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2 Development of a Model Framework for Terrestrial Carbon Flux Prediction: the Regional Carbon  
3 and Climate Analytics Tool (RCCAT) Applied to Non-tidal Wetlands

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15

16

17 Abstract

18 Wetlands play a pivotal role in carbon sequestration but emit methane (CH<sub>4</sub>), creating  
19 uncertainty in their net climate impact. Although process-based models offer mechanistic  
20 insights into wetland dynamics, they are computationally expensive, uncertain, and difficult to  
21 upscale. In contrast, data-driven models provide a scalable alternative by leveraging extensive  
22 datasets to identify patterns and relationships, making them more adaptable for large-scale  
23 applications. However, their performance can vary significantly depending on the quality and  
24 representativeness of the data, as well as the model design, which raises questions about their  
25 reliability and generalizability in complex wetland systems. To address these issues, we present  
26 a data-driven framework for upscaling wetland CO<sub>2</sub> and CH<sub>4</sub> emissions, across a range of  
27 machine learning models that vary in complexity, validated against an extensive observational  
28 dataset from the Sacramento-San Joaquin Delta. We show that artificial intelligence (AI)  
29 approaches, including Random Forests, gradient boosting methods (XGBoost, LightGBM),  
30 Support Vector Machines (SVM) and Recurrent Neural Networks (GRU, LSTM), outperform  
31 linear regression models, with RNNs standing out, achieving an R<sup>2</sup> of 0.72 for daily CO<sub>2</sub> flux  
32 predictions compared to 0.62 for linear regression, and an R<sup>2</sup> of 0.60 for CH<sub>4</sub> flux predictions  
33 compared to 0.54 for linear regression. Interestingly, linear regression performed better than  
34 random forest for methane flux, which highlights the necessity for comparison. Despite that,  
35 interannual variability is less well captured, with annual mean absolute error of 193 gC m<sup>-2</sup> yr<sup>-1</sup>  
36 for CO<sub>2</sub> fluxes and 11 gC-CH<sub>4</sub> m<sup>-2</sup> yr<sup>-1</sup> for CH<sub>4</sub> fluxes. By integrating vertically-resolved  
37 atmospheric, subsurface, and spectral reflectance information from readily available sources,



the model identifies key drivers of wetland CO<sub>2</sub> and CH<sub>4</sub> emissions and enables regional upscaling. These findings demonstrate the potential of AI methods for upscaling, providing practical tools for wetland management and restoration planning to support climate mitigation efforts.

## 1. Introduction

Wetlands provide a wide array of ecological, economic, and environmental benefits (Costanza et al., 2014). They play a crucial role in biodiversity conservation, water purification, flood control, and climate regulation (Grande et al., 2023; Sharma and Singh, 2021). Significant attention has been recently given to wetland restoration due to their ability to sequester carbon from the atmosphere (Lolu et al., 2020; Upadhyay et al., 2020). These ecosystems are highly effective at storing carbon in their soils because the anaerobic conditions in waterlogged soils suppress organic matter decomposition, allowing carbon to accumulate over time (Mitsch and Gosselink, 2015a). However, wetlands can also be significant sources of CH<sub>4</sub>, a potent greenhouse gas (Brix et al., 2001), leading to potentially net positive effects of wetlands on climate warming. The most accurate way to determine the carbon balance in natural ecosystems is through direct and continuous measurements of carbon and GHG sources and sinks (Baldocchi et al., 2001). This involves monitoring carbon dynamics using techniques such as eddy covariance (EC) towers (Aubinet et al., 2012), soil carbon stock assessments (Harrison et al., 2011), and lateral carbon transport measurements (Ciais et al., 2008). However, these measurements are time-consuming to carry out, costly, and require specialized instruments and expertise, limiting their application to a few representative sites globally (Hill et al., 2017; Kumar et al., 2017). The Ameriflux network offers roughly 500 EC sites comprising about 3600 site years of data, monitoring carbon fluxes across various ecosystems such as forests, grasslands, and wetlands (Pastorello et al., 2020). Eddy-covariance site footprints range in scale and are typically determined by the sensor height and atmospheric turbulence (Chu et al., 2021). Data from these Ameriflux sites could potentially be upscaled and used for estimating fluxes from non-monitored sites to obtain regional assessments of carbon balance for various ecosystem types, including wetlands.

In this study, we focus on nontidal wetlands due to the presence of a cluster of EC towers in a small region located in the Sacramento-San Joaquin Delta, including three sites, each with over a decade of continuous data. Reported sequestration rates in wetlands vary widely, influenced by factors such as climate, vegetation, and management. For instance, reported sequestration rates range from as low as 26 gC m<sup>-2</sup> yr<sup>-1</sup> in boreal rain-fed bogs (Villa and Bernal, 2018) to as high as 797 gC m<sup>-2</sup> yr<sup>-1</sup> in constructed wetlands with emergent *Phragmites* in the Netherlands (de Klein and van der Werf, 2014). Similarly, temperate wetlands in central Ohio exhibit a wide range of carbon sequestration rates depending on vegetation: forested depressional wetlands dominated by *Quercus palustris* sequester up to 473 gC m<sup>-2</sup> yr<sup>-1</sup>, while marshes dominated by *Typha* sequester around 210 gC m<sup>-2</sup> yr<sup>-1</sup> (Bernal and Mitsch, 2012). In Victoria, Australia, freshwater marshes show varying sequestration rates from 91 gC m<sup>-2</sup> yr<sup>-1</sup> in shallow marshes to 230 gC m<sup>-2</sup> yr<sup>-1</sup> in permanent open freshwater wetlands (Carnell et al., 2018). More relevant to



79 this study, in the San Francisco Bay-Delta region, nontidal managed wetlands dominated by  
80 *Schoenoplectus* and *Typha* species sequester carbon at rates of approximately  $355 \pm 249$  gC-  
81  $\text{CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$ . This estimate is based on direct calculations using Ameriflux data from sites with  
82 over a decade of observations (US-Myb, US-Tw1, and US-Tw4). For this calculation we used  
83 full-year annual averages and their corresponding standard deviation to the annual mean, to  
84 highlight the significant inter-annual variability, with the standard deviation close to the mean.  
85 The unit reported for these Delta sites is in  $\text{gC-CO}_2 \text{ m}^{-2} \text{ yr}^{-1}$ , as the EC tower directly detects  
86  $\text{CO}_2$  exchange, which is convenient for GHG assessment purposes. It is worth noting that, at  
87 these sites, some years were a net  $\text{CO}_2$  source, due to site-specific disturbances such as  
88 caterpillar infestations, drought, or when vegetation cover was fully established (Anderson et al.,  
89 2018; Knox et al., 2017; Rey-Sanchez et al., 2021). See table S1 for more detailed information  
90 and references therein.

91 Although  $\text{CO}_2$  balance (photosynthesis minus community respiration) is an important component  
92 of carbon sequestration, in many wetland systems sequestration benefits are counterbalanced  
93 by  $\text{CH}_4$  emissions, a potent greenhouse gas, with a warming potential 27 times higher than  $\text{CO}_2$   
94 (Lee et al., 2023) that can often offset climate mitigation efforts.  $\text{CH}_4$  emission rates also vary  
95 substantially over time and across wetlands, from as low as  $0.23 \text{ gC-CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  in saltwater  
96 zones of estuarine environments (Abril and Iversen, 2002) to as high as  $270 \text{ gC-CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  in  
97 certain freshwater wetlands (Knox et al., 2021). For example, restored freshwater wetlands in  
98 Maryland dominated by grasses and sedges emit around  $142 \text{ gC-CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  (Stewart et al.,  
99 2024). Tropical wetlands in Costa Rica exhibit some of the highest emissions, with isolated and  
100 floodplain wetlands releasing between 220 and  $263 \text{ gC-CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  (Mitsch et al., 2013). The  
101 San Francisco Bay-Delta wetlands that have high carbon sequestration rates also release  $\text{CH}_4$   
102 at rates of  $35 \pm 13 \text{ gC-CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  (direct measurements from the eddy covariance tower data  
103 (Arias-Ortiz et al., 2021)). See table S2 for further information and reference therein. This dual  
104 role of wetlands in both sequestering carbon and emitting  $\text{CH}_4$  reveals the complex effect they  
105 have on the global greenhouse gas balance. Therefore, integrating  $\text{CO}_2$  and  $\text{CH}_4$  emissions is  
106 critical to assess the net climate benefits of wetland conservation and restoration initiatives.

