

Reply to comments on “Measurement report: High contribution of N_2O_5 uptake to particulate nitrate formation in NO_2 -limited urban areas” by Lin et al.

We would like to thank the editor and reviewers for their efforts in handling, reading, and critically reviewing our manuscript, which have helped us to further improve our manuscript. The comments on our paper are carefully addressed.

General Comments:

Lin et al. present an analysis of the controlling factors for particulate nitrate (pNO_3^-) production in Xiamen, Southeast China. Xiamen is notable compared to many other Chinese urban areas because N_2O_5 production there is NO_2 limited, in contrast to the O_3 limited conditions of other regions such as Beijing. They show that under these NO_2 limited conditions, N_2O_5 heterogeneous uptake contributes significantly to pNO_3^- . These findings are significant as the conditions in the study region may be increasingly relevant to other urban areas in China, especially as emissions controls continue to change NO_x , O_3 , and VOC loadings. Relatedly optimal emissions control strategies to reduce pNO_3^- and O_3 can be in conflict as elucidated in box model sensitivity simulations. Overall, this work provides useful new insights into pNO_3^- in the NO_2 limited regime for N_2O_5 production. The analysis is of a high quality, and conclusions are well supported. I believe this work will be a useful addition to the literature and will likely be well suited for publication in ACP following revision and response to the comments below.

Response: We are grateful for your thoughtful comments on the manuscript and we

have made revisions accordingly. Our point-to-point responses to each comment are as follows (reviewer's comments are in black font, our responses are in blue font and our revisions in the manuscript are italic font).

Main Comments:

1. Was aerosol surface area density measured? If so, I would encourage the authors to also present values for the N_2O_5 heterogeneous uptake coefficient ($\gamma_{\text{N}_2\text{O}_5}$) derived from the iterative box model. $\gamma_{\text{N}_2\text{O}_5}$ is known to depend on pNO_3^- concentrations and it could be quite interesting to see if that feedback impacts overall pNO_3^- formation from N_2O_5 . $\gamma_{\text{N}_2\text{O}_5}$ values would also help with interpretation of the analytical results and iterative model skill (e.g. why is kN_2O_5 so much higher in this work than in other urban areas as noted in Line 298, is this due to differences in surface area or $\gamma_{\text{N}_2\text{O}_5}$)

Response: Thanks for your suggestion. Yes, we monitored the aerosol surface area (SA) concentrations in the size range of 7-300 nm under dry conditions. Since we did not apply hygroscopicity parameters to correct the data, the reported SA concentration was underestimated. For the valid observation, the average SA concentration was $110 \mu\text{m}^2/\text{cm}^3$, corresponding to a nighttime average N_2O_5 uptake coefficient ($\gamma_{\text{N}_2\text{O}_5}$) of 0.223. The $\gamma_{\text{N}_2\text{O}_5}$ should be considered as an upper limit, and the actual $\gamma_{\text{N}_2\text{O}_5}$ could be lower. Compared with other Chinese sites ($10^{-2} - 10^{-1}$), this $\gamma_{\text{N}_2\text{O}_5}$ was relatively high (Li et al., 2025). Thus, the high uptake rate of N_2O_5 could be attributed to the elevated $\gamma_{\text{N}_2\text{O}_5}$. Since our findings indicate that kN_2O_5 has a relatively limited impact on pNO_3^- formation compared to PNO_3 , we propose conducting further investigation

into $\gamma\text{N}_2\text{O}_5$ and its feedback with pNO_3^- through targeted case studies in future work.

2. Some additional details on the VOC measurements and the fraction of NO_3 reactivity captured by the measured VOCs would be useful in the main text. Isoprene, styrene, and 2-butene have were shown to dominate VOC nitrate reactivity during winter in Beijing (Hu et al. 2023). Were those same species found to dominate NO_3 reactivity here, and are any unmeasured VOC expected to matter for NO_3 reactivity. More generally how do the specific VOC measured impact the discussion of pNO_3^- response to NO_x and VOCs.

Response: We replied to this comment in the following two aspects.

