

~~Amazonian Podzols – a carbon time bomb?~~ Vulnerability of soil organic carbon in Amazonian Podzols to changes in environmental conditions.

Naoise Nunan^a, Claire Chenu^b, Valérie Pouteau^b, André Soro^{a,e}, Kevin Potard^a, Célia Regina Montes^c, Patricia Merdy^d, Adolpho José Melfi^c, Yves Lucas^d

^a Institute of Ecology and Environmental Sciences, iEES Paris, Sorbonne Université, CNRS, IRD, INRA, UPEC, Univ Paris Diderot, 4 Place Jussieu, Paris 75005, France

^b UMR Ecosys, Université Paris-Saclay, INRAe, AgroParisTech, Palaiseau, 91120, France

^c IEE, LEST, Universidade de São Paulo, São Paulo 05508-010, Brazil

^d Université de Toulon, Aix Marseille Univ, CNRS, IM2NP, Toulon, France

^e present address: Renewable Materials Research Centre, Department of Wood and Forest Sciences, Université Laval, Québec, QC, Canada

Corresponding author:

Naoise Nunan: naoise.nunan@cnrs.fr; Orcid: 0000-0003-3448-7171

E-mail addresses and Orcid:

Naoise Nunan – naoise.nunan@cnrs.fr; Orcid: 0000-0003-3448-7171

Claire Chenu – claire.chenu@inrae.fr; Orcid: 0000-0001-9054-0489

Valérie Pouteau – valerie.pouteau@inrae.fr

André Soro – dogninema-andre.soro.1@ulaval.ca; Orcid: 0000-0002-1694-6234

Célia Regina Montes – crmlauar@usp.br; Orcid: 0000-0002-5173-1909

Patricia Merdy – patricia.merdy@im2np.fr; Orcid: 0000-0002-4540-6464

Adolpho José Melfi – ajmelfi@usp.br; Orcid: 0000-0001-5960-937X

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34 **Abstract**

35 It has recently been shown that the C stocks in Amazonian podzols are very large. They are much larger than was
36 previously thought, particularly in the Bh horizon, which has been estimated to contain in excess of 10Pg C for
37 Amazonia alone. It is predicted that changes in the regional climate will result in a drier soil moisture regime, which
38 may affect the C dynamics in these generally waterlogged soils. In order to determine the vulnerability to
39 decomposition of the organic C contained in the Amazonian podzols as a result of environmental changes, we
40 established a series of incubation experiments in which the effects of different environmental factors were measured.
41 The direct effect of drier soil moisture regimes was tested by incubating undisturbed cores from the Bh horizon at a
42 range of matric potentials. Contrary to what is usually found in soils, no significant difference in mineralisation was
43 found among matric potentials, suggesting that other factors control microbial mineralisation of this organic C. In a
44 second series of incubations, the effect of nitrogen additions, of anoxic conditions and of labile C substrate additions
45 were also tested on undisturbed cores of the Bh horizon. Samples incubated under oxic conditions produced more than
46 twice as much CO₂ as samples incubated under anoxic conditions, whilst the mineralisation rates of samples incubated
47 under oxic conditions with the addition of N increased more than four-fold relative to the anoxic samples. The addition
48 of labile C did not have a significant effect on C mineralisation. The data suggest that the large pool of C in Amazonian
49 Podzols may be vulnerable to increases in N and O₂ availability. An extrapolation of the data to the whole of the
50 Amazonian podzols suggests that changes in environmental conditions could result in an increased C-CO₂ flux to the
51 atmosphere of up to 0.41Pg C per annum. This is equivalent to 8% of the current net global C flux to the atmosphere.

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54 | **1. Introduction**

55 Hydromorphic Podzols are widespread in equatorial regions (Bernoux et al., 2002; Quesada et al., 2011). They
56 are characterised by a deep sandy horizon on top of clayey horizons. A small portion of the dissolved organic matter
57 from the upper organic horizon moves freely through the sandy horizon and accumulates at the transition with the clay
58 horizon below, forming a deep Bh horizon that can reach thicknesses of several meters (Montes et al., 2011; Sierra et

59 al., 2013; Doupoux et al., 2017). In the Amazon basin, the majority of the dissolved organic carbon is transferred to the
60 Amazon river via a perched water-table in the sandy E horizon above the impermeable Bh horizon (Doupoux et al.,
61 2017). Despite the C loss to the river network, the C stocks in the Amazonian Podzols have been estimated to exceed
62 13Pg C, the majority of which is contained in the deep Bh horizons (Montes et al., 2011). This C represents a significant
63 portion of the C stored in the Amazon basin; the total woody biomass of Amazonian forests having been estimated to be
64 between 121 and 126 Pg C (Malhi et al., 2006). Furthermore, ¹⁴C dating of the Bh horizon C suggests that the Bh C is
65 very old, reaching ages of up to 25 thousand years (Doupoux et al., 2017). The Bh horizons are also characterised by
66 organic matter with C/N ratios that can be exceptionally high, with values sometimes exceeding 80 (Montes et al.,
67 2023), and by the fact that they are cemented and relatively impermeable (Sierra et al., 2013; Montes et al., 2023). The
68 E horizon above it is therefore generally waterlogged, thus preventing oxygen from penetrating down the soil profile to
69 the Bh horizon.

70 The vulnerability of this C to changes in environmental conditions is still poorly characterised. The ¹⁴C ages in
71 the Bh horizons suggest that it is very stable and resistant to decomposition. However, this does not mean that it is not
72 vulnerable to decomposition if environmental conditions change. Old soil C can be mineralised quite rapidly, as was
73 shown, for example, by Fontaine et al. (2007): the simple addition of cellulose to subsoil samples of a Cambisol
74 stimulated the mineralisation of old organic C with an apparent ¹⁴C age of 2500 years. The decomposition of old,
75 millennial organic C was shown to be as responsive to warming as fast cycling C in high latitude soils (Vaughn and
76 Torn, 2019). These data suggest therefore that old organic C may not be intrinsically resistant to decomposition, but
77 rather that it is not decomposed under the prevailing conditions.

78 Regional climate models, which downscale global climate projections to regions of interest, all predict reductions
79 in precipitation levels and longer or more frequent periods of drought in the Amazon region (Avila-Diaz et al., 2020)
80 and, indeed, the Amazon experienced record-breaking droughts in 2023 and 2024 (Marengo, 2024). A significant
81 potential consequence of prolonged periods of drought is that the perched water table above the Bh horizon could dry
82 out, thus leading to its aeration. It is well established that decomposition rates in aerobic conditions are far greater than
83 those in anoxic conditions (Linn and Doran, 1984, Moyano et al., 2012), so we might expect significant increases in Bh
84 horizon mineralisation rates subsequent to such changes in conditions.

85 The projected climate changes, compounded by deforestation and fires, may also lead to changes in the structure
86 and composition of the forest (Esquivel-Muelbert et al., 2019; Flores et al., 2024) and, ultimately, to forest dieback
87 (Boulton et al., 2022; Flores et al., 2024). There are already precursor signs of resilience loss in the Amazon forest
88 related to reductions in mean annual precipitation and human land-use (Boulton et al., 2022). Forests on hydromorphic

89 podzols are particularly vulnerable to dieback due to their shallow rooting depth, which is probably a consequence of
90 the lack of nutrients in the E horizon, due to its very low exchange capacity, as well as the mostly waterlogged
91 conditions. According to Sierra et al. (2013), the high bulk density of this E horizon below the organic horizon may also
92 play a role. Such dramatic changes in forest dynamics are known to alter element cycling and potentially result in
93 nutrient losses, including N, from the vegetation and the surface organic horizon to the soil horizons below (Xiong et
94 al., 2011). The very high C/N ratio of the Bh horizon organic matter suggests that the mineralisation of this organic C is
95 constrained by N availability and increases in N flux from overlaying horizons, due to the decomposition of dead
96 biomass or increased N deposition (Galloway et al., 2008), may unlock the Bh horizon C.