107 To evaluate how wetlands contribute to the atmospheric radiation budget at larger scales, it is  
108 essential to quantify both GHG emissions and carbon sequestration, especially at sites where  
109 direct measurements are unavailable (Moomaw et al., 2018). Upscaling models serve this  
110 purpose by allowing estimation of sequestration and emission rates across larger spatial scales  
111 than those covered by the original data sources (Villa and Bernal, 2018) which provide GHG  
112 accounting and net climate benefit assessments for specific wetland sites (Nahlik and  
113 Fennessy, 2016). Moreover, it aids in targeting wetland restoration efforts that aim to optimize  
114 sequestration by identifying locations with the greatest potential for net carbon uptake.

115 Process-based models have traditionally been used to estimate sequestration and emissions  
116 (Mack et al., 2023; Zhang et al., 2002). Models such as DNDC (Li, 1996), DayCent (Parton et  
117 al., 1998), and *Ecosys* (Grant et al., 2017) have been applied to simulate biogeochemical  
118 processes in terrestrial ecosystems, including modeling  $\text{CH}_4$  emissions, carbon balances, and  
119 soil carbon and nitrogen cycling (Grant and Roulet, 2002; Weiler et al., 2018; Zhang et al.,  
120 2002). While these models can elucidate the processes that play a role in carbon dynamics,



121 they require extensive mechanistic parameterization to accurately represent the interactions in  
122 various ecosystems(Pastorello et al., 2020; Yin et al., 2023). This approach often necessitates  
123 site-specific information and data collection, making implementation over vast areas challenging  
124 (Saunois et al., 2024; Xu and Trugman, 2021). The extensive data needs associated with these  
125 process-rich models showcase the need for alternative approaches that can effectively upscale  
126 wetland emissions without such intensive resource demands.

127 Artificial Intelligence (AI) methods, such as machine learning and deep learning, have been  
128 widely applied in ecological modeling in recent years, alongside long-term, large-scale data  
129 collection efforts (Perry et al., 2022). Recent deep learning applications have demonstrated  
130 success in capturing the complex dynamics of carbon and methane fluxes in these systems  
131 (Ouyang et al., 2023; Yuan et al., 2022, 2024; Zou et al., 2024). The availability of open-source  
132 modeling platforms like TensorFlow and PyTorch has made advanced computational  
133 techniques, such as neural networks, more accessible, enabling the rapid development and  
134 deployment of a range of specialized modeling tasks (Xu et al., 2021). Despite several recent  
135 studies demonstrating the potential of machine learning for large-scale carbon cycling in  
136 wetland ecosystems, this remains a relatively young field. Moreover, carbon dynamics in  
137 wetland ecosystems are temporally variable and inherently nonlinear, making them particularly  
138 well-suited for testing machine learning approaches (Arora et al., 2019, 2022). We therefore  
139 emphasize the importance of evaluating and comparing various approaches within this domain  
140 and their potential for large-scale assessment.

141 A pervasive challenge in model development is the ability to balance complexity with  
142 generalizability. While more complex models can capture nonlinear relationships, they also  
143 increase the risk of overfitting, where the model performs well in the testing, but poorly on new  
144 conditions (Hastie, 2009; Tashman, 2000). Furthermore, it is also important to use a robust  
145 validation framework. For the application of upscaling, it is important that the model is able to  
146 extrapolate spatially. For this purpose, a leave-one-site-out (LOSO) validation approach is  
147 typically carried out, whereby the models are trained on data that excludes a single site, with the  
148 excluded site data saved for model testing (Bodesheim et al., 2018; Tramontana et al., 2016). It  
149 is also important to avoid data leakage, where information from the training set inadvertently  
150 appears in the testing set (Kaufman et al., 2012), a risk posed when splitting temporally  
151 adjacent data points that are close in value, potentially inflating performance statistics (Bergmeir  
152 and Benítez, 2012; Kaufman et al., 2012). For example, daily rates of change relative to a  
153 system where seasonal dynamics dominate, such as emissions of CH<sub>4</sub> emissions in vegetated  
154 wetlands (Knox et al., 2021).

155 In this study, we introduce a model framework for coastal nontidal wetland CO<sub>2</sub> and CH<sub>4</sub>  
156 emissions using several ‘off-the-shelf’ models. These models are trained and validated against  
157 observational data, and results are compared to find the most predictive model. The top  
158 performing model is then used to upscale carbon sequestration and CH<sub>4</sub> emissions in nontidal  
159 wetlands at regional scale. The San Francisco Bay-Delta serves as the area of interest, due to  
160 its network of EC towers that have been operating for a relatively long time and relevance to  
161 future wetland restoration efforts. We employ a suite of models, ranging widely in complexity: (1)  
162 linear regression; (2) Random Forests (Breiman, 2001), an ensemble method that constructs



multiple decision trees to reduce overfitting; (3) gradient boosting techniques such as LightGBM (Ke et al., 2017) and XGBoost (Chen and Guestrin, 2016), which are scalable tree boosting systems able to handle complex nonlinear relationships and variable interactions; (4) Support Vector Machines (SVM) (Cortes, 1995), a kernel-based technique that can approximate nonlinear boundaries between data points and (5) the Recurrent Neural Network (RNN) such as the Long Short-Term Memory (LSTM) neural network (Hochreiter, 1997), an advanced model designed to process sequential data and capture non-linear interactions over long-term dependencies. We also test a model with similar but simpler architecture, the Gated Recurrent Unit (GRU) (Chung et al., 2014), which uses fewer parameters. Linear regressions serve as a baseline to assess the applicability of the more sophisticated methods. Random Forests have been used to upscale northern wetland methane emissions (Peltola et al., 2019), gradient boosting methods have demonstrated success in ecological modeling (Ding, 2024; Räsänen et al., 2021; Zou et al., 2024), and LSTM neural networks have been successfully applied to model CO<sub>2</sub> and CH<sub>4</sub> fluxes in ecosystems (Yuan et al., 2022, 2024; Zou et al., 2024). Our proposed framework is designed to provide transparency, easy determination of model practicality and applicability, and contextualisation to model performances by comparing to a baseline model (i.e. linear regression).

180

## 181 2. Methods

Our ultimate aim is to establish a robust modeling framework for estimating wetland carbon fluxes in sites that are not monitored. To achieve this, we compare a range of models, from simple linear regression to advanced recurrent machine learning neural networks. Since the goal is to predict unseen sites, we emphasize cross-site predictability by validating and testing the models at sites not included in training. Doing so ensures predictions are applicable beyond the training sites and addresses challenges often associated with model generalizability (Meyer and Pebesma, 2022). This strategy serves several purposes:

1. **Performance Contextualization:** Starting with the simplest type of model provides a baseline for performance and helps evaluate the advantage (or lack thereof) for using more complex models.
2. **Practicality and Transparency:** Advanced models may offer better performance but often require significant effort to set up and may lack interpretability. By comparing models of varying complexity using the same input data, we assess whether the added complexity is justified.
3. **Feature Evaluation:** Training with different combinations of relevant features helps us to understand which features are dominating control, and the limitations of the data in terms of predictive capacity.

### 199 2.1 Model targets

The model targets two key variables: CO<sub>2</sub> (**FCO2**) and CH<sub>4</sub> (**FCH4**) surface emissions. Both variables follow a sign convention where positive values indicate emissions to the atmosphere



(source) and negative values indicate sequestration (sink). Both variables are available at half-hourly resolution through the Ameriflux database.

The models we developed all operate on a daily time scale, requiring target variables to be aggregated to the daily time scale. This approach assumes that sub-daily variations have a negligible non-linear contribution to longer time scales, an assumption supported by the dominant seasonal signal typically observed in flux data from these systems (Knox et al., 2021).

