



1 **Application and Evaluation of CRACMM V1.0** 2 **Mechanism in PM_{2.5} Simulation Over China**

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15 16 **Abstract**

17 Chemical mechanisms are one of the major sources of bias in chemical transport model
18 simulations, making their improvement a critical step towards enhancing model performance
19 and supporting air quality management and research. In this study, a newly developed chemical
20 mechanism, the Community Regional Atmospheric Chemistry Multiphase Mechanism
21 (CRACMM), integrated into the Community Multiscale Air Quality (CMAQ) modeling system,
22 was evaluated through comparison with two traditional chemical mechanisms, CB6r3_{ae7} and
23 Saprc07_{tic_ae7i}, for China. Sensitivity simulations related to precursor reactive organic carbon
24 (ROC) emissions were conducted to investigate the key driving factors of PM_{2.5} formation. The
25 results show slight differences in the correlation coefficient (R), mean bias (MB), and
26 normalized mean bias (NMB) values for the three chemical mechanisms when using the
27 traditional primary organic aerosol (POA) inventory. However, when using the full volatility
28 emission inventory, CRACMM shows improvements in R, MB, and NMB values in some
29 regions. CRACMM predicts higher PM_{2.5} concentrations during spring, summer and autumn,
30 mainly due to enhanced secondary organic aerosol (SOA) formation driven by increased
31 precursor emissions. Benzene–toluene–xylene (BTX) species and semi-volatile organic
32 compound (SVOC) emissions significantly contributed to PM_{2.5} formation in CRACMM. The
33 SOA from BTX emissions accounts for nearly 50% of the PM_{2.5} changes, while intermediate-



34 volatility organic compounds (IVOC) and SVOCs emissions mainly affect PM_{2.5}
35 concentrations through SOA formation. These results indicate that CRACMM, when using the
36 full volatile inventory, can effectively compensate for the underestimation of PM_{2.5} mass that
37 may occur with traditional POA treatment, particularly in regions with high photochemical
38 activity and abundant S/IVOC precursors.

39 **1. Introduction**

40 Exposure to airborne PM_{2.5} is associated with a variety of harmful health effects (2011; Liu et
41 al., 2024; Kim et al., 2015) and was reported to cause 4.14 million deaths worldwide annually
42 (95% confidence interval: 3.45 to 4.80) (Murray et al., 2020). Thus, a better understanding of
43 the PM_{2.5} formation mechanism is essential for formulating effective air pollution control
44 strategies.

45 The Chemical Transport Models (CTMs) serve as valuable tools for identifying factors
46 contributing to PM_{2.5} formation. They can reproduce a series of physical and chemical processes
47 that atmospheric pollutants undergo after being emitted into the atmosphere, driven by emission
48 inventory data and meteorological fields. Gas-phase chemical reaction mechanisms are an
49 essential part of the CTMs. Condensed mechanisms such as the Carbon-Bond (CB) mechanism
50 (Yarwood et al., 2005; Yarwood et al., 2010) and the Statewide Air Pollution Research Center
51 (SAPRC) mechanism (Carter, 2010, 1999; Carter, 2000) are widely used in CTMs. In these
52 mechanisms, various volatile organic compounds (VOCs) are lumped into functional groups.

53 The CB chemical mechanism has evolved since the 1970s, with CB7 being one of the most
54 recent versions. The carbon bonds are treated as a reaction unit, grouped by their bonding state,
55 without explicitly marking their location within the molecule. Aerosol Module 7 (aero7) is the
56 latest aerosol representation within the CMAQ model, developed by the U.S. Environmental
57 Protection Agency (EPA). Aero7 improves consistency in representing SOA formation
58 pathways between the CB- and SAPRC-based chemical mechanisms. It also updates
59 monoterpene SOA yields from photooxidation, adds uptake of water onto hydrophilic organics,
60 and includes consumption of inorganic sulfates (SO₄²⁻) when isoprene epoxydiol (IEPOX)
61 organosulfates are formed (Pye et al., 2013). Furthermore, it enhances computational efficiency
62 by using a volatility basis set (VBS) to parameterize SOA yields rather than using the Odum 2-
63 product fit (Zhang et al., 2021).



64 The SAPRC mechanism was first developed in the 1980s at the University of California,
65 Riverside. Saprc07tic is one of the most widely used versions (Xie et al., 2013), which
66 introduces refinements for isoprene chemistry. It is commonly used in regulatory and research
67 applications requiring a detailed representation of VOCs. Saprc07tic is a detailed, explicit
68 chemical mechanism that represents the atmospheric oxidation of a large number of VOCs,
69 providing a more nuanced representation of ozone formation and SOA production. Aero7 and
70 Aero7i modules contain the same major pathways to SOA, but AERO7i provides more
71 diagnostic information in terms of IEPOX SOA identification (Pye et al., 2017; Pye et al., 2013),
72 as well as some additional high-NO_x formation pathways to SOA (Pye et al., 2015).

73 CRACMM (Pye et al., 2023) is a state-of-the-art atmospheric chemistry mechanism and was
74 introduced with the release of CMAQ version 5.4 in October 2022 by the U.S. EPA. CRACMM
75 integrates gas-phase and particle-phase reactions, offering a wide-ranging representation of
76 atmospheric processes and enhanced capabilities for simulating multiphase chemistry in
77 regional air quality modeling. This approach helps models to more closely replicate
78 photochemical processes that occur in the atmosphere. CRACMM builds upon the well-
79 established Regional Atmospheric Chemistry Mechanism, version 2 (RACM2) framework
80 (Goliff et al., 2013) and incorporates enhanced representations of various organic compounds,
81 particularly monoterpenes and aromatics. Furthermore, CRACMM includes a built-in,
82 transparent mapping of emissions to mechanism species, ensuring conservation of emitted
83 carbon while tracking its transformation in products. CB6r3_ae7 and Saprc07tic_ae7i typically
84 exclude organic species with saturation concentrations (C_i^*) in the low-volatility organic
85 compound (LVOC, $0.3 < C_i^* \mu\text{g m}^{-3}$), SVOC ($0.3 \leq C_i^* < 300 \mu\text{g m}^{-3}$) and IVOC ($300 \leq C_i^* < 3 \times$
86 $10^6 \mu\text{g m}^{-3}$) ranges, which act as potential precursors to SOA formation (Chang et al., 2022).
87 SOA precursors beyond traditional, non-oxygenated volatile hydrocarbons such as S/IVOCs,
88 phenolic compounds, furans, and other oxygenated organic compounds are considered in
89 CRACMM (Pye et al., 2023).

90 Some previous studies have compared the model performance using various mechanisms in
91 CTMs. Luecken et al. (2019) compared CB6r3, CB5TU, and CB5 performances, and found
92 that CB6r3 performed best in simulating the vertical distribution of peroxyacyl nitrates.
93 Derwent et al. (2017) found that the condensed mechanisms (including CB6r3) and Master



94 Chemical Mechanism version 3.3.1 (MCMv3.3.1) have large differences in predicted hydroxyl
95 radical ($\text{HO}\cdot$) concentrations and their responses to NO_x and VOC reductions. For the SAPRC
96 mechanism, Kang et al., (2022) evaluated a highly condensed SAPRC chemical mechanism
97 and found that Saprc07 predicts slightly lower O_3 concentrations than the standard fix-
98 parameter version of the Saprc11. CRACMM predictions were compared with RACM2 and
99 CB6r3 over the northeastern US during summer. The results showed that CRACMM tends to
100 predict more O_3 than CB6r3 but less than RACM2 (Place et al., 2023). However, the $\text{PM}_{2.5}$
101 predictions by the newly developed CRACMM mechanism and their sensitivities to precursors
102 have not yet been investigated. In this study, $\text{PM}_{2.5}$ predictions from the CRACMM mechanism
103 were evaluated with surface observations comprehensively, covering different seasons and
104 regions. Results derived by CRACMM are compared with two well-established chemical
105 mechanisms, Saprc07tic_ae7i and CB6r3_ae7. The differences in $\text{PM}_{2.5}$ and SOA drivers
106 between CRACMM and the two existing mechanisms are further explored. The results of this
107 study provide a solid foundation for the further application of CRACMM in understanding and
108 regulating air pollution in China and globally.