97 As indicated above, the mineralisation of old, deep soil C can be stimulated by inputs of fresh, labile organic
98 matter (Fontaine et al., 2007). This phenomenon is known as the “priming effect”. The priming effect is believed to
99 arise due the inputs of fresh organic matter causing microbial communities to mine for N by decomposing organic
100 matter or due to soil organic matter being co-metaboilised during with the frsh inputs (Blagodatskaya & Kuzyakov,
101 2008). The death of plant biomass and its subsequent decomposition willis likely to release significant amounts of labile
102 organic matter thatwhich maycould stimulate the mineralisation of the Bh horizon organic C through a priming effect.

103 Previous studies have suggested that large quantities of C could be released to the atmosphere from these Bh
104 horizons under certain conditions (Sierra et al., 2013; Montes et al., 2023). However, in these studies the structural
105 integrity of samples was not preserved, which is likely to have allowed far greater oxygenation of the samples than
106 would occur under natural circumstances. Furthermore, the disruption of the physical structure of soils is known to
107 stimulate the mineralisation of organic C (Rovira and Greacen 1957; Salomé et al., 2010). Here, we incubated
108 undisturbed soil cores in order to obtain more realistic estimations of the vulnerability of organic C in hydromorphic
109 Podzols, particularly that in the Bh horizon, to a range of potential future disruptions to the present environmental
110 conditions. The conditions tested were changes in moisture status, and in O₂ and N availability. We also tested the effect
111 of the addition of a cocktail of labile organic molecules that sought to mimic the arrival of soluble, labile organic matter
112 from the soil surface. The hypotheses were fourfold. The first was that reductions in moisture content from saturation to
113 a matric potential of approximately -31.6 hPa would result in increases in CO₂ emissions, due to increased O₂
114 availability in the pore space (Moyano et al., 2012; Sierra et al., 2017), but that further decreases in moisture content
115 would result in lower CO₂ emissions, due to reductions in the diffusion of C-substrates towards enzymes or the
116 diffusion of enzymes towards insoluble C-substrates (Davidson et al., 2014). The second hypothesis was that anoxic
117 conditions were responsible for the slow decomposition rates (Sierra et al., 2017; Davidson et al., 2014) and that
118 decomposition would be stimulated by increases in O₂ levels. The third hypothesis was that decomposition is N limited,

119 | as indicated by the characteristically high C:N ratios of Podzol Bh horizons (Montes et al., 2023) and, therefore, that the
120 | addition of mineral N would stimulate decomposition. The fourth and final hypothesis was that the addition of readily
121 | available sources of C would result in a priming effect (Fontaine et al., 2007) that releases CO₂ from Bh horizon organic
122 | matter.

123 | **2. Materials and methods**

124 | Samples were taken from three sites circa one hundred meters apart in the region of Cabeça do Cachorro,
125 | Amazonas state, Brazil, near the town of São Gabriel da Cachoeira (Fig S1). The profiles at the three sites were typical
126 | Amazonian Podzols. They were made up of a waterlogged O horizon of about 15cm, an E horizon that was also
127 | waterlogged and slightly less than a meter thick, a silt-loam Bh horizon that was slightly over a meter thick underneath
128 | which there was a C horizon. Bulk soil and undisturbed soil cores were collected from the OH, E, Bh and C horizons to
129 | a depth of 3m using an augur inserted into a metal tube that was used to prevent the sandy, waterlogged E horizon from
130 | collapsing into the bore hole. This sampling procedure was necessary due to the perched water table above the Bh
131 | horizon. The samples were placed in medical sample containers and closed. They were thus maintained in anoxic
132 | conditions prior to use. One undisturbed sample was taken from each site x horizon per incubation treatment (see
133 | below) and for the establishment of moisture release characteristics, resulting in 10 undisturbed samples being taken
134 | from each horizon at each site, and a total of 120 samples. The total C and N contents of the soils were
135 | analysed/determined by elemental analysis, and pH measurements were carried out in a 1:5 soil:water mixture of the
136 | bulk soil.

137 | Soil moisture retention curves were established on three replicate samples from each horizon using a suction table
138 | and ceramic pressure plates (Eijkelkamp). All samples used in the subsequent incubation experiments were adjusted to
139 | the desired matric potential using a suction table and pressure plates.

140 | Two microcosm incubation experiments were set up in order to measure~~which the~~ CO₂ emissions from samples in
141 | response to changes in environmental conditions. ~~were measured over a period of slightly more than two months (68~~
142 | ~~days for the first incubation and 72 days for the second) at 28°C in the dark.~~ The first incubation measured the response
143 | to differences in moisture content in O, E, Bh and C horizons. This incubation lasted 68 days. The second incubation
144 | measured changes in CO₂ emission in response to O₂, mineral N or substrate-C availability in the Bh horizon and lasted
145 | for 72 days. Both sets of incubations were carried out at 28°C in the dark. There were three replicate microcosms for
146 | every treatment, resulting in a total of 60 microcosms (4 horizons x 5 moisture contents x 3 replicates) for the first
147 | incubation and 12 microcosms (4 treatments x 3 replicates) for the second incubation. In both incubation experiments
148 | the samples were placed on sample holders in 1 L air-tight jars that were sealed with rubber gaskets and firmly closed

149 with spring-lock catches. The glass lids of the jars were fitted with a septum that allowed for headspace sampling. The
150 concentration of CO₂ in the headspace of sample microcosms was analysed on 18 and 23 occasions during the first and
151 second incubations, respectively. The headspace of the microcosms was flushed with CO₂-free air at regular intervals to
152 ensure that the conditions did not become anoxic, except for the anoxic treatment in the second incubation experiment
153 (see below). The moisture content of the samples was adjusted gravimetrically when necessary. The CO₂ concentrations
154 in the microcosm headspaces were determined by gas chromatography (Agilent 3000A) and the isotopic signature of
155 the CO₂ by gas chromatography coupled to an Isochrome III isotope ratio mass spectrometer (Micromass-GVI Optima).
156 The amount of soil organic C and ¹³C-labelled substrate mineralised in the samples that received a cocktail of ¹³C-
157 labelled substrates (see below) was determined by isotopic mass balance (e.g. Ruamps et al., 2011).

158 In the first incubation experiment the three replicate samples from each horizon were incubated at matric
159 potentials of 0, -5, -31.6, -316 or -1585 kPa. The matric potentials were chosen in order to have a broad range of
160 potentials from saturation (0 kPa) to the permanent wilting point (-1585 kPa). The range of matric potentials was
161 centered on -31.6 kPa because respiration maxima are generally reached at approximately this potential (Moyano et al.,
162 2012). In the second experiment a series of treatments were imposed in order to determine whether microbial
163 decomposition of the Bh horizon organic matter was limited by O₂, N or energy availability. The samples were first pre-
164 incubated for two weeks under oxic conditions in order to ensure that there were no differences in soil respiration
165 among the samples chosen for the different treatments. After the pre-incubation, four treatments were imposed: an
166 anoxic treatment, an O₂ treatment, an O₂ + N treatment and an O₂ + simple organic substrate treatment.

167 The anoxic treatment was established by replacing the microcosm headspace with N₂. In the oxic treatments
168 samples were incubated in the presence of ambient O₂ levels, alone or with the addition of N (1.6 mg g⁻¹ soil, in order to
169 bring the C/N ratio of the soil to approximately 20) or a cocktail of substrates (15 mg g⁻¹ soil C). The cocktail was made
170 up of ¹³C-labelled 30% glucose, 50% vanillin and 20% pyruvic acid. These proportions were chosen to approximate the
171 soil solution of forest soils (Kaiser et al., 2001). On the 34th day of incubation, a second addition of N (half the amount
172 previously added) and of the substrate cocktail was carried out. The other two treatments received water. The matric
173 potential of the samples was set at -1585 kPa.

