

# Application and Evaluation of CRACMM V1.0 Mechanism in PM<sub>2.5</sub> Simulation Over China

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## Abstract

Chemical mechanisms are one of the major sources of bias in chemical transport model simulations, making their improvement a critical step towards enhancing model performance and supporting air quality management and research. In this study, a newly developed chemical mechanism, the Community Regional Atmospheric Chemistry Multiphase Mechanism (CRACMM), integrated into the Community Multiscale Air Quality (CMAQ) modeling system, was evaluated through comparison with two traditional chemical mechanisms, Carbon Bond 6 version r3 with aero7 treatment of SOA (CB6r3\_ae7) and State Air Pollution Research Center version 07tc with extended isoprene chemistry and aero7i treatment of SOA (Saprc07tic\_ae7i), for China. Sensitivity simulations related to precursor reactive organic carbon (ROC) emissions were conducted to investigate the key driving factors of PM<sub>2.5</sub> formation. The results indicate that, when using the traditional primary organic aerosol (POA) inventory, the differences among the three chemical mechanisms are within 0 - 0.14 for the R, 0 - 10 μg/m<sup>3</sup> for the MB, and within 10% for the NMB values. However, when the full-volatility emission inventory is applied in January, CRACMM exhibits improved performance in the Pearl River Delta (PRD) region. The MB is reduced by 3.0-7.8 μg/m<sup>3</sup>. In addition, the NMB decreases by 17–23%, and the root mean square error (RMSE) is reduced by 1- 6 μg/m<sup>3</sup> compared with simulations using the traditional POA inventory across the four months. CRACMM predicts higher PM<sub>2.5</sub>

34 concentrations during spring, summer and autumn, mainly due to enhanced secondary organic  
35 aerosol (SOA) formation driven by increased precursor emissions. Benzene–toluene–xylene  
36 (BTX) species and semi-volatile organic compound (SVOC) emissions significantly  
37 contributed to PM<sub>2.5</sub> formation in CRACMM. The SOA from BTX emissions accounts for  
38 nearly 50% of the PM<sub>2.5</sub> changes, while intermediate-volatility organic compounds (IVOC) and  
39 SVOCs emissions mainly affect PM<sub>2.5</sub> concentrations through SOA formation. These results  
40 indicate that CRACMM, when using the full-volatility inventory, can effectively compensate  
41 for the underestimation of PM<sub>2.5</sub> mass that may occur with traditional POA treatment,  
42 particularly in regions with high photochemical activity and abundant S/IVOC precursors.

### 43 **1. Introduction**

44 Exposure to airborne PM<sub>2.5</sub> is associated with a variety of harmful health effects (Liu et al.,  
45 2024; Tsai et al., 2025; Kim et al., 2015) and was reported to cause 4.14 million deaths  
46 worldwide annually (95% confidence interval: 3.45 to 4.80) (Murray et al., 2020), underscoring  
47 the highly need for effective mitigation of PM<sub>2.5</sub> pollution. Thus, a better understanding of the  
48 PM<sub>2.5</sub> formation mechanism is essential for formulating effective air pollution control strategies.  
49 Chemical transport models such as the Community Multiscale Air Quality (CMAQ) model  
50 (Byun and Schere, 2006) have been widely applied in China to investigate air quality issues,  
51 including seasonal PM<sub>2.5</sub> and ozone distributions and their formation mechanisms. The CMAQ  
52 version 5.02 and 5.3.2, both versions are equipped with the gas-phase mechanism of State Air  
53 Pollution Research Center version99 (Saprc99) and Carbon Bond 6 version (CB6). These two  
54 chemical mechanisms were employed to simulate air quality over China during 2013–2019  
55 (Mao et al., 2022). In that study, PM<sub>2.5</sub> were examined in the North China Plain (NCP), Yangtze  
56 River Delta (YRD), and Pearl River Delta (PRD), while it was overestimated in the Chengyu  
57 Basin (CY) and Fen-Wei Plain (FWP) regions with NMB and NME values greatly exceeded  
58 the suggested criteria. In addition, the FWP and CY also showed lower R than the benchmark  
59 (R<0.4) suggested by Emery et al during heavy polluted episode (Emery et al., 2017). During  
60 2014–2017, the Comprehensive Air Quality Model with Extension (CAMx) (Yarwood et al.,  
61 2007) version 6.2, version 7.1, CMAQv5.0.2, and CMAQv5.3.2 models performed quite well  
62 in PM<sub>2.5</sub> mean performance. The bias and error terms of the four models were resemblant small:

63 -0.29, -0.07, -0.04, and -0.11 for NMB and 0.51, 0.48, 0.53, and 0.52 for NME, 0.58, 0.55,  
64 0.60, and 0.39, for R respectively (Meng et al., 2026), the CAMx6.2, CAMx7.1, and  
65 CMAQv5.0.2 simulations covered the entire country, whereas CMAQv5.3.2 was evaluated  
66 over eastern China. (Meng et al., 2026). According to Kang et al's research, the model shows  
67 good performance for PM<sub>2.5</sub> in most areas, except for the PRD region, where the mean fractional  
68 biases (MFB) for PM<sub>2.5</sub> using Saprc serious mechanisms are slightly outside the recommended  
69 range (Boylan and Russell, 2006). According to Huang's research, evaluation results of PM<sub>2.5</sub>  
70 simulation by SOAP3 in CAMx in BTH, YRD, PRD, FWP, SCB and PRD areas, the R values  
71 are 0.31, 0.45, 0.38, 0.48, and 0.15 in July, respectively. In November, the R values are also low  
72 for PRD and FWP, 0.30 and 0.48, respectively (Huang et al., 2024).

73 Existing studies indicate that different chemical mechanisms vary significantly in terms of the  
74 number of species, reaction complexity, and the coverage of chemical pathways, could lead to  
75 noticeable differences in regional air quality model performance. With this realization, we  
76 further examined the limitations of the commonly used chemical mechanisms (CB, Saprc and  
77 Regional Atmospheric Chemical Mechanism (RACM)). In CB mechanism, model species  
78 represent the concentrations of constituent groups regardless of the molecule to which they are  
79 attached. Initially, this approach conserved carbon atoms, required relatively few species, and  
80 both led to lower computational cost. However, as the mechanism evolved, its grouping  
81 increasingly resembled aggregated molecule schemes, since both molecular structure and total  
82 molecular weight significantly influence atmospheric chemistry (Yarwood et al., 2005). The  
83 disadvantage of CB mechanism is that chemical expression of free radicals is insufficient (Kang  
84 et al., 2016; Sarwar et al., 2008). Comparative analysis of different CB6 mechanism variants  
85 shows that differences in reaction pathways can lead to significant deviations in model  
86 predictions of ozone, NO<sub>x</sub>, and formaldehyde (Cao et al., 2021).

87 The Saprc series (Saprc -90, Saprc -99, Saprc -07) aggregate VOCs by molecule or functional  
88 group, representing roughly 400 categories. Condensed versions are widely used in urban and  
89 regional air quality models, but simplifications can limit the accuracy of organic chemistry  
90 representation (Stockwell et al., 2012). Comparative analyses of different Saprc variants reveal  
91 discrepancies in predicting ozone, radical species, and oxidative products, indicating  
92 uncertainties arising from mechanism and emission inventories (Kang et al., 2025). For RACM,

93 it was developed for broader applications, where NO<sub>x</sub> is lower and slower-reacting organics  
94 are more important. RACM version 2 includes 118 species and 356 reactions. Even though this  
95 mechanism is computationally efficient, however, it may inadequately capture detailed organic  
96 chemistry and secondary organic aerosol (SOA) formation, resulting with potential  
97 uncertainties in air quality predictions (Stockwell et al., 2012).

98 In air quality models, gas-phase chemical mechanisms are typically coupled with aerosol  
99 modules to simulate interactions between the gas and particulate phases. Aerosol Module 7  
100 (aero7) is the latest aerosol representation within the CMAQ model (Byun and Schere, 2006),  
101 developed by the U.S. Environmental Protection Agency (EPA). Aero7 improves consistency  
102 in representing SOA formation pathways between the CB- and SAPRC-based chemical  
103 mechanisms. It also updates monoterpene SOA yields from photooxidation, adds uptake of  
104 water onto hydrophilic organics, and includes consumption of inorganic sulfates (SO<sub>4</sub><sup>2-</sup>) when  
105 isoprene epoxydiol (IEPOX) organosulfates are formed (Pye et al., 2013). Furthermore, it  
106 enhances computational efficiency by using a volatility basis set (VBS) to parameterize SOA  
107 yields rather than using the Odum 2-product fit (Zhang et al., 2021a).

108 Most of the above chemical mechanisms and SOA treatment exhibit substantial limitations in  
109 simulating SOA formation (Wang et al., 2019; Zhang et al., 2021; Zhang et al., 2022; Zhang et  
110 al., 2023; Huang et al., 2024). First, Semi-Volatile Organic Compounds (SVOC) and  
111 Intermediate-Volatility Organic Compounds (IVOC) species are not explicitly represented,  
112 leading to an underestimation of SOA contribution from those precursors. Even in 2D-VBS  
113 mechanisms, which includes S/IVOC species, the gas-to-particle conversion of oxidative  
114 products are mainly characterized using empirical parameters (e.g., yields and volatility  
115 distributions). Moreover, the 2D-VBS framework does not explicitly track individual chemical  
116 reactions; instead, it relies on parameterizations derived from environmental chamber  
117 experiments or model calibration data, making it difficult to resolve specific chemical pathways  
118 of SOA formation (Chang et al., 2022). Second, the use of fixed-yield empirical  
119 parameterization limits the representation of multigenerational oxidative processes and gas-  
120 particle partitioning dynamics (Chang et al., 2022). Third, POA is commonly assumed to be  
121 non-volatile and non-reactive, thereby neglecting its potential for re-evaporation and  
122 subsequent oxidation to form SOA. These simplifications overly idealize SOA formation

123 processes missing various key reactions and resulted in an inaccurate representation of the  
124 complex aging of organic aerosols in the atmosphere (Huang et al., 2024).

125 The Community Regional Atmospheric Chemistry Multiphase Mechanism (CRACMM) is the  
126 latest chemical mechanism developed under the leadership of scientists at the US EPA.  
127 CRACMM is built upon version 2 of RACM (Goliff et al., 2013) and incorporates state-of-the-  
128 science developments, including autoxidation, aromatic chemistry, oxygenated hydrocarbons,  
129 organic nitrates, and halogen chemistry. These advances enhance the representation of  
130 atmospheric chemical transformations, enabling more realistic simulations of key air pollutants  
131 such as O<sub>3</sub>, PM<sub>2.5</sub>, and various hazardous species, e.g. formaldehyde (Skipper et al., 2024). In  
132 addition, CRACMM integrates a full-volatility organic framework and explicitly accounts for  
133 multigenerational oxidative processes, thereby improving the physicochemical representation  
134 of secondary organic aerosol (SOA) formation and providing a more comprehensive  
135 description of SOA evolution (Ng et al., 2008).

136 CRACMM also provides detailed species mapping methodology between emission inventory  
137 and chemical mechanism (Pye et al., 2023), to ensure carbon conservation when tracking the  
138 transformation of carbon from emission sources to products (Ng et al., 2008). CB6r3\_ae7 and  
139 Saprc07tic\_ae7 do not consider certain SOA precursors such as L/S/IVOC (Chang et al., 2022),  
140 and CRACMM explicitly accounts for SOA precursors beyond traditional non-oxygenated  
141 volatile hydrocarbons, including phenolic compounds, furans, and other oxygenated organic  
142 species (Pye et al., 2023). This makes CRACMM describing and simulating SOA in more  
143 precise and accurate manner. No doubt CRACMM a good chemical mechanism candidate to  
144 address the above listed shortcomings. CRACMMv1.0 was applied in the simulation over  
145 northeast U.S. in summer reported in the work by Place et al (2023), and the result showed  
146 ozone simulated values were better than RACM2\_ae6 chemical mechanism in terms of  
147 comparing with observation: average bias of RACM2\_ae6 is +4.2ppb, average bias of  
148 CRACMMv1.0 +2.1 ppb. A few studies have been conducted using CRACMM but they all  
149 focused on US CONUS domain (Place et al., 2023). CRACMM has been not applied on China  
150 domain so thorough evaluation is a must.

151 For computational cost considerations, the number of species and reactions directly affects  
152 model runtime. Mechanisms with larger numbers of species and reactions, such as

153 Saprc07tic\_ae7i, which contains the largest number of reactions and species among the standard  
154 CMAQ mechanisms, is the most computational expensive mechanism compared to simpler  
155 mechanisms like CB6r3\_ae7. In contrast, CB has fewer reactions and species. The primary  
156 difference between CB6r3\_ae7 Saprc07tic\_ae7i, and CRACMM lies in the chemistry module.  
157 Our tests indicate that CRACMM requires approximately 30–40% more computational time  
158 than CB6r3\_ae7, but 20–30% less than Saprc.

159 With advances in high-performance computing, CRACMM could also be applied in a global  
160 scale. Particularly, the Model for Prediction Across Scales (MPAS) global meteorological  
161 model has recently been successfully coupled with CMAQ, demonstrating the application of  
162 CRACMM in the global MPAS-CMAQ coupled model framework (Wong et al., 2024). Also,  
163 CRACMM is not restricted to CMAQ. The design of CRACMM follows a modular framework  
164 for gas-phase chemistry and SOA formation, which does not rely on model-specific  
165 assumptions. Therefore, in principle, CRACMM could be implemented in other regional  
166 models, although additional effort would be required to adapt emission mapping and aerosol–  
167 chemistry coupling.

