

1 **Reframing gullies as recharge zones in dryland landscapes of the Loess Plateau, China**

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18

19 **Abstract**

20 Large gullies in dryland landscapes are often indicators of land degradation by surface runoff.  
21 However, under conditions where gully systems are hydrologically arrested by restoration interventions  
22 that increase water residence time—most notably check dams and ponds—they may also function as  
23 hydrologically active zones of groundwater recharge and subsurface connectivity. In China’s Loess  
24 Plateau, we assess these functions in the Nianzhuang Catchment using a multi-indicator, process-based  
25 approach that integrates stable isotopes ( $\delta^2\text{H}$ ,  $\delta^{18}\text{O}$ ), chloride concentrations, and groundwater level  
26 fluctuations. Our results show that precipitation is the dominant source of recharge for shallow pore water  
27 within engineered gully zones, while deeper fissure water is replenished more slowly through percolation.  
28 Hydrological arrest through ecological engineering interventions acts as focal points for groundwater  
29 infiltration, enhancing recharge in otherwise limited dryland systems. Estimated annual recharge in the  
30 monitored gully-zone pore aquifer (238–241 mm) is equivalent to about 43% of the mean annual  
31 precipitation at the site, a site-specific recharge magnitude that far exceeds reported catchment-wide  
32 recharge rates observed in nearby tableland and hilly areas. Our results indicate that engineered gully  
33 systems can act as focused recharge zones rather than solely degraded landforms. By linking runoff  
34 convergence and ponding to measurable recharge responses, the study provides a process-based  
35 framework for assessing groundwater dynamics in managed semi-arid landscapes.

36

37 **Keywords:** surface water, spring water, pore water, fissure water, hydrological connectivity, groundwater  
38 recharge

39

40 **1. Introduction**

41 Groundwater recharge is a critical yet poorly understood component of hydrological cycles in  
42 dryland catchments (Li et al., 2024a). It is shaped by the precipitation regime, surface landcover  
43 heterogeneity, integrity of the subsurface regolith, characteristics of the underlying bedrock, and human  
44 interventions (Vries and Simmers, 2002; Owuor et al., 2016; Salek et al., 2018; Xu and Beekman, 2019;  
45 Zhang et al., 2020; Li et al, 2024b; Medici et al., 2024). While favorable subsurface flow pathways can  
46 locally enhance recharge, dryland regions are highly sensitive to even slight changes in precipitation, soil  
47 moisture, or runoff generation. This heightened sensitivity reflects their position along climatic ecotones

48 and the influence of complex land–atmosphere–biosphere feedbacks (Kuang et al., 2019; Al-Oqaili et al.,  
49 2020; He et al., 2020; Jin et al., 2019; Jia et al., 2024). Small changes in these processes can cascade  
50 across catchments at various scales, amplifying existing vulnerabilities to ecological and social systems  
51 (Nicholson, 2011; Huang et al., 2017; Berg et al., 2016). In these fragile landscapes, understanding  
52 groundwater replenishment processes is crucial for sustaining ecosystems, securing water, and guiding  
53 restoration and management (Gleeson et al., 2016; Jasechko and Perrone, 2021; Scanlon et al., 2006).

54 Despite a growing body of research on groundwater recharge in (semi-) arid regions, significant  
55 knowledge gaps remain in landscapes with pronounced spatial heterogeneity, such as slopes, hilltops,  
56 and gully systems, where infiltration pathways and recharge processes can diverge sharply over short  
57 distances (Tooth, 2012; Manna et al., 2018; Letz et al., 2021). Gully systems, often seen as signs of land  
58 degradation, may beneficially act as recharge zones, capturing and infiltrating surface runoff during  
59 episodic rainfall (Tan et al., 2017; Li et al., 2024a; Xue et al., 2025). This same topographic focusing  
60 enables the rapid downslope transport of contaminants, including agricultural nutrients, sediments, and  
61 associated pollutants (Lian et al., 2025; Qu et al., 2025). However, the role of gullies in promoting vertical  
62 infiltration into groundwater is highly dependent on local subsurface connectivity and permeability  
63 conditions. Moreover, their broader hydrological functions remain poorly quantified, especially under  
64 the influence of widespread human interventions such as check dams and artificial ponds. While these  
65 structures are typically designed to arrest land surface degradation, they can substantially alter surface–  
66 subsurface connectivity and reshape recharge dynamics in uncertain ways (Lamontagne et al., 2021;  
67 Huang et al., 2019; Wang et al., 2023).

68 Worldwide, loess covers approximately 6% of the land surface area, forming discontinuous east–  
69 west belts in the mid-latitude forest-steppe, steppe, and desert-steppe zones of both hemispheres (Liu,  
70 1985; Pécsi, 1990; Li et al., 2020). Among these, the Chinese Loess Plateau, the focus of our study  
71 accounts for approximately 7.4% of the global loess area (635,280 km<sup>2</sup>; Li et al., 2020). It serves as a  
72 globally important natural laboratory for studying soil erosion and groundwater recharge processes, due  
73 to its exceptionally thick loess deposits (Li et al., 2021), highly erodible soils, intense summer rainstorms,  
74 and long history of agricultural activity, which collectively make it one of the most severely eroded  
75 regions worldwide (Shi and Shao, 2000; Fu et al., 2011). Its distinctive stratigraphic structure,  
76 characterized by thick, low-permeability loess layers, fundamentally governs groundwater behavior  
77 (Qiao et al., 2017). Meanwhile, extensive human interventions aimed at erosion control, including large-

78 scale afforestation and gully engineering projects, have profoundly altered regional hydrological  
79 processes and the spatial redistribution of water (Wang et al., 2020; Zhao et al., 2024).

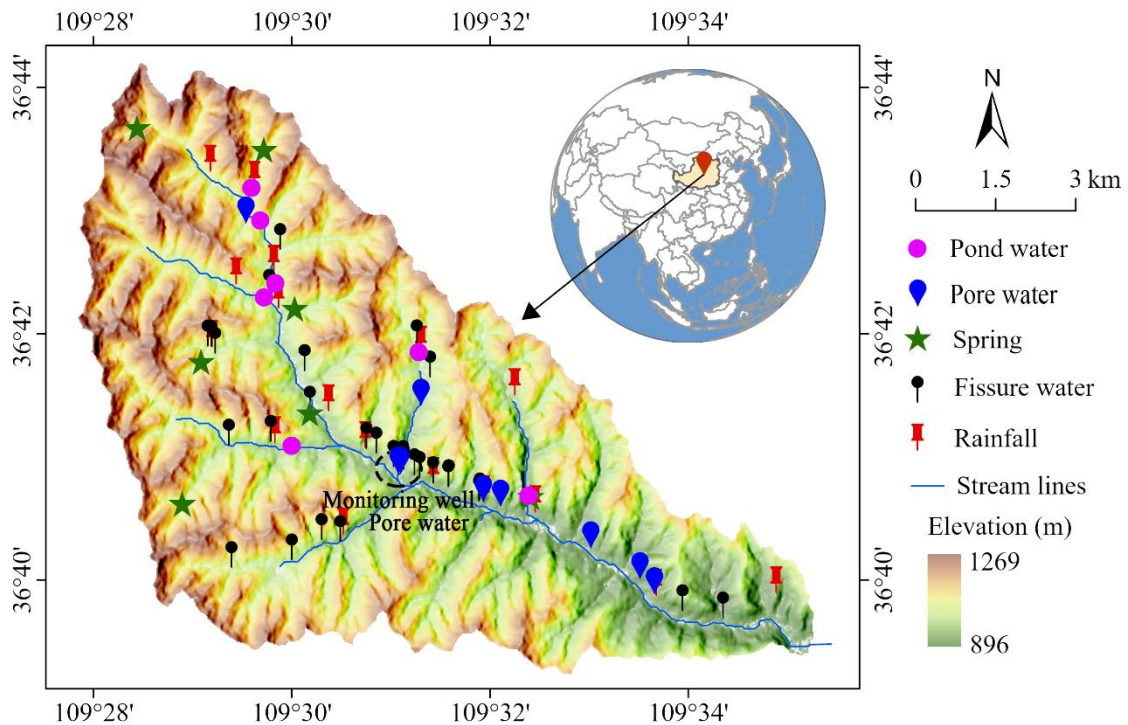
80 The setting for our investigation is a semi-arid landscape that has been shaped by severe soil erosion,  
81 extensively modified by engineered landforms; and it is now characterized by chronic water scarcity (Fu  
82 et al., 1999; Liu et al., 2017; Liu and Li, 2017; Li et al., 2021; Huang et al., 2024). Water scarcity  
83 manifests as declining groundwater levels, reduced streamflow, dried-up wells and springs, and limited  
84 irrigation capacity (Yu et al., 2025). In such vulnerable environments, understanding the sources and  
85 sustainability of groundwater recharge is critical for long-term water resource management (Ajjur and  
86 Baalousha, 2021; Meles et al., 2024). Groundwater, for example, is a lifeline for rural communities in  
87 the hilly–gully region, yet scientific attention has largely bypassed the gullies themselves. Most previous  
88 studies have focused on recharge processes in tablelands and loess-covered hills, highlighting slow  
89 “piston flow” as the dominant mechanism (Huang and Pang, 2011; Huang et al., 2013; Li et al., 2017;  
90 Lu, 2020; Wang et al., 2024). However, the deep-profile recharge mechanisms observed in these areas  
91 may not apply to the gully-dominated landscapes of the Loess Plateau (Wang et al., 2024; Qiao et al.,  
92 2017; Zhu et al., 2018). Moreover, the hydrological functions of widely distributed gully systems,  
93 especially under the influence of engineering structures such as check dams, remain insufficiently  
94 quantified, and their underlying processes have long remained in the research shadow (Liu et al., 2011).

95 Therefore, this study selects the Nianzhuang Catchment, a typical gully area on the Loess Plateau  
96 impacted by check dams, to establish a multi-method framework for assessing groundwater recharge by  
97 integrating stable isotope analysis ( $\delta^2\text{H}$  and  $\delta^{18}\text{O}$ ), chloride concentrations, water table fluctuations, and  
98 hydro-statistical modeling. Specifically, our goals are to: (1) characterize the isotopic and hydrochemical  
99 signatures of precipitation, surface water (ponds), shallow pore water, and deeper fissure water; (2)  
100 identify and trace hydraulic connections and flow paths of different water bodies; and (3) quantitatively  
101 estimate pore-water recharge rates. This integrated approach aims to advance understanding of  
102 groundwater dynamics in complex dryland terrains, reframes engineered gully systems as critical  
103 recharge zones in engineered dryland landscapes, providing actionable insights for sustainable  
104 groundwater management and ecological restoration in the Loess Plateau and similar semi-arid regions  
105 worldwide.

106

107 **2. Sampling site**

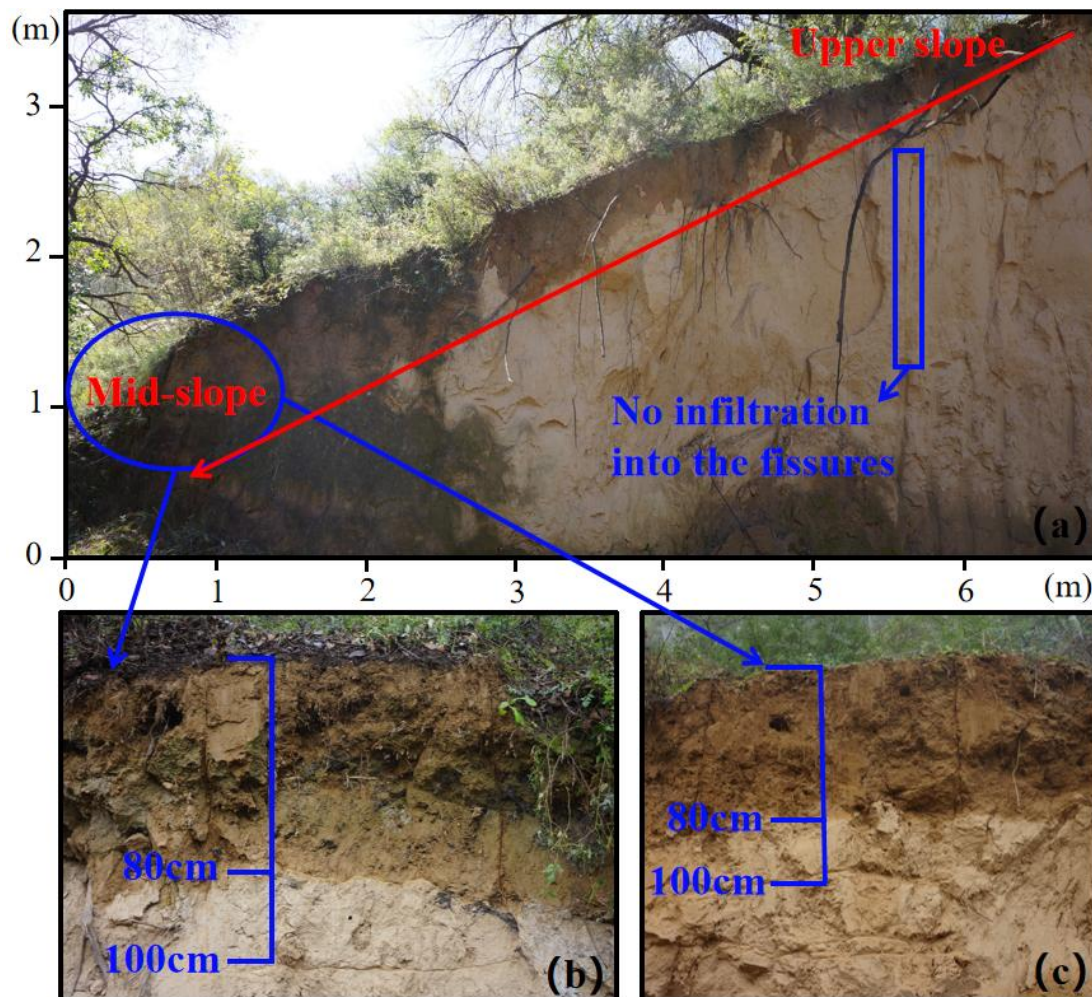
108 The Nianzhuang Catchment is located northwest of Yan'an City in Shaanxi Province, China  
109 (approximately 36°42'N, 109°31'E). As a tributary of the Yan River, which ultimately flows into the  
110 Yellow River, the catchment spans 53.94 km<sup>2</sup> and includes the well-studied Yangjuangou sub-catchment  
111 (3.11 km<sup>2</sup>; ~36°35'N, ~109°32'E), previously investigated in numerous hydrological and ecological  
112 studies (Fu et al., 1999; Fu et al., 2011; Fu et al., 2017; Liu and Li, 2017). Elevation ranges from 896 to  
113 1,269 m, with terrain gradually sloping from northwest to southeast (Fig. 1). The region experiences a  
114 semi-arid continental monsoon climate, with a mean annual precipitation of approximately 550 ± 100  
115 mm, concentrated between July and September (Liu et al., 2017).