109 **2. Methodology**

110 **2.1 Model configuration**

111 Model simulations were conducted using CMAQ v5.4 with a horizontal resolution of $36 \text{ km} \times$
112 36 km , covering mainland China (Figure 1). This domain includes five key city clusters with
113 notable air pollution levels: Beijing-Tianjin-Hebei (BTH), Yangtze River Delta (YRD), Pearl
114 River Delta (PRD), Fen-Wei Plain (FWP), and Sichuan Basin (SCB). Simulations were carried
115 out for the months of January, April, July, and October 2021, representing winter, spring,
116 summer, and autumn, respectively. The model includes 34 vertical layers, with the first layer
117 located approximately 35 meters above the ground. Each simulation was initialized with a 15-
118 day spin-up period before the start of each month. In addition to CRACMM version 1.0, two
119 other chemical mechanisms, CB6r3_ae7 and Saprc07tic_ae7i, were included for comparisons,
120 the number of reactions and gas- and particle-phase species in three different chemical
121 mechanisms used in CMAQ are shown in Figure S1. All three mechanisms are available in
122 CMAQ v5.4, with the "m3dry" deposition scheme selected. The initial and boundary conditions



for CRACMM and Saprc07tic_ae7i were mapped from the seasonal average hemispheric CMAQ output files distributed through the CMAS Data Warehouse. Meteorological input files were generated through an offline run of the Weather Research and Forecasting (WRF) model (<https://www.mmm.ucar.edu/models/wrf>) version 4.0 with configurations detailed in our previous studies (Huang et al., 2021a). The archived dataset, including the concentrations and model performance statistics of PM_{2.5} and its components, model configurations, and the locations of all observation sites, is available on Zenodo (Su et al., 2025).

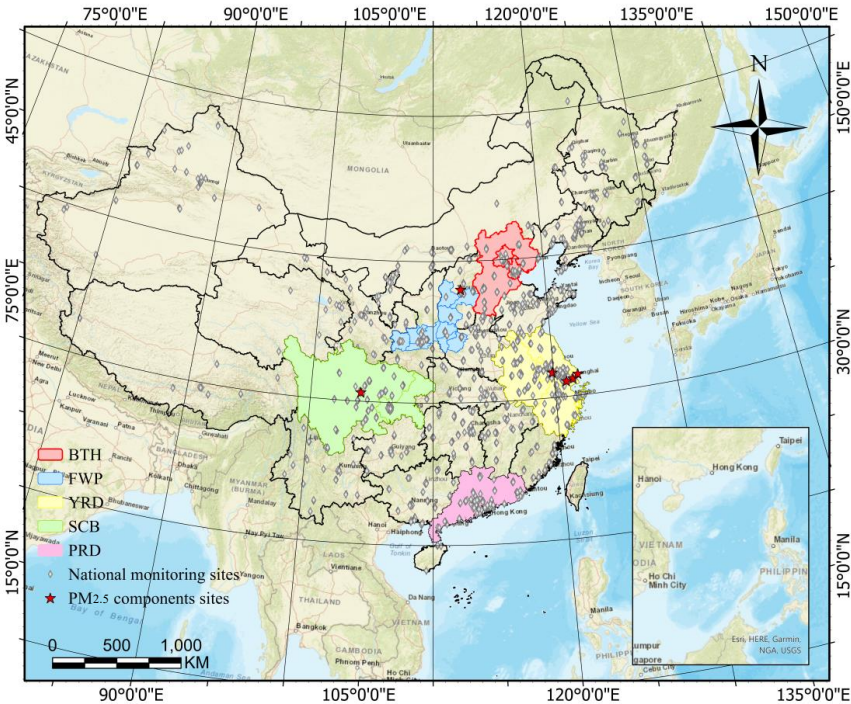


Figure 1. Model domain with five major city clusters (outlined in color), locations of national monitoring sites (grey diamonds), and six PM_{2.5} chemical components observation sites (red stars).

2.2 Emissions

2.2.1 Traditional Emissions Inventory

The 2019 anthropogenic Multi-resolution Emission Inventory for China (MEIC), developed by Tsinghua University, was utilized in this study (<http://www.meicmodel.org>). Biogenic emissions were estimated using the Model of Emissions of Gases and Aerosols from Nature



version 3.2 (MEGANv3.2, <https://bai.ess.uci.edu/megan/data-and-code/megan32>) (Guenther et al., 2012). Currently, MEIC supports VOC emission only for both CB6r3_ae7 and Saprc07tic_ae7i, but not the CRACMM mechanism. Therefore, anthropogenic VOC species were converted to CRACMM input species using a binary decision tree approach. This approach distinguishes between one-to-one and non-one-to-one mappings based on chemical species correspondence from Saprc07tic_ae7i and CB6r3_ae7 to CRACMM. The one-to-one mappings are further classified into explicit one-to-one (routine A) and lumped one-to-one (routine B), while the non-one-to-one mappings include many-to-one (routine C) and many-to-many (routine D) cases. In routine A, both Saprc07tic_ae7i and CB6r3_ae7 consist of a few species that can be mapped directly to CRACMM based on CAS number, e.g., HCHO (formaldehyde) is mapped to HCHO. In routine B, the mapping is based on lumped species categories or names. For example, OLE1 (alkenes other than ethene, with $\text{kOH} < 7 \times 10^4 \text{ ppm}^{-1} \text{ min}^{-1}$) in Saprc07tic_ae7i is mapped to OLI (internal alkenes) in CRACMM, OLE2 (alkenes with $\text{kOH} > 7 \times 10^4 \text{ ppm}^{-1} \text{ min}^{-1}$) is mapped to OLT (terminal alkenes), and ONIT is mapped to RNO3 based on the same name used for organic nitrates.

In routine C, new species could be added to the CRACMM mechanism, such as CSL (Cresols) and PHEN (Phenol and aromatic diols), which correspond to CRES (phenols and cresols) in the Saprc07tic_ae7i mechanism. In this case, we use the emission factor ratio to distribute the species. As MEIC does not provide species-level emission factors, data from the 2017 U.S. National Emission Inventory (NEI) (Pye et al., 2023) were utilized, which contain over 3,000 species with corresponding emission factors, source sectors, and CRACMM species mappings. The emission factors are averaged over all sources for different MEIC sectors (mobile sources, industrial sources, etc.). Another example is that both XYE (P-xylene and less reactive aromatics) and XYM (M-xylene and more reactive aromatics) are newly introduced species in the CRACMM mechanism, corresponding to XYL (Xylene and other aromatics) in the Saprc07tic_ae7i mechanism. According to the 2017 NEI, their ratio is 0.3:0.7. Therefore, 0.3 of XYL in Saprc07tic_ae7i is assigned to XYE in CRACMM, and 0.7 of XYL in Saprc07tic_ae7i is assigned to XYM in CRACMM.

Routine D is more complicated, but the mapping is still based on the emission factor ratio for proper mapping. For instance, in CRACMM, GLY represents both glyoxal and glycolaldehyde.



169 To construct this species from Saprc07tic_ae7i, GLY (representing glyoxal only) and part of
170 CCHO (glycolaldehyde and acetaldehyde) are mapped, such that GLY in CRACMM
171 corresponds to $GLY + CCHO \times 0.25$ in Saprc07tic_ae7i. Similarly, ACD (acetaldehyde) in
172 CRACMM corresponds to $CCHO \times 0.75$ in Saprc07tic_ae7i. For species lumped from multiple
173 species, only those with larger emission factors are considered. Table S1 outlines the
174 correspondence relationships for major species with substantial emissions. Since it is
175 challenging to compare VOC emissions in different mechanisms due to the lumping rules, we
176 only conducted an overall comparison of total emissions, as shown in Table S2.

177 2.2.2 POA Emissions

178 Two POA inventories were employed in this study: a traditional POA emissions inventory and
179 a full volatility inventory. CRACMM, Saprc07tic_ae7i, and CB6r3_ae7 all use the same dataset
180 for the traditional POA emissions inventory. This inventory applies a VBS profile based on
181 Woody et al. (2016) and Robinson et al. (2007), treating POA as semi-volatile with C_i^* values
182 ranging from 10^{-2} to 10^3 $\mu\text{g}/\text{m}^3$. The detailed species of POA included in each mechanism are
183 listed in Table S3.

184 In contrast, the full volatility inventory distributes POA emissions across a wider range of
185 volatility bins. Laboratory experiments have demonstrated that L/S/IVOC emissions, which are
186 largely absent in the traditional POA inventory, contribute to SOA formation much more
187 efficiently than VOCs, owing to their lower volatility. To capture these processes, the full
188 volatile inventory developed by Chang et al. (2022) was used, with species mapped from the
189 two-dimensional VBS (2D-VBS) mechanism to CRACMM based on their C_i^* and O:C values.
190 Since neither Saprc07tic_ae7i nor CB6r3_ae7 includes a representation of full volatile POA,
191 only CRACMM can utilize this comprehensive inventory. The methodology outlined by Chang
192 et al. (2022) includes emissions from various sources, along with their corresponding profiles,
193 volatility ranges, and emission amounts. For traditional POA inventory, the POA emission
194 amount was 2840 kt/y, while the new full volatile emission inventory includes emissions of
195 LVOC (1,342 kt/y), SVOC (1,169 kt/y), and IVOC (3,939 kt/y), resulting in a total of 6,450
196 kt/y. The new inventory fills a gap of 3,610 kt/y in L/S/IVOC emissions that were absent from
197 the traditional inventory. To thoroughly evaluate CRACMM and compare it with CB6r3_ae7
198 and Saprc07tic_ae7i, four simulation scenarios were designed, as shown in Table 1.