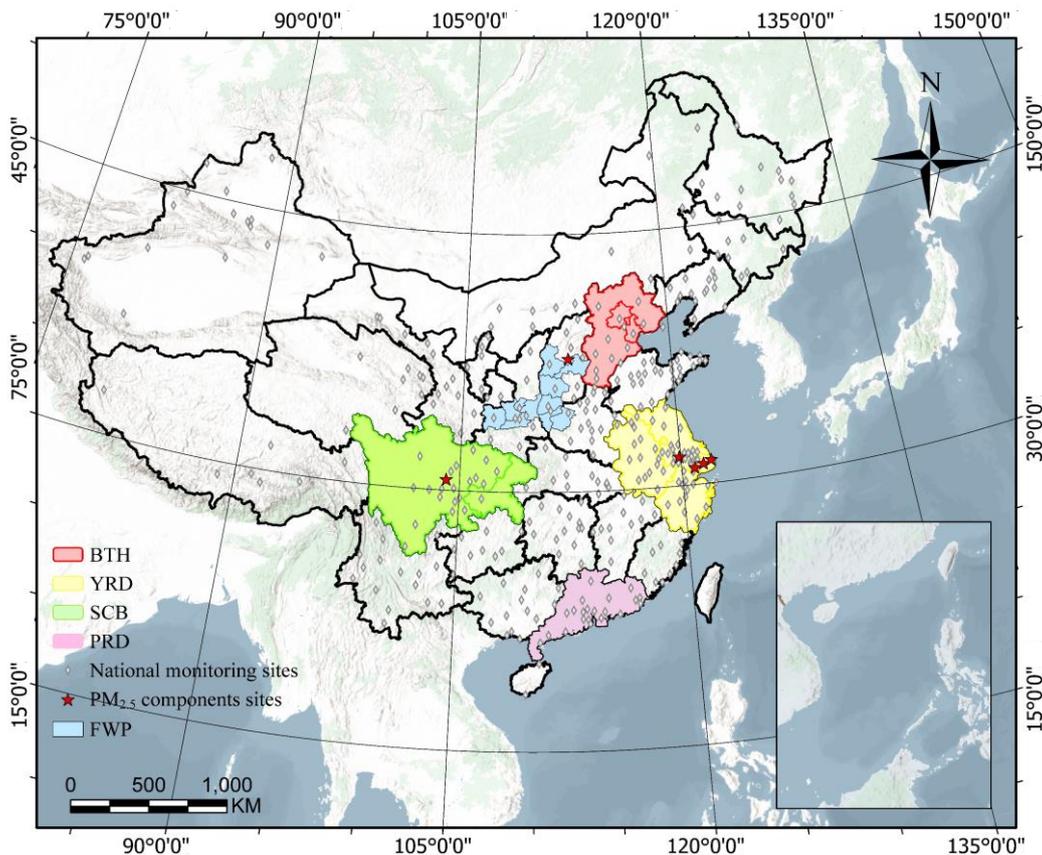
168 In this study, PM<sub>2.5</sub> predictions from the CRACMM mechanism were evaluated with surface  
169 observations comprehensively, covering different seasons and regions. Results derived by  
170 CRACMM are compared with two well-established chemical mechanisms, Saprc07tic\_ae7i  
171 and CB6r3\_ae7. The differences in PM<sub>2.5</sub> and SOA drivers between CRACMM and the two  
172 existing mechanisms are further explored. The results of this study provide a solid foundation  
173 for the further application of CRACMM in understanding and regulating air pollution in China  
174 and globally.

## 175 **2. Methodology**

### 176 **2.1 Model configuration**

177 Model simulations were conducted using CMAQ v5.4 with a horizontal resolution of 36 km ×  
178 36 km, covering mainland China (Figure 1). This domain includes five key city clusters with  
179 notable air pollution levels (Deng et al., 2022; Huang et al., 2024): Beijing-Tianjin-Hebei (BTH),  
180 Yangtze River Delta (YRD), Pearl River Delta (PRD), Fen-Wei Plain (FWP), and Sichuan  
181 Basin (SCB). Simulations were carried out for the months of January, April, July, and October

182 2021, representing winter, spring, summer, and autumn, respectively. The model includes 34  
183 vertical layers, with the first layer located approximately 35 meters above the ground. Each  
184 simulation was initialized with a 15-day spin-up period before the start of each month. In  
185 addition to CRACMM version 1.0, two other chemical mechanisms, CB6r3\_ae7 and  
186 Saprc07tic\_ae7i, were included for comparisons, the number of reactions and gas- and particle-  
187 phase species in three different chemical mechanisms used in CMAQ are shown in Figure S1.  
188 Simple flowcharts illustrating the different species represented in each mechanism and their  
189 sources and sinks are shown in Figure S2 for CRACMM and Figure S3 for CB6r3\_ae7. For  
190 Saprc07tic\_ae7i, it has been described in detail in a previous studies (Pye et al., 2015b). All three  
191 mechanisms are available in CMAQv5.4, with the "m3dry" deposition scheme selected. The  
192 initial and boundary conditions for CRACMM and Saprc07tic\_ae7i were mapped from the  
193 seasonal average hemispheric CMAQ output files distributed through the CMAS. CMAQ ready  
194 meteorological input files were created by the Meteorology-Chemistry Interface Processor  
195 (MCIP) (Otte and Pleim, 2010) version 5.4 processing through output files from an offline run  
196 of the Weather Research and Forecasting (WRF) model version 4.0 (Skamarock et al., 2019).  
197 WRF configuration was detailed in our previous studies (Huang et al., 2021a). The archived  
198 dataset, including the concentrations and model performance statistics of PM<sub>2.5</sub> and its  
199 components, model configurations, and the locations of all observation sites, is available on  
200 Zenodo (Su and Chen, 2025)



201

202 **Figure 1.** Model domain with five key city clusters (outlined in color), locations of national  
 203 monitoring sites (grey diamonds), and six PM<sub>2.5</sub> chemical components observation sites (red  
 204 stars).

## 205 **2.2 Emissions**

### 206 **2.2.1 Traditional Emissions Inventory**

207 The 2019 anthropogenic Multi-resolution Emission Inventory for China (MEIC), developed by  
 208 Tsinghua University, was utilized in this study (<http://www.meicmodel.org>). Its spatial  
 209 resolution is  $0.25^\circ \times 0.25^\circ$  and includes five sectors: power, industry, residential,  
 210 transportation, and agriculture, and the provincial VOC emissions in 2019 from MEIC,  
 211 categorized by sector for January, April, July, October are shown in Tables S1-S4. Biogenic  
 212 emissions were estimated using the Model of Emissions of Gases and Aerosols from Nature  
 213 version 3.2 (MEGANv3.2) (Guenther et al., 2012). Currently, MEIC supports VOC emission  
 214 only for both CB6r3\_ae7 and Saprc07tic\_ae7i, but not the CRACMM mechanism. Therefore,  
 215 anthropogenic VOC species were converted to CRACMM input species using a binary decision  
 216 tree approach. This approach distinguishes between one-to-one and non-one-to-one mappings

217 based on chemical species correspondence from Saprc07tic\_ae7i and CB6r3\_ae7 to CRACMM.  
218 The one-to-one mappings are further classified into explicit one-to-one (routine A) and lumped  
219 one-to-one (routine B), while the non-one-to-one mappings include many-to-one (routine C)  
220 and many-to-many (routine D) cases. In routine A, both Saprc07tic\_ae7i and CB6r3\_ae7 consist  
221 of a few species that can be mapped directly to CRACMM based on CAS number, e.g., HCHO  
222 (formaldehyde) is mapped to HCHO. In routine B, the mapping is based on lumped species  
223 categories or names. For example, OLE1 (alkenes other than ethene, with  $k_{OH} < 7 \times 10^4 \text{ ppm}^{-1}$   
224  $\text{min}^{-1}$ ) in Saprc07tic\_ae7i is mapped to OLI (internal alkenes) in CRACMM, OLE2 (alkenes  
225 with  $k_{OH} > 7 \times 10^4 \text{ ppm}^{-1} \text{ min}^{-1}$ ) is mapped to OLT (terminal alkenes), and ONIT is mapped to  
226 RNO3 based on the same name used for organic nitrates.

227 In routine C, new species could be added to the CRACMM mechanism, such as CSL (Cresols)  
228 and PHEN (Phenol and aromatic diols), which correspond to CRES (phenols and cresols) in  
229 the Saprc07tic\_ae7i mechanism. In this case, we use the emission factor ratio to distribute the  
230 species. As MEIC does not provide species-level emission factors, data from the 2017 U.S.  
231 National Emission Inventory (NEI) (Pye et al., 2023b) were utilized, which contain over 3,000  
232 species with corresponding emission factors, source sectors, and CRACMM species mappings.  
233 The emission factors are averaged over all sources for different MEIC sectors (mobile sources,  
234 industrial sources, etc.). Another example is that both XYE (P-xylene and less reactive  
235 aromatics) and XYM (M-xylene and more reactive aromatics) are newly introduced species in  
236 the CRACMM mechanism, corresponding to XYL (Xylene and other aromatics) in the  
237 Saprc07tic\_ae7i mechanism. According to the 2017 NEI, their ratio is 0.3:0.7. Therefore, 0.3  
238 of XYL in Saprc07tic\_ae7i is assigned to XYE in CRACMM, and 0.7 of XYL in  
239 Saprc07tic\_ae7i is assigned to XYM in CRACMM.

240 Routine D is more complicated, but the mapping is still based on the emission factor ratio for  
241 proper mapping. For instance, in CRACMM, GLY represents both glyoxal and glycolaldehyde.  
242 To construct this species from Saprc07tic\_ae7i, GLY (representing glyoxal only) and part of  
243 CCHO (glycolaldehyde and acetaldehyde) are mapped, such that GLY in CRACMM  
244 corresponds to  $\text{GLY} + \text{CCHO} \times 0.25$  in Saprc07tic\_ae7i. Similarly, ACD (acetaldehyde) in  
245 CRACMM corresponds to  $\text{CCHO} \times 0.75$  in Saprc07tic\_ae7i. For species lumped from multiple  
246 species, only those with larger emission factors are considered. Table S5 outlines the

247 correspondence relationships for major species with substantial emissions. Since it is  
248 challenging to compare VOC emissions in different mechanisms due to the lumping rules, we  
249 only conducted an overall comparison of total emissions, as shown in Table S6.

### 250 **2.2.2 POA Emissions**

251 Two POA inventories were employed in this study: a traditional POA emissions inventory and  
252 a full-volatility inventory. CRACMM, Saprc07tic\_ae7i, and CB6r3\_ae7 all use the same dataset  
253 for the traditional POA emissions inventory. This inventory applies a VBS profile based on  
254 Woody et al. (2016) and Robinson et al. (2007), treating POA as semi-volatile with  $C_i^*$  values  
255 ranging from  $10^{-2}$  to  $10^3$   $\mu\text{g}/\text{m}^3$ . The detailed species of POA included in each mechanism are  
256 listed in Table S7.

257 In contrast, the full-volatility inventory distributes POA emissions across a wider range of  
258 volatility bins. Laboratory experiments have demonstrated that L/S/IVOC emissions, which are  
259 largely absent in the traditional POA inventory, contribute to SOA formation much more  
260 efficiently than VOCs, owing to their lower volatility. To capture these processes, the full-  
261 volatility inventory developed by Chang et al. (2022) was used. In this study, all species mapped  
262 from the full-volatility inventory (implemented using the two-dimensional VBS (2D-VBS)  
263 framework) to CRACMM are classified into two lumped categories: (1) an alkane-like ROC  
264 group, (2) 15 CRACMM mechanism species representing oxygenated S/IVOCs. Since all  
265 species exist in both the particle and gas phases, the same mapping rules are applied. The  
266 mapping rules for the gas phase are summarized in Table S8.

267 For eight new alkane-like ROC species with high OA formation potential spanning the  
268 L/S/IVOC range and are grouped by  $\log_{10}(C_i^*)$  into ROCN1ALK, ROCP0ALK, VROCP1ALK,  
269 ROCP2ALK, ROCP3ALK, ROCP4ALK, ROCP5ALK, and ROCP6ALK. They are mapping  
270 from CSM1O2C00P, CS00O2C00P, CS01O2C00P, CS02O2C00P, CS03O2C00P,  
271 CS04O2C00P, CS05O2C00P and CS06O2C00P species in full-volatility emission inventory  
272 used in the 2D-VBS mechanism where numbers after CS indicate the negative(M) or positive  
273 (0)  $\log_{10}(C_i^*[\mu\text{g}/\text{m}^3])$  value and the number after 2C means  $10 \times nO : nC$  (e.g., CS05O2C00P is  
274  $C_i^* = 10^{-5}$   $\mu\text{g}/\text{m}^3$  with  $nO : nC = 0$ ).

275 For oxygenated L/S/IVOC, the species in 2D-VBS mechanism were lumped into 15 CRACMM  
276 mechanism species, spanning  $C_i^*$  values of  $10^{-2}$  to  $10^6$   $\mu\text{g}/\text{m}^3$  and  $nO : nC$  of 0.1 to 0.8:

277 ROCN2OXY2, ROCN2OXY4, ROCN2OXY8, ROCN1OXY1, ROCN1OXY3,  
278 ROCN1OXY6, ROCP0OXY2, ROCP0OXY4, ROCP1OXY1, ROCP1OXY3, ROCP2OXY2,  
279 ROCP3OXY2, ROCP4OXY2, ROCP5OXY1, and ROCP6XY1. 2D-VBS products of known  
280  $nC$  and  $nO$  were mapped to the available CRACMM model species, first by interpolating to the  
281 two nearest species in  $nO$ :  $nC$  space, and then to the two nearest species  $\log_{10}(C_i^*)$  points.  
282 Since neither Sapr07tic\_ae7i nor CB6r3\_ae7 includes a representation of full-volatility POA,  
283 only CRACMM can utilize this comprehensive inventory. The methodology outlined by Chang  
284 et al. (2022) includes emissions from various sources, along with their corresponding profiles,  
285 volatility ranges, and emission amount. The anthropogenic L/S/IVOC emission inventory using  
286 a volatility-binned approach with full coverage of both particle and gas phases was shown in  
287 Figure S4. For traditional POA inventory, the POA emission amount was 2840 kt/y, while the  
288 new full-volatility emission inventory includes emissions of Low-Volatility Organic  
289 Compounds (LVOC) (1,342 kt/y), SVOC (1,169 kt/y), and IVOC (3,939 kt/y), resulting in a  
290 total of 6,450 kt/y. The new inventory fills a gap of 3,610 kt/y in L/S/IVOC emissions that were  
291 absent from the traditional inventory. To thoroughly evaluate CRACMM and compare it with  
292 CB6r3\_ae7 and Sapr07tic\_ae7i, four simulation scenarios were designed, as shown in Table  
293 1.

294 In the CMAQ model, a potential combustion SOA (pcSOA) species is introduced to compensate  
295 for the fraction of SOA formed from the oxidation of combustion-related organic compounds  
296 that are not explicitly represented in the model (Murphy et al., 2017). Traditional chemical  
297 mechanisms often do not include SVOCs and IVOCs emitted from combustion sources, nor  
298 their associated oxidation pathways, which leads to a systematic underestimation of SOA levels  
299 in model simulations. To address this issue, CMAQ has incorporated the pcSOA species into  
300 the organic aerosol module since version 5.2, providing an empirical representation of this  
301 missing combustion-related SOA component (Murphy et al., 2017). This treatment is primarily  
302 intended for anthropogenic combustion sources, such as motor vehicle emissions, industrial  
303 combustion, and biomass burning. Because the CRACMM species framework explicitly  
304 accounts for multigenerational oxidation processes across different volatility ranges of VOCs,  
305 the empirical anthropogenic SOA source (pcSOA) implemented in the CB6r3\_ae7 and

306 Saprc07tic\_ae7i mechanisms was turned off in all simulations to avoid double counting with  
 307 the CRACMMv1.0 mechanism.