These target variables could then be used to calculate annual NECB (Net Ecosystem Carbon Balance;  $\text{gC m}^{-2} \text{yr}^{-1}$ ) and annual wetland net atmospheric radiative effect ( $\text{FCO}_2\text{e}$  ( $\text{CO}_2$ -equivalent flux)  $\text{gCO}_2\text{e m}^{-2} \text{yr}^{-1}$ ). The global warming potential (GWP) of non-fossil  $\text{CH}_4$  is 27.2 as per the latest IPCC assessment (Lee et al., 2023). For this study, we neglect contributions of lateral fluxes due to data limitations, and that lateral transport at these sites is assumed to be negligible due to the limited outflow from the wetlands (Miller et al., 2008).  $\text{FCO}_2\text{e}$  is defined as annually averaged  $\text{CO}_2$  and  $\text{CH}_4$  emissions, adjusted for the global warming potential (GWP) of each gas. A positive  $\text{FCO}_2\text{e}$  indicates that the ecosystem is contributing positively to atmospheric warming, and vice versa. Here we consider  $\text{CO}_2$  and  $\text{CH}_4$  emissions but neglect contributions from  $\text{N}_2\text{O}$  due to data limitations and because  $\text{N}_2\text{O}$  emissions are considered negligible in Delta wetlands (Windham-Myers et al., 2018).

## 2.2 Region of interest

The Sacramento-San Joaquin Delta was selected for this study due to its high density of EC towers and extensive long-term data. We selected sites for model training and validation where data was collected for at least a decade to capture interannual variability. Hence three restored wetland sites, US-Myb (Matthes et al., 2016), US-Tw1 (Valach et al., 2016), and US-Tw4 (Eichelmann et al., 2016) are selected in this study. While data from two other sites (i.e., US-Sne and US-Tw5) are available, the lack of sufficient temporal coverage and, in the case of US-Sne, not fully established vegetation cover, makes them less representative of a stable ecosystem. Focusing on sites with over a decade of continuous data allows for capturing long-term dynamics more effectively and provides sufficient time for the wetlands to reach a stable state. The dataset encompasses 35 full site-years of observations across the three sites within the Delta (Novick et al., 2018) (Table 2, Figure 1), with detailed mapping data sourced from the EcoAtlas Database (Workgroup, 2019) which provides land use and vegetation surveys across wetlands in California.

Table 2: Model training sites

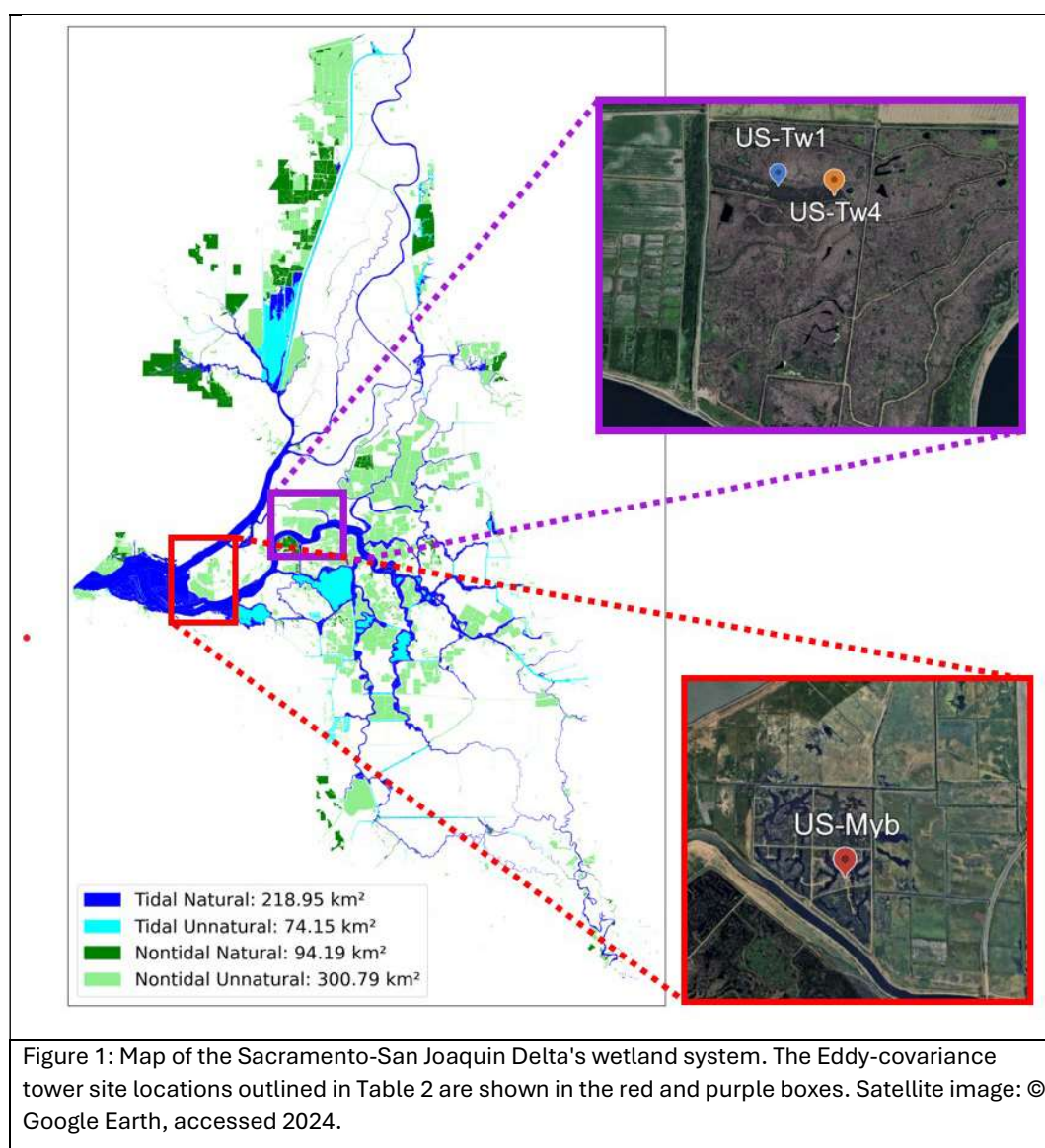
| Site Code | Site Name                         | Water Type | Salinity | Years of Data (Full) | Start Date |
|-----------|-----------------------------------|------------|----------|----------------------|------------|
| US-Myb    | Mayberry Wetland                  | Non-Tidal  | Fresh    | 13                   | 2010       |
| US-Tw1    | Twitchell Wetland West Pond       | Non-Tidal  | Fresh    | 12                   | 2011       |
| US-Tw4    | Twitchell Island East End Wetland | Non-Tidal  | Fresh    | 10                   | 2013       |

233



234 The sites are dominated by Tules (*Schoenoplectus*), Cattails (*Typha*), and invasive species such as  
235 *Phragmites*, which are perennial emergent plants well suited to wetland environments (López et  
236 al., 2016). The Delta itself is host to the largest estuarine system on the US Pacific coast, spanning  
237 approximately 3,000 km<sup>2</sup>, and contains a diverse network of wetland systems. Historically, much of  
238 the area was drained and converted for agriculture (Laćan and Resh, 2016; Lund et al., 2010), but  
239 recent restoration efforts have reclaimed select portions of the landscape for environmental  
240 benefits.

241





242

## 243 **2.3 Model features**

244 The application of this work focuses on upscaling carbon fluxes from similar wetlands at a regional  
245 scale. To achieve this, we aim to predict fluxes at unmonitored sites using widely available data  
246 that are expected to be key drivers of FCO<sub>2</sub> and FCH<sub>4</sub>. Since site-level measurements from EC  
247 towers are not available at a larger spatial scale, we focus on ecosystem drivers that can be  
248 accessed across broader spatial extents.