199 Additionally, the empirical representation of anthropogenic SOA source (pcSOA) (Murphy et
200 al., 2017) was turned off in CB6r3_ae7 and Saprc07tic_ae7i, as pcSOA is deprecated in
201 CRACMM.

202 **Table 1.** Description of simulation scenarios and their emissions

Scenarios	Mechanisms	POA emission inventory	Anthropogenic + Biogenic emission inventory
1	CB6r3_ae7	Traditional POA inventory	MEIC+MEGAN
2	Saprc07tic_ae7i	Traditional POA inventory	MEIC+MEGAN
3	CRACMM	Traditional POA inventory	MEIC+MEGAN
4	CRACMM	Full volatile inventory	MEIC+MEGAN

203 **2.3 Observational data and model performance evaluation**

204 Hourly concentrations of PM_{2.5} at national monitoring stations were obtained from the China
205 National Environmental Monitoring Centre (<http://air.cnemc.cn:18007>), which were then used
206 to evaluate model performance. Field observational data of PM_{2.5} chemical components
207 including NO₃⁻ (nitrate), SO₄²⁻, NH₄⁺ (ammonium), OC (organic carbon), and EC (elemental
208 carbon) at six super monitoring station sites were collected, as detailed in Figure 1 and Table
209 S4. Missing observation periods were excluded from the analysis. Model performance was
210 assessed using well-established statistical metrics, including the R, MB, NMB, root mean
211 square error (RMSE), normalized mean error (NME), and index of agreement (IOA). The
212 equations for calculating these metrics can be found in our previous study (Wang et al., 2024;
213 Huang et al., 2021a).

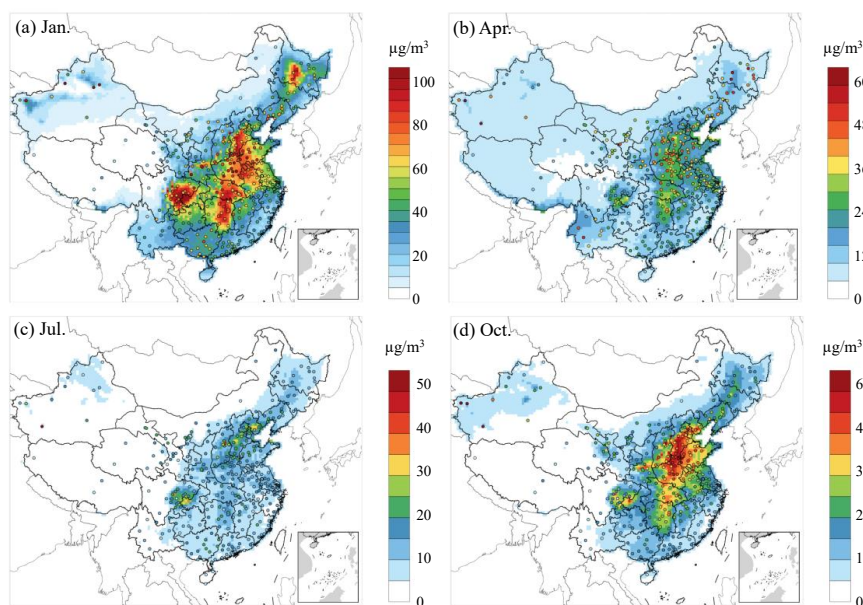
214 **3. Results and discussion**

215 **3.1 Overview of CMAQ-CRACMM model performance evaluation on PM_{2.5}**

216 Figure 2 depicts the spatial distribution of observed (dots) and simulated PM_{2.5} concentrations
217 for January, April, July, and October 2021, based on the CRACMM model with the full volatile
218 inventory. In January (Figure 2a), PM_{2.5} concentrations range from 5 to over 100 µg/m³, with
219 the highest values concentrated in the North China Plain (NCP) and parts of the SCB. These
220 elevated levels are primarily driven by relatively higher anthropogenic emissions and stagnant



221 meteorological conditions typical of winter. In April (Figure 2b), concentrations have
222 significantly decreased, ranging from 5 to 40 $\mu\text{g}/\text{m}^3$, with the most notable reductions observed
223 in northern regions. During July (Figure 2c), $\text{PM}_{2.5}$ concentrations were at their lowest, typically
224 ranging from 0 to 40 $\mu\text{g}/\text{m}^3$. This decline is mainly attributable to the increased precipitation
225 and the associated washout of pollutants. In addition, higher planetary boundary layer (PBL)
226 heights during the warm season, coupled with enhanced atmospheric mixing and dilution,
227 further contributed to the decrease in $\text{PM}_{2.5}$ concentrations. In October (Figure 2d), $\text{PM}_{2.5}$
228 concentrations rise again, ranging from 5 to 60 $\mu\text{g}/\text{m}^3$, with the highest concentrations observed
229 in the NCP and along the eastern coastal regions, which is attributed to the heating in later
230 autumn and unfavorable meteorological conditions. Overall, monthly variations in $\text{PM}_{2.5}$
231 concentrations are primarily driven by meteorological conditions and the distribution of
232 emission sources.



233
234 **Figure 2.** Monthly average $\text{PM}_{2.5}$ concentrations predicted (raster) by CRACMM and observed
235 (dots) in 2021 using the full volatility emission inventory.

236 The performance of the CMAQ model in simulating hourly $\text{PM}_{2.5}$ concentrations was evaluated
237 by comparing the model outputs with observations from national monitoring sites. In January,
238 CRACMM exhibited generally high R values in Northern China, indicating strong agreement



239 between observed and simulated $PM_{2.5}$ levels. In contrast, lower R values were observed in
240 southern regions. The model demonstrated a negative bias across most areas (Figure 4a),
241 suggesting a general underestimation of $PM_{2.5}$ concentrations, except in the YRD and SCB
242 regions, where positive biases are observed at many sites. This discrepancy is likely due, in part,
243 to the frequent dust storm events in January of that year. Although both observations and model
244 outputs from the major dust episodes on January 13–14 were excluded to minimize their
245 influence on the monthly evaluation, the missing of dust emissions, the complex meteorological
246 conditions associated with persistent northern dust layers may have contributed to an
247 underestimation of $PM_{2.5}$ concentrations by up to $30 \mu\text{g}/\text{m}^3$ in northern regions. Nevertheless,
248 the model still achieves high R values in the BTH, YRD, and FWP regions.

249 Results for April (Figures 3b and 4b) show generally strong correlations in the eastern regions,
250 while several monitoring sites in the south exhibit lower R values. Compared to January, the
251 MB is less pronounced. April also experiences dust storm events. In July, R values decline
252 across all regions relative to January and April, and most stations exhibit relatively small MB
253 values (Figures 3c and 4c). The strong influence of temperature and solar radiation on
254 photochemical processes during summer may result in more pronounced diurnal variations in
255 chemical composition, making the simulation of chemical processes more challenging.
256 Moreover, the chemical mechanisms may inadequately capture non-linear interactions and the
257 influence of SOA, further reducing the correlation. Additionally, synoptic-scale variations can
258 also affect the spatial distribution and concentration of key atmospheric species. R values are
259 improved in October (Figures 3d and 4d), with 90% of the sites achieving R values of 0.8 and
260 the MB is around $10 \mu\text{g}/\text{m}^3$ with higher evaluation in SCB and BTH regions. Overall,
261 wintertime observed peaks generally underestimated and lower summertime observed values
262 generally captured. The model demonstrates strong performance in January and October,
263 characterized by higher correlations and smaller biases. However, it had weaker performance
264 in April and July with lower correlations.

