



1     **Contrasting Inland-Coastal Aerosol Mixing States: An Entropy-Based**

2     **Metric for CCN Activity**

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## 24 Abstract

25 Simplified assumptions of aerosol hygroscopic mixing states in modeling studies  
26 often introduce substantial uncertainties in estimating cloud condensation nuclei (CCN)  
27 concentrations and their climatic impacts. This study systematically investigates the  
28 contrasting relationships between mixing states and CCN activity by comparing  
29 ambient measurements from inland and coastal sites. We show distinct seasonal  
30 variations of the particles mixing state. In winter, externally mixed particles dominated  
31 both sites, with comparable mixing state indices ( $\chi$ ) of  $0.38 \pm 0.12$  and  $0.39 \pm 0.09$   
32 respectively for coastal air masses and inland air. However, summer measurements  
33 showed pronounced differences: photochemical processes promoted significantly  
34 higher internal mixing in coastal aerosols ( $\chi = 0.69 \pm 0.19$ ), whereas inland  $\chi$  values only  
35 increased moderately to  $0.47 \pm 0.12$ . A universal logarithmic correlation was identified  
36 between the critical diameter ( $D_{\text{cri}}$ ) characterizing CCN activity and  $\chi$  ( $D_{\text{cri}} = -$   
37  $32.15 \ln(\chi) + 84.71$ , Pearson  $r = -0.74$ ), but with distinct decrement rates for coastal vs.  
38 inland aerosols. Our further quantitative analysis reveals a 0.1 increase in  $\chi$  enhanced  
39 winter CCN concentrations ( $N_{\text{CCN}}$ ) by 39–65% under typical cloud supersaturations,  
40 whereas this effect diminished to ~9% in summer. These results underscore that mixing  
41 states exert more pronounced control over  $N_{\text{CCN}}$  in diverse environments. Our work  
42 provides critical constraints for parameterizing fine aerosols CCN activity in climate  
43 models, thereby reducing uncertainties in aerosol–climate effect estimations.



## 44 1. Introduction

45 Atmospheric cloud condensation nuclei (CCNs) are complex mixtures of organic  
46 and inorganic components. Their chemical and physical properties make quantifying  
47 aerosol-cloud interactions challenging (Liu et al., 2018; Rosenfeld et al., 2019; Xu et  
48 al., 2022, 2024; Virtanen et al., 2025), introducing uncertainties into climate effect  
49 assessments (Charlson et al., 1992; Shrivastava et al., 2017; IPCC, 2021; Manavi et al.,  
50 2025; Chen et al., 2022). Accurate climate model predictions of aerosol impacts require  
51 understanding aerosol mixing states under different atmospheric conditions and their  
52 effects on CCN activity (Ching et al., 2016; Zheng et al., 2021). Current models often  
53 oversimplify mixing states by assuming pure internal or external mixing (Winkler, 1973;  
54 Zheng et al., 2021; Stevens et al., 2019; Riemer et al., 2019). This is problematic  
55 because mixing states directly determine particle hygroscopicity and CCN estimates  
56 (Wang et al., 2010; Ren et al., 2018). For example, CCN activity for internal-mixed  
57 aerosols rely more on inorganic components, while external mixtures are more sensitive  
58 to organic matter (Ren et al., 2018; Bhattu et al., 2015). Such simplifications can lead  
59 to significant errors, e.g., Sotiropoulou et al. (2007) found that mixing state assumptions  
60 caused two-fold  $N_{CCN}$  estimation errors in global models.

61 Systematic observations across diverse environments are critical because aerosol  
62 mixing states exhibit pronounced spatial-temporal variations (Ye et al., 2018; Liu et al.,  
63 2025; Hughes et al., 2018). For example, continental and coastal regions present  
64 contrasting scenarios (Ramachandran et al., 2016). The continental areas are dominated  
65 by anthropogenic emissions, where aerosol aging is driven by industrial and traffic-



66 related pollutants (Huang et al., 2014; Ren et al., 2023). Particles here undergo  
67 progressive internal mixing via photochemical reactions and coagulation, altering their  
68 hygroscopic properties (Ervens et al., 2010). While the coastal regions feature dynamic  
69 interactions between marine aerosols (e.g., sea salt) and continental pollutants (Schill  
70 et al., 2015; Collins et al., 2013; Cheung et al., 2020). Seasonal shifts in air mass sources  
71 (e.g., marine vs. continental dominance) create unique mixing state patterns (Xu et al.,  
72 2020, 2021a). For instance, summer photochemical processes in coastal areas can  
73 enhance internal mixing, while winter often retains more external mixing due to stable  
74 atmospheric conditions.

75       However, the aerosols in continental and coastal regions have distinct climate  
76 feedback mechanisms (Bellouin et al., 2019; Pan et al., 2022; Gong et al., 2023). The  
77 continental aerosols influence regional cloud formation, while coastal aerosols affect  
78 marine boundary layer clouds that are key components of global climate systems (Liu  
79 et al., 2018). But the current models lack regional-specific mixing state parameters and  
80 usually assume uniform mixing in both environments. This could lead to large  
81 uncertainties in predicting CCN concentrations, highlighting the need for site-specific  
82 observations.

