3$, $NH_4$, EC, POM, SOM, TOM, CM, and SS (in rows ordered from top to bottom).

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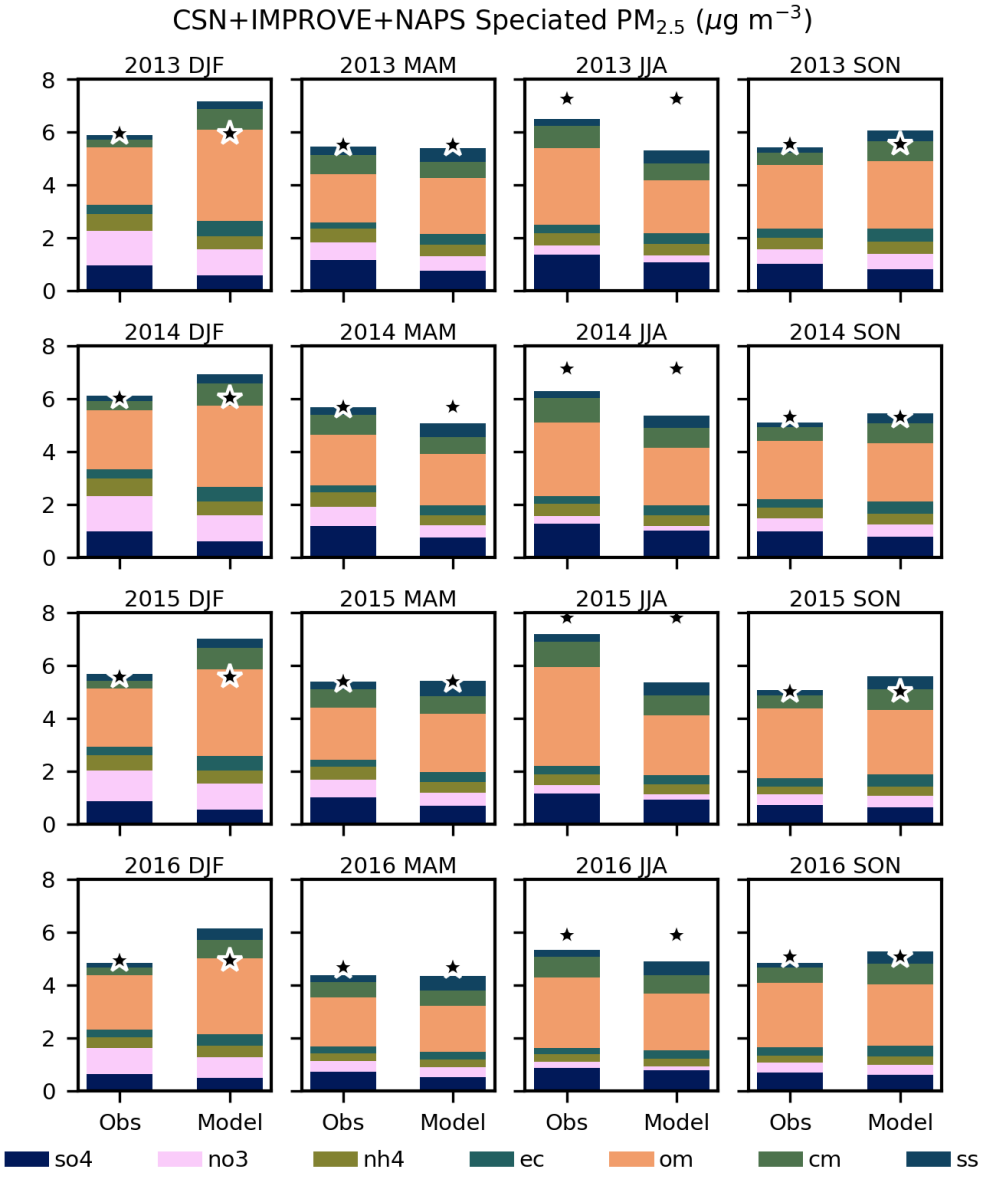


Figure 15. Stacked bar graphs of observed vs. RAQDPS-OP023-predicted domain-wide season-mean~~al~~ $PM_{2.5}$ chemical component concentrations ($\mu g \cdot m^{-3}$) based on combined CSN, IMPROVE, and NAPS $PM_{2.5}$ speciation daily measurements and hindcasts for four consecutive years. The top row corresponds to 2013 seasons, the next two rows below to 2014 and 2015 seasons, and the bottom row to 2016 seasons. Each row has four seasonal bar-graph pairs (observed and predicted), starting with winter (DJF) on the left, and then spring (MAM), summer (JJA), and autumn (SON) on the right. The stars mark the measured or predicted season-mean~~al~~ gravimetric $PM_{2.5}$ total mass.