174 Due to the difficulty and expense of the sampling exercise, there were not a sufficient number of undisturbed
175 cores for a fully factorial experimental design (i.e. there were no treatment combinations) and, therefore, ~~F~~treatment and
176 horizon differences in the cumulative amount of CO₂ evolved from the soil samples during the incubations were tested
177 by one-way ANOVA. Data were log-transformed prior to analysis, where necessary. Differences in soil properties
178 among horizons were also tested by one-way ANOVA. In order to estimate the annual CO₂ flux from the Bh horizon of

179 Amazonian Podzols to the atmosphere under the different conditions tested here, a first-order decay model with one
180 pool (Manzoni et al., 2012) was fitted to the cumulative CO₂ emission curves (Equation 1):

$$181 \quad -CO_2 = a(1 - e^{-at}) \quad (1)$$

182 where a is the pool of mineralisable C, α is the rate at which the organic C is mineralised and t is time. The model fitting
183 was done using the nls command in R (R Core Team (2022). R: A language and environment for statistical computing.
184 R Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>).

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187 **3. Results and discussion**

188 The soil properties were typical of Amazonian Podzols (Montes et al., 2011; Sierra et al., 2013; Doupoux et al.,
189 2017). The pH was acidic (<4.6) throughout the profile, without showing any significant differences among horizons
190 (Fig. 1). The C and N contents ranged from 2 to 248 and from 0.1 to 9.3 mg g⁻¹ soil, respectively, with significantly
191 higher values for both variables in the OH horizon (Fig. 1), and the C/N ratios were all >20 but the Bh horizon was
192 the most acidic (pH=4) and had showed by far the highest C/N-ratio at 53, which was significantly higher than in the
193 other horizons). No other statistically significant differences were observed.

194 **3.1 Effect of moisture status and horizon**

195 We first tested the effects of changes in moisture status on the mineralisation rates of the organic C in each of the
196 OH, E, Bh and C horizons. This was achieved by adjusting the matric potential of the samples (based on the water
197 retention curves, Fig S24) and measuring CO₂ emissions during a subsequent incubation at 28°C and under aerobic
198 conditions. There were no clear trends with matric potential in any of the horizons (Fig. 2), contrary to what might be
199 expected (Moyano et al., 2012; Sierra et al., 2017). Whilst this is surprising, the water retention curves (Fig S24) show
200 that sample moisture content varied little as a function of matric potential, suggesting that the saturation of the pore
201 network changed little and that the environmental conditions of microbial decomposers in the samples were relatively
202 unaffected. Sierra et al. (2017) showed that moisture effects on decomposition rates are strongly modulated by O₂
203 availability. As the changes in moisture content in this experiment were small, O₂ availability inside the cores may not
204 have changed much, thus reducing any matric potential effect on decomposition.

205 Overall, whilst the amounts of C mineralised from the Oh horizon were an order of magnitude higher ($P<0.001$)
206 than from the other horizons (Fig. 2A), reflecting the higher organic C content of this horizon. However, the high
207 variability in the data meant that the differences among horizons within each matric potential were not always
208 significant. At the matric potentials of -5 and -31.6 kPa significant differences were not found with the E and Bh

209 | horizons, respectively, and neither of these horizons showed a significant difference with the OH horizon at -316kPa.
210 | ~~€~~The specific mineralisation rates (the amount of CO₂ produced per unit organic C) were much significantly (P<0.001)
211 | higher in the E horizon at all the matric potentials, except for the driest (Fig. 2B). Here also, the high variability in the
212 | data meant that specific respiration rates in the E horizon within a matric potential were not significantly different from
213 | the other horizons. The specific mineralisation rate is indicative of the mineralisability of the organic C (Fierer et al.,
214 | 2003; Salomé et al., 2010). With the exception of the E horizon, the specific mineralisation rates were generally in the
215 | range of what has been found elsewhere in incubations of approximately two months (usually between 1-3%, e.g.
216 | Salomé et al., 2010; Autret et al., 2020; Kan et al., 2020), although the rate of the Bh horizon was at the lower end of
217 | this range. The specific mineralisation rates in the E horizon were an order of magnitude higher however. These very
218 | high rates may be linked to the low total C content of the E horizon (2.3 mg g⁻¹ soil), however, others have not found
219 | such high specific mineralisation rates, even in deep soil where total C contents were lower than what was observed
220 | here (Fierer et al., 2003; Salomé et al., 2010). It is more likely therefore, that the high specific mineralisation rates,
221 | suggesting that the organic C in this horizon was highly labile and readily available. The high decomposability of the
222 | organic C in the E horizon may be due to the nature of the organic matter or due to the sandy texture of the horizon. In
223 | general, organic C is less persistent in sandy soils, possibly due to lower rates of mineral-associated organic matter
224 | formation (Haddix et al., 2020) or to more oxic conditions. The lack of difference in the mineralisation rates across
225 | matric potentials suggests that oxygen availability did not play a major role in this particular experiment, possibly
226 | because the cores were in oxic conditions within the microcosms. The fact that the organic C in the E horizons ~~above~~
227 | ~~and below the Bh horizon~~ was more mineralisable than in the Bh horizon (Fig. 2B) suggests that the specific conditions
228 | in the Bh horizon had a strong limiting effect on the mineralisation, as the C in the Bh horizon was likely transferred
229 | from the E horizon above it. ~~A~~The second series of incubations was carried out in order to investigate the causes of the
230 | relatively low specific mineralisation rates in the Bh horizon.

231 | 3.2 Effects of N, C and O₂ availability

232 | The second incubation experiment was carried out to test whether the decomposition of the organic matter in the
233 | Bh horizon was N, O₂ or energy limited. The undisturbed samples were incubated for two weeks under oxic conditions
234 | and at 28°C prior to initiating the treatments, in order to ensure that there were no differences among the samples
235 | selected for each treatment (Fig 3A). There was a rapid divergence in mineralisation rates following the initiation of the
236 | treatments and, by the end of the incubation, there were significant differences (P<0.001) in the amounts of CO₂
237 | released from the samples subjected to the different treatments (Fig 3B). The samples that received N under oxic
238 | conditions mineralised approximately twice as much organic C as the oxic control (P<0.01) and the samples that

239 | received a cocktail of simple substrates ($P<0.05$), and in excess of four times as much C as the samples incubated under
240 | anoxic conditions ($P<0.001$; Figs 3 & S32 for O_2 contents in all treatments). No significant differences were observed
241 | between the oxic treatments with or without the addition of the simple substrate cocktail, even though the substrate
242 | cocktail was rapidly mineralised after both additions (Fig S43). Both oxic treatments without N mineralised
243 | significantly ($P<0.01$) more than the anoxic treatment. These data suggest that the mineralisation of the Bh horizon
244 | organic C is constrained by N and O_2 availability rather than by energetic deficiencies in the organic matter. The very
245 | large C/N ratio of the Bh horizon organic matter (Fig 1) tends to confirm that low N availability to soil microbial
246 | decomposers is a major limiting factor of organic matter decomposition in these soils. An analysis of the ^{14}C -age of the
247 | organic C in other Amazonian Podzols showed that there was a negative relationship between the N content of the of the
248 | organic matter and the the ^{14}C -age of the organic C (Montes et al., 2023), confirming that N limitation is a major factor
249 | in the dynamics of C in such soils.

250 | There was a CO_2 pulse after both additions of the substrate cocktail, but this was due to the mineralisation of the
251 | substrate-C that was added rather than an increase in the mineralisation of Bh horizon organic C (Figs 3 & S4). The lack
252 | of a priming effect may be due to the low pH of the soil (Fig 1). The priming effect is more common in soils with pHs
253 | between 5.5 and 7.5 but tends to be lower at the pH values found here (Wang and Kuzyakov, 2024). Furthermore, it has
254 | been shown that soils that are characterised by high levels of mineral associated organic C, as is the case in the Bh
255 | horizons of Podzols (Schmidt et al., 2000; Douppoux et al., 2017), also tend to be less prone to the priming effect (Chen
256 | et al., 2019). It should also be noted that a month after the first addition of the substrate cocktail, only 22% of the added
257 | C was mineralised, and only 15% was mineralised slightly more than a month after the second addition (Fig S4). These
258 | mineralisation rates are lower than what is usually found.- The mineralisation of glucose often exceeds 60% after a
259 | month's incubation (e.g. Hamer and Marschner, 2002), while that of pyruvate and vanilin can exceed 30% (Chenu et al.,
260 | 2025) and 20% (Juarez et al., 2013), respectively. These low mineralisation rates may also have been due to an N
261 | limitation, but this would have to be confirmed experimentally. Nitrogen limitation can arise due to microbial cells
262 | being unable to produce proteins, such as enzymes or membrane transport proteins, necessary for activity, as proteins
263 | are N rich molecules (Nunan et al., 2020).