(1) Thanks for your suggestion, we have provided additional details about the effect of VOCs on NO_3 reactivity in the revised Section 3.3. Based on our observed VOCs (**Table S5**), the NO_3 reactivity (k_{NO_3}) was calculated. The contribution of the observed VOC species to the NO_3 reactivity are presented in **Figure R1 (Figure S10 in the revised supplementary materials)**. Similar to previous observation in Beijing (Hu et al., 2023), the styrene, 2-butene, and isoprene were the dominant VOC species contributing to k_{NO_3} . In this work, we calculated the loss of N_2O_5 , as shown in **eq 4** in the supplementary material, the $k_{\text{NO}_3}/K_{\text{eq}}[\text{NO}_2]$ corresponds to the indirect chemical loss of N_2O_5 through NO_3 chemistry. The reaction rate of $k_{\text{NO}_3}/K_{\text{eq}}[\text{NO}_2]$ was calculated to be 0.000136 s^{-1} , which is much smaller than that of the $k_{\text{N}_2\text{O}_5}$ (0.00764 s^{-1}). This indicates that the influence of VOCs on pNO_3^- formation via N_2O_5 uptake through the consumption of its precursors NO_3 is minor, which supported the SHAP

analysis. For monoterpene species that are highly reactive with NO_3 radicals, no relevant data were available in our study to access their impact on kNO_3 . This limitation likely led to an underestimation of the calculated kNO_3 , as we have highlighted the underestimation in the supplementary material (**line 98-104**).

The supplements of NO_3 reactivity analysis in the main text (**line 319-324**) are as follows:

*“The total concentrations of the observed VOCs (TVOCs) showed a weak negative correlation with N_2O_5 uptake (**Figure 4e**). Similar to existing research (Hu et al., 2023), specific VOC species, such as styrene, 2-butene, and isoprene, can readily consume NO_3 radicals (**Figure S10**), thereby inhibiting N_2O_5 formation. However, the loss of N_2O_5 through the reaction between VOCs and NO_3 was relatively limited compared to its direct uptake, as determined by our calculations (Text S4), which supported the SHAP analysis.”*

(2) The response of pNO_3^- formation to VOCs reduction was considerably weaker than to NO_x variations. Therefore, we did not focus on the detailed effects of individual anthropogenic VOC species on nitrate production. Although unmeasured monoterpene may influence NO_3 reactivity and consequently pNO_3^- formation, these compounds are mainly emitted from biogenic sources, which are difficult to regulate through anthropogenic control. Thus, this aspect was also not discussed in detail in our study. In future work, we will select the periods with substantial indirect loss of N_2O_5 by NO_3 and perform a more detailed analysis of the impact of specific VOC species on pNO_3^- formation.

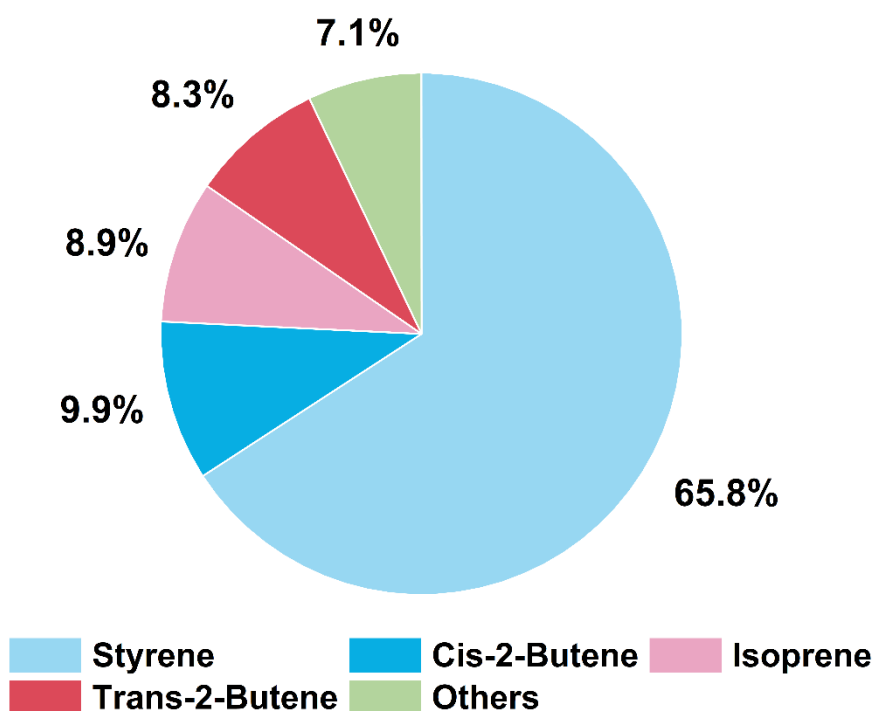


Figure R1. Contribution of observed VOCs to the total NO_3 reactivity (kNO_3).