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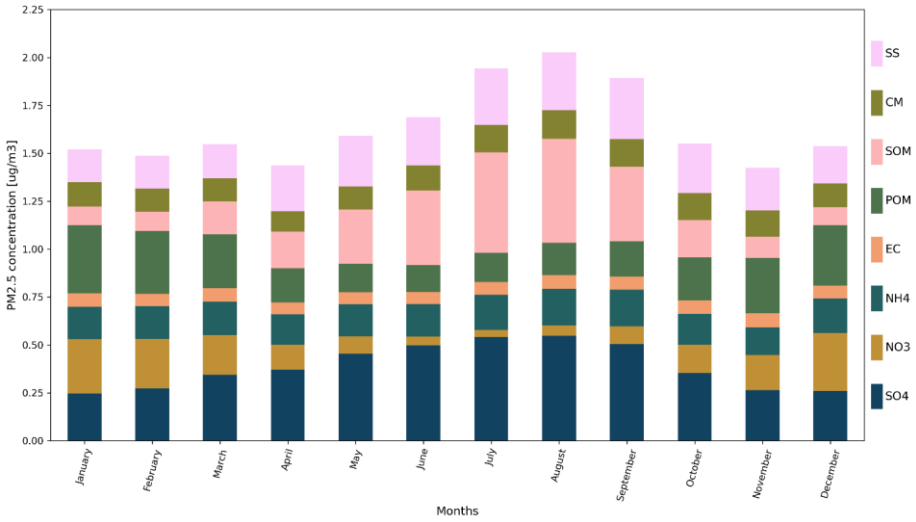


Figure 16. Time series of RAQDPS-OP023-predicted monthly mean variation of eight daily $PM_{2.5}$ chemical component concentrations ($\mu g \cdot m^{-3}$) area-weighted averaged over North American continent grid cells and 2013 to 2016 simulations. These components are $SO_4$, $NO_3$, $NH_4$, EC, POM, SOM, CM, and SS (shown ordered from bottom to top in stacked bar graphs).


Figure 17. Time series of RAQDPS-OP023-predicted season-mean diurnal (UTC) variation of eight hourly $PM_{2.5}$ chemical component concentrations ($\mu g \cdot m^{-3}$) area-weighted averaged over North American continent grid cells and 2013 to 2016 simulations for each of four seasons: (a) winter; (b) spring; (c) summer; and (d) autumn. These components are $SO_4$, $NO_3$, $NH_4$, CM, SS, EC, POM, and SOM (shown ordered from bottom to top in stacked bar graphs).

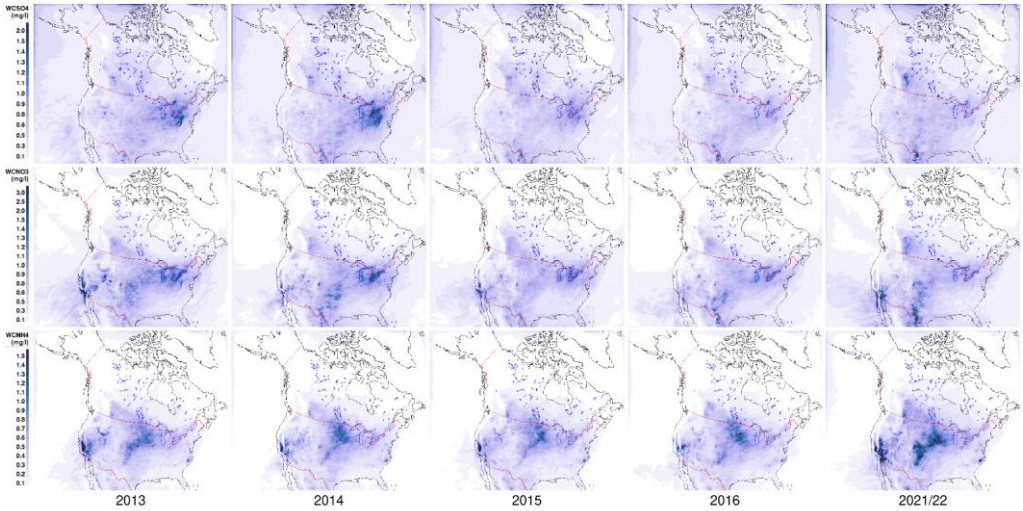


Figure 18. Spatial distribution of RAQDPS-OP023-predicted ~~mean~~ annual-mean concentrations in precipitation ($mg \cdot L^{-1}$) of (top row) $SO_4^{=}$, (middle row) $NO_3^{-}$, and (bottom row) $NH_4^{+}$ for five years (from left to right, 2013–2016 and 2021/22).