264 | In view of the very large quantities of organic C that are stored in the Bh horizon of the Amazonian Podzols
265 | (Montes et al., 2011), we sought to estimate the annual CO_2 flux from these horizons to the atmosphere under the
266 | different conditions tested here. We first fitted a first order decay model with a single pool to the respiration data (Fig
267 | S54 – best fit based on the Akaike Information Criterion) and extrapolated the mineralisation curves to a year. Montes et
268 | al. (2011) estimated that 78.8 % of the 13 Pg C in the Podzol profiles is found in the Bh horizon (10.45 Pg C), which we

269 | used to estimate the potential total C fluxes from the Bh Horizon of the Amazonian Podzols (Fig. 4). The ~~dramatic~~
270 | increase in CO₂ flux in the oxic treatment with N translates to an extra 0.41 Pg C yr⁻¹ being released to the atmosphere
271 | compared to the anoxic treatment ($P < 0.0501$). The other two oxic treatments also resulted in significant increases in the
272 | amount of C released ($P < 0.01$), ~~but these differences were not statistically significant.~~

273 | Global soil respiration estimates are subject to large uncertainties, due to the complex set of biogeochemical and
274 | biophysical processes that are involved. These uncertainties are one of the major causes of uncertainty in terrestrial
275 | ecosystem models (He et al., 2022). Nevertheless, ~~data-driven estimates, generally based on the global soil respiration~~
276 | ~~database (Jian et al., 2021) and using a variety of methods, have been carried out (Hashimoto et al., 2023). These~~
277 | ~~estimates suggest that global soil respiration ranges from 68 to 101 PgC yr⁻¹ (Hashimoto et al., 2023). The potential~~
278 | ~~increase in soil respiration from Amazonian Podzols under oxic and N replete conditions is therefore equivalent to 0.4~~
279 | ~~to 0.6% of the global soil respiration. However, soil respiration is made up of autotrophic respiration (the respiration of~~
280 | ~~plant roots) and heterotrophic respiration (respiration resulting from the decomposition of soil organic matter), and a~~
281 | recent study has estimated global soil heterotrophic respiration to be 48.8 ± 0.9 Pg C yr⁻¹ (He et al., 2022). The potential
282 | increase in CO₂ flux from Amazonian Podzols couldis therefore be equivalent to 0.8% of global soil heterotrophic
283 | respiration. ~~Bearing in mind that the most recent estimates put fossil CO₂ emissions at 9.9 ± 0.5 Pg C yr⁻¹ and the net~~
284 | ~~annual flux of C to the atmosphere at 5.2 Pg C yr⁻¹ (Friedlingstein et al., 2023), the potential increase in CO₂ emissions~~
285 | ~~from hydromorphic Podzols would be equivalent to a 4.3% increase in fossil CO₂ emissions or to a 7.9 % increase of~~
286 | ~~the net C flux to the atmosphere. These numbers are significant and unlikely to be captured in future climate model~~
287 | ~~projections. Should such fluxes occur, they would considerably exacerbate climate change.~~

288 | There are a number of uncertainties associated with the estimates put forward in this study. The first is that it is a
289 | laboratory study and, even though the samples were undisturbed, the experiment and the treatments are somewhat
290 | artificial. For example, the Bh horizons can often be found at depths greater than 1m (Doupoux et al., 2017) and the
291 | degree to which O₂ or N would reach it is uncertain. Ideally, an experiment testing similar treatments should be carried
292 | out *in situ* in order to determine the magnitude of the vulnerability of the organic C to N and O₂ availability, though this
293 | would be extremely difficult. The ~~second uncertainty~~first is the use of a first order decay model with only one pool to
294 | extrapolate to yearly CO₂ carbon fluxes. The use of single pool models has been criticised in the past because it can
295 | mask the presence of smaller C pools with faster turnover times (Davidson et al., 2000). This criticism is less relevant
296 | here because the pool is small and even if it did respond differently to the treatments, it would not change the overall
297 | trend of the results. Furthermore, one might expect this smaller pool to respond more rapidly to N availability and, if
298 | anything, increase the observed N effect. Nitrogen additions are often associated with an increase in enzymes that

299 catalyse carbohydrate hydrolysis but a decrease in oxidative enzymes that catalyse the breakdown of polyphenols
300 (Moorhead and Sinsabaugh, 2006), thus increasing the decomposition rates of the fast-cycling C pool and slowing down
301 the decomposition of the slow-cycling C pool. Although, the decomposition of fast-cycling C pools is likely less
302 sensitive to variations in O₂ levels than that of slow-cycling C pools (Lin et al., 2021), the smaller size of the pool is
303 unlikely to change the overall conclusions, as suggested above. The ~~third~~^{second} uncertainty lies in the duration of the
304 incubations (72 days), which may not have been long enough to detect the response of the slower-cycling C pools in
305 these soils. However, the parameters obtained from the first order decay model suggest that the treatments increased the
306 size of the mineralisable pool rather than the rate at which the pool was mineralised (Table S1). This suggests that the
307 treatments increased the amount of C that was readily mineralisable in the samples and, therefore, not detecting the
308 slow C pool's response may not be problematic as this change was detected. Furthermore, Montes et al. (2023)
309 measured CO₂ emissions from Bh horizon samples of Amazonian Podzols and found that the fast pool always
310 accounted for < 0.5% soil C, far less than in any of the treatments here, other than the anoxic treatment. A ~~third~~^{fourth}
311 uncertainty associated with the study is that the amount of N arriving in the Bh horizons annually may be lower than
312 what was added here, despite the fact that future N deposition projections indicate significant increases in reactive N
313 deposition (Galloway et al., 2004) and forest dieback might also result in higher soil N concentrations (Xiong et al.,
314 2011).

315 The CO₂ flux from the Amazonian Podzols may be further enhanced by increases in average temperature that will
316 occur with climate change, due to the very high C/N ratio of the Bh horizon: CO₂ fluxes from soils with high C/N ratios
317 show a positive response to temperature (Karhu et al., 2014). Overall, these data suggest that ~~the organic C contained ;~~
318 ~~under environmental conditions that fall within the window of climate model predictions;~~ⁱⁿ equatorial Podzols ~~is~~
319 ~~vulnerable to increases in N and O₂ availability and that could release~~ significant amounts of CO₂ ~~could be released~~ to
320 the atmosphere and thus exacerbate atmospheric CO₂ levels. ~~Should the estimates made here prove to be correct, then~~
321 ~~the response of equatorial Podzols to environmental changes would have to be included in earth system models.~~

322 1. Data availability

323 The data that support the findings of this study are openly available in Zenodo at
324 <https://zenodo.org/records/15824641>.

325

326 Author contribution

327 Naoise Nunan: Conceptualization, Investigation, Methodology, Data curation, Formal analysis, Writing – original
328 draft, Writing – review & editing, Funding acquisition.

329 Claire Chenu: Conceptualization, Investigation, Writing – review & editing.
330 Valérie Pouteau: Investigation, Data curation, Methodology, Writing – review & editing.
331 André Soro: Investigation, Data curation, Methodology, Writing – review & editing.
332 Kevin Potard: Investigation, Data curation, Methodology, Writing – review & editing.
333 Célia Regina Montes: Investigation, Writing – review & editing.
334 Patricia Merdy: Investigation, Writing – review & editing.
335 Adolpho José Melfi: Investigation, Writing – review & editing.
336 Yves Lucas: Conceptualization, Investigation, Writing – review & editing, Funding acquisition.