249 The models utilize a comprehensive set of features from two readily accessible datasets: the  
250 Western Land Data Assimilation System (WLDAS) (Erlingis et al., 2021) and satellite-derived  
251 products from MODIS (Justice et al., 2002) (Supplementary Table S3). WLDAS provides high-  
252 resolution hydrological and meteorological data at 1-km spatial and daily temporal resolutions,  
253 spanning from 1980 to the present. Key variables include soil moisture, soil temperature,  
254 precipitation, solar radiation, and water table depth. MODIS complements these inputs with  
255 remote sensing data at a spatial resolution of 250–500 meters and temporal intervals ranging  
256 from 4 to 16 days, providing vegetation indices such as the Normalized Difference Vegetation  
257 Index (NDVI) and Leaf Area Index (LAI). The broad spatial and temporal coverage of these  
258 datasets enables upscaling across various regions. By relying on publicly available data  
259 sources, the framework remains practical and adaptable, facilitating rapid implementation with  
260 the appropriate training data.

261

## 262 **2.4 Model suite**

263 To evaluate ML model performance in calculating FCO<sub>2</sub> and FCH<sub>4</sub>, we implemented a suite of  
264 seven models ranging from simple linear methods to more complex neural networks. These  
265 models have been used in various ecosystems to study fluxes and collectively represent a broad  
266 spectrum of methodological complexity. Table 3 summarizes the core characteristics and  
267 advantages of each approach.

268 Table 3: An overview of the models that are applied to wetland fluxes

| Model Name                    | Category                   | Description                                                        | Key Strengths                                              |
|-------------------------------|----------------------------|--------------------------------------------------------------------|------------------------------------------------------------|
| Linear Regression             | Regression                 | Fits a linear relationship between predictors and fluxes           | Simple baseline, easily interpretable (Breiman, 2001)      |
| Random Forest (Breiman, 2001) | Ensemble of Decision Trees | Aggregates multiple decision trees to enhance prediction stability | Robust to nonlinearity, reduces overfitting (Cortes, 1995) |



|                                             |                          |                                                                  |                                                                   |
|---------------------------------------------|--------------------------|------------------------------------------------------------------|-------------------------------------------------------------------|
| Support Vector Machine (Cortes, 1995) (SVM) | Kernel-Based Method      | Uses flexible kernels to find optimal separating hyperplanes     | Effective in high dimensions, adaptable kernels (Ke et al., 2017) |
| LightGBM (Ke et al., 2017)                  | Gradient Boosting        | Employs iterative boosting with efficient tree growth            | Fast, memory-efficient, handles large datasets                    |
| XGBoost (Chen and Guestrin, 2016)           | Gradient Boosting        | Improves boosting with regularization and efficient computations | Manages outliers, handles sparse data well                        |
| LSTM Neural Network (Hochreiter, 1997)      | Recurrent Neural Network | Captures temporal dependencies in sequential data inputs         | Ideal for time-series, learns long-term patterns                  |
| GRU Neural Network (Chung et al., 2014)     | Recurrent Neural Network | Similar to LSTM but streamlined with fewer parameters            | Efficient temporal modeling, lower complexity                     |

269

270 These models act to demonstrate a spectrum of model complexity and how that can be leveraged  
271 to improve flux prediction.

272

## 273 **2.5 Validation framework**

274 To evaluate the models' ability to generalize across sites, we employed a Leave-One-Site-Out  
275 (LOSO) cross-validation strategy. In LOSO, we train the models on data from all but one site, and  
276 test the models on the excluded site. This approach is repeated for each site in the dataset and  
277 then aggregated, ensuring that there are no spatio-temporal connections between the training and  
278 testing data. While few models are immune to overfitting, this approach minimizes the risk of doing  
279 so.

280 An integral part of our modeling approach is the strategic selection of input features to optimize the  
281 model's performance. We perform this selection by first selecting features that are expected to be  
282 important, guided by mechanistic considerations of wetland processes gained from fieldwork and  
283 insights from mechanistic models (Table S3). Since the total number of possible feature  
284 combinations is too large for an exhaustive search, we adopt a feed-forward selection (**FFS**)  
285 strategy. This method begins with a single feature and iteratively adds features that most improves  
286 the model's performance based on a chosen statistic. At each step, we evaluate the model's  
287 performance with each potential new feature and select the one that provides the greatest



improvement. This process continues until adding additional features no longer significantly enhances the model's performance. By using this approach, we efficiently identify the most influential predictors without the computational burden of testing all possible combinations.

## 2.6. Validation

As suggested above, each model was trained using data from two wetland sites and then validated on the third. Although the number of sites was limited, each site offered over a decade of observations accumulated to a daily time step, ensuring exposure to a range of environmental conditions representative of the wetland type and regional climate. For each excluded site, the model's predictions were compared against measured FCO<sub>2</sub> and FCH<sub>4</sub>. We aggregated performance metrics ( $R^2$ , Pearson correlation coefficient ( $r$ ), and RMSE) across all site predictions. This process was paired with the FFS method optimized to maximize  $R^2$ .

## 3. Results

### 3.1 Model Validation

We tested six modeling techniques of varying complexities (Table 3). Model performance scores for daily predictions are shown in Figure 2, demonstrating that nearly all machine learning models outperformed the linear regression baseline ( $R^2 = 0.62$  for FCO<sub>2</sub> and  $R^2 = 0.54$  for FCH<sub>4</sub>). For FCO<sub>2</sub>, LSTM and GRU achieved the highest  $R^2$  values (0.71 and 0.70, respectively), outperforming other methods. A similar result was found for FCH<sub>4</sub>, with LSTM and GRU both scoring  $R^2$  of 0.60. These results suggest that deep learning models can provide tangible benefits over linear regression methods for upscaling flux predictions. The LSTM model was selected for upscaling in this study as it scored highest consistently, though other ML models scored comparably, so we do not assert it as definitively the best model.

The feature selection process had access to 26 environmental features from WLDAS and 8 features from MODIS (see table S3 for full details). These variables encompass a wide range of atmospheric, soil, and vegetation characteristics, such as precipitation, temperature, soil moisture, and spectral indices, key environmental drivers known to influence carbon and methane flux dynamics (Mitsch and Gosselink, 2015b). Results revealed that model performance plateaued after including 4 features, meaning that only a small subset of predictors were needed to maximize predictive skill. Notably all ML models, regardless of complexity, selected the same initial feature for both FCO<sub>2</sub> and FCH<sub>4</sub>, which highlights temperature as a dominant environmental driver. As shown in Table 4 for the LSTM case, temperature-related variables emerged as the first selected feature for both FCO<sub>2</sub> (Air Temperature) and FCH<sub>4</sub> (Canopy Temperature) predictions, illustrating the importance of temperature in driving ecosystem activity. As an aside, while temperature variables may be most significant in the first step, many other features scored comparably. The subsequent inclusion of hydrological and



spectral reflectance variables led to modest improvements in  $R^2$  values, which is also mirrored by improvements in RMSE and  $r$ .

Table 4: Feed-forward feature selection process.

| Target Variable | Step | Chosen Feature            | $R^2$ | RMSE  | $r$  |
|-----------------|------|---------------------------|-------|-------|------|
| FCO2            | 1    | Air Temperature           | 0.66  | 1.62  | 0.82 |
| FCO2            | 2    | Water Table Depth         | 0.68  | 1.57  | 0.83 |
| FCO2            | 3    | Bare Soil Evaporation     | 0.70  | 1.54  | 0.84 |
| FCO2            | 4    | Blue Reflectance          | 0.71  | 1.50  | 0.85 |
| FCH4            | 1    | Canopy Temperature        | 0.52  | 0.054 | 0.73 |
| FCH4            | 2    | Near-Infrared Reflectance | 0.58  | 0.051 | 0.76 |
| FCH4            | 3    | Total Evapotranspiration  | 0.60  | 0.050 | 0.77 |

Figure 3 shows both FCO2 and FCH4 results, including time series and scatter plots comparing predictions to observations. Overall, the predicted values track the observations reasonably well. For FCO2, predictions tended to regress toward the mean, underestimating peak emissions at local maxima and overestimating at local minima. The ML models also displayed less interannual variability than the observations, common in machine learning approaches (Ouyang et al., 2023). For wetlands, this is likely due to limited subsurface process information included in the machine learning models. Still, the scatter plot shows strong performance for FCO2 ( $r = 0.84$ ,  $R^2 = 0.71$ ,  $RMSE = 1.49 \text{ gC-CO}_2 \text{ m}^{-2} \text{ day}^{-1}$ ), despite a noticeable spread around the 1:1 line.