116  
117 Fig. 1. The geographical location and sampling sites for rainfall, pond water, pore water, spring water,  
118 and fissure water in the Nianzhuang Catchment. The Nianzhuang Catchment is located in the hilly and  
119 gully region of the central Loess Plateau, with elevations ranging from 896 to 1269 m. The average depth  
120 of pore water wells is 8.0 ± 1.5 m (range: 4–10 m), while that of fissure water wells is 57.6 ± 29.2 m  
121 (range: 25–170 m). These sampling sites represent locations where both rainy and dry season samples  
122 were collected, and are all situated within the gully areas of the catchment.

123  
124 The catchment features highly dissected loess terrain, with characteristic soils and landforms such  
125 as Loess Liang (ridges), Loess Mao (mounds), and steep loess slopes (Cai et al., 2019). Gullies, often

126 “V”- or “U”-shaped, dominate the lower-lying regions and serve as important recharge zones. These  
 127 landforms, together with ancient landslides, minor collapses, and sinkholes, highlight the geomorphic  
 128 instability of the Loess Plateau landscape (Li et al., 2021). From May to October 2023, total rainfall  
 129 reached 420 mm, with 115 mm in September alone. Despite this substantial precipitation, field  
 130 observations revealed shallow infiltration depths on loess slopes even after heavy rainfall events of up to  
 131 41 mm. Infiltration was limited to 20–30 cm at hilltops and about 80 cm at mid-slope, with no distinct  
 132 preferential flow and largely unsaturated soil profiles (Fig. 2). These observations suggest that  
 133 groundwater recharge occurs mainly through surface or near-surface runoff converging into engineered  
 134 gully systems, underscoring their critical role as focused zones of groundwater recharge and key sites for  
 135 studying these processes.



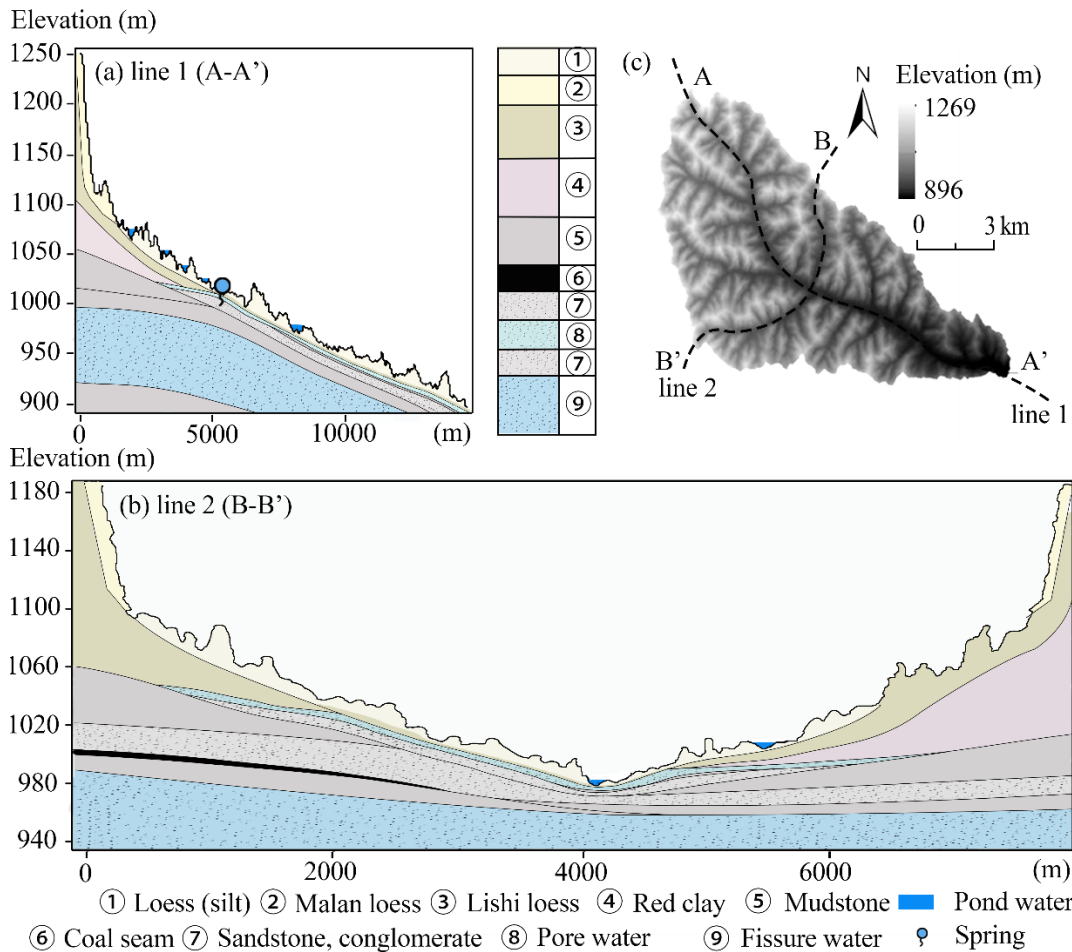
136  
 137 Fig. 2. The topographic profile of the Nianzhuang Catchment in the hilly region of the Loess Plateau.  
 138 Full profile from the top to mid-slope (a); two repeated mid-slope profiles (b, c). The photo was taken  
 139 after a 41 mm rainfall event over four days. Subsequent measurements showed that infiltration depths

140 reached only 20–30 cm at the top of the slope, compared to approximately 80 cm at the mid-slope  
141 positions.

142

143 The stratigraphy of the catchment reflects the typical layered structure of the Loess Plateau, which  
144 plays a key role in controlling groundwater recharge. In upland hilly areas, thick loess deposits overlie  
145 bedrock, with the Upper Pleistocene Malan Loess, a light grayish-yellow, loosely textured, and silt-rich  
146 unit (>60%), characterized by well-developed vertical joints and abundant hematite and goethite.  
147 Beneath it lies the Middle Pleistocene Lishi Loess, a grayish-yellow to light brown unit with prominent  
148 jointing and higher iron mineral content. Below the loess, the Neogene Red Clay appears as a distinctly  
149 reddish, calcareous nodule-bearing aquitard due to its low permeability. The entire sequence rests on  
150 Jurassic sandstone–conglomerate bedrock, composed mainly of quartz-rich fluvial–lacustrine deposits.

151 Loess thickness in the Liang and Mao regions often exceeds 150 meters, resulting in deep water  
152 tables and limited groundwater accessibility. In contrast, gully zones exhibit distinctly different  
153 hydrogeological characteristics. Here, thinner loess layers overlie Neogene and Jurassic formations,  
154 sometimes interbedded with coal seams up to 5 meters thick (Fig. 3a–c). The significant reduction in  
155 loess thickness, combined with the relatively high permeability of Neogene coarse sandstone and  
156 conglomerate (0.07–0.31 m/d), creates favorable conditions for infiltration and focused recharge. These  
157 dynamics are especially evident at gully heads, where surface runoff from adjacent uplands converges  
158 and infiltrates, forming efficient recharge zones. As a result, gully areas tend to have shallower water  
159 tables and more rapid water renewal, making them more suitable for domestic groundwater use. Springs  
160 frequently emerge at gully bottoms where lateral flow is facilitated at the loess–bedrock interface.  
161 Streams in this dry environment are largely intermittent.



163 Fig. 3. Hydrogeologic cross-section of the study area. Cross-section along Line 1 (Northwest-Southeast)

164 (a); cross-section along Line 2 (Southwest-Northeast) (b); location map of Line 1 and Line 2 within the

165 study area (c). The Malan Loess (11.7–12.6 Ka BP) and Lishi Loess (12.6–78.1 Ka BP) are two major

166 Quaternary loess stratigraphic units in China. Based on hydrogeological research, the stratigraphy of the

167 hilly region features a multi-layer structure from top to bottom: Upper Pleistocene Malan Loess, Middle

168 Pleistocene Lishi Loess, Neogene Red Clay and Mudstone (2.58–23.03 Ma BP), and Jurassic Sandstone

169 and Conglomerate (145–201.3 Ma BP). In the gully region, the stratigraphy includes Holocene loess (silt,

170 11.7 ka BP–present), Middle Pleistocene Lishi Loess, Neogene sandstone and mudstone, and Jurassic

171 sandstone and conglomerate, with some areas containing coal seams up to 5 meters thick.

172

173 Groundwater in the catchment can be broadly categorized into three types: pore water, spring water,

174 and fissure water. Pore water is stored in permeable sandstone and conglomerate aquifers beneath loess

175 and above mudstone or red clay. These aquifers are approximately 2–3 m thick, exhibit a sheet-like

176 distribution, and have low water yield. Conceptually, “pore water” here refers to groundwater in a

177 saturated aquifer, not to soil moisture. Fissure water occurs in fractured bedrock aquifers, which are  
178 spatially discontinuous due to irregular fracture development. The main water-bearing zones include  
179 cavities and jointed fissure networks, with an average aquifer thickness of about 6 m and moderate water  
180 yield. Hydraulic conductivity in these sandstone and conglomerate aquifers ranges from 0 to 0.47 m/d  
181 (Cai et al., 2019). Spring water emerges primarily at gully bases, especially in upper catchments, and  
182 originates from both pore and fissure sources, possibly supplemented by surface or pond water. Springs  
183 fed by pore water typically have low discharge rates (0–0.1 L/s) and low water yield, while those fed by  
184 fissure water exhibit moderate discharge rates (0.5–1.0 L/s) and moderate water yield.

185 Over recent decades, landscape rehabilitation through the Grain for Green Project and land  
186 reshaping under the Gully Land Consolidation Project have significantly altered the hydrological regime  
187 (Fu et al., 1999; Liu et al., 2017). Historically, surface runoff in the degraded catchment was flashy and  
188 episodic due to sparse vegetation. However, ecological restoration and small-scale engineering  
189 interventions, such as check dams, terraces, roads, and ponds, have moderated surface hydrology. Surface  
190 runoff, generated primarily during storm events, now contributes alongside delayed baseflow from  
191 groundwater recharge and interflow. The latter is often limited by the thick unsaturated zone in upland  
192 loess areas but may be enhanced in gully regions, where stratigraphy and land use favor infiltration  
193 (Wang et al., 2024; Gates et al., 2011). Gully areas also contain numerous check dams and ponds, with  
194 most water sourced from Hortonian overland flow of slope lands and direct rainfall. These small water  
195 bodies, often constructed for erosion control and water retention, influence local hydrological dynamics  
196 and may play a role in enhancing infiltration and recharge.

197

### 198 **3. Methods**

199 Our approach integrates stable isotope analysis ( $\delta^2\text{H}$  and  $\delta^{18}\text{O}$ ), chloride concentration analysis, and  
200 water table fluctuation monitoring to investigate groundwater recharge dynamics. The isotopic  
201 composition of water bodies reflects both their origins and the processes they undergo, such as  
202 evaporation, infiltration, and mixing (Wan and Liu, 2016; Kumar et al., 2019; Dasgupta et al., 2024).  
203 Precipitation, surface water, and groundwater typically exhibit distinct isotopic signatures due to these  
204 differing pathways (Gleeson et al., 2016; Kuang et al., 2019; Al-Oqaili et al., 2020). When isotopic  
205 patterns among water sources converge, it often indicates strong hydrological connectivity (Yang and

206 Wang, 2023). Because stable isotopes behave conservatively, they serve as effective tracers of water  
207 sources and flow paths (Gleeson et al., 2016; Al-Oqaili et al., 2020; Dasgupta et al., 2024). In parallel,  
208 water table fluctuation (WTF) monitoring provides a means of estimating recharge by observing changes  
209 in groundwater levels in response to precipitation events (Nachabe, 2002; Heppner and Nimmo, 2005;  
210 Gumuła-Kawęcka et al., 2022). By combining these complementary methods, this study aims to elucidate  
211 groundwater recharge pathways and quantify recharge rates in gully regions, thereby identifying key  
212 recharge zones and advancing our understanding of groundwater processes in the Loess Plateau.