83       Recent studies have used the mixing state index ( $\chi$ ) to characterize aerosol  
84 heterogeneity (Zheng et al., 2021; Ching et al., 2017; Yuan et al., 2023), but cross-  
85 environment comparisons remain limited. By integrating inland and coastal  
86 measurements, this study will focus on addressing two key gaps, (1) How continental  
87 vs. marine-dominated environments shape aerosol mixing states and CCN activity; (2)

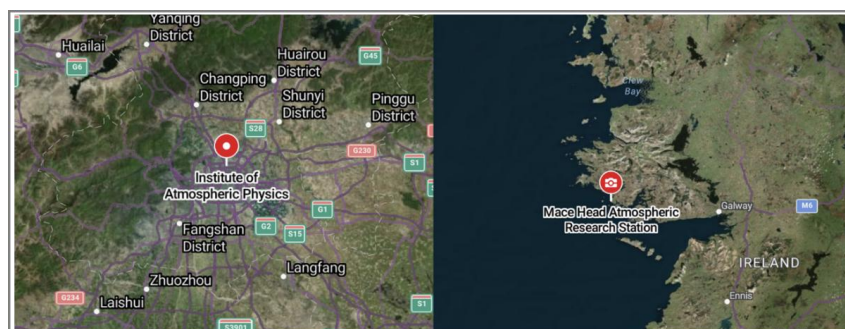


88 Whether  $\chi$ -based CCN parameterizations show regional dependencies, providing  
89 critical constraints for climate models.

## 90 2. Data and Methods

### 91 2.1 Field Campaigns

92 The inland atmospheric measurements were conducted for two periods from 16  
93 November to 6 December and 29 May to 13 June, respectively in urban Beijing, at the  
94 Institute of Atmospheric Physics, Chinese Academy of Sciences (IAP, 39.97° N,  
95 116.37° E). This urban site exhibited highly variable aerosol populations dominated by  
96 local anthropogenic sources including vehicular, cooking emissions, and residential  
97 heating. Coastal measurements were performed at the Mace Head atmospheric research  
98 station (MHD, 53.33° N, 9.90° W) from 1 November 2009 to 30 January 2010 and 11  
99 to 31 August 2010, which located on the west coast of Ireland. Aerosol particles here  
100 experience alternating influences from polluted continental and clean marine  
101 atmospheres. The map of the sites was shown in Figure 1. More details about the  
102 campaigns were given in Fan et al. (2020) and Xu et al. (2021a).



103  
104 **Fig 1.** Map of the sites in the Inland of the Institute of Atmospheric Physics (IAP) and



105 Coastal of Mace Head (MHD). (© Google Maps, <https://maps.google.com/>, last access:  
106 2 April 2025).

## 107 2.2 Instrumentation

### 108 Hygroscopicity measurements

109 The particle hygroscopicity at both sites was characterized using the humidified  
110 tandem differential mobility analyzer (HTDMA). The hygroscopic growth factor (Gf),  
111 defined as the ratio of the particle diameter at the fixed RH (90%) and dry diameter set  
112 in this study for 40, 80, 110, 150, 200 nm at IAP and 35, 50, 75, 110 and 165 nm at  
113 MHD, respectively. The Gf probability density function (Gf-PDF) was derived using  
114 the TDMAinv algorithm (Gysel et al., 2009).

115 Here for each particle size, the hygroscopicity parameter  $\kappa$  can be subsequently  
116 calculated using  $\kappa$ -Köhler theory (Petters and Kreidenweis, 2007):

$$117 \quad \kappa = (Gf^3 - 1) \cdot \left[ \frac{1}{RH} \exp\left(\frac{4\sigma_s M_w}{RT \rho_w D_d Gf}\right) - 1 \right] \quad (1)$$

118 where RH is the HTDMA relative humidity (90% set in the instrument),  $\sigma_{s/a}$  is the  
119 surface tension of pure water (0.072 mN m<sup>-1</sup>),  $M_w$  and  $\rho_w$  are the molecular weight and  
120 the density of pure water,  $R$  is the gas constant, and  $T$  is the absolute temperature,  $D_d$  is  
121 the droplet diameter.

122 Then, the  $\kappa$ -PDF is obtained and normalized as  $\int_0^\infty c(\kappa) d\kappa = 1$ , where  $c(\kappa)$  is  
123 normalized as  $\kappa$ -PDF. Further it was used to calculate the particle population  
124 heterogeneity (Calculation seen in Section 2.3).

### 125 Chemical components



126 For the inland atmospheric measurements, the non-refractory submicron aerosol  
127 (smaller than  $1\mu\text{m}$ , NR-PM<sub>1</sub>) chemical composition was quantitatively characterized  
128 using the Aerodyne High-Resolution Time-of-Flight Aerosol Mass Spectrometer (HR-  
129 ToF-AMS) (DeCarlo et al., 2006), including sulfate ( $\text{SO}_4^{2-}$ ), nitrate ( $\text{NO}_3^-$ ), ammonium  
130 ( $\text{NH}_4^+$ ), chloride (ChL) and organics (Org). The black carbon (BC) mass concentration  
131 was determined from the light absorption with a seven-wavelength aethalometer (AE33,  
132 Magee Scientific Corp.).

133 Measurements of PM<sub>1</sub> in the coastal atmosphere were also performed by the HR-  
134 ToF-AMS, including major inorganic salts (non-sea-salt sulfate, nss- $\text{SO}_4^{2-}$ ;  
135 methanesulfonic acid, MSA;  $\text{NO}_3^-$ ;  $\text{NH}_4^+$ ) and organic matter. The instrument operation  
136 and calibration have been described in previous studies (Ovadnevaite et al., 2014; Xu  
137 et al., 2019).

#### 138 **Aerosol number size distribution and CCN number concentration**

139 Particle number size distributions (PNSD) were measured using an integrated  
140 system consisting of a Differential Mobility Analyzer (DMA; model 3081, TSI Inc.)  
141 coupled with a Condensation Particle Counter (CPC; model 3772, TSI Inc.). During the  
142 measurements at IAP, the PNSD covered the size range of 10-550 nm with a 5-minute  
143 time resolution. It scanned size range of 20-500 nm at MHD with a 10-minute temporal  
144 resolution. The CCN number concentrations were quantified at both sites using a  
145 Droplet Measurement Technologies CCN counter (DMT-CCNc) (Lance et al., 2006).  
146 The instrument's supersaturation (SS) settings were carefully calibrated before and after  
147 each campaign using ammonium sulfate aerosol following Rose et al. (2008).