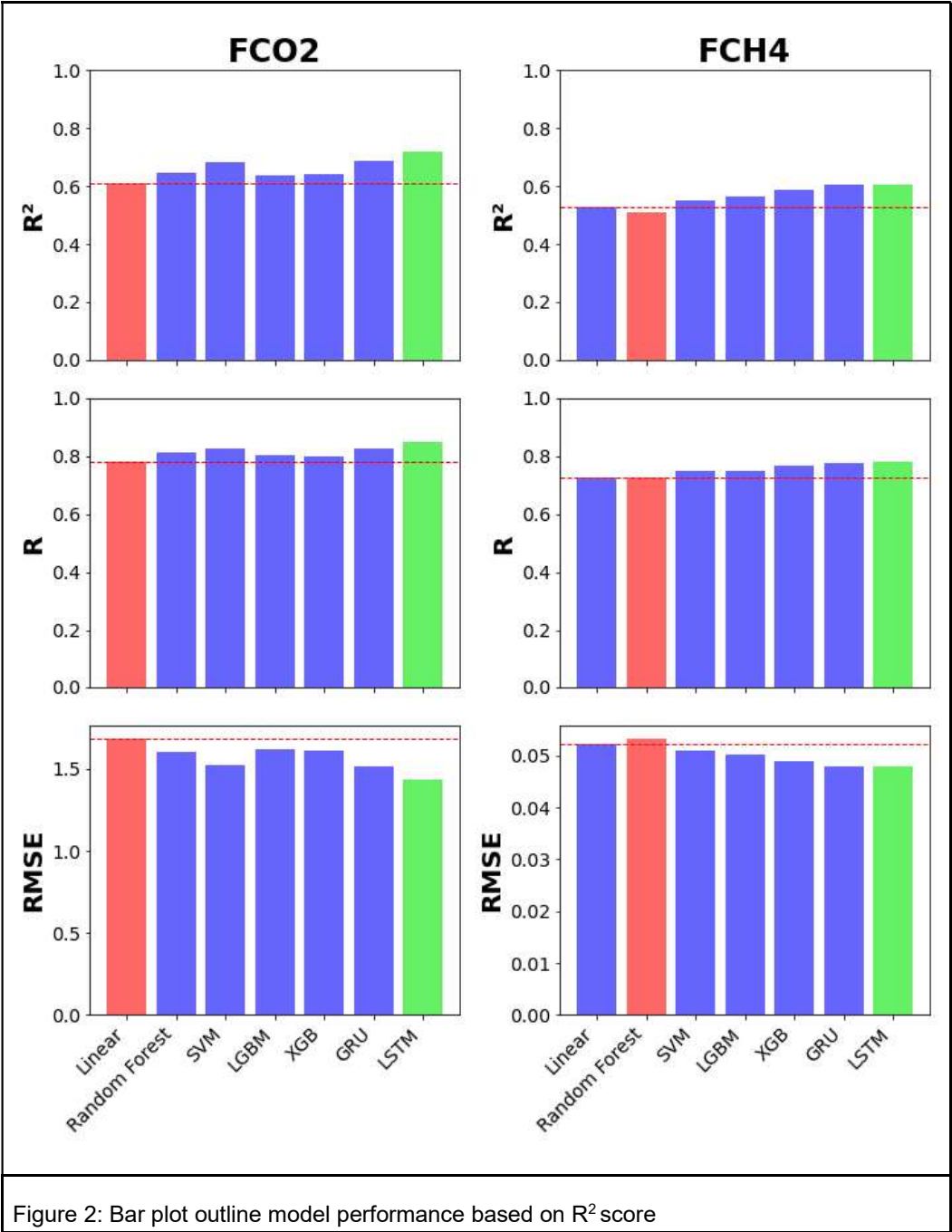
FCH4 predictions exhibited similar behavior, with lower interannual variability than the observations. At the US-Myb site, for example, observed FCH4 were initially high (aside from the first year, when vegetation cover had yet to be fully established) but declined over time, stabilizing at lower values. The ML models captured this shift to some extent, predicting higher fluxes early in the time series and then modulating to lower levels later on. However, predictions did not fully replicate the magnitude of the observed downward annual trend, introducing bias



345 into the scatter plots at higher and lower extreme values. This phenomenon is known as  
346 regression to the mean, observed in similar machine learning studies (Ouyang et al., 2023).  
347 Consequently, the FCH<sub>4</sub> model performance was weaker than the FCO<sub>2</sub> model ( $R^2 = 0.61$ ,  $r =$   
348  $0.78$ ,  $RMSE = 0.05 \text{ g C-CH}_4 \text{ m}^{-2} \text{ day}^{-1}$ ), indicating that the processes controlling FCH<sub>4</sub> in  
349 younger wetlands like US-Myb may require more detailed subsurface information (such as soil  
350 organic C, oxygen, or redox information) to be accurately modeled.

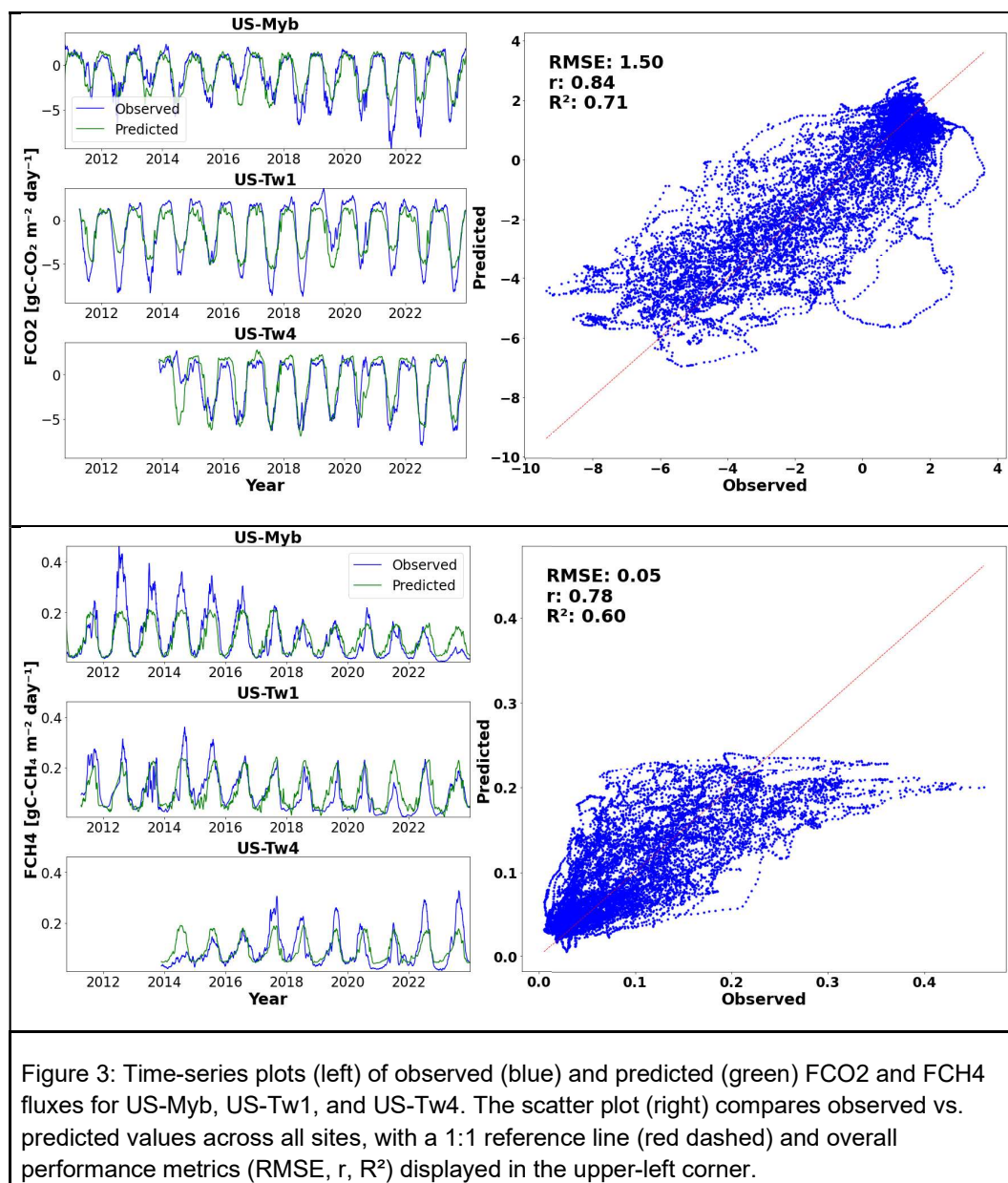
351 The annual bar plots presented in Figure 4 highlight the model's difficulty in capturing the  
352 interannual variability of carbon fluxes across the study sites. While the average FCO<sub>2</sub> and  
353 FCH<sub>4</sub> predictions are generally aligned with observed average values with small overall mean  
354 bias, the model struggles to reproduce the observed year-to-year variability. Although direct  
355 subsurface measurements are available at certain sites, at the regional scale their limited spatial  
356 and temporal coverage currently limits integration into models designed for regional upscaling  
357 over inter-annual timescale. For example, while spatial maps of wetland soil organic carbon  
358 exist (Uhran et al., 2022), using only three sites for training purposes would provide just three  
359 corresponding data points, limiting model training. The LOSO validation approach revealed that  
360 deep learning models, particularly LSTM and GRU, consistently outperformed traditional linear  
361 regression and other machine learning methods for both FCO<sub>2</sub> and FCH<sub>4</sub> predictions. While  
362 nonlinear models demonstrated clear advantages, the magnitude of improvement was relatively  
363 modest, reflecting the inherent challenges of capturing site-specific inter-annual dynamics of  
364 wetland emissions. To improve model performance, additional techniques such as feature  
365 transformations or attention mechanisms could be implemented. However, the primary goal of  
366 this model suite is to ensure reproducible results with 'off-the-shelf' models, which serves as a  
367 foundation for more advanced, nuanced approaches.