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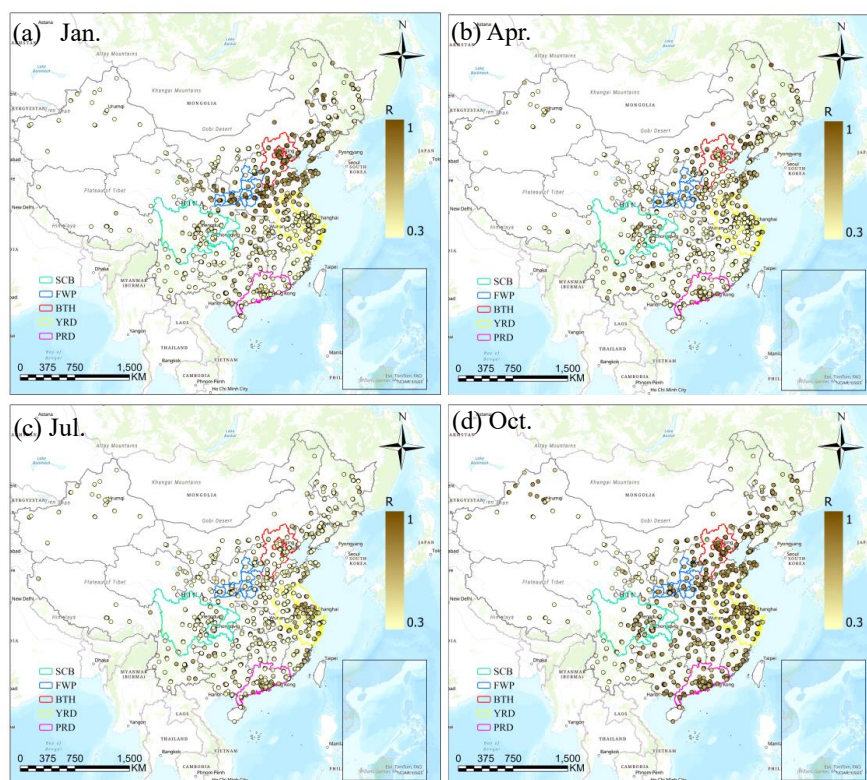
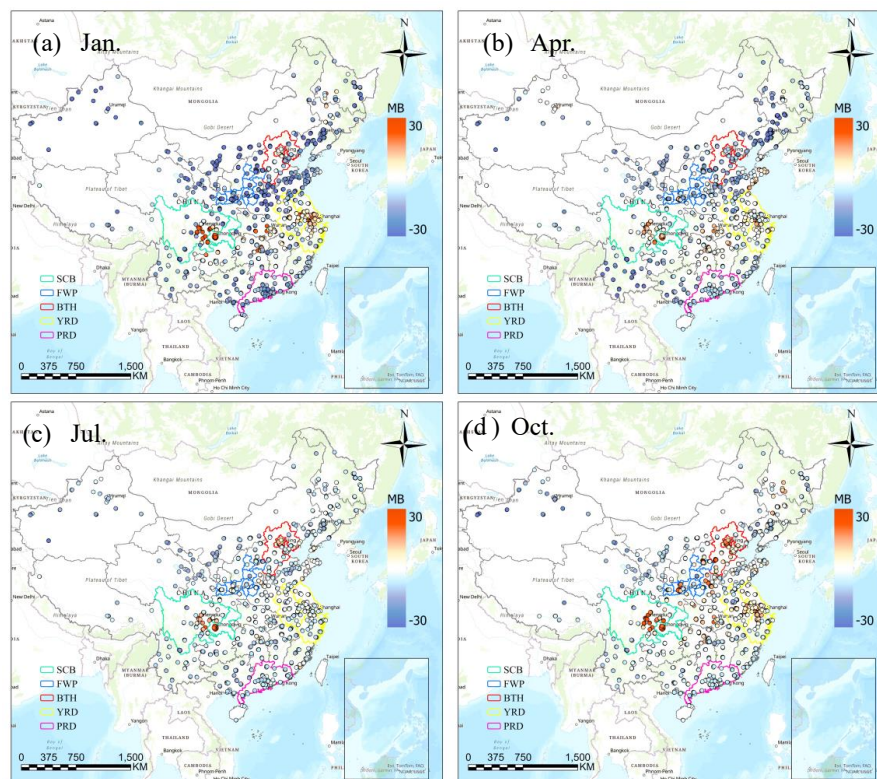


Figure 3. R values between predicted and observed PM_{2.5} concentrations using CRACMM with the full volatile emission inventory for January, April, July, and October of 2021.

Figures S2 and S3 compare the R and MB values between CRACMM (with the full-volatile inventory) and CB6r3_ae7 (with the traditional inventory). In January, CRACMM demonstrates notable improvements in R values at several sites in the PRD and YRD regions, with increases ranging from 0.2 to 0.4 (Figure S2a), while changes at most other sites remain relatively minor. Regarding MB, the most pronounced differences also occur in January: some locations in the BTH and YRD regions show higher MB values—up to 10 µg/m³—with CRACMM, whereas other regions display reduced MB values (Figure S3a). In April (Figures S2b and S3b), CRACMM achieves higher R values at certain sites in the YRD, while slightly lower correlations are observed in the PRD compared to CB6r3_ae7. MB values remain elevated in the SCB and parts of the YRD region for CRACMM. For July (Figures S2c and S3c), R values from CRACMM are generally comparable to those from CB6r3_ae7. However, MB values tend to decrease across most regions, indicating a potential improvement in bias



281 performance during summer. In October, CRACMM shows moderate increases in R values—
 282 by approximately 0.1 at most sites (Figure S2d). MB values are lower in the PRD region but
 283 higher in the FWP region compared to CB6r3_ae7 (Figure S3d).
 284 In evaluating the CMAQ model's performance for hourly $PM_{2.5}$ concentrations, CRACMM
 285 generally shows good correlations with observed data in January and October. However,
 286 discrepancies arise in April and July, likely due to complex meteorological and chemical
 287 conditions such as dust storms and increased photochemical activity. The model tends to
 288 underestimate peak $PM_{2.5}$ concentrations during winter but captures lower summer
 289 concentrations more accurately. Comparisons between CRACMM (with the full-volatile
 290 inventory) and CB6r3_ae7 (using the traditional inventory) highlight improvements in R values
 291 in the PRD and parts of the YRD regions in January for CRACMM, although performance
 292 declines in July.



293
 294 **Figure 4.** The MB values between predicted and observed $PM_{2.5}$ concentrations using
 295 CRACMM with full volatile inventory for January, April, July, and October of 2021.



PM_{2.5} components in six selected cities using CRACMM with the full volatile inventory were evaluated. The data of certain cities have been used in previous studies(Wang et al., 2024). Details for the selected cities are provided in Table S4, Tables S5-S8 present the statistical performance of PM_{2.5} components—including NO₃⁻, SO₄²⁻, NH₄⁺, OC, and EC—across four months. Figures S7-S10 demonstrate that CRACMM effectively captured the overall peaks and troughs of observed PM_{2.5} concentrations in January. The model also successfully simulated the heavy pollution period from January 20th to 25th in Taiyuan, with results similar to our previous study (Wang et al., 2024). The three ions were well simulated in both Changzhou and Pudong in the YRD region, particularly in Changzhou, where the overall trends of OC and EC showed strong consistency with observations. However, for EC in Pudong, the model struggled to capture the hourly peak values accurately.

Table 2. Monthly averaged metrics of PM_{2.5} evaluation and the number of monitoring sites.

Months	R	IOA	NMB	NME	No.
January	0.46	0.61	-18.58%	43.23%	1388
April	0.41	0.52	-29.07%	47.39%	1402
July	0.36	0.49	-32.39%	42.59%	1309
October	0.68	0.72	-11.48%	44.95	1358
Recommend benchmark	>0.60	>0.70	<±45%	<±55%	/

As shown in Table 2, based on observations from over 1,300 monitoring sites across China for CRACMM, the correlation metrics (R and IOA) met the recommended benchmark (Huang et al., 2021b) in October, while they fell short of the benchmark in January, April, and July. In contrast, the bias metrics (NMB and NME) satisfied the recommended benchmark across all four representative months, indicating that the model performed well in controlling overall bias. The evaluation metrics for different regions and months are provided in Tables S9–S12. The heatmaps in Figure S4 illustrate the variations in R values across the five key regions—YRD, SCB, PRD, FWP, and BTH—for the three chemical mechanisms over four representative months (January, April, July, and October). Corresponding heatmaps for MB and NMB are presented in Figures S5 and S6, respectively. In the YRD region, all mechanisms show relatively stable R values across the four months (Figure S4). The CRACMM simulation using the traditional inventory consistently results in lower MB (Figure S5) and NMB (Figure S6) values compared to CB6r3_ae7 and



321 Saprc07tic_ae7i throughout the year. When the full volatile inventory is incorporated, MB
322 improves in April and July but worsens in January and October. Similarly, NMB values indicate
323 higher modeled concentrations in all four months with the full volatile inventory compared to
324 the traditional one. This trend is consistent with Figures S5, and 5, where the YRD region shows
325 higher modeled concentrations from CRACMM using the full volatile inventory than from both
326 CB6r3_ae7 and Saprc07tic_ae7i across all months.