308 **Table 1.** Description of simulation scenarios and their emissions

Scenarios	Mechanisms	POA emission inventory	Anthropogenic + Biogenic emission inventory
1	CB6r3_ae7	Traditional POA inventory	MEIC+MEGAN
2	Saprc07tic_ae7i	Traditional POA inventory	MEIC+MEGAN
3	CRACMM	Traditional POA inventory	MEIC+MEGAN
4	CRACMM	Full-volatility inventory	MEIC+MEGAN

### 309 **2.3 Observational data and model performance evaluation**

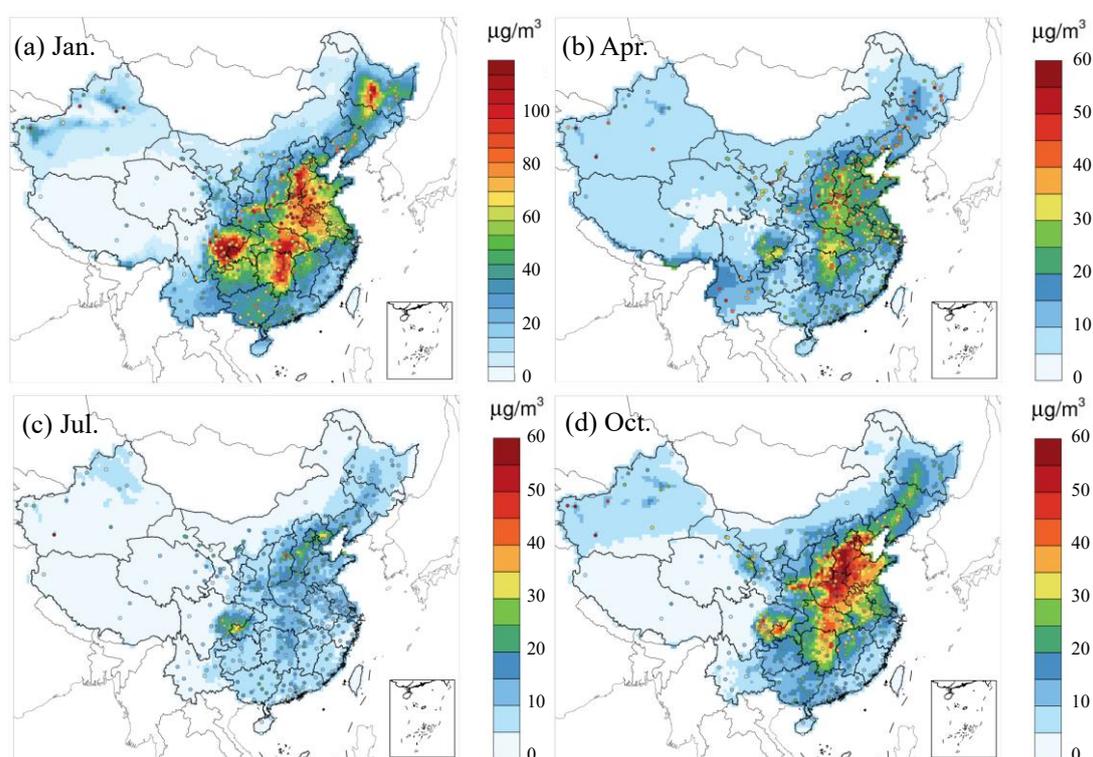
310 Hourly concentrations of PM<sub>2.5</sub> at national monitoring stations were obtained from the China  
 311 National Environmental Monitoring Centre (<http://air.cnemc.cn:18007>), which were then used  
 312 to evaluate model performance. Field observational data of PM<sub>2.5</sub> chemical components  
 313 including NO<sub>3</sub><sup>-</sup> (nitrate), SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup> (ammonium), OC (organic carbon), and EC (elemental  
 314 carbon) at six super monitoring station sites were collected, as detailed in Figure 1 and Table  
 315 S9. Missing observation periods were excluded from the analysis. Model performance was  
 316 assessed using well-established statistical metrics, including the R, MB, NMB, root mean  
 317 square error (RMSE), normalized mean error (NME), and index of agreement (IOA). The  
 318 formulas for each individual parameter are presented in Table S10. In these equations,  $\bar{C}_m$  and  
 319  $\bar{C}_o$  represent the mean modeled and observed concentrations over all samples, respectively;  
 320  $C_m$  and  $C_o$  denote the modeled and observed values for the  $i$ -th sample; and  $N$  is the total  
 321 number of valid samples. A combined analysis of these statistical indicators enables a  
 322 comprehensive assessment of model performance and reliability, providing a basis for further  
 323 model refinement and interpretation of the simulation results.

## 324 **3. Results and discussion**

### 325 **3.1 Overview of CMAQ-CRACMM model performance evaluation on PM<sub>2.5</sub>**

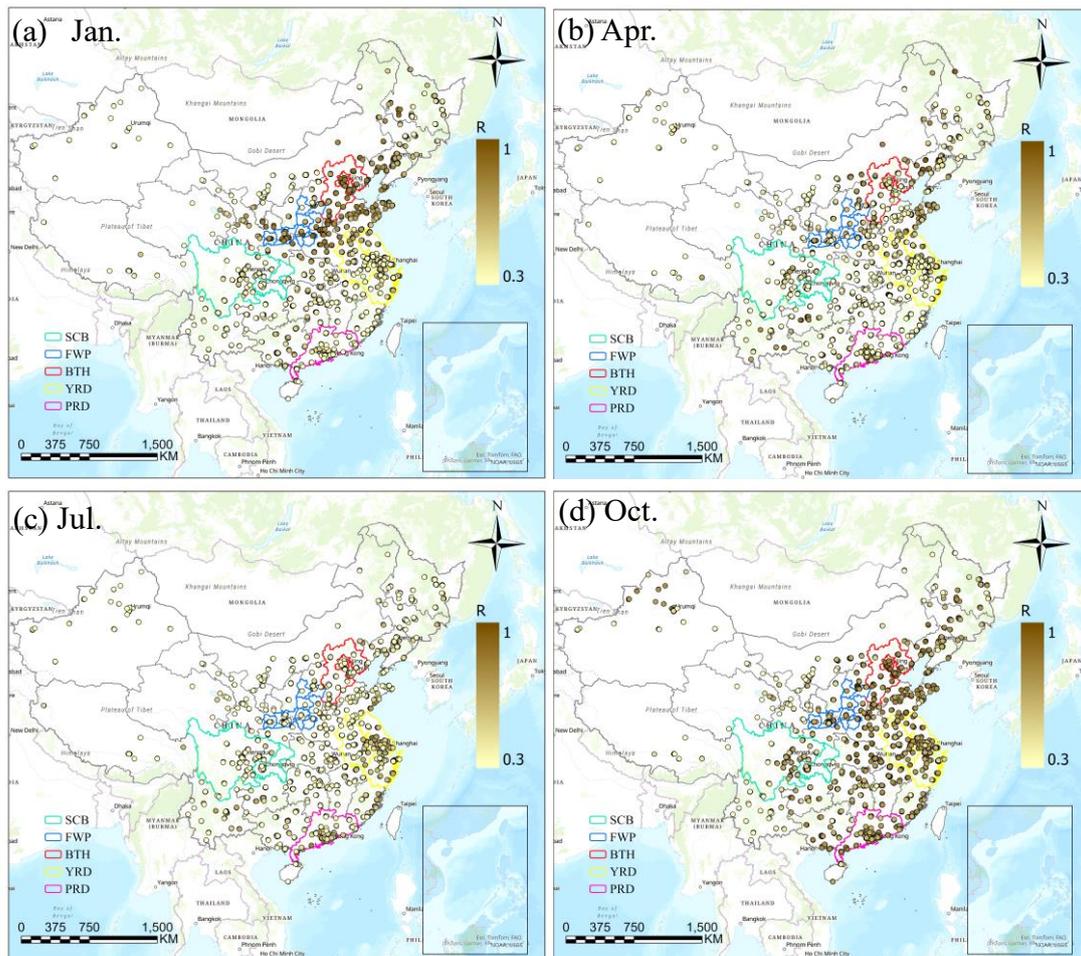
326 Figure 2 depicts the spatial distribution of observed (dots) and simulated PM<sub>2.5</sub> concentrations  
 327 for January, April, July, and October 2021, based on the CRACMM model with the full-  
 328 volatility inventory. In January (Figure 2a), PM<sub>2.5</sub> concentrations range from 5 to over 100  
 329 µg/m<sup>3</sup>, with the highest values concentrated in the NNCP and parts of the SCB. These elevated  
 330 levels are primarily driven by relatively higher anthropogenic emissions and stagnant

331 meteorological conditions typical of winter. In April (Figure 2b), concentrations have  
332 significantly decreased, ranging from 5 to 40  $\mu\text{g}/\text{m}^3$ , with the most notable reductions observed  
333 in northern regions. During July (Figure 2c),  $\text{PM}_{2.5}$  concentrations were at their lowest, typically  
334 ranging from 0 to 40  $\mu\text{g}/\text{m}^3$ . This decline is mainly attributable to the increased precipitation  
335 which washed out pollutants, in the southern and eastern parts of China compared to other  
336 months (Figure S5) In addition, higher planetary boundary layer (PBL) heights during the warm  
337 season in the south, particularly in the YRD region of China than other months (Figure S6),  
338 enhancing atmospheric mixing and dilution, and led to the decrease in  $\text{PM}_{2.5}$  concentrations. In  
339 October (Figure 2d),  $\text{PM}_{2.5}$  concentrations rise again, ranging from 5 to 60  $\mu\text{g}/\text{m}^3$ , with the  
340 highest concentrations observed in the NCP and along the eastern coastal regions, which is  
341 attributed to the heating in later autumn and unfavorable meteorological conditions. Overall,  
342 monthly variations in  $\text{PM}_{2.5}$  concentrations are primarily driven by meteorological conditions  
343 and the distribution of emission sources.



345 **Figure 2.** Monthly average  $\text{PM}_{2.5}$  concentrations predicted (raster) by CRACMM and observed  
346 (dots) in 2021 using the full-volatility emission inventory. Note that the color scale for panel  
347 (a) differs from panels (b–d) to highlight variations in the data.

348 The performance of the CMAQ model in simulating hourly PM<sub>2.5</sub> concentrations was evaluated  
349 by comparing the model outputs with observations from national monitoring sites. In January,  
350 CRACMM shows higher R value ( $R > 0.7$ ) over northern and eastern China (e.g., BTH, FWP),  
351 whereas lower R values (approximately 0.4) are found in southern regions (e.g., PRD) (Figure  
352 3a). The model generally underestimates PM<sub>2.5</sub> across most areas (Figure 4a), except for the  
353 YRD and SCB regions, where positive biases occur at many sites. This spatial contrast is partly  
354 attributable to dust-related influences, although data from the major dust episode on 13-14  
355 January were excluded from the monthly evaluation, the absence of explicit dust emissions and  
356 the associated complex meteorological conditions likely contributed to PM<sub>2.5</sub> underestimation  
357 of up to  $\sim 30 \mu\text{g}/\text{m}^3$  in northern China. Results for April (Figures 3b and 4b) show generally  
358 good correlations in the eastern regions, while several monitoring sites in the south exhibit  
359 lower R values. Compared to January, the MB is less pronounced. April also experiences dust  
360 storm events. In July, R values decline across all regions relative to January and April, and most  
361 stations exhibit relatively small MB values (Figures 3c and 4c). The strong influence of  
362 temperature and solar radiation on photochemical processes during summer may result in more  
363 pronounced diurnal variations in chemical composition, making the simulation of chemical  
364 processes more challenging (Seinfeld et al., 1998). Moreover, the chemical mechanisms may  
365 inadequately capture non-linear interactions and the influence of SOA (Harrison et al., 2022),  
366 further reducing the correlation. Additionally, synoptic-scale variations can also affect the  
367 spatial distribution and concentration of key atmospheric species (Zhu et al., 2023). R values  
368 are improved in October (Figures 3d and 4d), with 90% of the sites achieving R values of 0.8  
369 and the MB is around  $10 \mu\text{g}/\text{m}^3$  with higher evaluation in SCB and BTH regions. Overall,  
370 wintertime observed peaks generally are underestimated and lower summertime observed  
371 values generally are well captured. The model demonstrates strong performance in January and  
372 October, characterized by higher correlations and smaller biases. However, it had weaker  
373 performance in April and July with lower correlations.



374

375 **Figure 3.** R values between predicted and observed PM<sub>2.5</sub> concentrations using CRACMM with  
 376 the full-volatility emission inventory for January, April, July, and October of 2021.

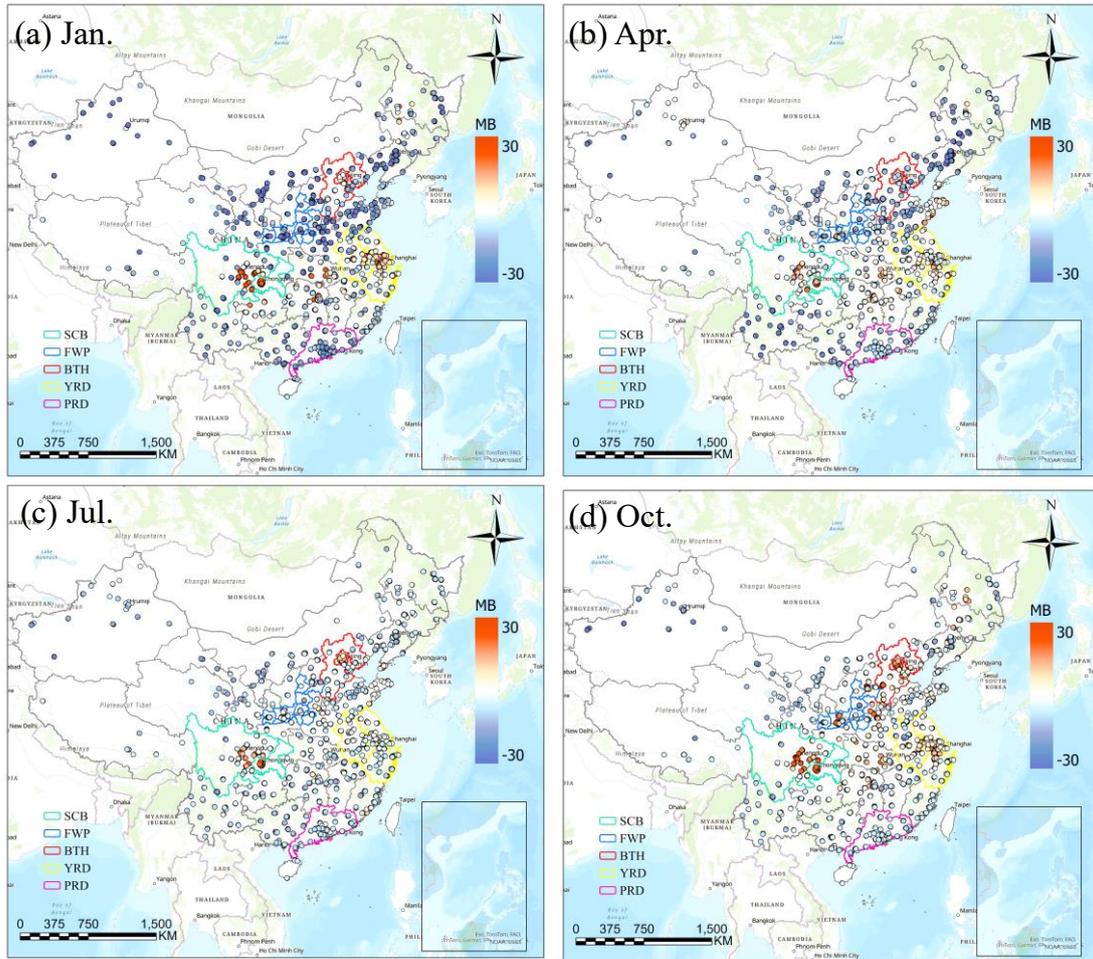
377 Figures S7 and S8 compare the R and MB values between CRACMM (with the full-volatility  
 378 inventory) and CB6r3\_ae7 (with the traditional inventory). In January, CRACMM  
 379 demonstrates notable improvements in R values at several sites in the PRD and YRD regions,  
 380 with increases ranging from 0.2 to 0.4 (Figure S7a), while changes at most other sites remain  
 381 relatively minor (0-0.1). Regarding MB, the most pronounced differences also occur in January:  
 382 some locations in the BTH and YRD regions show higher MB values—up to 10 µg/m<sup>3</sup>—with  
 383 CRACMM, whereas other sites display reduced MB values (Figure S8a). In April (Figures S7b  
 384 and S8b), CRACMM achieves higher R values (0 - 0.2) at certain sites in the YRD, while  
 385 slightly lower correlations (0-0.16) are observed in the PRD compared to CB6r3\_ae7. MB  
 386 values remain elevated in the SCB and parts of the YRD region for CRACMM. For July  
 387 (Figures S7c and S8c), R values from CRACMM are generally comparable to those from  
 388 CB6r3\_ae7. However, MB values tend to decrease across most regions, indicating a potential

389 improvement in bias performance during summer. In October, CRACMM shows moderate  
390 increases in R values—by approximately 0.1 at most sites (Figure S7d). MB values are lower  
391 in the PRD region but higher in the FWP region compared to CB6r3\_ae7 (Figure S8d).