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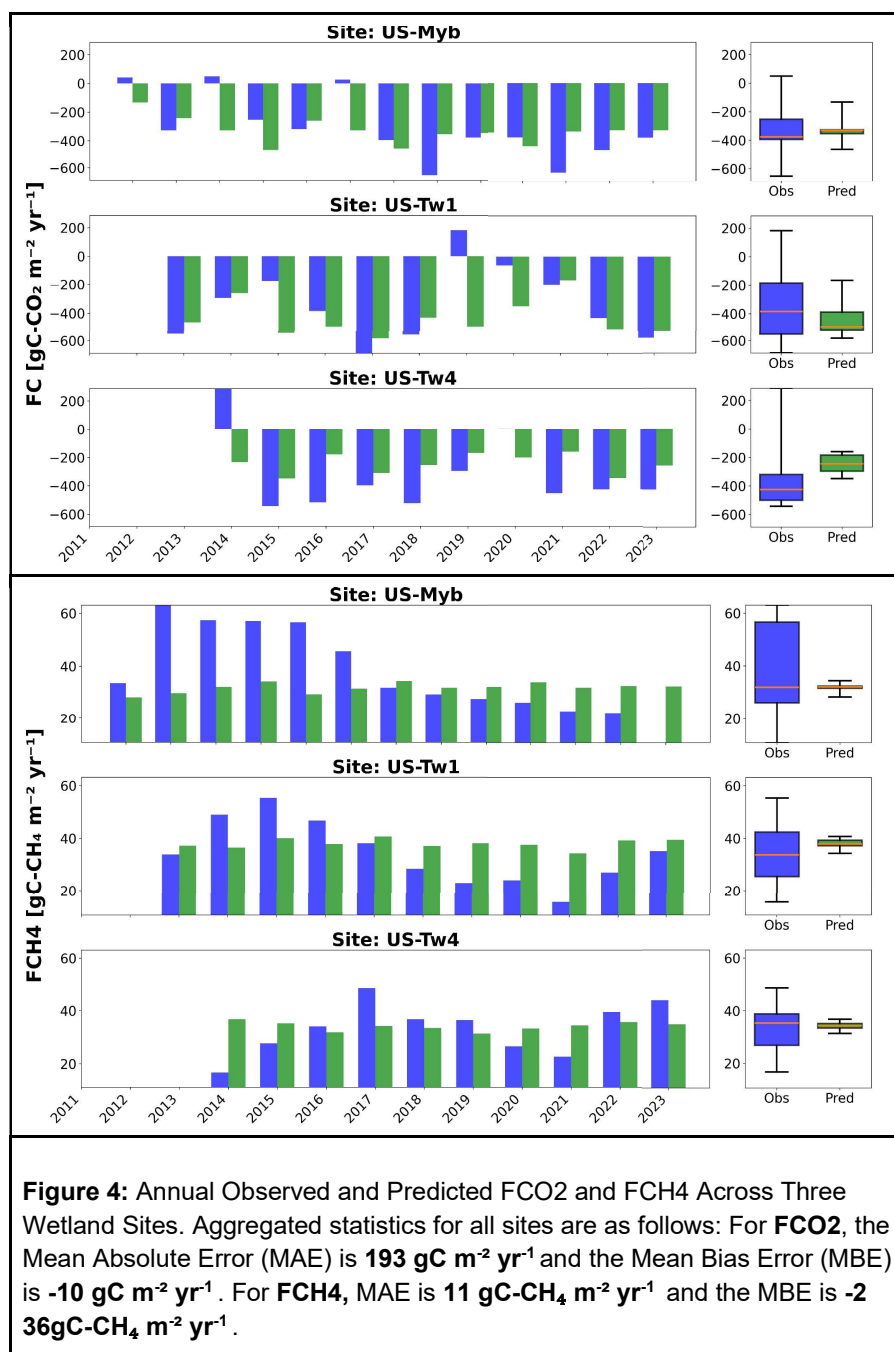
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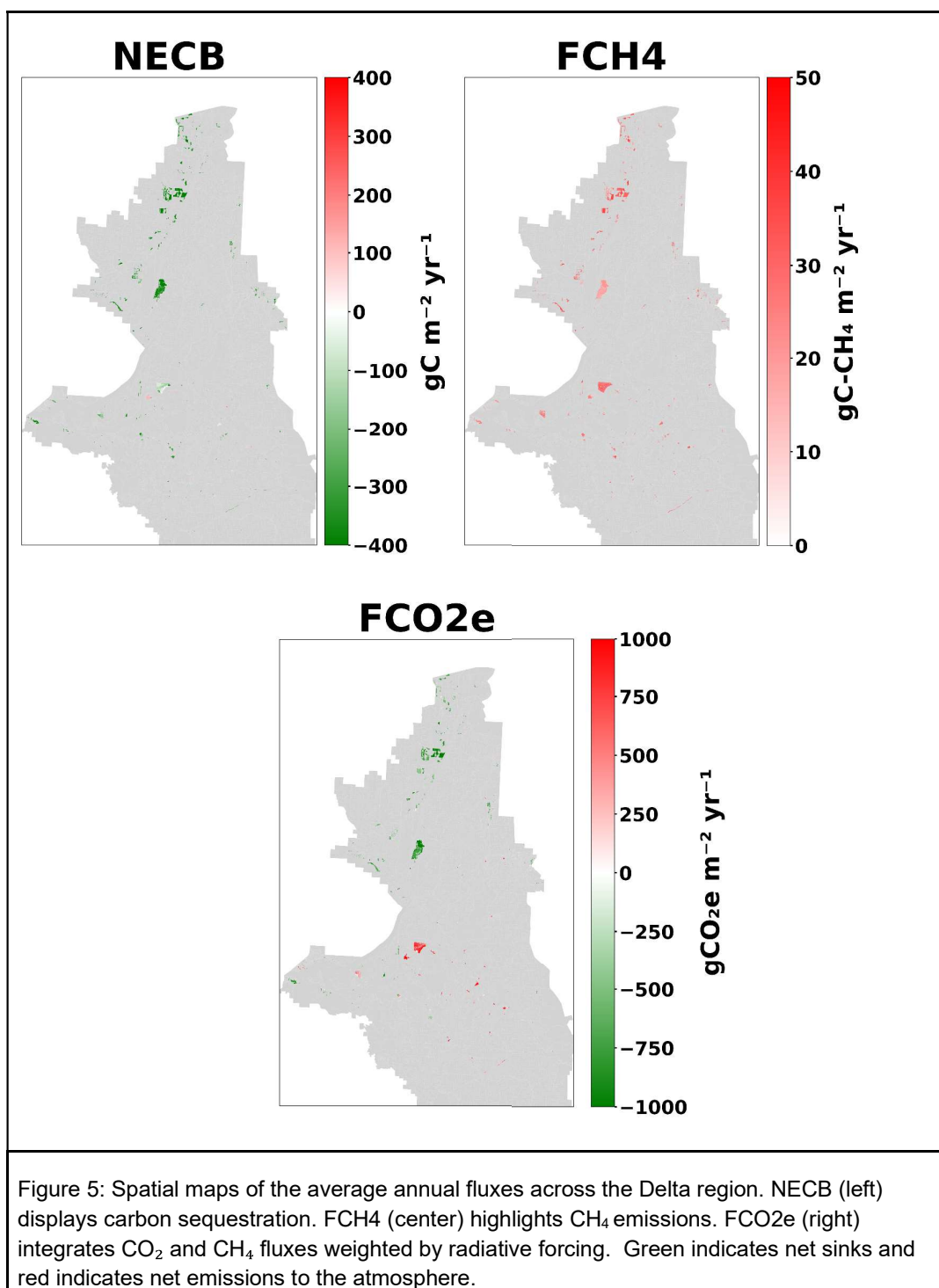
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### 380 **3.2. Model Application: Upscaling**