213

### 214 **3.1. Field measurements of hydrological data**

215 Precipitation was collected from October 24, 2023, to October 24, 2024, using a weather station  
216 situated in an open field within the catchment. Continuous groundwater level data were recorded from  
217 September 24, 2023, to December 20, 2024. Groundwater pressure and temperature were monitored  
218 using Onset HOBO U20-001-03 sensors (20 m range), with a pressure accuracy of  $\pm 0.3\%$  FS ( $\pm 2.55$  kPa)  
219 and a resolution of  $< 0.085$  kPa, and a temperature accuracy of  $\pm 0.44$  °C with a resolution of 0.1 °C. The  
220 sensor was calibrated to atmospheric pressure before installation to ensure accurate measurement of  
221 absolute static water pressure, and water table levels were calculated based on the measured pressure  
222 data. The conversion relationship between water pressure and groundwater level is given by  $Y =$   
223  $0.86 \times X - 22.1$  where  $Y$  represents the groundwater level and  $X$  represents the water pressure. The  
224 conversion between water pressure and groundwater level is based on the principle of hydrostatics. The  
225 hydrostatic pressure  $P$  at the sensor is related to the height of the overlying water column  $h$  by  $P = \rho gh$ ,  
226 where  $\rho$  is the water density and  $g$  is the gravitational acceleration. In unconfined aquifer, the pressure  
227 measured by the sensor corresponds directly to the static pressure exerted by the overlying water column.  
228 From this, the water column height  $h$  can be calculated, and combined with the sensor's installation  
229 elevation, the depth to the groundwater table can be determined. Notably, the monitoring well is located  
230 in the pore water layer of the gully region. The well is hand-dug (1.1 m wide, 10 m deep) and is unaffected  
231 by human activities.

232 Soil physical properties were assessed using the cutting ring method, based on undisturbed soil cores  
233 collected at five depth intervals: 0–10, 10–20, 20–30, 30–40, and 40–50 cm. Quadruplicate samples were  
234 taken near groundwater monitoring wells in the gully using pre-weighed cutting cylinders. The samples  
235 were immediately transported to the laboratory for analysis. Bulk density, capillary porosity, non-

236 capillary porosity, total porosity, and field water capacity were determined following the LY/T 1215-  
237 1999 standard for forest soil water-physical properties. Soil particle size distribution was analyzed using  
238 a laser particle size analyzer at the College of Natural Resources and Environment, Northwest A&F  
239 University. Soil texture classification followed the USDA system: sand (0.05–2 mm), silt (0.002–0.05  
240 mm), and clay (<0.002 mm) (Dane et al., 2002).

241

### 242 **3.2 Water sampling**

243 A total of 181 water samples were collected from various locations in rainy season (September 2023,  
244 99 samples) and dry season (April 2024, 82 samples); see Fig. 1. Rainy season samples included 48 from  
245 rainfall, 7 from pond water (water retention reservoirs), 9 from spring water, 9 from pore water, and 26  
246 from fissure water. During the dry season, samples included 31 from rainfall, 6 from pond water, 8 from  
247 pore water, 29 from fissure water, and 8 from spring water.

248 Pore water was collected from several shallow, hand-dug wells measuring approximately 1.1 meters  
249 in diameter and 4–10 meters in depth. Fissure water was sampled from deeper, narrow-diameter wells  
250 (0.2 meters wide, 25–170 meters deep). In areas with numerous deep wells, we employed random  
251 sampling to ensure representative coverage of fissure water sources. To minimize the risk of collecting  
252 stagnant water, all pore and fissure water wells were purged for 10–15 minutes prior to sampling. Spring  
253 water was collected directly from natural discharge points, although most springs in the region exhibit  
254 low flow rates, typically less than 0.1 L/s, occasionally reaching up to 0.2 L/s.

255 A total of 18 bulk rainfall collectors were randomly and evenly distributed across the 54 km<sup>2</sup> study  
256 area, and samples were collected immediately following rainfall events. For nighttime precipitation,  
257 samples were collected the next morning at 6:00 AM. During the study period, we collected two types of  
258 precipitation samples: (1) spatial samples from individual events (18 in the rainy season and 15 in the  
259 dry season) across the catchment, capturing spatial variability; and (2) sequential events at a fixed station  
260 (30 in the rainy season and 16 in the dry season), characterizing seasonal inputs.

261 During sampling, 100 mL collection bottles were rinsed two to three times with the sample water,  
262 then slowly filled to minimize air exposure. After filling, the bottles were tightly sealed with screw caps  
263 and further secured with Parafilm to prevent evaporation and contamination. All samples were  
264 immediately stored in a portable cooler at 4°C and transported to the laboratory for isotopic and chloride  
265 concentration analysis.

266

### 267 3.3. Isotopic analysis

268 The  $\delta^2\text{H}$  and  $\delta^{18}\text{O}$  values of the water samples were determined using a Los Gatos Research liquid  
269 water isotope analyzer (Model 912-0032, LGR Inc., California, USA) at the Institute of Water-Saving  
270 Agriculture in Arid Areas of China, Northwest A&F University. Each sample was injected six times in  
271 the following sequence: three standard injections, followed by six natural sample injections, and then  
272 three additional standard injections. The isotope ratios were calculated using the average composition  
273 from injections 4 through 6.

274 Isotope values are expressed in delta ( $\delta$ ) notation, which represents the relative difference in isotope  
275 ratio between a sample and the Vienna Standard Mean Ocean Water (VSMOW) reference. The  
276 measurement precision was  $\pm 0.5\text{‰}$  for  $\delta^2\text{H}$  and  $\pm 0.1\text{‰}$  for  $\delta^{18}\text{O}$ . The delta values were calculated using  
277 the following equations:

$$278 \quad \delta^{18}\text{O} = \left( \frac{R_{\text{sample}}}{R_{\text{standard}}} \right) - 1 \quad (1)$$

$$279 \quad \delta^2\text{H} = \left( \frac{R_{\text{sample}}}{R_{\text{standard}}} \right) - 1 \quad (2)$$

280 where  $R_{\text{sample}}$  and  $R_{\text{standard}}$  are the ratios of heavy to light isotopes ( $^{18}\text{O}/^{16}\text{O}$  or  $^2\text{H}/^1\text{H}$ ) in the sample  
281 and the standard, respectively. Results are expressed in per mil (‰).

282

### 283 3.4. Mixing process of different water bodies

284 Inverse transit time proxies (ITTPs) were calculated to assess differences in water transit times and  
285 mixing processes across various water bodies (Tetzlaff et al., 2009). ITTPs are defined as the ratio of the  
286 standard deviation of  $\delta^{18}\text{O}$  in the water sample (e.g., pond water, spring water, pore water, or fissure  
287 water) to that in precipitation over the same time period:

$$288 \quad \text{ITTP} = \frac{\sigma_{\delta^{18}\text{O}}(\text{sample})}{\sigma_{\delta^{18}\text{O}}(\text{precipitation})} \quad (3)$$

289 This ratio captures the attenuation of seasonal isotopic variability in  $\delta^{18}\text{O}$  as water moves through  
290 the landscape. In general, ITTP values less than 1 indicate substantial damping of the precipitation signal,  
291 consistent with longer water residence times, greater mixing, and larger storage volumes. Conversely,  
292 values approaching 1 suggest minimal damping and rapid flow paths.

293 However, interpretation of ITTPs must also account for fractionation processes. In particular,  
294 evapotranspiration (ET) selectively removes lighter isotopes ( $^{16}\text{O}$ ), enriching the remaining water in

295 heavier isotopes ( $^{18}\text{O}$ ). This enrichment can artificially increase the variance of  $\delta^{18}\text{O}$  in near-surface or  
296 shallow soil water compartments, inflating ITTP values even in systems with relatively slow transit times  
297 (Tetzlaff et al., 2009). This is especially relevant in arid and semi-arid regions, where ET can dominate  
298 the water balance during dry seasons.

299

### 300 **3.5. Hydraulic connectivity estimation**

301 Structural Equation Modeling (SEM) has been widely applied in water science to evaluate complex  
302 relationships among hydrological, geological, and anthropogenic variables, particularly in studies of  
303 groundwater contamination and water quality degradation (Wu, 2010; Lupi et al., 2019; Xie et al., 2025).  
304 In this study, SEM is used explicitly as an exploratory, hypothesis-generating tool to assess potential  
305 hydrological connectivity among water sources based on dual-isotope ( $\delta^2\text{H}$ – $\delta^{18}\text{O}$ ) data from rainfall,  
306 pond water, spring water, pore water, and fissure water. SEM is not a mass-conserving or process-based  
307 flow model, nor is it used here to infer volumetric fluxes, recharge rates, or source apportionment. Instead,  
308 it serves as a statistical consistency check on hypothesized connectivity, identifying direct and indirect  
309 associations among water bodies that are evaluated in conjunction with tracer evidence and hydrometric  
310 observations.

311 Within the SEM framework, path relationships are primarily explained through two types of effects:  
312 The direct effect refers to the immediate impact of one variable on another through a single path, typically  
313 quantified as a standardized regression coefficient. Total effect represents the overall impact of one  
314 variable on another through all possible paths (including both direct and indirect), calculated as the sum  
315 of the direct effect and all indirect effects. Comparing direct and total effects allows identification of  
316 intermediary linkages and dominant association structures within the hypothesized connectivity network.

317 Given the potential for isotopic signatures to be altered by evaporation, mixing, or other non-  
318 conservative processes, results must be interpreted with caution. Pathways with p-values  $> 0.05$  were  
319 excluded during model refinement, and the final model met standard goodness-of-fit criteria (degrees of  
320 freedom  $< 3$ , RMSEA  $< 0.05$ , CFI  $> 0.95$ , NFI  $> 0.95$ ). SEM analysis was conducted using SPSS Amos  
321 26.0 (IBM SPSS, Chicago, Illinois, USA).

322 In addition, we applied variance partitioning to evaluate the relative contributions of different water  
323 sources to pore and fissure water. This method decomposes the total variance in isotopic composition  
324 into components attributable to individual sources (e.g., precipitation, pond water, spring water), offering

325 a complementary estimate of source influence. While useful, this approach remains subject to the same  
326 limitations as SEM, particularly the challenges of isotopic overlap and limited resolution in environments  
327 affected by mixing and evaporation (Lai et al., 2022).

328 To further constrain recharge pathways, we incorporated chloride ion ( $\text{Cl}^-$ ) as a conservative tracer.  
329 Unlike stable isotopes,  $\text{Cl}^-$  behaves conservatively with respect to fractionation, and when interpreted  
330 alongside isotopes, it can further explain mixing and recharge pathways. Chloride concentrations in all  
331 water samples were analyzed using an ion chromatograph (DIONEX ICS-1100) at the College of Natural  
332 Resources and Environment, Northwest A&F University, China. Each sample was analyzed in triplicate,  
333 with charge balance errors maintained below 5% to ensure analytical accuracy. This rigorous approach  
334 enhances the reliability of chloride data, supporting its integration with isotopic indicators in source  
335 attribution.

336

### 337 3.6. Groundwater recharge

338 Groundwater recharge in the gully zone is quantified using the water table fluctuation (WTF)  
339 method, which infers recharge and discharge events from temporal changes in groundwater levels (Healy  
340 and Cook, 2002; Gumuła-Kawęcka et al., 2022). We recognize that recharge can originate from three  
341 hydrological sources: (1) surface water; (2) the unsaturated zone (3) and the saturated zone (Scanlon et  
342 al., 2022; Wang et al., 2024). Among these, estimates based on the saturated zone are generally most  
343 reliable, as recharge from the unsaturated zone reflects potential inputs that may never reach the water  
344 table (Beven and Germann, 2013; Huang et al., 2019). The WTF method is widely used for estimating  
345 saturated zone recharge due to its high temporal resolution and conceptual simplicity (Xu et al., 2024).  
346 Based on previous site-specific studies (Wang et al., 2024), this method is well-suited for our analysis.

347 The water table fluctuation method assumes that changes in the groundwater table result solely from  
348 recharge or discharge, assuming a constant specific yield ( $S_y$ ) over time (Healy and Cook, 2002; Obuobie  
349 et al., 2012). The formula is as follows:

$$350 \quad R_i = S_y \frac{\Delta H_i}{\Delta t} \quad (4)$$

351 where,  $R$  is the groundwater recharge (mm),  $S_y$  is the specific yield of the aquifer,  $\Delta H_i$  (where  
352  $\Delta H_i > 0$ ) is rise in groundwater level between days  $i - 1$  and  $i$ , and  $\Delta t$  is the time interval (one day).  
353 Specific yield, which represents effective drainable porosity of the shallow gully aquifer system, rather

354 than soil water release from the unsaturated zone. In this study,  $S_y$  was estimated using two  
355 complementary approaches: (1) an empirical method based on soil texture to constrain plausible ranges  
356 of drainable porosity, and (2) the test pit method, in which  $S_y$  is calculated as the difference between total  
357 porosity and field water capacity (Liang, 2016). These approaches provide an estimate of aquifer-scale  
358 effective specific yield appropriate for shallow unconfined groundwater systems in loess-derived gully  
359 environments.

360 Empirical values for soil texture are referenced in Table A1. The test pit method for estimating  $S_y$  is  
361 described as follows:

$$362 \quad S_y = TP - FWC \quad (5)$$

363 where,  $TP$  (total porosity) and  $FWC$  (field water capacity) were measured using the cutting ring  
364 method.

365 We applied two methods to calculate the daily groundwater table increments ( $\Delta H_i$ ). The RISE  
366 method assumes that recharge occurs only when the groundwater table elevation increases between two  
367 consecutive days (Gumuła-Kawęcka et al., 2022). Thus,  $\Delta H_i = H_i - H_{i-1}$  if  $H_i > H_{i-1}$ , otherwise,  
368  $\Delta H_i = 0$ . The master recession curve (MRC) method assumes that, in the absence of recharge, the  
369 groundwater table declines daily by a specific amount ( $\Delta H_{MRCi}$ ). This amount represents a simplified  
370 approximation of discharge processes in the aquifer, particularly lateral outflow to nearby surface water  
371 bodies. MRC establishes a functional relationship between a daily decrement of the water table ( $\Delta H_{MRCi}$ )  
372 and the water table elevation ( $H_{i-1}$ ) during periods without recharge.

$$373 \quad \Delta H_{MRCi} = A \cdot H_{i-1} + B \quad (6)$$

374 where, the coefficients A and B were fitted for each piezometer based on data from periods of  
375 continuous groundwater table decreases lasting longer than two weeks. The daily water table increment  
376 due to recharge was then calculated as:  $\Delta H_i = H_i - H_{i-1} + \Delta H_{MRCi}$  if  $H_i > (H_{i-1} - \Delta H_{MRCi})$ ,  
377 otherwise,  $\Delta H = 0$ .