337 **Competing interests**

338 The authors declare that they have no conflict of interest

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351 **References**

- 352 Avila-Diaz, A., Benezoli, V., Justino, F., Torres, R., and Wilson, A.: Assessing current and future trends of climate
353 extremes across Brazil based on reanalyses and earth system model projections, *Clim. Dyn.*, 55, 1403–1426,
354 <https://doi.org/10.1007/s00382-020-05333-z>, 2020.
- 355 Bernoux, M., Da Conceição Santana Carvalho, M., Volkoff, B., and Cerri, C. C.: Brazil’s Soil Carbon Stocks, *Soil Sci.*
356 *Soc. Am. J.*, 66, 888–896, <https://doi.org/10.2136/sssaj2002.8880>, 2002.
- 357 Boulton, C. A., Lenton, T. M., and Boers, N.: Pronounced loss of Amazon rainforest resilience since the early 2000s,
358 *Nat. Clim. Change*, 12, 271–278, <https://doi.org/10.1038/s41558-022-01287-8>, 2022.

359 Davidson, E. A., Trumbore, S. E., and Amundson, R.: Biogeochemistry - Soil warming and organic carbon content,
360 Nature, 408, 789–790, 2000.

361 Doupoux, C., Merdy, P., Montes, C. R., Nunan, N., Melfi, A. J., Pereira, O. J. R., and Lucas, Y.: Modelling the genesis
362 of equatorial podzols: age and implications for carbon fluxes, Biogeosciences, 14, 2429–2440,
363 <https://doi.org/10.5194/bg-14-2429-2017>, 2017.

364 Esquivel-Muelbert, A., Baker, T. R., Dexter, K. G., Lewis, S. L., Brienen, R. J. W., Feldpausch, T. R., Lloyd, J.,
365 Monteagudo-Mendoza, A., Arroyo, L., Álvarez-Dávila, E., Higuchi, N., Marimon, B. S., Marimon-Junior, B. H.,
366 Silveira, M., Vilanova, E., Gloor, E., Malhi, Y., Chave, J., Barlow, J., Bonal, D., Davila Cardozo, N., Erwin, T.,
367 Fauset, S., Hérault, B., Laurance, S., Poorter, L., Qie, L., Stahl, C., Sullivan, M. J. P., Ter Steege, H., Vos, V. A.,
368 Zuidema, P. A., Almeida, E., Almeida De Oliveira, E., Andrade, A., Vieira, S. A., Aragão, L., Araujo-Murakami, A.,
369 Arets, E., Aymard C, G. A., Baraloto, C., Camargo, P. B., Barroso, J. G., Bongers, F., Boot, R., Camargo, J. L.,
370 Castro, W., Chama Moscoso, V., Comiskey, J., Cornejo Valverde, F., Lola Da Costa, A. C., Del Aguila Pasquel, J.,
371 Di Fiore, A., Fernanda Duque, L., Elias, F., Engel, J., Flores Llampazo, G., Galbraith, D., Herrera Fernández, R.,
372 Honorio Coronado, E., Hubau, W., Jimenez-Rojas, E., Lima, A. J. N., Umetsu, R. K., Laurance, W., Lopez-
373 Gonzalez, G., Lovejoy, T., Aurelio Melo Cruz, O., Morandi, P. S., Neill, D., Núñez Vargas, P., Pallqui Camacho, N.
374 C., Parada Gutierrez, A., Pardo, G., Peacock, J., Peña-Claros, M., Peñuela-Mora, M. C., Petronelli, P., Pickavance,
375 G. C., Pitman, N., Prieto, A., Quesada, C., Ramírez-Angulo, H., Réjou-Méchain, M., Restrepo Correa, Z.,
376 Roopsind, A., Rudas, A., Salomão, R., Silva, N., Silva Espejo, J., Singh, J., Stropp, J., Terborgh, J., Thomas, R.,
377 Toledo, M., Torres-Lezama, A., Valenzuela Gamarra, L., Van De Meer, P. J., Van Der Heijden, G., et al.:
378 Compositional response of Amazon forests to climate change, Glob. Change Biol., 25, 39–56,
379 <https://doi.org/10.1111/gcb.14413>, 2019.

380 Fierer, N., Allen, A. S., Schimel, J. P., and Holden, P. A.: Controls on microbial CO₂ production: a comparison of
381 surface and subsurface soil horizons, Glob. Change Biol., 9, 1322–1332, 2003.

382 Flores, B. M., Montoya, E., Sakschewski, B., Nascimento, N., Staal, A., Betts, R. A., Levis, C., Lapola, D. M.,
383 Esquivel-Muelbert, A., Jakovac, C., Nobre, C. A., Oliveira, R. S., Borma, L. S., Nian, D., Boers, N., Hecht, S. B.,
384 Ter Steege, H., Arieira, J., Lucas, I. L., Berenguer, E., Marengo, J. A., Gatti, L. V., Mattos, C. R. C., and Hirota, M.:
385 Critical transitions in the Amazon forest system, Nature, 626, 555–564, [https://doi.org/10.1038/s41586-023-06970-](https://doi.org/10.1038/s41586-023-06970-0)
386 0, 2024.

387 Fontaine, S., Barot, S., Barre, P., Bdioui, N., Mary, B., and Rumpel, C.: Stability of organic carbon in deep soil layers
388 controlled by fresh carbon supply, Nature, 450, 277–U10, 2007.

389 Friedlingstein, P., O’Sullivan, M., Jones, M. W., Andrew, R. M., Bakker, D. C. E., Hauck, J., Landschützer, P., Le
390 Quéré, C., Luijkx, I. T., Peters, G. P., Peters, W., Pongratz, J., Schwingshackl, C., Sitch, S., Canadell, J. G., Ciais, P.,
391 Jackson, R. B., Alin, S. R., Anthoni, P., Barbero, L., Bates, N. R., Becker, M., Bellouin, N., Decharme, B., Bopp, L.,
392 Brasika, I. B. M., Cadule, P., Chamberlain, M. A., Chandra, N., Chau, T.-T.-T., Chevallier, F., Chini, L. P., Cronin,
393 M., Dou, X., Enyo, K., Evans, W., Falk, S., Feely, R. A., Feng, L., Ford, D. J., Gasser, T., Ghattas, J., Gkritzalis, T.,
394 Grassi, G., Gregor, L., Gruber, N., Gürses, Ö., Harris, I., Hefner, M., Heinke, J., Houghton, R. A., Hurtt, G. C., Iida,
395 Y., Ilyina, T., Jacobson, A. R., Jain, A., Jarníková, T., Jersild, A., Jiang, F., Jin, Z., Joos, F., Kato, E., Keeling, R. F.,
396 Kennedy, D., Klein Goldewijk, K., Knauer, J., Korsbakken, J. I., Körtzinger, A., Lan, X., Lefèvre, N., Li, H., Liu, J.,
397 Liu, Z., Ma, L., Marland, G., Mayot, N., McGuire, P. C., McKinley, G. A., Meyer, G., Morgan, E. J., Munro, D. R.,
398 Nakaoka, S.-I., Niwa, Y., O’Brien, K. M., Olsen, A., Omar, A. M., Ono, T., Paulsen, M., Pierrot, D., Pockock, K.,
399 Poulter, B., Powis, C. M., Rehder, G., Resplandy, L., Robertson, E., Rödenbeck, C., Rosan, T. M., Schwinger, J.,
400 Séférian, R., et al.: Global Carbon Budget 2023, *Earth Syst. Sci. Data*, 15, 5301–5369,
401 <https://doi.org/10.5194/essd-15-5301-2023>, 2023.

402 Galloway, J. N., Townsend, A. R., Erisman, J. W., Bekunda, M., Cai, Z., Freney, J. R., Martinelli, L. A., Seitzinger, S. P.,
403 and Sutton, M. A.: Transformation of the Nitrogen Cycle: Recent Trends, Questions, and Potential Solutions,
404 *Science*, 320, 889–892, <https://doi.org/10.1126/science.1136674>, 2008.

405 Hashimoto, S., Ito, A., and Nishina, K.: Divergent data-driven estimates of global soil respiration, *Commun. Earth*
406 *Environ.*, 4, 460, <https://doi.org/10.1038/s43247-023-01136-2>, 2023.

407 He, Y., Ding, J., Dorji, T., Wang, T., Li, J., and Smith, P.: Observation□based global soil heterotrophic respiration
408 indicates underestimated turnover and sequestration of soil carbon by terrestrial ecosystem models, *Glob. Change*
409 *Biol.*, 28, 5547–5559, <https://doi.org/10.1111/gcb.16286>, 2022.