381 After selecting LSTM as the model of choice, it was retrained using all available data from the  
382 three sites for upscaling.. The Sacramento-San Joaquin Delta contains roughly 700km<sup>2</sup> of  
383 wetland area, including tidal and nontidal regions. The upscaling domain encompasses  
384 approximately 25 km<sup>2</sup> of nontidal wetlands in the region, dominated by vegetation types relevant  
385 to the training sites, specifically Tules, Cattails, and Phragmites. The assumption is that the  
386 training sites used in this study are representative of the broader conditions in the Delta, but we  
387 acknowledge that local variability in carbon dynamics, such as those caused by microclimates  
388 prevalent in the area, may not be fully captured during the ML model training. Improvements to  
389 the model might be achieved if additional site data covering a wider range of environmental  
390 conditions were incorporated. The feature data used to optimize the model were spatially  
391 interpolated onto the regional model grid and the model applied to yield flux estimations.  
392 Although relatively modest in spatial extent, these wetlands are of particular interest given their  
393 role in carbon sequestration and potential climate mitigation and as targets for conservation and  
394 restoration.

395 Figure 5 displays spatial maps of annual flux estimates of Net Ecosystem Carbon Balance  
396 (NECB), methane flux (FCH<sub>4</sub>), and the CO<sub>2</sub> equivalent flux rate (FCO<sub>2e</sub>) in the study domain .  
397 The results show that carbon sequestration, indicated by negative NECB (green) values, are  
398 notably stronger in the more northern parts of the domain, which is likely driven by air  
399 temperature (dominant feature), with this region of California being subject to distinctive  
400 microclimates. In contrast, FCH<sub>4</sub> shows a comparatively uniform spatial pattern, which was also  
401 observed in the model validation. Similar to NECB, the FCO<sub>2e</sub> distribution shows strong  
402 latitudinal dependence, with a net CO<sub>2e</sub> sink in northern zones and a tendency toward  
403 emissions in the southern portion of the study area, though localized heterogeneity exists.

404 Figure 6 shows averaged fluxes in the upscaling domain over the full study period. The results  
405 highlight the Delta as an overall carbon sink, with NECB averaging approximately -380 gC m<sup>-2</sup>  
406 yr<sup>-1</sup>, indicating persistent sequestration across multiple years. CH<sub>4</sub> fluxes average 28 gC-CH<sub>4</sub>  
407 m<sup>-2</sup> yr<sup>-1</sup>, and shows little spatial variability. Values are consistent with those previously reported  
408 in the region (Arias-Ortiz et al., 2021). Integrating these fluxes into a CO<sub>2</sub>-equivalent metric, this  
409 regional wetland system remains a net sink of CO<sub>2e</sub>, with approximately 400 gCO<sub>2e</sub> m<sup>-2</sup> yr<sup>-1</sup>  
410 sequestered on average in the upscaling domain.



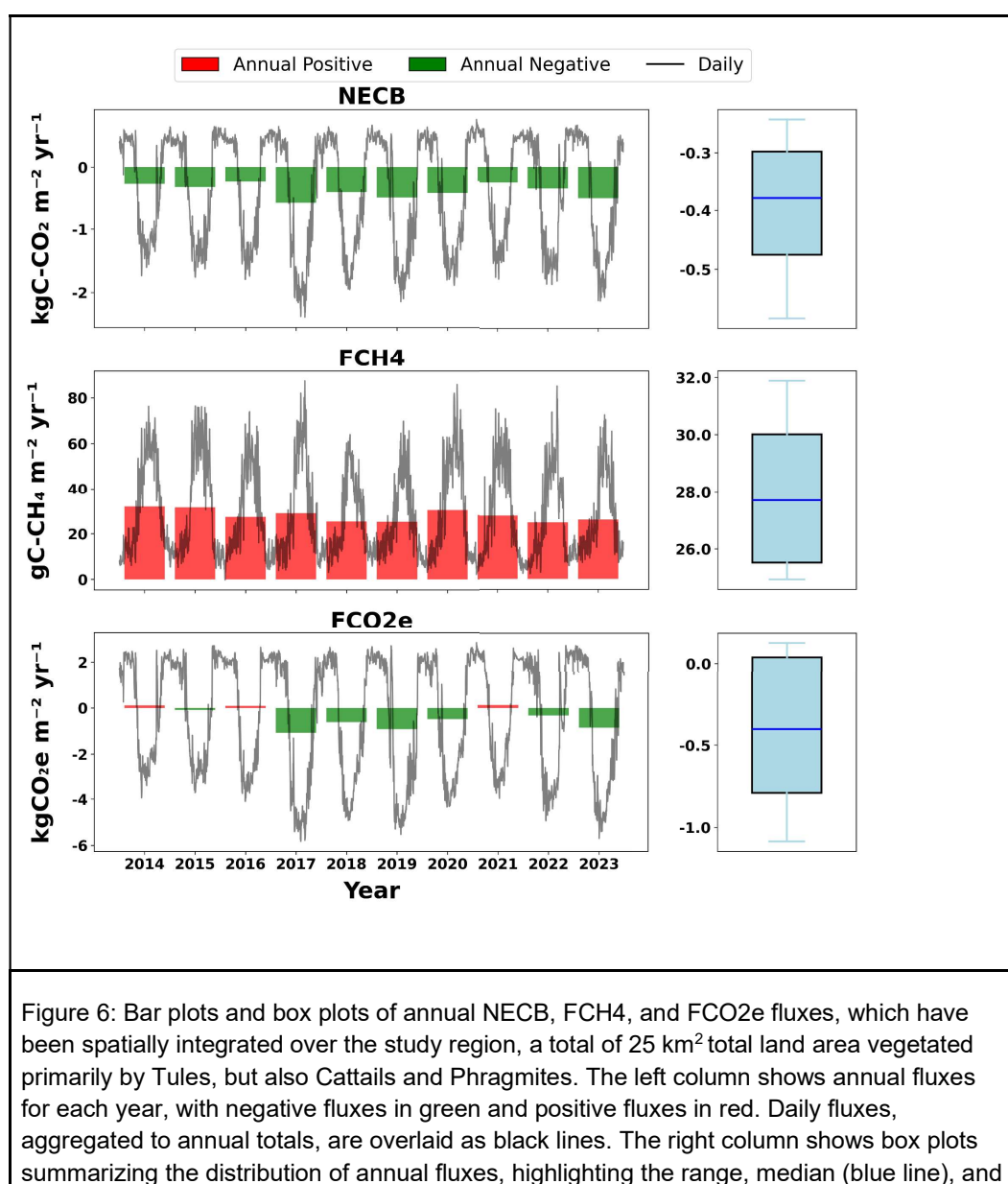


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spread of values. Each row represents a different flux variable: (a) NECB, (b) FCH<sub>4</sub>, and (c) FCO<sub>2e</sub>.