Minor Comments:

1. L19 and 29: The meaning of NO_2 -limited in the abstract may not be clear to the reader as these regimes have not yet been introduced or defined.

Response: Thanks for your comment. We have revised the abstract to more clearly indicate that the meaning of NO_2 -limited. The modifications in the revised manuscript (line 19-21) are as follows:

“However, the relative contributions of pNO_3^- formation pathways in urban areas remain poorly quantified, particularly under the NO_2 -limited regime that governs its formation (as defined by the NO_2/O_3 ratio), which hinders effective particulate pollution control.”

2. L63: The meaning of this sentence isn't clear. Are you saying that when N_2O_5 dominates pNO_3^- , N_2O_5 production is typically NO_2 limited or aerosol surface area is large.

Response: Yes, this is exactly what we intended to express. For clearer expression, we have revised the manuscript (**line 63-65**) as follows:

“However, the N_2O_5 uptake served as the dominant pathway for pNO_3^- formation, typically under NO_2 -limited conditions (e.g., reduced emissions during the pandemic) or under large aerosol surface areas (e.g., severe particulate pollution episodes).”

3. L78 and elsewhere: I would encourage making sure the terminology distinguishing various effects is clear throughout the manuscript. I understand that the intended meaning is that VOC reduction will decrease the removal of NO_3 by VOCs, leading to higher N_2O_5 production rates and therefore more pNO_3^- production from N_2O_5 heterogeneous reactions. However, the phrasing “enhancing N_2O_5 uptake” implies to me an increase in the first order N_2O_5 heterogeneous rate ($k_{\text{N}_2\text{O}_5}$) which is independent of VOC. (also lines 279, 281)

Response: Thanks for your suggestion. We have modified the corresponding phrasing in the revised manuscript (e.g., **line 77-79** and **line 294-296**) to clearly distinguish between “ N_2O_5 uptake processes” and “ pNO_3^- production via N_2O_5 uptake”.

The modifications in the main text are as follows:

“A recent study has revealed that under O_3 -limited conditions for N_2O_5 formation (Zhang et al., 2023), reducing NO_x emissions had negligible effects, while reducing

VOCs decreased the consumption of NO₃ by VOCs, thereby enhancing pNO₃⁻ formation from N₂O₅ uptake.” (line 77-79)

“The steep slope of the positive correlation between P(NO₃) and SHAP values indicated that P(NO₃) strongly enhances pNO₃⁻ formation via N₂O₅ uptake.” (line 294-296)

4. Line 131: while the R² is good the slopes seem like they are far from 1. Please give values for these slopes and discuss implications.

Response: Thanks for your comment. The mean slopes of observed versus simulated N₂O₅ and ClNO₂ were 0.50 and 0.64, respectively, indicating that both N₂O₅ and ClNO₂ were underestimated in the simulations. This underestimation was mainly attributed to the model configuration in the multiphase chemical box model.

First, in the multiphase chemical box model, both dilution and dry deposition processes were included and constrained by the boundary layer height ($k_{dilution} = k_{dilution,base} \times \frac{BLH_i}{BLH_{max}}$, $k_{deposition} = \frac{k_{deposition,base}}{BLH_i}$). The rates of dilution and dry deposition may be overestimated. During the nighttime, when the boundary layer height is lower, the dry deposition rate becomes larger while the diffusion rate decreases, leading to lower simulated N₂O₅. In addition, transport process could contribute to ambient N₂O₅ levels. The absence of transport part in the box model may also contribute to the underestimation of simulated N₂O₅.

Second, a 3-day spin up was set before each model simulation to allow intermediate species to reach a stable concentration. Consequently, in addition to the

observed VOCs, some secondary chemical species formed from these VOCs were present in the model. These species could also react with NO_3 , thereby reducing the precursors of N_2O_5 and contributing to the underestimation of N_2O_5 in the simulation.

The underestimation of N_2O_5 also led to an underestimation of ClNO_2 . Correspondingly, the pNO_3^- production via N_2O_5 uptake would be underestimated. In the revised manuscript (line 142-147), we have addressed the underestimation caused by the model simulation and discussed its implications for the estimated pNO_3^- production via N_2O_5 uptake.