392 In evaluating the CMAQ model's performance for hourly PM<sub>2.5</sub> concentrations, CRACMM  
393 generally shows good correlations with observed data in January and October. However,  
394 discrepancies arise in April, likely due to chemical conditions such as dust storms, springtime  
395 dust events are frequently observed in China, particularly in northern regions, driven by strong  
396 surface winds and synoptic-scale transport (Huo et al., 2025). For July, larger discrepancies are  
397 likely associated with more complex meteorological conditions, as reflected by the relatively  
398 poor performance of wind speed, particularly in the SCB region. In contrast, the simulation  
399 performance of relative humidity and temperature shows only minor differences relative to the  
400 other three months (Tables S11–S13). In addition, photochemical activity typically peaks in the  
401 summer months (June-August) due to stronger solar radiation and higher temperatures (Gu et  
402 al., 2022), which further exacerbates these discrepancies.

403 The model tends to underestimate peak PM<sub>2.5</sub> concentrations during winter but captures lower  
404 summer concentrations more accurately. Comparisons between CRACMM (with the full-  
405 volatility inventory) and CB6r3\_ae7 (using the traditional inventory) highlight improvements  
406 in R values in the PRD and parts of the YRD regions in January for CRACMM, although  
407 performance declines in July.

408



409

410 **Figure 4.** The MB values between predicted and observed PM<sub>2.5</sub> concentrations using  
 411 CRACMM with full-volatility inventory for January, April, July, and October of 2021.

412 The simulation performance of PM<sub>2.5</sub> components in six selected cities using CRACMM with  
 413 the full-volatility inventory was evaluated. Some observations were from the previous  
 414 study(Wang et al., 2024). Tables S14-S17 summarize the statistical values of PM<sub>2.5</sub> components,  
 415 including NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>, OC, and EC, for the four selected months.. Figures S13-S16  
 416 demonstrate that CRACMM effectively captured the overall peaks and troughs of observed  
 417 PM<sub>2.5</sub> concentrations in January. The model also successfully simulated the heavy pollution  
 418 period from January 20th to 25th in Taiyuan, with results similar to our previous study (Wang  
 419 et al., 2024). The three ions were well simulated in both Changzhou and Pudong in the YRD  
 420 region, particularly in Changzhou, where the overall trends of OC and EC showed strong  
 421 consistency with observations. However, for EC in Pudong, the model struggled to capture the  
 422 hourly peak values accurately.

423

424 **Table 2.** Monthly averaged metrics of PM<sub>2.5</sub> evaluation and the number of monitoring sites.

Months	R	IOA	NMB	NME	No.
January	0.46	0.61	-19%	43%	1388
April	0.41	0.52	-29%	47%	1402
July	0.36	0.49	-32%	43%	1309
October	0.68	0.72	-12%	45%	1358
Recommend benchmark(Huang et al., 2021b)	>0.60	>0.70	<±45%	<±55%	/

425 As shown in Table 2, the results of performance evaluation based on observations from over  
 426 1,300 monitoring sites across China for CRACMM shows that the correlation metrics (R and  
 427 IOA) met the recommended benchmark (Huang et al., 2021b) in October, while they fell short  
 428 of the benchmark in January, April, and July. In contrast, the bias metrics (NMB and NME)  
 429 satisfied the recommended benchmark across all four representative months, indicating that the  
 430 model performed well in controlling overall bias. The evaluation metrics for different regions  
 431 and months are provided in Tables S18–S21. Previous studies have reported similar R values.  
 432 For example, the evaluation results of PM<sub>2.5</sub> simulations using CMAQ over five key regions,  
 433 which were the same five regions used in our study, showed that the R were generally below  
 434 0.6 in July (Huang et al., 2024). In addition, a multi-model intercomparison study involving  
 435 CAMx6.2, CAMx7.1, CMAQv5.0.2 and CMAQv5.3.2 reported R values of 0.58, 0.55, 0.60,  
 436 and 0.39, respectively, for mean PM<sub>2.5</sub> concentrations in China during 2014–2017. The  
 437 CAMx6.2, CAMx7.1, and CMAQv5.0.2 simulations covered the entire country, whereas  
 438 CMAQv5.3.2 was evaluated over eastern China (Meng et al., 2026). The heatmaps in Figure  
 439 S9 illustrate the variations in R values across the five key regions—YRD, SCB, PRD, FWP,  
 440 and BTH—for the three chemical mechanisms over four representative months. Corresponding  
 441 heatmaps for MB, NMB and RMSE are presented in Figures S10, S11 and S12, respectively.  
 442 In the YRD region, all mechanisms show relatively stable R values across the four months  
 443 (Figure S9). The CRACMM simulation using the traditional inventory consistently results in  
 444 lower MB (Figure S10) and NMB (Figure S11) values compared to CB6r3\_ae7 and  
 445 Saprc07tic\_ae7i throughout the year, RMSE values range from for 0-2 µg/m<sup>3</sup>(Figure S12).  
 446 When the full-volatility inventory is incorporated, MB improves in April and July but worsens  
 447 in January and October. Similarly, NMB values indicate higher modeled concentrations in all

448 four months with the full-volatility inventory compared to the traditional one. This trend is  
449 consistent with Figure 5, where the YRD region shows higher modeled concentrations from  
450 CRACMM using the full-volatility inventory than from both CB6r3\_ae7 and Sapr07tic\_ae7i  
451 across all months.

452 In the SCB region, R values remain relatively consistent across the three mechanisms (Figure  
453 S9a–c). A slight improvement is observed in July, where R increases from 0.22 with the  
454 traditional inventory to 0.27 with the full-volatility inventory (Figure S9d). However, in  
455 October, R decreases from 0.62 to 0.56 after switching to the full-volatility inventory. The  
456 increase in R value in July suggests that the traditional inventory may underestimate key  
457 precursors (e.g., S/IVOCs), while the full-volatility inventory better captures these species  
458 active at higher temperatures, improving model–observation agreement. In contrast, the  
459 October decrease in R may reflect uncertainties in representing some gas- or particle-phase  
460 organics under cooler conditions. Regarding MB, CRACMM generally shows reduced values  
461 in most months when using the traditional inventory, except for April. With the full-volatility  
462 inventory, MB decreases further in October, while slight increases are observed in the other  
463 months. The trends in NMB follow a similar pattern to those in MB. It could be due to  
464 overestimation of certain intermediate- or low-volatility species under specific conditions.

465 In the PRD region, CRACMM exhibits notable performance improvements in January. As  
466 shown in Figures S9a–c, the R value increases from 0.20 with CB6r3\_ae7 and Sapr07tic\_ae7i  
467 to 0.35 when using CRACMM with the traditional POA inventory. When the full-volatility  
468 inventory is applied (Figure S9d), the R value further increases to 0.50. Concurrently, the MB  
469 improves significantly, decreasing from  $-19.6 \mu\text{g}/\text{m}^3$  to  $-11.8 \mu\text{g}/\text{m}^3$ , and the NMB is reduced  
470 from  $-43\%$  to  $-26\%$ . These results indicate that CRACMM, particularly with the full-volatility  
471 inventory, achieves both higher correlation and lower bias for  $\text{PM}_{2.5}$  simulations in the PRD  
472 region during January. Under the traditional POA inventory, CRACMM tends to underestimate  
473  $\text{PM}_{2.5}$  concentrations in PRD during January (Figure S8a). However, after switching to the full-  
474 volatility inventory, simulated concentrations exceed those predicted by CB6r3\_ae7 (Figure 5a),  
475 primarily due to increased contributions from SOA (Figure 8b). In comparison, during April  
476 and July, the CRACMM simulations using the traditional emissions inventory showed lower R  
477 values than those of CB6r3\_ae7 and Sapr07tic\_ae7i, and the correlation further declined when

478 using the full-volatility emissions inventory. Although CRACMM features a more  
479 comprehensive design for gas-phase oxidation mechanisms, the intense photochemical activity  
480 in July and the rapid oxidation of high concentrations of IVOC precursors may have introduced  
481 more complex SOA formation pathways and product distributions, thereby increasing modeling  
482 uncertainties and weakening the agreement with observations. By October, model performance  
483 had improved. Across all months, CRACMM combined with the full-volatility emissions  
484 inventory consistently outperformed the other mechanisms in terms of MB and NMB,  
485 highlighting the critical role of this inventory in addressing the underestimation of PM<sub>2.5</sub>  
486 associated with traditional POA treatment.

487 In the FWP region, both the R (Figure S9a–c) and MB (Figure S9a–c) values show minimal  
488 variation across the three mechanisms, indicating limited sensitivity to the chemical mechanism  
489 alone. However, notable changes in MB and NMB are observed with the incorporation of the  
490 full-volatility inventory in April, July and October, which align with the higher PM<sub>2.5</sub>  
491 concentrations shown in Figure 5, compared to CB6r3\_ae7. In April and July, MB values shift  
492 from -12.7 µg/m<sup>3</sup> and -5.6 µg/m<sup>3</sup> to -2.4 µg/m<sup>3</sup> and 1.2 µg/m<sup>3</sup>, respectively, with NMB showing  
493 a similar trend. These changes suggest an improvement in model agreement when the full-  
494 volatility inventory is employed. In contrast, both MB and NMB increase in October, indicating  
495 that the full-volatility inventory leads to higher simulated concentrations during this month.

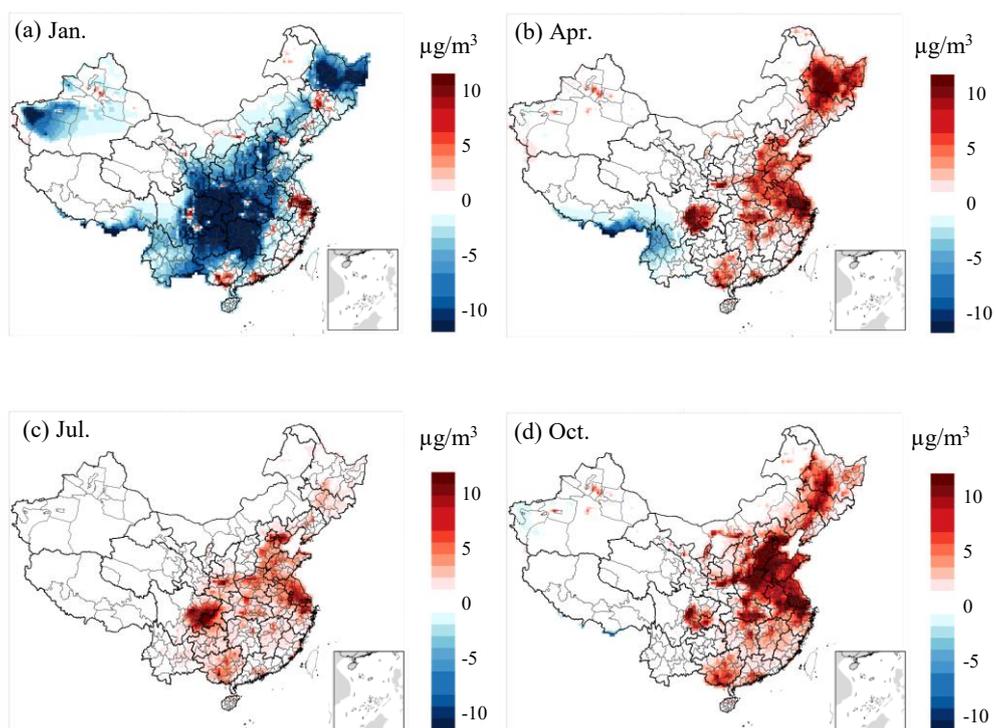
496 In the BTH region, The R (Figure S9a–c) and MB (Figure S10a–c) values remain largely  
497 consistent across the three mechanisms. CRACMM with the traditional POA inventory shows  
498 a decrease in R values from 0.44 to 0.38 compared to CB6r3\_ae7 in July. After incorporating  
499 the full-volatility inventory, the R in BTH experiences a more significant drop, falling to 0.24.  
500 Notably, MB and NMB indicate that the modeled results are lower for CRACMM with the  
501 traditional POA inventory in January, while they are higher in the other months. Additionally,  
502 BTH exhibited the highest IVOC emissions in the inventory (Chang et al., 2022), and  
503 uncertainties in emission estimates may have further contributed to this result.

504 Overall, the differences in R and MB values across the three chemical mechanisms are  
505 relatively small when the traditional POA inventory is used. However, for CRACMM, the MB  
506 in January indicates a stronger underestimation, primarily due to differences in POA (as shown  
507 in Figure 8a), leading to lower modeled concentrations compared to CB6r3\_ae7 and

508 Sapr07tic\_ae7i. With the incorporation of the full-volatility inventory, the MB shifts toward  
509 higher modeled concentrations in the subsequent three months. Notably, more pronounced  
510 differences are observed in January in the PRD region and in July in the BTH region.