## 2.3 Calculation the heterogeneity for aerosol particles

To characterize the heterogeneous distribution of the hygroscopic and non-hygroscopic components in populations (Chen et al., 2022), we calculated the mixing state index ( $\chi$ ) using the  $\kappa$ -PDF, following the methodology of Yuan et al. (2023). Two surrogate groups in a population of  $N$  aerosol particles were assumed (Zheng et al., 2021). One surrogate group consists the non-hygroscopic species with  $\kappa_{NH}$  of 0.01 and another group contains the hygroscopic species with  $\kappa_H$  of 0.6 (Yuan et al., 2023; Ching et al., 2017). At the coastal MHD site, we accounted for the enhanced hydrophilicity of marine aerosols by additionally testing  $\kappa_H$  values of 0.7 and 0.8 (Fig. S1). While these variations in  $\kappa_H$  introduced a mean uncertainty of 8% in  $\chi$  values, it did not significantly affect the seasonal or site comparisons. The volume fraction of two surrogate groups can be calculated based on the total  $\kappa$  according to the Zdanovskii–Stokes–Robinson (ZSR) mixing rule (Zdanovskii, 1948; Stokes et al., 1966).

The mixing state index  $\chi$  is defined as the affine ratio of the average particle species diversity ( $D\alpha$ ) and population species diversity ( $D\gamma$ ) as:

$$\chi = \frac{D\alpha - 1}{D\gamma - 1} \quad (2)$$

The average per-particle species diversity  $D\alpha$  can be calculated as follows. First, the mixing entropies at bin  $i$  ( $H_i$ ) are determined according to equation (3),

$$H_i = -P_{i,NH} \times \ln P_{i,NH} - P_{i,H} \times \ln P_{i,H} \quad (3)$$

where  $P_{i,NH}$  and  $P_{i,H}$  are the volume fraction of each group for the  $\kappa$ -PDF with  $X$  bins at bin  $i$  ( $i=1,2,\dots,X$ ), and can be determined from the  $P_{i,NH} + P_{i,H} = 1$  and  $P_{i,NH} \times \kappa_{NH} + P_{i,H} \times \kappa_H = \kappa_i$ . Here  $\kappa_{NH} = 0.01$ ,  $\kappa_H = 0.6$ ;  $\kappa_i$  represents the





170 hygroscopicity parameter at bin  $i$ .

171 Based on the assumption that particles in the same diameter have the same mixing

172 entropy  $H_\alpha = \sum_{j=1}^N P_j \times H_j$ ,  $P_j = \frac{V_j}{V_{total}} = \frac{1}{N}$ ; the per-particle mixing entropies  $H_\alpha$  is

173 determined according to equation (4),

174 
$$H_\alpha = \sum_{i=1}^X H_i \times c(\kappa)_i \times \Delta\kappa \quad (4)$$

175 where  $c(\kappa)_i$  is the probability density of the normalized  $\kappa$ -PDF at bin  $i$ , and  $\Delta\kappa$

176 represents the bin width. Then, the average per-particle species diversity  $D_\alpha$  can be

177 determined as  $D_\alpha = e^{H_\alpha}$ ;

178 The bulk population species diversity  $D_\gamma$  can be calculated as follows. First, the

179 aerosol population of the mixing entropy can be calculated as equation (5):

180 
$$H_\gamma = -P_{NH} \times \ln P_{NH} - P_H \times \ln P_H \quad (5)$$

181 where  $P_{NH}$  and  $P_H$  are the volume fraction of the non-hygroscopic and hygroscopic

182 components in the population, and can be calculated by equation (6) and (7):

183 
$$P_{NH} = \sum_{i=1}^X P_{i,NH} \times c(\kappa)_i \times \Delta\kappa \quad (6)$$

184 
$$P_H = \sum_{i=1}^X P_{i,H} \times c(\kappa)_i \times \Delta\kappa \quad (7)$$

185 Then, the bulk population species diversity  $D_\gamma$  can be determined as  $D_\gamma = e^{H_\gamma}$ .

186 Here, the definition of surrogate species as supersets encompassing hygroscopicity

187 heterogeneity implies that the heterogeneity parameter  $\chi$  ranges from 0 to 1. When the

188 mixing index  $\chi$  approaches 0, it indicates a completely segregated state where

189 hygroscopic and non-hygroscopic species reside in distinct particles. While for the case

190 the mixing index  $\chi$  to be 1 represents that the non-hygroscopic and hygroscopic species

191 distributing homogeneously throughout the aerosol population.

