



High-resolution Emission Inventory Development and Co-emission Hotspot Identification of Air Pollutants and Greenhouse Gases in Central Plains Region, China

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Abstract. A high-resolution inventory provides scientific basis for numerical simulations and control
10 strategies. Under the background of synergistic carbon reduction and pollution control, constructing a
carbon-pollutant co-emission inventory is of great significance for regional air quality improvement.
Taken Henan Province in the Central Plains region as an example, the most polluted regions in China, an
update emission inventory was developed. The study presents results showing that in 2022, the total
emissions of SO₂, NO_x, CO, PM₁₀, PM_{2.5}, VOCs, NH₃, BC, OC, CO₂, CH₄, and N₂O in Central China,
15 particularly Henan Province, were 408.7, 1336.2, 4647.3, 901.1, 440.0, 759.3, 672.7, 47.4, 90.3,
540462.0, 12462.0 and 42.9 kt respectively. The emissions were predominantly attributed to industrial
combustion, electricity generation, motor vehicles, and agricultural activities. Significant spatial
heterogeneity was observed. Northern heavy industrial cities exhibited high carbon and pollution
intensities with carbon emission 1.75-3.7 times higher than the provincial average. In contrast, central
20 transportation hubs were primarily characterized by elevated emissions of NO_x and VOCs. Southern
agricultural areas showed low carbon but high NH₃ emissions. Temporally, emissions of SO₂ and PM_{2.5}
peaked during winter, whereas NH₃ increased during the summer agricultural season. High-emission
grids were predominantly concentrated in urban agglomerations of the north-central region, especially
around Zhengzhou, Jiaozuo, and Anyang. Hotspot analysis revealed that 5% of high-emission grids
25 accounted for more than 50% of total emissions, indicating a highly uneven spatial distribution. These
results highlight that understanding the region-specific emission characteristics of different regions is
critical for developing mitigation strategies in future.



1 Introduction

Atmospheric pollutants are primarily emitted due to fundamental human activities, including energy consumption and economic operations (Kanakidou, 2024; Huang et al., 2024). Notably, pollutants such as SO₂, NO_x, VOCs, PM_{2.5} and NH₃, play a key role in regional atmospheric pollution formation (Kanakidou, 2024), which can significantly impair visibility, ecology and public health (Manisalidis et al., 2020). Meanwhile, the global warming effects of greenhouse gases (GHGs), as the main emission drivers of climate change, have been widely noticed and recognized (IPCC, 2021). Although air pollutants and GHGs differ in their environmental impacts with air pollutants primarily posing a local, short-term health and ecological risk, while GHGs contribute to global-scale, long-term warming effect, they are fundamentally both gaseous originating from highly overlapping emission sources, including fossil energy combustion, industrial process, transportation, and residential activities (Xu et al., 2025; Liu et al., 2023; Liu et al., 2021; MEEPRC, 2021).

Globally, the air pollution situation in China remains concerning and the PM_{2.5} and ozone pollution have persisted at high concentrations for a long time (IQAIR, 2024; Jin et al., 2021). According to the Bulletin of China's Ecological and Environmental Conditions 2023, 136 out of 339 cities at prefecture level and above in the country have exceeded national Grade II air quality standards (MEEPRC, 2024a; Wang et al., 2025). Meanwhile, China is a major contributor to global GHG emissions. According to the Carbon Emissions Report 2023 released by International Energy Agency, the total carbon emissions in China reach 12.6 billion tons in 2023 with a growth rate of 4.7% (IEA, 2024). Therefore, more and more scholars at home and abroad are highly concerned about the quantification of their emissions including Chinese scholars as well.

Domestic and international researches have made significant progress, and a series of national and global scale emission inventories have been constructed, including MEIC (Li et al., 2018), EDGAR (Muntean et al., 2018), REAS (Kurokawa et al., 2013), and CHRED (Cai et al., 2018). Meanwhile, a series of emission inventories in China have been constructed at the regional (Xu et al., 2020; Zhao et al., 2024; Wu et al., 2022), provincial (Zhou et al., 2020; Jiang et al., 2020a), urban (Liu et al., 2018), and industry (Gao et al., 2019; Hua et al., 2016; Jiang et al., 2020b) levels have also made significant progress. In particular, the development of air pollutant emission inventories has been rapid



advancements in spatial resolution and temporal distribution (Chen et al., 2024), and some studies have successfully constructed gridded data with a precision of $0.1^\circ \times 0.1^\circ$ or even higher (Zheng et al., 2021). However, GHGs emission inventories typically emphasize aggregate total emission of the five major sectors, including energy, industrial process, agriculture, forestry, and waste at national or provincial levels, rather than high-resolution emission inventories (Geng et al., 2011; Yu and Tan, 2023; Xue et al., 2021; He et al., 2024). At present, China has attached great importance to the work of reducing pollution and carbon emissions in a coordinated manner as a national strategy (CPGPRC, 2022; Yu et al., 2020). A series of policy documents have been issued successively which cover the national or provincial strategy (CPGRC, 2024a) and key industries (CPGRC, 2024b; MTPRC, 2024). As the emission reduction potential of traditional end-of-pipe measures gradually narrowing (Kong et al., 2024; Cao et al., 2025), it is necessary to develop integrated emission inventory of carbon and pollutants at city level or even enterprise level (Wu et al., 2024). It not only aids to explore the emission reduction potential from the front-end and also enhances the identification of GHG spatial grid synergies.

In the dual context of high-value for air pollutant and GHG emissions pressure in China, Central China, especially represented by Henan Province is typical and representative of high pollutant concentrations and high carbon emission. Henan Province is a major contributor to $\text{PM}_{2.5}$ and O_3 pollution double high value area: in 2023, $\text{PM}_{2.5}$ concentration exceeds the national standard by 28%, and O_3 exceeds the national standard by 12%; the proportion of good air quality day was lower than the national average of 85.5% (MEEPRC, 2024a). Also, Henan Province has a share of coal in energy consumption close to 55.3% in 2023, and the energy-related CO_2 emission in this region was accounted for approximately 5% of the total emission in China (HBS, 2024). Therefore, we selected Henan Province as the study area to develop a integrated emission inventory of carbon and pollutants based on a unified source classification system with enterprise-level point source support as possible. It hopes to features enterprise-level point source support, $3 \text{ km} \times 3 \text{ km}$ spatial grid resolution, and detailed temporal allocation.

The objectives of this study are to (1) calculate the emissions of air pollutants and GHGs based on uniform sources, including nine conventional air pollutants (SO_2 , NO_x , CO , PM_{10} , $\text{PM}_{2.5}$, VOCs, NH_3 , BC, and OC) and three GHGs (CO_2 , CH_4 , and N_2O); (2) analyze the emission characteristics of source



and city contributions; (3) construct a $3 \text{ km} \times 3 \text{ km}$ spatially gridded emissions and monthly assignments;
85 (4) compare with other studies and conduct uncertainty analysis; and (5) identify high-emission grids
using hotspot analysis methods.

2 Data and methodology

2.1 Study domain

This study centered on Henan Province as its primary research site, with 2022 serving as the
90 reference period. Situated at the crossroads of central and eastern China, the region stretches along the
middle and lower reaches of the legendary Yellow River, occupying a strategically important position in
the country's geographical landscape. It governs 18 administrative divisions, including Zhengzhou,
Kaifeng, Luoyang, Pingdingshan, Anyang, Hebi, Xinxiang, Jiaozuo, Puyang, Xuchang, Luohe,
Sanmenxia, Nanyang, Shangqiu, Xinyang, Zhoukou, Zhumadian, and the Jiyuan Demonstration Area
95 (see Figure 1).

Henan Province, a central Chinese region, exemplifies high population density, industrial and
agricultural prominence, and key transportation networks. With a landmass of roughly 167000 square
kilometers, Henan Province makes up about 1.73% of China's total territory (HPPG, 2023). In 2022, its
GDP reached 6.13 trillion yuan, ranking fifth nationally (HBS, 2023; SCPRC, 2023). With a permanent
100 population of 98.72 million in 2022, it ranks third in China (NBS, 2023). The economic landscape is
dominated by a split of 8.6% in the primary sector, 38.3% in the secondary, and 53.1% in the tertiary
sector (HBS, 2023). The region's energy structure has long been dominated by coal, with standard coal
consumption reaching 214 million tons in 2022, accounting for 62.7% of the province's total energy
consumption (HBS, 2023). The five major high-energy-consuming industries of petrochemicals,
105 chemicals, building materials, iron and steel, and non-ferrous metals account for 53.84% of the total
output value of large-scale industrial products in Henan Province (HBS, 2023). Henan had 42.6 million
live pigs in stock in 2022, ranking second nationally, while Zhengzhou, the provincial capital, had 4.57
million registered motor vehicles, ranking sixth among Chinese cities (HBS, 2023; HPDEE, 2022).