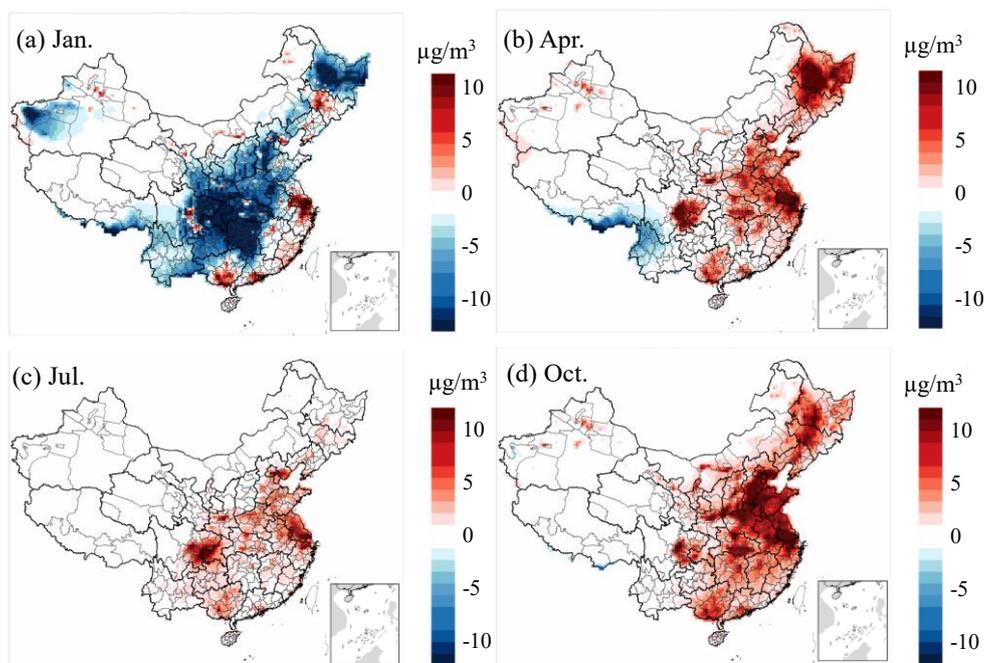
### 511 **3.2 Comparisons of model predicted PM<sub>2.5</sub> between CRACMM and other mechanisms**

512 Figures 5 and 6 illustrate the differences in model outputs between CRACMM with full-  
513 volatility inventory and CB6r3\_ae7, as well as between CRACMM and Sapr07tic\_ae7i. In  
514 January, CRACMM predicts lower PM<sub>2.5</sub> concentrations across central and northern China  
515 compared to both CB6r3\_ae7 (Figure 5a) and Sapr07tic\_ae7i (Figure 6a), with the  
516 differences—up to 10 µg/m<sup>3</sup>—observed in central and north of China. While CRACMM  
517 simulates higher PM<sub>2.5</sub> concentrations in the PRD and YRD regions. For the remaining  
518 months—April, July, and October—CRACMM with full-volatility inventory generally predicts  
519 higher PM<sub>2.5</sub> levels than the other two mechanisms (Figures 5b–d, Figures 6b–d). When  
520 CRACMM and CB6r3\_ae7 are configured with the traditional POA inventory, as shown in  
521 Figure S17, the differences of PM<sub>2.5</sub> concentrations are reduced in April, July, and October  
522 compared with full-volatility POA inventory. But CRACMM still predicts lower PM<sub>2.5</sub> levels  
523 than CB6r3\_ae7 in January. A likely explanation is that the lower photochemical activity leads  
524 to reduced SOA formation, as the enhanced SOA pathways are less active during the winter  
525 months.



526

527 **Figure 5.** Differences in model-predicted PM<sub>2.5</sub> concentrations between CRACMM (full-  
 528 volatility inventory) and CB6r3\_ae7.



529

530 **Figure 6.** Differences in model-predicted PM<sub>2.5</sub> concentrations between CRACMM (full-  
 531 volatility inventory) and Saprc07tic\_ae7i.

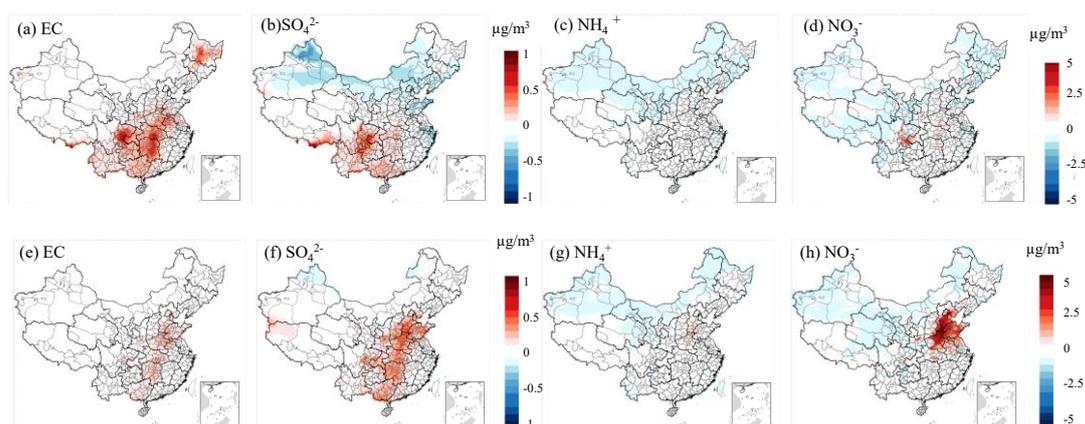
532 **3.3 Comparisons of model predicted PM<sub>2.5</sub> chemical components between CRACMM**  
533 **and other mechanisms**

534 Analysis of Figures 5, 6, and S17 indicates that the most significant differences among the  
535 model simulations occur in January and October, whereas the discrepancies between chemical  
536 mechanisms are substantially smaller in April and July—likely due to the overall lower  
537 pollutant concentrations during these months, which may reduce the sensitivity to mechanistic  
538 differences. Consequently, the subsequent analysis focuses on a detailed comparison of PM<sub>2.5</sub>  
539 component variations between CRACMM and CB6r3\_ae7 for January and October.

540 **3.3.1 Inorganic aerosol**

541 SO<sub>4</sub><sup>2-</sup>, NO<sub>3</sub><sup>-</sup>, and NH<sub>4</sub><sup>+</sup> are the dominant secondary inorganic components in PM<sub>2.5</sub>. Nitrogen  
542 dioxide (NO<sub>2</sub>) and sulfur dioxide (SO<sub>2</sub>) can fully dissolve into cloud water or aerosol liquid  
543 phases and subsequently oxidize to form nitrate and sulfate. Ammonium salts are produced  
544 through the neutralization reactions of these acidic species with atmospheric ammonia (NH<sub>3</sub>).  
545 EC primarily originates from the incomplete combustion of carbonaceous fuels, especially  
546 under oxygen-limited conditions. It is commonly emitted from sources such as vehicle exhaust,  
547 industrial combustion, and biomass burning. From a chemical mechanism perspective,  
548 CRACMM retains the inorganic chemistry framework of RACM2 but incorporates updated  
549 rate constants for several reactions. Specifically, the rate expressions for 26 inorganic reactions  
550 were revised in CRACMM compared to RACM2 (Pye et al., 2023b). Overall, differences in  
551 inorganic component predictions among CRACMM, CB6r3\_ae7, and Saprc07tic\_ae7i are  
552 relatively minor. As shown in Figure 7, predicted concentrations of major inorganic species in  
553 January and October are comparable between CRACMM and CB6r3\_ae7, with differences  
554 ranging from -1 to 1 μg/m<sup>3</sup> for EC and SO<sub>4</sub><sup>2-</sup>, and -5 to 5 μg/m<sup>3</sup> for NH<sub>4</sub><sup>+</sup> and NO<sub>3</sub><sup>-</sup>. These  
555 results suggest that the variation in simulated inorganic aerosol concentrations is only  
556 marginally affected by differences in the inorganic chemistry schemes.

557



558

559 **Figure 7.** Differences in model-predicted PM<sub>2.5</sub> components—(a)EC, (b)SO<sub>4</sub><sup>2-</sup>, (c)NH<sub>4</sub><sup>+</sup>,  
 560 (d)NO<sub>3</sub><sup>-</sup> for January, and (e)EC, (f)SO<sub>4</sub><sup>2-</sup>, (g)NH<sub>4</sub><sup>+</sup>, (h)NO<sub>3</sub><sup>-</sup> for October—between CRACMM  
 561 (with full-volatility inventory) and CB6r3\_ae7. Subplots (a–b) and (e–f) are plotted using the  
 562 same color scale, as are subplots (c–d) and (g–h).

### 563 3.3.2 Organic aerosol

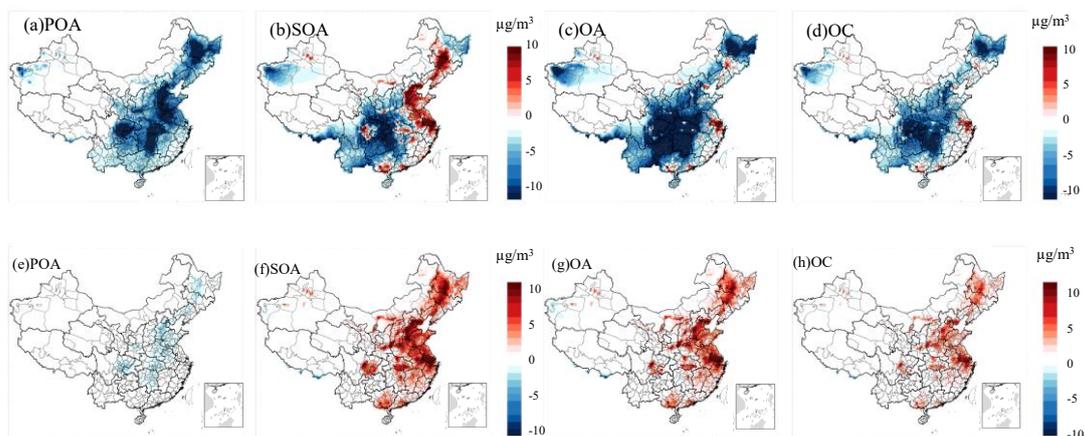
564 In January, CB6r3\_ae7 consistently predicts higher POA concentrations than CRACMM under  
 565 both the traditional and full-volatility inventory configurations, with the most pronounced  
 566 differences occurring in east China (Figure 8a and Figure S18a). In CRACMM, POA aging is  
 567 represented using a modified 2D-VBS framework (Murphy et al., 2017), where  $C_i^*$  range from  
 568  $10^{-2}$  to  $10^3$   $\mu\text{g}/\text{m}^3$ ,  $C_i^*$  represents the effective saturation concentration, which characterizes the  
 569 volatility of organic compounds and influences gas-particle partitioning. A significant portion  
 570 of the alkane-like L/SVOC mass contributing to ambient OA comes from the direct emissions  
 571 of low-volatility species (e.g., AROCN2ALK, AROCN1ALK, AROCP0ALK, AROCP1ALK,  
 572 AROCP2ALK, AROCP3ALK, where the numbers indicate negative (N) or positive (P)  
 573  $\log_{10}(C_i^*[\mu\text{g}/\text{m}^3])$  value. When species reside in the gas-phase as a vapor, it is prefixed with a  
 574 “V” and when in the particle phase, a prefix “A” is used. For example, VROCP2ALK is an  
 575 alkane-like vapor species with  $C_i^*$  of  $10^2$   $\mu\text{g}/\text{m}^3$ , and AROCP2ALK is a particulate species of  
 576 the same volatility. and their oxidation products (e.g., AROCN2OXY2, AROCP0OXY2,  
 577 AROCP1OXY1, AROCP2OXY2, AROCP3OXY2), these species follow a similar naming  
 578 convention as the L/S/IVOC alkanes, where numbers after N and P indicate negative or positive  
 579  $\log_{10}(C_i^*)$  value and the value ends in  $10 \times nO:nC$  (e.g., ROCN2OXY2 is  $C_i^* = 10^{-2}$   $\mu\text{g}/\text{m}^3$  with

580  $nO:nC = 0.2$ ). By contrast, CB6r3\_ae7 adopts a semi-volatile POA approach in which primary  
581 emissions (e.g., LVPO1, SVPO1–3, IVPO1) and their oxidation products (e.g., LVOO1,  
582 LVOO2, SVOO1, SVOO2, SVOO3) partition between gas and particle phases across a  $C_i^*$   
583 range of  $10^{-1}$  to  $10^3$   $\mu\text{g}/\text{m}^3$ . This framework aligns with the 1.5D-VBS scheme proposed by Koo  
584 et al. (2014). Details of the POA species and their properties are provided in Table S7. The most  
585 significant differences in simulated POA concentrations occur in January, likely due to  
586 enhanced partitioning of SVOC to the particle phase under low wintertime temperatures.  
587 Additionally, differences in multigenerational oxidation aging and volatility treatment between  
588 the two mechanisms contribute to the simulation discrepancies. In October, although POA  
589 concentrations in CRACMM remain lower than in CB6r3\_ae7, the difference is less  
590 pronounced compared to January for both POA inventory (Figures 8e and S12e).

591 To better understand the drivers of SOA formation in the two mechanisms, we analyzed the  
592 spatial distribution of SOA concentrations under both traditional and full-volatility POA  
593 inventories. In January, CRACMM predicts higher SOA levels in the BTH and parts of the SCB  
594 regions compared to CB6r3\_ae7 when using the traditional POA inventory (Figure S18b).  
595 Under the full-volatility inventory, CRACMM also shows increased SOA concentrations in the  
596 YRD and PRD regions (Figure 8b). These increases correspond with high IVOC emissions in  
597 BTH and YRD, consistent with the spatial patterns reported by Chang et al. (2022), although  
598 their data reflect monthly averages across January and July. The SOA enhancement in YRD and  
599 PRD under the full-volatility inventory highlights the critical role of IVOC emissions in these  
600 areas. In CB6r3\_ae7, SOA is primarily formed from the oxidation of traditional VOC sources,  
601 such as isoprene, monoterpenes, sesquiterpenes, benzene, toluene, xylene, alkanes, and PAHs  
602 (Carlton et al., 2010; Pye and Pouliot, 2012).

603 In contrast, CRACMM incorporates additional SOA precursor systems, including phenol and  
604 aromatic diols, pinon aldehyde, oxygenated IVOCs, furanone, and other compounds. As a result,  
605 in regions with elevated anthropogenic emissions, CRACMM generally simulates higher SOA  
606 concentrations. However, in the SCB region, SOA levels remain lower, possibly due to the  
607 reduced reactivity of these new precursors under the lower ambient temperatures typical of this  
608 region. In terms of overall OA concentrations, CRACMM generally predicts lower values than  
609 CB6r3\_ae7 across most regions, except for some part of YRD region (Figure 8c and Figure

610 S18c). The spatial distribution of OC is similar to that of OA, with CRACMM also showing  
 611 lower concentrations than CB6r3\_ae7 (Figure 8d and Figure S18d).  
 612 In October (Figures 8f–h and S18e–h), SOA remains the dominant contributor to the differences  
 613 in PM<sub>2.5</sub> concentrations. Under the full-volatility inventory, CRACMM predicts significantly  
 614 higher SOA concentrations compared to CB6r3\_ae7 (Figure 8f), resulting in elevated OA levels  
 615 (Figure 8g). The spatial pattern of OC concentrations (Figure 8h) closely resembles that of OA.  
 616 When the traditional POA inventory is applied, the differences in SOA, OC, and OA  
 617 concentrations between the two mechanisms are minimal (Figures S18e–h). The most  
 618 pronounced increases with the full-volatility inventory are observed in the YRD, PRD, and SCB  
 619 regions, attributable to the inclusion of a more comprehensive set of SOA precursors.