The modifications in the main text are as follows:

*“As shown in **Figure S3**, the model performed well in simulating the trends of N_2O_5 and ClNO_2 with R^2 of 0.88 and 0.49, respectively. However, a systematic underestimation existed in the simulated N_2O_5 and ClNO_2 concentrations, which likely resulted from the model configuration including overestimated physical removal rates, elevated concentration of intermediate VOC species, or uncertainties in transport processes. Consequently, the simulated pNO_3^- formation from N_2O_5 uptake in this study could be regarded as a lower limit.”*

5. Line 135: NO_3^- from N_2O_5 can also partition to the gas phase as HNO_3 . I don't think this is an important point for this analysis, but it is not clear that this effect would lead to an overestimation of the $\text{OH} + \text{NO}_2$ pathway.

Response: Thank you for the note. We have removed the relevant content from the main text.

6. Fig 3: Panel A. Doesn't the right y-axis show the percent contribution not the ratio?

Response: Thank you for the note. The right y-axis of Fig 3. Panel A is the percentage of N₂O₅ uptake to nitrate formation (%). We have adjusted the **Figure 3** in the revised manuscript.

7. Supplement L56 and L65: Were N₂O₅ and ClNO₂ calibrated through the full 2 meter stainless steel inlet used for the ambient observations? If not, was an inlet loss rate determined. N₂O₅ loss on that length of stainless steel could be substantial.

Response: We apologize for the incorrect description in the previous version of the supplementary material. A long perfluoroalkoxy (PFA) tube with a length of nearly 2 meters and a 1/4 -inch inner diameter was used for sampling, not the 2-meter stainless steel one. In order to minimize the effect of particles deposited on the surface of the sampling inlet, the tube was cleaned by deionized water and dried by nitrogen flow once a week. In the calibration process, the standard gas was also delivered to the instrument through the PFA tube, consistent with the configuration for the field measurement. To better clarify the operation and calibration of the CIMS instrument, we have moved the relevant content from the supplementary material to the main text. The revised text (**line 112-127**) is provided below.

“A nearly 2-meter long perfluoroalkoxy (PFA) tube with a 1/4-inch inner diameter was used for sampling. The total sampling flow rate was set as 10 standard liters per minute (SLPM), of which only 2SLPM was diverted to the CIMS. A nitrogen (N₂) flow

(99.999%, 2.7 SLPM), carrying methyl iodide (CH_3I) vapor released from a heated permeation tube, passed through a soft X-ray source (Tofwerk AG, P-type) to generate reagent ions I^+ . The I^+ was combined with the target gas in an ion molecule reaction (IMR) chamber and then detected by the ToF-CIMS. Ambient N_2O_5 and ClNO_2 were detected as the $\text{I}(\text{N}_2\text{O}_5)^+$ and $\text{I}(\text{ClNO}_2)^+$ clusters at 235 and 208 m/z . The detailed calibration procedures of N_2O_5 and ClNO_2 are described in **Text S2**, following established methods (Wang et al., 2022c; Wang et al., 2022b; Thaler et al., 2011). Briefly, N_2O_5 was generated from the reaction between O_3 and excessive NO_2 , while ClNO_2 was synthesized via the reaction of Cl_2 (6 ppm in N_2) with a moist mixture of NaNO_2 and NaCl . The calibration curves for N_2O_5 and ClNO_2 at different RH are shown in **Figure S2**, with mean sensitivities of 0.110 ± 0.063 and 0.055 ± 0.018 ncps/ppb, respectively. The instrument background was determined by introducing dry N_2 into the inlet for 20 min. Based on three times the standard deviation (3σ) of the background signal, the typical 1-minute detection limits for N_2O_5 and ClNO_2 were estimated to be 1.3 and 0.61 ppt, respectively.”

8. Supplement L62: IClNO_2^+ is at m/z 208

Response: Thank you for the note. We have corrected it.

9. Supplement L82: At what averaging time?

Response: Thank you for the comment. The averaging time is 1 minute and we have added it in the revised main text (**line 125-127**) as follows.

“Based on three times the standard deviation (3σ) of the background signal, the typical 1-minute detection limits for N_2O_5 and $ClNO_2$ were estimated to be 1.3 and 0.61 ppt, respectively.”

10. Figure S2: These sensitivities are notably quite low compared to typical Iodide CIMS instruments. Also, the LODs quoted in line L82 seem very good given the poor sensitivity. Can you expand further on how these values were derived.

Response: We replied to this comment in the following two aspects.