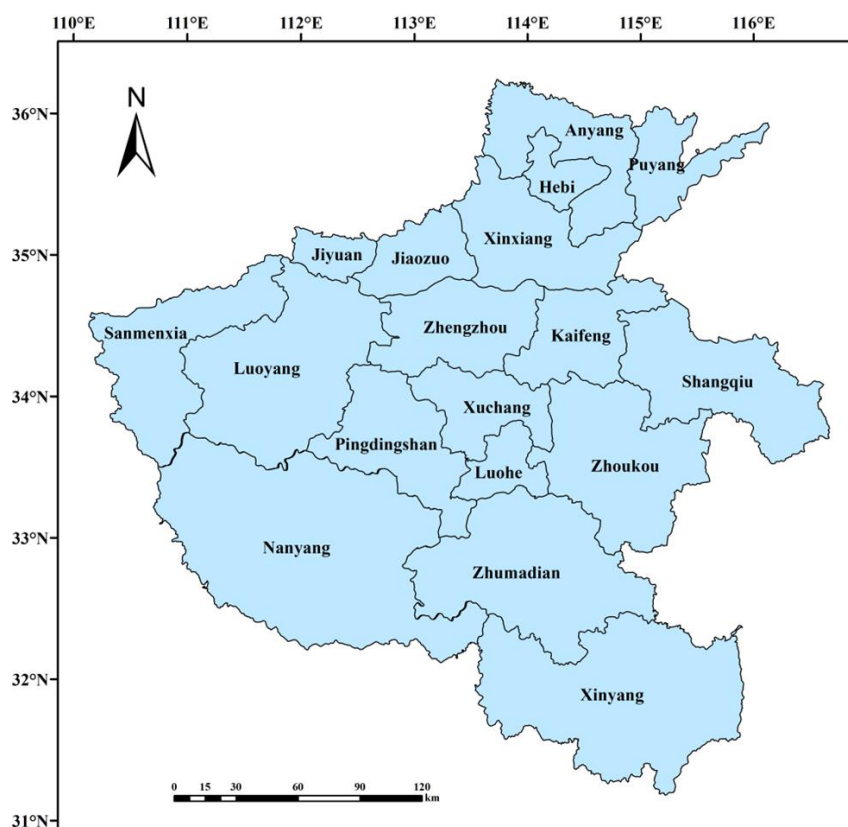


Figure 1. Study domain.

2.2 Source categorization

To develop a trustworthy emissions inventory, as a fundamental requirement, it's vital to have a sound and thorough system for classifying the emission sources (Zhang et al.,2023). Distinct emission inventories traditionally cataloged GHGs and air pollutants separately. To facilitate the integration of the compilation processes for both types of emission inventory, this study focuses on the unified source classification framework. By refer to the “Technical Guidelines for the Compilation of Comprehensive Emission Inventories of Air Pollutants and GHGs (for Trial Implementation)” (hereinafter referred to as the “Technical Guidelines”) (MEEPRC,2024b), the unified classification framework of seven emission source of electricity and thermal power, industry, mobile source and oil storage and transportation, fugitive dust, resident, agriculture, and waste treatment was adopted. A detailed list of these specific



source types and associated activity data can be found in Table S1.

2.3 Methodology and data collection

The research utilized two main techniques for gauging emissions: the emission factor and the mass balance methodologies. Data on activities were gathered from a blend of on-site investigations, documented information from ecological and environmental authorities, and significant statistical compilations, including the Henan Provincial Statistical Yearbook, the China Statistical Yearbook, and the China Agricultural Yearbook. The compiled datasets included pivotal elements such as energy usage, industrial output, pollution abatement facilities, regional dispersion, and vehicle numbers, with comprehensive data outlined in Table S1.

Emission factors were chosen using pertinent literature and local assessments, emphasizing applicability to Henan Province and adjacent regions. The corresponding factor data can be found in Tables S2 through S7.

2.3.1 Air pollutant emission calculation

The emissions of conventional air pollutants (SO_2 , NO_x , CO , VOCs , PM_{10} , $\text{PM}_{2.5}$, NH_3) are calculated using the emission factor method with the following formula:

$$E = A \times EF \quad (1)$$

Here, E denotes the total emissions from each source within Henan Province, A represents the activity level data, EF stands for the emission factor value.

For electricity and thermal power source as well as industrial combustion, a more detailed mass balance method can be used for SO_2 emission, with the following formula:

$$E_{\text{SO}_2} = \sum_{p,q} 2 \times S \times F_{p,q} \times C_q \times (1 - \eta_{\text{SO}_2}) \quad (2)$$

where, E is the emission of SO_2 , 2 is the molecular weight ratio of sulfur dioxide (SO_2 molecular weight 64) to sulfur (S molecular weight 32), S is the sulfur content in the fuel, F is the fuel consumption, C is the sulfur conversion rate, and η_{SO_2} is the sulfur removal rate.

BC and OC emissions are derived from $\text{PM}_{2.5}$ levels and their respective fractions, calculated as follows:

$$E_i = E_{\text{PM}_{2.5}} \times f_i \quad (3)$$



Where E_i is the emission of BC or OC; i represents BC or OC; $E_{PM_{2.5}}$ is the $PM_{2.5}$ emission; and f_i denotes the BC or OC fraction within $PM_{2.5}$.

2.3.2 GHG emission calculation

(1) CO₂ emission calculation method

CO₂ emissions from electricity and thermal power source are calculated using the following formula:

$$E_{CO_2} = \sum_i AC_i \times NCV_i \times CC_i \times COF_i \times \frac{44}{12} \quad (4)$$

where, E_{CO_2} is the CO₂ emission; i is the fuel type; AC is the fuel consumption; NCV is the net calorific value of the fuel; CC is the carbon content per unit calorific value of the fuel; COF is the carbon oxidation factor of the fuel; and $44/12$ is the molecular weight ratio of CO₂ to C.

The CO₂ emission calculation methods for industrial source, mobile source and oil storage and transportation, residential source, and agricultural source are similar to those for electricity and thermal power source, all based on fuel consumption and emission factors.

(2) CH₄ emission calculation method

The CH₄ emission calculation formula for agricultural livestock and poultry manure management systems is:

$$E_{M-CH_4} = EF_{M-CH_4} \times AP \quad (5)$$

where, E_{M-CH_4} is the CH₄ emission factor per unit livestock, and AP is the annual livestock population.

For waste treatment source, the CH₄ emission formula for domestic wastewater treatment is:

$$E_{CH_4} = (TOW \times EF) - R \quad (6)$$

where, TOW is the total organic loading in domestic wastewater, EF is the emission factor, and R is the CH₄ recovery rate.

The CH₄ emission formula for municipal solid waste landfill is:

$$E_{CH_4} = (MSW_T \times MSW_F \times L_0 - R) \times (1 - OX) \quad (7)$$

where, MSW_T is the total municipal solid waste generation, MSW_F is the fraction sent to landfills, L_0 is the methane generation potential, R is the recovered methane, and OX is the oxidation factor.

CH₄ emissions from mobile source and oil storage and transportation and residential source are calculated using Eq. (1)



(3) N₂O Emission Calculation Method

N₂O emissions from agricultural livestock and poultry farms are divided into direct and indirect emissions. The direct emission formula is:

$$E_{M-N_2O-dir} = \sum_i AP \times Nex \times \frac{MS_i}{100} \times EF_i \times \frac{44}{28} \quad (8)$$

180 where, Nex is the annual nitrogen excretion per animal unit, MS_i is the fraction of manure managed using system i , EF_i is the direct N₂O emission factor for management system i , and AP represents livestock population (using annual throughput for pigs, broilers, and beef cattle, and average inventory for dairy cattle and laying hens).

The indirect emission formula is:

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$$E_{M-N_2O-ind} = \sum_i Nex \times MS_i \times \frac{Frac_{gas}}{100} \times EF_i \times \frac{44}{28} \times AP \times 10^{-3} \quad (9)$$

where $Frac_{gas}$ is the fraction of nitrogen volatilized as NH₃ and NO_x, and EF_i is the emission factor for N₂O formation from volatilized nitrogen.

The N₂O emission calculation formula for domestic wastewater treatment in waste treatment source is:

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$$E_{N_2O} = N_E \times EF_E \times \frac{44}{28} \quad (10)$$

where N_E is the nitrogen content in wastewater, EF_E is the N₂O emission factor for wastewater treatment, and $44/28$ is the molecular weight conversion factor.

N₂O emissions from electricity and thermal power and mobile source and oil storage and transportation are calculated using Eq. (1)

195 2.4 Temporal and spatial allocation

This research aimed to bolster the precision of air quality modeling at a high resolution, and to gain a deeper understanding of how air pollutants and greenhouse gases are released together. To achieve this, we crafted a detailed gridded emission inventory that captured both time-based and spatial-based nuances. Based on annual emissions, the results were further allocated by month and mapped to a 3 km × 3 km spatial grid.

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For temporal distribution, monthly variation was determined based on activity profiles of different emission source. Key monthly activity parameters were collected through sector-specific surveys and



field investigations. These parameters included electricity generation, traffic volume, satellite fire counts, residential activity patterns, construction schedules, and precipitation. Using these data, weighted
 205 monthly coefficients were developed to reflect the temporal variation of different source. These coefficients were then applied to allocate annual emissions across individual month, resulting in a time-resolved emission profile.

Although high-resolution spatial allocation methods for air pollutants are relatively well established, grid-based mapping of GHG emissions remains limited. In particular, refined spatial distribution of
 210 GHGs has lacked systematic methodological approaches. In this study, spatial allocation was conducted using ArcGIS software, with emissions mapped to a $3 \text{ km} \times 3 \text{ km}$ grid. The allocation process emphasized enhancing the representation of point-source. For major sectors such as power, steel, cement, and chemicals, geographic coordinates of 7390 enterprises were collected. These covered 23 key industries, including cement, glass, and iron and steel. Line source were allocated based on road networks and traffic
 215 volume, while area source were distributed using land use and population density data. The final gridded inventory demonstrates high spatial resolution and a high coverage rate of point source. The uncertainty of the grid spatial distribution is mainly due to the deviation of proxy data such as population density and cultivated land area in spatial distribution, as well as the inaccuracy of the geographical coordinates of pollution sources. Detailed temporal and spatial allocation parameters are provided in Tables S8–S9.