410 Jian, J., Vargas, R., Anderson-Teixeira, K., Stell, E., Herrmann, V., Horn, M., Kholod, N., Manzon, J., Marchesi, R.,
411 Paredes, D., and Bond-Lamberty, B.: A restructured and updated global soil respiration database (SRDB-V5), *Earth*
412 *Syst. Sci. Data*, 13, 255–267, <https://doi.org/10.5194/essd-13-255-2021>, 2021.

413 Karhu, K., Auffret, M. D., Dungait, J. A. J., Hopkins, D. W., Prosser, J. I., Singh, B. K., Subke, J.-A., Wookey, P. A.,
414 Ågren, G. I., Sebastià, M.-T., Gouriveau, F., Bergkvist, G., Meir, P., Nottingham, A. T., Salinas, N., and Hartley, I.
415 P.: Temperature sensitivity of soil respiration rates enhanced by microbial community response, *Nature*, 513, 81–84,
416 <https://doi.org/10.1038/nature13604>, 2014.

417 Lin, Y., Campbell, A. N., Bhattacharyya, A., DiDonato, N., Thompson, A. M., Tfaily, M. M., Nico, P. S., Silver, W. L.,
418 and Pett-Ridge, J.: Differential effects of redox conditions on the decomposition of litter and soil organic matter,
419 *Biogeochemistry*, 154, 1–15, <https://doi.org/10.1007/s10533-021-00790-y>, 2021.

420 Linn, D. M. and Doran, J. W.: Effect of Water-Filled Pore Space on Carbon Dioxide and Nitrous Oxide Production in
421 Tilled and Nontilled Soils, *Soil Sci. Soc. Am. J.*, 48, 1267–1272,
422 <https://doi.org/10.2136/sssaj1984.03615995004800060013x>, 1984.

423 Malhi, Y., Wood, D., Baker, T. R., Wright, J., Phillips, O. L., Cochrane, T., Meir, P., Chave, J., Almeida, S., Arroyo, L.,
424 Higuchi, N., Killeen, T. J., Laurance, S. G., Laurance, W. F., Lewis, S. L., Monteagudo, A., Neill, D. A., Vargas, P.
425 N., Pitman, N. C. A., Quesada, C. A., Salomao, R., Silva, J. N. M., Lezama, A. T., Terborgh, J., Martinez, R. V., and
426 Vinceti, B.: The regional variation of aboveground live biomass in old-growth Amazonian forests, *Glob. Change*
427 *Biol.*, 12, 1107–1138, <https://doi.org/10.1111/j.1365-2486.2006.01120.x>, 2006.

428 Marengo JA, Cunha AP, Espinoza JC, Fu R, Schöngart J, Jimenez JC, Costa MC, Ribeiro JM, Wongchuig S, Zhao S.:
429 The Drought of Amazonia in 2023-2024. *Am. J. Climate Change* 13(03): 567-597, 2024.

430 Montes, C. R., Lucas, Y., Pereira, O. J. R., Achard, R., Grimaldi, M., and Melfi, A. J.: Deep plant-derived carbon
431 storage in Amazonian podzols, 8, 113–120, <https://doi.org/10.5194/bg-8-113-2011>, 2011.

432 Montes, C. R., Merdy, P., Da Silva, W. T. L., Ishida, D., Melfi, A. J., Santin, R. C., and Lucas, Y.: Mineralization of soil
433 organic matter from equatorial giant podzols submitted to drier pedoclimate: A drainage topochronosequence study,
434 *CATENA*, 222, 106837, <https://doi.org/10.1016/j.catena.2022.106837>, 2023.

435 Moorhead, D. L. and Sinsabaugh, R. L.: A theoretical model of litter decay and microbial interaction, *Ecol. Monogr.*,
436 76, 151–174, [https://doi.org/10.1890/0012-9615\(2006\)076\[0151:ATMOLD\]2.0.CO;2](https://doi.org/10.1890/0012-9615(2006)076[0151:ATMOLD]2.0.CO;2), 2006.

437 Moyano, F. E., Vasilyeva, N., Bouckaert, L., Cook, F., Craine, J., Curiel Yuste, J., Don, A., Epron, D., Formanek, P.,
438 Franzluebbers, A., Ilstedt, U., Kätterer, T., Orchard, V., Reichstein, M., Rey, A., Ruamps, L., Subke, J.-A., Thomsen,
439 I. K., and Chenu, C.: The moisture response of soil heterotrophic respiration: interaction with soil properties,
440 *Biogeosciences*, 9, 1173–1182, <https://doi.org/10.5194/bg-9-1173-2012>, 2012.

441 Quesada, C. A., Lloyd, J., Anderson, L. O., Fyllas, N. M., Schwarz, M., and Czimczik, C. I.: Soils of Amazonia with
442 particular reference to the RAINFOR sites, *Biogeosciences*, 8, 1415–1440, <https://doi.org/10.5194/bg-8-1415-2011>,
443 2011.

444 Rovira, A. and Greacen, E.: The effect of aggregate disruption on the activity of microorganisms in the soil, *Aust. J.*
445 *Agric. Res.*, 8, 659, <https://doi.org/10.1071/AR9570659>, 1957.

446 Ruamps, L. S., Nunan, N., and Chenu, C.: Microbial biogeography at the soil pore scale, *Soil Biol. Biochem.*, 43, 280–
447 286, <https://doi.org/10.1016/j.soilbio.2010.10.010>, 2011.

448 Salome, C., Nunan, N., Pouteau, V., Lerch, T. Z., and Chenu, C.: Carbon dynamics in topsoil and in subsoil may be
449 controlled by different regulatory mechanisms, *Glob. Change Biol.*, 16, 416–426, 2010.

450 Sierra, C. A., Jiménez, E. M., Reu, B., Peñuela, M. C., Thuille, A., and Quesada, C. A.: Low vertical transfer rates of
451 carbon inferred from radiocarbon analysis in an Amazon Podzol, *Biogeosciences*, 10, 3455–3464,
452 <https://doi.org/10.5194/bg-10-3455-2013>, 2013.

453 Vaughn, L. J. S. and Torn, M. S.: ¹⁴C evidence that millennial and fast-cycling soil carbon are equally sensitive to
454 warming, *Nat. Clim. Change*, 9, 467–471, <https://doi.org/10.1038/s41558-019-0468-y>, 2019.

455 Xiong, Y., D’Atri, J. J., Fu, S., Xia, H., and Seastedt, T. R.: Rapid soil organic matter loss from forest dieback in a
456 subalpine coniferous ecosystem, *Soil Biol. Biochem.*, 43, 2450–2456, <https://doi.org/10.1016/j.soilbio.2011.08.013>,
457 2011.

458 Zhou, L., Zhou, X., Zhang, B., Lu, M., Luo, Y., Liu, L., and Li, B.: Different responses of soil respiration and its
459 components to nitrogen addition among biomes: a meta-analysis, *Glob. Change Biol.*, 20, 2332–2343,
460 <https://doi.org/10.1111/gcb.12490>, 2014.