327 In the SCB region, R values remain relatively consistent across the three mechanisms (Figure
328 S4a–c). A slight improvement is observed in July, where R increases from 0.22 with the
329 traditional inventory to 0.27 with the full volatile inventory (Figure S4d). However, in October,
330 R decreases from 0.62 to 0.56 after switching to the full volatile inventory. The increase in R
331 value in July suggests that the traditional inventory may underestimate key precursors (e.g.,
332 S/IVOCs), while the full-volatility inventory better captures these species active at higher
333 temperatures, improving model–observation agreement. In contrast, the October decrease in R
334 may reflect uncertainties in representing some gas- or particle-phase organics under cooler
335 conditions. Regarding MB, CRACMM generally shows reduced values in most months when
336 using the traditional inventory, except for April. With the full volatile inventory, MB decreases
337 further in October, while slight increases are observed in the other months. The trends in NMB
338 follow a similar pattern to those in MB. It could be due to overestimation of certain
339 intermediate- or low-volatility species under specific conditions.

340 In the PRD region, CRACMM exhibits notable performance improvements in January. As
341 shown in Figures S4a–c, the R value increases from 0.20 with CB6r3_ae7 and Saprc07tic_ae7i
342 to 0.35 when using CRACMM with the traditional POA inventory. When the full volatile
343 inventory is applied (Figure S4d), the R value further increases to 0.50. Concurrently, the MB
344 improves significantly, decreasing from $-19.6 \mu\text{g}/\text{m}^3$ to $-11.8 \mu\text{g}/\text{m}^3$, and the NMB is reduced
345 from -43% to -26% . These results indicate that CRACMM, particularly with the full volatile
346 inventory, achieves both higher correlation and lower bias for $\text{PM}_{2.5}$ simulations in the PRD
347 region during January. Under the traditional POA inventory, CRACMM tends to underestimate
348 $\text{PM}_{2.5}$ concentrations in PRD during January (Figure S3a). However, after switching to the full
349 volatile inventory, simulated concentrations exceed those predicted by CB6r3_ae7 (Figure 5a),
350 primarily due to increased contributions from SOA (Figure 8b). In comparison, during April



351 and July, the CRACMM simulations using the traditional emissions inventory showed lower R
352 values than those of CB6r3_ae7 and Saprc07tic_ae7i, and the correlation further declined when
353 using the full-volatility emissions inventory. Although CRACMM features a more
354 comprehensive design for gas-phase oxidation mechanisms, the intense photochemical activity
355 in July and the rapid oxidation of high concentrations of IVOC precursors may have introduced
356 more complex SOA formation pathways and product distributions, thereby increasing modeling
357 uncertainties and weakening the agreement with observations. By October, model performance
358 had improved. Across all months, CRACMM combined with the full-volatility emissions
359 inventory consistently outperformed the other mechanisms in terms of MB and NMB,
360 highlighting the critical role of this inventory in addressing the underestimation of PM_{2.5}
361 associated with traditional POA treatment.

362 In the FWP region, both the R (Figure S4a–c) and MB (Figure S5a–c) values show minimal
363 variation across the three mechanisms, indicating limited sensitivity to the chemical mechanism
364 alone. However, notable changes in MB and NMB are observed with the incorporation of the
365 full volatile inventory in April, July and October, which align with the higher PM_{2.5}
366 concentrations shown in Figure 5, compared to CB6r3_ae7. In April and July, MB values shift
367 from -12.7 µg/m³ and -5.6 µg/m³ to -2.4 µg/m³ and 1.2 µg/m³, respectively, with NMB showing
368 a similar trend. These changes suggest an improvement in model agreement when the full
369 volatile inventory is employed. In contrast, both MB and NMB increase in October, indicating
370 that the full volatile inventory leads to higher simulated concentrations during this month.

371 In the BTH region, The R (Figure S4a–c) and MB (Figure S5a–c) values remain largely
372 consistent across the three mechanisms. CRACMM with the traditional POA inventory shows
373 a decrease in R values from 0.44 to 0.38 compared to CB6r3_ae7 in July. After incorporating
374 the full volatile inventory, the R in BTH experiences a more significant drop, falling to 0.24.
375 Notably, MB and NMB indicate that the modeled results are lower for CRACMM with the
376 traditional POA inventory in January, while they are higher in the other months. Additionally,
377 BTH exhibited the highest IVOC emissions in the inventory (Chang et al., 2022), and
378 uncertainties in emission estimates may have further contributed to this result.

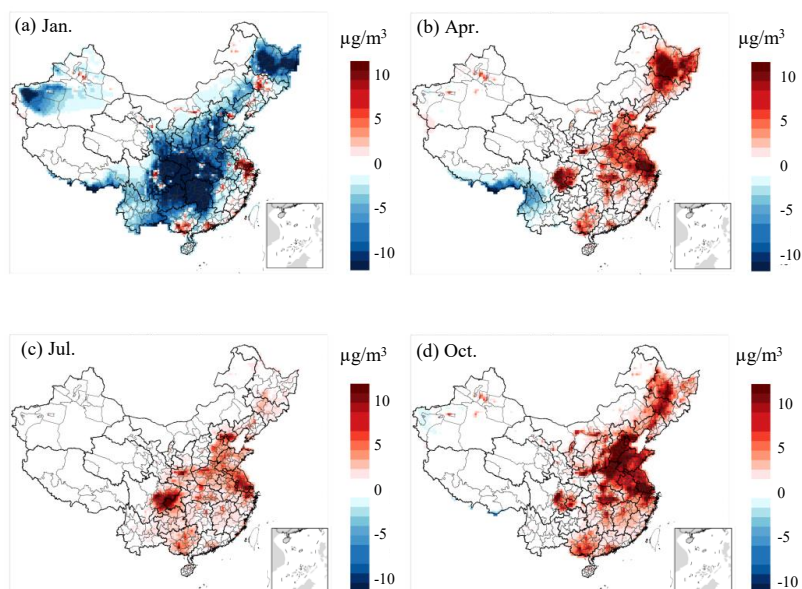
379 Overall, the differences in R and MB values across the three chemical mechanisms are
380 relatively small when the traditional POA inventory is used. However, for CRACMM, the MB



381 in January indicates a stronger underestimation, primarily due to differences in POA (as shown
382 in Figure 8a), leading to lower modeled concentrations compared to CB6r3_ae7 and
383 Saprc07tic_ae7i. With the incorporation of the full volatile inventory, the MB shifts toward
384 higher modeled concentrations in the subsequent three months. Notably, more pronounced
385 differences are observed in January in the PRD region and in July in the BTH region.

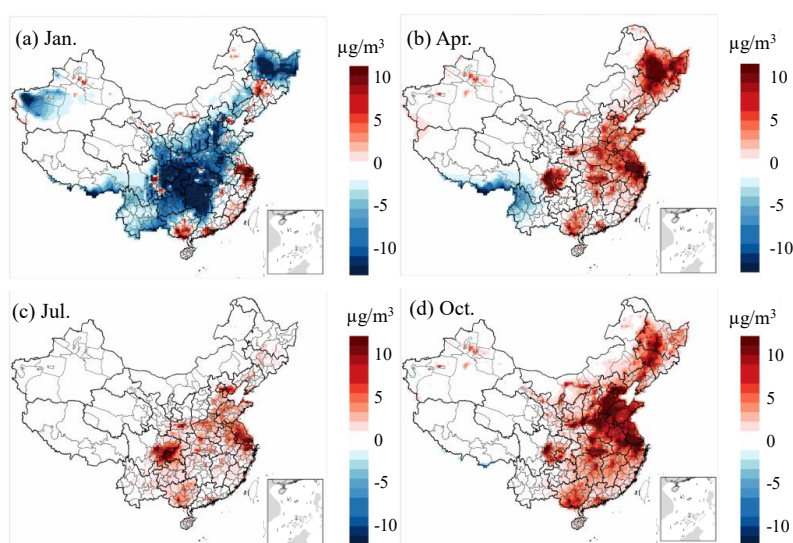
386 **3.2 Comparisons of model predicted PM_{2.5} between CRACMM and other mechanisms**

387 Figures 5 and 6 illustrate the differences in model outputs between CRACMM with full volatile
388 inventory and CB6r3_ae7, as well as between CRACMM and Saprc07tic_ae7i. In January,
389 CRACMM predicts lower PM_{2.5} concentrations across central and northern China compared to
390 both CB6r3_ae7 (Figure 5a) and Saprc07tic_ae7i (Figure 6a), with the differences—up to 10
391 µg/m³—observed in central and north of China. While CRACMM simulates higher PM_{2.5}
392 concentrations in the PRD and YRD regions. For the remaining months—April, July, and
393 October—CRACMM with full volatile inventory generally predicts higher PM_{2.5} levels than
394 the other two mechanisms (Figures 5b–d, Figures 6b–d). When CRACMM and CB6r3_ae7 are
395 configured with the traditional POA inventory, as shown in Figure S11, the differences of PM_{2.5}
396 concentrations are reduced in April, July, and October compared with full volatile POA
397 inventory. But CRACMM still predicts lower PM_{2.5} levels than CB6r3_ae7 in January. A likely
398 explanation is that the lower photochemical activity leads to reduced SOA formation, as the
399 enhanced SOA pathways are less active during the winter months.