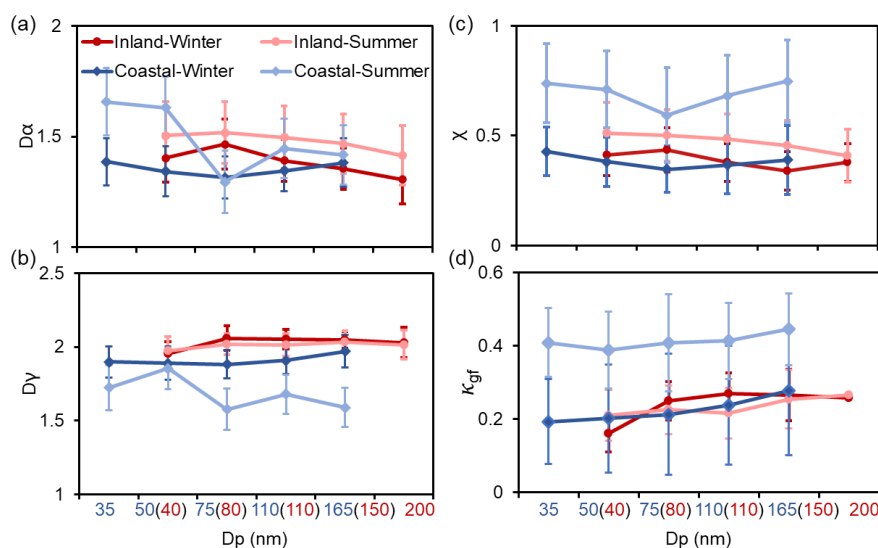


### 192 3. Result and Discussion

#### 193 3.1 Comparison of the heterogeneity in the inland and coastal atmosphere

194 To characterize the hygroscopic heterogeneity of atmospheric aerosols, Figure 2  
195 depicts variations in mixing state metrics ( $D\alpha$ ,  $D\gamma$ ,  $\chi$ ) and the hygroscopic parameter  
196 ( $\kappa_{gf}$ ) across particle size distributions. For inland aerosols,  $D\alpha$  and  $\chi$  decrease with  
197 increasing particle diameter, accompanied by higher  $\kappa_{gf}$  values. This trend indicates that  
198 inland particle populations tend to homogenize into hygroscopic compositions through  
199 primary particle aging or secondary formation processes (Liu et al., 2025; Chen et al.,  
200 2022; Zhong et al., 2022). In contrast, coastal particles exhibit a non-monotonic pattern:  
201  $D\alpha$  and  $\chi$  decrease for Aitken-mode particles ( $<100$  nm) but increase for accumulation-  
202 mode particles. The  $\kappa_{gf}$  shows consistent size-dependent increases in both winter and  
203 summer campaigns.

204 Notably, the mixing state metrics exhibit a pronounced minimum at 75 nm  
205 particles, influenced by distinct mechanisms: winter minima reflect the high sea salt  
206 fraction, while summer minima are driven by anthropogenic organic matter (Cheung et  
207 al., 2020; Xu et al., 2021a). Lower winter  $\chi$  values—coupled with broader  $\kappa$ -PDF  
208 distributions—indicate stronger external mixing and compositional diversity compared  
209 to summer (Fig. S2). Seasonal  $\chi$  and  $\kappa_{gf}$  disparities are more pronounced at the coastal  
210 site, primarily driven by the seasonal alternation of marine and anthropogenic emission  
211 sources.

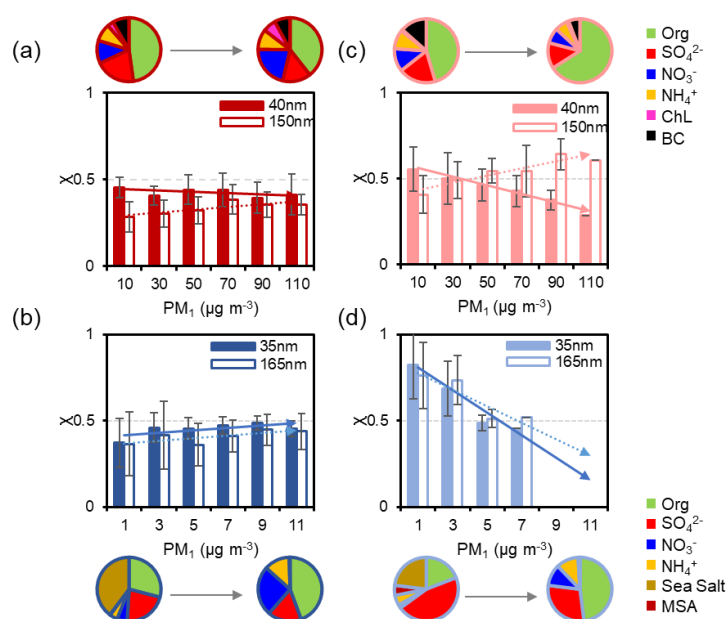


**Fig 2.** Mean values of the  $D\alpha$  (a),  $D\gamma$  (b),  $\chi$  (c) and  $\kappa_{gf}$  (d) for aerosols of five diameters during winter and summer periods in Inland (IAP) and Coastal (Mace Head) sites.

Ultrafine particles (40 nm inland vs. 35 nm coastal, Aitken mode) and larger particles (150 nm inland vs. 165 nm coastal, accumulation mode) were selected to investigate distinct evolutionary processes of aerosol heterogeneity (Fig. 3 and Fig. S3). With the increasing of PM concentration during winter, the variation in  $\chi$  values exhibit only minor both at the inland and coastal sites, generally fluctuating between approximately  $-0.04$  and  $0.08$  (Fig. 3a and b). Inland accumulation-mode particles show a modest increase in  $\chi$ , corresponding with a higher proportion of inorganic salts. Conversely, at coastal sites, the composition fraction shifts from a sea-salt dominance toward organic matter, accompanied by a  $\sim 20\%$  increase in nitrate content (Fig. 3b). In summer, the variation of  $\chi$  with PM concentration becomes markedly pronounced at both inland and coastal stations. For example,  $\chi$  for 40 nm particles decreases as PM



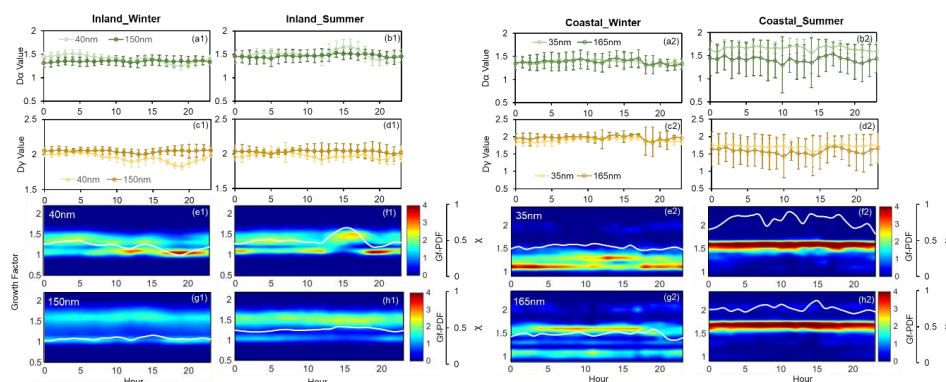
226 increases at inland sites (Fig. 3c). The elevated particle heterogeneity mainly arises  
227 from the locally primary emissions and photochemically driven new particle formation.  
228 In contrast,  $\chi$  for 150 nm particles increases from  $\sim 0.40$  to  $\sim 0.60$  with rising PM,  
229 reflecting enhanced secondary formation and internal mixing during pollution process  
230 that render the particle population more homogeneous. At coastal sites,  $\chi$  declines with  
231 rising PM by approximately 0.37 for 35 nm particles and 0.24 for 165 nm particles,  
232 mirroring the shift in chemical composition makeup from inorganic dominance to  
233 greater organic content (Fig. 3d).