220 2.5 Identification method of hotspot regions for co-control

To pinpoint areas where CO_2 and key air pollutants overlap, this research leveraged a hotspot analysis. We employed ArcGIS to conduct and visualize this analysis. Initially, using our pre-existing $3 \text{ km} \times 3 \text{ km}$ gridded emission inventory, we mapped out the spatial distribution of CO_2 alongside four pollutants of target—namely SO_2 , NO_x , $\text{PM}_{2.5}$, and VOCs—across the entire grid network of Henan
 225 Province. Each pollutant, including CO_2 , was ranked independently according to grid-level emission intensity. Grids within the top 20% were classified into four hotspot categories: the top 5% as Category 1, 6%–10% as Category 2, 11%–15% as Category 3, and 16%–20% as Category 4. A bi-directional matching rule was then applied to identify co-hotspots. For example, if a grid ranked within the top 20% for both CO_2 and SO_2 emissions, it was identified as a CO_2 – SO_2 co-hotspot. The same method was used
 230 to identify co-hotspot zones for CO_2 with NO_x , $\text{PM}_{2.5}$, and VOCs. These results provide spatial insights



into areas with high levels of both carbon and pollutant emissions.

3 Results and discussions

3.1 Emissions and source contribution

As shown in Table 1 and Figure 2, the data illustrated the composition of atmospheric pollutants and GHGs emitted from the seven primary source categories in Henan Province. In 2022, the cumulative emissions were staggering, with SO₂ contributing 408.7 kt, NO_x at 1336.2 kt, CO at 4647.3 kt, PM₁₀ at 901.1 kt, PM_{2.5} at 440.0 kt, VOCs at 759.3 kt, NH₃ at 672.7 kt, BC at 47.4 kt, OC at 90.3 kt, CO₂ at 540462.0 kt, CH₄ at 12462.0 kt, and N₂O at 42.9 kt.

SO₂ emissions are mainly from industrial source, accounting for 83.5% of the total emission. Among the industrial source, industrial combustion were the most dominant sub-category, accounting for 47.2%, followed by cement manufacturing (12.1%) and chemical raw material manufacturing (11.6%). Electricity and thermal power constituted the second most important source category, accounting for 5.0% of the total. The internal emission from this sector was almost exclusively attributable to coal combustion, which contributed 99.5% of the emissions. This is closely related to Henan's coal-dominated energy structure. In 2022, coal accounted for 62.7% of the province's total energy consumption, significantly higher than the national average (HBS, 2023). The extensive use of coal and high-sulfur fuels has led to the industrial and power sectors being the major emission source of SO₂, a conclusion that aligns with the results of previous studies (Tong et al., 2025; Ren et al., 2021).

NO_x emissions are primarily attributed to mobile source and oil storage and transportation, industrial source, and electricity and thermal power, account for 39.8%, 36.2%, and 21.9% of the total emission, respectively. Within the mobile source and oil storage and transportation, On-road mobile dominates significantly, contributing to 31.1% of the total emissions. As a national transportation hub, Henan has experienced increasing road traffic pressure. By 2022, the total length of urban roads reached 19532 km, 1.81 times that of 2012 (HBS, 2013; HBS, 2023). During the same period, the number of heavy-duty trucks increased by approximately 1.52 times (HBS, 2013; HBS, 2023). The intensive operation of diesel trucks, especially in logistics centers such as Zhengzhou and Luoyang, has significantly increased NO_x emissions. Within industrial sources, industrial combustion (64.2%)



dominated the cement industry. Same to SO_2 , within electric thermal sources, coal combustion contributed up to 99.2% of NO_x emission (Sun et al., 2023).

260 CO emissions are dominated by industrial sources (57.1%), followed by mobile source and oil storage and transportation (16.9%) and residential sources (16.7%). Among the subcategories of industrial source, industrial combustion accounts for 19.1%, ferrous metal production for 17.3%, and cement manufacturing for 9.5%. Within the mobile source and oil storage and transportation category, on-road mobile is the primary contributor, accounting for 93.1% of the total emission from this sector
 265 (He et al., 2024). The increasing number of heavy-duty diesel trucks has intensified transportation activity, thereby significantly contributing to CO emissions (Keramydas et al., 2019). Residential CO emissions mainly originate from firewood use (7.6%) and household coal combustion (6.5%), indicating that traditional energy use in rural and peri-urban areas remains a key emission source. The transition to clean energy alternatives in these regions remains insufficient and requires further promotion.

270 PM_{10} and $\text{PM}_{2.5}$ were mainly from fugitive dust, industrial, and residential. Fugitive dust source contributed 49.0% of PM_{10} , while industrial source accounted for 21.2%. Within sub-sources of emissions from dust source, road dust (23.0%) and construction dust (25.9%) were predominant, showing that significant contributions of transportation and urban construction to particulate emissions. From 2012 to 2022, the length of urban roads in this region increased by 8734 km (an 81.0% rise), and the total
 275 area under building construction grew by 64.4% (HBS, 2013; HBS, 2023). Urban expansion and infrastructure development have significantly increased dust emissions from roads and construction sites (Zhou et al., 2023). On the other hand, $\text{PM}_{2.5}$ was mainly derived from industrial source, dust source and residential, accounting for 43.7%, 24.4% and 18.0%, respectively. In the residential sector, biomass combustion contributes to 5.6% of $\text{PM}_{2.5}$ emissions, primarily resulting from the use of firewood and
 280 crop residues in rural areas. This indicates that the adoption of clean energy alternatives remains insufficient, positioning biomass combustion as a significant contributor.

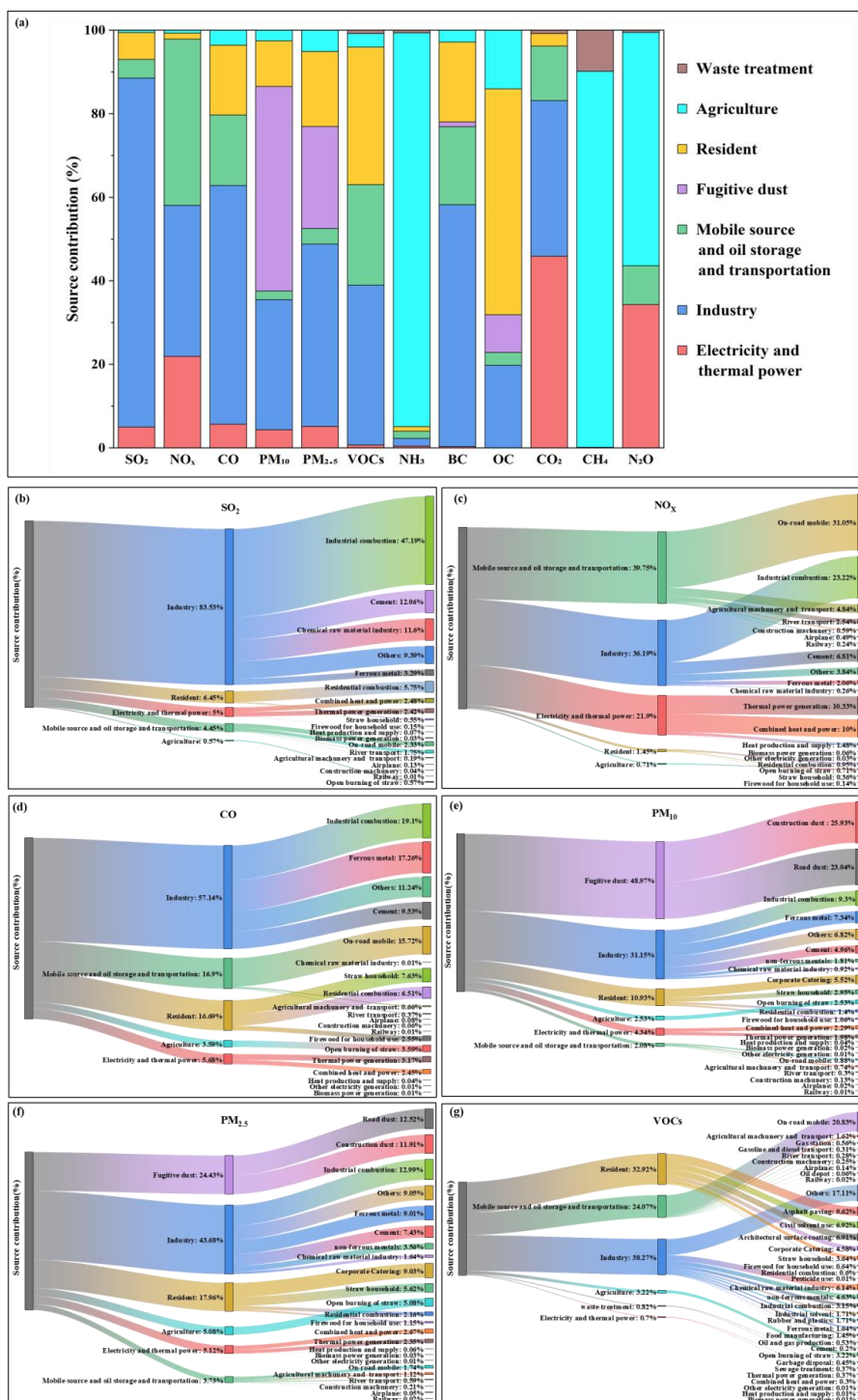
The primary sources of VOCs emissions are predominantly industrial, residential, and mobile source and oil storage and transportation source, accounting for 38.3%, 32.3%, and 24.1% of the total emissions, respectively. Within sub-sources of emissions from industrial source, chemical manufacturing
 285 (Zhou et al., 2021), non-ferrous metals, and industrial combustion together account for 14.0% of VOCs emissions. This reflects the dominance of traditional heavy industries in Henan. The relatively high

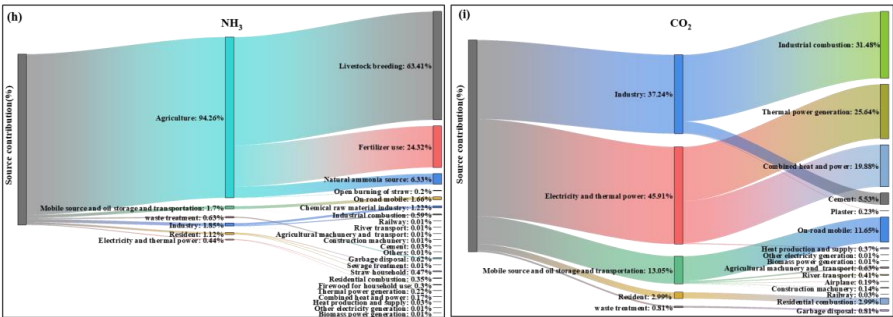


Table 1. Emissions of air pollutants and GHGs in Henan Province by source category, 2022 (Unit: kt).