400

401 **Figure 5.** Differences in model-predicted $PM_{2.5}$ concentrations between CRACMM (full
402 volatile inventory) and CB6r3_ae7.



403

404 **Figure 6.** Differences in model-predicted $PM_{2.5}$ concentrations between CRACMM (full
405 volatile inventory) and Sapr07tic_ae7i.



406 **3.3 Comparisons of model predicted PM_{2.5} chemical components between CRACMM**
407 **and other mechanisms**

408 Analysis of Figures 5, 6, and S11 indicates that the most significant differences among the
409 model simulations occur in January and October, whereas the discrepancies between chemical
410 mechanisms are substantially smaller in April and July—likely due to the overall lower
411 pollutant concentrations during these months, which may reduce the sensitivity to mechanistic
412 differences. Consequently, the subsequent analysis focuses on a detailed comparison of PM_{2.5}
413 component variations between CRACMM and CB6r3_ae7 for January and October.

414 **3.3.1 Inorganic aerosol**

415 SO₄²⁻, NO₃⁻, and NH₄⁺ are the dominant secondary inorganic components in PM_{2.5}. Nitrogen
416 dioxide (NO₂) and sulfur dioxide (SO₂) can fully dissolve into cloud water or aerosol liquid
417 phases and subsequently oxidize to form nitrate and sulfate. Ammonium salts are produced
418 through the neutralization reactions of these acidic species with atmospheric ammonia (NH₃).
419 EC primarily originates from the incomplete combustion of carbonaceous fuels, especially
420 under oxygen-limited conditions. It is commonly emitted from sources such as vehicle exhaust,
421 industrial combustion, and biomass burning. From a chemical mechanism perspective,
422 CRACMM retains the inorganic chemistry framework of RACM2 but incorporates updated
423 rate constants for several reactions. Specifically, the rate expressions for 26 inorganic reactions
424 were revised in CRACMM compared to RACM2 (Pye et al., 2023). Overall, differences in
425 inorganic component predictions among CRACMM, CB6r3_ae7, and Saprc07tic_ae7i are
426 relatively minor. As shown in Figure 7, predicted concentrations of major inorganic species in
427 January and October are comparable between CRACMM and CB6r3_ae7, with differences
428 ranging from -1 to 1 µg/m³ for EC and SO₄²⁻, and -5 to 5 µg/m³ for NH₄⁺ and NO₃⁻. These
429 results suggest that the variation in simulated inorganic aerosol concentrations is only
430 marginally affected by differences in the inorganic chemistry schemes.

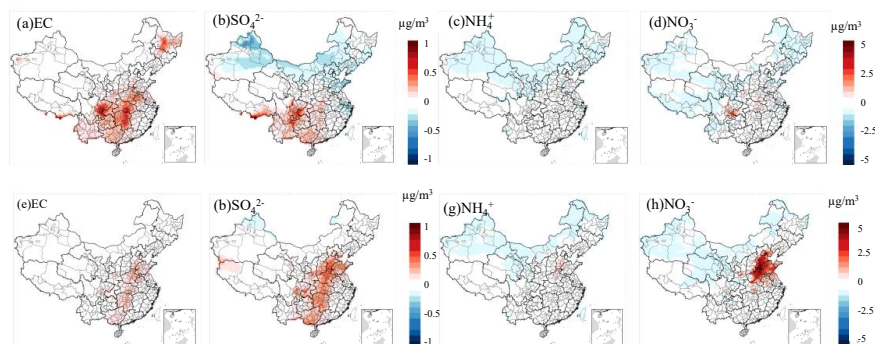


Figure 7. Differences in model-predicted $\text{PM}_{2.5}$ components—(a)EC, (b) SO_4^{2-} , (c) NH_4^+ , (d) NO_3^- for January, and (e)EC, (f) SO_4^{2-} , (g) NH_4^+ , (h) NO_3^- for October—between CRACMM (with full volatile inventory) and CB6r3_ae7. Figures (a–d) share the same scale, as do figures (e–h).

3.3.2 Organic aerosol

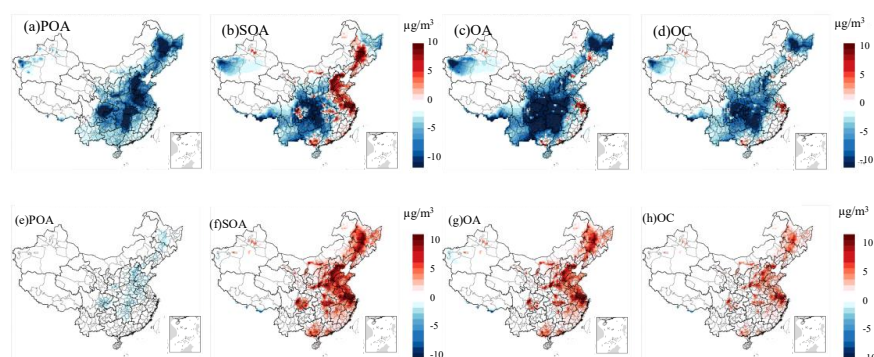
In January, CB6r3_ae7 consistently predicts higher POA concentrations than CRACMM under both the traditional and full-volatile inventory configurations, with the most pronounced differences occurring in east China (Figure 8a and Figure S12a). In CRACMM, POA aging is represented using a modified 2D-VBS framework (Murphy et al., 2017), where C_i^* range from 10^{-2} to $10^3 \mu\text{g}/\text{m}^3$. A significant portion of the alkane-like L/SVOC mass contributing to ambient OA comes from the direct emissions of low-volatility species (e.g., AROCN2ALK, AROCN1ALK, AROCP0ALK, AROCP1ALK, AROCP2ALK, AROCP3ALK) and their oxidation products (e.g., ROCN2OXY2, ROCP0OXY2, ROCP1OXY1, ROCP2OXY2, ROCP3OXY2). By contrast, CB6r3_ae7 adopts a semi-volatile POA approach in which primary emissions (e.g., LVPO1, SVPO1–3, IVPO1) and their oxidation products (e.g., LVOO1, LVOO2, SVOO1, SVOO2, SVOO3) partition between gas and particle phases across a C_i^* range of 10^{-1} to $10^3 \mu\text{g}/\text{m}^3$. This framework aligns with the 1.5D-VBS scheme proposed by Koo et al. (2014). Details of the POA species and their properties are provided in Table S3. The most significant differences in simulated POA concentrations occur in January, likely due to enhanced partitioning of SVOC to the particle phase under low wintertime temperatures. Additionally, differences in multigenerational oxidation aging and volatility treatment between



the two mechanisms contribute to the simulation discrepancies. In October, although POA concentrations in CRACMM remain lower than in CB6r3_ae7, the difference is less pronounced compared to January for both POA inventory (Figures 8e and S12e). To better understand the drivers of SOA formation in the two mechanisms, we analyzed the spatial distribution of SOA concentrations under both traditional and full-volatile POA inventories. In January, CRACMM predicts higher SOA levels in the BTH and parts of the SCB regions compared to CB6r3_ae7 when using the traditional POA inventory (Figure S12b). Under the full-volatile inventory, CRACMM also shows increased SOA concentrations in the YRD and PRD regions (Figure 8b). These increases correspond with high IVOC emissions in BTH and YRD, consistent with the spatial patterns reported by Chang et al. (2022), although their data reflect monthly averages across January and July. The SOA enhancement in YRD and PRD under the full-volatile inventory highlights the critical role of IVOC emissions in these areas. In CB6r3_ae7, SOA is primarily formed from the oxidation of traditional VOC sources, such as isoprene, monoterpenes, sesquiterpenes, benzene, toluene, xylene, alkanes, and PAHs (Carlton et al., 2010; Pye and Pouliot, 2012). In contrast, CRACMM incorporates additional SOA precursor systems, including phenol and aromatic diols, pinon aldehyde, oxygenated IVOCs, furanone, and other compounds. As a result, in regions with elevated anthropogenic emissions, CRACMM generally simulates higher SOA concentrations. However, in the SCB region, SOA levels remain lower, possibly due to the reduced reactivity of these new precursors under the lower ambient temperatures typical of this region. In terms of overall OA concentrations, CRACMM generally predicts lower values than CB6r3_ae7 across most regions, except for some part of YRD region (Figure 8c and Figure S12c). The spatial distribution of OC is similar to that of OA, with CRACMM also showing lower concentrations than CB6r3_ae7 (Figure 8d and Figure S12d). In October (Figures 8f–h and S12f–h), SOA remains the dominant contributor to the differences in PM_{2.5} concentrations. Under the full-volatile inventory, CRACMM predicts significantly higher SOA concentrations compared to CB6r3_ae7 (Figure 8f), resulting in elevated OA levels (Figure 8g). The spatial pattern of OC concentrations (Figure 8h) closely resembles that of OA. When the traditional POA inventory is applied, the differences in SOA, OC, and OA concentrations between the two mechanisms are minimal (Figures S12f–h). The most



483 pronounced increases with the full-volatile inventory are observed in the YRD, PRD, and SCB
484 regions, attributable to the inclusion of a more comprehensive set of SOA precursors.