234  
235 **Fig 3.** Variation of the average  $\chi$  of 40 nm and 150 nm in inland and 35 nm and 165 nm  
236 in coastal site with the particle mass concentration in Inland-winter (a), Inland-summer  
237 (b), Coastal-winter (c) and Coastal-summer (d). The pie charts represent the average  
238 mass fraction during four field campaigns.



Figure 4 illustrates pronounced diurnal variations in mixing state metrics ( $D\alpha$ ,  $D\gamma$ , Gf-PDF,  $\chi$ ) between inland and coastal atmospheres. In the inland atmosphere, winter exhibited steeper declines in  $D\alpha$  and  $\chi$  during evening rush hours than summer, indicating a higher fraction of non-hygroscopic particles (40 nm) from fresh traffic emissions (Fig. 4a1). Concurrently, reduced  $D\gamma$  values suggest that the bulk population consists of uniformly distributed less-hygroscopic (LH) components (Fig. 4c1). Aitken mode particles showed bimodal and broader Gf-PDF distributions, corresponding to cooking activities (11:00–13:00 LT) and traffic peaks (17:00–20:00 LT) (Cai et al., 2020). Midday photochemical aging promoted more internally mixed aerosols (Yang et al., 2012; Liu et al., 2025), as evidenced by increasing  $D\alpha$  at the urban site (Fig. 4b1). Conversely, accumulation-mode particles showed minimal diurnal variations, suggesting stable relative proportions of LH and more-hygroscopic (MH) components in inland aerosols across seasons.



**Fig 4.** The variation of  $D\alpha$ ,  $D\gamma$ , Gf-PDF, and  $\chi$  during winter and summer periods for 40 nm and 150 nm aerosols in Inland (a1-h1) and for 35 nm and 165 nm aerosols Coastal site (a2-h2).



255 For the coastal atmosphere, the mixing state metrics ( $D\alpha$ ,  $D\gamma$ , and  $\chi$ ) of Aitken and  
256 accumulation mode particles in winter exhibited analogous diurnal patterns,  
257 characterized by a descending trend at nightfall. This corresponds to an enhanced modal  
258 distribution of near-hydrophobic (NH) particles at 35 nm and more-hygroscopic (MH)  
259 particles at 165 nm. In summer,  $D\alpha$  and  $D\gamma$  both trended downward during daytime,  
260 with the decline of  $D\gamma$  being more pronounced. A conspicuous seasonal discrepancy  
261 between Aitken and accumulation mode particles was observed in this region (Fig. 4a2–  
262 h2), where the mixing state index  $\chi$  increased incrementally from winter to summer.  
263 Specifically, the mean  $\chi$  for 35 nm particles escalated from 0.42 to 0.80, and for 165  
264 nm particles, it rose from 0.39 to 0.76. This trend demonstrates a strong alignment with  
265 the spread factor documented by Xu et al. (2021a, b).

266 The Gf-PDF diurnal profiles of Aitken mode particles displayed a bimodal and  
267 broadened distribution, corresponding to a less-hygroscopic (LH) mode of biogenic  
268 origin during nighttime and a more-hygroscopic (MH) mode dominated by sea salt  
269 (comprising 55% number fraction) during daytime. Analogously, accumulation mode  
270 particles exhibited bimodal distributions with a higher proportion of MH mode during  
271 daytime, primarily attributed to the prevalence of sea salt and non-sea-salt sulfate (nss-  
272 sulfate) in the coastal atmosphere (Xu et al., 2020). In contrast, summer observations  
273 revealed that Gf-PDFs of both Aitken and accumulation mode particles transitioned to  
274 unimodal distributions, signifying more homogeneous mixing of LH and MH  
275 components within individual particles. This uniformity is linked to processes including  
276 sulfuric acid condensation, admixture of sulfate with biogenic organic matter (Xu et al.,



277 2021a), as well as photochemical oxidation and atmospheric aging (Jimenez et al.,  
278 2009).

### 279 3.2 Dependence of the aerosol properties on the mixing state

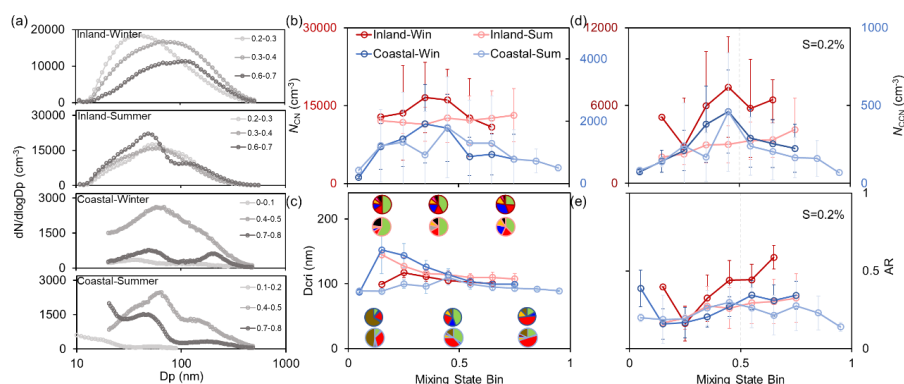
280 The mixing state of particle populations undergoes dynamic transformations  
281 during atmospheric aging, profoundly influencing their CCN activity. Unlike prior  
282 studies that assumed mixing states based on chemical component fractions (Yang et al.,  
283 2012; Padró et al., 2012; Ren et al., 2018), this work employs the entropy-derived  
284 mixing state index  $\chi$ , which quantifies the distribution of hygroscopic and non-  
285 hygroscopic species (Zheng et al., 2021; Ching et al., 2017). We systematically  
286 investigate how aerosol properties evolve with changing  $\chi$ . Figure 5 illustrates the  
287 dependency of aerosol characteristics on  $\chi$  (ranging from 0 to 1 in 0.1 increments),  
288 presenting key insights into particle size and chemical composition—two fundamental  
289 determinants of CCN activity (Ren et al., 2018).