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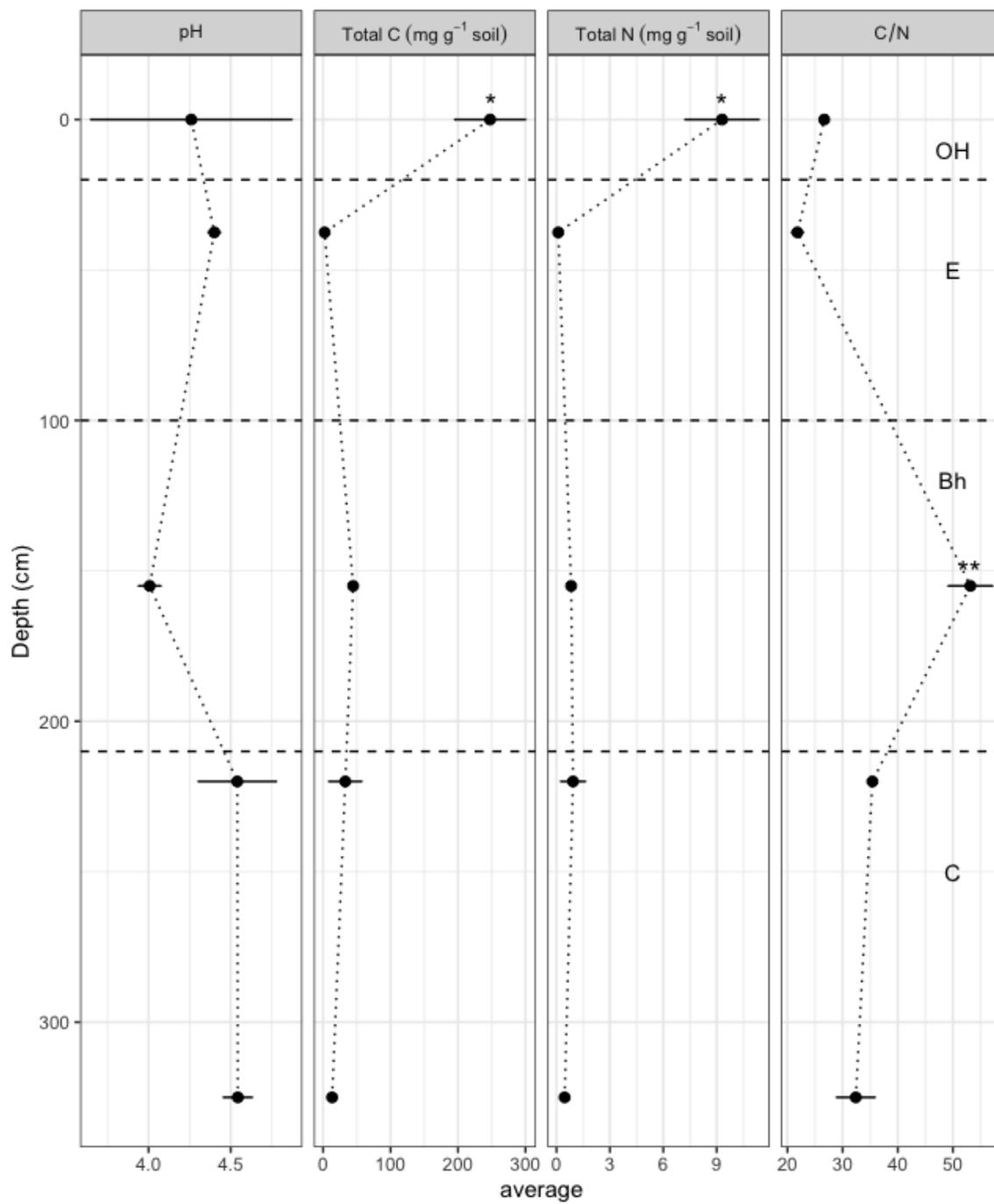
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496 Figure 1 Properties of different horizons of the Podzols. The bars indicate the standard deviation of
 497 the mean where it is larger than the size of the symbols. Stars indicate horizons with values that are
 498 significantly (*<0.05, **<0.01) higher than in other horizons.

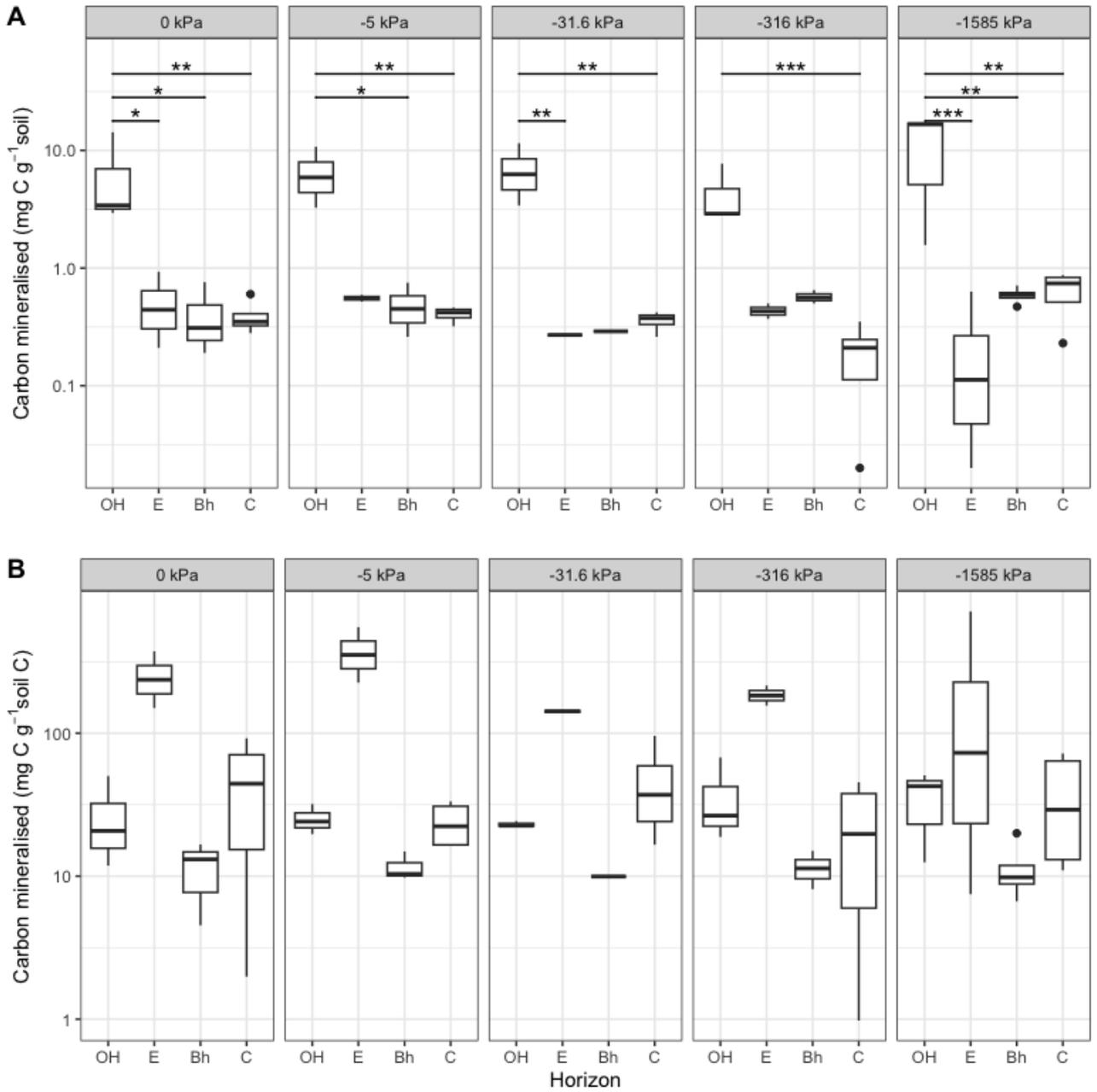
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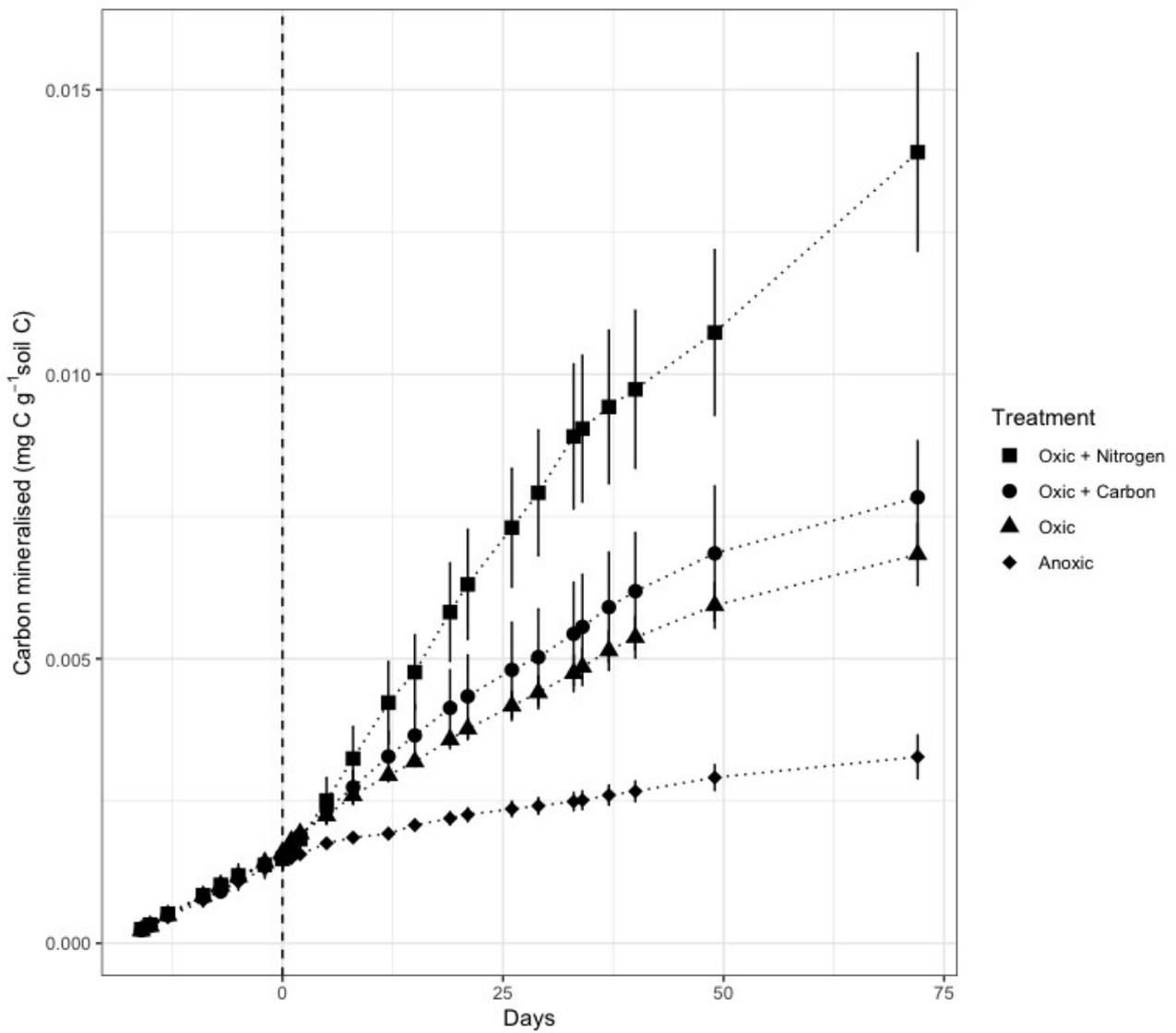
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Figure 2 Total C mineralisation in the different Podzol horizons at different matric potentials during the 68 day incubation. Mineralisation is expressed per g soil (A) and per g soil C (B). Note that the y-axes are in log scale. Across all matric potentials, the carbon mineralised (per g soil) from the OH horizon was significantly ($P < 0.001$) higher than in all other horizons. Differences within