485
486 **Figure 8.** Differences in model-predicted $PM_{2.5}$ components—(a) POA, (b) SOA, (c) OA, and
487 (d) OC for January, and (e) POA, (f) SOA, (g) OA, and (h) OC for October—between
488 CRACMM (with full volatile inventory) and CB6r3_ae7. Figures (a–d) share the same scale,
489 as do figures (e–h).

490 Overall, in January, the primary differences between CRACMM and CB6r3_ae7 stem from
491 lower POA concentrations in CRACMM, primarily due to semi-volatile partitioning and
492 reduced aging of semi-volatile POA species at lower temperatures. SOA concentrations are
493 elevated in eastern China but reduced over the SCB, reflecting both the slower oxidation of
494 additional SOA precursors under winter conditions in the SCB and the greater availability of
495 these precursors in the eastern region. In October, the key differences in model predictions are
496 primarily driven by the POA inventory used. The full-volatility inventory yields higher SOA
497 concentrations than the traditional inventory, largely due to the inclusion of L/S/IVOCs, which
498 are efficient SOA precursors.

499 3.4 Sensitivity study on $PM_{2.5}$ and SOA responses to changes of precursors

500 In this section, CMAQ simulations with emission perturbations are conducted to identify the
501 key drivers of $PM_{2.5}$ formation in January, when $PM_{2.5}$ concentrations are notably high. A series
502 of emission sensitivity simulations were performed within CMAQ to assess the role of
503 precursor ROC systems in $PM_{2.5}$ formation using CRACMM with the full volatile inventory
504 across China. These sensitivity simulations involved running zeroed emission scenarios for



January (i.e., setting emissions of a specific chemical class or sector to zero) to examine how $PM_{2.5}$ concentrations respond to changes in emissions. A subset of these sensitivity simulations was also conducted using the CB6r3_ae7 and Saprc07tic_ae7i mechanisms. A detailed list of all the zeroed emission simulations is provided in Tables 3 and S13. Due to the non-linear nature of $PM_{2.5}$ production in response to ROC perturbations, these simulations offer an initial evaluation of how $PM_{2.5}$ formation responds to reduced ROC emissions, providing valuable insights into how chemical systems behave under varying emission conditions in the three mechanisms.

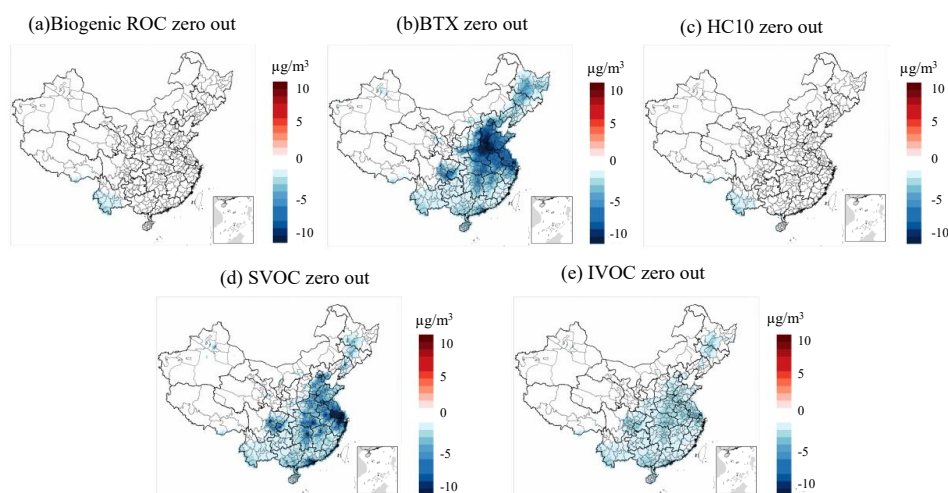
Table 3. List of emission reductions relative to the base simulations in CMAQ-CRACMM.

Chemical Mechanism	Emission Reduction
CRACMM	Benzene, toluene and xylene-like emissions set to zero
CRACMM	Biogenic-ROC emissions set to zero
CRACMM	IVOC emissions set to zero
CRACMM	SVOC emissions set to zero
CRACMM	HC10 zero out (decane and species of similar reactivity)

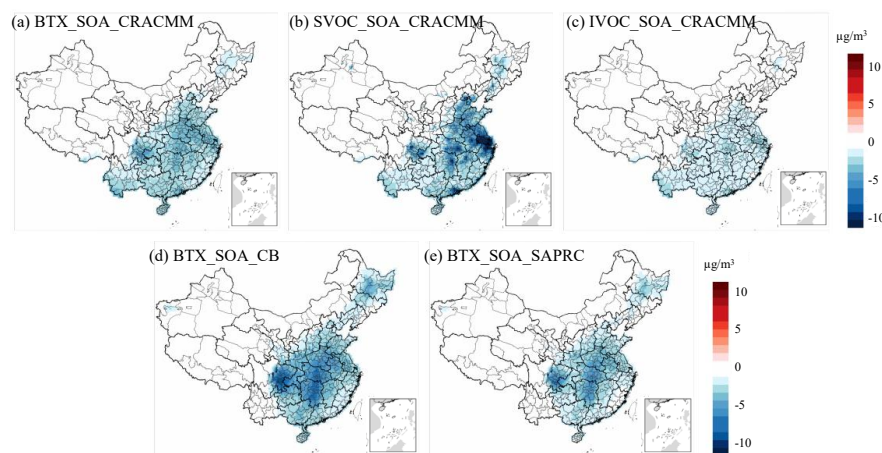
Figure 9 shows domain-wide differences in average $PM_{2.5}$ concentrations between the base CRACMM simulation and a series of zeroed emission simulations. A similar spatial pattern in $PM_{2.5}$ response was observed for zeroed biogenic and HC10 emissions (Figures S13a, c in percentage and Figures 9a, c in concentration), with CRACMM predicting a modest 1% change in $PM_{2.5}$ and less than $3 \mu g/m^3$ in many parts of China. This can be attributed to the low SOA yield by mass (0.09 g/g) for HC10 compounds (Pye et al., 2023), and the generally low winter emissions of biogenic ROC, as shown in previous studies (Pye et al., 2023). Zeroing BTX emissions resulted in average $PM_{2.5}$ concentration changes of -20% to 0% (Figure S13b) and -10 to $0 \mu g/m^3$ (Figure 9b), particularly in the YRD and BTH regions, where BTX emissions are highest. The high $PM_{2.5}$ formation potential of BTX compounds is attributed to their overall emission abundance and high SOA yield by mass (~ 0.5 g/g). Moreover, ozone levels in urban areas with significant BTX emissions also decrease in the BTX zero-out scenario (Figure S14). This effect is particularly evident in the PRD region, where the reduction reaches up to $10 \mu g/m^3$. This aligns with the findings of Place et al. (2023). One factor is the removal of BTX emissions, which serve as precursors for SOA formation. The second factor is that zeroing BTX emissions leads to a decrease in O_3 , which weakens atmospheric oxidizing capacity and reduces SOA



530 formation. However, it is important to note that Place's study was conducted in summer.



531
532 **Figure 9.** Changes in $PM_{2.5}$ concentrations between each zero-out scenario and its
533 corresponding base simulation: (a) biogenic ROC emission, (b) BTX emission, (c) HC10
534 emission, (d) SVOC emission, and (e) IVOC emission with CRACMM.



535
536 **Figure 10.** Changes in SOA concentrations between each zero-out scenario and its
537 corresponding base simulation: (a) BTX emissions, (b) SVOC emissions, (c) IVOC emissions
538 with CRACMM, (d) BTX emissions with CB6r3_ae7, (e) zeroed BTX emissions with
539 Saprc07tic_ae6.