(1) In **Figure S2**, the sensitivities appear lower due to the normalization of N_2O_5 and $ClNO_2$ signals applied in the calibration curves. The normalized signals of N_2O_5 and $ClNO_2$ are calculated as $N_2O_5(ncps) = \frac{(IN_2O_5)^-}{I^-(H_2O)^-}$ and $ClNO_2(ncps) = \frac{(ClNO_2)^-}{I^-(H_2O)^-}$, respectively. The signal intensity of $[I^- + (IH_2O)^-]$ was approximately on the order of 10^5 counts. Consequently, compared with the signal $(IN_2O_5)^-$, the normalized signals were quite low. Thus, the sensitivities appeared relatively low. In our work, the mean sensitivities of N_2O_5 and $ClNO_2$ were 0.110 ± 0.063 and 0.055 ± 0.018 ncps/ppb, respectively, which are comparable to those reported in existing research (see in **Figure R2**). To avoid misunderstanding, we have revised the description of **Figure S2** (line 172-174) in the revised supplementary material) to emphasize that the signals represent normalized results, and we have presented the corresponding sensitivities in the main text (line 122-124).

The modifications are presented below.

“In panels (a) and (b), the signals of N_2O_5 and $ClNO_2$ are normalized signals,

which were calculated according equation $N_2O_5(ncps) = \frac{(IN_2O_5)^-}{\Gamma + (IH_2O)^-}$ and equation $ClNO_2(ncps) = \frac{(IClNO_2)^-}{\Gamma + (IH_2O)^-}$, respectively.” (line 172-174 in the revised supplementary material)

“The final calibration curves for N_2O_5 and $ClNO_2$ at different RH are shown as **Figure S2** with mean sensitivities of 0.110 ± 0.063 and 0.055 ± 0.018 ncps/ppb, respectively.” (line 122-124)

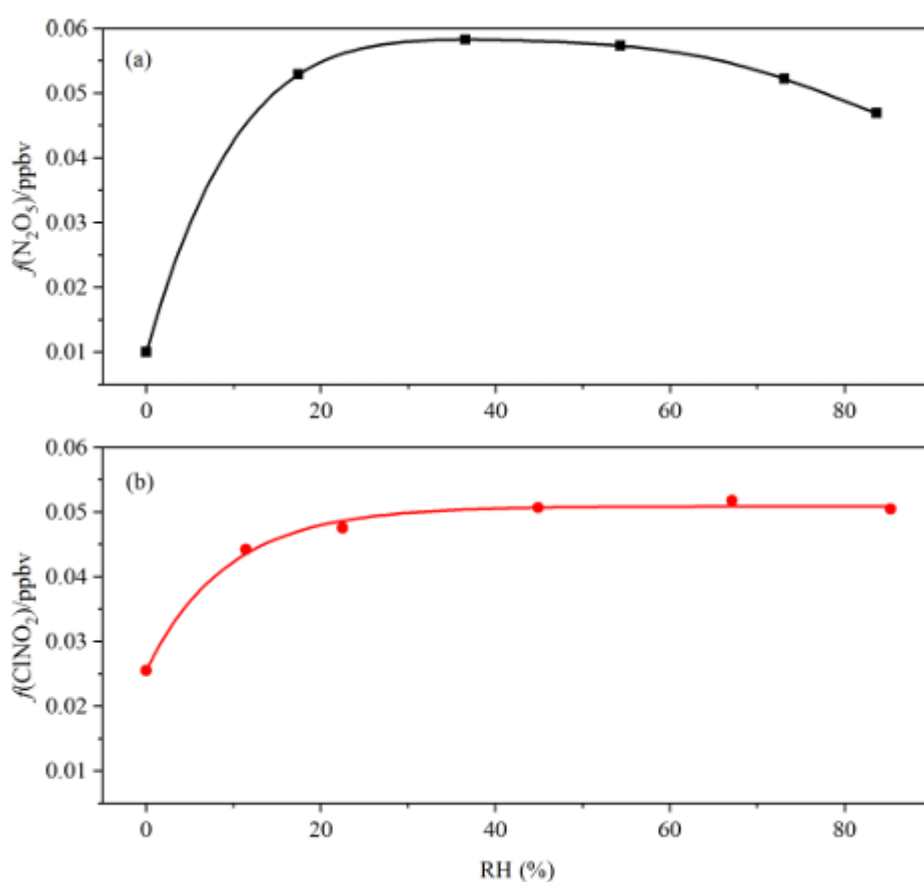


Figure R2. CIMS sensitivities as a function of RH for N_2O_5 and $ClNO_2$ reported in the existing study by Wang et al (Wang et al., 2022a).