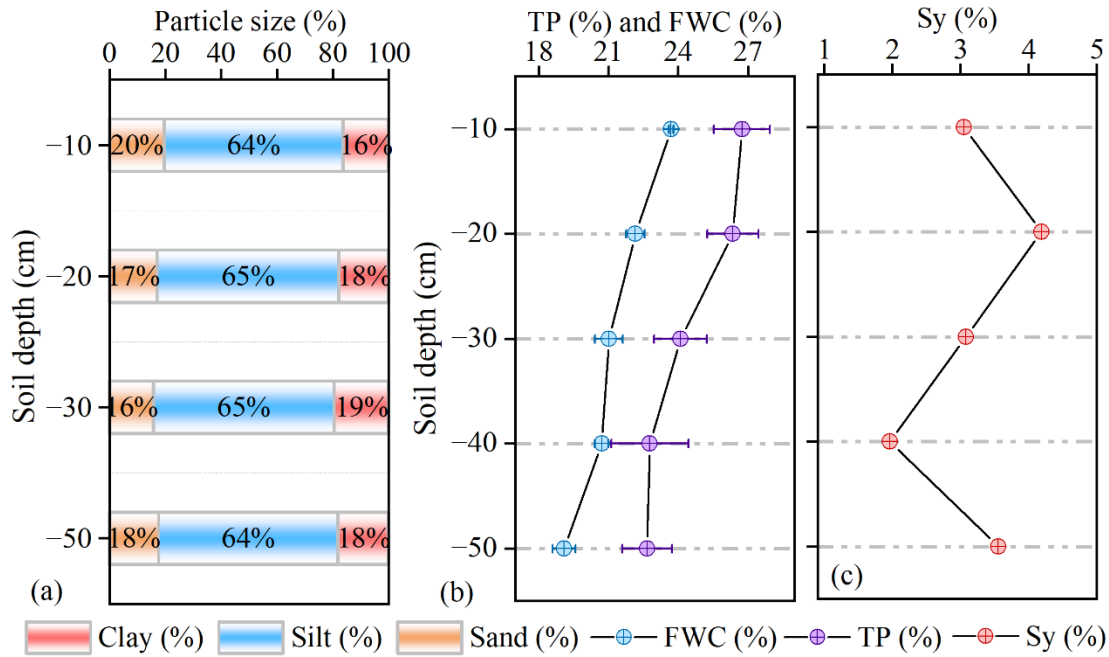
378 Notably, we used water dynamics from October 24, 2023, to October 24, 2024, to calculate pore  
379 water recharge, as this period exhibited clear groundwater fluctuations, making it more representative.

380

## 381 **4. Results**

### 382 **4.1. Soil properties**

383 The upper 50 cm of the soil profile is composed primarily of silt ( $64.6 \pm 0.6\%$ ), with smaller but  
 384 nearly equal proportions of clay ( $18.0 \pm 1.3\%$ ) and sand ( $17.4 \pm 1.6\%$ ), classifying the loess as silt loam  
 385 according to the International Union of Soil Sciences (IUSS) scheme (Fig. 4a). Total porosity and field  
 386 water capacity decreased slightly with depth, averaging  $24.5 \pm 1.9\%$  and  $21.3 \pm 1.7\%$ , respectively (Fig.  
 387 4b). Specific yield remained relatively consistent within the 10–50 cm depth interval, averaging  
 388  $3.2 \pm 0.8\%$ , falling within the expected empirical range for silt loam soils (2%–7%) (Fig. 4c).



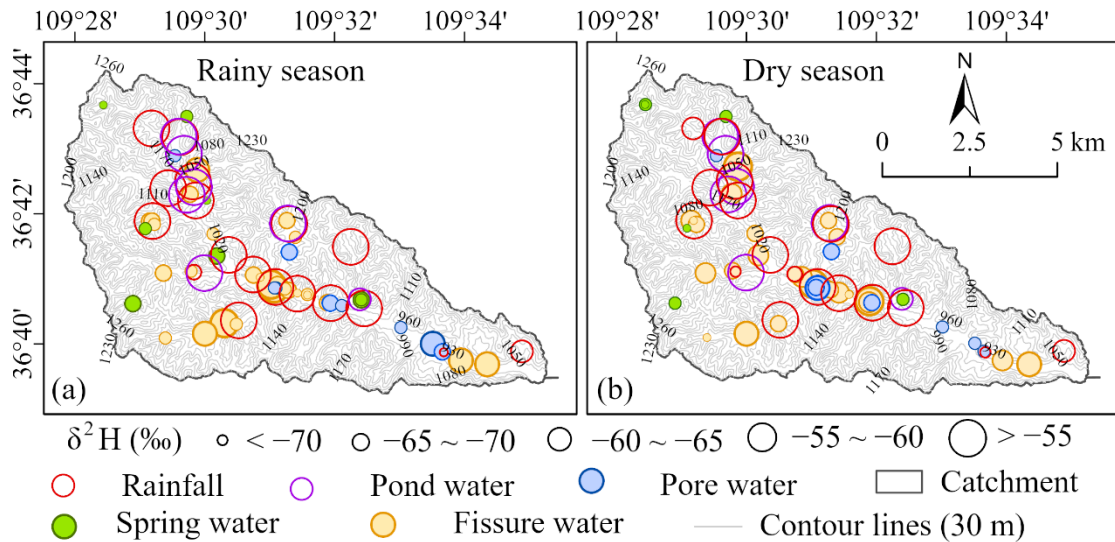
389 Fig. 4. Vertical variation in soil texture and water retention characteristics in the gully region of the Loess  
 390 Plateau. (a) Soil particle size distribution by depth, showing relatively uniform composition across layers  
 391 (10–50 cm), dominated by silt (64–65%), with moderate clay (16–20%) and low sand (16–20%) content.  
 392 This fine-textured profile supports high moisture retention and slows infiltration, promoting delayed  
 393 recharge. (b) Depth profiles of total porosity (TP) and field water capacity (FWC) reveal decreases with  
 394 depth to 40 cm, with FWC reaching ~27%, suggesting greater water-holding capacity in subsoil layers  
 395 and enhanced buffering of infiltrated water. (c) Vertical variations in the Specific Yield (Sy) across  
 396 different soil layers. Collectively, these physical properties reflect a vertically stratified soil system where  
 397 near-surface layers regulate infiltration pulses, and deeper layers act as long-term storage, shaping the  
 398 timing and magnitude of subsurface recharge.

400

#### 401 4.2. Hydrological signatures of rainfall, surface water, and groundwater sources

402 Pond water and rainfall exhibit similar spatial isotopic patterns, with more positive  $\delta^2\text{H}$  values ( $\delta^2\text{H} >$

403  $-55\text{‰}$ ) than spring water, pore water, and fissure water (Fig. 5a, b). These values are line with the notion  
 404 of direct rainfall and Hortonian runoff are the primary source of pond water. In contrast, the  $\delta^2\text{H}$  values  
 405 of pore, spring, and fissure water show little seasonal variation and are consistently more negative ( $\delta^2\text{H}$   
 406  $< -55\text{‰}$ ), than mean rainfall, indicating longer residence times and reduced evaporative influence.



407  
 408 Fig. 5. The spatial distributions of  $\delta^2\text{H}$  values during the (a) rainy season and (b) dry season for rainfall,  
 409 pond water, spring, pore water, and fissure water in the gully region of the Loess Plateau. To highlight  
 410 spatial differences among water sources,  $\delta^2\text{H}$  values were classified into five intervals:  $< -70\text{‰}$ ,  $-70$  to  
 411  $-65\text{‰}$ ,  $-65$  to  $-60\text{‰}$ ,  $-60$  to  $-55\text{‰}$ , and  $> -55\text{‰}$ . Sampling points are color-coded by water type: red  
 412 for rainfall, purple for pond water, blue for pore water, green for spring water, and orange for fissure  
 413 water.

414

415 The  $\delta^2\text{H}$ - $\delta^{18}\text{O}$  relationships and box plots for each water source reveal key insights into the  
 416 dominant hydrological processes occurring during the rainy season (Fig. 6). Firstly, rainfall follows a  
 417 Local Meteoric Water Line (LMWL) of  $\delta^2\text{H} = 7.7 \cdot \delta^{18}\text{O} + 8.9$  ( $R^2 = 0.95$ ), which is closely aligned, though  
 418 slightly offset, from the Global Meteoric Water Line (GMWL:  $\delta^2\text{H} = 8 \cdot \delta^{18}\text{O} + 10$ ) (rainy season; Fig. 6a,  
 419 c). This alignment indicates that precipitation in the region has a typical meteoric origin. Additionally,  
 420 minimal evaporative enrichment occurred prior to collection. The relatively wide interquartile range of  
 421 rainfall  $\delta^{18}\text{O}$  values suggests that precipitation was derived from storm systems with considerable  
 422 isotopic variability, reflecting differences in rainfall intensity, air mass origin, and temperature. However,  
 423 this variability remains moderate compared with global patterns that span extreme rainfall events and  
 424 broader climatic gradients.

425 Pond water, in contrast, exhibits a clear evaporation signature with a shallower slope of  $\delta^2\text{H} =$   
426  $5.6 \cdot \delta^{18}\text{O} - 17.1$  (rainy season;  $R^2 = 0.95$ ),  $\delta^2\text{H} = 4.6 \cdot \delta^{18}\text{O} - 20.7$  (dry season;  $R^2 = 0.74$ ). This signature  
427 aligns with expectations for surface water bodies, where open exposure facilitates fractionation. The box  
428 plot shows strong evaporative enrichment, with median values shifted significantly toward more positive  
429  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  compared to rainfall (Fig. 6a, c). Pond water maintains a relatively consistent slope and  
430 range across seasons, reinforcing its stable evaporative signature and less dynamic recharge behavior.

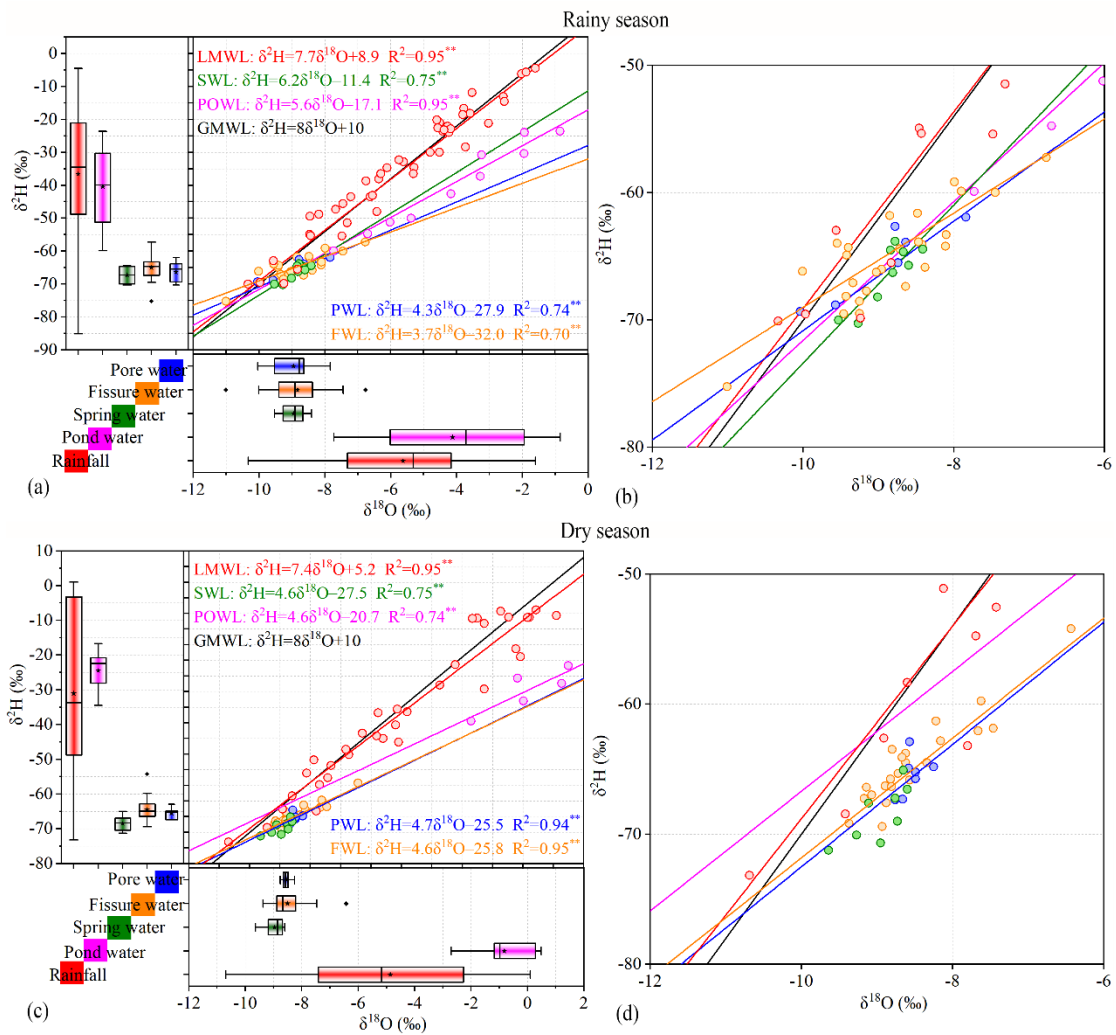
431 Spring water shows a clear seasonal transition in its isotopic composition. In the rainy season, its  
432 evaporation line ( $\delta^2\text{H} = 6.2 \cdot \delta^{18}\text{O} - 11.4$ ;  $R^2 = 0.75$ ) falls closer to the LMWL, suggesting that spring  
433 discharge is augmented by recent rainfall, likely delivered through rapid infiltration and shallow  
434 subsurface flow pathways during high-intensity events (Fig. 6a, b). However, the isotopic values of  
435 spring water are substantially more depleted than those of precipitation, indicating that older water stored  
436 in the porous subsurface aquifer dominates the overall spring flow composition composed of new and  
437 relatively old water.