Emission source	SO ₂	NO _x	CO	PM ₁₀	PM _{2.5}	VOCs	NH ₃	BC	OC	CO ₂	CH ₄	N ₂ O
Electricity and thermal power	20.5	292.6	264.0	39.1	22.5	5.3	2.9	0.1	0.0	248146.7	-	14.7
Industry	341.4	483.6	2655.4	280.7	192.2	290.7	12.4	27.5	17.8	201241.1	-	--
Mobile source												
and oil storage	18.2	531.2	785.2	18.8	16.4	182.8	11.5	8.9	2.8	70551.0	4.0	4.0
and transportation												
Fugitive dust	-	-	-	441.2	107.5	-	-	0.6	8.1	-	-	-
Resident	26.3	19.3	775.8	98.5	79.0	250.0	7.5	9.1	48.9	16133.0	13.1	
Agriculture	2.3	9.5	166.9	22.8	22.4	24.5	634.1	1.3	12.7	-	11219.5	24.0
Waste treatment	-	-	-	-	-	6.3	4.3	-	-	4390.3	1225.4	0.2
Total	408.7	1336.2	4647.3	901.-1	440.0	759.3	672.7	47.4	90.3	540462.0	12462.0	42.9

Annotation: The original data for the OC emissions from electricity and thermal power is 0.03 kt, which has been rounded to 0.0 kt.





290 **Figure 2. Source contributions of air pollutants and GHG emissions. (a) Overall source distribution across pollutants; (b-i) Sub-category source distribution across pollutants: (b) SO₂, (c) NO_x (d) CO, (e) PM₁₀, (f) PM_{2.5}, (g) VOCs, (h) NH₃, (i) CO₂.**

proportion of residential source emissions is attributed to the source classification methodology, which incorporates civil solvent use, architectural surface coating, asphalt paving, and pesticide use within the residential category. On the residential side, asphalt paving (9.6%) and building surface coating (6.9%) are the primary subcategories. These emissions are closely linked to large-scale urban construction, where coatings, adhesives, and asphalt materials are widely used (Liang et al., 2021). Mobile source and oil storage and transportation emerge as the predominant contributor. This substantial contribution can be attributed to Henan Province's extensive vehicle fleet, with Zhengzhou alone having 4.57 million vehicles in stock in 2022, ranking sixth nationally (HBS, 2023; HPDEE, 2022). As a province with 98.72 million permanent residents and serving as a crucial transportation hub connecting northern and southern China (HBS, 2023), Henan experiences substantial traffic volumes from both local commuting and long-distance freight transport. The province's well-developed highway network and rapid economic growth have driven increased vehicle ownership and usage, consequently elevating VOCs emissions from mobile source and oil storage and transportation.

NH₃, CH₄, and N₂O emissions are mainly from agricultural source, accounting for 94.3%, 90.0%, and 55.8% of total emissions, respectively. Livestock farming is the main contributor to all three. Henan is a major livestock province, with 42.6 million live pigs in stock in 2022, ranking second nationally (HBS, 2023). Large amounts of manure release NH₃, CH₄, and N₂O through volatilization, anaerobic decomposition, and denitrification, forming the key agricultural emission pathways. The province also has abundant agricultural land, with a stable arable area of 14.7 million hectares. As a major grain and



agricultural production base, Henan applies fertilizers intensively, particularly nitrogen fertilizers, which significantly contribute to increased NH_3 and N_2O emissions. In terms of N_2O , 34.3% of emission is attributed to electricity and thermal power source, with nearly all (99.9%) originating from coal
315 combustion. This highlights once again the province's dependence on coal within its energy structure.

CO_2 emissions are primarily from electricity and thermal power (45.9%) and industrial source (37.2%), together contributing 83.1% of total emissions. This highlights Henan's heavy dependence on coal-based energy and energy-intensive industries. Almost all emissions from the power sector (99.5%) originate from coal combustion, supported by six major coalfields and annual coal consumption
320 exceeding 200 million tones. This coal-dominated energy structure is the primary driver of CO_2 emissions in the power sector (Wang et al., 2024). Industrial CO_2 emissions are mainly predominantly generated by high-emission sectors such as steel, chemicals, and building materials. Specifically, coal-fired industrial boilers account for 84.5% of industrial emissions, while cement production contributes 14.9%. mobile source and oil storage and transportation contribute 13.1% of total CO_2 emissions, with On-road
325 mobile comprising 89.2% of this proportion. As a key transportation hub, Henan experiences intensive freight activity, particularly from heavy-duty diesel trucks, which constitute the main source of CO_2 emission from mobile source and oil storage and transportation (Zhang et al., 2025).

3.2 Emission characteristics by city

Influenced by its industrial structure and functional positioning, the emission patterns of pollutant
330 and GHGs emission across the study region exhibit clear geographical characteristics and synergistic variation trends (Figure 3 and Figure 4). The northern region, encompassing Anyang, Hebi, and Jiaozuo, is dominated by energy-intensive heavy industries such as steel production, non-ferrous metals processing, and building materials manufacturing. The extensive use of coal in these activities leads to high emission of SO_2 , NO_x , and $\text{PM}_{2.5}$, as well as significantly elevated levels of CO_2 emission intensity.
335 Anyang exemplifies this pattern with 2022 emissions of 77259 tons SO_2 , 153763 tons NO_x , and 59738 tons $\text{PM}_{2.5}$. Its CO_2 emissions totaled 513.6 Mt, yielding a carbon intensity of 20444 t/ 10^8 yuan GDP-1.75 times the provincial average of 11679 t/ 10^8 yuan GDP. Source apportionment analysis indicates that industrial source contributes 95.5% of SO_2 and 74.6% of NO_x emissions. Combined industrial source and electricity and thermal power account for over 91% of regional CO_2 emissions. Similar emission



340 profiles observed in Jiaozuo and Hebi indicate that this region as a concentrated hotspot for both air
pollutants and carbon emissions.

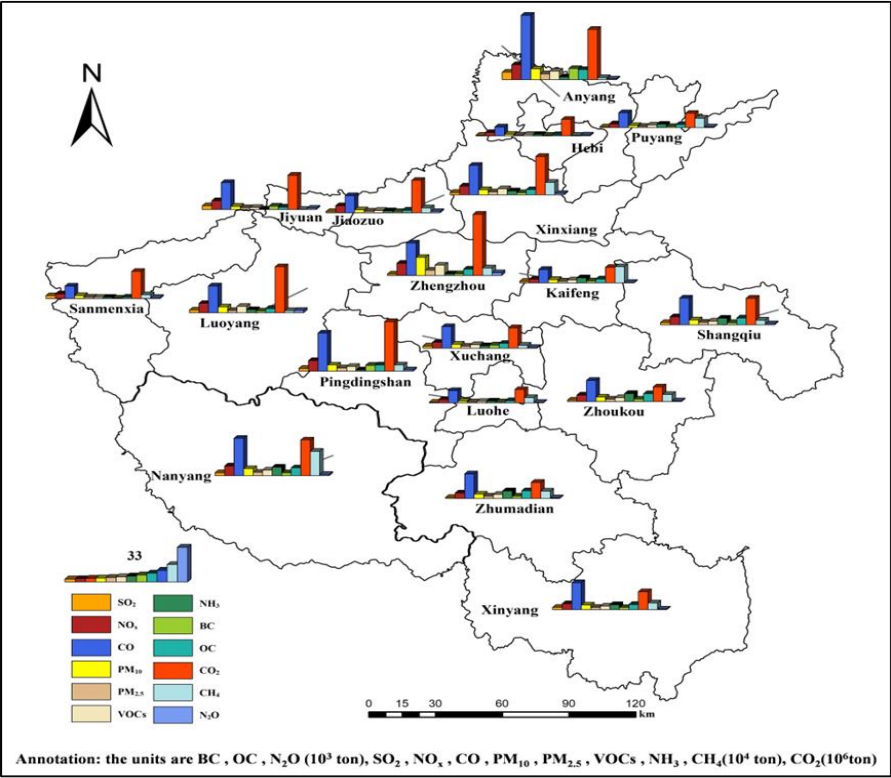
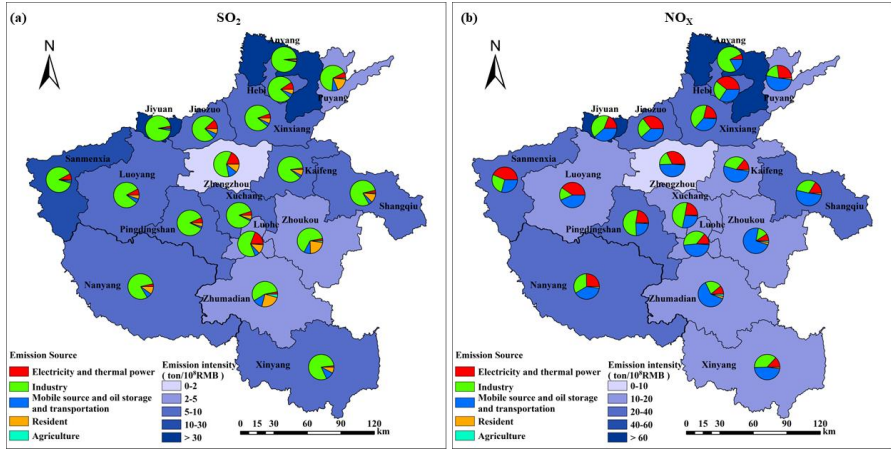


Figure 3. Map of air pollutants and GHG emissions in each city.