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#### 418 **4. Discussion**

419 This study demonstrates the development and evaluation of a data-driven framework to upscale  
420 terrestrial CO<sub>2</sub> and CH<sub>4</sub> flux estimates for non-tidal wetlands in the Sacramento-San Joaquin  
421 Delta. By systematically comparing models of varying complexity, including linear regression,  
422 ensemble methods, gradient boosting algorithms, and recurrent neural networks (RNNs), we  
423 presented a transparent assessment of model performance. The goals were to identify the  
424 model that best predicts CO<sub>2</sub> and CH<sub>4</sub> fluxes and critically appraise whether incremental  
425 complexity is justified by improvements in predictive capacity. Relevant cited works have  
426 included many different machine learning approaches for predicting emissions. This work aims  
427 to unify modelling efforts by establishing a standard framework for developing robust data-  
428 driven models, particularly for upscaling purposes.

429 Our results indicate that non-linear and more advanced models generally outperformed simple  
430 linear regression approaches. Among all tested models, the Long Short-Term Memory (LSTM)  
431 and Gated Recurrent Unit (GRU) neural networks provided the highest overall skill in predicting  
432 both CO<sub>2</sub> and CH<sub>4</sub> fluxes at daily timescales. This improvement was marginal but consistent,  
433 supporting the notion that time-series models, which inherently capture temporal dependencies  
434 and non-linearities, can provide tangible benefits over linear methods and traditional machine  
435 learning algorithms.

436 However, while these deep learning models performed best, the performance gains were not as  
437 large as might be expected given their significantly higher complexity and computational  
438 demands. Similar outcomes have been noted in other ecological modeling applications, where  
439 advanced machine learning methods yield improvements that are statistically significant yet  
440 modest in terms of performance gains relative to linear models (Oh et al., 2022; Wood, 2022).

441 The deep learning models provided reasonable estimates of daily fluxes but struggled to  
442 replicate the full range of interannual variability observed in the field measurements, which is a  
443 common issue for data-driven models in this field (Nelson et al., 2024). This limited ability to  
444 capture long-term trends and extremes mirrors common challenges in machine learning-based  
445 modeling, where the absence of explicit mechanistic understanding limits extrapolation beyond  
446 the conditions represented in the training data. The difficulty in reproducing interannual  
447 fluctuations was particularly evident for CH<sub>4</sub> fluxes, an outcome consistent with the high spatial  
448 and temporal complexity of CH<sub>4</sub> cycling in wetland environments and the limited availability of



449 subsurface parameters (e.g., oxygen concentration, redox conditions, substrate availability) that  
450 drive CH<sub>4</sub> production. This may not be surprising as the number of annual cycles available in  
451 the training set was only 35 years.

452 The observed regression to the mean and the reduced dynamic range in model predictions may  
453 reflect insufficient representation of key environmental drivers in the feature set or inadequate  
454 temporal coverage and variability in the training data. While publicly available datasets such as  
455 WLDAS and MODIS were effective at providing spatially and temporally comprehensive inputs,  
456 the lack of direct subsurface and soil biogeochemical measurements likely limited the model's  
457 ability to capture critical internal processes that are likely causing the observed differences  
458 between years. Although the feed-forward selection process for the model features had access  
459 to an extensive pool of relevant features, results indicated that only a small subset of features  
460 was necessary to maximise performance. This suggests that, while there are many features that  
461 control CO<sub>2</sub> and methane, their contribution to predictive accuracy may be redundant or  
462 captured indirectly by other variables. The exclusion of particular features, such as the water  
463 table depth for FCH<sub>4</sub>, illustrates the trade-off between mechanistic intuition and data-driven  
464 optimization. Strong correlations between features and limited independent variability can lead  
465 to features being left out that would typically be considered ecologically relevant.

466 After applying the chosen model (LSTM) to calculate CO<sub>2</sub> and CH<sub>4</sub> fluxes, we estimated NECB  
467 and CO<sub>2</sub>-equivalent fluxes for similar wetland settings across the Delta region. The results show  
468 spatial heterogeneity and pinpoint regions that act as stronger net carbon sinks, as well as  
469 areas where CH<sub>4</sub> emissions may offset climate benefits of net carbon sequestration. Such  
470 insights support targeted conservation and restoration strategies aimed at maximizing net  
471 carbon sequestration benefits, facilitating ongoing efforts to restore and manage wetlands to  
472 contribute to net-zero emission goals.

473 A key advantage of the chosen approach is its reliance on readily available, open-source data  
474 streams and standard computational resources. The framework can be deployed efficiently  
475 without specialized hardware, making it accessible to resource-limited organizations,  
476 practitioners, and researchers.

477 The primary objectives of this study were to identify a suitable model, contextualize model  
478 performance by comparing to a baseline linear regression, and highlight trade-offs between  
479 complexity, interpretability, and accuracy. By explicitly testing multiple models ranging from  
480 simple linear regressions to advanced recurrent neural networks, we demonstrated that  
481 complexity alone does not guarantee a substantial increase in predictive power. Instead,  
482 complexity should be adopted judiciously, based on the magnitude of performance gains, the  
483 cost of model implementation, and the level of interpretability.

484 We suggest that future modeling efforts should focus on deriving mechanistically relevant  
485 predictors (Ouyang et al., 2023), and incorporating hybrid modeling approaches (Yao et al.,  
486 2023) that combine the strengths of process-based and machine learning methods. Attention  
487 mechanisms (Yuan et al., 2022), advanced architectures (e.g., Transformers (Vaswani, 2017)),



or physics-informed machine learning (Raissi et al., 2019) may also help address model performance limitations.

## 5. Conclusions

This study provides a transparent, methodical demonstration of an artificial intelligence approach to modeling wetland carbon dioxide (CO<sub>2</sub>) and methane (CH<sub>4</sub>) emissions, using a suite of “off-the-shelf” tools and establishing a standardized benchmarking protocol for model performance evaluation. In the study region (the Sacramento–San Joaquin Delta), inter-model comparisons revealed modest but appreciable performance differences when comparing advanced models with a linear regression baseline. While there are tangible benefits to employing machine learning for these purposes, it is likely that the gap between simpler models and more sophisticated models will widen as data quantity and quality continues to increase. Ultimately, this study lays the groundwork for regional scale model benchmark testing, facilitating the development of more advanced modeling approaches that can guide wetland management, restoration planning, and climate mitigation strategies.

## Code and data availability

The current version of the **RCCAT model** is available on GitHub at <https://github.com/ashbre2/RCCAT> under the **MIT License**. The exact version of the model used to produce the results presented in this paper has been archived on Zenodo under <https://doi.org/10.5281/zenodo.14933820> (Brereton, 2025)

## Author contributions

AB developed the model, performed the analysis, and led the writing of the manuscript. ZM, BA, AP, WR, and KY contributed to model development and provided expertise in carbon flux modeling. QZ contributed expertise in machine learning methods. TA provided domain-specific knowledge of the region of interest. YZ and YX contributed expertise in hydrological processes. All co-authors reviewed, provided input on the manuscript drafts, and approved the final version.

## Competing interests

The authors declare that they have no conflict of interest.

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522 significantly.
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