438 During the dry season, the isotopic slope flattens and deviates further from the Local Meteoric Water  
439 Line (LMWL), reflecting increased evaporative influence or prolonged residence times (Fig. 6c, d). This  
440 seasonal shift suggests that as rainfall inputs decline, spring discharge becomes increasingly composed  
441 of slow-draining, older water that has undergone greater isotopic modification, either through mixing or  
442 evaporation in near-surface storage zones. Collectively, these patterns suggest that spring water acts as a  
443 dynamic integrator of recharge processes, rapidly responding to event-driven infiltration during the rainy  
444 season, yet also reflecting the delayed mobilization of older water stored in the subsurface. This behavior  
445 may be partly explained by a piston-like displacement mechanism, where incoming rainfall pushes pre-  
446 existing groundwater toward discharge zones.

447 Pore and fissure water show remarkably similar isotopic signatures during the rainy season. Pore  
448 water, again sampled from a porous subsurface aquifer, follows a fitted line of  $\delta^2\text{H} = 4.3 \cdot \delta^{18}\text{O} - 27.9$   
449 (rainy season;  $R^2 = 0.74$ ), while fissure water, likely drawing from the same aquifer but through  
450 weathered bedrock pathways, fits  $\delta^2\text{H} = 3.7 \cdot \delta^{18}\text{O} - 32.0$  (rainy season;  $R^2 = 0.70$ ). These slopes are  
451 significantly flatter than those of rainfall, pond, or spring water, a pattern interpreted as evidence of  
452 evaporation prior to recharge or mixing with evaporated surface water. However, the box plots of the  
453 isotope data present a different picture. Both pore and fissure waters are systematically more depleted in  
454  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  than precipitation, and their narrow interquartile ranges suggest a relatively uniform

455 isotopic composition (Fig. 6a-d).

456 These depleted and uniform isotopic compositions indicate recharge dominated by isotopically light  
457 rainfall events, primarily intense summer monsoon storms, rather than evaporative enrichment. Depleted  
458 groundwater signatures thus reflect recharge from these summer events, with percolation delayed to  
459 cooler seasons due to soil storage and reduced evaporation. The flat regression slopes and narrow  
460 interquartile ranges are reconciled by longer residence times and mixing in the subsurface aquifer, which  
461 acts as a hydrological buffer that dampens seasonal isotopic variability (Table A2).

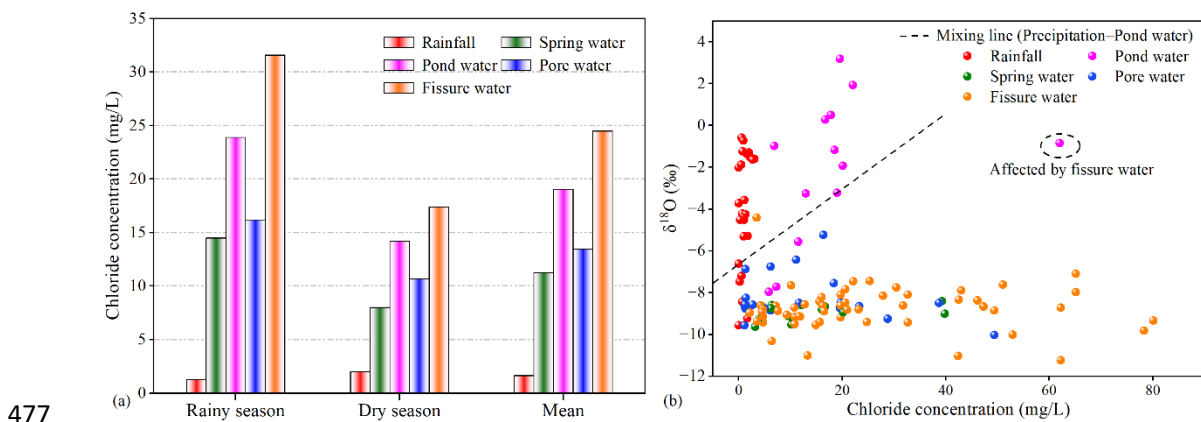


462  
463 Fig. 6. Dual stable isotopic compositions of rainfall, pond water, spring water, pore water, and fissure  
464 water during the rainy season and dry season in the gully region of the Loess Plateau. The black line  
465 represents the global meteoric water line (GMWL,  $\delta^2\text{H}=10 + 8\delta^{18}\text{O}$ ). GMWL is the global meteoric water  
466 line of Craig, LMWL is the local meteoric water line, SWL is the spring water line, POWL is the pond  
467 water line, FWL is the fissure water line, and PWL is the pore water line. Panels (b) and (d) are magnified  
468 views of (a) and (c), respectively, highlighting the isotopic compositions of pore water, fissure water, and

469 spring water (x-axis:  $-12$  to  $-6\%$ ; y-axis:  $-80$  to  $-50\%$ ).

470

471 Complementing the isotope data,  $\text{Cl}^-$  levels in pore water consistently fall between those of  
472 precipitation and pond water across both seasons (Fig. 7a), and the correlation pattern between chloride  
473 concentration and  $\delta^{18}\text{O}$  supports a mixed recharge origin for pore water (Fig. 7b). This trend aligns with  
474 the isotopic evidence from the rainy season and supports the interpretation that pond water contributes  
475 to pore water recharge via vertical percolation through the vadose zone, particularly during high-rainfall  
476 periods when infiltration capacity is exceeded.

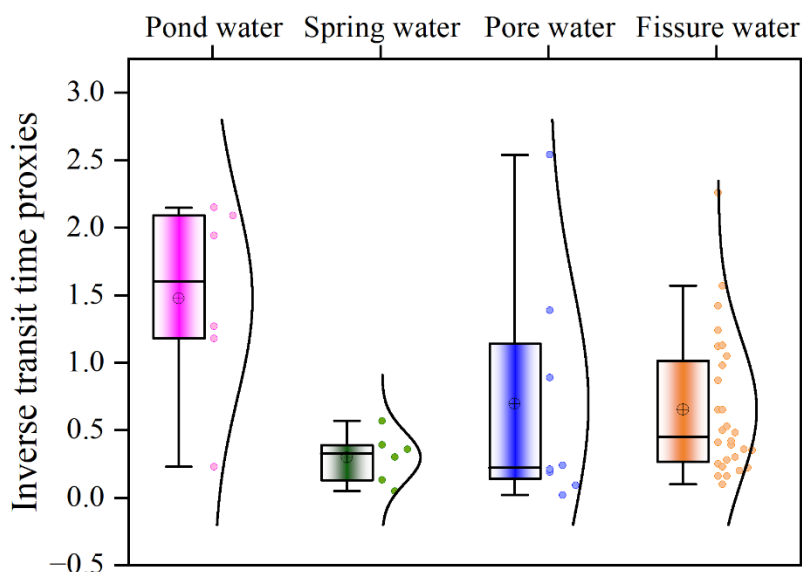


477

478 Fig. 7. Chloride concentration of various water sources in the rainy and dry seasons (a), and the spatial  
479 relationship between chloride concentration and  $\delta^{18}\text{O}$  for different water sources (b).

480

481 The inverse transit time proxies (ITTPs) broadly support the dual-isotope interpretations of water  
482 source dynamics. Pond water exhibited the highest ITTP values ( $1.5 \pm 0.7$ ), indicating rapid turnover and  
483 limited subsurface storage. These elevated values likely reflect inputs from direct rainfall and overland  
484 flow, as well as evaporative enrichment, which increases isotopic variability and can artificially shorten  
485 the apparent residence time. In contrast, pore water ( $0.7 \pm 0.3$ ) and fissure water ( $0.6 \pm 0.5$ ) showed lower  
486 ITTPs, consistent with longer residence times, greater subsurface mixing, and attenuation of seasonal  
487 isotopic signals due to delayed recharge. Spring water had the lowest ITTPs ( $0.3 \pm 0.2$ ), reflecting slow  
488 subsurface transport and integration of older water sources (Fig. 8).



489

490 Fig. 8. Boxplots and kernel density estimates of inverse transit time proxies (ITTPs) for pond water,  
 491 spring water, pore water, and fissure water. Higher ITTP values indicate shorter water transit times since  
 492 precipitation, while lower values suggest longer residence and greater isotopic damping. Pond water  
 493 exhibited the highest and most consistent ITTPs (median  $\approx 1.5$ ), implying rapid recharge from recent  
 494 rainfall or stormflow. Spring water showed the lowest ITTPs ( $\approx 0.3$ ), consistent with longer subsurface  
 495 flow paths and storage. Pore and fissure water displayed intermediate and more variable ITTPs, reflecting  
 496 mixing between recent and older water sources, as well as seasonal differences in infiltration and soil  
 497 moisture replenishment.

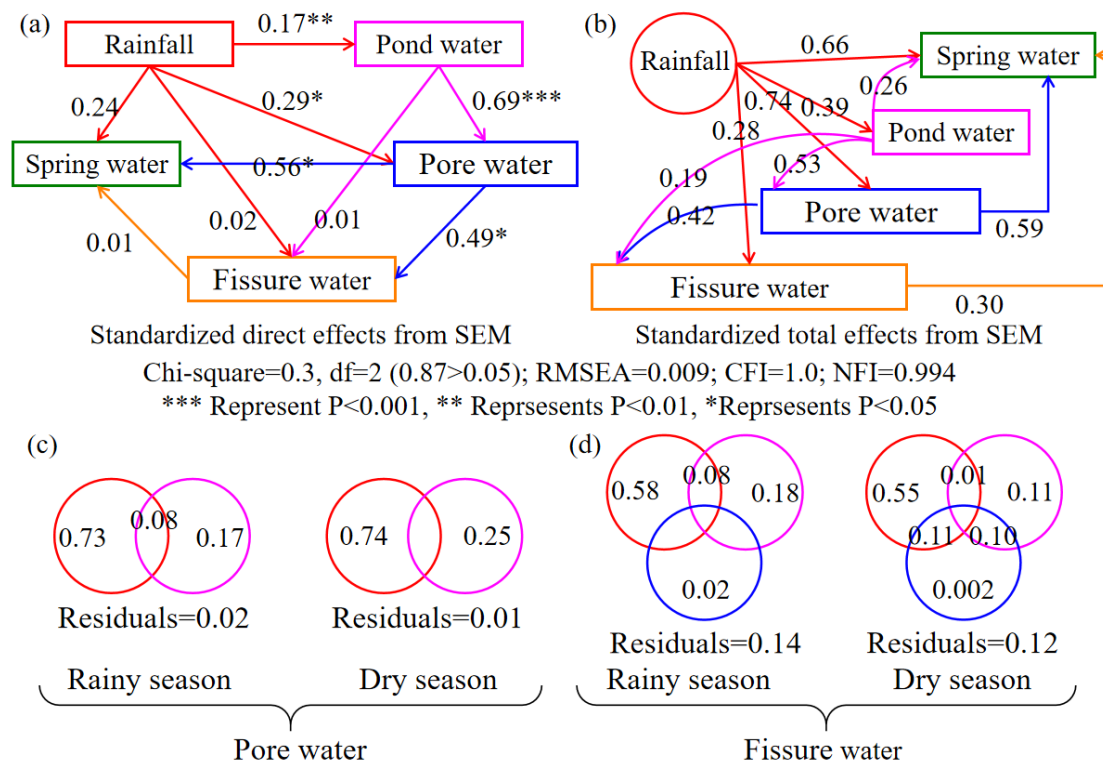
498

499 **4.3. Hydrological linkages and recharge efficiency**

500 The SEM analysis reveals significant hydrological linkages among different water bodies in the  
 501 catchment, with particularly well-defined pathways connecting rainfall, pond water, pore water, and  
 502 fissure water (Fig. 9a, b). Rainfall contributes over 73% to pore water recharge, far exceeding the <17%  
 503 contribution from pond water (Fig. 9c). However, the SEM results indicate that the direct effect of pond  
 504 water on pore water is stronger than that of rainfall (Fig. 9b). This apparent contradiction likely stems  
 505 from the strong statistical association between rainfall and pond water, as pond water is primarily derived  
 506 from rainfall and shares similar isotopic signatures.

507 These findings underscore the importance of integrating multiple lines of evidence rather than  
 508 relying solely on SEM outputs. For example, chloride concentrations in pore water more closely resemble  
 509 those of pond water, suggesting that pond water may contribute to pore water recharge under specific

510 spatial or temporal conditions (Fig. 7a). The spatial distributions of chloride concentration and  $\delta^{18}\text{O}$   
 511 further show that part of the pore water plots near the precipitation–pond water mixing line, providing  
 512 evidence that pond water can contribute to pore water recharge, particularly in localized areas or during  
 513 discrete recharge events (Fig. 7b). At deeper levels, the linkage between pore water and fissure water is  
 514 supported by their nearly identical isotope values and similar ITTPs, suggesting a shared subsurface  
 515 origin and a strong hydrological connection (Fig. 9a, b). In contrast, although there is some hydrological  
 516 connectivity between pore water and spring water, differences in isotopic slopes and residence times may  
 517 lead to an overestimation of their interaction. However, their nearly identical spatial distributions of  
 518 chloride concentrations and  $\delta^{18}\text{O}$  provide more direct and reliable evidence of connectivity.