509 each matric potential are shown in the graph (*<0.05, **<0.01, ***<0.001). The carbon mineralised
510 per g soil C from the E horizon was significantly ($P<0.001$) higher than in all other horizons, but
511 showed no significant differences within each matric potential.

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519 | Figure 3. Soil organic C mineralisation in Podzol Bh horizons prior to and after imposition of
520 | treatments (dashed line indicates day at which treatments commenced). Bars indicate standard error
521 | of the mean where error is larger than the size of the symbols. At the end of the incubation, the
522 | amount of C mineralised in the oxic+N treatment was significantly higher than the oxic and the
523 | oxic+C treatments ($P<0.05$) as well as the anoxic treatment ($P<0.001$). Both oxic treatments
524 | without N were also significantly ($P<0.01$) higher than the anoxic treatment.

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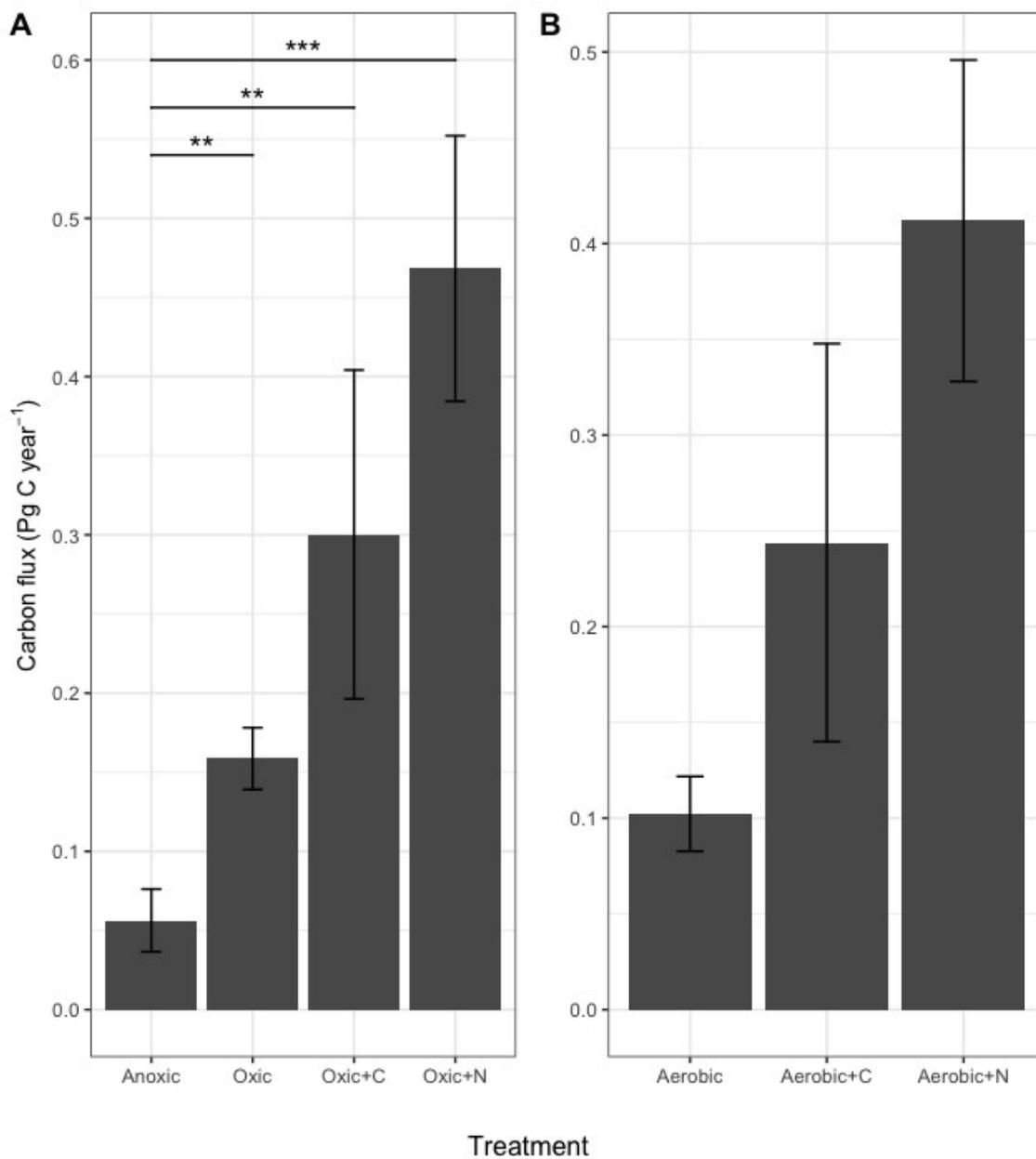
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535 | Figure 4 Estimated annual carbon flux from Bh horizons to the atmosphere (A) and estimated
 536 | increase in C flux to the atmosphere if the present anoxic conditions were to change (B). The bars
 537 | indicate the standard error of the mean. Only the oxic+-N and the oxic+C treatments resulted in
 538 | significantly (** $P < 0.015$) more C flux than the anoxic treatment.

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