290 As  $\chi$  increases, the peak diameter ( $D_{\text{peak}}$ ) of the particle number size distribution  
291 (PNSD) shifts toward larger sizes (Fig. 5a and Fig. S4), while peak concentrations occur  
292 within the intermediate  $\chi$  range (0.3–0.6). This trend indicates that CN number  
293 concentration ( $N_{\text{CN}}$ ) first increases, driven by primary emissions and new particle  
294 formation, then decreases due to mixing and aging processes (Fig. 5b). Notably, inland  
295 summer  $N_{\text{CN}}$  exhibits a sustained slight increase, linked to frequent new particle  
296 formation events and subsequent particle growth.

297 The critical diameter ( $D_{\text{cri}}$ )—defined as the minimum size for activation at a given



supersaturation—depends on the mass fraction of soluble components (Petters and Kreidenweis, 2007). Using a typical cloud supersaturation of 0.2% as a case study, Fig. 5c shows that  $D_{\text{cri}}$  decreases with increasing soluble species (e.g., sulfate, nitrate) in the inland atmosphere. In contrast, coastal  $D_{\text{cri}}$  exhibits nonlinear variations with  $\chi$ : high external mixing (low  $\chi$ ) elevates  $D_{\text{cri}}$  due to dominant organic components, reducing sea salt particle fractions. As  $\chi$  increases, the mass fraction of non-sea-salt sulfate (nss-sulfate) rises, enhancing activation potential by decreasing  $D_{\text{cri}}$ .



**Fig 5.** Variation of the average particle number size distribution (PNSD) with the mixing state index  $\chi$  (a), variation of the  $N_{\text{CN}}$  with the  $\chi$  (b), variation of the  $D_{\text{cri}}$  and mass fraction of chemical composition with the  $\chi$  (c), variation of the  $N_{\text{CCN}}$  and activation ratio (AR) at  $S=0.2\%$  with the  $\chi$  (d-e).

The dependence of CCN activity at 0.2% supersaturation on mixing state index  $\chi$  reveals distinct inter-atmospheric differences, as shown in Fig. 5d-e. In the inland atmosphere,  $N_{\text{CCN}}$  at  $S=0.2\%$  demonstrates a monotonic increasing trend with  $\chi$ , attributed to the synergistic effects of rising  $N_{\text{CN}}$  and decreasing  $D_{\text{cri}}$  (Fig. S5). By contrast, coastal  $N_{\text{CCN}}$  follows a pattern analogous to  $N_{\text{CN}}$ , with peak concentrations





315 shifting toward higher  $\chi$  values. This highlights the dominant role of particle size effects  
316 in enhancing CCN concentrations under marine-influenced conditions (Perkins et al.,  
317 2022).

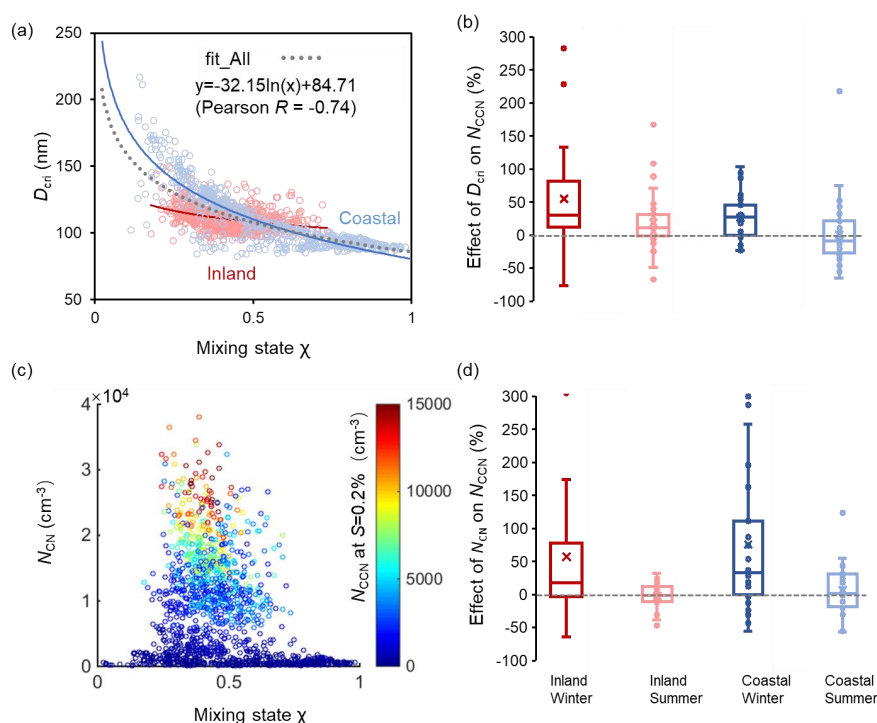
318 Two distinct  $D_{\text{cri}}-\chi$  trends underpin these disparities: one remains stable, driven by  
319 the inherent hygroscopicity of sea salt, while the other exhibits steep  $D_{\text{cri}}$  declines  
320 associated with anthropogenic pollution as internal mixing intensifies. These  
321 discrepancies are further manifested in the nonlinear  $D_{\text{cri}}-\chi$  relationship. The activation  
322 ratio (AR)—quantifying aerosol cloud droplet formation potential at fixed  
323 supersaturation—also varies by site (Fig. 5e). Notably, inland winter AR shows a  
324 marked increase with  $\chi$ , likely due to enhanced  $N_{\text{CCN}}$  from the elevated inorganic  
325 fraction under higher mixing states (Fig. 3). Conversely, the inorganic fraction  
326 decreases during other sampling periods, dampening AR growth.