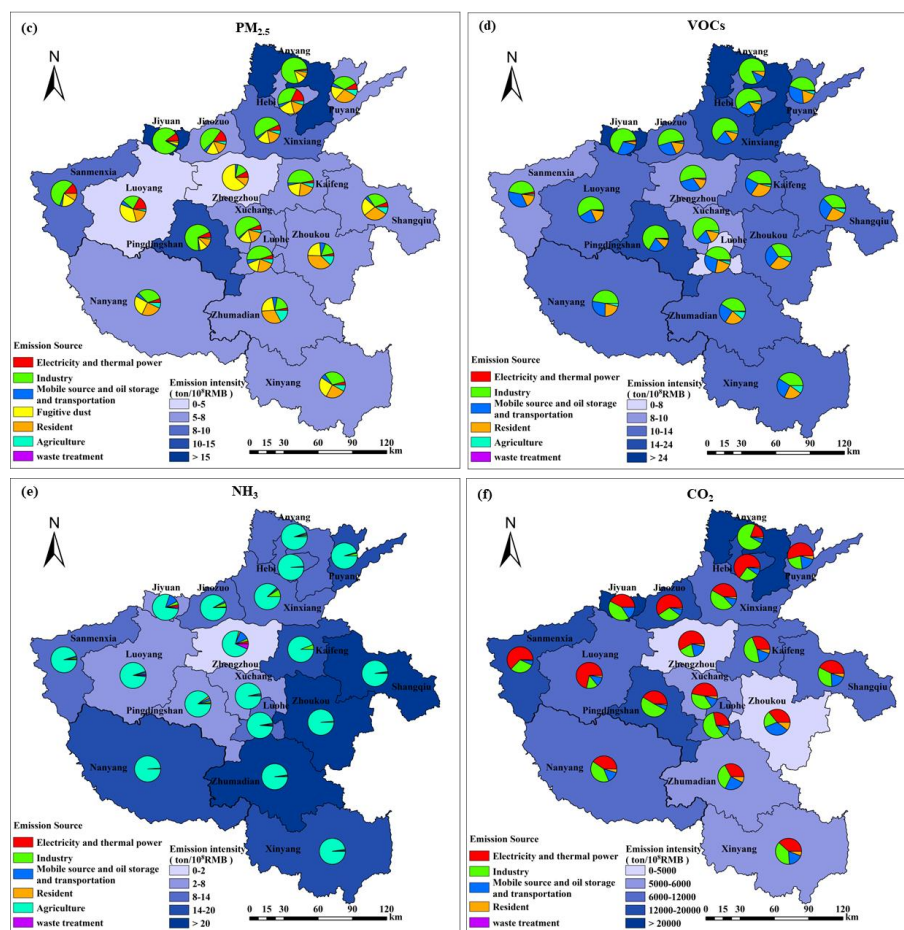


Figure 4. Maps of emission intensity per unit of GDP (ton/10⁸ RMB) for (a) SO₂, (b) NO_x, (c) PM_{2.5}, (d) VOCs, (e) NH₃, and (f) CO₂ across cities. Pie charts in each map indicate the source contributions of each city.

As for the central region, with Zhengzhou as its core, serves as the province's primary hub for political, economic, and transportation. This area exhibits high NO_x and VOCs emissions along with the highest CO₂ emission levels in the province, indicating a clear pattern of pollution dominated by motor vehicle emission. Zhengzhou recorded 125643 tons of NO_x emissions in 2022, ranking second in the province, with mobile source and oil storage and transportation contributing 61.0%. The city's VOCs emissions reached 106270 tons, ranking first provincially, with mobile source and oil storage and transportation accounting for over 30% of the total. This differs significantly from the provincial emission structure dominated by industrial source, reflecting the typical emission characteristics of



megacities with dense traffic networks. Despite generating the province's largest CO₂ emissions (625.7 Mt), the carbon intensity remains relatively low at 4837 t/10⁸yuan GDP, attributed to substantial economic output representing 21.1% of provincial GDP (HBS, 2023). The city's vehicle fleet consists of 4.6 million units, accounting 23% of the provincial total, and ranks first in the province in terms of passenger vehicles, heavy-duty trucks, and new registrations (ZBS, 2023). This extensive transportation infrastructure and logistics operations make road transport the primary contributor to air pollution and carbon emissions both.

In addition, the western region, including Luoyang, Pingdingshan, and Jiyuan, utilizes its abundant coal resources to support the development of energy and chemical industries, leading to increased SO₂ emissions and a high carbon dioxide intensity. Jiyuan exemplifies this high-carbon emission model, with 378855 tons of SO₂ emission and 349.7 Mt of CO₂ emission in 2022. Its carbon intensity of 438378 t/10⁸yuan GDP is 3.7 times higher than that the provincial average. Emission source analysis reveals that industrial source contribute 94.0% of SO₂ emission, while industrial source (41.7%) and electricity and thermal power source (42.2%) together account for 83.9% of CO₂ emission.

In contrast, the southern region, encompassing Zhumadian, Zhoukou, and Xinyang, serves as a major agricultural production area characterized by larger NH₃ emission from farming activities but relatively lower carbon emission totals and intensities. Zhumadian illustrates this pattern with primary industry comprising 17.6% of GDP—the province's highest proportion. The substantial livestock populations, including 6.74 million pigs and 53.6 million poultry, generate 76792 tons of NH₃ emissions (ZBS, 2023), with agricultural source accounting for 96.5%. Conversely, regional CO₂ emission amount total only 163.3 million tons, resulting in an emission intensity of 5013 t/10⁸yuan GDP—markedly lower than that of northern industrial cities. This disparity reflects the significantly lower energy demand associated with agricultural production system compared to industrial ones.

Finally, for the eastern region, which includes Shangqiu and Xuchang, integrates industrial production, agricultural activities, and transportation functions, leading to moderate overall pollutant emissions and intermediate CO₂ intensities. Xuchang's emission profile exemplifies this multi-source pattern: industrial sources contribute 87% of SO₂ emission (22452 tons); NO_x emissions (58434 tons) are predominantly from industrial source (48.9%) and mobile source and oil storage and transportation (27.2%); PM_{2.5} emissions (20842 tons) are mainly generated by industrial (55.0%) and fugitive dust



385 (16.1%); NH₃ emissions (23510 tons) are overwhelmingly attributed to agricultural activities (94.6%) of;
 and VOCs emissions originate from a variety of sectors, including from industrial (52.1%), residential
 (24.9%), and mobile source and oil storage and transportation with petroleum operations (19.2%). This
 complex, multi-source emission structure adds the complexity of integrated pollution control strategies
 for this region.

390 3.3 Comparison and evaluation of emission inventory

3.3.1 Comparison with others

To evaluate the reliability of the emission inventory developed in this study, we conducted a
 comparative analysis with several existing studies and databases, as summarized in Table 2.

First, regarding GHGs, the estimated CO₂ emissions in this study are notably higher than those
 395 reported by MEIC and CEADs. This difference primarily stems from variations in activity data source
 and spatial resolution. While all studies employ the emission factor approach, our estimates are derived
 from individual enterprise-level accounting of 7390 designated size industrial enterprises, incorporating
 detailed terminal energy consumption data from power and heating sectors, using one-to-one enterprise-
 level accounting rather than aggregated statistical data. In contrast, MEIC and CEAD rely on provincial
 400 energy balance sheets. That being said, CHRED's CO₂ emissions figures for Henan Province are a tad
 higher than what we've come up with. We reckon this discrepancy boils down to differences in what each
 of us is counting as part of the whole. CHRED's throwing both direct and indirect CO₂ emissions into
 the mix, while we're just looking at the direct stuff.

In terms of air contaminants, our findings show that the projected discharges of SO₂, NO_x, PM₁₀
 405 and PM_{2.5} are actually more substantial than the figures provided by the MEIC. This can be attributed to
 the treatment of key emission source categories through enterprise-specific activity data. For industrial
 source, emissions were calculated through one-to-one enterprise-level accounting, particularly for high-
 emission industries such as cement, steel, and chemicals.