540 As shown in Figures 10a, d, and e, the impact of zeroing BTX species on SOA formation for
541 the three mechanisms accounts for only approximately 50% of the total $PM_{2.5}$ change observed



542 in Figures 9b, S16a and c. This suggests that the changes in $PM_{2.5}$ concentrations resulting from
543 the removal of BTX emissions are not solely due to SOA formation but may also involve other
544 pathways or chemical processes. In contrast, the influence of S/IVOC species on $PM_{2.5}$
545 concentrations in the CRACMM mechanism is primarily driven by SOA formation. This
546 conclusion is supported by the spatial distribution and concentration differences shown in
547 Figures 9d, e, and Figures 10b, c, which exhibit nearly identical patterns. These similarities
548 indicate that the effects of S/IVOC emissions on $PM_{2.5}$ are mainly driven by SOA production.
549 The largest $PM_{2.5}$ response was observed when emissions from SVOC sources were excluded
550 from the simulation (Figure S13d), primarily because SVOCs have the highest yield, exceeding
551 1.0 g/g. The percentage changes in $PM_{2.5}$ range from -40% to 0%, with a concentration
552 reduction of more than $-10 \mu\text{g}/\text{m}^3$ (Figure 9d), particularly in the YRD region, where SVOC
553 emissions are substantial. For IVOCs, the reduction is about -5% across much of China, except
554 in the western regions where IVOC emissions are very low (Figure S13e), resulting in a
555 reduction of less than $-5 \mu\text{g}/\text{m}^3$ (Figure 9e).

556 A similar $\Delta PM_{2.5}$ response in percentage (Figure S15) and concentration change (Figure S16)
557 was observed when biogenic and BTX emissions were zeroed in simulations using
558 CB6r3_ae7_ae7 and Saprc07tic_ae7i. However, the CRACMM simulation with zeroed
559 biogenic emissions (Figure 9a) showed a more pronounced and widespread decrease in $PM_{2.5}$
560 compared to both CB6r3_ae7_ae7 and Saprc07tic_ae7i. This difference can be attributed to the
561 inclusion of new S/IVOC species in MEGAN, which are not accounted for in the other two
562 mechanisms. Additionally, zeroing BTX emissions had a greater impact in CB6r3_ae7,
563 particularly in central China, compared to Saprc07tic_ae7i and CRACMM. A possible reason
564 for this is that CRACMM reduces the number of lumped species in BTX and enhances the
565 representation of aromatic IVOC species, such as single-ring aromatics $\log_{10}(Ci^*) \approx 5$
566 (ROCP5ARO) and $\log_{10}(Ci^*) \approx 6$ (ROCP6ARO). These species are included in CB6r3_ae7
567 under categories like m-xylene and other more reactive aromatics (XYM), as well as less
568 reactive aromatics (XYE). As a result, CRACMM incorporates fewer species in BTX emissions
569 compared to CB6r3_ae7.

570 3.5 Uncertainty Analysis and limitations

571 3.5.1 Limitations of VOC Speciation Mapping



572 In the speciation mapping process, explicit species were directly mapped across the CRACMM,
573 CB6r3_ae7, and Saprc07tic_ae7i mechanisms, as these species are explicitly represented and
574 therefore allow for one-to-one mapping. Consequently, their spatial distributions were assumed
575 to be identical across the three mechanisms, and uncertainties associated with their emission
576 estimates were considered negligible. In contrast, lumped species present greater complexity
577 due to differences in the VOC species included within each lumped mechanism species. Given
578 that a direct one-to-one mapping between lumped species is not feasible, the mapping was
579 performed by matching the dominant lumped VOC species across mechanisms based on their
580 relative emission magnitudes. Overall, CRACMM incorporates a more comprehensive set of
581 VOC species than either CB6r3_ae7 or Saprc07tic_ae7i. The total mapped emissions associated
582 with each mechanism-specific inventory are summarized in Table S2. Nevertheless,
583 uncertainties remain due to regional differences in emission profiles. In particular, the total
584 emissions and source sector distributions of VOC species in Chinese emission inventories may
585 differ from those represented in the NEI. Such discrepancies introduce additional uncertainty
586 into the speciation mapping process.

587 **3.5.2 Uncertainty in Mapping L/S/IVOC Emissions**

588 For L/S/IVOCs, the 2D-VBS framework was applied to aggregate predicted products into a
589 reduced set of 15 representative CRACMM mechanism species. These representative species
590 span a wide range of saturation concentrations (Ci^*) from 10^{-2} to $10^6 \mu\text{g m}^{-3}$ and O:C ratios (nO :
591 nC) from 0.1 to 0.8. The original 2D-VBS inventory is structured along a two-dimensional grid
592 defined by $\log_{10}(Ci^*)$ and O:C ratio. During the mapping process to CRACMM, if a species in
593 the 2D-VBS inventory could not be directly assigned to a CRACMM species, it was reassigned
594 to the most proximate species in $\log_{10}(Ci^*)$ and/or O:C ratio. This reassignment process, while
595 necessary, inevitably introduces additional uncertainty. The L/S/IVOC emission inventory used
596 in this study is based on the work of Chang et al (Chang et al., 2022). For uncertainties
597 associated with this inventory, readers are referred to the analysis by Chang et al., which
598 discusses uncertainties related to emission factors, species classification, and spatial
599 distribution.

600 **4. Conclusions**



601 This study introduces the newly mapped VOC and POA inventories (both traditional and full
602 volatile) for CRACMM and presents the first comprehensive evaluation of $PM_{2.5}$ predictions
603 using the newly developed CRACMM chemical mechanism. The performance of CRACMM
604 with CB6r3_ae7, and Saprc07tic_ae7i are compared, and results demonstrate that CMAQ with
605 CRACMM provides reliable predictions of $PM_{2.5}$ and its components across China during the
606 months of January, April, July, and October 2021, although there are discrepancies in some
607 complex regions.

608 In conclusion, the comparison of the three chemical mechanisms using the traditional POA
609 inventory reveals that differences in R and MB values are generally small. However, with the
610 replacement of the full volatile inventory, CRACMM tends to predict lower $PM_{2.5}$
611 concentrations in January across most regions of China except PRD and YRD. In the other
612 months, CRACMM predicts higher concentrations than CB6r3_ae7 and Saprc07tic_ae7i when
613 the full volatile inventory is incorporated. The differences in $PM_{2.5}$ concentrations in January,
614 are primarily attributed to lower POA concentrations, which are influenced by semi-volatile
615 partitioning and reduced aging of semi-volatile POA species under lower temperatures. In
616 contrast, CRACMM simulates elevated SOA concentrations in eastern China due to enhanced
617 precursor availability, while reduced SOA formation is observed in the SCB, where winter
618 conditions slow the oxidation of precursors. The inclusion of the full volatile inventory in
619 CRACMM results in higher SOA concentrations in October, driven by increased precursor
620 availability. Overall, CRACMM demonstrates improved performance in terms of R and MB,
621 particularly in January and October for the PRD region, but performs less well in April and
622 July, particularly in the BTH region, compared to CB6r3_ae7. Additionally, CRACMM with
623 the full volatile inventory increase in simulated $PM_{2.5}$ concentrations, resulting in smaller
624 deviations from observation across many regions, highlighting the importance of including
625 S/IVOC emissions in the chemical mechanism. Emission perturbation simulations using
626 CMAQ further emphasize the significant role of various emission species, particularly BTX
627 and SVOC, in driving $PM_{2.5}$ formation. The SOA contribution from BTX emissions accounts
628 for nearly 50% of the $PM_{2.5}$ changes, while S/IVOC emissions primarily influence $PM_{2.5}$
629 through SOA formation. BTX emissions had a more significant impact in CB6r3_ae7,
630 particularly in central China, partly due to the fewer VOC species included in the lumped BTX



631 of the CRACMM mechanism. Future assessments of O₃ predictions with CRACMM will offer
632 additional constraints on the gas and aerosol chemistry that contributes to PM_{2.5} formation.

633 **Data availability.** The model simulation is based on the CMAQ v5.4 developed by the U.S.
634 EPA, and the code is publicly available at <https://doi.org/10.5281/zenodo.7218076>. All input
635 data to reproduce the results and figures in this paper, has been archived on Zenodo
636 (<https://doi.org/10.5281/zenodo.16791307>. Su et al., 2025) and is freely accessible.

637 **Supplement.** The supplement related to this article is available online at

638 **Author contributions.** LL and YJW designed the study. QFS conducted the modeling work,
639 formal analysis and drafted the original manuscript. YFC contributed to data curation and
640 emissions preparation. DCW contributed to model revision, manuscript editing and reviewing.
641 HOTP, GS, BM, BP and YJW provided critical review and editing of the manuscript. LH and
642 YJW contributed to modelling setup. LL supervised the project and secured funding.

643 **Competing interests.** The authors declare no conflicts of interest.

644 **Disclaimer:** The views expressed in this article are those of the authors and do not necessarily
645 represent the views or the policies of the U.S. Environmental Protection Agency (EPA).

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