### 327 **3.3 Impact of the mixing state on the CCN activity**

328 To better interpret the impact of mixing state on CCN concentrations, Fig. 6  
329 quantifies the relative change in  $N_{\text{CCN}}$  at  $S=0.2\%$  as mixing state index  $\chi$  increases,  
330 contextualizing how CN concentration and chemical compositions (i.e.,  $D_{\text{cri}}$ ) evolve  
331 with mixing and aging across particle populations.  $D_{\text{cri}}$  demonstrates heightened  
332 sensitivity to minor  $\chi$  fluctuations at low mixing states ( $\chi < 0.5$ ; Fig. 6a), whereas further  
333 increases in internal mixing (higher  $\chi$ ) exert negligible influence on  $D_{\text{cri}}$  for already  
334 internally mixed particles. This behavior suggests that the  $D_{\text{cri}}-\chi$  relationship may  
335 enable a novel parameterization for  $D_{\text{cri}}$  estimation, a framework that is not yet reported



in prior literature.



337

**Fig 6.** Dependency of the critical diameter ( $D_{\text{cri}}$ ) on the  $\chi$  (a), relative change of CCN number concentration ( $N_{\text{CCN}}$ ) at supersaturation  $S = 0.2\%$  with the reduction in  $D_{\text{cri}}$  (b); Dependency of the CN number concentration ( $N_{\text{CN}}$ ) on the  $\chi$ , different colors represent the  $N_{\text{CCN}}$  (c), relative change of  $N_{\text{CCN}}$  with the change in  $N_{\text{CN}}$  (d).

Coastal aerosol data points (blue dots) span a broad  $D_{\text{cri}}$  range (80–220 nm) with  $\chi$  varying from 0.1 to 1, reflecting alternating influences of highly hygroscopic inorganic salts (sea salt, sulfate) and less-hygroscopic organic matter. In contrast, inland aerosols—dominated by anthropogenic pollutants—exhibit a narrower  $D_{\text{cri}}$  range (90–150 nm). Both environments show negative  $D_{\text{cri}}$ - $\chi$  correlations, but with distinct functional forms: coastal aerosols feature an exceptional logarithmic fit ( $D_{\text{cri}} = -$



348  $42.98\ln(\chi) + 80.36$ ,  $R^2 = 0.75$ ; Fig. 6a blue line), while inland aerosols (red line) yield a  
349 shallower slope ( $-12.04$ ). Pooling all data, we derive a generalized parameterization:  
350  $D_{\text{cri}} = -32.15\ln(\chi) + 84.71$  (Pearson  $r = -0.74$ ,  $R^2 = 0.54$ ).

351 Box plot analyses (Fig. S6) show mixing state reduces  $D_{\text{cri}}$  by 2.2–6.8% across  
352 campaigns, with the steepest winter decline.  $\chi$  impacts on  $N_{\text{CN}}$  differ starkly between  
353 environments: positive effects in polluted inland air (+9%) versus negative effects in  
354 coastal regions (-2%). Inland aerosols, frequently perturbed by primary emissions and  
355 new particle formation, exhibit elevated  $N_{\text{CN}}$  (peaking at  $\chi = 0.2$ – $0.7$ ), while coastal  $N_{\text{CN}}$   
356 remains  $\sim 5000 \text{ cm}^{-3}$  across all  $\chi$ .

357 To isolate the impacts of critical diameter ( $D_{\text{cri}}$ ) and condensation nuclei number  
358 concentration ( $N_{\text{CN}}$ ) on CCN activity, we categorized data into two groups: C1 (particles  
359 within specific  $N_{\text{CN}}$  ranges) evaluates  $N_{\text{CCN}}$  variations driven by  $D_{\text{cri}}$ - $\chi$  relationships,  
360 while C2 (particles within fixed  $D_{\text{cri}}$  intervals) assesses  $N_{\text{CN}}$ - $\chi$  effects (Fig. 6b). Relative  
361 changes (RC) in  $D_{\text{cri}}$ ,  $N_{\text{CN}}$ , and  $N_{\text{CCN}}$  with  $\chi$  were calculated by comparing successive  $\chi$   
362 increments ( $\chi_{i+1}$  vs.  $\chi_i$ ,  $i=0, 0.1 \dots 1$ ) within defined  $N_{\text{CN}}/D_{\text{cri}}$  windows.

363 Notably,  $\chi$  exerts more pronounced effects on  $N_{\text{CCN}}$  for externally mixed aerosols.  
364 For example, coastal winter aerosols (high external mixing;  $\chi_{\text{mean}} = 0.38 \pm 0.12$ ) showed  
365  $N_{\text{CCN}}$  RCs of 23% (C1) and 72% (C2), whereas coastal summer aerosols (high internal  
366 mixing;  $\chi_{\text{mean}} = 0.69 \pm 0.19$ ) exhibited negligible effects (-2.5% in C1, 0.9% in C2). Inland  
367 atmospheres, despite smaller seasonal  $\chi$  variations, showed analogous trends: winter  
368  $N_{\text{CCN}}$  RCs (55% in C1, 57% in C2 for external mixing) exceeded summer values for  
369 more internally mixed populations (Fig. 6d). These results confirm that hygroscopic



370 heterogeneity strongly influences  $N_{CCN}$  under external mixing, aligning with prior work  
371 (Ching et al., 2017).