620  
 621 **Figure 8.** Differences in model-predicted PM<sub>2.5</sub> components—(a) POA, (b) SOA, (c) OA, and  
 622 (d) OC for January, and (e) POA, (f) SOA, (g) OA, and (h) OC for October—between  
 623 CRACMM (with full-volatility inventory) and CB6r3\_ae7.

624 Overall, in January, the primary differences between CRACMM and CB6r3\_ae7 stem from  
 625 lower POA concentrations in CRACMM, primarily due to semi-volatile partitioning and  
 626 reduced aging of semi-volatile POA species at lower temperatures. SOA concentrations are  
 627 elevated in eastern China but reduced over the SCB, reflecting both the slower oxidation of  
 628 additional SOA precursors under winter conditions in the SCB and the greater availability of  
 629 these precursors in the eastern region. In October, the key differences in model predictions are  
 630 primarily driven by the POA inventory used. The full-volatility inventory yields higher SOA  
 631 concentrations than the traditional inventory, largely due to the inclusion of L/S/IVOCs, which

632 are efficient SOA precursors.

### 633 **3.4 Sensitivity study on PM<sub>2.5</sub> and SOA responses to changes of precursors**

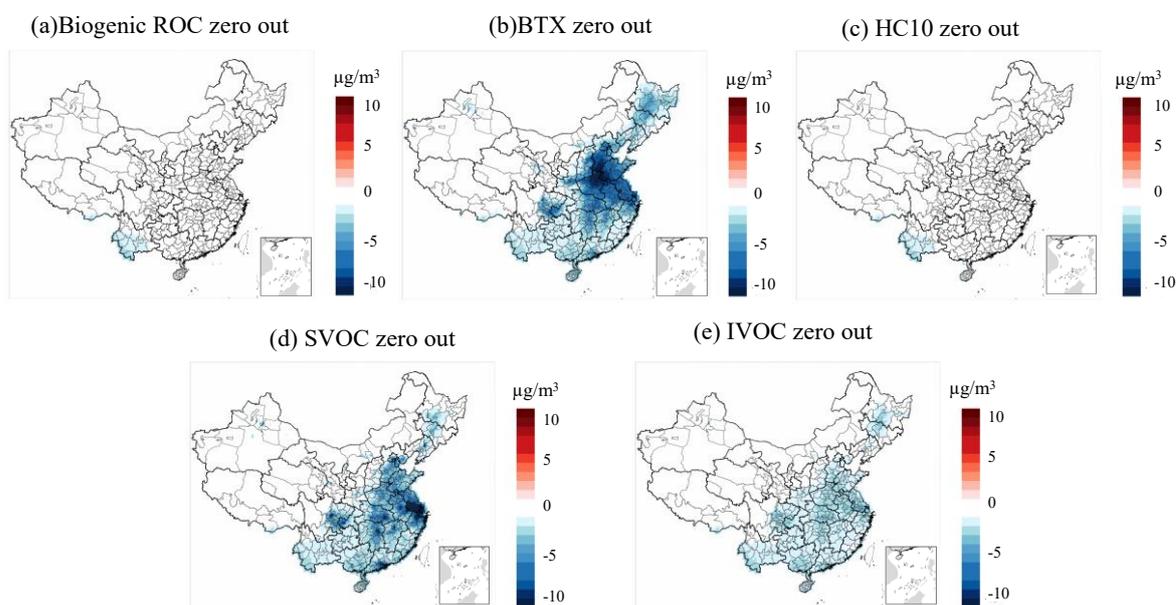
634 In this section, CMAQ simulations with emission perturbations are conducted to identify the  
635 key drivers of PM<sub>2.5</sub> formation in January, when PM<sub>2.5</sub> concentrations are notably high. A series  
636 of emission sensitivity simulations were performed within CMAQ to assess the role of  
637 precursor ROC systems in PM<sub>2.5</sub> formation using CRACMM with the full-volatility inventory  
638 across China. These sensitivity simulations involved running zeroed emission scenarios for  
639 January (i.e., setting emissions of a specific chemical class or sector to zero) to examine how  
640 PM<sub>2.5</sub> concentrations respond to changes in emissions. A subset of these sensitivity simulations  
641 was also conducted using the CB6r3\_ae7 and Sapr07tic\_ae7i mechanisms. A detailed list of  
642 all the zeroed emission simulations is provided in Tables 3 and S22. Due to the non-linear nature  
643 of PM<sub>2.5</sub> production in response to ROC perturbations, these simulations offer an initial  
644 evaluation of how PM<sub>2.5</sub> formation responds to reduced ROC emissions, providing valuable  
645 insights into how chemical systems behave under varying emission conditions in the three  
646 mechanisms.

647 **Table 3.** List of emission reductions relative to the base simulations in CMAQ-CRACMM.

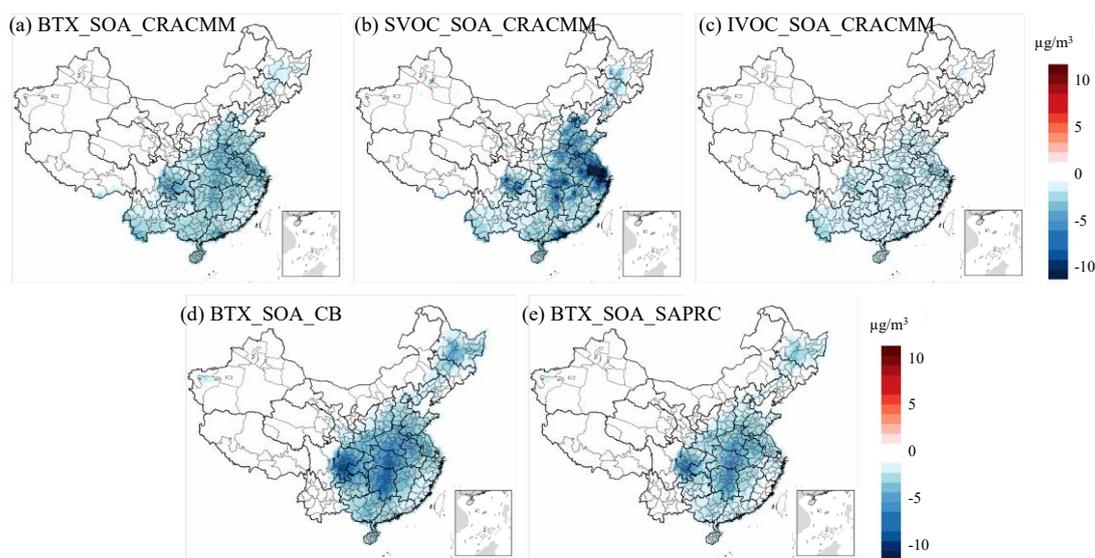
Chemical Mechanism	Emission Reduction
CRACMM	Benzene, toluene and xylene-like emissions set to zero
CRACMM	Biogenic-ROC emissions set to zero
CRACMM	IVOC emissions set to zero
CRACMM	SVOC emissions set to zero
CRACMM	HC10 zero out (decane and species of similar reactivity)

648 Figure 9 shows domain-wide differences in average PM<sub>2.5</sub> concentrations between the base  
649 CRACMM simulation and a series of zeroed emission simulations. A similar spatial pattern in  
650 PM<sub>2.5</sub> response was observed for zeroed biogenic and HC10 emissions (Figures S19a, c in  
651 percentage and Figures 9a, c in concentration), with CRACMM predicting a modest 1% change  
652 in PM<sub>2.5</sub> and less than 3 µg/m<sup>3</sup> in many parts of China. This can be attributed to the low SOA  
653 yield by mass (0.09 g/g) for HC10 compounds (Pye et al., 2023), and the generally low winter  
654 emissions of biogenic ROC, as shown in previous studies (Pye et al., 2023). Zeroing BTX  
655 emissions resulted in average PM<sub>2.5</sub> concentration changes of -20% to 0% (Figure S19b) and -  
656 10 to 0 µg/m<sup>3</sup> (Figure 9b), particularly in the YRD and BTH regions, where BTX emissions are

657 highest. The high PM<sub>2.5</sub> formation potential of BTX compounds is attributed to their overall  
658 emission abundance and high SOA yield by mass (~0.5 g/g). Moreover, ozone levels in urban  
659 areas with significant BTX emissions also decrease in the BTX zero-out scenario (Figure S20).  
660 This effect is particularly evident in the PRD region, where the reduction reaches up to 10 µg/m<sup>3</sup>.  
661 This aligns with the findings of Place et al. (2023). One factor is the removal of BTX emissions,  
662 which serve as precursors for SOA formation. The second factor is that zeroing BTX emissions  
663 leads to a decrease in O<sub>3</sub>, which weakens atmospheric oxidizing capacity and reduces SOA  
664 formation. However, it is important to note that Place’s study was conducted in summer.



665  
666 **Figure 9.** Changes in PM<sub>2.5</sub> concentrations between each zero-out scenario and its  
667 corresponding base simulation: (a) biogenic ROC emission, (b) BTX emission, (c) HC10  
668 emission, (d) SVOC emission, and (e) IVOC emission with CRACMM.



669  
 670 **Figure 10.** Changes in SOA concentrations between each zero-out scenario and its  
 671 corresponding base simulation: (a) BTX emissions, (b) SVOC emissions, (c) IVOC emissions  
 672 with CRACMM, (d) BTX emissions with CB6r3\_ae7, (e) zeroed BTX emissions with  
 673 Saprc07tic\_ae6.

674 As shown in Figures 10a, d, and e, the impact of zeroing BTX species on SOA formation for  
 675 the three mechanisms accounts for only approximately 50% of the total  $PM_{2.5}$  change observed  
 676 in Figures 9b, S16a and c. This suggests that the changes in  $PM_{2.5}$  concentrations resulting from  
 677 the removal of BTX emissions are not solely due to SOA formation but may also involve other  
 678 pathways or chemical processes. In contrast, the influence of S/IVOC species on  $PM_{2.5}$   
 679 concentrations in the CRACMM mechanism is primarily driven by SOA formation. This  
 680 conclusion is supported by the spatial distribution and concentration differences shown in  
 681 Figures 9d, e, and Figures 10b, c, which exhibit nearly identical patterns. These similarities  
 682 indicate that the effects of S/IVOC emissions on  $PM_{2.5}$  are mainly driven by SOA production.  
 683 The largest  $PM_{2.5}$  response was observed when emissions from SVOC sources were excluded  
 684 from the simulation (Figure S19d), primarily because SVOCs have the highest yield, exceeding  
 685 1.0 g/g. The percentage changes in  $PM_{2.5}$  range from -40% to 0%, with a concentration  
 686 reduction of more than  $-10 \mu\text{g}/\text{m}^3$  (Figure 9d), particularly in the YRD region, where SVOC  
 687 emissions are substantial. For IVOCs, the reduction is about -5% across much of China, except  
 688 in the western regions where IVOC emissions are very low (Figure S19e), resulting in a  
 689 reduction of less than  $-5 \mu\text{g}/\text{m}^3$  (Figure 9e).

690 A similar  $\Delta\text{PM}_{2.5}$  response in percentage (Figure S21) and concentration change (Figure S22)  
691 was observed when biogenic and BTX emissions were zeroed in simulations using  
692 CB6r3\_ae7\_ae7 and Saprc07tic\_ae7i. However, the CRACMM simulation with zeroed  
693 biogenic emissions (Figure 9a) showed a more pronounced and widespread decrease in  $\text{PM}_{2.5}$   
694 compared to both CB6r3\_ae7\_ae7 and Saprc07tic\_ae7i. This difference can be attributed to the  
695 inclusion of new S/IVOC species in MEGAN, which are not accounted for in the other two  
696 mechanisms. Additionally, zeroing BTX emissions had a greater impact in CB6r3\_ae7,  
697 particularly in central China, compared to Saprc07tic\_ae7i and CRACMM. A possible reason  
698 for this is that CRACMM reduces the number of lumped species in BTX and enhances the  
699 representation of aromatic IVOC species, such as single-ring aromatics  $\log_{10}(Ci^*) \approx 5$   
700 (ROCP5ARO) and  $\log_{10}(Ci^*) \approx 6$  (ROCP6ARO). These species are included in CB6r3\_ae7  
701 under categories like m-xylene and other more reactive aromatics (XYM), as well as less  
702 reactive aromatics (XYE). As a result, CRACMM incorporates fewer species in BTX emissions  
703 compared to CB6r3\_ae7.

### 704 **3.5 Uncertainty Analysis and limitations**

#### 705 **3.5.1 Limitations of VOC Speciation Mapping**

706 In the speciation mapping process, explicit species were directly mapped across the CRACMM,  
707 CB6r3\_ae7, and Saprc07tic\_ae7i mechanisms, as these species are explicitly represented and  
708 therefore allow for one-to-one mapping. Consequently, their spatial distributions were assumed  
709 to be identical across the three mechanisms, and uncertainties associated with their emission  
710 estimates were considered negligible. In contrast, lumped species present greater complexity  
711 due to differences in the VOC species included within each lumped mechanism species. Given  
712 that a direct one-to-one mapping between lumped species is not feasible, the mapping was  
713 performed by matching the dominant lumped VOC species across mechanisms based on their  
714 relative emission magnitudes. Overall, CRACMM incorporates a more comprehensive set of  
715 VOC species than either CB6r3\_ae7 or Saprc07tic\_ae7i. The total mapped emissions associated  
716 with each mechanism-specific inventory are summarized in Table S6. Nevertheless,  
717 uncertainties remain due to regional differences in emission profiles. In particular, the total  
718 emissions and source sector distributions of VOC species in Chinese emission inventories may  
719 differ from those represented in the NEI. Such discrepancies introduce additional uncertainty

720 into the speciation mapping process.

### 721 **3.5.2 Uncertainty in Mapping L/S/IVOC Emissions**

722 For the species which can be mapped on a one-to-one basis in Table S8, the associated  
723 uncertainty is assumed to be zero. For species mapping involving the same  $\log_{10}(C_i^*)$  values,  
724 such as VROCP1OXY1, VROCP0OXY4, and VROCN2OXY4, some uncertainty may be  
725 introduced due to the proximity in volatility space. Anthropogenic L/S/IVOCs emission  
726 inventories for China (Figure S4) contains (a) particle-phase emissions with full-volatility  
727 coverage and (b) gas-phase emission inventories (Chang et al., 2022). Based on uncertainties  
728 of activity data and emission factors for each sector, the uncertainty of OA emissions can be  
729 quantified using a Monte Carlo method. According to Chang et al.(2022), the overall  
730 uncertainties at the 95% confidence interval for LVOC, SVOC, IVOC and VOC are (-40%,  
731 +43%), (-35%, 38%), (-33%, +33%) and (-21%, +28%), respectively. The overall uncertainty  
732 for L/S/IVOC is (-25%, +30%). Uncertainties across sectors tend to partially offset each other,  
733 resulting in a total emission uncertainty that is often smaller than that of the individual sectors.  
734 S/IVOC emissions from domestic Volatile Chemical Products (VCPs) have the largest  
735 uncertainties (-81%, +143%), followed by open biomass burning (-58%, +81%) and industrial  
736 VCPs (-50%, +65%). Although the emission factors are based on local experiments, emissions  
737 from domestic fossil fuel and biomass burning still have considerable uncertainties (-38%,  
738 +62%) and (-38%, +51%), respectively.