In contrast, the estimated emissions of NH₃, CO and VOCs are slightly lower than those reported in
 410 MEIC. the open burning of crop residues was estimated using satellite-based fire hotspot data. The
 number of fire points was used to adjust burning ratios across different regions, thereby generating less
 region-specific emission factors. The discrepancy in CO and VOCs can be primarily attributed to the use



Table 2. Comparisons with other study.

Region	Year	GDP/Billion CNY	Pollutant/kt								Reference		
			SO ₂	NO _x	CO	PM ₁₀	PM _{2.5}	VOC _s	NH ₃	CO ₂		CH ₄	N ₂ O
Henan	2022	62106	408.7	1336.2	4647.3	901.1	440.0	759.3	672.7	540462.0	12461.97	42.93	This study
Henan	2020	54691	232.7	938.0	5784.6	342.1	265.2	1181.7	789.0	427679.2			MEIC
Henan	2020	62106								572790.0			CHRED
Henan	2020	54691								483738.0			CEADs
Guangdong	2020	88521	181.0	985.0			191.00	976.0	390.0				Li et al., 2023
Shanxi	2017	55216	131.0	381.7		279.5	737.7	206.1					Bo et al., 2023

Annotation: MEIC (Multi-resolution Emission Inventory for China, developed by Tsinghua University, http://meicmodel.org.cn/?page_id=560); CHRED (China High-Resolution Emission Database, Developed by the Department of Earth System Science, Tsinghua University, College of Environmental Sciences and Engineering, Peking University; Institute of Atmospheric Environment, Chinese Research Academy of Environmental Sciences, <https://www.cityghg.com/toArticleDetail?id=203>); CEADs (China Emission Accounts and Datasets, Developed by Institute of Climate Change and Sustainable Development, Tsinghua University, <https://www.ceads.net/data/province/>)



of localized emission factors for mobile source and oil storage and transportation. These factors were
415 adjusted using local meteorological data, vehicle age structure, and fuel standards. By employing region-
specific parameters, our approach mitigates the risk of overestimation that may arise from using national
average emission factors.

Additionally, to compare regional emission levels, we analyzed emission inventory data from other
provinces, such as Guangdong and Shaanxi (Li et al., 2023; Bo et al., 2023). The results indicate that
420 Henan Province generally exhibits higher emissions of most pollutants. It can be explained by differences
in industrial and energy structures. As a major agricultural and industrial province in Central China, our
study region relies heavily on coal and maintains intensive agricultural activities, which contribute to
elevated emissions of SO₂, NO_x, NH₃, and CH₄. In contrast, Guangdong benefits from a cleaner energy
structure and an economy dominated by the service sector, thereby reducing emissions. Although
425 Shaanxi possesses a substantial industrial base, with the primary sector accounting for 7.8%, the
secondary sector for 47.6%, and the tertiary sector for 44.6% (SPBS, 2023), its industrial structure differs
significantly from that of Henan. Coal mining constitutes a large proportion of Shaanxi's industrial
structure, with annual raw coal production consistently ranking third nationally. Additionally, Shaanxi
has a lower vehicle ownership compared to Henan, resulting in emission differences between the two
430 provinces.

3.3.2 Uncertainty analysis

Uncertainties in emission inventories arise mainly from the volatility and incompleteness of data on
emission factor and activity data of emission source, as well as errors and under-representation of
monitoring instruments. These issues are inherent challenges in constructing emission inventory (Super
435 et al., 2020). This study employs Monte Carlo simulation to assess emission estimate uncertainties. The
uncertainty ranges of emissions, simulated mean value and emission at 95% confidence intervals for each
emission source and total source are presented in Table 3. Among all pollutants, the emissions of NH₃,
VOCs, and N₂O show relatively high uncertainty. This is mainly due to the limited representativeness of
emission factors and difficulties in collecting reliable activity data. For NH₃, the main source of
440 uncertainty lies in activity data, with waste treatment showing the highest uncertainty (-73%, 75%),
followed by agriculture (-54%, 52%). The primary reason is the complexity of emission processes and



Table 3. Uncertainty of the emission inventory.

Category	Mobile source and oil					
	Electricity and thermal power	Industry	storage and transportation	Fugitive dust	Resident	Agriculture
SO ₂	(-18%, 16%)	(-45%, 46%)	(-57%, 55%)	-	(-36%, 38%)	(-33%, 34%)
NOx	(-32%, 33%)	(-30%, 29%)	(-28%, 28%)	-	(-28%, 28%)	(-72%, 75%)
CO	(-28%, 31%)	(-33%, 34%)	(-60%, 60%)	-	(-49%, 45%)	(-105%, 99%)
PM ₁₀	(-38%, 40%)	(-28%, 28%)	(-42%, 2%)	(-54%, 51%)	(-58%, 55%)	(-70%, 70%)
PM _{2.5}	(-39%, 41%)	(-44%, 45%)	(-62%, 57%)	(-46%, 48%)	(-38%, 38%)	(-73%, 67%)
VOCs	(-27%, 27%)	(-57%, 55 %)	(-65%, 63%)	-	(-34%, 36%)	(-52%, 56%)
NH ₃	(-37%, 34%)	(-37%, 41%)	(-31%, 28%)	-	(-31%, 32%)	(-54%, 52%)
CO ₂	(-41%, 42%)	(-46%, 42%)	(-36%, 37%)	-	(-22%, 21%)	-
CH ₄	-	-	(-47%, 47%)	-	(-40%, 39%)	(-28%, 30%)
N ₂ O	(-39%, 38%)	-	(-80%, 71%)	-	-	(-102%, 95%)



challenges in accurate monitoring. In waste treatment, different treatment technologies and variable waste composition contribute to emission uncertainties (Zheng et al., 2012). In agricultural activities, varying fertilization practices and changing climatic conditions significantly influence NH_3 volatilization, while farmers often lack detailed fertilization records (Lessmann et al., 2025). These factors collectively contribute to the substantial uncertainty in emission data. For VOCs, uncertainty mainly arises from the applicability of emission factors. In sectors such as mobile source and oil storage and transportation, emission levels vary widely due to vehicle conditions, fuel quality, and equipment aging, resulting in an uncertainty of up to -65% and 63%. For N_2O , uncertainty is primarily concentrated in agricultural source. Data Gaps in data regarding livestock manure management—including treatment methods, storage duration, and application practices—make it challenging to accurately quantify emissions, particularly for small-scale farms (Hu et al., 2024).

3.4 Characteristic of monthly variation and gridded spatial emission

3.4.1 Temporal variation

SO_2 , CO , PM_{10} and $\text{PM}_{2.5}$ exhibit elevated levels during winter months (January and December) (Figure 5). This is mainly due to the increase in pollutant emissions resulting from the large amount of coal and biomass combustion during the heating period in winter, while the poor atmospheric dispersion conditions in winter make it easy for pollutants to accumulate, thus leading to higher emissions. NH_3 and N_2O , which have emission in the summer months, especially June and July, and lowest -emission in the winter months (January and December). Primarily caused by elevated NH_3 and N_2O releases during summer months due to intensive agricultural practices, such as the application of nitrogen fertilizers and animal excreta management. In addition, high summer temperatures increase the volatilization of ammonia and soil microbial activity, further exacerbating emissions of these two pollutants. In winter, when agricultural activities are reduced and temperatures are lower, emissions of NH_3 and N_2O are significantly lower. Emission of CH_4 and VOCs are relatively stable throughout the year, with small fluctuations reflecting their stable sources of emissions. The highest -emission of NO_x are found in August, and the lowest emission are found in April and May. This results from higher NO_x levels in summer, driven by increased vehicle emissions and intensified photochemical reactions in high temperatures. CO_2 is highest in December and lowest in April. Winter CO_2 emissions predominantly stem



from fossil fuel burning. The substantial demand for heating during winter leads to an increase in fossil fuel consumption, causing emissions to reach their peak. In contrast, the climate in April is mild; heating requirements have ceased while cooling demands have not yet commenced, resulting in energy consumption returning to baseline levels.