372 Mixing state impacts on  $N_{CCN}$  are most pronounced during winter in both  
373 environments, attributed to heightened winter  $D_{crit}$  sensitivity to  $\chi$ : a 0.1  $\chi$  increase  
374 reduces  $D_{crit}$  by 5.2% (winter), boosting  $N_{CCN}$  by 39%, versus 2.4%  $D_{crit}$  reduction  
375 (summer) yielding only 6%  $N_{CCN}$  enhancement. Concomitantly, winter  $N_{CN}-\chi$  effects on  
376  $N_{CCN}$  reach 65%, far exceeding summer responses.

377 Contrasting with prior evaluation methods that oversimplify mixing states (Ren et  
378 al., 2018; Xu et al., 2021b), the entropy-based framework adopted herein enables  
379 explicit quantification of CCN activity evolution in response to mixing state transitions.  
380 Inland winter aerosols are presumably shaped by intense urban pollution sources—  
381 including traffic emissions, residential heating, and cooking activities—thereby  
382 enriching the externally mixed particle fraction (Fan et al., 2020; Xie et al., 2020).  
383 Analogously, coastal winter aerosols exhibit dominant external mixing, consisting of  
384 near-hydrophobic and hydrophilic particle mixtures (Xu et al., 2021a). As illustrated in  
385 Fig. S2, winter aerosol populations display bimodal or multimodal  $\kappa$ -PDF distributions,  
386 evidencing high-degree external mixing with chemically diverse compositions. These  
387 results collectively highlight the pivotal role of mixing state heterogeneity in  
388 modulating CCN activity across environments.

#### 389 4. Conclusions

390 The mixing state of aerosol populations undergoes complex transformations  
391 during atmospheric aging, altering the distribution of hygroscopic and non-hygroscopic



392 components and thus influencing CCN activity (Xu et al., 2021a; Ching et al., 2017).  
393 This study derived a mixing state index from field-measured hygroscopicity probability  
394 density functions, systematically investigating its impacts on CCN activity in inland  
395 and coastal environments. Results provide field evidence that aerosol mixing states  
396 generally reside between purely internal and external extremes (Chen et al., 2022),  
397 highlighting a dual regulatory mechanism of mixing state on CCN activity. As  $\chi$   
398 increases, CN number concentrations ( $N_{\text{CN}}$ ) first rise—driven by primary emissions and  
399 new particle formation—then decline due to condensation and coagulation during aging.  
400 Additionally, a logarithmic decreasing relationship between critical diameter ( $D_{\text{cri}}$ ) and  
401  $\chi$  was identified for both inland and coastal particles, parameterized as  $D_{\text{cri}} = -32.15\ln(\chi)$   
402  $+ 84.71$  (Pearson  $R = -0.74$ ,  $R^2 = 0.54$ ). This offers a practical approach to estimate  $D_{\text{cri}}$   
403 from  $\chi$ , serving as a general framework for integrating mixing state effects on CCN  
404 activity in atmospheric models.

405 Entropy-based analyses confirm the pivotal role of mixing state in regulating  $N_{\text{CCN}}$ ,  
406 especially for externally mixed aerosols: a 0.1  $\chi$  increase can enhance  $N_{\text{CCN}}$  by 39–65%.  
407 Current models often oversimplify aerosol mixing states as purely internal or external  
408 (Stevens et al., 2019; Bauer et al., 2013), the latter being particularly sensitive to organic  
409 matter (Ren et al., 2018; Bhattu et al., 2015). Such simplifications introduce significant  
410 biases in  $N_{\text{CCN}}$  estimation (Riemer et al., 2019; Ching et al., 2019). The  $\chi$ - $D_{\text{cri}}$   
411 parameterization proposed here offers a novel approach to reduce model complexity in  
412 representing aerosol hygroscopicity and CCN activation, enabling more accurate  
413 simulations of aerosol CCN capacity. This advancement improves our understanding of



414 aerosol-cloud interactions (IPCC, 2021; Rosenfeld et al., 2019), critical for refining  
415 climate effect assessments.

#### 416 **Data availability**

417 All data used in the study are available at <https://doi.org/10.3974/geodb.2019.06.11.V1>  
418 (Fan et al., 2019) and <http://doi.org/10.17632/3dx6pnx869.1> (Xu et al., 2021a).

#### 419 **Author contributions**

420 RH and JR conceived the conceptual development of the paper. JR, FZ and WX directed  
421 and performed the experiments with YW and LC. FZ provided the dataset in the inland  
422 site. JO, DC and CO provided the dataset in the coastal site. JR conducted the data  
423 analysis and wrote the draft. All authors edited and commented on the various sections  
424 of the paper.

#### 425 **Competing interests**

426 The contact author has declared that none of the authors has any competing interests.

#### 427 **Supporting Information**

428 Additional analysis results that were applied in this study. Sensitivity of the  
429 hygroscopic parameter for the group of the hygroscopic species on the mixing state  
430 index  $\chi$  (Figure S1), mean values of the  $\kappa$ -PDF for aerosols of five diameters (Figure  
431 S2), time series of the average per-particle species diversity  $D\alpha$ , the bulk population  
432 species diversity  $D\gamma$ , and their affine ratio  $\chi$  (Figure S3), variation of the peak diameter  
433 ( $D_{\text{peak}}$ ) with the mixing state index (Figure S4), diurnal variation of  $\chi$  and CN



434 concentration during winter and summer periods for 40 nm and 150 nm aerosols in  
435 inland and for 35 nm and 165 nm aerosols in coastal site (Figure S5), relative change  
436 of the critical diameter and CN concentration with the mixing state index  $\chi$  (Figure S6)  
437 (PDF).

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