## 739 **4. Conclusions**

740 This study introduces the newly mapped VOC and POA inventories (both traditional and full-  
741 volatility) for CRACMM and presents the first comprehensive evaluation of PM<sub>2.5</sub> predictions  
742 using the newly developed CRACMM chemical mechanism. The performance of CRACMM  
743 with CB6r3\_ae7, and Sapr07tic\_ae7i are compared, and results demonstrate that CMAQ with  
744 CRACMM provides reliable predictions of PM<sub>2.5</sub> and its components across China during the  
745 months of January, April, July, and October 2021, although there are discrepancies in some  
746 complex regions.

747 In conclusion, the comparison of the three chemical mechanisms using the traditional POA  
748 inventory reveals that differences in R and MB values are generally small. However, with the

749 replacement of the full-volatility inventory, CRACMM tends to predict lower PM<sub>2.5</sub>  
750 concentrations in January across most regions of China except PRD and YRD. In the other  
751 months, CRACMM predicts higher concentrations than CB6r3\_ae7 and Saprc07tic\_ae7i when  
752 the full-volatility inventory is incorporated. The differences in PM<sub>2.5</sub> concentrations in January,  
753 are primarily attributed to lower POA concentrations, which are influenced by semi-volatile  
754 partitioning and reduced aging of semi-volatile POA species under lower temperatures. In  
755 contrast, CRACMM simulates elevated SOA concentrations in eastern China due to enhanced  
756 precursor availability, while reduced SOA formation is observed in the SCB, where winter  
757 conditions slow the oxidation of precursors. The inclusion of the full-volatility inventory in  
758 CRACMM results in higher SOA concentrations in October, driven by increased precursor  
759 availability. Overall, CRACMM demonstrates improved performance in terms of R and MB,  
760 particularly in January and October for the PRD region, but performs less well in April and  
761 July, particularly in the BTH region, compared to CB6r3\_ae7. Additionally, CRACMM with  
762 the full-volatility inventory increase in simulated PM<sub>2.5</sub> concentrations, resulting in smaller  
763 deviations from observation across many regions, highlighting the importance of including  
764 S/IVOC emissions in the chemical mechanism. Emission perturbation simulations using  
765 CMAQ further emphasize the significant role of various emission species, particularly BTX  
766 and SVOC, in driving PM<sub>2.5</sub> formation. The SOA contribution from BTX emissions accounts  
767 for nearly 50% of the PM<sub>2.5</sub> changes, while S/IVOC emissions primarily influence PM<sub>2.5</sub>  
768 through SOA formation. BTX emissions had a more significant impact in CB6r3\_ae7,  
769 particularly in central China, partly due to the fewer VOC species included in the lumped BTX  
770 of the CRACMM mechanism. Future assessments of O<sub>3</sub> predictions with CRACMM will offer  
771 additional constraints on the gas and aerosol chemistry that contributes to PM<sub>2.5</sub> formation.

772 **Data availability.** The model simulation is based on the CMAQ v5.4 developed by the U.S.  
773 EPA, and the code is publicly available at <https://doi.org/10.5281/zenodo.7218076>. Biogenic  
774 emissions were estimated using the MEGANv3.2 model (Guenther et al., 2012b), which is  
775 available at <https://bai.ess.uci.edu/megan/data-and-code/megan32>. All input data to reproduce  
776 the results and figures in this paper, has been archived on Zenodo ([10.5281/zenodo.18222704](https://doi.org/10.5281/zenodo.18222704)).  
777 (Su and Chen, 2025)) and is freely accessible.

778 **Supplement.** The supplement related to this article is available online at

779 **Author contributions.** LL and YJW designed the study. QFS conducted the modeling work,

780 formal analysis and drafted the original manuscript. YFC contributed to data curation and  
781 emissions preparation. DCW contributed to model revision, manuscript editing and reviewing.  
782 HOTP, GS, BM, BP and YJW provided critical review and editing of the manuscript. LH and  
783 YJW contributed to modelling setup. LL supervised the project and secured funding.

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## 794 **References**

- 795 Boylan, J. W. and Russell, A. G.: PM and light extinction model performance metrics, goals,  
796 and criteria for three-dimensional air quality models, *Atmospheric Environment*, 40,  
797 4946-4959,doi: <https://doi.org/10.1016/j.atmosenv.2005.09.087>, 2006.
- 798 Byun, D. and Schere, K. L.: Review of the Governing Equations, Computational Algorithms,  
799 and Other Components of the Models-3 Community Multiscale Air Quality (CMAQ)  
800 Modeling System, *Applied Mechanics Reviews*, 59, 51-77,doi: 10.1115/1.2128636 %J  
801 *Applied Mechanics Reviews*, 2006.
- 802 Cao, L., Li, S., and Sun, L.: Study of different Carbon Bond 6 (CB6) mechanisms by using a  
803 concentration sensitivity analysis, *Atmos. Chem. Phys.*, 21, 12687-12714,doi:  
804 10.5194/acp-21-12687-2021, 2021.
- 805 Carlton, A. G., Bhave, P. V., Napelenok, S. L., Edney, E. O., Sarwar, G., Pinder, R. W.,  
806 Pouliot, G. A., and Houyoux, M.: Model Representation of Secondary Organic Aerosol in  
807 CMAQv4.7, *Environ. Sci. Technol.*, 44, 8553-8560,doi: 10.1021/es100636q, 2010.
- 808 Carter, W.: Implementation Of The Saprc-99 Chemical Mechanism Into The Models-3  
809 Framework, 2000.
- 810 Carter, W. P. L.: DOCUMENTATION OF THE SAPRC-99 CHEMICAL MECHANISM FOR  
811 VOC REACTIVITY ASSESSMENT VOLUME 1 OF 2 DOCUMENTATION TEXT,
- 812 Carter, W. P. L.: Development of the SAPRC-07 chemical mechanism, *Atmos. Environ.*, 44,  
813 5324-5335,doi: 10.1016/j.atmosenv.2010.01.026, 2010.
- 814 Chang, X., Zhao, B., Zheng, H., Wang, S., Cai, S., Guo, F., Gui, P., Huang, G., Wu, D., Han,  
815 L., Xing, J., Man, H., Hu, R., Liang, C., Xu, Q., Qiu, X., Ding, D., Liu, K., Han, R.,  
816 Robinson, A. L., and Donahue, N. M.: Full-volatility emission framework corrects  
817 missing and underestimated secondary organic aerosol sources, *One Earth*, 5, 403-  
818 412,doi: 10.1016/j.oneear.2022.03.015, 2022.
- 819 Deng, C., Tian, S., Li, Z., and Li, K.: Spatiotemporal characteristics of PM(2.5) and ozone

820 concentrations in Chinese urban clusters, *Chemosphere*, 295, 133813,doi:  
821 10.1016/j.chemosphere.2022.133813, 2022.

822 Derwent, R.: Intercomparison of chemical mechanisms for air quality policy formulation and  
823 assessment under North American conditions, *J. Air Waste Manag. Assoc.*, 67, 789-  
824 796,doi: 10.1080/10962247.2017.1292969, 2017.

825 Emery, C., Liu, Z., Russell, A. G., Odman, M. T., Yarwood, G., and Kumar, N.:  
826 Recommendations on statistics and benchmarks to assess photochemical model  
827 performance, *Journal of the Air & Waste Management Association (1995)*, 67, 582-  
828 598,doi: 10.1080/10962247.2016.1265027, 2017.

829 Goliff, W. S., Stockwell, W. R., and Lawson, C. V.: The regional atmospheric chemistry  
830 mechanism, version 2, *Atmospheric Environment*, 68, 174-185,doi:  
831 <https://doi.org/10.1016/j.atmosenv.2012.11.038>, 2013.

832 Gu, J., Chen, Z., Zhang, N., Peng, S., Cui, W., Huo, G., and Chen, F.: Characterization of  
833 Atmospheric Fine Particles and Secondary Aerosol Estimated under the Different  
834 Photochemical Activities in Summertime Tianjin, China, *International journal of  
835 environmental research and public health*, 19,doi: 10.3390/ijerph19137956, 2022.

836 Guenther, A. B., Jiang, X., Heald, C. L., Sakulyanontvittaya, T., Duhl, T., Emmons, L. K., and  
837 Wang, X.: The Model of Emissions of Gases and Aerosols from Nature version 2.1  
838 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions,  
839 *Geosci. Model Dev.*, 5, 1471-1492,doi: 10.5194/gmd-5-1471-2012, 2012.

840 Harrison, R. M., Beddows, D. C. S., Tong, C., and Damayanti, S.: Non-linearity of secondary  
841 pollutant formation estimated from emissions data and measured precursor-secondary  
842 pollutant relationships, *npj Climate and Atmospheric Science*, 5, 71,doi: 10.1038/s41612-  
843 022-00297-9, 2022.

844 Huang, L., Wang, Q., Wang, Y., Emery, C., Zhu, A., Zhu, Y., Yin, S., Yarwood, G., Zhang, K.,  
845 and Li, L.: Simulation of secondary organic aerosol over the Yangtze River Delta region:  
846 The impacts from the emissions of intermediate volatility organic compounds and the  
847 SOA modeling framework, *Atmospheric Environment*, 246, 118079,doi:  
848 <https://doi.org/10.1016/j.atmosenv.2020.118079>, 2021a.

849 Huang, L., Zhu, Y., Zhai, H., Xue, S., Zhu, T., Shao, Y., Liu, Z., Emery, C., Yarwood, G.,  
850 Wang, Y., Fu, J., Zhang, K., and Li, L.: Recommendations on benchmarks for numerical  
851 air quality model applications in China – Part 1: PM<sub>2.5</sub> and chemical species, *Atmos.  
852 Chem. Phys.*, 21, 2725-2743,doi: 10.5194/acp-21-2725-2021, 2021b.

853 Huang, L., Wu, Z. a., Liu, H., Yarwood, G., Huang, D., Wilson, G., Chen, H., Ji, D., Tao, J.,  
854 Han, Z., Wang, Y., Wang, H., Huang, C., and Li, L.: An improved framework for  
855 efficiently modeling organic aerosol (OA) considering primary OA evaporation and  
856 secondary OA formation from VOCs, IVOCs, and SVOCs, *Environmental Science:  
857 Atmospheres*,doi: 10.1039/d4ea00060a, 2024.

858 Huo, Q., Yin, Z., Ma, X., and Wang, H.: Distinctive dust weather intensities in North China  
859 resulted from two types of atmospheric circulation anomalies, *Atmos. Chem. Phys.*, 25,  
860 1711-1724,doi: 10.5194/acp-25-1711-2025, 2025.

861 Kang, M., Zhang, H., and Ying, Q.: Effectiveness of emission controls on atmospheric  
862 oxidation capacity and air pollutant concentrations: uncertainties due to chemical  
863 mechanisms and inventories, *Atmos. Chem. Phys.*, 25, 11453-11467,doi: 10.5194/acp-25-

864 11453-2025, 2025.

865 Kang, M., Hu, J., Zhang, H., and Ying, Q.: Evaluation of a highly condensed SAPRC  
866 chemical mechanism and two emission inventories for ozone source apportionment and  
867 emission control strategy assessments in China, *Sci. Total Environ.*, 813, 151922,doi:  
868 10.1016/j.scitotenv.2021.151922, 2022.

869 Kang, Y.-H., Oh, I., Jeong, J.-H., Bang, J.-H., Kim, Y.-K., Kim, S., Kim, E., Hong, J.-H., and  
870 Lee, D.-G. J. J. o. E. S. I.: Comparison of CMAQ ozone simulations with two chemical  
871 mechanisms (SAPRC99 and CB05) in the Seoul metropolitan region, *25*, 85-97, 2016.

872 Kim, K.-H., Kabir, E., and Kabir, S.: A review on the human health impact of airborne  
873 particulate matter, *Environment International*, 74, 136-143,doi:  
874 <https://doi.org/10.1016/j.envint.2014.10.005>, 2015.

875 Koo, B., Knipping, E., and Yarwood, G.: 1.5-Dimensional volatility basis set approach for  
876 modeling organic aerosol in CAMx and CMAQ, *Atmos. Environ.*, 95, 158-164,doi:  
877 10.1016/j.atmosenv.2014.06.031, 2014.

878 Liu, J., Niu, X., Zhang, L., Yang, X., Zhao, P., and He, C.: Exposure risk assessment and  
879 synergistic control pathway construction for O<sub>3</sub>-PM<sub>2.5</sub> compound pollution in China,  
880 *Atmospheric Environment: X*, 21, 100240,doi:  
881 <https://doi.org/10.1016/j.aeaoa.2024.100240>, 2024.

882 Luecken, D. J., Yarwood, G., and Hutzell, W. T.: Multipollutant modeling of ozone, reactive  
883 nitrogen and HAPs across the continental US with CMAQ-CB6, *Atmos. Environ.*, 201,  
884 62-72,doi: 10.1016/j.atmosenv.2018.11.060, 2019.

885 Mao, J., Li, L., Li, J., Sulaymon, I. D., Xiong, K., Wang, K., Zhu, J., Chen, G., Ye, F., Zhang,  
886 N., Qin, Y., Qin, M., and Hu, J.: Evaluation of Long-Term Modeling Fine Particulate  
887 Matter and Ozone in China During 2013–2019, Volume 10 - 2022,doi:  
888 10.3389/fenvs.2022.872249, 2022.

889 Meng, F., Du, X., Tang, W., He, J., Li, Y., Wang, X., Yu, S., Tang, X., Xing, J., Xie, M., Zeng,  
890 L., and Dong, H.: Evaluation of Regional Atmospheric Models for Air Quality  
891 Simulations in the Winter Season in China, *Atmosphere*, 17, 1,doi:  
892 10.3390/atmos17010001, 2026.

893 Murphy, B. N., Woody, M. C., Jimenez, J. L., Carlton, A. M. G., Hayes, P. L., Liu, S., Ng, N.  
894 L., Russell, L. M., Setyan, A., Xu, L., Young, J., Zaveri, R. A., Zhang, Q., and Pye, H. O.  
895 T.: Semivolatile POA and parameterized total combustion SOA in CMAQv5.2: impacts  
896 on source strength and partitioning, *Atmos. Chem. Phys.*, 17, 11107-11133,doi:  
897 10.5194/acp-17-11107-2017, 2017.