(2) As for the LOD, it was calculated based on the standard deviation of the background signal and the sensitivity. The background signals of the CIMS instrument were determined by introducing dry N_2 into the inlet for a duration of 20 min. According

to three times the standard deviation (3σ) of the background signal, the typical detection limit of N_2O_5 and ClNO_2 for 1 min were estimated. In the revised main text, we have added the details (**line 125-127**) as follows:

“The background signals of the CIMS instrument ascertained by introducing dry N_2 into the inlet for a duration of 20 min. According to three times the standard deviation (3σ) of the background signal, the typical detection limit of N_2O_5 and ClNO_2 for 1 min were estimated to be 1.3 and 0.61 ppt, respectively.”

11. Supplement L85: What time resolution data was used for the iterative box model

Response: Thank you for the comment. The time resolution of the input data for the iterative box model is one hour. We have added this detail in the revised supplementary material (**line 82-83**).

The modifications are as below.

“Notably, the input data for the iterative box model have a time resolution of 1 hour.”

References:

- Hu, H., Wang, H., Lu, K., Wang, J., Zheng, Z., Xu, X., Zhai, T., Chen, X., Lu, X., Fu, W., Li, X., Zeng, L., Hu, M., Zhang, Y., and Fan, S.: Variation and trend of nitrate radical reactivity towards volatile organic compounds in Beijing, China, *Atmos. Chem. Phys.*, 23, 8211-8223, <https://doi.org/10.5194/acp-23-8211-2023>, 2023.
- Li, J., Zhai, T., Chen, X., Wang, H., Xie, S., Chen, S., Li, C., Gong, Y., Dong, H., and Lu, K.: Direct measurement of N_2O_5 heterogeneous uptake coefficients on atmospheric aerosols in southwestern China and evaluation of current parameterizations, *Atmos. Chem. Phys.*, 25, 6395-6406, <https://doi.org/10.5194/acp-25-6395-2025>, 2025.
- Thaler, R. D., Mielke, L. H., and Osthoff, H. D.: Quantification of Nitryl Chloride at Part Per Trillion Mixing Ratios by Thermal Dissociation Cavity Ring-Down Spectroscopy, *Analytical Chemistry*, 83, 2761-2766, <https://doi.org/10.1021/ac200055z>, 2011.

Wang, H., Peng, C., Wang, X., Lou, S., Lu, K., Gan, G., Jia, X., Chen, X., Chen, J., Wang, H., Fan, S., Wang, X., and Tang, M.: N₂O₅ uptake onto saline mineral dust: a potential missing source of tropospheric ClNO₂ in inland China, *Atmos. Chem. Phys.*, 22, 1845-1859, <https://doi.org/10.5194/acp-22-1845-2022>, 2022a.

Wang, H. C., Peng, C., Wang, X., Lou, S. R., Lu, K. D., Gan, G. C., Jia, X. H., Chen, X. R., Chen, J., Wang, H. L., Fan, S. J., Wang, X. M., and Tang, M. J.: N₂O₅ uptake onto saline mineral dust: a potential missing source of tropospheric ClNO₂ in inland China, *Atmospheric Chemistry and Physics*. 22, 1845-1859, <https://doi.org/10.5194/acp-22-1845-2022>, 2022b.

Wang, H. C., Yuan, B., Zheng, E., Zhang, X. X., Wang, J., Lu, K. D., Ye, C. S., Yang, L., Huang, S., Hu, W. W., Yang, S. X., Peng, Y. W., Qi, J. P., Wang, S. H., He, X. J., Chen, Y. B., Li, T. G., Wang, W. J., Huangfu, Y. B., Li, X. B., Cai, M. F., Wang, X. M., and Shao, M.: Formation and impacts of nitryl chloride in Pearl River Delta, *Atmospheric Chemistry and Physics*. 22, 14837-14858, <https://doi.org/10.5194/acp-22-14837-2022>, 2022c.

Zhang, Y. N., Wang, H. L., Huang, L. B., Qiao, L. P., Zhou, M., Mu, J. S., Wu, C., Zhu, Y. J., Shen, H. Q., Huang, C., Wang, G. H., Wang, T., Wang, W. X., and Xue, L. K.: Double-Edged Role of VOCs Reduction in Nitrate Formation: Insights from Observations during the China International Import Expo 2018, *Environmental Science & Technology*. 57, 15979-15989, <https://doi.org/10.1021/acs.est.3c04629>, 2023.