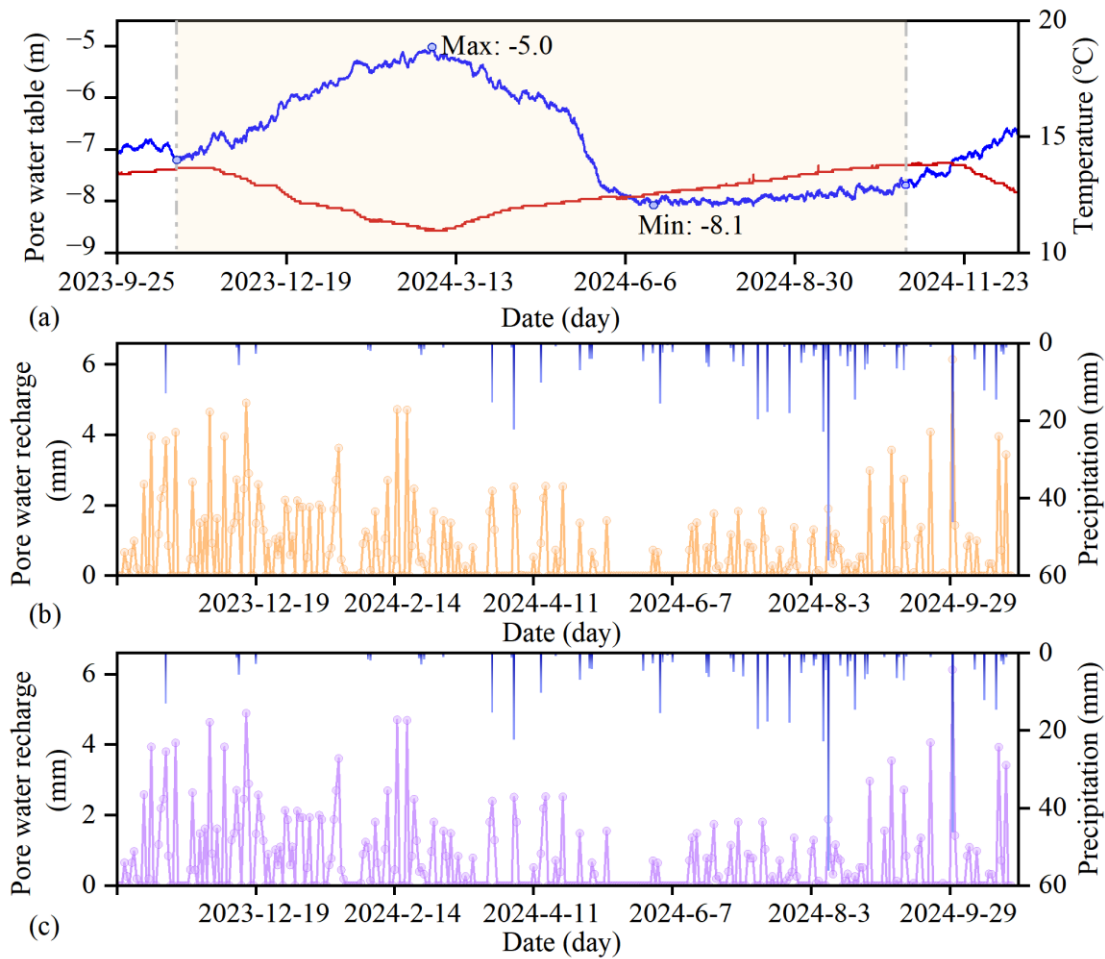
519  
 520 Fig. 9. Structural equation modeling (SEM) and variance partitioning results illustrating hydraulic  
 521 connectivity among water sources in the gully region of the Loess Plateau. Panels (a) and (b) show the  
 522 standardized direct (a) and total effects (b) among rainfall, pond water, pore water, spring water, and  
 523 fissure water, based on  $\delta^{18}\text{O}$  and  $\delta^2\text{H}$  data. In SEM, the total effect includes both direct pathways (a; e.g.,  
 524 rainfall → pore water) and indirect pathways mediated by other variables (b; e.g., rainfall → pond water  
 525 → pore water). Arrows indicate hypothesized water flow pathways, with line thickness proportional to  
 526 effect size. Asterisks denote statistical significance (\*P < 0.05, \*\*P < 0.01, \*\*\*P < 0.001). The model fit  
 527 is excellent ( $\chi^2 = 0.3$ , df = 2, RMSEA = 0.009, CFI = 1.0, NFI = 0.994), supporting the robustness of these

528 inferred connections. Panels (c) and (d) present variance partitioning results showing the relative  
529 contributions of source waters to pore water and fissure water during the rainy and dry seasons,  
530 respectively. In panel (c), rainfall (red) and pond water (pink) explain a large portion of pore water  
531 variability, with some shared explanatory power and modest residuals. In panel (d), fissure water reflects  
532 a more complex origin, with contributions from rainfall (red), pond water (pink), and pore water (blue),  
533 and greater overlap and residuals, especially during the dry season.

534

535         Although the model results initially suggest that rainfall, mediated through pond water, is the  
536 primary source of pore water recharge, discrepancies among the different indicators call for a more  
537 critical interpretation of the evidence. The contradictions observed across isotopic, chloride, and ITTP  
538 data underscore the need for further quantitative validation. To address the contradictions observed in the  
539 SEM and variance partitioning results, we apply the water table fluctuation method to independently  
540 estimate the recharge rate from rainfall to pore water. Groundwater level fluctuations in the gully system  
541 revealed clear seasonal recharge dynamics, with an initial rise in the pore water table beginning on  
542 October 24, 2023, followed by a decline through early spring (March 1, 2024) and a gradual recovery  
543 starting June 20, 2024 (Fig. 10a). Between October 24, 2023, and October 24, 2024, cumulative recharge  
544 was estimated at  $238.0 \pm 6.0$  mm (RISE) and  $241.4 \pm 6.0$  mm (MRC), with 159 and 167 recharge days,  
545 respectively (Fig. 10b, c). Given that annual precipitation totaled approximately 550 mm, “site-scale”  
546 recharge was approximately 43–44%, underscoring the significance of focused infiltration in sustaining  
547 shallow aquifer recharge within the gully environment.

548         Site-scale recharge is lower than the precipitation-to-pore water contribution estimated by the  
549 variance decomposition method and may more accurately reflect “actual” recharge, as statistical  
550 estimates can be biased by similarities in isotopic signatures. Combined with dual-isotope and SEM  
551 analyses, the WTF results support a conceptual model where storm runoff is captured and redistributed  
552 through loess soils and retention structures, enabling both shallow and deeper subsurface flow. These  
553 flow paths link pore water, fissure water, and spring discharge across the complex gully landscape,  
554 reflecting both vertical and lateral connectivity within the groundwater system.



555 — Pore water table — Temperature — Precipitation — RISE — MRC  
 556 Fig. 10. Temporal dynamics of pore water table depth, temperature, precipitation, and recharge in the  
 557 gully region of the Loess Plateau. (a) Daily time series of pore water table depth (blue line) and surface  
 558 temperature (red line) from September 2023 to November 2024. The water table fluctuates seasonally,  
 559 rising from  $\sim -8.1$  m in late summer to a maximum of  $\sim -5.0$  m in early spring (March 2024), indicating  
 560 delayed infiltration and cool-season recharge. (b) Daily precipitation (blue bars) and modeled pore water  
 561 recharge estimates using the MRC methods. (c) Daily precipitation (blue bars) and modeled pore water  
 562 recharge estimates using the RISE methods. Most recharge events occur from October to April, even  
 563 when rainfall is not especially high, while warm-season precipitation contributes little to recharge, likely  
 564 due to increased evaporative losses and shallow soil retention. Together, these patterns suggest strong  
 565 seasonal control on recharge processes, with effective infiltration primarily occurring during cooler, low-  
 566 evaporation periods.

567

## 568 **5. Discussion**

### 569 **5.1. Isotopic compositions of various water bodies in the gully region**

570 In hydrological studies, the isotopic composition of water bodies reflects both sources and changes  
571 in hydrological processes (Wan and Liu, 2016; Kumar et al., 2019; Dasgupta et al., 2024). Precipitation,  
572 surface water, and groundwater usually exhibit differences in isotopic characteristics due to variations in  
573 evaporation, infiltration, and mixing (Gleeson et al., 2016; Kuang et al., 2019; Al-Oqaili et al., 2020).  
574 However, similar isotopic distributions among different water bodies often indicate a strong hydrological  
575 connection in their water sources (Yang and Wang, 2023).

576 Our study found that rainfall and pond water have similar spatial isotopic distribution patterns,  
577 indicating that pond water primarily originates from rainfall. This reflects the region's geographical and  
578 hydrological characteristics. In this severely eroded gully region, the local government has constructed  
579 an extensive network of ponds and check dams to capture hillside runoff (Liu et al., 2017; Xue et al.,  
580 2025). As a result, most precipitation from the hills converges into these gully ponds and check dams. A  
581 lower pond water line slope indicates evaporation fractionation during retention. Evaporation  
582 preferentially removes lighter isotopes ( $^1\text{H}$  and  $^{16}\text{O}$ ), enriching heavier isotopes ( $^2\text{H}$  and  $^{18}\text{O}$ ) and shifting  
583 the pond water's isotopic composition from its precipitation source (Aragu et al., 1998; Zhang and Wu,  
584 2009; Gleeson et al., 2016). This similarity is primarily due to the fact that pond water originates from  
585 rainfall and its runoff in the hilly-gully region (Liu et al., 2011; Ji et al., 2024). Additionally, the isotopic  
586 values of most groundwater in the gully areas are more depleted compared to those of rainfall and pond  
587 water, likely due to the recharge mechanisms and residence times of different groundwater types, and the  
588 inherent isotopic characteristics of their primary recharge sources (Ouali et al., 2024). The depleted  
589 signatures in groundwater reflect preferential capture of isotopically light summer monsoon events, with  
590 effective percolation delayed to cooler seasons due to transient soil storage and minimized evaporation,  
591 consistent with observed water table rises predominantly from October to April. Nevertheless, these  
592 values fall within the range of precipitation isotopic values, leaning towards the more negative end. This  
593 suggests two complementary mechanisms: (1) the thin unsaturated zone (<10 meters) provides  
594 preferential pathways for rapid infiltration of precipitation, minimizing evaporative fractionation, and (2)  
595 groundwater is likely recharged primarily by intense precipitation events (e.g., summer storms) with  
596 inherently more negative isotopic signatures (Liu, 2024). Together, these processes explain the observed  
597 isotopic characteristics of groundwater.

598 To further investigate the complexity of different types of groundwater in the gully area, we  
599 conducted hydrological and geological surveys, collecting water samples from spring water, pore water,  
600 and fissure water. The results show that the isotopic values of spring water, pore water, and fissure water  
601 are closely clustered, indicating a strong interconnection among the different types of groundwater within  
602 the hydrological cycle. This is likely due to their shared geological and hydrological environments  
603 (Bouwer, 2002; Li et al., 2021; Zhang et al., 2022). Our study found that the water line slope of pore  
604 water and fissure water was higher in the dry season than in the rainy season, with values falling between  
605 the slopes of pond water from the rainy and dry seasons. To investigate the cause of these results, we  
606 analyzed groundwater level dynamics, which showed that water tables were lower at the end of the rainy  
607 season but rebounded afterward, making dry-season tables higher. This suggests that rainy-season  
608 groundwater mixes with evaporatively enriched water, lowering its slope, while dry-season groundwater  
609 is recharged by delayed rainfall and pond water, increasing its slope. These findings reveal that isotopic  
610 composition is influenced by both current and prior hydrological conditions. They also highlight the  
611 complexity of evaporation fractionation regulated by water mixing and demonstrate the significant  
612 impact of pond construction on groundwater recharge and regional hydrology in the gully regions.

613

## 614 **5.2. Groundwater recharge processes in the gully region**

615 In recent years, discussions of groundwater recharge sources on the Loess Plateau have largely  
616 focused on tableland and hilly areas characterized by thick loess deposits, whereas gully regions have  
617 received comparatively limited attention (Li et al., 2017; Xiang, 2020; Lu, 2020). For instance, Liu et al.  
618 (2011) demonstrated that groundwater near valley bottoms in hilly loess areas can be replenished by a  
619 combination of precipitation, runoff, and surface water. Our results are broadly consistent with these  
620 earlier findings, but extend them by providing multiple lines of site-specific evidence. Based on stable  
621 isotope signatures and chloride concentrations, we independently identify precipitation and surface water  
622 as the primary sources of groundwater recharge in gully systems. Furthermore, by applying a structural  
623 equation model (SEM), we quantitatively evaluate the relative importance of different recharge pathways,  
624 demonstrating that surface water (particularly pond water) plays a key mediating role in transferring  
625 precipitation inputs to subsurface pore water. Building on these results, we classify groundwater in the  
626 study area into three functional types, spring water, pore water, and fissure water, and propose a  
627 progressive, multi-stage recharge framework: (1) direct recharge of pond water by precipitation and

628 indirect recharge of pore water by precipitation; (2) focused recharge from pond water to pore water; and  
629 (3) downward percolation from pore water to fissure water. This framework highlights the complexity of  
630 groundwater flow and recharge processes in gully-dominated landscapes and underscores the significant  
631 influence of human interventions, such as ponds and check dams, on modifying hydrological  
632 connectivity and recharge dynamics.