898 Murray, C. J. L., Aravkin, A. Y., and Zheng, P.: Global burden of 87 risk factors in 204  
899 countries and territories, 1990–2019: a systematic analysis for the Global Burden  
900 of Disease Study 2019, *The Lancet*, 396, 1223-1249,doi: 10.1016/S0140-6736(20)30752-  
901 2, 2020.

902 Ng, N. L., Kwan, A. J., Surratt, J. D., Chan, A. W. H., Chhabra, P. S., Sorooshian, A., Pye, H.  
903 O. T., Crounse, J. D., Wennberg, P. O., Flagan, R. C., and Seinfeld, J. H.: Secondary  
904 organic aerosol (SOA) formation from reaction of isoprene with nitrate radicals (NO<sub>3</sub>),  
905 *Atmospheric Chemistry and Physics*, 8, 4117-4140,doi: 10.5194/acp-8-4117-2008, 2008.

906 Otte, T. L. and Pleim, J. E.: The Meteorology-Chemistry Interface Processor (MCIP) for the  
907 CMAQ modeling system: updates through MCIPv3.4.1, *Geosci. Model Dev.*, 3, 243-

908 256,doi: <http://10.5194/gmd-3-243-2010>, 2010.

909 Place, B. K., Hutzell, W. T., Appel, K. W., Farrell, S., Valin, L., Murphy, B. N., Seltzer, K. M.,  
910 Sarwar, G., Allen, C., Piletic, I. R., D'Ambro, E. L., Saunders, E., Simon, H., Torres-  
911 Vasquez, A., Pleim, J., Schwantes, R. H., Coggon, M. M., Xu, L., Stockwell, W. R., and  
912 Pye, H. O. T.: Sensitivity of northeastern US surface ozone predictions to the  
913 representation of atmospheric chemistry in the Community Regional Atmospheric  
914 Chemistry Multiphase Mechanism (CRACMMv1.0), *Atmos. Chem. Phys.*, 23, 9173-  
915 9190,doi: 10.5194/acp-23-9173-2023, 2023a.

916 Place, B. K., Hutzell, W. T., Appel, K. W., Farrell, S., Valin, L., Murphy, B. N., Seltzer, K. M.,  
917 Sarwar, G., Allen, C., Piletic, I. R., D'Ambro, E. L., Saunders, E., Simon, H., Torres-  
918 Vasquez, A., Pleim, J., Schwantes, R. H., Coggon, M. M., Xu, L., Stockwell, W. R., and  
919 Pye, H. O. T.: Sensitivity of northeastern US surface ozone predictions to the  
920 representation of atmospheric chemistry in the Community Regional Atmospheric  
921 Chemistry Multiphase Mechanism (CRACMMv1.0), *Atmospheric Chemistry and  
922 Physics*, 23, 9173-9190,doi: 10.5194/acp-23-9173-2023, 2023b.

923 Pye, H. O. T. and Pouliot, G. A.: Modeling the Role of Alkanes, Polycyclic Aromatic  
924 Hydrocarbons, and Their Oligomers in Secondary Organic Aerosol Formation, *Environ.  
925 Sci. Technol.*, 46, 6041-6047,doi: 10.1021/es300409w, 2012.

926 Pye, H. O. T., Luecken, D. J., Xu, L., Boyd, C. M., Ng, N. L., Baker, K. R., Ayres, B. R.,  
927 Bash, J. O., Baumann, K., Carter, W. P. L., Edgerton, E., Fry, J. L., Hutzell, W. T.,  
928 Schwede, D. B., and Shepson, P. B.: Modeling the Current and Future Roles of  
929 Particulate Organic Nitrates in the Southeastern United States, *Environ. Sci. Technol.*, 49,  
930 14195-14203,doi: 10.1021/acs.est.5b03738, 2015a.

931 Pye, H. O. T., Luecken, D. J., Xu, L., Boyd, C. M., Ng, N. L., Baker, K. R., Ayres, B. R.,  
932 Bash, J. O., Baumann, K., Carter, W. P. L., Edgerton, E., Fry, J. L., Hutzell, W. T.,  
933 Schwede, D. B., and Shepson, P. B.: Modeling the Current and Future Roles of  
934 Particulate Organic Nitrates in the Southeastern United States, *Environmental Science &  
935 Technology*, 49, 14195-14203,doi: 10.1021/acs.est.5b03738, 2015b.

936 Pye, H. O. T., Pinder, R. W., Piletic, I. R., Xie, Y., Capps, S. L., Lin, Y.-H., Surratt, J. D.,  
937 Zhang, Z., Gold, A., Luecken, D. J., Hutzell, W. T., Jaoui, M., Offenberg, J. H.,  
938 Kleindienst, T. E., Lewandowski, M., and Edney, E. O.: Epoxide Pathways Improve  
939 Model Predictions of Isoprene Markers and Reveal Key Role of Acidity in Aerosol  
940 Formation, *Environ. Sci. Technol.*, 47, 11056-11064,doi: 10.1021/es402106h, 2013.

941 Pye, H. O. T., Murphy, B. N., Xu, L., Ng, N. L., Carlton, A. G., Guo, H., Weber, R., Vasilakos,  
942 P., Appel, K. W., Budisulistiorini, S. H., Surratt, J. D., Nenes, A., Hu, W., Jimenez, J. L.,  
943 Isaacman-VanWertz, G., Misztal, P. K., and Goldstein, A. H.: On the implications of  
944 aerosol liquid water and phase separation for organic aerosol mass, *Atmos. Chem. Phys.*,  
945 17, 343-369,doi: 10.5194/acp-17-343-2017, 2017.

946 Pye, H. O. T., Place, B. K., Murphy, B. N., Seltzer, K. M., D'Ambro, E. L., Allen, C., Piletic,  
947 I. R., Farrell, S., Schwantes, R. H., Coggon, M. M., Saunders, E., Xu, L., Sarwar, G.,  
948 Hutzell, W. T., Foley, K. M., Pouliot, G., Bash, J., and Stockwell, W. R.: Linking gas,  
949 particulate, and toxic endpoints to air emissions in the Community Regional Atmospheric  
950 Chemistry Multiphase Mechanism (CRACMM), *Atmospheric Chemistry and Physics*, 23,  
951 5043-5099,doi: 10.5194/acp-23-5043-2023, 2023a.

952 Pye, H. O. T., Place, B. K., Murphy, B. N., Seltzer, K. M., D'Ambro, E. L., Allen, C., Piletic,  
953 I. R., Farrell, S., Schwantes, R. H., Coggon, M. M., Saunders, E., Xu, L., Sarwar, G.,  
954 Hutzell, W. T., Foley, K. M., Pouliot, G., Bash, J., and Stockwell, W. R.: Linking gas,  
955 particulate, and toxic endpoints to air emissions in the Community Regional Atmospheric  
956 Chemistry Multiphase Mechanism (CRACMM), *Atmos. Chem. Phys.*, **23**, 5043-  
957 5099, doi: 10.5194/acp-23-5043-2023, 2023b.

958 Robinson, A. L., Donahue, N. M., Shrivastava, M. K., Weitkamp, E. A., Sage, A. M.,  
959 Grieshop, A. P., Lane, T. E., Pierce, J. R., and Pandis, S. N.: Rethinking Organic  
960 Aerosols: Semivolatile Emissions and Photochemical Aging, *Science*, **315**, 1259-  
961 1262, doi: 10.1126/science.1133061, 2007.

962 Sarwar, G., Luecken, D., Yarwood, G., Whitten, G. Z., Carter, W. P. J. J. o. a. m., and  
963 climatology: Impact of an updated carbon bond mechanism on predictions from the  
964 CMAQ modeling system: Preliminary assessment, **47**, 3-14, 2008.

965 Seinfeld, J. H., Pandis, S. N. J. A. c., and physics: From air pollution to climate change, **1326**,  
966 1998.

967 Skamarock, W. C., Klemp, J., Dudhia, J., Gill, D. O., Liu, Z., and Berner, J.: A Description  
968 of the Advanced Research WRF Model Version 4, doi: <https://doi.org/10.5065/1dfh-6p97>,  
969 2019.

970 Skipper, T. N., D'Ambro, E. L., Wiser, F. C., McNeill, V. F., Schwantes, R. H., Henderson, B.  
971 H., Piletic, I. R., Baublitz, C. B., Bash, J. O., Whitehill, A. R., Valin, L. C., Mouat, A. P.,  
972 Kaiser, J., Wolfe, G. M., St. Clair, J. M., Hanisco, T. F., Fried, A., Place, B. K., and Pye,  
973 H. O. T.: Role of chemical production and depositional losses on formaldehyde in the  
974 Community Regional Atmospheric Chemistry Multiphase Mechanism (CRACMM),  
975 *Atmospheric Chemistry and Physics*, **24**, 12903-12924, doi: 10.5194/acp-24-12903-2024,  
976 2024.

977 Stockwell, W. R., Lawson, C. V., Saunders, E., and Goliff, W. S.: A Review of Tropospheric  
978 Atmospheric Chemistry and Gas-Phase Chemical Mechanisms for Air Quality Modeling,  
979 *Atmosphere*, **3**, 1-32, doi: 10.3390/atmos3010001, 2012.

980 Su, Q. and Chen, Y.: Zenodo [data set], doi: 10.5281/zenodo.18222704, 2025.

981 Tsai, Y. G., Wang, J. Y., Yang, K. D., Yang, H. Y., Yeh, Y. P., Chang, Y. J., Lee, J. H., Wang, S.  
982 L., Huang, S. K., and Chan, C. C.: Long-term PM(2.5) exposure impairs lung growth and  
983 increases airway inflammation in Taiwanese school children, *ERJ open research*, **11**, doi:  
984 10.1183/23120541.00972-2024, 2025.

985 Wang, Y., Pavuluri, C. M., Fu, P., Li, P., Dong, Z., Xu, Z., Ren, H., Fan, Y., Li, L., Zhang, Y.-  
986 L., and Liu, C.-Q.: Characterization of Secondary Organic Aerosol Tracers over Tianjin,  
987 North China during Summer to Autumn, *ACS Earth and Space Chemistry*, **3**, 2339-  
988 2352, doi: 10.1021/acsearthspacechem.9b00170, 2019.

989 Wang, Y., Ning, M., Su, Q., Wang, L., Jiang, S., Feng, Y., Wu, W., Tang, Q., Hou, S., Bian, J.,  
990 Huang, L., Lu, G., Manomaiphiboon, K., Kaynak, B., Zhang, K., Chen, H., and Li, L.:  
991 Designing regional joint prevention and control schemes of PM<sub>2.5</sub> based on source  
992 apportionment of chemical transport model: A case study of a heavy pollution episode, *J.*  
993 *Clean. Prod.*, **455**, 142313, doi: 10.1016/j.jclepro.2024.142313, 2024.

994 Wong, D. C., Willison, J., Pleim, J. E., Sarwar, G., Beidler, J., Bullock, R., Herwehe, J. A.,  
995 Gilliam, R., Kang, D., Hogrefe, C., Pouliot, G., and Foroutan, H.: Development of the

996 MPAS-CMAQ coupled system (V1.0) for multiscale global air quality modeling, *Geosci.*  
997 *Model Dev.*, 17, 7855-7866,doi: 10.5194/gmd-17-7855-2024, 2024.

998 Woody, M. C., Baker, K. R., Hayes, P. L., Jimenez, J. L., Koo, B., and Pye, H. O. T.:  
999 Understanding sources of organic aerosol during CalNex-2010 using the CMAQ-VBS,  
1000 *Atmos. Chem. Phys.*, 16, 4081-4100,doi: 10.5194/acp-16-4081-2016, 2016.

1001 Xie, Y., Paulot, F., Carter, W. P. L., Nolte, C. G., Luecken, D. J., Hutzell, W. T., Wennberg, P.  
1002 O., Cohen, R. C., and Pinder, R. W.: Understanding the impact of recent advances in  
1003 isoprene photooxidation on simulations of regional air quality, *Atmos. Chem. Phys.*, 13,  
1004 8439-8455,doi: 10.5194/acp-13-8439-2013, 2013.

1005 Yarwood, G., Morris, R. E., and Wilson, G. M.: Particulate Matter Source Apportionment  
1006 Technology (PSAT) in the CAMx Photochemical Grid Model, *Air Pollution Modeling*  
1007 *and Its Application XVII*, Boston, MA, 2007//, 478-492,

1008 Yarwood, G., Rao, S., Yocke, M., and Whitten, G.: Updates to the carbon bond chemical  
1009 mechanism: CB05 final report to the US EPA, RT-0400675, 2005.

1010 Yarwood, G., Jung, J., Whitten, G. Z., Heo, G., Mellberg, J., and Estes, M.: UPDATES TO  
1011 THE CARBON BOND MECHANISM FOR VERSION 6 (CB6),

1012 Zhang, J., He, X., Ding, X., Yu, J. Z., and Ying, Q.: Modeling Secondary Organic Aerosol  
1013 Tracers and Tracer-to-SOA Ratios for Monoterpenes and Sesquiterpenes Using a  
1014 Chemical Transport Model, *Environmental Science & Technology*, 56, 804-813,doi:  
1015 10.1021/acs.est.1c06373, 2022.

1016 Zhang, J., He, X., Gao, Y., Zhu, S., Jing, S., Wang, H., Yu, J. Z., and Ying, Q.: Estimation of  
1017 Aromatic Secondary Organic Aerosol Using a Molecular Tracer—A Chemical Transport  
1018 Model Assessment, *Environ. Sci. Technol.*,doi: 10.1021/acs.est.1c03670, 2021a.

1019 Zhang, J., He, X., Gao, Y., Zhu, S., Jing, S., Wang, H., Yu, J. Z., and Ying, Q.: Assessing  
1020 Regional Model Predictions of Wintertime SOA from Aromatic Compounds and  
1021 Monoterpenes with Precursor-specific Tracers, *Aerosol and Air Quality Research*, 21,doi:  
1022 10.4209/aaqr.210233, 2021b.

1023 Zhang, J., Liu, J., Ding, X., He, X., Zhang, T., Zheng, M., Choi, M., Isaacman-VanWertz, G.,  
1024 Yee, L., Zhang, H., Misztal, P., Goldstein, A. H., Guenther, A. B., Budisulistiorini, S. H.,  
1025 Surratt, J. D., Stone, E. A., Shrivastava, M., Wu, D., Yu, J. Z., and Ying, Q.: New  
1026 formation and fate of Isoprene SOA markers revealed by field data-constrained modeling,  
1027 *npj Climate and Atmospheric Science*, 6, 69,doi: 10.1038/s41612-023-00394-3, 2023.

1028 Zhu, S., Wang, Z., Qu, K., Xu, J., Zhang, J., Yang, H., Wang, W., Sui, X., Wei, M., and Liu,  
1029 H.: Spatial Characteristics and Influence of Topography and Synoptic Systems on PM<sub>2.5</sub>  
1030 in the Eastern Monsoon Region of China, *Aerosol and Air Quality Research*, 23,  
1031 220393,doi: 10.4209/aaqr.220393, 2023.

1032