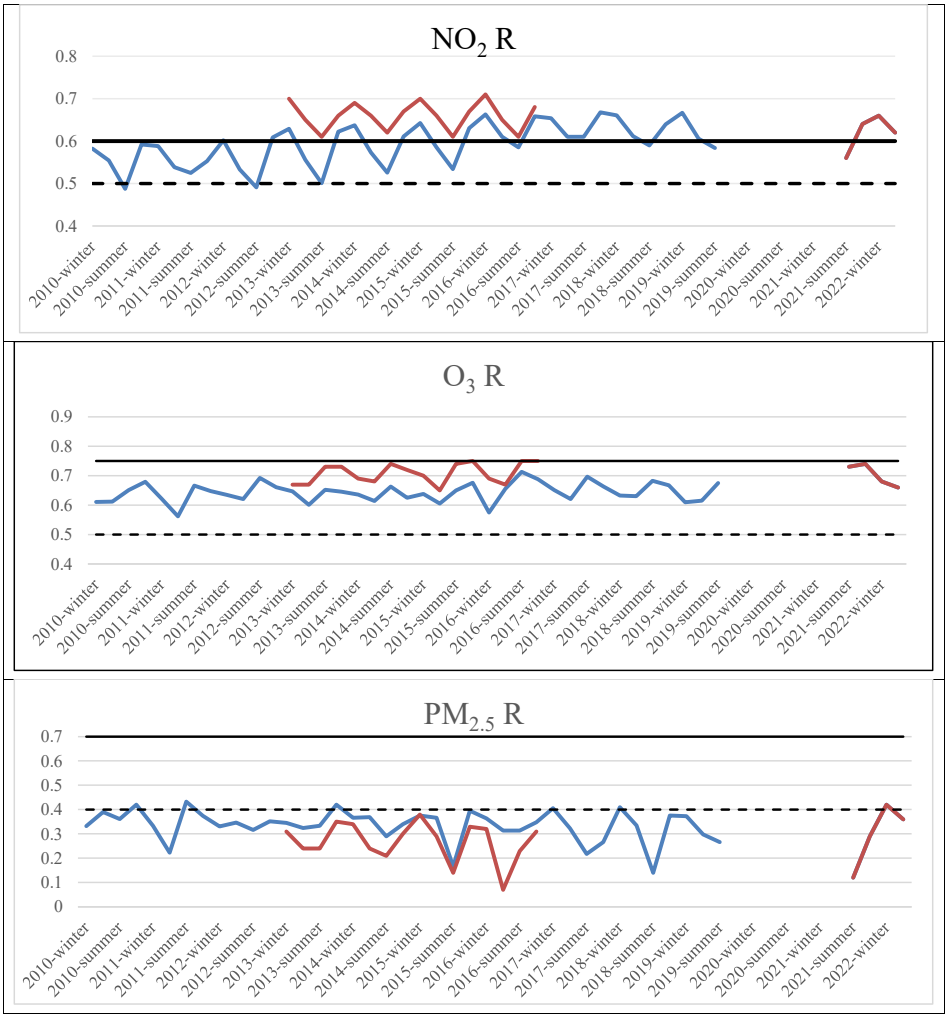


Figure 19. Time series of season-mean~~al~~ correlation coefficient (R) scores for surface (top) $NO_2$, (middle) $O_3$, and (bottom) $PM_{2.5}$ hourly abundances for all available North American NRT measurement stations for the periods Jan. 2010–June 2019 and June 2021-May 2022. The solid blue line denotes scores for the operational RAQDPS at the time while the solid ~~orange~~ red line denotes scores for the RAQDPS-OP023 forecasts and hindcasts. The dashed and solid black lines mark the "acceptable" and "good" benchmark thresholds, respectively, taken from Emery et al. (2017) for $O_3$ and $PM_{2.5}$ and from Zhai et al. (2024) for $NO_2$.