633 In the deep-profile unsaturated zones of the hilly region on the Loess Plateau, previous studies have  
634 used chloride mass balance and tritium peak methods to estimate groundwater recharge from  
635 precipitation, typically accounting for 2%–22% of annual rainfall, with water residence times in the  
636 unsaturated zone lasting several years or even hundreds of years (Huang and Pang, 2011; Tan et al., 2016;  
637 Li et al., 2017; Wang et al., 2024). However, these studies did not address the catchment role of gully  
638 regions. Field observations and past studies show that precipitation rarely directly infiltrates thick loess  
639 in hilly areas (Xu et al., 1993; Li, 2001). Instead, it forms surface runoff that converges into engineered  
640 gullies and accumulates in ponds or other water bodies (perched water), serving as a concentrated  
641 recharge source for groundwater (Yu et al., 2025), reflecting the sustained and delayed impact of gully  
642 runoff on groundwater recharge, which is consistent with the results of this study. In summary, while  
643 hillslope-scale studies describe a “dispersed recharge” mode, where precipitation percolates slowly  
644 through thick unsaturated zones, this study identifies a “concentrated recharge” mode in engineered  
645 gullies, driven by runoff convergence and regulated by check dams via ponding. These fundamentally  
646 distinct modes, differing in hydrological processes, spatial scales, and recharge efficiencies, collectively  
647 enhance the understanding of groundwater recharge mechanisms on the Loess Plateau.

648 Notably, our study does not consider confined water. Tan et al. (2016) indicated that the groundwater  
649 in the high mountain-hilly loess aquifer does not originate from the upwelling of ancient regional  
650 groundwater, and there is no evidence of deep confined water beneath the loess strata in the high  
651 mountain-hills. Additionally, our findings represent only the groundwater recharge results in the gully  
652 regions for two reasons: (1) The hydrological system is complex, with significant variations across  
653 different landscapes of the Loess Plateau (Li et al., 2019; Li et al., 2021). For example, Li et al. (2019)  
654 found that groundwater dominates the hydrological system in the tableland on the Loess Plateau, where  
655 surface water (streams) is recharged by groundwater because river channels are deeper than the bedrock.  
656 (2) Our data collection focused on gully because the “Loess Liang” and “Loess Mao” hillside areas are  
657 covered by thick loess with minimal water sources.

658

### 659 **5.3. Groundwater recharge rates in the gully region**

660 In many parts of the world, identifying the sources of groundwater recharge and its renewability is  
661 essential for effective water resource management (Ajjur and Baalousha, 2021; Meles et al., 2024). In  
662 the hilly-gully region of the Loess Plateau, where groundwater is considered a crucial source of safe  
663 water, understanding the origins and recharge of aquifers provides valuable information for water  
664 resource planners (Wang et al., 2006; Liu et al., 2011; Wang et al., 2024). This knowledge is essential  
665 and should be shared with regions facing similar challenges.

666 Groundwater recharge can be quantified from three hydrological sources: surface water, the  
667 unsaturated zone, and the saturated zone (Scanlon et al., 2022). Recharge estimates based on the saturated  
668 zone are generally more reliable than those from the unsaturated zone, as the latter represents potential  
669 recharge that may not ultimately reach the groundwater table (Beven and Germann, 2013; Huang and  
670 Shao, 2019). The groundwater table fluctuation method is widely used for estimating saturated zone  
671 recharge due to its high temporal resolution and intuitive nature (Gumuła-Kawęcka et al., 2022; Xu et  
672 al., 2024). In our study area, ITTPs estimated similar transit times for both pore water and fissure water.  
673 Therefore, we used the groundwater table fluctuation method to assess the recharge of pore water in the  
674 gully region. The total recharge from 2023 to 2024 was estimated at  $241.4 \pm 6.0$  mm and  $238 \pm 6.0$  mm  
675 using the MRC and RISE methods, respectively. Under constant specific yield conditions, the MRC  
676 method typically estimates higher groundwater recharge and recharge days than RISE, as it accounts for  
677 groundwater table decline due to lateral outflow and other discharge processes in the absence of recharge  
678 (Heppner and Nimmo, 2005). Our findings support this pattern. Furthermore, the key parameter for  
679 estimating groundwater recharge using the water table fluctuation method is specific yield ( $S_y$ ), which  
680 depends on soil properties and water table depth (Liang et al., 2015). Shallow soil measurements (0–50  
681 cm) using the test pit method (total porosity minus field capacity) yielded  $S_y \approx 0.03$ , consistent with high  
682 capillary retention in near-surface loess (Wang et al., 2024). However, for water tables deeper than 2 m  
683 (as in this study, typically 4–10 m), the test pit method provides a reliable estimate of aquifer-scale  
684 drainable porosity (Nachabe, 2002; Shah and Ross, 2009; Liang et al., 2015). Accordingly, we adopted  
685  $S_y = 0.032$ , aligned with values of  $\sim 0.03$  reported for similar loess-derived unconfined aquifers on the  
686 Loess Plateau (Wang et al., 2023). Sensitivity analysis indicates that recharge estimates vary by  
687 approximately  $\pm 25\%$  across the plausible  $S_y$  range ( $0.032 \pm 0.008$ ), reflecting uncertainty in effective

688 drainable porosity within shallow gully aquifers.

689 Research on groundwater recharge in the Loess Plateau has mainly focused on deep-profile  
690 unsaturated zones in the tableland and hilly areas, with tracer methods estimating recharge between 9 to  
691 100 mm (Huang and Pang, 2011; Li et al., 2017; Xiang et al., 2019; Lu, 2020; Wang et al., 2024). In  
692 contrast, our study in the gully region indicates recharge of up to 240 mm, much higher than previous  
693 estimates on deep-profile unsaturated zones. This difference reflects several factors: 1) Unsaturated zone  
694 thickness: In the gully region, the unsaturated zone is generally less than 10 m thick, much shallower  
695 than in tableland and hilly areas (mean thickness of 92.2 m), making infiltration easier and promoting  
696 effective recharge. 2) Gully topography and hydrology, characterized by well-developed channels,  
697 concentrated runoff, and widespread ponds and check dams, promote focused infiltration (Liu et al., 2017;  
698 Li et al., 2021; Xue et al., 2025). 3) Research methods: Tracer methods reflect long-term recharge rates  
699 and are better suited for thicker unsaturated zones (Huang and Pang, 2011; Lu, 2020; Li et al., 2017). In  
700 contrast, the water table fluctuation method directly captures short-term recharge dynamics and works  
701 better in thinner unsaturated zones. Moreover, this method also better captures surface water-  
702 groundwater interactions and focused recharge effects (Gumuła-Kawęcka et al., 2022). These findings  
703 underscore the importance of studying recharge in gully regions, filling a research gap in the Loess  
704 Plateau's geomorphology and providing new ecohydrological insights. However, the robustness of our  
705 findings requires further exploration. On one hand, due to the limited spatial distribution of sampling  
706 points, the current results primarily reflect the hydrological characteristics of engineered gullies, and  
707 their representativeness at the regional scale requires validation through future expansion of the  
708 monitoring network. On the other hand, the study period did not encompass extreme precipitation or  
709 drought events, which may significantly alter surface flow convergence conditions and vadose zone  
710 water transport mechanisms, thereby substantially impacting recharge processes. Future work should  
711 strengthen dynamic monitoring and simulation analysis under extreme hydrological scenarios.

712

#### 713 **5.4. Revised conceptual model**

714 To convey our evolving understanding of the spatial structure and dynamics in the Gully Region,  
715 we developed a conceptual model that traces precipitation's transformation into subsurface water, from  
716 runoff capture and surface ponding in dammed gully reaches, through infiltration in the unsaturated zone,  
717 to recharge in both shallow porous aquifer and deeper bedrock fissure systems (Figure 11). This

718 conceptual reframing is grounded in the stark contrasts between hydrological processes active on hilly  
719 uplands and managed gully systems. In the hilly uplands, previous studies have shown that thick loess  
720 deposits, often exceeding 90 m (including low-permeability aquifers), combined with steep slopes ( $>15^\circ$ )  
721 severely restrict vertical infiltration (Zhu et al., 2018; Huang et al., 2019; Huang et al., 2024).  
722 Compounded by short-duration, high-intensity rainfall events that provide insufficient moisture for deep  
723 profile wetting, rainfall is converted rapidly into surface runoff (Li et al., 2021). Our new work shows  
724 that runoff is systematically funneled downslope into gully systems as a consequence of ecological  
725 engineered check dams and retention ponds that intercept and concentrate overland flow. Most  
726 infiltration occurs after surface water accumulates in fill zones of engineered gullies. Ponds and perched  
727 water bodies subsequently serve as localized recharge foci.

728 Crucially, gully systems possess distinct hydrogeological characteristics: the loess mantle is much  
729 thinner (typically  $< 25$  m), and the soils are dominated by silt loam textures with moderate specific yield  
730 (0.02–0.05) and high field capacity (21–28%). These properties promote transient water storage and  
731 enable temporally delayed and depth-partitioned infiltration. Based on our integrated analyses of stable  
732 isotopes, chloride concentrations, and inverse transit time proxies, we find that engineered gullies  
733 function not as passive erosional features but as active, managed recharge conduits. This  
734 conceptualization captures a critical spatial transition, from runoff generation in the hilly uplands to  
735 focused recharge in gully zones, emphasizing the pivotal role of gully systems in regulating groundwater  
736 recharge across the Loess Plateau landscape.

737 Combined hydrological monitoring and multi-indicator analysis further reveal that following the  
738 rainy season, infiltration depths on hilly slopes are typically shallow (less than 1 m), while groundwater  
739 levels in gully areas exhibit pronounced rises exceeding 2 m (Fig. 11). Recharge estimates based on the  
740 water table fluctuations reach up to approximately 240 mm at the monitored gully reach, far surpassing  
741 values observed in deep unsaturated zones of tablelands and hills (Huang and Pang, 2011; Li et al., 2017;  
742 Lu, 2020; Wang et al., 2024). The results of this study reinforce the role of engineered gully reaches as  
743 focal points for groundwater recharge and further quantify site-scale pore-water recharge equivalent to  
744  $\sim 43\%$  of mean annual precipitation, a finding that highlights the efficiency of focused infiltration under  
745 managed conditions. This value reflects spatially focused recharge under conditions of runoff  
746 convergence and ponding, and should not be interpreted as representative of hillslope or catchment-wide  
747 recharge rates.

748 Liu et al. (2011) found that groundwater near valleys in the hilly loess area is replenished by  
749 precipitation, runoff, and surface water. Moreover, fissure water exhibits more depleted isotopic  
750 signatures and higher chloride concentrations, indicating deeper percolation of pore water or mixing with  
751 older recharge sources (Fig. 11). These patterns, supported by ITTPs and statistical (SEM-based)  
752 connectivity indicators, reveal a hierarchical recharge sequence: event-driven infiltration enters a porous  
753 shallow aquifer, some of which slowly percolates into deeper fissure zones. This hierarchical mechanism  
754 is facilitated by the combination of thin loess mantles, engineered interventions (e.g., check dams and  
755 ponds), and delayed hydrological responses.

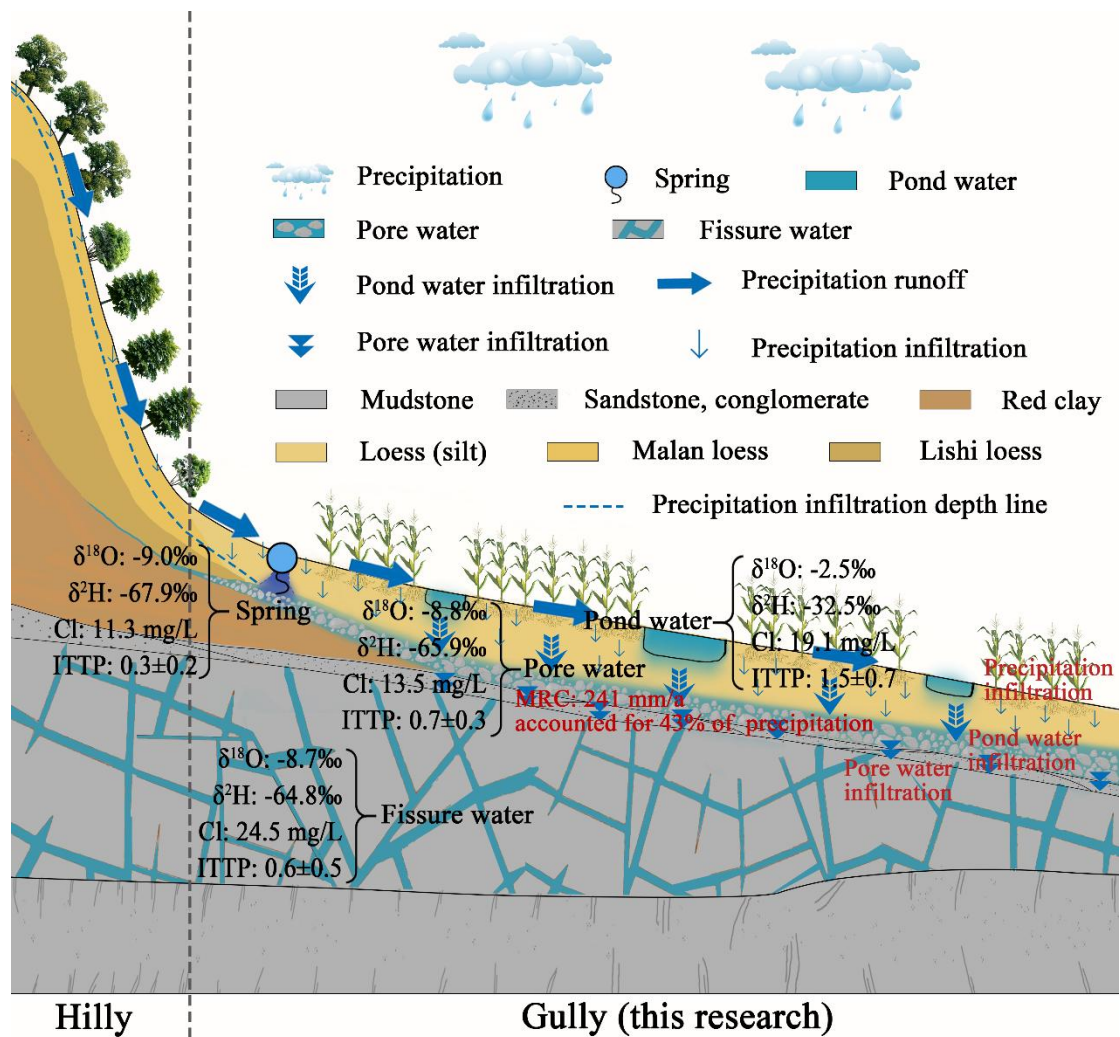
756 By integrating multiple lines of evidence, this conceptual model redefines engineered gullies as  
757 selective recharge corridors whose hydrological function emerges from the interaction between  
758 geomorphic structure and human intervention. It challenges the traditional view of gullies as purely  
759 erosional landforms and emphasizes their dual hydrological function: acting both as runoff conveyance  
760 channels and as transient reservoirs that store and redistribute water across space and time. This recharge  
761 capacity is jointly governed by topographic convergence, reduced loess thickness, and the presence of  
762 engineered structures such as check dams and retention ponds that increase residence time.