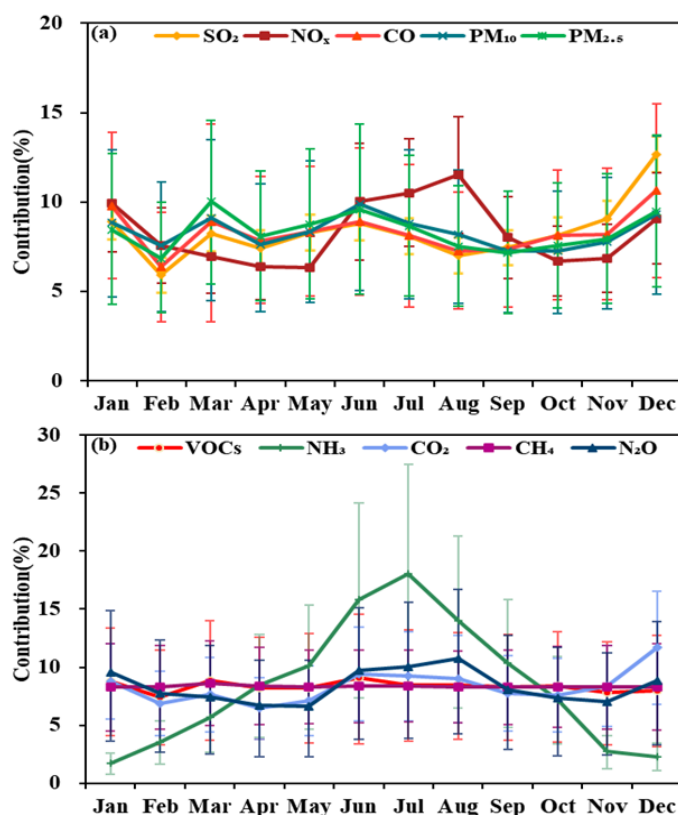


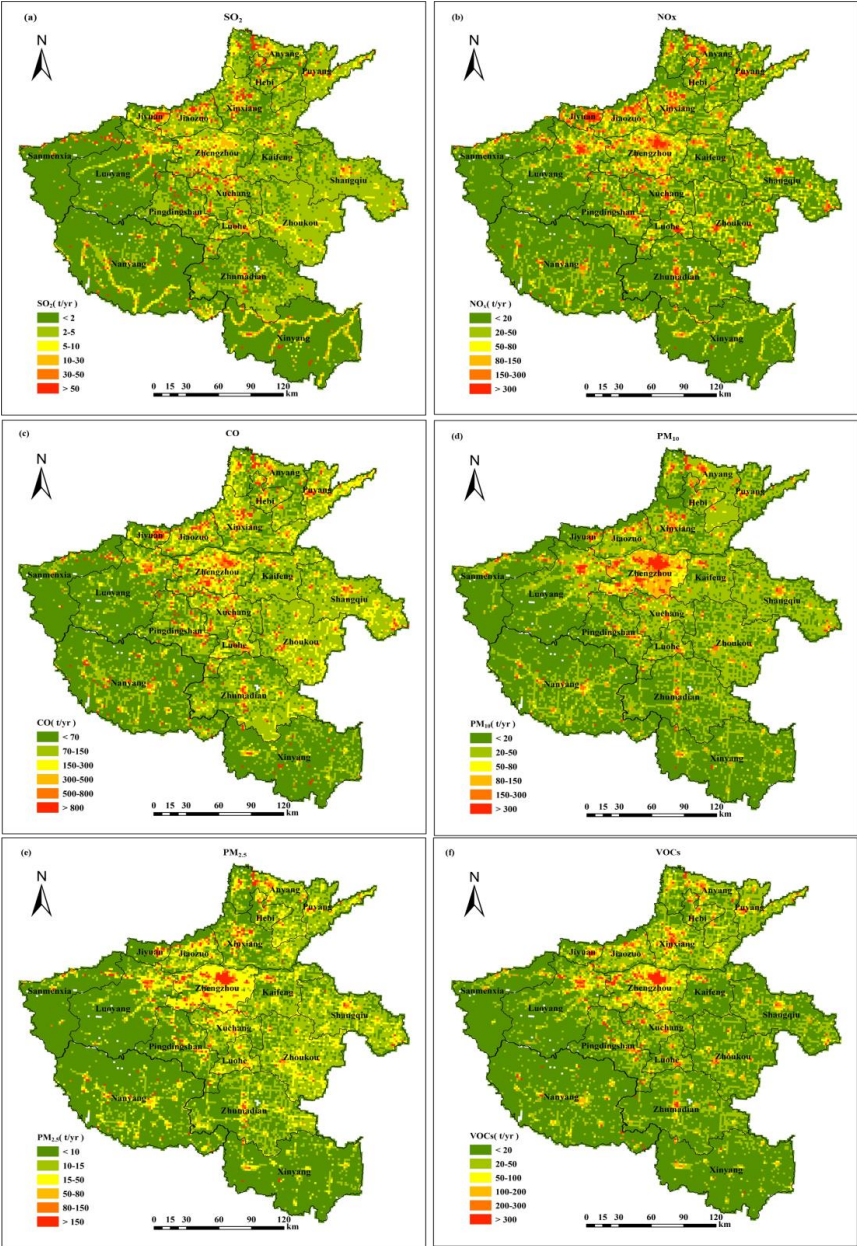
Figure 5. Monthly variations of air pollutants and GHG emissions in 2022 with uncertainty boundaries.

3.4.2 Spatial distribution

High-emission areas of SO₂, NO_x, CO, PM₁₀, PM_{2.5}, VOCs, and CO₂ are mainly concentrated in cities such as Zhengzhou, Luoyang, Jiaozuo, Xinxiang, Anyang, and Hebi, along with their surrounding regions (Figure 6). These regions are the province's primary industrial hubs and population hotspots, boasting notably higher emission levels than the rest of the areas. The layout of sulfur SO₂ and NO_x hotspots is tightly tied to the positions of coal-burning power stations, steelworks, and cement plants.



These industrial facilities are predominantly situated in heavy industrial cities like Anyang, Jiaozuo, Hebi, and Luoyang. PM₁₀ and PM_{2.5} emissions are primarily concentrated in Zhengzhou, Luoyang, Jiaozuo, 485 Xinxiang, and Xuchang. In the Zhengzhou metropolitan area, particulate emission exhibits a clear spatial distribution pattern, which is influenced not only influenced by industrial and fugitive dust source but



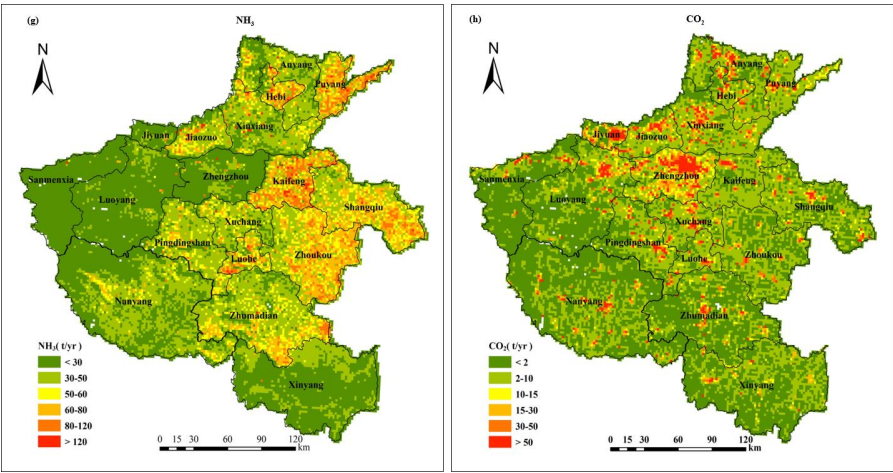


Figure 6. 3 km × 3 km gridded spatial distribution of air pollutants and GHG emissions.

also closely related to heavy urban traffic and high vehicle density. High-emission zones of CO and VOCs are observed in Zhengzhou and its surrounding cities, including Kaifeng, Xinxiang, and Xuchang. These cities have high road traffic density and account for more than half of the province's total vehicle ownership. In addition, the presence of petrochemical facilities, logistics centers, and fuel stations further contributes to elevated CO and VOCs emissions.

The spatial distribution of CO₂ is strongly correlated with energy consumption. Industrial and densely populated cities, such as Zhengzhou, Luoyang, Jiaozuo, Xinxiang, Anyang, and Hebi, exhibit relatively high levels of gridded CO₂ emission density. Collectively, these six cities contribute to over 51% of the province's total industrial energy consumption, amounting to more than 72 million tons of standard coal equivalent. Coal and electricity remain the dominant energy sources, leading to substantial carbon emissions.

In contrast, NH₃ exhibits a spatial pattern dominated by agricultural activities. Pronounced emissions concentrate across central, eastern plains with intense farming. Large-scale application of nitrogen fertilizer and livestock farming are the major emission sources of NH₃. Elevated NH₃ levels are also observed around Zhengzhou, primarily attributed to vehicle emissions from urban traffic.

3.5 Identification and spatial characterization of synergistic emission hotspots

Figure 7 demonstrates the contribution of major pollutants in Henan Province in 2022 (SO₂, NO_x,



PM_{2.5}, VOCs) and CO₂ synergistic hotspot areas contribute to the total emissions. From both pollutant and CO₂ emissions, co-commitment types typically contribute more to total emissions. Especially, category 1 (the grid of the top 5% of the emission intensity) occupies a large share of pollutants or CO₂ emissions, accounting for more than 50% of the total emissions. This indicates that the synergistic hotspots identified using gridded emission data and the top 20% quartile ranking effectively captured the spatially most concentrated emission areas with high precision. As can be seen in Figure 8, the synergistic hotspot areas of CO₂ and various pollutants in Henan Province exhibit significant spatial clustering characteristics, mainly concentrated in the north-central region. This concentration is particularly pronounced in the urban agglomeration centered around Zhengzhou, indicating that that this region represents overlapping zone of carbon emissions and multiple pollutant emissions, making it a critical region for synergistic management. Specifically, the synergistic hotspots of SO₂ and CO₂ are highly concentrated in the north-central region, with dense hotspot grids are observed in Zhengzhou and