763 Crucially, the model offers insight into the multifunctionality of ecological engineering, particularly  
764 check dams and ponds, in enhancing groundwater recharge, and supporting ecosystem restoration across  
765 the Loess Plateau. This study proposes a cascade-type recharge framework for engineered gully systems,  
766 highlighting the role of engineered gullies as convergence pathways that locally focus infiltration and  
767 groundwater recharge. Rather than invoking preferential flow within the soil matrix, this framework  
768 emphasizes topographic convergence, stratigraphic thinning, and engineered ponding as the dominant  
769 mechanisms that promote spatially concentrated recharge within gully zones. While this process is  
770 demonstrated using site-specific tracer and water-table observations, its broader relevance at the  
771 catchment scale remains conceptual and warrants further investigation. Furthermore, water movement  
772 within the silted loess layer of the gully system remains dominated by a piston flow pattern (Yu et al.,  
773 2025). By identifying the pivotal role of gully systems in stormwater detention, delayed infiltration, and  
774 depth-partitioned recharge, this study establishes a mechanistically grounded conceptual basis improving  
775 water resource allocation, infrastructure planning, and groundwater sustainability in arid and semi-arid  
776 regions.

777 However, with the reconstruction of gully systems and ecological restoration, attention must also

778 be given to the potential risks of pollutant migration (Yu et al., 2020). The hydrological functions of  
779 gullies may enhance the movement of pollutants into groundwater, especially in areas with intensive  
780 human activities, where pollutants can enter engineered gullies through surface runoff and subsequently  
781 infiltrate the groundwater system. During ecological restoration, excessive human intervention or soil  
782 improvement measures may lead to the accumulation and dispersion of pollutants, which may  
783 compromise groundwater security (Liu et al., 2017). Therefore, the protection and rational reconstruction  
784 of gully systems should not only focus on their hydrological functions but also consider potential  
785 environmental risks, particularly the pathways of pollutant migration. These findings therefore  
786 underscore the need to evaluate gully-based restoration strategies within an integrated water-quality and  
787 groundwater-protection framework (Obuobie et al., 2012; Zhao et al., 2019; Zhao and Wang, 2021; Xue  
788 et al., 2025).

789 The study demonstrates how hydrologically arrested gully systems can function as critical “recharge  
790 windows” for groundwater in arid areas. This underscores the importance of strategically identifying and  
791 managing gully networks in watershed management, while avoiding excessive filling or hardening to  
792 preserve their hydrological functions. In ecological restoration projects, directing surface runoff toward  
793 engineered gullies under controlled conditions can efficiently convert limited precipitation into  
794 groundwater storage, thereby enhancing regional water retention capacity. Beyond advancing theoretical  
795 understanding of regional hydrological processes, this conceptual model provides a process-based  
796 foundation for developing spatially targeted models of groundwater recharge in managed dryland  
797 landscapes.



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Fig. 11. Hydraulic connections between different water bodies in the hilly-gully region of the Loess Plateau. The study area consists of hilly and gully regions. In the hilly area, the stratigraphic sequence from top to bottom is Malan loess, Lishi loess, red clay, sandstone, and mudstone. Rainfall infiltration within the Malan loess is less than 1 m, and the area is mainly covered by vegetation. In the gully area, the stratigraphy from top to bottom includes loess (silt), sandstone and conglomerate, and mudstone. Pore water is found within the sandstone and conglomerate, while fissure water occurs in bedrock fractures (mudstone). Numerous check dams or ponds are distributed throughout the gully area. The vertical separation between the pore water and pond water ranges from 3 to 5 m. Corn is the main crop cultivated in this region. Most springs in the study area are located at the junction of the hilly and gully regions and are discharged from pore water.

### 5.5. Limitations and future research directions

This study underscores the limitations of relying on a single indicator to infer localized groundwater

812 recharge pathways, as doing so may lead to oversimplified or potentially misleading interpretations of  
813 complex hydrological processes. While stable isotope signatures suggest precipitation contributes to pore  
814 aquifer recharge, they do not provide clear evidence of direct recharge from pond water to either pore or  
815 fissure groundwater during the dry season. In contrast, the spatial distribution of chloride concentrations  
816 offers compelling support for focused pond water leakage into the shallow groundwater system.

817 Without explicit mass-balance constraints, structural equation modeling may not independently or  
818 quantitatively represent actual groundwater flow processes. In contrast, the water-table fluctuation  
819 method, which directly measures changes in groundwater levels, provides a more empirically grounded  
820 estimate of total recharge. Each approach nevertheless offers distinct strengths: water-table fluctuations  
821 resolve the timing and magnitude of recharge, whereas isotopic, hydrochemical, and modeling analyses  
822 yield critical insights into recharge sources and flow pathways. By leveraging the complementarity and  
823 mutual corroboration of these methods, our study robustly demonstrates the pivotal role of gully areas in  
824 groundwater recharge.

825 Despite effort to address uncertainties, limitations remain in terms of spatial and temporal sampling  
826 density. For example, the lack of long-term tracers such as groundwater age, combined with limited  
827 observations of groundwater level fluctuations, constrains our ability to assess recharge dynamics over  
828 multi-year timescales. Additionally, the current sampling design includes only two campaigns during the  
829 rainy and dry seasons, which may be insufficient to fully capture the seasonal variability of ITTP values.  
830 This, in turn, may affect the accuracy of groundwater renewal frequency estimates and the strength of  
831 inferred hydrological connections. In arid and semi-arid regions, groundwater recharge is typically  
832 triggered by infrequent, high-intensity rainfall events. However, existing sampling strategies based on  
833 seasonal intervals often lack the temporal resolution necessary to capture these short-lived, event-driven  
834 recharge processes effectively.

835 Future research should address these issues through several improvements: (1) conducting higher-  
836 frequency, event-scale sampling to systematically monitor rainfall, spring discharge, and pond water  
837 level dynamics, thus capturing the influence of key hydrological events on recharge processes; (2)  
838 expanding the spatial coverage of pore and fissure well monitoring to improve the accuracy of regional  
839 recharge pattern identification; and (3) incorporating additional environmental tracers (e.g.,  $^3\text{H}$ ,  $^{22}\text{Na}$ ) to  
840 trace flow paths and estimate recharge lag times. In addition, systematic observation of event-scale  
841 hydrological processes should be strengthened by establishing a high-frequency, event-driven monitoring

842 network to better capture the nonlinear coupling among rainfall, surface runoff, and groundwater  
843 dynamics. This approach would significantly improve our understanding of rapid infiltration events and  
844 associated recharge mechanisms.

845 From a methodological perspective, integrating statistical techniques, such as SEM and variance  
846 decomposition analysis, with process-based physical models like MODFLOW and HYDRUS can  
847 provide mechanistically constrained insights into recharge pathways. Compared to correlation-based  
848 statistical methods, physical models offer greater precision in characterizing groundwater flow and  
849 recharge processes across both temporal and spatial dimensions, helping to reduce uncertainties  
850 associated with non-conservative tracer behavior and the absence of mass balance constraints. Regarding  
851 measurements, hydrometric instrumentation within check dams and beneath pond beds could further  
852 quantify the recharge effects of various engineering interventions under specific hydrological conditions.  
853 Additionally, integrating isotopic data with mean transit time modeling, combined with targeted  
854 catchment-wide field monitoring and improved spatial analysis, could help elucidate recharge pathways,  
855 quantify temporal dynamics, and enhance process-level understanding of groundwater recharge  
856 throughout the catchment of such complex dryland landscapes. Collectively, these efforts will contribute  
857 to a stronger theoretical foundation and offer practical guidance for the precise management of water  
858 resources, the design of ecologically appropriate engineering interventions, and the implementation of  
859 effective landscape rehabilitation strategies.

860

## 861 **6. Conclusion**

862 Through integrated analysis of stable isotopes, chloride concentrations, water-table fluctuations, and  
863 inverse transit time proxies, this study provides multiple, convergent lines of evidence that engineered  
864 gully reaches on the Loess Plateau function as hydrologically significant recharge zones, rather than  
865 solely as products of accelerated erosion and degradation. Precipitation-driven runoff supports  
866 substantial recharge to shallow pore aquifers, with site-scale recharge magnitudes equivalent to  
867 approximately 43% of mean annual precipitation at the monitored gully reach. Although evaporative  
868 fractionation limits the ability of stable isotopes alone to resolve direct recharge from ponded surface  
869 water, chloride concentrations provide independent evidence consistent with mixing between pond water  
870 and pore water, complementing the isotopic patterns. Together, these indicators indicate likely hydraulic

871 connectivity, while not constituting a mass-balanced quantification of recharge sources. Recharge within  
872 shallow gully-zone aquifers is spatially concentrated and temporally selective, governed by topographic  
873 convergence, loess stratigraphy, and ecological engineering structures, particularly check dams and  
874 ponds, which increase surface-water residence time and promote focused infiltration.

875 These findings offer a process-based framework for developing hydrological indicators of  
876 landscape function and restoration performance managed in dryland environments. Specifically, recharge  
877 magnitude, isotopic damping, and solute accumulation patterns may serve as diagnostic indicators for  
878 identifying effective recharge zones and tracking hydrological response to intervention. To refine these  
879 indicators, future studies should incorporate high-frequency monitoring, event-based sampling, and  
880 multi-tracer approaches. Collectively, this work challenges conventional views of gullies as hydrological  
881 liabilities and demonstrates context-dependent role as targeted recharge assets in engineered dryland  
882 systems, refining conceptual understanding of dryland groundwater recharge and informing the  
883 evaluation and design of ecological engineering strategies in similar semi-arid landscapes.

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887 **Author contributions:**

888 ZXJ: Conceptualization, Methodology, Formal analysis, Investigation, Visualization, Data curation,  
889 Validation, Writing-original draft. ADZ: Conceptualization, Formal analysis, Validation, Visualization,  
890 Writing-review & editing. LW: Conceptualization, Funding acquisition, Project administration,  
891 Resources, Supervision, Visualization, Writing-review & editing.

892 **Data availability:**

893 All datasets are available at Zenodo: <https://zenodo.org/records/18596600>.

894 **Conflicts of Interest:**

895 The contact author has declared that none of the authors has any competing interests.

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899

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1156 **Appendix A**

1157 1. The specific yield of different soil textures is shown.

1158 Table A1. The specific yield of different texture of soil (adapted from Liang, 2016)

Texture	Average Specific Yield	Minimum Specific Yield	Maximum Specific Yield	Coefficient of Variation (%)
Clay	0.02	0.00	0.05	59
Silt	0.08	0.03	0.19	60
Sandy Clay	0.07	0.03	0.12	44
Fine Sand	0.21	0.10	0.28	32
Medium Sand	0.26	0.15	0.32	18
Coarse Sand	0.27	0.20	0.35	18
Gravelly Sand	0.25	0.20	0.35	21
Fine Gravel	0.25	0.21	0.35	18
Medium Gravel	0.23	0.13	0.26	14
Coarse Gravel	0.22	0.12	0.26	20

1159

1160 2. The isotopic composition ( $\delta^2\text{H}$  and  $\delta^{18}\text{O}$ ) of various water sources in the rainy and dry seasons is shown.

1161 Table A2. Isotopic composition ( $\delta^2\text{H}$  and  $\delta^{18}\text{O}$ ) of various water sources in the rainy and dry seasons

	Rainy season		Dry season	
	$\delta^2\text{H}$	$\delta^{18}\text{O}$	$\delta^2\text{H}$	$\delta^{18}\text{O}$
Rainfall	$-36.6 \pm 20.4\text{‰}$	$-5.6 \pm 2.3\text{‰}$	$-31.0 \pm 23.2\text{‰}$	$-4.9 \pm 3.0\text{‰}$
Pond water	$-40.5 \pm 13.1\text{‰}$	$-4.1 \pm 2.3\text{‰}$	$-24.5 \pm 6.9\text{‰}$	$-0.8 \pm 1.3\text{‰}$
Spring water	$-67.3 \pm 2.6\text{‰}$	$-9.0 \pm 0.4\text{‰}$	$-68.4 \pm 2.2\text{‰}$	$-9.0 \pm 0.4\text{‰}$
Pore water	$-66.3 \pm 3.1\text{‰}$	$-9.0 \pm 0.6\text{‰}$	$-65.4 \pm 3.8\text{‰}$	$-8.5 \pm 0.6\text{‰}$
Fissure water	$-65.0 \pm 3.8\text{‰}$	$-8.8 \pm 0.9\text{‰}$	$-64.5 \pm 5.5\text{‰}$	$-8.5 \pm 0.9\text{‰}$

1162