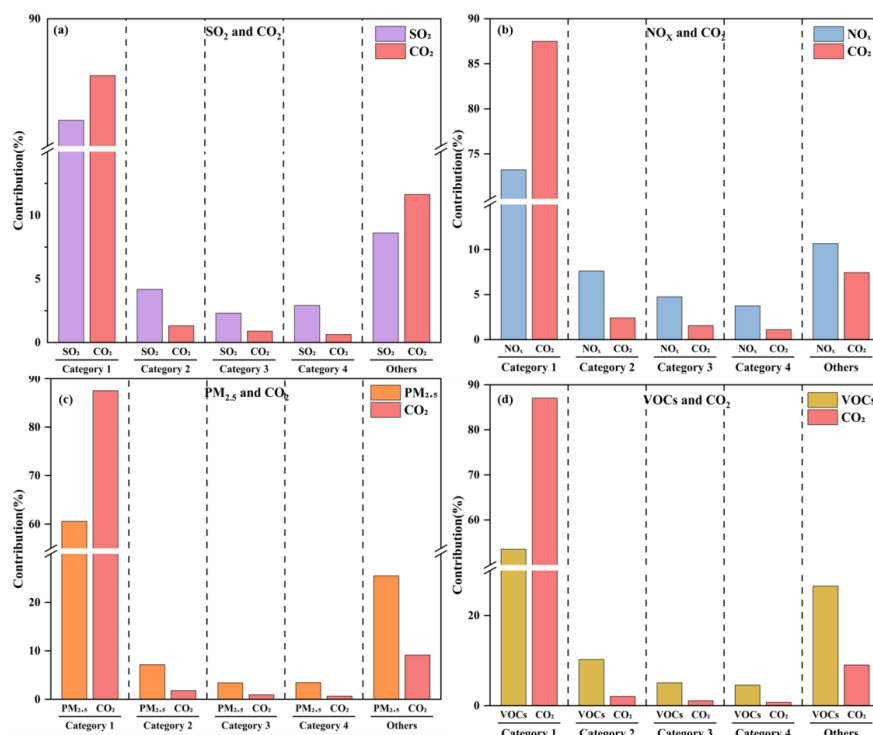
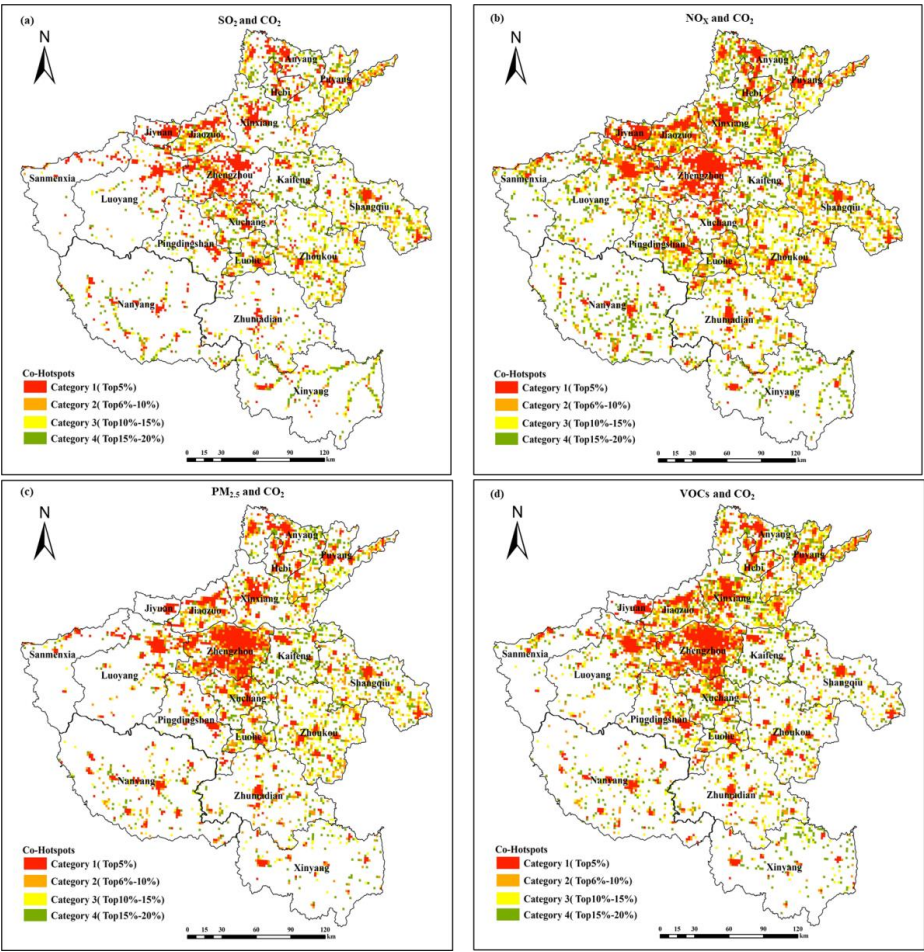


Figure 7. Emission contributions (%) of five categories to different emission co-hotspot. (a) SO₂ and CO₂, (b) NO_x and CO₂, (c) PM_{2.5} and CO₂, (d) VOCs and CO₂. Categories represent the grad of grid based on descending order of emissions: Category 1 (top 5% grids), Category 2 (6-10%), Category 3 (11-15%), Category 4 (16-20%), and Others.



Category 4 (16-20%), and Others (bottom 80%).



525 **Figure 8. Maps of the co-hotspots of (a) SO₂ and CO₂, (b) NO_x and CO₂, (c) PM_{2.5} and CO₂, and (d) VOCs and CO₂ in Henan Province.**

its neighboring cities. This distribution reflects the co-driving effect of heavy industrial activities, predominantly from coal-fired power plants, iron and steel, cement, and other high-energy-consuming industries, on SO₂ and carbon emissions. In contrast, the synergistic hotspots of NO_x and CO₂ form a distinct belt-like distribution within the urban agglomerations, along the major transportation arteries and industrial areas. This patterns suggests that their emission sources are influenced by the combined effects of industrial combustion and motor vehicle traffic, exhibiting typical multi-source superposition characteristics. The synergistic hotspots of PM_{2.5} and CO₂ demonstrate a broader distribution, in addition



to the dense hotspots in the urban core area, extending beyond the densely populated urban core to
 535 surrounding township areas. This wider dispersion indicates that their emissions are jointly influenced
 by a variety of sources, including industry, transportation, and residential life (e.g., heating in winter and
 bulk coal combustion), showcasing strong spatial diffusivity and composability. In contrast, the
 distribution of synergistic hotspots of VOCs and CO₂ is relatively dispersed. While there are still
 concentrated distributions in urban core areas, industrial parks, and along transportation arteries, the
 540 overall number of high-level hotspots is limited, and their spatial contiguity is weak. This indicates that
 the spatial coupling between VOCs emissions and CO₂ is relatively low, primarily originating from non-
 combustion industrial processes such as chemicals, petrochemicals, solvent use, and transportation.

4 Conclusions

In 2022, the emissions of major species in Henan Province were: SO₂ 408.7 kt (-43%, 44%), NO_x
 545 at 1336.2kt (-29%, 29%), CO at 4647.3 kt (-46%, 46%), PM₁₀ at 901.1 kt (-51%, 48%), PM_{2.5} at 440.0
 kt (-44%, 45%), VOCs at 759.3 kt (-56%, 55%), NH₃ at 672.7 kt (-54%, 52%), CO₂ at 540462.0 kt (-
 42%, 42%), CH₄ at 12462.0 kt (-38%, 39%), and N₂O at 42.9 kt (-64%, 58%). with industry, electricity
 and thermal power, mobile source and oil storage and transportation, and agriculture as the primary
 emission source. Pollutants and GHGs exhibited distinct spatial distribution patterns: the northern heavy
 550 industrial region was characterized by both high carbon intensity and elevated pollution levels, with
 carbon intensity ranging from 1.75 to 3.7 times the provincial average; the central transportation hub was
 dominated by NO_x and VOCs emissions; the western energy-chemical region showed prominent SO₂
 emission; the southern agricultural region was primarily influenced by NH₃ emission and exhibited low
 carbon emission; and the eastern multifunctional region displayed a multi-source coexistence pattern.

555 Temporally, SO₂ and PM_{2.5} peak during winter due to increased heating demand, whereas NH₃ and
 N₂O emissions are higher during summer agricultural season. Spatially, high-emission areas are
 predominantly located in the urban agglomerations of central and northern regions, particularly in
 Zhengzhou, Jiaozuo, and Anyang. Synergistic hotspot analysis indicates that the top 5% high-emission
 grid cells account for more than 50% of total emissions, with these areas primarily concentrated in major
 560 urban agglomerations.



5 Implication

This study reveals pronounced spatial heterogeneity in emissions across different cities and an extremely uneven spatial distribution pattern in the study region. This pattern is consistent with observations from other provinces in the North China Plain, suggesting that regional clustering of emission hotspot is a widespread phenomenon (Wu et al., 2024; Cai et al., 2018). Furthermore, distinct emission profile variations exist among cities due to differences in economic structures: northern heavy industrial cities exhibit both high carbon intensity and elevated pollution levels, presenting substantial potential for coordinated pollution and carbon reduction; central transportation hubs are dominated by NO_x and VOCs emissions, making them suitable for coordinated control strategy of mobile source; and southern agricultural regions show relatively low carbon emissions but higher NH₃ emission, necessitating comprehensive agricultural emission management approaches. These findings emphasize the need for region-specific, synergistic emission control strategies.

Consequently, future research and policy implementation should prioritize several key areas to promote synergistic emission reduction. First, systematically identifying high-emission hotspots can provide a solid scientific foundation for integrated air quality and climate governance. This underscores the need for enhanced collaborative emission reduction efforts that simultaneously address multiple pollutants and GHGs in these priority areas. Second, given the distinct emission patterns across regions, it is essential to develop region-specific control strategies achieve coordinated reductions of both air pollutants and GHGs. Third, high-resolution emission inventories should be leveraged to formulate more precise and coordinated management policies, thereby enhancing the overall effectiveness of synergistic emission reduction initiatives.

Data availability. The authors do not have permission to share data.

Author contributions. S. Yin conceived, coordinated the study and revised the manuscript. J. Li performed emission inventories, analyzed characteristics, and drafted the manuscript. C. Su provided research concepts and calculation assistance. J. Wei and M. Guan contributed to inventory development, and C. Yu provided manuscript feedback.

Competing interests. The contact author has declared that none of the authors has any competing interests.



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